Response to reviewers for revised manuscript bg-2016-64

Referee # 1 (Orit Sivan)

Referee's comment: Suggest to accept as is.

Author's reply: We thank the referee for her insightful review that improved the quality of this paper.

Referee # 2 (anonymous)

Referee's comment: The authors have done an excellent job in revising the paper in response to the reviewers' comments. Most importantly in my view, they have justified the assumptions and reasons for excluding organoclastic iron reduction in the methane zone, making it clear that the purpose of the study is to explore the potential for iron-dependent AOM rather than determining actual Fe-AOM rates. I only find one minor issue:

I still respectfully disagree with the use of the closed-system Rayleigh distillation model to determine fractionation factors for sulfate-AOM (l. 428-441). A remark is now included that determination of accurate factors require modelling, which is fine, but I am confused by the remark that "diffusion could be slower than oxidation". This true in the sense that some methanogenesis occurs in the AOM zone, but AOM rates in the SMTZ are about an order of magnitude higher than methane production rates there (Fig. 6), so most of the methane must be supplied by diffusion – this is by no means a "quasi-closed system" as claimed in the text.

Just a remark: I my comments to the 1st version I had missed the fact that hydrogenotrophic methanogenesis in the model depends on the DIC concentration and not on the hydrogen concentration. This is a strange formulation since methanogens in sediments usually deplete hydrogen to the energetic threshold (see papers by Hoehler et al.) and thus, expectedly, are hydrogen limited (with hydrogen supply limited by the degradation of organic matter and, possibly, pyrite formation). Thus, the modelling of two separate methanogenic pathways is questionable, but this has no real influence on the paper now, because the authors have chosen to present the combined methane production.

Author's reply: We thank the referee for the compliment and for the critical remark regarding the modeling of hydrogenotrophic methanogenesis. We acknowledge the referee's view and have removed our previous interpretation of a "quasi-closed system" from the manuscript (lns 431-436).

Referee # 3 (W.-L. Hong)

Referee's comment: The authors have addressed all points I raised previously. It's now almost flawless. I would however like to make a final note on the numerical issue (my very first point in the review). I believe ReacTran is doing a good job numerically but numerical error is simply inevitable in many cases. It's something all modelers should acknowledge and be honest about. A better understanding about how large such error might be is therefore very important. TVD is such a widely used scheme for solving transport-reaction equations and is also a good numerical scheme. It is however not immune from numerical issues if a proper discretization is not chosen. I encourage the authors to refer to studies such as Wang and Hutter (2001) (DOI: 10.1002/fld.197) or Toth and Odstrcil (1996) (doi:10.1006/jcph.1996.0197) which comparing different numerical schemes including TVD. As the authors replied: "The ReacTran package accounts for numerical dispersion using total variation diminishing (TVD) slope limiters". This statement does not guarantee everything. The only way to check numerical issues is to run the model with some clean scenario (such as advection of a square function).

Author's reply: We thank the referee for the positive remarks and his helpful input regarding numerical errors. We will carefully consider the studies mentioned by the referee in our future modeling work.

Anaerobic oxidation of methane alters sediment records of sulfur, iron and phosphorus in the Black Sea

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- 13 **Abstract.** The surface sediments in the Black Sea are underlain by extensive deposits of iron (Fe) oxide-rich lake
- sediments that were deposited prior to the inflow of marine Mediterranean Sea waters ca. 9000 years ago. The
- subsequent downward diffusion of marine sulfate into the methane-bearing lake sediments has led to a multitude of
- diagenetic reactions in the sulfate-methane transition zone (SMTZ), including anaerobic oxidation of methane
- 17 (AOM) with sulfate. While the sedimentary cycles of sulfur (S), methane and Fe in the SMTZ have been extensively
- studied, relatively little is known about the diagenetic alterations of the sediment record occurring below the SMTZ.
- 19 Here we combine detailed geochemical analyses of the sediment and pore water with multicomponent diagenetic
- 20 modeling to study the diagenetic alterations below the SMTZ at two sites in the western Black Sea. We focus on the
- dynamics of Fe, S and phosphorus (P) and demonstrate that diagenesis has strongly overprinted the sedimentary
- burial records of these elements. In line with previous studies in the Black Sea, we show that sulfate-mediated AOM
- 23 substantially enhances the downward diffusive flux of sulfide into the deep limnic deposits. During this downward
- 24 sulfidization, Fe oxides, Fe carbonates and Fe phosphates (e.g. vivianite) are converted to sulfide phases, leading to
- an enrichment in solid phase S and the release of phosphate to the pore water. Below the sulfidization front, high
- 26 concentrations of dissolved ferrous Fe (Fe²⁺) lead to sequestration of downward diffusing phosphate as authigenic
- 27 vivianite, resulting in a transient accumulation of total P directly below the sulfidization front.
- Our model results further demonstrate that downward migrating sulfide becomes partly re-oxidized to sulfate due to
- 29 reactions with oxidized Fe minerals, fueling a cryptic S cycle and thus stimulating slow rates of sulfate-driven AOM
- 30 (~ 1 100 pmol cm⁻³ d⁻¹) in the sulfate-depleted limnic deposits. However, this process is unlikely to explain the
- 31 observed release of dissolved Fe²⁺ below the SMTZ. Instead, we suggest that besides organoclastic Fe oxide
- 32 reduction and reactivation of less reactive Fe oxides by methanogens, AOM coupled to the reduction of Fe oxides
- may also provide a possible mechanism for the high concentrations of Fe²⁺ in the pore water at depth. Our results
- reveal that methane plays a key role in the diagenetic alterations of Fe, S and P records in Black Sea sediments. The
- downward sulfidization into the limnic deposits is enhanced through sulfate-driven AOM with sulfate and AOM with
- 36 Fe oxides may provide a deep source of dissolved Fe²⁺ that drives the sequestration of P in vivianite below the
- 37 sulfidization front.

