

Dynamics of riverine CO₂ in the Yangtze River fluvial network and their implications for carbon evasion

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Abstract: Understanding riverine carbon dynamics is critical for not only better estimates of various carbon fluxes but also evaluating their significance in the global carbon budget. As an important pathway of global land-ocean carbon exchange, the Yangtze River has received less attention regarding its vertical carbon evasion compared with lateral transport. Using long-term water chemistry data, we calculated CO₂ partial pressure ($p\text{CO}_2$) from pH and alkalinity and examined its spatial and temporal dynamics and the impacts of environmental settings. With alkalinity ranging from 415 to >3400 $\mu\text{eq L}^{-1}$, the river waters were supersaturated with dissolved CO₂, generally 2–20 folds the atmospheric equilibrium (i.e., 390 μatm). Changes of $p\text{CO}_2$ were collectively controlled by carbon inputs from terrestrial ecosystems, hydrological regime, and rock weathering. High $p\text{CO}_2$ values were observed spatially in catchments with abundant carbonate presence and seasonally in the wet season when recent-fixed organic matter was exported into the river network. In-stream processing of organic matter facilitated CO₂ production and sustained the high $p\text{CO}_2$, although the alkalinity presented an apparent dilution effect with water discharge. The decreasing $p\text{CO}_2$ from the smallest headwater streams through tributaries to the mainstem channel illustrates the significance of direct terrestrial carbon inputs in controlling riverine CO₂. With a basin-wide mean $p\text{CO}_2$ of 2662 ± 1240 μatm , substantial CO₂

evasion from the Yangtze River fluvial network is expected. Future research efforts are needed to quantify the amount of CO₂ evasion and assess its biogeochemical implications for watershed-scale carbon cycle. In view of the Yangtze River's relative importance in global carbon export, its CO₂ evasion would be significant for global carbon budget.

30 **Keywords:** CO₂ partial pressure ($p\text{CO}_2$); riverine carbon cycle; spatial and temporal patterns; CO₂ evasion; Yangtze River

1. Introduction

Inland waters, including rivers, streams, lakes, wetland, and reservoirs, have recently been
35 recognized as active components of the global carbon cycle, transporting, storing, and processing huge amounts of terrestrially-derived carbon (Aufdenkampe et al., 2011; Cole et al., 2007; Raymond et al., 2013; Richey et al., 2002; Weyhenmeyer et al., 2015; Borges et al., 2015). With a higher CO₂ partial pressure ($p\text{CO}_2$) than the atmospheric equilibrium (i.e., 390 μatm), inland waters are mostly net carbon sources to the atmosphere. Published studies show that the
40 annually degassed CO₂ from inland waters is estimated to almost entirely compensate the total annual carbon uptake by ocean systems (Wanninkhof et al., 2013; Regnier et al., 2013). Global estimates of CO₂ evasion from rivers and streams range from 0.56 to 1.8 PgC yr⁻¹ (Aufdenkampe et al., 2011; Raymond et al., 2013; Lauerwald et al., 2015). It is apparent that these results vary considerably and are associated with great uncertainties. The most recent estimate of 0.65 PgC
45 yr⁻¹ by Lauerwald et al. (2015) accounts for only 36% of the efflux estimated by Raymond et al. (2013). While both studies have used the same hydrochemical database (GloRiCh), it should be

noted that Raymond et al. (2013) used all the calculated $p\text{CO}_2$ values whereas Lauerwald et al. (2015) used only 18% of the sampling locations.

Among the numerous factors contributing to current CO_2 evasion uncertainties, a principal reason is the absence of a spatially explicit $p\text{CO}_2$ data set that covers the full spectrum of the global river and stream network.

Existing global maps of CO_2 evasion from fluvial network are typically generated on the basis of incomplete spatial coverage of $p\text{CO}_2$, in which Asian rivers are heavily underrepresented (e.g., Aufdenkampe et al., 2011; Battin et al., 2009; Lauerwald et al., 2015; Raymond et al., 2013).

Due to lack of direct *in situ* measurements, simplified extrapolation is normally used to predict $p\text{CO}_2$ in and CO_2 evasion from Asian river systems. Consequently, the estimation accuracy is problematic and even erroneous. For example, for the Yellow River in East Asia, while the calculated $p\text{CO}_2$ from river water chemistry is $2800 \mu\text{atm}$ (Ran et al., 2015a), the modeled $p\text{CO}_2$ by Lauerwald et al. (2015) is 30% lower (i.e., $<2000 \mu\text{atm}$). A much lower estimate of <700

μatm can be derived from the $p\text{CO}_2$ map produced in Raymond et al. (2013). Such great discrepancies are largely because riverine $p\text{CO}_2$ is highly site-specific and affected by a wide range of environmental factors (e.g., Abril et al. 2015; Teodoru et al., 2015).

Asian rivers are significant contributors to global carbon flux as a result of high soil erosion and particulate organic carbon export, accounting for 40% of the global carbon flux from land to sea (Schlünz

and Schneider, 2000; Hope et al., 1994). Estimating the amount of CO_2 degassed from Asian rivers is critical for global CO_2 evasion assessments. Recent work in Mekong and Yellow rivers has demonstrated high $p\text{CO}_2$ and CO_2 effluxes (Alin et al., 2011; Ran et al., 2015b), further

highlighting the necessity of incorporating the currently underrepresented Asian rivers into global carbon budget assessments.

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As an important carbon contributor to the West Pacific Ocean, the Yangtze River has received widespread attention in fluvial carbon export at various spatial and temporal scales. Studies of flux estimates of different carbon species date back to the early 1980s (Cauwet and Mackenzie, 1993; Gan et al., 1983; Milliman et al., 1984; Wang et al., 2012; Zhang et al., 2014; Ittekkot, 1988).

75 Intensive observations covering seasonal variability show that the Yangtze River transports approximately 20 Mt of carbon per year into the oceans (Wu et al., 2007; Bao et al., 2015).

Contrary to the long history of lateral export measurements, however, few studies have examined the vertical carbon exchange between the river system and the atmosphere (Li et al., 2012; Zhao et al., 2013; Chen et al., 2008). This is by nature largely due to the differences in sampling

80 strategy. Unlike the lateral export that only involves measurements on the mainstem or at specific sites near the river mouth, quantifying basin-wide CO₂ evasion requires a spatially explicit *p*CO₂ data set encompassing the entire fluvial network. Any attempts of using limited local measurements to up-scale to the watershed scale are challenging and subject to large uncertainties. This has impacted the understanding of the riverine carbon cycle within the

85 Yangtze River watershed as well as its links to the atmosphere and ocean systems.

By using long-term water chemistry data measured in the Yangtze River watershed, we calculated the riverine *p*CO₂ from pH and alkalinity. In combination with hydrologic and

geologic information, the objectives of this study were to 1) investigate the spatial and temporal
90 patterns of $p\text{CO}_2$ under ‘natural’ processes before significant human perturbations, mainly dam
impoundment and land-use change since the 1990s; 2) to explore the couplings between $p\text{CO}_2$
and environmental settings by investigating environmental and geomorphologic controls. Based
on the obtained $p\text{CO}_2$, we further evaluated its implications for CO_2 evasion. In view of the
Yangtze River’s role in global fluvial export of water, sediment, and carbon (Syvitski et al.,
95 2005; Wang et al., 2012), its contribution to the global CO_2 evasion from river systems is likely
significant. This $p\text{CO}_2$ database is thus helpful to examine the spatial distribution of global
riverine $p\text{CO}_2$ and to refine estimates of global CO_2 evasion.

2. Material and methods

100 2.1 The Yangtze River basin

With a length of 6380 km, the Yangtze River is the longest river in China and the third longest in
the world. The river originates on the Tibetan Plateau and flows eastward through the Sichuan
Basin and the Middle-Lower Reach Plains, before emptying into the East China Sea (Fig. 1a). Its
drainage area is 1.81 million km^2 . The Yangtze River basin is mainly overlain by sedimentary
105 rocks that are composed of marine carbonates, evaporites, and continental deposits. Carbonate
sedimentary rocks are widely distributed within the watershed and are particularly abundant in
the Wujiang, Yuanjiang, and Hanjiang tributary catchments (Fig. 1b). Siliciclastic sedimentary
rocks are also widely present in the basin while metamorphic rocks are mainly scattered in the
middle-lower reach (Fig. 1b). The Yangtze River is joined by a number of large tributaries,

including the Yalongjiang, Daduhe, Minjiang, Jialingjiang, Wujiang, Yuanjiang, Xiangjiang, Hanjiang, and Ganjiang rivers (Fig. 1a).

