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# Dynamics of riverine $CO_2$ in the Yangtze River fluvial network and their implications for carbon evasion

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- Abstract: Understanding riverine carbon dynamics is critical for not only better estimates of various carbon fluxes but also evaluating their significance in the global carbon budget. As an important pathway of global land-ocean carbon exchange, the Yangtze River has received less attention regarding its vertical carbon evasion than lateral transport. Using long-term water chemistry data, we calculated CO<sub>2</sub> partial pressure (pCO<sub>2</sub>) from pH and alkalinity and examined its spatial and temporal dynamics and the impacts of environmental settings. With alkalinity ranging from 415 to  $>3400 \mu mol L^{-1}$ , the river waters were supersaturated with dissolved CO<sub>2</sub>, generally 2–20 folds the atmospheric equilibrium (i.e., 390 μatm). Changes of pCO<sub>2</sub> were collectively controlled by terrestrial ecosystems, hydrological regime, and rock weathering. High pCO<sub>2</sub> values were observed spatially in catchments with abundant carbonate presence and seasonally in the wet season when recent-fixed organic matter was exported into the river network. In-stream processing of organic matter facilitated CO<sub>2</sub> production and sustained the high  $pCO_2$ , although the alkalinity presented an apparent dilution effect with lower alkalinity concentrations in higher flow periods. The decreasing pCO<sub>2</sub> from the smallest headwater streams through tributaries to the mainstream illustrates the significance of direct terrestrial carbon input in controlling riverine carbon. With a basin-wide mean pCO<sub>2</sub> of 2662±1240 µatm, substantial

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 $CO_2$  evasion from the Yangtze River fluvial network is expected. Future research efforts are thus needed to quantify the amount of  $CO_2$  evasion and assess its biogeochemical implications for watershed-scale carbon cycle. In view of the Yangtze River's relative importance in global carbon export, its  $CO_2$  evasion would be significant for global carbon budget.

**Keywords:**  $CO_2$  partial pressure ( $pCO_2$ ); riverine carbon cycle; spatial and temporal patterns;  $CO_2$  evasion; Yangtze River

# 1. Introduction

Inland waters, including rivers, streams, lakes, wetland, and reservoirs, have recently been recognized as active components of the global carbon cycle, transporting, storing, and processing huge amounts of terrestrially-derived carbon (Aufdenkampe et al., 2011;Cole et al., 2007;Raymond et al., 2013;Richey et al., 2002;Weyhenmeyer et al., 2015;Borges et al., 2015). With a higher CO<sub>2</sub> partial pressure (*p*CO<sub>2</sub>) than the atmospheric equilibrium (i.e., 390 μatm), inland waters are mostly net carbon sources to the atmosphere. Published studies show that the annually degassed CO<sub>2</sub> from inland waters is estimated to almost entirely compensate the total annual carbon uptake by ocean systems (Wanninkhof et al., 2013;Regnier et al., 2013). Global estimates of CO<sub>2</sub> evasion from rivers and streams range from 0.56 to 1.8 PgC yr<sup>-1</sup> (Aufdenkampe et al., 2011;Raymond et al., 2013;Lauerwald et al., 2015). It is apparent that these results vary considerably and are associated with great uncertainties. The most recent estimate of 0.65 PgC yr<sup>-1</sup> by Lauerwald et al. (2015) accounts for only 36% of the efflux estimate made by Raymond et al. (2013), although they used the same water chemistry data set. Among the numerous factors

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contributing to current  $CO_2$  evasion uncertainties, a principal reason is the absence of a spatially explicit  $pCO_2$  data set that covers the full spectrum of the global river and stream network.

Existing global maps of CO<sub>2</sub> evasion from fluvial network are typically generated on the basis of incomplete spatial coverage of pCO<sub>2</sub>, in which Asian rivers are heavily underrepresented (e.g., Aufdenkampe et al., 2011; Battin et al., 2009; Lauerwald et al., 2015; Raymond et al., 2013). Due to lack of direct in situ measurements, simplified extrapolation is normally used to predict pCO<sub>2</sub> in and CO<sub>2</sub> evasion from Asian river systems. Consequently, the estimation accuracy is problematic and even erroneous. For example, for the Yellow River in East Asia, while the calculated pCO<sub>2</sub> from river water chemistry records is 2800 µatm (Ran et al., 2015a), the modeled pCO<sub>2</sub> by Lauerwald et al. (2015) is 30% lower (i.e., <2000 μatm). A much lower estimate of  $<700 \,\mu$ atm can be derived from the pCO<sub>2</sub> map produced in Raymond et al. (2013). Such great discrepancies are largely because riverine pCO<sub>2</sub> is highly site-specific and affected by a wide range of environmental factors (e.g., Abril et al. 2015; Teodoru et a., 2015). Asian rivers are significant contributors to global carbon export, accounting for 40% of the global carbon flux from land to sea (Schl ünz and Schneider, 2000; Hope et al., 1994). Estimating the amount of CO<sub>2</sub> degassed from Asian rivers is critical for global CO<sub>2</sub> evasion assessments. Recent work in Mekong and Yellow rivers has demonstrated high pCO<sub>2</sub> and CO<sub>2</sub> effluxes (Alin et al., 2011;Ran et al., 2015b), further highlighting the necessity of incorporating the currently underrepresented Asian rivers into global carbon budget assessments.

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As an important carbon contributor to the West Pacific Ocean, the Yangtze River has received widespread attention in fluvial carbon export at various spatial and temporal scales. Studies of 70 flux estimates of different carbon species date back to the early 1980s (Cauwet and Mackenzie, 1993; Gan et al., 1983; Milliman et al., 1984; Wang et al., 2012; Zhang et al., 2014; Ittekkot, 1988). Intensive observations covering seasonal variability show that the Yangtze River transports approximately 20 Mt of carbon per year into the oceans (Wu et al., 2007; Bao et al., 2015). Contrary to the long history of lateral export measurements, however, few studies have examined the vertical carbon exchange between the river system and the atmosphere (Li et al., 2012; Zhao et al., 2013; Chen et al., 2008). This is by nature largely due to the differences in sampling strategy. Unlike the lateral export that only involves measurements on the mainstream or at specific sites near the river mouth, quantifying basin-wide CO<sub>2</sub> evasion requires a spatially explicit pCO<sub>2</sub> data set encompassing the entire fluvial network. Any attempts of using limited 80 local measurements to up-scale to the watershed scale are challenging and subject to large uncertainties. This has impacted the understanding of riverine carbon cycle within the Yangtze River watershed as well as its links to the atmosphere and ocean systems.