1 Introduction

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Anaerobic oxidation of methane (AOM), a process initially regarded as a biogeochemical curiosity, functions as an 39 40 important sink for oceanic methane (CH₄) by consuming > 90 % of all CH₄ produced in marine sediments (Knittel 41 and Boetius, 2009; Reeburgh, 2007). Although recent studies indicate that the biological oxidation of CH₄ could be 42 coupled to various additional electron acceptors such as nitrate and nitrite (Ettwig et al., 2010; Raghoebarsing et al., 2006) as well as metal oxides (Beal et al., 2009; Egger et al., 2015b; Riedinger et al., 2014; Scheller et al., 2016; 43 Segarra et al., 2013; Sivan et al., 2011), sulfate (SO₄²⁻) is commonly thought to be the dominant electron acceptor in 44 anoxic marine systems (Knittel and Boetius, 2009; Reeburgh, 2007). 45 46 Nevertheless, a coupling between anaerobic CH₄ oxidation and iron (Fe) oxide reduction (Fe-AOM) could have a 47 significant impact on sedimentary Fe cycling and related processes such as phosphorus (P) diagenesis, because of the 8:1 Fe-CH₄ stoichiometry of the reaction (Beal et al., 2009; Egger et al., 2015a; Rooze et al., 2016). Environmental 48 49 conditions that favor Fe-AOM in marine systems are still poorly understood. The required co-occurrence of pore 50 water CH₄ and abundant reducible Fe oxides suggests that Fe-AOM may occur in sediments that receive a relatively 51 high input of Fe oxides compared to the in-situ production of sulfide, which could allow a portion of Fe oxides to 52 escape the conversion to authigenic Fe sulfides and to remain preserved in the methanogenic sediments below the zone of SO₄²⁻ reduction (Egger et al., 2015b; Riedinger et al., 2014; Rooze et al., 2016). In addition, perturbations 53 inducing transient diagenesis such as anthropogenic eutrophication or climate change may also create diagenetic 54 55 environments that are likely favorable for Fe-AOM, as they provide a mechanism for the burial of Fe oxide-rich 56 deposits below sulfidic sediment layers (Egger et al., 2015b; Riedinger et al., 2014). 57 The Black Sea represents a good example of a sedimentary system in which transient diagenesis associated with 58 postglacial sea-level rise has led to the accumulation of sulfidic sediments above Fe oxide-rich deposits. Here, the 59 establishment of a connection to the Mediterranean Sea through the shallow Bosporus around 9000 years ago 60 (Degens and Ross, 1974; Soulet et al., 2011) led to the inflow of marine waters into a freshwater basin, resulting in 61 permanent salinity/density stratification and in the development of euxinic conditions (i.e. free dissolved sulfide 62 present in the bottom water), making the current Black Sea the largest permanently anoxic basin on Earth. In the absence of oxygen and metal oxides, SO_4^{2-} reduction is the dominant benthic mineralization process of organic 63 matter in Black Sea surface sediments below the chemocline (~ 100 m depth) (Jørgensen et al., 2001; Thamdrup et 64 al., 2000). At present, SO_4^{2-} penetrates through the modern coccolith ooze (Unit I) and the marine sapropel (Unit II) 65 sediments and a few meters into the Upper Pleistocene freshwater deposits (Unit III) (Arthur and Dean, 1998; 66 Degens and Ross, 1974; Jørgensen et al., 2004). Below the SO₄²-bearing zone, methanogenesis takes over as the 67 dominant process of organic matter degradation, resulting in the buildup of CH₄ in the pore water at depth. 68 Interactions between the cycles of sulfur (S) and CH₄ in Black Sea sediments have been extensively studied during 69 recent years (Holmkvist et al., 2011b; Jørgensen et al., 2001, 2004; Knab et al., 2009; Leloup et al., 2007) and AOM 70 coupled to SO₄² reduction (SO₄-AOM) was found to account for an estimated 7-18 % of total SO₄² reduction in 71 these sediments (Jørgensen et al., 2001). The production of sulfide in the sulfate-methane transition zone (SMTZ) as 72 73 a result of SO₄-AOM represents the main source of pore water sulfide at depth in the sediment. This intensified production of sulfide drives an enhanced downward diffusive flux of sulfide into the deep limnic deposits of Unit III, 74

forming a distinct diagenetic sulfidization front recognized as a black band or a series of bands owing to the

conversion of Fe oxides to Fe sulfides (Berner, 1974; Jørgensen et al., 2004; Neretin et al., 2004). 76 77 At present, the impact of the downward-migrating sulfidization front on sedimentary P, a key nutrient for marine phytoplankton, and the potential role of Fe-mediated AOM in the deep limnic deposits remain largely unknown. A 78 79 buildup of ferrous Fe (Fe²⁺) in the pore water at depth as found in previous studies (Holmkvist et al., 2011b; Jørgensen et al., 2004; Knab et al., 2009), could indicate ongoing Fe reduction in the CH₄-bearing deep limnic 80 81 sediments and thus a potential coupling between AOM and Fe oxide reduction. The sediment records investigated up 82 to now, however, do not extend deep enough to allow the sedimentary cycling of Fe and related biogeochemical 83 processes below the sulfidization front to be investigated. In particular, the presence of abundant dissolved Fe²⁺ combined with a potential release of pore water phosphate (HPO₄²⁻) during reductive dissolution of Fe oxides may be 84 conducive to the formation of reduced Fe(II)-P minerals such as vivianite (Fe₃(PO)₄*8H₂O) below the sulfidization 85 86 front (Egger et al., 2015a; Hsu et al., 2014; März et al., 2008; Sivan et al., 2011). Post-depositional diagenetic 87 alterations as a result of downward sulfidization could therefore overprint burial records of P in the Upper 88 Pleistocene deposits. 89 In this study, we combine detailed geochemical analyses of the sediment and pore water with multicomponent 90 diagenetic modeling to study the diagenetic alterations below the lake-marine transition at two sites in the western Black Sea. Focusing on the dynamics of S, Fe and P, we demonstrate that AOM coupled to SO₄² reduction enhances 91 92 the downward sulfidization and associated dissolution of Fe oxides, Fe carbonates and vivianite, supporting earlier 93 findings of an SO₄-AOM enhanced downward sulfidization in Black Sea sediments (Jørgensen et al., 2001). Below the sulfidization front, downward diffusing HPO₄²- precipitates as vivianite by reaction with the abundant dissolved 94 Fe²⁺. We propose that organoclastic Fe oxide reduction, reactivation of less reactive Fe oxides by methanogens 95 96 (Sivan et al., 2016) and/or AOM coupled to the reduction of Fe oxides are the key processes explaining the high

concentrations of dissolved Fe²⁺ at depth in the sediment. Trends in total S and P with depth are significantly altered

by the above-mentioned reactions, highlighting that diagenesis may strongly overprint burial records of these

99 elements below a lake-marine transition.

2 Materials and methods

2.1 Sample collection

102 **2.1.1** Gravity core sampling

Sediment samples were taken at two slope sites in the western Black Sea during a cruise in June 2013 with R/V Pelagia. Gravity cores containing ~ 7 m of sediment were collected at sites 4 (43°40.6' N, 30°7.5' E; 377 meters below sea surface (mbss)) and 5 (43°42.6' N, 30°6.1' E; 178 mbss) (Fig. 1), both situated below the current chemocline (~ 100 m water depth). The core liners were pre-drilled with 2 cm diameter holes in two rows of 10 cm resolution on opposing sides of the tube, offset by 5 cm and taped prior to coring. Upon recovery, the liners were cut into 1 m sections, transferred to a temperature-controlled container set at in-situ bottom water temperature (11 °C)

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and secured vertically. Subsequently, the taped holes were cut open and a cut-off syringe was inserted horizontally directly after opening each hole.

From one series of holes, 10 mL of wet sediment was extracted at 20 cm resolution and immediately transferred into a 65 mL glass bottle filled with saturated NaCl solution for CH₄ analysis. The NaCl solution was topped up after addition of the sample, ensuring that no air bubbles remained. Each bottle was sealed with a black rubber stopper and a screw cap and was subsequently stored upside-down at room temperature. From the second series of holes, 20 mL sediment was extracted at 20 cm resolution, sealed with parafilm that was tightly closed with an elastic band, and directly inserted into a nitrogen (N₂)-purged glove box. Subsequently, the sediment was transferred into a 50 mL centrifuge tube and centrifuged (4500 rpm; 30 min). The supernatant from each centrifuged sample was filtered through 0.45 µm pore size disposable filters via 20 mL plastic syringes in the glove box and collected in 15 mL centrifuge tubes. The sediment fraction was stored frozen (-20 °C) for solid phase analysis. Filtered pore water samples were sub-sampled under N₂ for analysis of dissolved HPO₄²⁻, ammonium (NH₄⁺), dissolved inorganic carbon (DIC), Fe, manganese (Mn), SO₄²⁻ and sulfide (Σ H₂S = H₂S + HS⁻) (see section 2.2) Additional samples of 10 mL of sediment were collected at approximately 50 cm resolution and transferred into pre-weighed 15 mL glass vials to determine porosity from gravimetric water loss.