Figure 1

Except the headwater region characterized by high elevation and cold climate (annual mean temperature $<4^{\circ}\text{C}$), the remaining watershed is affected by a subtropical monsoon climate with the annual mean temperature in the middle-lower reach varying from 16 to 18°C (Chen et al., 2002). Rainfall is the major source of water discharge, whereas snowfall supply is only significant in the ice-covered upstream mountainous areas. With a mean precipitation of 1100 mm yr^{-1} , the precipitation is spatially highly variable, decreasing from 1644 mm yr^{-1} in the lower reach, to 1396 mm yr^{-1} in the middle reach, and 435 mm yr^{-1} in the upper reach (Chetelat et al., 2008). Approximately 60% of the annual precipitation falls during the wet season from June to September. Affected by summer monsoon, the wet season generally occurs earlier in the middle and lower reaches than in the inland upper reach. Water discharge from the upper to the lower reach presents a strong seasonal variability (Fig. 2). Monthly peak discharge occurs in July and can be 5–7 times greater than the lowest discharge in the dry season (October to May). The mean discharge at Datong station is $28,200\text{ m}^3\text{ s}^{-1}$ (see its location in Fig. 3b), and consequently the Yangtze River annually discharges 889 km^3 of water into the ocean (Yang et al., 2002).

Figure 2

2.2 Water chemistry data

Concentrations of alkalinity, major ions, and dissolved silica measured at 359 stations in the Yangtze River watershed (Fig. 1a) during the period 1960s–1985 were retrieved from the

Hydrological Yearbooks, which were yearly produced by the Yangtze River Conservancy Commission (YRCC) for internal use. Concomitant environmental variables measured at each sampling event, including pH, water temperature, and discharge, were also extracted from the yearbooks. The water samples for pH and temperature measurement were taken in the same period as these for ion analysis. The sampling frequency ranged from 1 to 14 times per month depending on flow conditions. While sampling at some stations during 1966–1975 was less frequent, ~80% of the 359 stations have been continuously sampled for at least 10 years, starting from the early 1970s. To avoid severe river pollution by human activity, only the samples collected prior to 1985 were used. In addition, samples with a pH lower than 6.5 were manually discarded (498 measurements; predominantly in the lower reach) because the calculated $p\text{CO}_2$ would be greatly biased due to contributions of noncarbonated alkalinity such as organic acid anions (Abril et al., 2015; Hunt et al., 2011). Because reservoir trapping and increased water residence time can remarkably alter the physical and biogeochemical properties of running water (Kemenes et al., 2011; Barros et al., 2011), the stations located inside or shortly below reservoirs were also intentionally removed. Given the tidal influences, mainstem stations downstream of Datong, 626 km inland from the coast, were also excluded, as were the stations in the delta region that were affected either by tides or by intersections with other rivers via artificial canals. Based on these selection criteria, 339 stations, including 13 mainstem stations and 326 tributary stations, were retained and 47,809 water chemistry measurements in total were compiled. The discarded samples due to $\text{pH} < 6.5$ accounted for approximately 1% of the considered measurements. No sampling station was excluded solely because it had $\text{pH} < 6.5$ samples only.

Chemical analyses of water samples were performed under the authority of YRCC following the standard procedures and protocols described by Alekin et al. (1973) and the American Public Health Association (1985). While pH and temperature were measured in the field, the alkalinity was determined by acid titration. Detailed sampling and analysis procedures were presented in Chen et al. (2002). One important issue regarding historical records is data reliability. No assessment reports on quality assurance and quality control are available in the hydrological yearbooks. An effective evaluation approach is to compare the hydro-chemical differences for samples collected at the same station but by different agencies. The Wuhan station on the Yangtze mainstem has also been monitored under the United Nations GEMS/Water Programme since 1980 (only yearly means available at <http://www.unep.org/gemswater>). The pH value from the yearbooks agreed well with that measured by the GEMS/Water Programme with <1.8% differences, while the alkalinity discrepancy between the two data sets is larger (Table 1). The yearbooks report a slightly higher alkalinity than the GEMS/Water Programme results by 7.6–13.9%, indicating that the yearbook reports are reliable for $p\text{CO}_2$ calculation. High data quality of the yearbook reports can also be validated from comparison of major dissolved elements measured by the two agencies at Wuhan station (see Chen et al., 2002).

Table 1

2.3 Calculation of $p\text{CO}_2$

The conventional method of calculating $p\text{CO}_2$ from pH and alkalinity was used. With ~90% of the pH values ranging from 7.1 to 8.3 suggestive of natural process for the Yangtze River,

bicarbonates were assumed equivalent to alkalinity (Amiotte-Suchet et al., 2003), accounting for 96% of the total alkalinity. As a result of low dissolved organic carbon (i.e., $<250\ \mu\text{M}$; Liu et al., 2016), impact of organic acids on alkalinity is predicted to be small. The $p\text{CO}_2$ was then calculated using CO2SYS program (Lewis and Wallace, 1998). However, using this method would produce biased extreme values that are unrealistic in natural river systems (Hunt et al., 2011; Weyhenmeyer et al., 2015). We thus reported median values per sampling station instead of means to avoid the impact of erroneous extreme results. The results were summarized in the Supplement (Table S1).

3. Results

3.1 Spatio-temporal variability of alkalinity and $p\text{CO}_2$

Except the excluded measurements, pH in the Yangtze River waters varied from 6.5 to 9.2 with 96% of the pH measurements ranging from 7.3 to 8.3 (Table 2). Higher pH values (i.e., >7.8) were spatially measured in the headwater streams and the Hanjiang catchments (see Fig. 1a for location). In comparison, the tributaries in the southern part of the watershed exhibited relatively low pH values. For the mainstem channel (Table 2), the median pH showed a significant downstream decrease from 8.29 to 7.55 ($r^2 = 0.77$; $p < 0.001$). The alkalinity varied from 415 to $>3400\ \mu\text{eq L}^{-1}$ (Fig. 3a). Higher alkalinity (i.e., $>2500\ \mu\text{eq L}^{-1}$) was observed in the upper reach and the upper part of the middle reach (Fig. 3a), in particular the carbonate-rich tributary catchments (e.g., the Jialingjiang, Wujiang, and Hanjiang rivers). In contrast, the lower part of

the middle reach (mainly the Ganjiang River) and the lower reach showed a lower alkalinity of $<2000 \mu\text{eq L}^{-1}$. The average alkalinity over the whole watershed was $2210 \pm 1023 \mu\text{eq L}^{-1}$.

Figure 3 and Table 2

The calculated $p\text{CO}_2$ varied by a magnitude of 2 with the highest $p\text{CO}_2$ being $24,432 \mu\text{atm}$. At 95% of the stations, the $p\text{CO}_2$ was higher than $1000 \mu\text{atm}$, generally 2–20 folds the atmospheric $p\text{CO}_2$. Only one station in the upper reach showed a median $p\text{CO}_2$ lower than the atmosphere. In the mainstem, the $p\text{CO}_2$ increased from $\sim 700 \mu\text{atm}$ at the uppermost station to $3800 \mu\text{atm}$ at Nanjing near the river mouth (Table 2). Averaged over all stations, the basin-wide $p\text{CO}_2$ was $2662 \pm 1240 \mu\text{atm}$. To better illustrate its spatial variability, we modeled the $p\text{CO}_2$ for the whole stream network using the Kriging interpolation method in ArcGIS 10.1 (Esri, USA) with the assumption that the station-based $p\text{CO}_2$ was representative of the surrounding streams. Similar to alkalinity, the $p\text{CO}_2$ presented significant spatial variations (Fig. 3b). The Yangtze mainstem near the headwater region and the Yalongjiang catchment showed the lowest $p\text{CO}_2$, generally $<1000 \mu\text{atm}$. In comparison, the carbonate-rich tributaries in the southern part of the watershed had high $p\text{CO}_2$ values. With carbonates occupying 83% of the catchment, the Wujiang River presented the highest median $p\text{CO}_2$ than other tributaries, averaging $3550 \pm 1356 \mu\text{atm}$. In the lower reach, the $p\text{CO}_2$ was $3988 \pm 1244 \mu\text{atm}$ on average, which is inconsistent with its relatively low alkalinity of $<2000 \mu\text{eq L}^{-1}$ (Fig. 3a). It is worth noting that the $p\text{CO}_2$ in Hanjiang catchment was lower than expected, given its high alkalinity ($>2500 \mu\text{eq L}^{-1}$). Differences in pH in these catchments are likely a principal cause of these inconsistencies.

In addition, the $p\text{CO}_2$ also showed strong temporal variability. Fig. 4 presents an example of $p\text{CO}_2$ changes at Datong station on the mainstem channel. Despite considerable inter-annual variations that could change by a factor of 5, the annual $p\text{CO}_2$ declined steadily during the >20-year-long sampling period ($r^2 = 0.18$; $p < 0.05$) (Fig. 4a). This trend is pronounced even if the anomalously high values in the late 1960s are excluded. Indeed, more than half of the evaluated stations, mainly in the middle-lower reach, showed a significant decreasing trend at the 95% confidence level. In contrast, gradual increases were observed at some tributary stations in the upper reaches. Seasonally, the $p\text{CO}_2$ in the wet season was on average 30% higher than that in the dry season (Fig. 4b), and greater fluctuation ranges could be observed in the wet season.