By using long-term water chemistry data measured in the Yangtze River basin, we calculated the riverine pCO<sub>2</sub> from pH and alkalinity. In combination with hydrologic and geologic information, the objectives of this study were to 1) investigate the spatial and temporal patterns of  $pCO_2$  under 'natural' processes before significant human perturbations, mainly dam impoundment and land use changes since the 1990s; 2) to explore the couplings between pCO<sub>2</sub> and environmental

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settings by investigating environmental and geomorphologic controls. Based on the obtained  $pCO_2$ , we further evaluated its implications for  $CO_2$  evasion. In view of the Yangtze River's role in global fluvial export of water, sediment, and carbon (Syvitski et al., 2005; Wang et al., 2012), its  $CO_2$  evasion would be globally substantial. This  $pCO_2$  database is thus helpful to examine the spatial distribution of global riverine  $pCO_2$  and to refine global  $CO_2$  evasion from river systems.

#### 95 **2. Material and methods**

# 2.1 The Yangtze River basin

With a length of 6380 km, the Yangtze River is the longest river in China and the third longest in the world. The river originates on the Tibetan Plateau and flows eastward through the Sichuan Basin and the Middle-Lower Reach Plains, before emptying into the East China Sea (Fig. 1a). Its drainage area is 1.81 million km<sup>2</sup>. The Yangtze River basin is mainly overlain by sedimentary rocks that are composed of marine carbonates, evaporites, and continental deposits. Carbonates are widely distributed within the watershed and are particularly abundant in the Wujiang, Yuanjiang, and Hanjiang tributary catchments (Fig. 1b). Silicates are also widely present in the basin while metamorphic rocks are mainly scattered in the middle-lower reach (Fig. 1b). The Yangtze River is joined by a number of large tributaries, including the Yalongjiang, Daduhe, Minjiang, Jialingjiang, Wujiang, Yuanjiang, Xiangjiang, Hanjiang, and Ganjiang rivers (Fig. 1a).

# Figure 1

Except the headwater region characterized by high elevation and cold climate (annual mean temperature <4 °C), the remaining watershed is affected by a subtropical monsoon climate with

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the annual mean temperature in the middle-lower reach varying from 16 to 18 °C (Chen et al., 2002). Rainfall is the major source of water discharge, whereas snowfall supply is only significant in the ice-covered upstream mountainous areas. With a mean precipitation of 1100 mm yr<sup>-1</sup>, the precipitation is spatially highly variable, decreasing from 1644 mm yr<sup>-1</sup> in the lower reach, to 1396 mm yr<sup>-1</sup> in the middle reach, and 435 mm yr<sup>-1</sup> in the upper reach (Chetelat et al., 2008). Approximately 60% of the annual precipitation falls during the wet season from June to September. Affected by summer monsoon, the wet season generally occurs earlier in the middle and lower reaches than in the inland upper reach. The water discharge from the upper to the lower reach presents a strong seasonal variability (Fig. 2). Monthly peak discharge occurs in July and can be 5–7 times greater than the lowest discharge in the dry season. The mean discharge at Datong station is 28,200 m<sup>3</sup> s<sup>-1</sup> (see its location in Fig. 3b), and consequently the Yangtze River annually discharges 889 km<sup>3</sup> of water into the ocean (Yang et al., 2002).

# Figure 2

# 2.2 Water chemistry data

Concentrations of alkalinity, major ions, and dissolved silica measured at 359 stations in the Yangtze River watershed (Fig. 1a) during the period 1960s–1985 were retrieved from the Hydrological Yearbooks, which were yearly produced by the Yangtze River Conservancy Commission (YRCC) for internal use. Other environmental variables concurrently measured at each sampling event, including pH, water temperature, and water discharge, were also extracted from the yearbooks for this study. The water samples for pH and temperature measurement were taken in the same period as these for ion analysis. The sampling frequency ranged from 1 to 14

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times per month depending on flow conditions. Sampling at some stations during the period 1966–1975 was less frequent. About 80% of the 359 stations have been continuously sampled for at least 10 years, starting from the early1970s. To avoid severe river pollution by human activity, only the samples collected prior to 1985 were used. In addition, samples with the pH lower than 6.5 were manually discarded because the calculated *p*CO<sub>2</sub> would be greatly biased due to contributions of noncarbonated alkalinity such as organic acid anions (Abril et al., 2015; Hunt et al., 2011). Because reservoir trapping and increased water residence time can remarkably alter the physical and biogeochemical properties of running water (Kemenes et al., 2011; Barros et al., 2011), the stations located inside or shortly below reservoirs were also intentionally removed. Given the tidal influences, mainstream stations downstream of Datong, 626 km inland from the coast, were also excluded, as were the stations in the delta region that were affected either by tides or by intersections with other rivers via artificial canals. Based on these selection criteria, 339 stations, including 13 mainstream stations and 326 tributary stations, were retained and 47,809 water chemistry measurements in total were compiled.

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Chemical analyses of water samples were performed under the authority of YRCC following the standard procedures and protocols described by Alekin et al. (1973) and the American Public Health Association (1985). While pH and temperature were measured in the field, the alkalinity was determined by acid titration. Detailed sampling and analysis procedures were presented in Chen et al. (2002). One important issue regarding historical records is data reliability. No assessment reports on quality assurance and quality control are available in the hydrological

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yearbooks. An effective evaluation approach is to compare the hydro-chemical differences for samples collected at the same station but by different agencies. The Wuhan station on the Yangtze mainstream has also been monitored under the United Nations GEMS/Water

Programme since 1980 (only yearly means available at <a href="http://www.unep.org/gemswater">http://www.unep.org/gemswater</a>). The pH value from the yearbooks agreed well with that measured by the GEMS/Water Programme with <1.8% differences, while the alkalinity discrepancy between the two data sets is larger (Table 1). The yearbooks report a slightly higher alkalinity than the GEMS/Water Programme results by 7.6–13.9%, indicating that the yearbook reports are reliable for pCO<sub>2</sub> calculation. High data quality of the yearbook reports can also be validated from comparison of major dissolved elements measured by the two agencies at Wuhan station (see Chen et al., 2002).

#### Table 1

# 2.3 Calculation of pCO<sub>2</sub>

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The conventional method of calculating  $pCO_2$  from pH and alkalinity was used. With ~90% of the measured pH ranging from 7.1 to 8.3 suggestive of natural process for the Yangtze River, bicarbonates were assumed equivalent to alkalinity (Amiotte-Suchet et al., 2003) because 96% of the alkalinity was dominated by bicarbonates. This approximation has been frequently used in Chinese river systems (Yao et al., 2007;Li et al., 2012;Ran et al., 2015a). The  $pCO_2$  was then calculated using CO2SYS program (Lewis and Wallace, 1998). However, using this method would produce biased extreme values that are unrealistic in natural river systems (Hunt et al., 2011;Weyhenmeyer et al., 2015). We thus reported median values per sampling station instead of means to avoid the impact of erroneous extreme results.