2.1.2 Multicore sampling

To sample the surface sediment, sediment cores (30-60 cm of sediment and at least 10 cm of overlying water) were recovered using an octopus multicorer (core diameter 10 cm). After recovery, the cores were stoppered at the base and at the top and immediately transported to a temperature-controlled container (11 °C). One multicore from each cast was pre-drilled with 2 cm diameter holes in two rows at 10 cm resolution on opposing sides of the tube, offset by 5 cm, and taped prior to coring. These holes were sampled for CH₄ as described for the gravity cores. Another core was directly inserted into a N₂-purged glove box through an airtight hole in the base. A bottom water sample was collected using a 20 mL plastic syringe and the remaining bottom water was removed with a Tygon tube. Subsequently, the core was sliced anoxically with decreasing resolution at depth, i.e. 0.5 cm resolution for the first 0-2 cm, 1 cm resolution between 2-10 cm, 2 cm resolution between 10-20 cm and 4 cm resolution for the rest of the core (> 20 cm). For each slice a sub-sample was placed in a pre-weighed 15 mL glass vial for water content and solid phase analysis and stored under N₂ in airtight jars at -20 °C. A second sub-sample was transferred to a 50 mL centrifuge tube and centrifuged (4500 rpm; 30 min). Both the supernatant water from each centrifuged sample and the bottom water sample were subsequently processed as described for the gravity cores.

Visual alignment of the pore water profiles from the multicores with those of the gravity cores showed that the first ~ 20 to 30 cm of sediment was lost during long coring. At site 5, the sediment in the multicore consisted of a gray and homogeneous turbidite below 1.5 cm depth. The depth for the gravity core at site 5 was thus corrected for the loss of the marine deposits, which were previously reported to be about 50 cm thick at a site in close proximity to site 5 (43°42.63' N, 30°6.12' E; 181 mbss) (Jørgensen et al., 2004)

2.2 Pore water subsampling

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A sub-sample of 0.5 mL was immediately transferred into a glass vial containing 1.5 mL of 8 M NaOH solution for 144 145 analysis of dissolved sulfide. Sub-samples for total dissolved Fe and Mn, which are assumed to represent Fe(II) and 146 Mn(II), were acidified with 10 μL 35 % suprapur HCl per mL of sub-sample. Note, however, that the dissolved (< 0.45 µm) Fe and Mn pools likely consist of a mixture of truly dissolved (aqueous), as well as organically complexed, 147 colloidal and nanoparticulate Fe and Mn species (Raiswell and Canfield, 2012). Another 1 mL of pore water for 148 HPO_4^{2-} analysis was acidified with 4 μL 5 M HCl. Pore water SO_4^{2-} was analyzed with ion chromatography (IC) in a 149 10-fold diluted sample (0.15 mL of pore water with 1.35 mL of de-oxygenated UHQ water). Sub-samples for DIC 150 151 analysis (0.5 mL) were collected in glass vials (4.9 mL) to which 4.4 mL of 25 g/L NaCl solution was added, making 152 sure that no headspace remained. Aliquots of the remaining pore water were used for the measurement of alkalinity (determined onboard by titrating 1 mL of untreated sub-sample with 0.01 M HCl; results presented in the 153 Supplementary Information only) and NH₄⁺. All sub-samples were stored at 4 °C and brought to room temperature 154 just before analysis. Subsampling for sulfide was performed immediately after filtration and all other subsampling 155 156 was performed within 4 hours of core recovery. Pore water sub-samples for HPO₄², DIC and sulfide were directly analyzed colorimetrically onboard on two separate 157 QuAAtro (SEAL Analytical, Germany) auto analyzers. HPO₄²⁻ was measured at 880 nm after the formation of 158 molybdophosphate-complexes (Murphy and Riley, 1962). Samples for DIC were acidified online after being 159 160 oxidized by H₂O₂ and analyzed as described by Stoll et al. (2001). To keep the dissolved sulfide in the non-volatile 161 HS form under alkaline conditions, 1.5 mL of 8 mM NaOH was added to the sulfide samples, which were 162 subsequently analyzed using the methylene blue method as described by Grasshoff (1969). Sub-samples for 163 dissolved Fe and Mn were analyzed onshore by ICP-OES (Perkin Elmer Optima 3000 Inductively Coupled Plasma -Optimal Emission Spectroscopy). For the analysis of pore water CH₄, a volume of 10 mL N₂ was injected into the 164 165 CH₄ serum flasks (while a needle inserted through the septum allowed 10 mL of water to escape) to create a 166 headspace from which a subsample was collected with a gas-tight syringe. Subsequently, CH₄ concentrations were determined in the home laboratory after injection into a Thermo Finnigan Trace GC gas chromatograph (Flame 167 Ionization Detector). δ¹³C-CH₄ and δD-CH₄ (D, deuterium) were analyzed by Continuous Flow Isotope Ratio Mass 168 Spectrometry (CF-IRMS) as described in detail in (Brass and Röckmann, 2010) and (Sapart et al., 2011). 169

2.3 Bulk sediment analysis

- Sediment samples were freeze-dried, powdered and ground in an agate mortar in an argon (Ar)-filled glove box and split into oxic and anoxic fractions. Samples from the oxic fraction were used for total elemental and organic carbon (C_{org}) analyses under normal atmospheric conditions, whereas anoxic splits for sediment P and Fe speciation were kept under an inert, oxygen-free Ar or N_2 atmosphere at all times to avoid oxidation artefacts (Kraal and Slomp,
- 175 2014; Kraal et al., 2009).

2.3.1 Total elemental composition and organic carbon

A split of ~ 125 mg of freeze-dried sediment was dissolved overnight in 2.5 mL HF (40 %) and 2.5 mL of HClO₄/HNO₃ mixture, in a closed Teflon bomb at 90 °C. The acids were then evaporated at 160 °C and the resulting gel was dissolved overnight in 1 M HNO₃ at 90 °C. Total elemental concentrations in the 1 M HNO₃ solutions were determined by ICP-OES. A second split of 0.3 g freeze-dried sediment was used to determine the C_{org} content using an elemental analyzer (Fison Instruments model NA 1500 NCS) after carbonate removal from the sediment with two washes with 1 M HCl (4 h and 12 h) followed by two washes with UHQ water and subsequent drying of the samples (Van Santvoort et al., 2002).