Figure 4

3.2 Correlations with hydro-geochemical variables

Fig. 5 presents two representative examples showing responses of alkalinity and $p\text{CO}_2$ to hydrological regimes. Changes of alkalinity at both stations reflected a clear dilution effect. High alkalinity concentrations were measured in low flow periods when groundwater was the major contributor to runoff (Figs. 5a and 5c). Checking all stations indicated that the alkalinity at 98% of the stations decreased exponentially with increasing water discharge after the onset of the wet season. In contrast, the $p\text{CO}_2$ presented diverse relationships with water changes (Figs. 5b and 5d). There was no discernible dependence of $p\text{CO}_2$ on flow in the mainstem, while a positive correlation was widely observed in small tributaries. Although only two stations were plotted

here, these diverse responses of alkalinity and $p\text{CO}_2$ to flow changes were widespread within the watershed, in particular for $p\text{CO}_2$ between mainstem and small tributaries.

Figure 5

In order to elucidate the impacts of rock weathering on $p\text{CO}_2$, we selected three typical tributary catchments with differing rock compositions (Table 3). The Wujiang catchment is mainly underlain by carbonate sedimentary rocks (83%) and the Ganjiang catchment by siliciclastic sedimentary rocks (65%), whereas the Jialingjiang catchment lies in the middle regarding the areal coverage of the two rocks (Table 3 and Fig. 1b). As the most typical weathering products of carbonate and siliciclastic sedimentary rocks, we plotted Ca^{2+} and dissolved silica (expressed as SiO_2) against $p\text{CO}_2$, respectively (Fig. 6). For the three catchments with contrasting rock compositions, the $p\text{CO}_2$ showed different responses to Ca^{2+} and SiO_2 . In Wujiang catchment, the log-transformed $p\text{CO}_2$ (i.e., $\lg(p\text{CO}_2)$) presented a significant negative correlation with Ca^{2+} concentration ($p < 0.001$) (Fig. 6). This negative correlation became less apparent with decreasing carbonate coverage in Jialingjiang and Ganjiang catchments. In contrast, while the $\lg(p\text{CO}_2)$ exhibited a positive correlation with SiO_2 in Jialingjiang and Ganjiang catchments characterized by high coverage of siliciclastic sedimentary rocks, no clear relation between $\lg(p\text{CO}_2)$ and SiO_2 was detected in Wujiang catchment (Fig. 6). However, when plotting $p\text{CO}_2$ against Ca^{2+} and SiO_2 for the entire Yangtze River watershed, there was no discernable relationship between $p\text{CO}_2$ and both variables (Fig. S1 in the Supplement).

Figure 6 and Table 3

4. Discussion

4.1 Uncertainty analysis of $p\text{CO}_2$

255 As an important parameter for CO_2 evasion estimation, an accurate riverine $p\text{CO}_2$ is essential to quantify CO_2 evasion and explore its biogeochemical implications for carbon cycle at different scales. Compared with direct measurement by means of membrane equilibration or headspace technique, the conventional $p\text{CO}_2$ calculation from alkalinity has been criticized for causing biases (Long et al., 2015; Hunt et al., 2011). Huge overestimations (i.e., >100%) have been
260 reported in rivers with organic-rich and acidic waters due to combined effects of high organic acids and low buffering capacity of carbonate systems at low pH (Abril et al., 2015).

Unfortunately, there were no organic carbon information in the yearbooks, and measurements of dissolved organic carbon (DOC) in the Yangtze River started in the early 1980s. Its DOC ranging from 130 to 180 μM was relatively low compared with other major world rivers (Bao et
265 al., 2015; Wang et al., 2012). Our recent sampling also shows that the mean DOC is 160 μM for the mainstem and 200 μM for major tributaries (Liu et al., 2016). Given the neutral to basic pH range and the alkalinity variations, we believe the impact of organic acids is minimal, although a slight overestimation may have occurred as suggested by Abril et al. (2015). Our recent $p\text{CO}_2$ measurements in the mainstem and major tributaries using a membrane contactor (Qubit DCO_2
270 System, Qubit Biology Inc., Canada) also indicate that the calculated $p\text{CO}_2$ results are consistent with the measured values with only ~8% differences (Liu et al., 2016).

Furthermore, this $p\text{CO}_2$ calculation method is sensitive to pH changes. High accuracy of pH measurements is critical to reduce the associated uncertainty. Similar to other water chemistry

records (i.e., Butman and Raymond, 2011; Lauerwald et al., 2015; Weyhenmeyer et al., 2015), the retrieved pH was reported with a precision of one decimal place. If the uncertainties in pH measurement accuracy are assumed to 0.1 pH units, the calculated $p\text{CO}_2$ would be underestimated by 26% or overestimated by 21%. To minimize human-induced disturbances in the chemical equilibrium of natural waters, we excluded the samples with $\text{pH} < 6.5$ and treated them as being significantly polluted. This arbitrary exclusion may have generated biased estimates of $p\text{CO}_2$ for the whole river network in general and some natural rivers characteristic of low pH in particular (Wallin et al., 2014). Considering the higher alkalinity than the GEMS/Water Programme results, the propagated uncertainty ranges from 14% (underestimation) to 27% (overestimation). As China's major industrial and agricultural regions, impact of human activity within the catchment, including sewage inputs and chemical fertilizer usage to a lesser extent, may have altered its chemical compositions and pH. In view of the small number of discarded measurements (1% of the total) and the high buffering capacity of carbonate alkalinity and low DOC contents, the calculated $p\text{CO}_2$ is reasonable and can be used for further CO_2 evasion estimation.

4.2 Environmental impacts on alkalinity and $p\text{CO}_2$

Export of alkalinity in river systems was affected by hydrological regime with a clear dilution effect (Fig. 5). The average alkalinity was 35% lower in the wet season than in the dry season. In both the mainstem and the tributaries, the higher alkalinity during low flow periods in the dry season (Figs. 5a and 5c) illustrated the contribution of groundwater recharge in providing

abundant alkalinity. With widespread carbonate presence, groundwater in the Yangtze River watershed was rich in dissolved inorganic carbon (DIC). Recent studies show that the alkalinity of typical karst groundwater in the watershed is in the range of 3300–4200 $\mu\text{eq L}^{-1}$ (Li et al., 2010b; Li et al., 2010a). With reduced relative contribution of groundwater in the wet season, the high alkalinity was diluted by local rain events that carried lower DIC contents. Spatially, the dilution effect was more pronounced in the upper reach than the middle-lower reach. This may have revealed the response of alkalinity production to land cover. Catchments with a higher forest cover normally exhibit a stronger dilution effect than cropland catchments (Raymond and Cole, 2003). While cropland was the major land use type in the middle-lower reach accounting for 53.5% of the total catchment area, forest cover in the upper Yangtze River watershed was much higher (37.3%) than the middle-lower reach (30.4%; data are from Data Center for Resources and Environmental Sciences for the 1980s).

Riverine dissolved CO_2 originates primarily from terrestrial ecosystem respiration, groundwater input, and in-stream processing of land-derived organic matter (Wallin et al., 2013; Lynch et al., 2010). Different from alkalinity showing a clear dilution effect, the stable $p\text{CO}_2$ in the Yangtze mainstem likely reflected the impact of different biogeochemical processes (Fig. 5b). Compared to the dry season in which the $p\text{CO}_2$ was mainly controlled by DIC inputs from groundwater, the elevated $p\text{CO}_2$ in the wet season suggested the influence of organic carbon transport and decomposition. Owing to strong erosion and leaching of recent-fixed organic matter, its organic carbon content in the wet season is significantly higher and the age much younger (Wang et al.,

2012;Zhang et al., 2014). Rapid mineralization of the labile fraction of organic carbon can increase the $p\text{CO}_2$. For instance, approximately 60% of the recent-fixed carbon entering the Yangtze River can be quickly degraded in the wet season, while the degradation ratio in the dry
320 season is only 31% (Wang et al., 2012). On the other hand, the increasing $p\text{CO}_2$ with flow in tributaries indicated enhanced supply of fresh dissolved CO_2 during high flow periods (Fig. 5d). For tributaries with more homogeneous catchment environments, decomposition of soil organic matter can provide abundant dissolved CO_2 (Liu et al., 2016;Li et al., 2012), generating a positive $p\text{CO}_2$ response to flow changes. Presence of wetlands and floodplains also affects river
325 biogeochemistry (Teodoru et al., 2015). Affected by dam impoundment, the catchment upstream of Yunxian station is characteristic of widespread wetlands and floodplains. Consequently, the enhanced connectivity between river and wetlands/floodplains along aquatic continuum, especially during wet seasons, may have maintained the high $p\text{CO}_2$ levels (Abril et al., 2014). For $p\text{CO}_2$ in the mainstem, it is likely because the increased dissolved CO_2 inputs by soil organic
330 matter decomposition from one region has been counteracted by low $p\text{CO}_2$ waters derived from other regions. This is highly possible given the heterogeneous catchment setting in terms of vegetation cover, soil type, and rainfall intensity. Furthermore, the large catchment implies a long travel time of land-derived organic carbon during fluvial delivery (3–5 months). Coupled with limited floodplains along the mainstem channel (see discussion below), direct inputs of CO_2
335 from soil respiration would be relatively low whereas strong CO_2 evasion in lower-order turbulent tributaries might have already exhausted dissolved CO_2 . Therefore, its $p\text{CO}_2$ dynamics appeared to be independent of hydrograph.