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#### 3. Results

# 3.1 Spatio-temporal variability of alkalinity and pCO<sub>2</sub>

Except the excluded measurements, pH in the Yangtze River waters varied from 6.5 to 9.2 with 96% of the pH measurements ranging from 7.3 to 8.3 (Table 2). Higher pH values (i.e., >7.8) were spatially measured in the headwater streams and the Hanjiang catchments (see Fig. 1a for location). In comparison, the tributaries in the southern part of the watershed exhibited relatively lower pH values. For the mainstream (Table 2), the median pH showed a significant downstream decrease from 8.29 to 7.55 ( $r^2 = 0.77$ ; p < 0.001). The alkalinity varied from 415 to >3400 µmol  $L^{-1}$  (Fig. 3a). Higher alkalinity (i.e., >2500 µmol  $L^{-1}$ ) was observed in the upper reach and the upper part of the middle reach (Fig. 3a), in particular the carbonate-rich tributary catchments (e.g., the Jialingjiang, Wujiang, and Hanjiang rivers). In contrast, the lower part of the middle reach (mainly the Ganjiang River) and the lower reach showed a lower alkalinity of <2000 µmol  $L^{-1}$ . The average alkalinity over the whole watershed was  $2210\pm1023$  µmol  $L^{-1}$ .

# Figure 3 and Table 2

The calculated  $pCO_2$  varied by a magnitude of 2 with the highest  $pCO_2$  being 24,432  $\mu$ atm. At 95% of the stations, the  $pCO_2$  was higher than 1000  $\mu$ atm, generally 2–20 folds the atmospheric  $pCO_2$ . Only one station in the upper reach showed a median  $pCO_2$  lower than the atmosphere. In the mainstream, the  $pCO_2$  increased from ~700  $\mu$ atm at the uppermost station to 3800  $\mu$ atm at Nanjing near the river mouth (Table 2). Averaged over all stations, the basin-wide  $pCO_2$  was  $2662\pm1240~\mu$ atm. To better illustrate its spatial variability, we modeled the  $pCO_2$  for the whole

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stream network using the Kriging interpolation method in ArcGIS 10.1 (Esri, USA) with the assumption that the station-based  $p\text{CO}_2$  was representative of the surrounding streams. Similar to alkalinity, the  $p\text{CO}_2$  presented significant spatial variations (Fig. 3b). The Yangtze mainstream near the headwater region and the Yalongjiang catchment showed the lowest  $p\text{CO}_2$ , generally <1000  $\mu$ atm. In comparison, the carbonate-rich tributaries in the southern part of the watershed had high  $p\text{CO}_2$  values. With carbonates occupying 83% of the catchment, the Wujiang River presented the highest median  $p\text{CO}_2$  than other tributaries, averaging 3550±1356  $\mu$ atm. In the lower reach, the  $p\text{CO}_2$  was 3988±1244  $\mu$ atm on average, which is inconsistent with its relatively low alkalinity of <2000  $\mu$ mol L<sup>-1</sup> (Fig. 3a). It is worth noting that the  $p\text{CO}_2$  in Hanjiang catchment was lower than expected, given its high alkalinity (>2500  $\mu$ mol L<sup>-1</sup>). Differences in pH in these catchments are likely a principal cause of these inconsistencies.

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In addition, the  $pCO_2$  also showed strong temporal variability. Fig. 4 presents an example of  $pCO_2$  changes at Datong station on the mainstream. Despite considerable inter-annual variations that could change by a factor of 5, the annual  $pCO_2$  declined steadily during the >20-year-long sampling period ( $r^2 = 0.18$ ; p < 0.05) (Fig. 4a). This trend is pronounced even if the anomalously high values in the late 1960s are excluded from analysis. Indeed, more than half of the evaluated stations, mainly in the middle-lower reach, showed a significant decreasing trend at the 95% confidence level. In contrast, gradual increases were observed at some tributary stations in the upper reaches. Seasonally, the  $pCO_2$  in the wet season was on average 30% higher than that in the dry season (Fig. 4b), and greater fluctuation ranges could be observed in the wet season.

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215 Figure 4

#### 3.2 Correlations with hydro-geochemical variables

Fig. 5 presents two representative examples showing responses of alkalinity and  $pCO_2$  to hydrological regimes. Changes of alkalinity at both stations reflected a clear dilution effect. High alkalinity concentrations were measured in low flow periods when groundwater was the major contributor to runoff (Figs. 5a and 5c). Checking all stations indicated that the alkalinity at 98% of the stations decreased exponentially with increasing water discharge after the onset of the wet season. In contrast, the  $pCO_2$  presented diverse relationships with water changes (Figs. 5b and 5d). There was no discernible dependence of  $pCO_2$  on flow in the mainstream, while a positive correlation was widely observed in small tributaries. Although only two stations were plotted here, these diverse responses of alkalinity and  $pCO_2$  to flow changes were widespread within the watershed, in particular for  $pCO_2$  between mainstream and small tributaries.

#### Figure 5

In order to elucidate the impacts of rock weathering on  $pCO_2$ , we selected three typical tributary catchments with differing rock compositions (Table 3). The Wujiang catchment is mainly underlain by carbonates (83%) and the Ganjiang catchment by silicates (65%), whereas the Jialingjiang catchment lies in the middle regarding the areal coverage of the two rocks (Table 3 and Fig. 1b). As the most typical weathering products of carbonate and silicate rocks, we plotted  $Ca^{2+}$  and dissolved silica (expressed as  $SiO_2$ ) against  $pCO_2$ , respectively (Fig. 6). For the three catchments with contrasting rock compositions, the  $pCO_2$  showed different responses to  $Ca^{2+}$  and  $SiO_2$ . In Wujiang catchment, the log-transformed  $pCO_2$  (i.e.,  $lg(pCO_2)$ ) presented a significant

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negative correlation with  $Ca^{2+}$  concentration (p < 0.001) (Fig. 6). This negative correlation became less apparent with decreasing carbonate coverage in Jialingliang and Ganjiang catchments. In contrast, while the  $\lg(pCO_2)$  showed positive correlation with  $SiO_2$  in Jialingjiang and Ganjiang catchments characterized by high silicate coverage, no clear relation between  $\lg(pCO_2)$  and  $SiO_2$  was detected in Wujiang catchment (Fig. 6).