2.3.2 Sediment P fractionation

To determine the solid phase partitioning of P, aliquots of 0.1 g dried sediment were subjected to the SEDEX sequential extraction procedure after Ruttenberg (1992), as modified by Slomp et al. (1996b), but including the first MgCl₂ step (Table 1). Sediment P was fractionated as follows: i) exchangeable-P ("P_{exch}", extracted by 1 M MgCl₂, pH 8, 0.5 h), ii) Fe-associated P ("P_{Fe}", extracted by citrate-bicarbonate-dithionite (CDB), buffered to pH 7.5 with Na citrate/Na bicarbonate, 8 h, followed by 1 M MgCl₂, pH 8, 0.5 h), iii) authigenic Ca-P ("P_{authi Ca-P}", including carbonate fluorapatite, biogenic hydroxyapatite and CaCO₃-bound P, extracted by 1 M Na acetate solution, buffered to pH 4 with acetic acid, 6 h, followed by 1 M MgCl₂, pH 8, 0.5 h), iv) detrital Ca-P ("P_{detr}", extracted by 1 M HCl, 24 h) and v) organic P ("P_{org}", after ashing at 550 °C for 2 h, extracted by 1 M HCl, 24 h). The MgCl₂ washes in steps ii and iii were to ensure that any HPO₄²⁻ re-adsorbed during CDB or acetate extraction was removed and included in the pools of Fe-associated P and authigenic Ca-P, respectively. Sediments were shielded from oxygen inside an Ar-filled glovebox until step 3 of the SEDEX procedure to eliminate the potential conversion of Ca-P to Fe-bound P due to pyrite oxidation upon oxygen exposure (Kraal and Slomp, 2014; Kraal et al., 2009). Dissolved HPO₄²⁻ in the CDB solution was analyzed by ICP-OES. For all other solutions, HPO₄²⁻ was determined colorimetrically (Strickland and Parsons, 1972) on a Shimadzu spectrophotometer using the ammonium heptamolybdate – ascorbic acid method.

2.3.3 Sediment Fe fractionation

Sediment Fe was fractionated into i) carbonate associated Fe ("Fe_{carb}", including siderite and ankerite, extracted by 1 M Na-acetate brought to pH 4.5 with acetic acid, 24 h), ii) easily reducible (amorphous) oxides ("Fe_{ox1}", including ferrihydrite and lepidocrocite, extracted by 1 M hydroxylamine-HCl, 24 h), iii) reducible (crystalline) oxides ("Fe_{ox2}", including goethite, hematite and akagenéite, extracted by Na-dithionite buffer, pH 4.8, 2 h) and iv) Fe in recalcitrant oxides (mostly magnetite, "Fe_{mag}", extracted by 0.2 M ammonium oxalate / 0.17 M oxalic acid solution, 2 h), according to Poulton and Canfield (2005), using a 50 mg aliquot of dried sediment (Table 1). An additional aliquot of 50 mg was subjected to an adapted sequential extraction procedure after Claff et al. (2010), separating labile Fe(II) ("Fe(II)_{HCI}") and Fe(III) ("Fe(III)_{HCI}") using 1 M HCl (4 h) from crystalline Fe oxide minerals ("Fe(II)_{CDB}", Na-dithionite buffer, pH 4.8, 4 h) and from pyrite ("Fe_{pyrite}", concentrated nitric acid, 2 h), for all multicores as well as for the long core at site 4 (Table 1).

- 211 At site 4 (multicore only) and 5 (multicore and gravity core), aliquots of 0.5 g dried sediment were used to sequentially determine the amount of FeS (acid volatile sulfur, "AVS", using 6 M HCl) and FeS2 (chromium
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- 213 reducible sulfur, "CRS", using acidic chromous chloride solution) via the passive diffusion method described by
- 214 (Burton et al., 2008) using iodometric titration of the ZnS formed in the alkaline Zn acetate traps to quantify AVS
- 215 and CRS (Table 1).

2.4 Diagenetic model

217 2.4.1 General form

- 218 A multicomponent transient diagenetic model was developed for site 4 based on existing diagenetic models (Reed et
- 219 al., 2011a, 2011b; Rooze et al., 2016) to gain a better understanding of the transient diagenesis in Black Sea
- sediments and to investigate the potential for Fe-AOM as a source of pore water Fe²⁺ at depth. The model describes 220
- 221 the cycling of dissolved and particulate chemical species in a 1D sediment column (Berner, 1980). A total of 25
- 222 different chemical species (Table 2) were subjected to a suite of biogeochemical reactions (Table 3) and vertical
- 223 transport through burial, as well as molecular diffusion for dissolved species (Boudreau, 1997; Soetaert et al., 1996;
- Wang and Van Cappellen, 1996). The general diagenetic equations for solid (Eq. (1)) and dissolved species (Eq. (2)) 224
- 225 are, respectively,

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$$(1 - \phi) \frac{\partial c_S}{\partial t} = -(1 - \phi) v \frac{\partial c_S}{\partial x} + \sum R_S$$
 (1)

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$$\phi \frac{\partial c_{aq}}{\partial t} = \phi D' \frac{\partial^2 c_{aq}}{\partial x^2} - \phi u \frac{\partial c_{aq}}{\partial x} + \sum R_{aq}$$
 (2)

- where C_S is the concentration of the solid species (mol L⁻¹; mass per unit volume of solids), C_{aa} the concentration of 228
- the dissolved species (mol L⁻¹; mass per unit volume of pore water), t is time (yr), ϕ the sediment porosity, x the 229
- distance from the sediment-water interface (cm), D' the diffusion coefficients of dissolved species in the sediment 230
- (cm² yr⁻¹) adjusted for the considered setting (Supplementary Table S1) (Boudreau, 1997) and corrected for the 231
- tortuosity in the porous medium (Boudreau, 1996) (see Supplementary Information). $\sum R_S$ and $\sum R_{aa}$ are the net 232
- reaction rates of the solid and dissolved species from the chemical reactions they participate in (Table 3), and v and 233
- u the advective velocities (cm yr⁻¹) of the solid and the dissolved species, respectively. Porosity and advective 234
- 235 velocities were described by depth-dependent functions to account for sediment compaction (Meysman et al., 2005;
- 236 Reed et al., 2011a) (see Supplementary Information and Supplementary Fig. S1).
- 237 Reactions considered by the model and corresponding reaction equations are given in Tables 3 and 4, respectively,
- 238 and are divided into primary redox reactions and other biogeochemical reactions, including various mineral
- 239 formation and dissolution reactions (Reed et al., 2011a, 2011b; Rooze et al., 2016). Corresponding reaction
- 240 parameters were mostly taken from the literature or, if these were not available or no fit to the data could be obtained
- 241 with existing parameter ranges, constrained using the extensive geochemical dataset for site 4 (Table 5). A model
- 242 sensitivity analysis for key parameters is provided in the Supplementary Information (Supplementary Figs. S2 and
- 243 S3).
- 244 To account for differences in reactivity and crystallinity between different species, organic matter and Fe oxides are
- 245 divided into three different pools, representing highly reactive (α), less reactive (β) and non-reactive (i.e. inert) (γ)

phases. For the Fe oxides, only the α phase is used by organoclastic Fe reduction (Table 3), while the β phase is also 246 247 used by Fe-AOM. This assumption was made to test whether the pore water and sediment profiles observed in the 248 Black Sea can be reproduced with Fe-AOM as the main Fe reduction pathway at depth. In addition, it allows an 249 assessment of the potential impact of Fe-AOM on sedimentary CH₄ cycling. Note that, as a consequence of the 250 exclusion of organoclastic Fe reduction at depth, the model results should not be interpreted as proof for Fe-AOM 251 but rather imply that it is a possible mechanism. 252 The succession of oxidants during organic matter decomposition (Froelich et al., 1979) is described by means of Monod kinetics (Table 4), whereby those oxidants with the highest metabolic free energy yield are used 253 preferentially until they become limiting and the oxidant with the next highest energy yield is used (Berg et al., 2003; 254 Boudreau, 1996; Reed et al., 2011b; Rooze et al., 2016; Wang and Van Cappellen, 1996). Oxidants considered by the 255 model are (in descending order of energy yield) O₂, nitrate (NO₃⁻), Fe oxides and SO₄²⁻. Once these oxidants are 256 exhausted, organic matter remineralization occurs by methanogenesis. Corresponding limiting concentrations for the 257 258 oxidants are taken from (Reed et al., 2011a) (Table 5). In addition, an attenuation factor, Ψ, is used to slow down anaerobic organic matter degradation through SO_4^{2-} reduction and methanogenesis, thus allowing for better 259 260 preservation of organic matter under anoxic bottom water conditions (Moodley et al., 2005; Reed et al., 2011a, 261 2011b).