The spatial distribution of alkalinity overlapped well with the outcrops of carbonate sedimentary rocks (Figs. 1b and 3a), with ~60% of the high alkalinity concentrations measured in carbonate catchments. Using Ca^{2+} as a proxy of rock weathering, the strong correlation between Ca^{2+} and alkalinity suggested the dominant role of weathering in controlling alkalinity and DIC export (Fig. 7). This is consistent with the significant impact of weathering on alkalinity as observed in other rivers (Raymond and Cole, 2003; Humborg et al., 2010). Particularly, given the higher susceptibility of carbonates to weathering than silicates (Goudie and Viles, 2012), the abundant carbonate presence in Wujiang catchment helped to sustain its high alkalinity and $p\text{CO}_2$ (Table 3). However, the negative correlation in Fig. 6a is contradictory to the common belief that carbonate dissolution will likely cause an elevated $p\text{CO}_2$ (Marcé et al., 2015; Teodoru et al., 2015). Given the significant correlation between Ca^{2+} and alkalinity, the decreasing $p\text{CO}_2$ with increasing Ca^{2+} is probably due to pH variability that may have offset the impact of weathering-induced DIC inputs in controlling $p\text{CO}_2$ (Fig. S2). A slight pH increase would result in a reduced $p\text{CO}_2$ as this calculation method is sensitive to pH fluctuations (Laruelle et al., 2013). The positive correlation between $p\text{CO}_2$ and SiO_2 in Jialingjiang and Ganjiang catchments demonstrated the impact of DIC export by silicate weathering. Despite the high silicate weathering rate in Ganjiang catchment, its alkalinity represented only one third of that in the other two catchments (Table 3). Apparently, its high $p\text{CO}_2$ of $2642 \pm 626 \mu\text{atm}$ was primarily due to its low pH (~6% lower). Overall, the catchments with more carbonate presence presented higher $p\text{CO}_2$ values (Figs. 1 and 3b). Because weathering products are typical for groundwater,

this also suggests that riverine $p\text{CO}_2$ has a strong groundwater signature. Different from the positive response of $p\text{CO}_2$ to discharge at Yunxian station reflecting the importance of connectivity between river and wetlands/floodplains (Fig. 5d), the decreasing $p\text{CO}_2$ at Xiajiang station with discharge is indicative of the impact of groundwater input on riverine carbon dynamics (Figs. S3 and 6f). Particularly, in dry seasons with groundwater dominating the runoff, the SiO_2 can explain ~25% of the $p\text{CO}_2$ variability in the sub-catchments covered mainly with siliciclastic sediment rocks, comparable to the results by Humborg et al. (2010) in Sweden. The indiscernible $p\text{CO}_2\text{-Ca}^{2+}$ and $p\text{CO}_2\text{-SiO}_2$ relationship for the entire watershed may be attributed to the spatial heterogeneity in lithology that has obscured the signature (Fig. S1). While both positive and negative relationships existed in sub-catchments with predominant carbonate or siliciclastic sediment rocks (Fig. 6), these relationships may have counteracted each other when all data points were plotted together.

Figure 7

Because $p\text{CO}_2$ was calculated from alkalinity, its spatial variability reflected largely the export of the latter. The inconsistencies between $p\text{CO}_2$ and alkalinity in Hanjiang catchment were likely caused by dam operation (Fig. 3). By altering the physical and biogeochemical properties of flowing water, dam trapping could cause a greatly declined $p\text{CO}_2$ as a result of photosynthetic CO_2 fixation and increased pH (Ran et al., 2015a). The Danjiangkou Reservoir (storage: 17.5 km^3) on the upper Hanjiang River was constructed in 1968. Unfortunately, the retrieved data for the Hanjiang River started from the 1970s, rendering it impossible to compare the $p\text{CO}_2$ differences between pre- and post-dam periods. An indirect evidence is that an elevated pH

within the reservoir has been measured (7.95–8.33; Li et al., 2009) relative to the 1970s (7.84±0.15). In the lower reach near the estuary (Fig. 3b), more pronounced net-heterotrophy and human activity could explain its high $p\text{CO}_2$. Settling down of particulate organic matter coupled with nutrient-rich water plume from offshore can accelerate CO_2 production. Chen et al. (2008) concluded that aerobic respiration of heterotrophic ecosystems was the primary determinant of the high $p\text{CO}_2$ in the inner Yangtze estuary. Moreover, the lower Yangtze River watershed was highly populated. Inputs of acids from agricultural fertilizer, sewage, and acid deposition have also decreased pH and shifted the carbonate system towards CO_2 (Duan et al., 2007; Chen et al., 2002), generating high $p\text{CO}_2$ values regardless of its relatively low alkalinity.

4.3 Geomorphological controls on alkalinity and $p\text{CO}_2$

To illustrate the geomorphological controls, the used 339 stations were aggregated by stream order based on their spatial positions. Both alkalinity and $p\text{CO}_2$ showed a decreasing trend from the smallest headwater streams through tributaries to the Yangtze mainstem (Fig. 8). The average decrease of alkalinity and $p\text{CO}_2$ were 94 $\mu\text{eq L}^{-1}$ and 266 μatm , respectively. Higher alkalinity and $p\text{CO}_2$ in the headwater streams reveal the significance of direct terrestrial inputs of organic carbon and dissolved CO_2 in controlling riverine carbon cycle. Over the study period, the Yangtze River watershed suffered severe soil erosion, averaging 2167 $\text{t km}^{-2} \text{ yr}^{-1}$ (Wang et al., 2007b). Huge amounts of carbon were transported into the river system via erosion (Wu et al., 2007). Decomposition of the terrestrial-origin organic carbon has resulted in the CO_2 excess in the headwater streams (Li et al., 2012).

Figure 8

The decreasing $p\text{CO}_2$ with increasing stream order imply continued CO_2 evasion along the river continuum and reduced supply of fresh CO_2 . Except the three lakes connected to the mainstem (Fig. 1a), the Yangtze River network is largely confined to its channel. Without large floodplains supplying labile organic matter to sustain high $p\text{CO}_2$ as in the Amazon River (Mayorga et al., 2005), its $p\text{CO}_2$ decreased progressively from the headwaters towards the mainstem channel. In addition, it is interesting to note that the $p\text{CO}_2$ in the highest three orders was equivalent ($\sim 1800 \mu\text{atm}$; Fig. 8). Instead of continuous decline, the stable $p\text{CO}_2$ suggests a balance between CO_2 evasion and supply of fresh CO_2 from upstream catchments or aquatic respiration. Contrary to the headwater streams with close contact with terrestrial ecosystems, the downstream large streams and rivers are far away from rapid fresh CO_2 input. Moreover, these large streams and rivers are generally characterized by comparatively low gas transfer velocities owing to weakened turbulence and mixing with benthic substrates (Butman and Raymond, 2011; Borges et al., 2015), which can effectively inhibit CO_2 degassing and therefore maintain the balance.

It is important to note, however, that the delineated 8 stream orders may not necessarily represent the actual stream network. Limited by spatial resolution, the smallest headwater streams might have been missed from the identified river network. In addition, these headwater streams are also generally absent of sampling stations. With much closer biogeochemical interactions with land ecosystems, these missed headwater streams tend to have higher $p\text{CO}_2$ (Benstead and Leigh,

2012;Aufdenkampe et al., 2011;Butman and Raymond, 2011). Thus, the actual $p\text{CO}_2$ gradient along the stream order may be sharper if a higher $p\text{CO}_2$ in the headwater streams is included.