# Figure 6 and Table 3

#### 4. Discussion

# 4.1 Uncertainty analysis of pCO<sub>2</sub>

As an important parameter for  $CO_2$  evasion estimation, an accurate riverine  $pCO_2$  is essential to 245 quantify CO<sub>2</sub> evasion and explore its biogeochemical implications for carbon cycle at different scales. Compared with direct measurement by means of membrane equilibration or headspace technique, the conventional pCO<sub>2</sub> calculation from alkalinity has been criticized for causing biases (Long et al., 2015; Hunt et al., 2011). Huge overestimations (i.e., >100%) have been reported in rivers with organic-rich and acidic waters due to combined effects of high organic acids and low buffering capacity of carbonate systems at low pH (Abril et al., 2015). 250 Unfortunately, there were no organic carbon information in the yearbooks, and measurements of dissolved organic carbon (DOC) in the Yangtze River started in the early 1980s. Its DOC ranging from 1.3 to 1.5 mg L<sup>-1</sup> was relatively low compared with other major world rivers (Bao et al., 2015; Wang et al., 2012). Our recent sampling also shows that the mean DOC is 1.9 mg  $L^{-1}$ for the mainstream and 2.4 mg L<sup>-1</sup> for major tributaries (Liu et al., 2016). Given the neutral to 255 basic pH range and the alkalinity variations, we believe the impact of organic acids is minimal,

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although a slight overestimation may have occurred as suggested by Abril et al. (2015). Our recent  $pCO_2$  measurements in the mainstream and major tributaries using a membrane contactor (Qubit DCO<sub>2</sub> System, Qubit Biology Inc., Canada) also indicate that the calculated  $pCO_2$  results are consistent with the measured values with only ~8% differences (Liu et al., 2016).

Furthermore, this  $pCO_2$  calculation method is sensitive to pH changes. High accuracy of pH measurements is critical to reduce the associated uncertainty. Similar to other water chemistry records (i.e., Butman and Raymond, 2011; Lauerwald et al., 2015; Weyhenmeyer et al., 2015), the retrieved pH was reported with a precision of one decimal place. If the uncertainties in pH measurement accuracy are assumed to 0.1 pH units, the calculated  $pCO_2$  would be underestimated by 26% or overestimated by 21%. To minimize human-induced disturbances in the chemical equilibrium of natural waters, we excluded the samples with pH<6.5 and treated them as being significantly polluted. Taking into account the higher alkalinity than the GEMS/Water Programme results, the propagated uncertainty ranges from 14% (underestimation) to 27% (overestimation). The Yangtze River basin is China's major industrial and agricultural regions, and influences of human activity, such as sewage inputs and chemical fertilizer usage to a lesser extent, may have altered the river water's chemical compositions and pH. As a result, 498 measurements were discarded from the analysis. Overall, for the Yangtze River with a high buffering capacity of carbonate alkalinity and low DOC concentrations, the calculated  $pCO_2$  is reasonable and can be used for further  $CO_2$  evasion estimation.

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# 4.2 Environmental impacts on alkalinity and pCO<sub>2</sub>

Export of alkalinity in river systems was affected by hydrological regime with a clear dilution effect (Fig. 5). The average alkalinity was 35% lower in the wet season than in the dry season. In both the mainstream and the tributaries, the higher alkalinity during low flow periods in the dry season (Figs. 5a and 5c) illustrated the contribution of groundwater recharge in providing abundant alkalinity. With widespread carbonate presence, groundwater in the Yangtze River watershed was rich in dissolved inorganic carbon (DIC). Recent studies show that the alkalinity of typical karst groundwater in the watershed is in the range of 3300-4200 µmol L<sup>-1</sup> (Li et al., 2010b;Li et al., 2010a). With reduced relative contribution of groundwater in the wet season, the high alkalinity was diluted by local rain events that carried lower DIC contents. Spatially, the dilution effect was more pronounced in the upper reach than the middle-lower reach. This may have revealed the response of alkalinity production to land cover. Catchments with a higher forest cover normally exhibit a stronger dilution effect than cropland catchments (Raymond and Cole, 2003). While cropland was the major land use type in the middle-lower reach accounting for 53.5% of the total catchment area, forest cover in the upper Yangtze River watershed was much higher (37.3%) than the middle-lower reach (30.4%; data are from Data Center for Resources and Environmental Sciences for the 1980s).

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Riverine dissolved  $CO_2$  originates primarily from terrestrial ecosystem respiration, groundwater input, and in-stream processing of land-derived organic matter (Wallin et al., 2013;Lynch et al., 2010). Different from alkalinity showing a clear dilution effect, the stable  $pCO_2$  in the Yangtze

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mainstream likely reflected the impacts of different biogeochemical processes in maintaining its  $pCO_2$  (Fig. 5b). Compared to the dry season in which the  $pCO_2$  was mainly controlled by DIC inputs from groundwater, the elevated pCO<sub>2</sub> in the wet season suggested the influence of organic carbon transport and decomposition. Its organic carbon content in the wet season is higher and the age is younger due primarily to strong erosion and leaching of recent-fixed organic matter (Wang et al., 2012; Zhang et al., 2014). Rapid mineralization of the labile fraction of organic carbon can increase the pCO<sub>2</sub>. For instance, approximately 60% of the recent-fixed carbon entering the Yangtze River can be quickly degraded in the wet season, while the degradation ratio in the dry season is only 31% (Wang et al., 2012). On the other hand, the increasing  $pCO_2$ with flow in tributaries (Fig. 5d) indicated enhanced supply of fresh dissolved CO<sub>2</sub> during high flow periods. For tributaries with more homogeneous catchment environments, decomposition of soil organic matter can provide abundant dissolved CO<sub>2</sub> (Liu et al., 2016;Li et al., 2012), thus generating a positive  $pCO_2$  response to flow changes. From this perspective, the stable  $pCO_2$  in the mainstream implied that the enhanced dissolved CO<sub>2</sub> input by soil organic matter decomposition from one region has likely been counteracted by low  $pCO_2$  waters derived from other regions. Consequently, the  $pCO_2$  dynamics appeared to be independent of hydrograph. This is highly possible given the spatial heterogeneity of the watershed environment in terms of vegetation cover, soil type, and rainfall intensity.

The spatial distribution of alkalinity overlapped well with the carbonate outcrops (Figs. 1b and 3a), with ~60% of the high alkalinity concentrations measured in carbonate catchments. Using

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Ca<sup>2+</sup> as a proxy of rock weathering, the strong correlation between Ca<sup>2+</sup> and alkalinity suggested the dominant role of weathering in controlling alkalinity and DIC export (Fig. 7). This is consistent with the significant impact of weathering on alkalinity as observed in other rivers (Raymond and Cole, 2003;Humborg et al., 2010). Particularly, given the higher susceptibility of carbonates to weathering than silicates (Goudie and Viles, 2012), the abundant carbonate presence in Wujiang catchment helped to sustain its high alkalinity and pCO<sub>2</sub> (Table 3). The negative response of pCO<sub>2</sub> to Ca<sup>2+</sup> in Fig. 6 indicated that an elevated pH has probably occurred, offsetting the weathering-induced DIC inputs in affecting pCO<sub>2</sub>. A slight pH increase would result in a reduced pCO<sub>2</sub> as this calculation method is sensitive to pH fluctuations (Laruelle et al., 2013). The positive correlation between pCO<sub>2</sub> and SiO<sub>2</sub> in Jialingjiang and Ganjiang catchments demonstrated the impact of DIC export by silicate weathering. Despite the high silicate weathering rate in Ganjiang catchment, its alkalinity represented only one third of that in the other two catchments (Table 3). Apparently, its high pCO<sub>2</sub> of 2642±626  $\mu$ atm was primarily due to its low pH (~6% lower). Overall, the catchments with more carbonate presence presented higher pCO<sub>2</sub> values (Figs. 1 and 3b).