Cycling of S is simulated using five different chemical species, i.e. Fe monosulfides (FeS), pyrite (FeS₂), elemental S (S₀), dissolved sulfide and pore water SO_4^{2-} (Table 2), combined in a network of various biogeochemical reactions (Table 3). The CH₄ cycle includes CH₄ production from organic matter and from DIC (i.e. CO₂), as well as CH₄ oxidation coupled to the reduction of O₂, SO_4^{2-} and Fe(OH)₃ (Table 3). For AOM a bimolecular rate equation was used (Table 4), which is the most common way to parameterize AOM in reactive transport models (Regnier et al., 2011) and allows the use of largely unknown half-saturation constants, in particular for the putative Fe-AOM pathway, to be avoided. Although Mn-oxides have also been suggested to be a thermodynamically favorable electron acceptor for AOM (Beal et al., 2009), they were not included in the model because of the relatively low Mn concentrations (~ 15 µmol g⁻¹ for total sedimentary Mn and < 30 µM for dissolved Mn²⁺; Supplementary Fig. S2 and S3) when compared to Fe and the likely presence of most of the Mn in the form of Mn-carbonates.

The P forms included in the model are pore water HPO₄²⁻, authigenic Ca-P, organic P and detrital P, as well as Febound P, i.e. P associated with Fe oxides and P in vivianite (Table 2). The removal of dissolved Fe²⁺ through formation of the Fe minerals FeS, siderite (FeCO₃) and vivianite is also included in the model (Table 3). Mass balances for all chemical species included in the model are given in Supplementary Table S2.

The boundary conditions at the sediment surface were specified as time-dependent depositional fluxes for the particulate components and as fixed bottom water concentrations for the dissolved species, while a zero gradient boundary condition was set for all chemical species at the base of the model domain (Fig. 2 and Supplementary Table S3). To avoid potential interferences of the lower boundary conditions with the model results in the upper sediments (see Supplementary Fig. S6), the model depth was set to 3000 cm and divided into 500 grid cells. The thickness of the upper layer was set at 1 cm, and the thickness of the following grid layers increased exponentially to ~ 6 cm at 800 cm depth and to ~18 cm at 3000 cm depth. In this paper, only the upper 800 cm are shown. However, all profiles extending over the full depth range are provided in the Supplementary Information file (Supplementary

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Fig. S5 and Fig. S7). The model code was written in R using the marelac geochemical dataset package (Soetaert et al., 2010) and the ReacTran package (Soetaert and Meysman, 2012) to calculate the transport in porous media. The set of ordinary differential equations was subsequently solved numerically with the Isoda integrator algorithm (Hindmarsh, 1983; Petzoldt, 1983)

2.4.2 Transient scenario

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The model applied in this study simulates the sediment deposition during the last 25000 years. A constant mass accumulation rate of 0.06 g cm⁻² yr⁻¹ over the Holocene was assumed. In order to reduce the computing time for the freshwater period, a higher mass accumulation rate of 1 g cm⁻² yr⁻¹ was used between 25000 and 10000 years before present (B.P.) and all fluxes were corrected accordingly (i.e. multiplied with a factor of 16.67). Inflow of Mediterranean saltwater into the Black Sea basin was modelled assuming an initial salinity of 1 for the freshwater lake and a linear increase to a salinity of 22 between 8500 and 1500 years B.P. (Fig. 2). Such a salinization scenario results in a good fit to the chloride (Cl⁻) profile (Fig. 3) and compares well with a previous salinity reconstruction suggesting a linear increase in salinity of 1 to 22 between 9000 ± 500 years B.P. and 2000 ± 500 years B.P. (Soulet et al., 2010). Bottom water salinity was converted to Cl⁻ and SO₄²⁻ using the molecular weights and seawater density derived from the marelac geochemical dataset package (Soetaert et al., 2010) (Supplementary Table S1). A shift from oxic towards euxinic conditions around 7600 years B.P., with a peak in organic matter loading around 5300 years B.P. and constant elevated organic matter fluxes after 2700 years B.P. was assumed, following a recent study comprising data from seven sediment cores collected from the Black Sea (Eckert et al., 2013) (Fig. 2). In addition, the input of organic matter was assumed to increase again in the last century, reflecting anthropogenic eutrophication of waters on the adjacent continental shelf as previously reported (Capet et al., 2013; Kemp et al., 2009). With the development of anoxic and sulfidic bottom-water conditions, depositional fluxes of reactive Fe oxides were assumed to be zero (Fig. 2). In contrast, fluxes of Fe sulfides are high under euxinic conditions and dominated by FeS2.

3 Results

3.1 Pore water profiles

Pore water profiles of SO_4^{2-} show a linear decrease from ~ 17 mM at the sediment water interface to a depth of ~ 230 cm at both sites, below which CH_4 starts to accumulate in the pore water (Fig. 3). Bubble formation and degassing of CH_4 during gravity coring could not be avoided because of the high concentrations of CH_4 in the limnic deposits above the saturation of ca 1.3 mM CH_4 at atmospheric pressure (calculated for a salinity of 22 and a temperature of 25 °C using the algorithm from Mogollón et al. (2013)). Observations of increased outgassing with depth during coring suggest that the low CH_4 concentrations in the deeper sediments at both sites are due to enhanced outgassing with increasing levels of CH_4 . Pore water profiles of NH_4^+ at both sites are similar and concentrations increase to ~ 3 mM at depth, suggesting that actual CH_4 concentrations at both sites could be comparable. Most of the CH_4 values thus only indicate the presence or absence of CH_4 and are not a quantitative measure (indicated as open diamonds in Fig. 4). Note that the upper ~ 300 cm of sediment at site 5 are likely less affected by CH_4 outgassing. Modeled pore