4.4 Implications for riverine CO_2 evasion

425 As mentioned earlier, riverine carbon transport has been a significant component of carbon cycle. Quantifying riverine carbon export is essential to better evaluate global carbon budget and elucidate the magnitude of carbon exchange between different carbon pools. For the estimation of CO_2 evasion, riverine $p\text{CO}_2$ denotes CO_2 concentration gradient across the water-air interface and thus the potential of CO_2 exchange. Prior studies indicate that elevated riverine $p\text{CO}_2$ can
430 enhance CO_2 evasion due to a steeper concentration gradient and a greater CO_2 availability for degassing (Long et al., 2015;Billett and Moore, 2008). When assessing global-scale CO_2 evasion, however, the spatial distribution of $p\text{CO}_2$ is heavily skewed towards Northern America, Europe, and Australia (e.g., Lauerwald et al., 2015; Raymond et al., 2013), while data for Asian rivers are extremely lacking. This absence of an equally distributed $p\text{CO}_2$ database has made it challenging
435 to accurately estimate global CO_2 evasion. The role of Asian rivers in global carbon export explicitly demonstrates that under-representation of Asian rivers would cause huge biases.

Comparing the Yangtze River with other rivers shows that its $p\text{CO}_2$ is higher than most world rivers (Table 4). The average $p\text{CO}_2$ of 2662 μatm suggests that the Yangtze River waters are
440 potentially a prominent carbon source for the atmosphere. Large CO_2 evasion fluxes have been reported by several small-scale studies in the upper reach and the estuary (Zhai et al., 2007;Chen

et al., 2008;Li et al., 2012), as also shown in Table 4. Nonetheless, a systematic estimation of CO₂ evasion from the whole Yangtze River network, including mainstem and its tributaries of all orders, remains lacking. This has further hampered the assessment of its CO₂ evasion in a wider context linking the watershed's land-atmosphere and land-ocean carbon exchanges.

Table 4

Accelerated human activity is another urgent issue to be considered when investigating its riverine *p*CO₂ and CO₂ evasion. Approximately 50,000 dams, including the world's largest reservoir (i.e., the Three Gorges Reservoir; TGR), have been constructed in recent decades (Xu and Milliman, 2009). Assessing the impacts of dam-triggered changes to flow regime and biogeochemical processes on *p*CO₂ and CO₂ evasion is particularly important for deeper insights into its riverine carbon cycle (Table 4). For example, while the *p*CO₂ at Datong station declined continuously before the TGR impoundment (Fig. 4a; Wang et al., 2007a), our recent field survey shows that it has recovered from 1440 µatm in the 1980s to present 1700 µatm (see Fig. 4a). As for CO₂ degassing, recent work in the TGR indicates that its CO₂ evasion fluxes are different from natural rivers and are higher than other temperate reservoirs (Table 4; Zhao et al., 2013). Future research efforts are warranted to conduct systematic monitoring and evasion estimation. Given the Yangtze River's role in global carbon export, a comprehensive assessment of CO₂ evasion is also meaningful for global carbon budget.

Conclusions

By using long-term water chemistry data measured in the Yangtze River watershed during the period 1960s–1985, we calculated its $p\text{CO}_2$ from pH and alkalinity. The pH in the Yangtze River waters varied from 6.5 to 9.2 and the alkalinity ranged from 415 to >3400 $\mu\text{eq L}^{-1}$ with high alkalinity concentrations occurring in carbonate-rich tributary catchments. Except one station in the upper reach showing a lower $p\text{CO}_2$ than the atmosphere, the Yangtze River waters were supersaturated with dissolved CO_2 , generally 2–20 folds the atmospheric equilibrium. Averaged over all stations, the basin-wide $p\text{CO}_2$ was $2662 \pm 1240 \mu\text{atm}$. As an important parameter for CO_2 evasion estimation, its $p\text{CO}_2$ was characterized by significant spatial and temporal variability, which was collectively controlled by carbon inputs from terrestrial ecosystems, hydrological regime, and rock weathering. High $p\text{CO}_2$ values were observed spatially in catchments with abundant carbonate presence and seasonally in the wet season when recent-fixed organic matter was flushed into the river network. Decomposition of organic matter by microbial activity in aquatic systems facilitated CO_2 production and sustained the high $p\text{CO}_2$ values in the wet season, although the alkalinity presented a significant dilution effect with water discharge. In addition, the $p\text{CO}_2$ decreased with increasing stream orders from the smallest headwater streams through tributaries to the mainstem channel. A higher $p\text{CO}_2$ in the headwater streams illustrated the influence of direct inputs of terrestrially-derived organic matter and weathering products via erosion and flushing on riverine carbon dynamics.

The substantially higher $p\text{CO}_2$ than the atmosphere indicated a potential of significant CO_2 emissions from the Yangtze River fluvial network. Quantifying the amount of CO_2 evasion

should be a top priority, upon which its biogeochemical implications for watershed-scale carbon cycle can be assessed in association with carbon burial and downstream export. Given the

485 extensive and intensive human disturbances within the watershed since the 1990s, special attention must be paid to the resulting changes to riverine $p\text{CO}_2$ and CO_2 evasion. A comparative analysis involving CO_2 evasion before large-scale human impacts and recent degassing estimates (e.g., Li et al., 2012; Liu et al., 2016) will be able to examine the anthropogenic perturbations of the river-atmosphere CO_2 fluxes due to damming and land-use change. Considering the Yangtze
490 River's relevance to global carbon export, quantifying its CO_2 evasion is also of paramount importance for better assessments of global carbon budget.

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700 Table 1. Comparison of alkalinity ($\mu\text{eq L}^{-1}$) and pH at Wuhan station between the GEMS/Water Programme results and the hydrological yearbooks, expressed as mean \pm standard error.

Item	1980	1981	1982	1983	1984	1984
GEMS/Water Programme						
Alkalinity	2050 \pm 286	2004 \pm 188	2000 \pm 232	1838 \pm 252	2200 \pm 247	1992 \pm 219
pH	7.83 \pm 0.16	7.73 \pm 0.24	8.04 \pm 0.09	8.06 \pm 0.05	8.00 \pm 0.09	7.88 \pm 0.06
Hydrological yearbooks						
Alkalinity	2310 \pm 314	2187 \pm 236	2274 \pm 268	2033 \pm 304	2383 \pm 277	2306 \pm 238
pH	7.93 \pm 0.09	7.87 \pm 0.09	8.01 \pm 0.09	7.94 \pm 0.08	7.93 \pm 0.10	7.98 \pm 0.08

Table 2. Riverine pH, alkalinity, and $p\text{CO}_2$ in the Yangtze River basin (median±standard deviation).

River/tributary	Station	pH	Alkalinity	$p\text{CO}_2$
			$\mu\text{eq L}^{-1}$	μatm
Mainstem	Benzilan	8.29±0.11	2352±435	681±156
	Shigu	8.18±0.48	2544±438	846±262
	Jingjiangjie	8.11±0.12	2905±362	916±202
	Dukou	8.22±0.12	2399±429	826±197
	Longjie	8.23±0.17	2185±396	786±226
	Huatan	8.17±0.15	2237±418	882±287
	Pingshan	8.13±0.10	2215±407	1001±235
	Zhutuo	7.88±0.19	2299±349	2405±781
	Cuntan	8.08±0.11	2173±311	1087±319
	Yichang	7.95±0.15	2343±300	1653±469
	Luoshan	7.76±0.11	2280±248	2380±691
	Wuhan	7.93±0.11	2060±263	1521±497
	Datong	7.84±0.14	1919±312	1711±806
	Nanjing ^a	7.56±0.16	2339±339	3796±1623
	Nanjing ^b	7.54±0.18	2296±357	3793±2186
Major tributaries ^c				
Yalongjiang	Xiaodeshi	8.02±0.22	2576±465	1567±715
Daduhe	Fuluzhen	7.66±0.23	1909±289	2577±1620
Minjiang	Gaochang	8.02±0.15	1816±327	1020±525
Tuojiang	Lijiawan	8.01±0.11	2705±507	1504±572
Jialingjiang	Beibei	8.11±0.14	2289±509	1196±244
Wujiang	Wulong	8.01±0.14	2420±279	1361±508
Yuanjiang	Taoyuan	7.61±0.25	1822±480	2801±2144
Xiangjiang	Xiangtan	7.76±0.44	1739±331	2349±2521
Hanjiang	Xiaoshicun	7.93±0.13	2262±480	1715±536
Ganjiang	Waizhou	7.44±0.44	880±236	2205±2048
Yangtze basin ^d	1% percentile	7.03	556	788
	10% percentile	7.35	842	1236
	50% percentile	7.71	2237	2455
	90% percentile	8.05	3305	4344
	99% percentile	8.28	4437	6163

^aaffected by high tides.

^baffected by low tides.

^cMedian values of the data for the lowermost station on the mainstem of the specific tributary.

^dStatistics based on the measurements at the used 339 stations.

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Table 3. Hydro-geochemical features of the Wujiang (Wulong station), Jialingjiang (Wusheng station), and Ganjiang (Xiajiang station) catchments.