335 Figure 7

Because  $pCO_2$  was calculated from alkalinity, its spatial variability reflected largely the export of the latter. The inconsistencies between  $pCO_2$  and alkalinity in Hanjiang catchment were likely caused by dam operation (Fig. 3). By altering the physical and biogeochemical properties of flowing water, dam trapping could cause a greatly declined  $pCO_2$  as a result of photosynthetic  $CO_2$  fixation and increased pH (Ran et al., 2015a). The Danjiangkou Reservoir (storage: 17.5

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km<sup>3</sup>) on the upper Hanjiang River was constructed in 1968. Unfortunately, the retrieved data for the Hanjiang River started from the 1970s, rendering it impossible to compare the  $pCO_2$  differences between pre- and post-dam periods. An indirect evidence is that an elevated pH within the reservoir has been measured (7.95–8.33; Li et al., 2009) relative to the 1970s (7.84±0.15). In the lower reach near the estuary (Fig. 3b), heterotrophic ecosystems and human activity could explain its high  $pCO_2$ . Settling down of particulate organic matter coupled with nutrient-rich water plume from offshore can accelerate  $CO_2$  production. Chen et al. (2008) concluded that aerobic respiration of heterotrophic ecosystems was the primary determinant of the high  $pCO_2$  in the inner Yangtze estuary. Moreover, the lower Yangtze River basin was highly populated. Inputs of acids from agricultural fertilizer, sewage, and acid deposition have also decreased pH and shifted the carbonate system towards  $CO_2$  (Duan et al., 2007; Chen et al., 2002), generating high  $pCO_2$  values regardless of its relatively low alkalinity.

# 4.3 Geomorphological controls on alkalinity and pCO<sub>2</sub>

To illustrate the geomorphological controls, the used 339 stations were aggregated by stream order based on their spatial positions. Both alkalinity and  $pCO_2$  showed a decreasing trend from the smallest headwater streams through tributaries to the Yangtze mainstream (Fig. 8). The average decrease of alkalinity and  $pCO_2$  were 94  $\mu$ mol L<sup>-1</sup> and 266  $\mu$ atm, respectively. Higher alkalinity and  $pCO_2$  in the headwater streams reveal the significance of direct terrestrial inputs of organic carbon and dissolved  $CO_2$  in controlling riverine carbon cycle. Over the study period, the Yangtze River watershed suffered severe soil erosion, averaging 2167 km<sup>-2</sup> yr<sup>-1</sup> (Wang et al.,

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2007b). Huge amounts of carbon were transported into the river system via erosion (Wu et al., 2007). Consequently, soil respiration and decomposition of the terrestrial-origin organic carbon have resulted in the CO<sub>2</sub> excess in the headwater streams (Li et al., 2012).

365 Figure 8

The decreasing  $pCO_2$  with increasing stream order imply continued  $CO_2$  evasion along the river continuum and reduced supply of fresh  $CO_2$ . Except the three lakes connected to the mainstream (Fig. 1a), the Yangtze River network is largely confined to its channel. Without large floodplains supplying labile organic matter to sustain high  $pCO_2$  as in the Amazon River (Mayorga et al., 2005), its  $pCO_2$  decreased progressively from the headwaters towards the mainstream. In addition, it is interesting to note that the  $pCO_2$  in the highest three orders was equivalent (~1800  $\mu$  matm; Fig. 8). Instead of continuous decline, the stable  $pCO_2$  suggests a balance between  $CO_2$  evasion and supply of fresh  $CO_2$  from upstream catchments or aquatic respiration. Contrary to the headwater streams with close contact with terrestrial ecosystems, the downstream large streams and rivers are far away from rapid fresh  $CO_2$  input. Moreover, these large streams and rivers are generally characterized by low gas transfer velocity owing to weakened turbulence and mixing with benthic substrates (Butman and Raymond, 2011;Borges et al., 2015). This can effectively inhibit  $CO_2$  degassing and therefore maintain the balance.

It is important to note, however, that the delineated 8 stream orders may not necessarily represent the actual stream network. Limited by spatial resolution, the smallest headwater streams might have been missed from the identified river network. In addition, these headwater streams are also

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generally absent of sampling stations. With much closer biogeochemical interactions with land ecosystems, these missed headwater streams tend to have higher  $pCO_2$  (Benstead and Leigh, 2012;Aufdenkampe et al., 2011). Thus, the actual  $pCO_2$  gradient along the stream order may be sharper if a higher  $pCO_2$  in the headwater streams is included.

# 4.4 Implications for riverine CO<sub>2</sub> evasion

As mentioned earlier, riverine carbon transport has been a significant component of carbon cycle. Quantifying riverine carbon export is essential to better evaluate global carbon budget and elucidate the magnitude of carbon exchange between different carbon pools. For the estimation of  $CO_2$  evasion, riverine  $pCO_2$  denotes  $CO_2$  concentration gradient across the water-air interface and thus the potential of  $CO_2$  exchange. Prior studies indicate that elevated riverine  $pCO_2$  can enhance  $CO_2$  evasion due to a steeper concentration gradient and a greater  $CO_2$  availability for degassing (Long et al., 2015;Billett and Moore, 2008). When assessing global-scale  $CO_2$  evasion, however, the spatial distribution of  $pCO_2$  is heavily skewed towards Northern America, Europe, and Australia (e.g., Lauerwald et al., 2015; Raymond et al., 2013), while data for Asian rivers are extremely lacking. This absence of an equally distributed  $pCO_2$  database has made it challenging to accurately estimate global  $CO_2$  evasion. The role of Asian rivers in global carbon export explicitly demonstrates that under-representation of Asian rivers would cause huge biases.