- water concentrations of CH₄ on the other hand, show a steep increase below the SMTZ, comparable to the gradient
- observed at site 5, and build up to concentrations of ~ 15 mM at depth (Supplementary Fig. S5).
- 320 The SMTZ is located around 230 cm depth in the sediment and is characterized by the removal of both pore water
- 321 SO₄²⁻ and dissolved CH₄. In this zone, SO₄-AOM drives the production of dissolved sulfide, DIC and alkalinity
- 322 (Supplementary Fig. S5) and diffusion of these pore water constituents away from the SMTZ (Fig. 3). Below the
- sulfide diffusion front, Fe²⁺ accumulates in the pore water. Dissolved HPO₄²⁻ reaches a maximum around the depth
- 324 where sulfide levels drop below the detection limit of 1 μmol L⁻¹, followed by a steep decrease with depth.
- Concentrations of pore water Mn²⁺ are more than an order of magnitude lower than those of dissolved Fe²⁺, and
- decrease from the sediment surface until ~ 200 cm depth, below which they slightly increase again (Supplementary
- 327 Fig. S5)

- 328 The smooth pore water profiles of δ¹³C-CH₄ and δD-CH₄ suggest that the isotopic composition of pore water CH₄
- 329 (available for site 5 only) is less affected by the CH₄ loss and reveals a biological origin in the limnic deposits, with
- hydrogenotrophic carbonate reduction, i.e. microbial reduction of CO₂ to CH₄ as the main methanogenic pathway for
- the range of CH₄ isotope ratios observed in these sediments (Fig. 4) (Whiticar, 1999). Upward diffusing CH₄ shows a
- gradual depletion in δ^{13} C-CH₄ from \sim -74 % at depth to \sim -96 % around the SMTZ, followed by subsequent
- progressive ¹³C enrichment towards the sediment surface. δD-CH₄ shows a small enrichment from -226 ‰ at depth
- to \sim -208 % at the SMTZ and a strong shift towards high δ D-CH₄ values of up to \sim 113 %.

3.2 Solid phase profiles

- A pronounced excursion in sedimentary C_{org} at site 4 in combination with a shift from gray clay deposits to micro-
- laminated black sediments indicates that the lake-marine transition, i.e. the transition between the marine sapropel
- Unit II and the deep limnic sediments of Unit III (Arthur and Dean, 1998; Degens and Ross, 1974), is located around
- a sediment depth of ~ 90 cm at site 4 (Fig. 5). At site 5, Unit I and Unit II were lost due to a turbidite, explaining the
- low concentrations of C_{org} in the upper sediments.
- 341 Concentrations of solid S increase with decreasing depth from 20 μmol g⁻¹ below 300 cm (sulfidization front) to ~
- 342 400 μmol g⁻¹ in the upper 100 cm at both sites and are dominated by FeS₂ (Fig. 5). Iron oxides show a decrease from
- $\sim 100 \ \mu mol \ g^{-1}$ at depth to $\sim 50 \ \mu mol \ g^{-1}$ in the sediments between $100-300 \ cm$ and a further decrease to $\sim 10 \ \mu mol$
- g^{-1} closer to the sediment surface. Amorphous Fe oxides (Fe_{ox1}) and more crystalline oxides (Fe_{ox2}) both account for
- half the total amount of Fe oxides, with a small contribution of recalcitrant oxides (Fe_{mag}) (Supplementary Fig. S4).
- 346 The results from the two different Fe extractions applied in this study (Table 1) generally compare well
- 347 (Supplementary Fig. S4). Note, however, that the Fe oxides in Fig. 5 represent the results from the extraction after
- Poulton and Canfield (2005). Results from the Fe extractions modified from Claff et al. (2010) are provided in the
- 349 Supplementary Information only. Sedimentary Mn content is relatively low at all three sites, ranging from ~ 5-10
- μ mol g⁻¹ in the marine sediments to $\sim 15 \mu$ mol g⁻¹ in the deep limnic deposits of Unit III (Supplementary Fig. S4).
- Sediments below the sulfidization front are characterized by high Fe carbonate contents of $\sim 100 \ \mu mol \ g^{-1}$. The sharp
- depletion in Fe carbonate around the sulfidization front could only be reproduced in the model by assuming Fe

- 353 carbonate dissolution by dissolved sulfide (Table 3). These results suggest a conversion of reactive Fe from
- carbonate toward sulfide phases in the presence of abundant dissolved sulfide.
- Units I and II show high concentrations of organic P, which accounts for ~ 30 % of total P in these sediments (Fig.
- 356 5). Low organic P and high concentrations of detrital P in the upper sediments at site 5 are due to the turbidite. The
- 357 limnic deposits of Unit III are generally depleted in organic P (< 6 % of total P) and enriched in detrital P.
- Authigenic Ca-P shows little variation in the sediments of Unit III, accounting for ~ 20 to 30 % of total P at the two
- sites. The contribution of Fe-associated P, on the other hand, is reduced in the limnic deposits of Unit III exposed to
- the downward diffusing sulfide (~ 20 %) when compared to the sediments below the sulfidization front (~ 30 %).
- Concentrations of exchangeable P are $\leq 2 \mu \text{mol g}^{-1}$ for sediments above the SMTZ and $\leq 1 \mu \text{mol g}^{-1}$ for sediments at
- depth (data not shown).
- Modeled SO_4^{2-} reduction rates show two distinct peaks of ~ 200 pmol SO_4^{2-} cm⁻³ d⁻¹ in the sediments of Unit II and
- in the sediments around the SMTZ (Fig. 6). Rates of CH₄ production are highest (~ 30 pmol CH₄ cm⁻³ d⁻¹) in the
- organic-rich marine deposits of Unit II and in the limnic deposits below the SMTZ. The sediments around the SMTZ
- are further characterized by high rates of SO₄-AOM (~ 200 pmol cm⁻³ d⁻¹), whereas sediments directly below the
- 367 sulfidization front show enhanced rates of S_0 disproportionation (~ 15 pmol cm⁻³ d⁻¹). Organoclastic SO_4^{2-} reduction
- 368 provides the main source for pore water sulfide in the organic-rich marine deposits, while SO₄-AOM and S₀
- disproportionation are the dominant sources of dissolved sulfide in sediments around the SMTZ and directly below
- the sulfidization front, respectively. Rates of Fe-AOM are generally low (< 0.04 pmol CH₄ cm⁻³ d⁻¹) and restricted to
- 371 the limnic deposits only.

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3.3 Temporal evolution

- 373 The temporal evolution in pore water and solid phase constituents illustrates the impact of the lake-marine transition
- on the sediment geochemistry (Fig. 7). Concentrations of pore water Cl⁻ and SO₄²⁻ increase with the intrusion of
- marine Mediterranean Sea waters, accompanied by a decrease in dissolved CH₄ and accumulation of pore water
- sulfide in the shallower sediments. Dissolved Fe²⁺ becomes restricted to non-sulfidic pore waters at depth, while
- 377 HPO₄²⁻ and solid S start to accumulate in the presence of dissolved sulfide. Iron oxides decrease in the surface
- 378 sediments as well as in the sediments at depth. Vivianite, on the other hand, becomes increasingly enriched in
- sediments below the downward diffusing sulfide front.