Control station	Control area km ²	Water discharge m ³ s ⁻¹	pH	Alkalinity	<i>p</i> CO ₂	Ca ²⁺	SiO ₂	Sedimentary rock types (% of area)		
				μeq L ⁻¹	μatm	μmol L ⁻¹	μmol L ⁻¹	Carbonate	Siliciclastic	Igneous + metamorphic
Wulong	80,536	1570	7.72±0.14	3021±527	3537±1247	1145±278	59±31	82.9	14.8	2.3
Wusheng	80,550	793	7.80±0.21	2484±948	2671±490	1005±170	94±30	30.4	55.3	14.3
Xiajiang	62,387	1644	7.34±0.08	953±266	2642±626	242±91	105±18	9.1	64.7	26.2

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Table 4. Comparison of *p*CO₂ and CO₂ evasion among world large rivers and typical reservoirs in the Yangtze River basin.

River	Country	Climate	<i>p</i> CO ₂	CO ₂ evasion	Reference
			μatm	mol m ⁻² yr ⁻¹	
Yangtze network	China	Subtropical monsoon	2662±1240	/	This study
Upper Yangtze	China	Subtropical monsoon	2100	57	Li et al., 2012
Lower Yangtze	China	Subtropical monsoon	1297±901	14.2–54.4	Wang et al., 2007a
Yangtze estuary	China	Subtropical monsoon	650–1440	15.5–34.2	Zhai et al., 2007
Amazon	Brazil	Tropical	3929	162.2	Lauerwald et al. 2015
Ottawa	Canada	Temperate	1200	14.2	Telmer and Veizer, 1999
Hudson	USA	Temperate	1125±403	5.8–13.5	Raymond et al., 1997
York estuary	USA	Temperate	1070±867	6.3	Raymond et al., 2000
Mississippi	USA	Temperate	1335±130	98.5±32.5	Dubois et al., 2010
Yukon	Canada	Subarctic	582–705	11.6–21.2	Lauerwald et al., 2015
Yellow	China	Arid and semiarid	2810±1985	312.4±149.2	Ran et al., 2015b
Xijiang (Pearl)	China	Subtropical monsoon	2600	69.2–130	Yao et al., 2007
Mekong (>100 m wide rivers)	SE Asia	Tropical monsoon	703–1597	32–138	Alin et al., 2011
Godavari estuary	India	Tropical monsoon	<500–33,000	52.6	Sarma et al., 2011
Global rivers			2400	131.2	Lauerwald et al. 2015
<i>Typical reservoirs in the Yangtze River basin</i>					
Wujiang cascade reservoirs			38–3300	-3.3–32.5	Wang et al., 2011
Three Gorges Reservoir (TGR)			/	35.1	Zhao et al., 2013

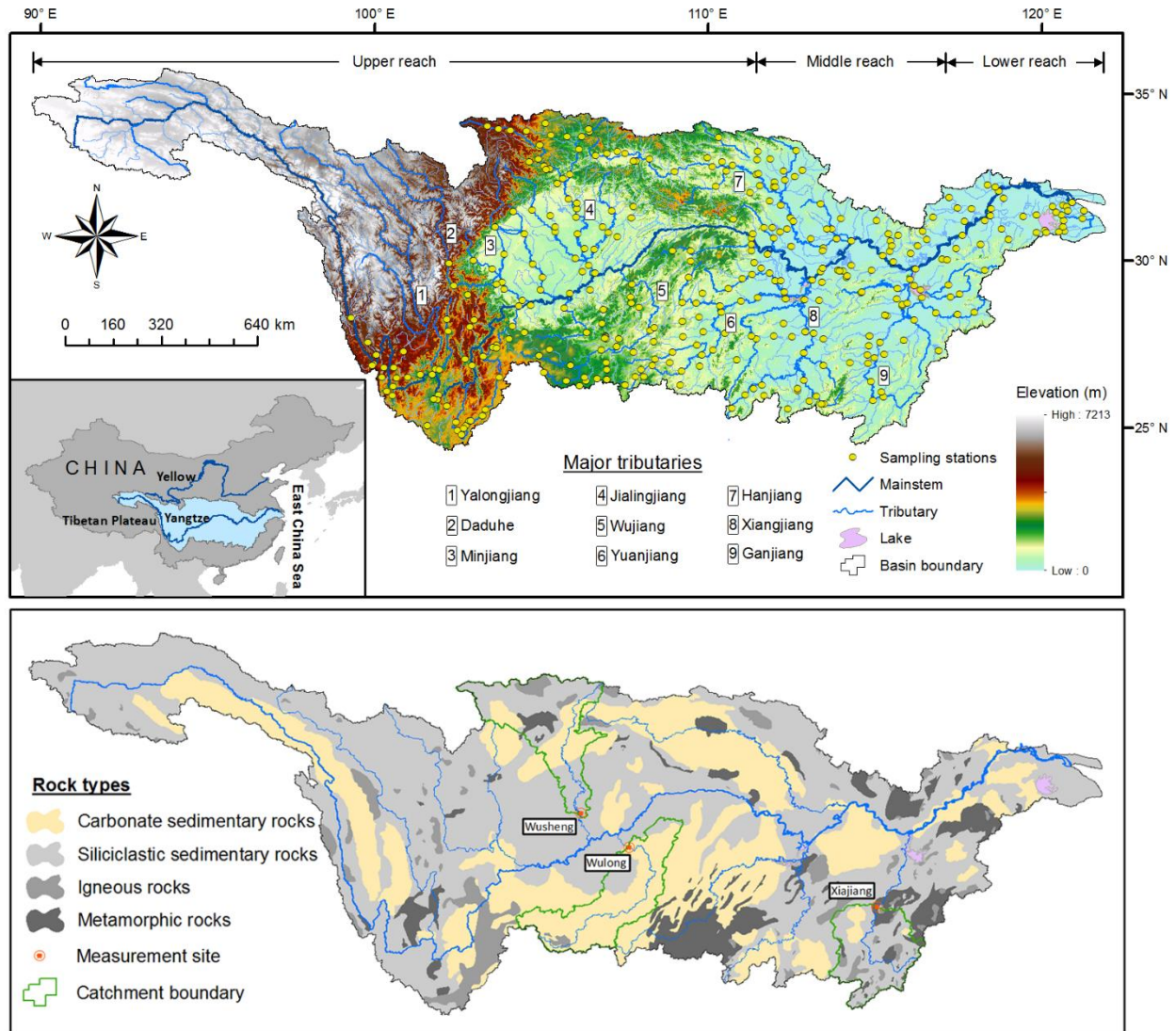


Fig. 1. Maps of the Yangtze River basin showing sampling stations (top) and rock compositions (bottom). Rock information is modified from Chen et al. (2002) and Chetelat et al. (2008).

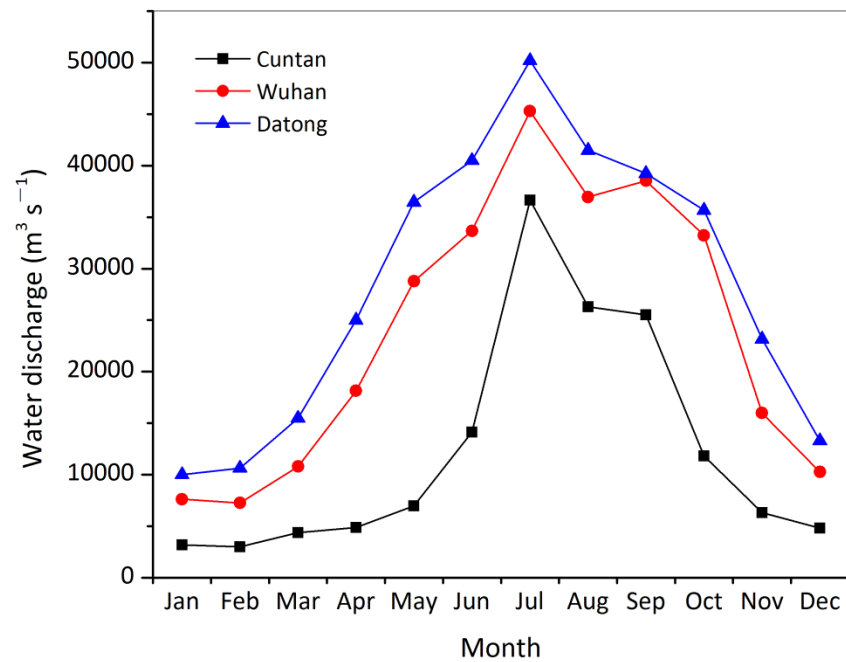


Fig. 2. Monthly variations in water discharge of the Yangtze River at Cuntan (upper reach), Wuhan (middle reach), and Datong stations (lower reach).