Comparing the Yangtze River with other rivers shows that its  $pCO_2$  is higher than most world rivers (Table 4). The average  $pCO_2$  of 2662 µatm suggests that the Yangtze River waters are

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potentially a prominent carbon source for the atmosphere. Large CO<sub>2</sub> evasion fluxes have been reported by several small-scale studies in the upper reach and the estuary (Zhai et al., 2007;Chen et al., 2008;Li et al., 2012), as also shown in Table 4. Nonetheless, a systematic estimation of CO<sub>2</sub> evasion from the whole Yangtze River network, including mainstream and its tributaries of all orders, remains lacking. This has further hampered the assessment of its CO<sub>2</sub> evasion in a wider context linking the watershed's land-atmosphere and land-ocean carbon exchanges.

410 **Table 4** 

Accelerated human activity is another urgent issue to be considered when investigating its riverine *p*CO<sub>2</sub> and CO<sub>2</sub> evasion. Approximately 50,000 dams, including the world's largest reservoir (i.e., the Three Gorges Reservoir; TGR), have been constructed in recent decades (Xu and Milliman, 2009). Assessing the impacts of dam-triggered changes to flow regime and biogeochemical processes on *p*CO<sub>2</sub> and CO<sub>2</sub> evasion is particularly important for deeper insights into its riverine carbon cycle (Table 4). For example, while the *p*CO<sub>2</sub> at Datong station declined continuously before the TGR impoundment (Fig. 4a; Wang et al., 2007a), our recent field survey shows that it has recovered from 1440 μatm in the 1980s to present 1700 μatm (see Fig. 4a). As for CO<sub>2</sub> degassing, recent work in the TGR indicates that its CO<sub>2</sub> evasion fluxes are different from natural rivers and are higher than other temperate reservoirs (Table 4; Zhao et al., 2013). Future research efforts are warranted to conduct systematic monitoring and evasion estimation. Given the Yangtze River's role in global carbon export, a comprehensive assessment of CO<sub>2</sub> evasion is also meaningful for global carbon budget.

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# 425 Conclusions

The degassing of CO<sub>2</sub> from inland waters was recently incorporated into assessments of the global carbon budget. Because direct degassing measurements are time-consuming and unrealistic at large spatial scales, CO<sub>2</sub> efflux from aquatic systems is generally estimated from the water-air gradient of *p*CO<sub>2</sub> and gas transfer velocity. As an important parameter for calculating CO<sub>2</sub> evasion, riverine *p*CO<sub>2</sub> is affected by a variety of factors and could vary substantially over space and time. Based on water chemistry data measured in the Yangtze River basin during the period 1960s–1985, we calculated the *p*CO<sub>2</sub> from pH and alkalinity. The pH in the Yangtze River waters varied from 6.5 to 9.2 and the alkalinity ranged from 415 to >3400 μmol L<sup>-1</sup> with high alkalinity concentrations occurring in carbonate-rich tributary catchments.

Except one station in the upper reach showing a lower *p*CO<sub>2</sub> than the atmosphere, the Yangtze River waters were supersaturated with dissolved CO<sub>2</sub>, generally 2–20 folds the atmospheric equilibrium. Averaged over all stations, the basin-wide *p*CO<sub>2</sub> was 2662±1240 μatm.

The observed spatial and temporal variations of  $pCO_2$  were collectively controlled by terrestrial ecosystems, hydrological regime, and rock weathering. High  $pCO_2$  values were observed spatially in catchments with abundant carbonate presence and seasonally in the wet season when recent-fixed organic matter was flushed into the river network. Decomposition of organic matter by microbial activity in aquatic systems facilitated  $CO_2$  production and sustained the high  $pCO_2$  values in the wet season, although the alkalinity presented a significant dilution effect with water discharge. In addition, the  $pCO_2$  decreased with increasing stream orders from the smallest

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headwater streams through tributaries to the mainstream. A higher  $pCO_2$  in the headwater streams illustrated the influence of direct inputs of terrestrially-derived organic matter and weathering products via erosion and flushing on riverine carbon dynamics.

450 The substantially higher  $pCO_2$  than the atmosphere indicated a potential of significant  $CO_2$  evasion from the Yangtze River fluvial network. Estimating the amount of  $CO_2$  evasion should be a top priority, upon which its biogeochemical implications for watershed-scale carbon cycle can be assessed in association with carbon burial and downstream export. Given the significant anthropogenic perturbations in recent decades, special attention must be paid to the resulting changes to riverine  $pCO_2$  and  $CO_2$  evasion. Considering the Yangtze River's relevance to global carbon export, quantifying its  $CO_2$  evasion is of paramount importance for better assessments of global carbon budget.

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Table 1. Comparison of alkalinity (meq L<sup>-1</sup>) and pH at Wuhan station between the GEMS/Water Programme results and the hydrological yearbooks, expressed as mean±standard error.

| <del>-</del>           | 1000            | 1001          | 1000          | 1000          | 1001          | 1001            |  |
|------------------------|-----------------|---------------|---------------|---------------|---------------|-----------------|--|
| Item                   | 1980            | 1981          | 1982          | 1983          | 1984          | 1984            |  |
| GEMS/Water Programme   |                 |               |               |               |               |                 |  |
| Alkalinity             | $2.05\pm0.29$   | $2.00\pm0.19$ | 2.01±0.23     | $1.84\pm0.25$ | $2.20\pm0.25$ | $1.99\pm0.22$   |  |
| pН                     | $7.83 \pm 0.16$ | $7.73\pm0.24$ | $8.04\pm0.09$ | $8.06\pm0.05$ | $8.00\pm0.09$ | $7.88 \pm 0.06$ |  |
| Hydrological yearbooks |                 |               |               |               |               |                 |  |
| Alkalinity             | $2.31\pm0.31$   | $2.19\pm0.24$ | $2.27\pm0.27$ | $2.03\pm0.30$ | $2.38\pm0.28$ | $2.31\pm0.24$   |  |
| pН                     | $7.93\pm0.09$   | $7.87 \pm 09$ | 8.01±0.09     | $7.94\pm0.08$ | 7.93±0.10     | $7.98\pm0.08$   |  |

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Table 2. Riverine pH, alkalinity, and  $pCO_2$  in the Yangtze River basin (median $\pm$ standard deviation).