4. Discussion

4.1 Coupled S, CH₄ and Fe dynamics

382 4.1.1 Organoclastic SO₄²⁻ reduction

- Model-derived areal rates of total SO_4^{2-} reduction of ~ 0.24 mmol SO_4^{2-} m⁻² d⁻¹ (Table 6), i.e. the total amount of
- SO_4^{2-} reduced per square meter of sea floor, compare well with calculated diffusive fluxes of SO_4^{2-} into the sediment
- at sites 4 and 5 (~ 0.21 and 0.20 mmol SO_4^{2-} m⁻² d⁻¹, respectively) and are in good agreement with previous SO_4^{2-} flux
- estimates of 0.17 to 0.28 mmol SO₄²⁻ m⁻² d⁻¹ for sediments of the western Black Sea (Jørgensen et al., 2001). In the

model, organoclastic SO_4^{2-} reduction accounts for > 65 % of total organic matter degradation in the upper 800 cm of 387 sediment, supporting previous conclusions that SO₄² reduction represents the dominant mineralization process of 388 organic matter in sediments below the chemocline (Jørgensen et al., 2001; Thamdrup et al., 2000). The remaining < 389 25 % of organic matter remineralization is due to methanogenesis. The relative contribution of SRR to organic 390 matter remineralization, however, likely is significantly higher when taking into account the high SRR in the 391 uppermost sediment layers (Jørgensen et al., 2001), which are not captured by our model. 392 The depth-dependent rate profile of SO_4^{2-} reduction shows two distinct peaks of ~ 70 and 230 pmol SO_4^{2-} cm⁻³ d⁻¹ 393 associated with organoclastic SO₄²⁻ reduction in the organic matter rich marine deposits of Unit I and Unit II. These 394 395 rates are at the low end of reported values from Black Sea sediments (0.1 - 20 nmol cm⁻³ d⁻¹) (Holmkvist et al., 2011b; Jørgensen et al., 2001, 2004; Knab et al., 2009; Leloup et al., 2007). Our model further demonstrates that the 396 two SRR peaks in the sediments of Unit I and Unit II are not reflected in the pore water profile of SO_4^{2-} . This finding 397 is in line with earlier work showing that the SO₄² gradient in Black Sea sediments is primarily affected by SO₄-398 AOM in the SMTZ (Jørgensen et al., 2001). The shorter diffusion distance (the diffusion time to ~ 200 cm is about 5 399 times longer than to ~ 90 cm, i.e. ~ 300 years vs. ~ 60 years) and higher porosity in Unit I and II (Supplementary Fig. 400 S1) both dampen the effect of SO_4^{2-} reduction in the marine deposits on the SO_4^{2-} profile (see also Jørgensen et al., 401 (2001)). Thus, our results support previous conclusions that SRR estimates based on pore water profiles of SO₄²⁻ (i.e. 402 net SO_4^{2-} consumption) alone may underestimate the actual SO_4^{2-} turnover (i.e. gross SO_4^{2-} reduction) in marine 403 sediments (Jørgensen, 1978; Jørgensen et al., 2001). 404

4.1.2 SO₄-AOM

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Pore water profiles of SO₄², CH₄, sulfide and DIC reveal a distinct SMTZ around 230 cm depth at both sites, where 406 SO₄-AOM with upward diffusing CH₄ results in the concomitant removal of pore water SO₄²- and CH₄ and in the 407 accumulation of dissolved sulfide and DIC in the pore waters of these sediments (Fig. 3). The depth of the SMTZ 408 409 and the steep increase in CH₄ to > 3 mM below the SMTZ found in this study are consistent with earlier observations 410 in sediments of the western Black Sea (Henkel et al., 2012; Holmkvist et al., 2011b; Jørgensen et al., 2001, 2004; Knab et al., 2009; Leloup et al., 2007). The location of the SMTZ, however, has progressed downwards in the last ca. 411 9000 years, following the inflow of SO₄²-rich salt water into the Black Sea basin (Fig. 7) (see also Henkel et al., 412 413 Calculated diffusive fluxes of SO₄²⁻ and CH₄ to the SMTZ (~ 0.2 mmol SO₄²⁻ m⁻² d⁻¹ and 0.08 mmol CH₄ m⁻² d⁻¹) and 414 modelled areal rates of SO₄-AOM (~ 0.16 mmol m⁻² d⁻¹) suggest that AOM accounts for ~ 40 to 70 % of the total 415 SO_4^{2-} reduction in these sediments, with the remaining ~ 30 to 60 % attributed to organoclastic SO_4^{2-} reduction. Such 416 417 a high contribution of AOM exceeds the range of previous estimates that included experimentally measured SRR close to the sediment surface (~ 7 to 18 %) (Jørgensen et al., 2001, 2004). Around the SMTZ, SO₄-AOM is 418 responsible for ~ 97 % of the total SO_4^{2-} reduction (Fig. 6 and Table 6), thus enhancing the downward diffusive flux 419 of sulfide into the deep limnic deposits of Unit III. Our model suggests that without this additional source of sulfide 420 through SO₄-AOM, the sulfidization front would currently be located around 150 cm depth in the sediment (Fig. 8). 421

- The consumption of upward diffusing CH₄ by SO₄²-driven AOM leads to a progressive enrichment of ¹³C and D in
- 423 the residual CH₄ above the SMTZ (Fig. 4) due to the preferential oxidation of isotopically light CH₄ during SO₄-
- 424 AOM (Alperin et al., 1988; Martens et al., 1999; Whiticar, 1999). Interestingly, pore water CH₄ above the SMTZ
- shows unusually high δD-CH₄ values that fall outside of the common range observed for pore water δD-CH₄ (e.g.
- Whiticar et al., 1999). Future studies, however, are needed to resolve the cause of the strong D-enrichment of
- dissolved CH₄ above the SMTZ in Black Sea sediments.
- 428 Modeled concentrations of CH₄ indicate that the measurements above the sulfidization front at site 5 are likely less
- affected by outgassing during core recovery (Fig. 4) and can thus be used to derive kinetic isotope fractionation
- factors for carbon (ε_C) and hydrogen (ε_H) associated with SO₄-AOM at the SMTZ using the Rayleigh distillation
- function (Crowe et al., 2011; Egger et al., 2015b; Rayleigh, 1896; Whiticar, 1999). Corresponding estimates for ε_C of
- $\sim 8 \%$ (R² = 0.972) and $\epsilon_{\rm H}$ of $\sim 58 \%$ (R² = 0.982) are at the lower end of previously documented values in marine
- and brackish-marine environments (8-38 % for ε_C and 100-324 % for ε_H) (Alperin et al., 1988; Egger et al., 2015b;
- Holler et al., 2009; Martens et al., 1999; Reeburgh, 2007). Note, however, that the Rayleigh distillation function only
- 435 applies to closed systems (Rayleigh, 1896). These estimates should therefore be interpreted as an approximation, as
- more accurate estimates would require isotope modeling (e.g. Alperin et al., 1988).
- At the base of the SMTZ, upward diffusing CH₄ reveals an initial depletion in δ^{13} C-CH₄ (Fig. 4). Such a shift to 13 C-
- depleted CH₄ together with a decrease in its concentration could indicate an enzyme-mediated equilibrium C isotope
- exchange during SO₄-AOM at low SO₄²⁻ concentrations (< 0.5 mM) (Holler et al., 2012; Yoshinaga et al., 2014). The
- effect of such mechanisms on deuterated CH₄ is likely limited.