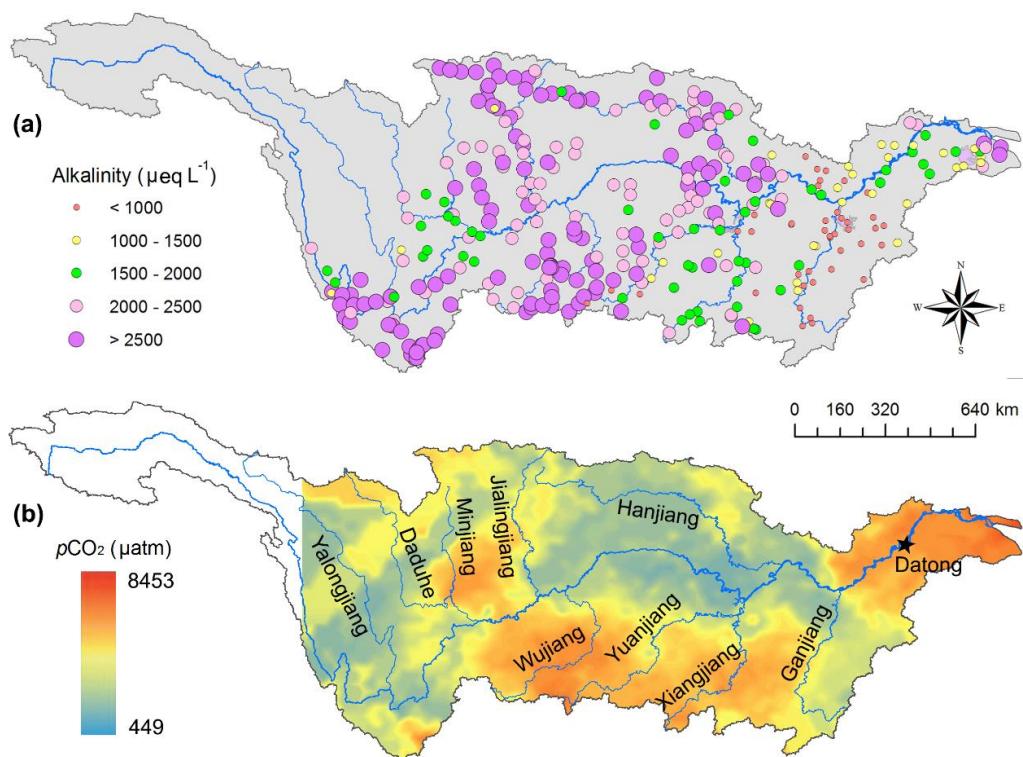


Fig. 3. Spatial distribution of alkalinity (a) and $p\text{CO}_2$ (b) in the Yangtze River basin. The headwater region in (b) was not interpolated due to insufficient stations.

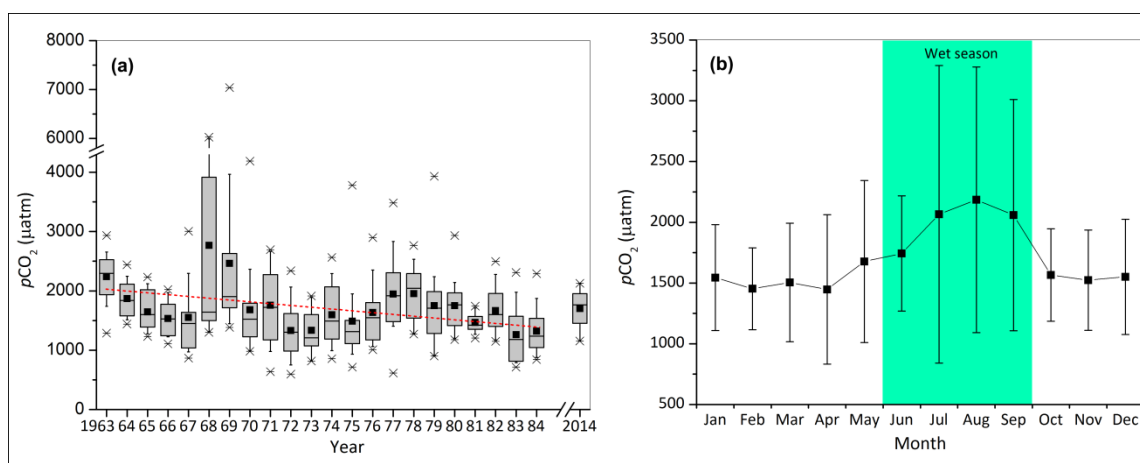


Fig. 4. Temporal variations of $p\text{CO}_2$ at Datong station. (a) box-and-whisker plot shows significant inter-annual changes; (b) seasonal variations. The dash line in (a) represents linear regression and the values for 2014 are derived from Liu et al. (2016). Error bars denote standard deviation.

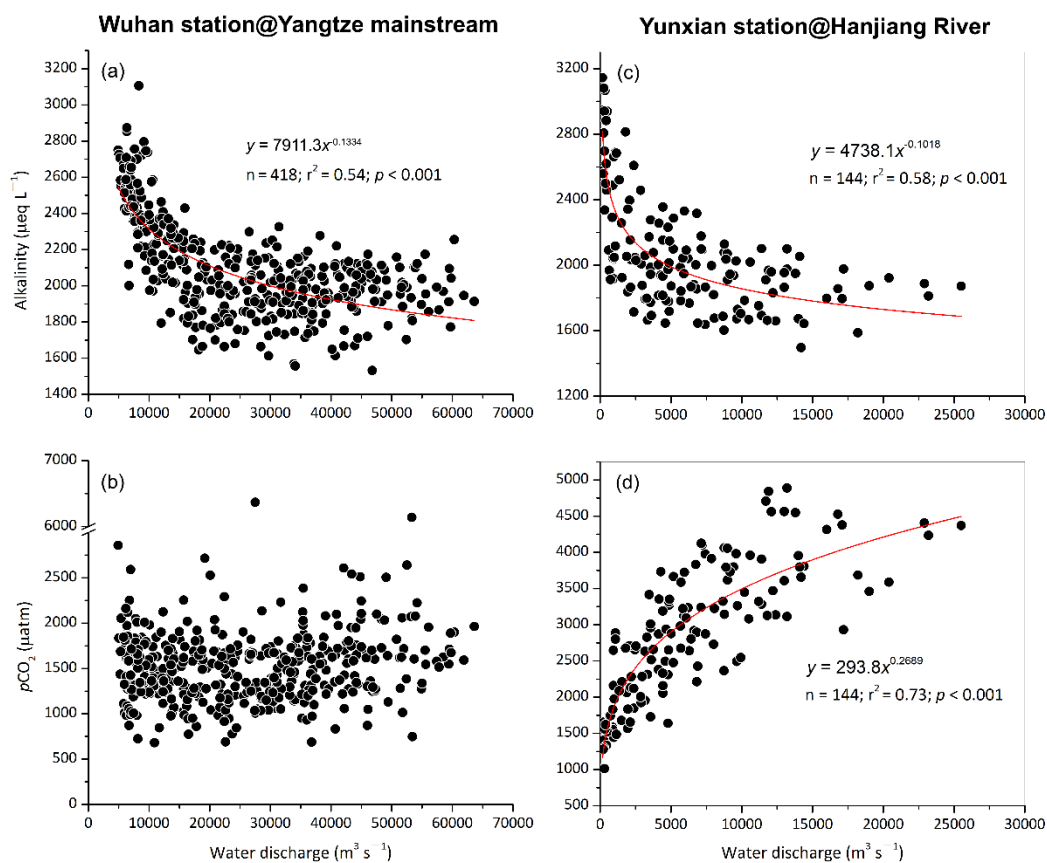


Fig. 5. Correlations between water discharge and instantaneous alkalinity and $p\text{CO}_2$: the mainstem at Wuhan station (a and b) and the Hanjiang River at Yunxian station (c and d).