| River/tributary            | Station              | рН              | Alkalinity           | $p\mathrm{CO}_2$ |
|----------------------------|----------------------|-----------------|----------------------|------------------|
| Kivei/tiloutary            | Station              | pm              | μmol L <sup>-1</sup> | μatm             |
|                            | Benzilan             | $8.29\pm0.11$   | 2352±435             | 681±156          |
|                            | Shigu                | $8.18\pm0.48$   | 2544±438             | 846±262          |
|                            | Jingjiangjie         | $8.11\pm0.12$   | 2905±362             | 916±202          |
|                            | Dukou                | $8.22\pm0.12$   | 2399±429             | 826±197          |
|                            | Longjie              | 8.23±0.17       | 2185±396             | 786±226          |
|                            | Huatan               | $8.17 \pm 0.15$ | 2237±418             | 882±287          |
|                            | Pingshan             | 8.13±0.10       | 2215±407             | 1001±235         |
| Mainstream                 | Zhutuo               | $7.88\pm0.19$   | 2299±349             | 2405±781         |
| Manistream                 | Cuntan               | $8.08\pm0.11$   | 2173±311             | 1087±319         |
|                            | Yichang              | 7.95±0.15       | 2343±300             | 1653±469         |
|                            | Luoshan              | 7.76±0.11       | 2280±248             | 2380±691         |
|                            | Wuhan                | 7.93±0.11       | 2060±263             | 1521±497         |
|                            | Datong               | $7.84 \pm 0.14$ | 1919±312             | 1711±806         |
|                            | Nanjing <sup>a</sup> | 7.56±0.16       | 2339±339             | 3796±1623        |
|                            | Nanjing <sup>b</sup> | $7.54 \pm 0.18$ | 2296±357             | 3793±2186        |
| Major tributaries          | c                    |                 |                      |                  |
| Yalongjiang                | Xiaodeshi            | $8.02\pm0.22$   | 2576±465             | 1567±715         |
| Daduhe                     | Fuluzhen             | $7.66 \pm 0.23$ | 1909±289             | 2577±1620        |
| Minjiang                   | Gaochang             | $8.02\pm0.15$   | 1816±327             | 1020±525         |
| Tuojiang                   | Lijiawan             | 8.01±0.11       | 2705±507             | 1504±572         |
| Jialingjiang               | Beibei               | $8.11\pm0.14$   | 2289±509             | 1196±244         |
| Wujiang                    | Wulong               | $8.01\pm0.14$   | 2420±279             | 1361±508         |
| Yuanjiang                  | Taoyuan              | $7.61\pm0.25$   | 1822±480             | 2801±2144        |
| Xiangjiang                 | Xiangtan             | $7.76 \pm 0.44$ | 1739±331             | 2349±2521        |
| Hanjiang                   | Xiaoshicun           | 7.93±0.13       | 2262±480             | 1715±536         |
| Ganjiang                   | Waizhou              | $7.44 \pm 0.44$ | 880±236              | 2205±2048        |
| Yangtze basin <sup>d</sup> | 1% percentile        | 7.03            | 556                  | 788              |
|                            | 10% percentile       | 7.35            | 842                  | 1236             |
|                            | 50% percentile       | 7.71            | 2237                 | 2455             |
|                            | 90% percentile       | 8.05            | 3305                 | 4344             |
|                            | 99% percentile       | 8.28            | 4437                 | 6163             |

<sup>&</sup>lt;sup>a</sup>affected by high tides.

<sup>&</sup>lt;sup>b</sup>affected by low tides.

<sup>&</sup>lt;sup>c</sup>Median values of the data for the lowermost station on the mainstream of the specific tributary.

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<sup>d</sup>Statistics based on the measurements at the used 339 stations.

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Table 3. Hydro-geochemical features of the Wujiang (Wulong station), Jialingjiang (Wusheng station), and Ganjiang (Xiajiang station) catchments.

| Control  | Control<br>area | Water<br>discharge | рН            | Alkalinity         | $p\mathrm{CO}_2$ | Ca <sup>2+</sup>   | SiO <sub>2</sub>     | Roc       | k types (% | of area)    |
|----------|-----------------|--------------------|---------------|--------------------|------------------|--------------------|----------------------|-----------|------------|-------------|
| station  | km <sup>2</sup> | $m^3 s^{-1}$       | pm            | $\mu$ mol $L^{-1}$ | μatm             | $\mu$ mol $L^{-1}$ | μmol L <sup>-1</sup> | Carbonate | Silicate   | Igneous +   |
|          |                 |                    |               |                    |                  |                    |                      |           |            | metamorphic |
| Wulong   | 80,536          | 1570               | $7.72\pm0.14$ | 3021±527           | 3537±1247        | 1145±278           | 59±31                | 82.9      | 14.8       | 2.3         |
| Wusheng  | 80,550          | 793                | $7.80\pm0.21$ | $2484\pm948$       | 2671±490         | 1005±170           | 94±30                | 30.4      | 55.3       | 14.3        |
| Xiajiang | 62,387          | 1644               | $7.34\pm0.08$ | 953±266            | 2642±626         | 242±91             | 105±18               | 9.1       | 64.7       | 26.2        |

Table 4. Comparison of  $pCO_2$  and  $CO_2$  evasion among world large rivers and typical reservoirs in the Yangtze River basin.

| River   | Country | Climate             | $pCO_2$        | CO <sub>2</sub> evasion | Reference               |  |  |
|---|---------|---------------------|----------------|-------------------------|-------------------------|--|--|
| Kivei   | Country | Cilliate            | μatm           | $mol m^{-2} yr^{-1}$    |                         |  |  |
| Yangtze network                               | China   | Subtropical monsoon | 2662±1240      | /                       | This study              |  |  |
| Upper Yangtze                                 | China   | Subtropical monsoon | 2100           | 57                      | Li et al., 2012         |  |  |
| Lower Yangtze                                 | China   | Subtropical monsoon | $1297 \pm 901$ | 14.2-54.4               | Wang et al., 2007a      |  |  |
| Yangtze estuary                               | China   | Subtropical monsoon | 650-1440       | 15.5-34.2               | Zhai et al., 2007       |  |  |
| Amazon  | Brazil  | Tropical            | 3929           | 162.2                   | Lauerwald et al. 2015   |  |  |
| Ottawa  | Canada  | Temperate           | 1200           | 14.2                    | Telmer and Veizer, 1999 |  |  |
| Hudson  | USA     | Temperate           | $1125 \pm 403$ | 5.8-13.5                | Raymond et al., 1997    |  |  |
| York estuary                                  | USA     | Temperate           | $1070\pm\!867$ | 6.3                     | Raymond et al., 2000    |  |  |
| Mississippi                                   | USA     | Temperate           | 1335±130       | 98.5±32.5               | Dubois et al., 2010     |  |  |
| Yukon   | Canada  | Subarctic           | 582-705        | 11.6-21.2               | Lauerwald et al., 2015  |  |  |
| Yellow  | China   | Arid and semiarid   | 2810±1985      | 312.4±149.2             | Ran et al., 2015b       |  |  |
| Xijiang (Pearl)                               | China   | Subtropical monsoon | 2600           | 69.2-130                | Yao et al., 2007        |  |  |
| Mekong (>100 m                                | SE Asia | Tropical monsoon    | 703-1597       | 32-138                  | Alin et al., 2011       |  |  |
| wide rivers)                                  |         |                     |                |                         |                         |  |  |
| Godavari estuary                              | India   | Tropical monsoon    | <500-33,000    | 52.6                    | Sarma et al., 2011      |  |  |
| Global rivers                                 |         |                     | 2400           | 131.2                   | Lauerwald et al. 2015   |  |  |
| Typical reservoirs in the Yangtze River basin |         |                     |                |                         |                         |  |  |
| Wujiang cascade                               |         |                     | 38-3300        | -3.3-32.5               | Wang et al., 2011       |  |  |
| reservoirs                                    |         |                     |                |                         |                         |  |  |
| Three Gorges                                  |         |                     | /              | 35.1                    | Zhao et al., 2013       |  |  |
| Reservoir (TGR)                               |         |                     |                |                         |                         |  |  |