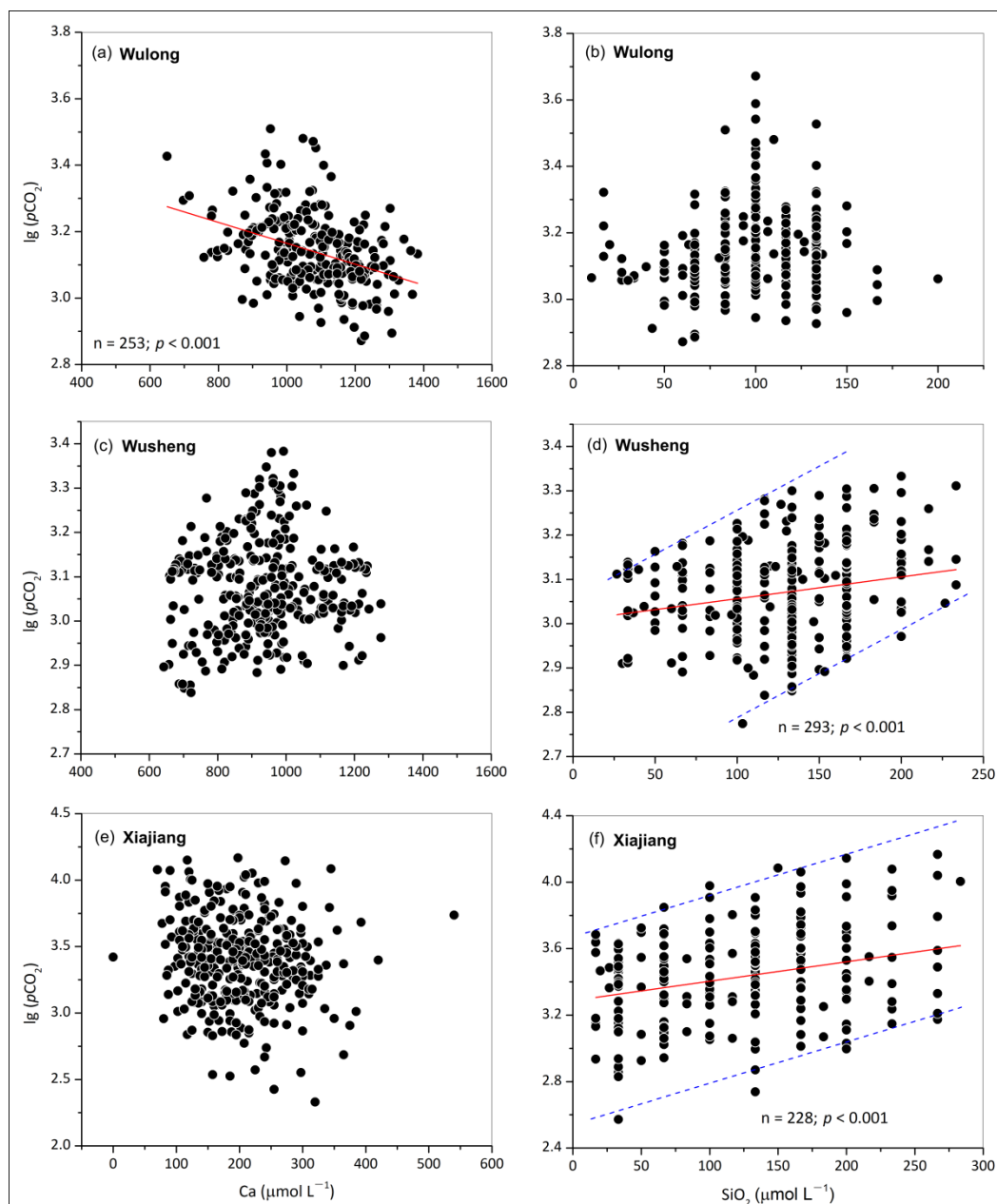


Fig. 6. Responses of $p\text{CO}_2$ to rock weathering products in three typical catchments with distinct rock compositions: a–b: Wujiang River (Wulong station); b–c: Jialiangjiang River (Wusheng station); e–f: Ganjiang River (Xiajiang station). The solid lines represent linear regression.

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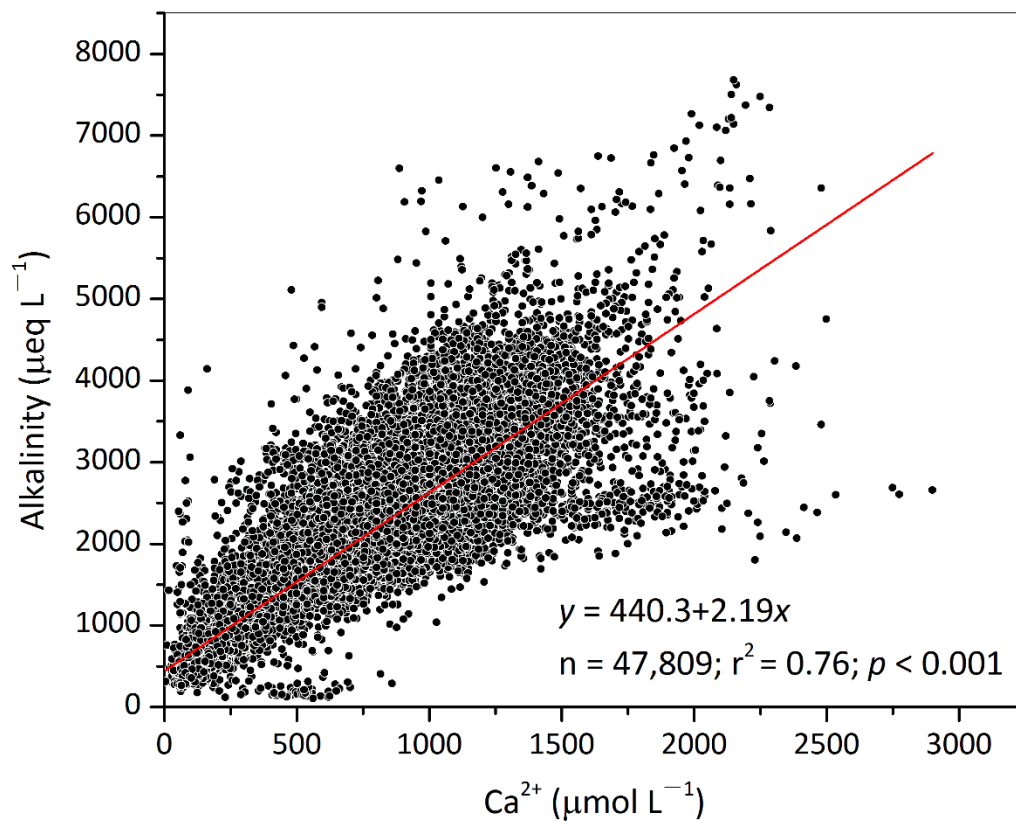


Fig. 7. Strong correlation between chemical weathering, using Ca^{2+} as a proxy, and alkalinity.

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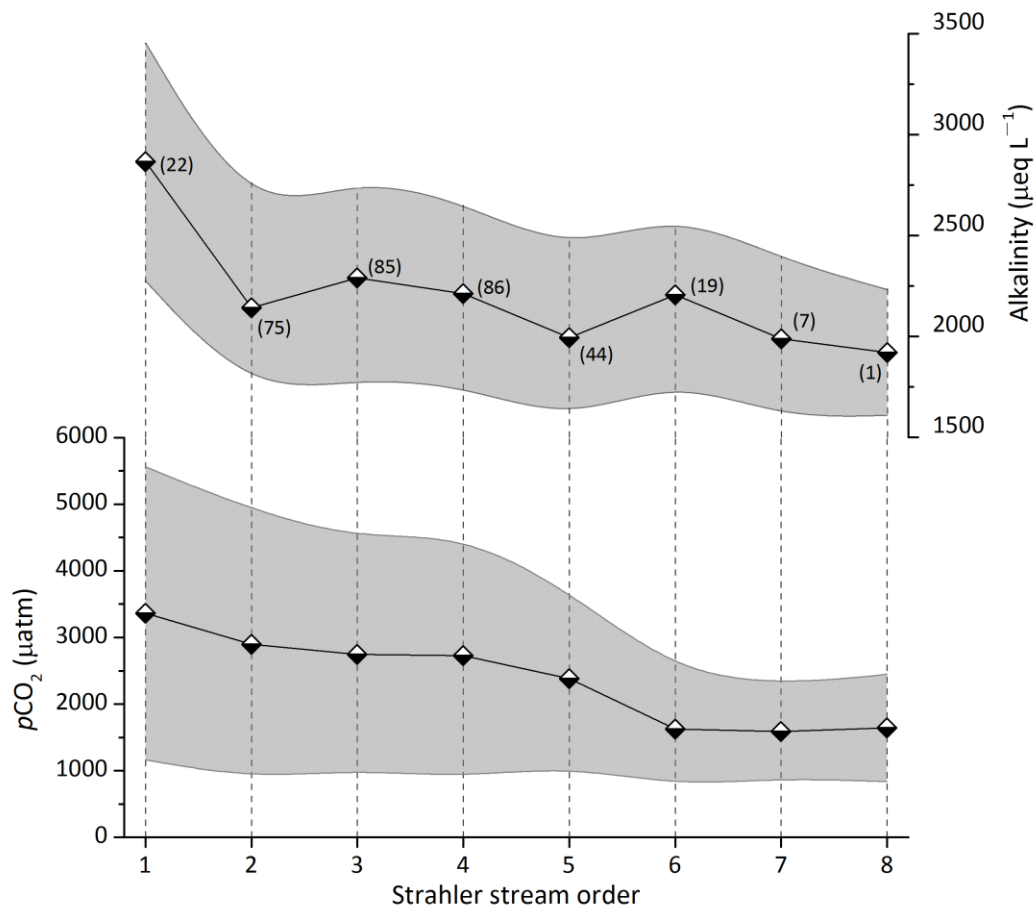


Fig. 8. Decreasing alkalinity (top) and $p\text{CO}_2$ (bottom) with increasing Strahler stream order. The grey shade denotes standard deviation and the numbers in parentheses represent the number of stations aggregated for each stream order.