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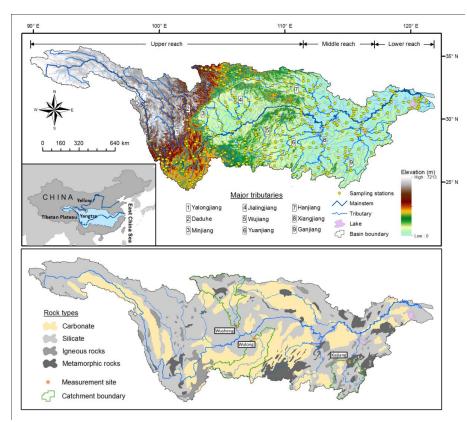


Fig. 1. Maps of the Yangtze River basin showing sampling stations (top) and rock compositions (bottom). Rock information is modified from Chen et al. (2002) and Chetelat et al. (2008).

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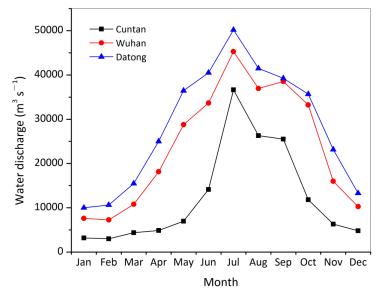


Fig. 2. Monthly variations in water discharge of the Yangtze River at Cuntan (upper reach), Wuhan (middle reach), and Datong stations (lower reach).





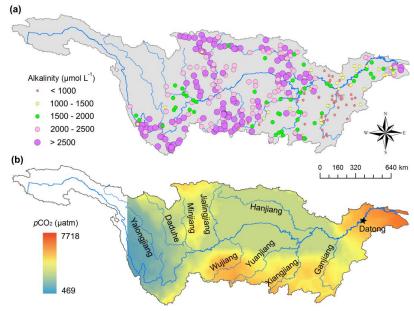


Fig. 3. Spatial distribution of alkalinity (a) and  $pCO_2$  (b) in the Yangtze River basin. The headwater region in (b) was not interpolated due to insufficient stations.

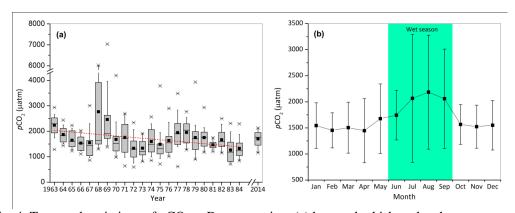
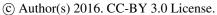


Fig. 4. Temporal variations of  $pCO_2$  at Datong station. (a) box-and-whisker plot shows significant inter-annual changes; (b) seasonal variations. The dash line in (a) represents linear regression and the values for 2014 are derived from Liu et al. (2016). Error bars denote standard deviation.







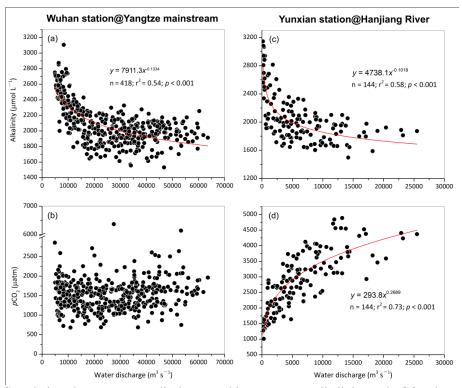


Fig. 5. Correlations between water discharge and instantaneous alkalinity and  $pCO_2$ : the mainstream at Wuhan station (a and b) and the Hanjiang River at Yunxian station (c and d).

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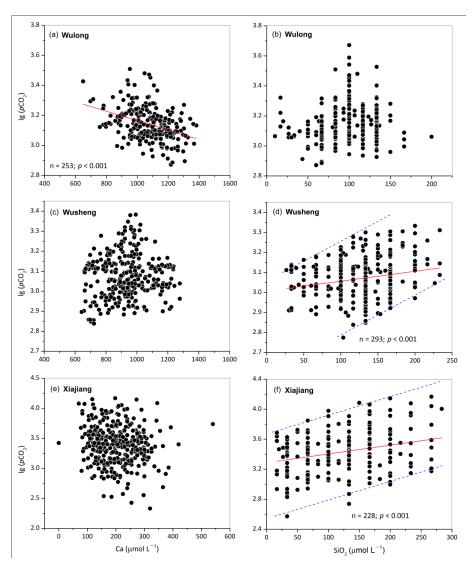


Fig. 6. Responses of *p*CO<sub>2</sub> to rock weathering products in three typical catchments with distinct rock compositions: a–b: Wujiang River (Wulong station); b–c: Jialiangjiang River (Wusheng station); e–f: Ganjiang River (Xiajiang station). The solid lines represent linear regression.

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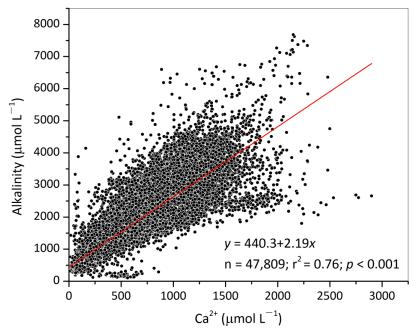


Fig. 7. Strong correlation between chemical weathering, using  $\text{Ca}^{2+}$  as a proxy, and alkalinity.

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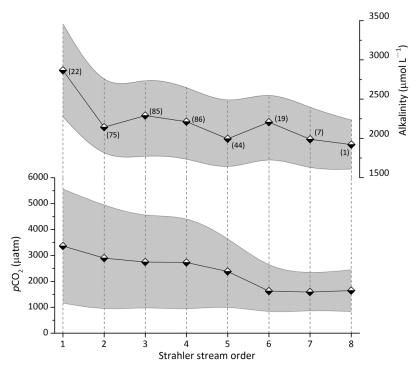


Fig. 8. Decreasing alkalinity (top) and  $pCO_2$  (bottom) with increasing Strahler stream order. The grey shade denotes standard deviation and the numbers in parentheses represent the number of stations aggregated for each stream order.