441 4.1.3 Cryptic S cycling

- Earlier studies showed evidence for ongoing SO_4^{2-} reduction (< 1 nmol cm⁻³ d⁻¹) within the SO_4^{2-} -depleted (< 0.5
- 443 mM) limnic deposits below the SMTZ in sediments of the Black Sea (Holmkvist et al., 2011b; Knab et al., 2009;
- Leloup et al., 2007), Baltic Sea (Holmkvist et al., 2011a, 2014; Leloup et al., 2009) and Alaskan Beaufort Sea
- 445 (Treude et al., 2014) likely driven by SO_4^{2-} production from re-oxidation of dissolved sulfide with oxidized Fe
- minerals. In this mechanism, Fe oxides enhance the recycling of sulfide to SO₄²⁻ in a cryptic S cycle (Holmkvist et
- al., 2011a; Treude et al., 2014) thereby fueling SO₄²-driven AOM in Fe oxide-rich sediments. In this cryptic S cycle,
- dissolved sulfide is oxidized to zero-valent sulfur (S₀), a key intermediate in AOM, which is subsequently
- disproportionated to SO₄² and sulfide by associated Deltaproteobacteria (Holmkvist et al., 2011a; Milucka et al.,
- 2012; Sivan et al., 2014; Treude et al., 2014). The additional SO₄²⁻, produced during S₀ disproportionation, may then
- be re-used by the methanotrophic archaea as an electron acceptor for SO₄-AOM (Milucka et al., 2012).
- Our model results suggest slow rates of ongoing SO_4^{2-} reduction of < 0.2 nmol cm⁻³ d⁻¹ (Fig. 6) within the limnic
- deposits below the SMTZ exposed to dissolved sulfide (Table 6), in line with estimated SRR based on ³⁵SO₄²⁻
- incubation experiments with Black Sea sediments from below the SMTZ of ~ 0.1-0.5 nmol cm⁻³ d⁻¹ (Knab et al.,
- 455 2009; Leloup et al., 2007). Below the sulfidization front, SRR show a distinct peak of ~ 5 pmol cm⁻³ d⁻¹. Active SO_4^{2-}
- reduction in these SO_4^{2-} -depleted sediments requires deep SO_4^{2-} formation to maintain low net rates of SO_4^{2-}
- reduction. In the model, S_0 disproportionation is the only potential source of pore water SO_4^{2-} at depth (Table 3).

Formation of S₀, in turn, occurs exclusively by oxidation of dissolved sulfide during the reductive dissolution of Fe 458 oxides, explaining the distinct S_0 disproportionation peak of ~ 15 pmol cm⁻³ d⁻¹ around the sulfidization front (Fig. 6). 459 Thus, based on the model assumptions, we conclude that Fe oxides increase the transformation of sulfide to SO₄² via 460 formation and subsequent disproportionation of S₀ in these sediments, as suggested previously (Holmkvist et al., 461 2011b; Knab et al., 2009; Leloup et al., 2007). Such recycling of SO_4^{2-} stimulates slow rates of SO_4 -AOM in the 462 sediments below the SMTZ, explaining the low background rates of SO₄² reduction in the SO₄² depleted limnic 463 deposits. These results support recent findings of indirect Fe stimulated SO₄²⁻ driven AOM in laboratory experiments 464 (Sivan et al., 2014), and highlight that Fe oxides could play a significant role as stimulators of AOM and S recycling 465 in natural environments. 466

4.2 Fe reduction below the sulfidization front

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Below the sulfidization front, Fe^{2^+} starts to accumulate in the pore water (Fig. 3). Although previous studies have also reported an increase of dissolved Fe^{2^+} around the depth where sulfide levels drop below the detection limit (Holmkvist et al., 2011b; Jørgensen et al., 2004; Knab et al., 2009), the source of this pore water Fe^{2^+} has remained unknown. One possible explanation could be that the elevated Fe^{2^+} concentrations at depth represent remnant Fe^{2^+} accumulated during the Black Sea "Lake" phase (Knab et al., 2009). In our model, Fe^{2^+} shows a broad peak of \sim 300 μ M until \sim 300 cm depth in the sediment during the initial Lake phase, assuming organoclastic Fe reduction as the only Fe reduction pathway (data not shown). The removal of Fe^{2^+} through authigenic formation of reduced Fe(II) minerals, however, prevents the accumulation of substantial amounts of Fe^{2^+} in the pore water below \sim 300 cm sediment depth during the Lake phase (Fig. 8). We therefore conclude that the high concentrations of dissolved Fe^{2^+} below the sulfidization front are most likely indicative of active Fe reduction in these sediments.

4.2.1 Fe reduction through cryptic S cycling

In theory, a cryptic S cycle, as described in section 4.1.3, could result in net accumulation of dissolved Fe²⁺ if the 479 sulfide consumption from reaction with ferric Fe outweighs the production of sulfide from SO_4^{2-} reduction. Modeled 480 Fe^{2+} indeed shows a peak of < 100 μ M directly below the sulfidization front, assuming no active Fe reduction in the 481 limnic deposits (Fig. 8). Model simulations further indicate that, based on the reaction network used in this study 482 (Table 3), cryptic S cycling could result in a build up of pore water Fe²⁺ of ~ 300 µM at depth in the sediment 483 provided there was no precipitation of reduced Fe(II) minerals (Supplementary Fig S2). However, concentrations of 484 dissolved Fe²⁺ are too low compared to the measurements and confined to sediments between 300 – 400 cm depths 485 486 only. The diagenetic model developed in this study therefore suggests that cryptic S cycling is unlikely to explain the high concentrations (~ 800 μM) of dissolved Fe²⁺ observed in the deep limnic deposits. 487

4.2.2 Organoclastic Fe reduction

In the model, the reduction of Fe oxides coupled to organic matter degradation only occurs with the easily reducible α phase in order to allow for the burial of the more crystalline β phase at depth (Table 3). Since the α phase is efficiently reduced in the upper few centimeters during organoclastic Fe reduction, no easily reducible Fe oxides are

being buried into the deep sediments in the diagenetic model. Organoclastic Fe reduction therefore does not occur within the modeled deep limnic deposits that exclusively contain more crystalline (β) and refractory (γ) Fe oxides (Fig. 5). Instead, we assume that CH₄ represents a plausible electron donor for the reduction of more crystalline Fe oxides in the organic-poor deep sediments with relatively refractory old organic matter (< 0.8 wt %). The exclusion of organoclastic Fe reduction at depth in the model provides an estimate of an upper constraint on the potential importance of Fe-AOM in Black Sea sediments. As a result of this assumption, however, the model results cannot be used to conclude whether Fe-AOM is more likely than organoclastic Fe reduction.

An increasing body of geochemical evidence and laboratory incubation experiments shows that Fe-AOM might be occurring in a variety of different aquatic environments (Amos et al., 2012; Beal et al., 2009; Crowe et al., 2011; Egger et al., 2015b; Riedinger et al., 2014; Scheller et al., 2016; Segarra et al., 2013; Sivan et al., 2011; Wankel et al., 2012). In addition, several studies have shown that Fe-reducing microorganisms are able to outcompete methanogens for common substrates (e.g. acetate and H₂), thus reducing the concentrations of these common primary electron donors to levels that are too low for methanogens to grow (Achtnich et al., 1995; Lovley and Phillips, 1987; Lovley et al., 1989). These results, together with the observed capability of methanogens to switch

from CH₄ production to Fe reduction (Bodegom et al., 2004; Bond and Lovley, 2002; Liu et al., 2011; Reiche et al., 2008; Sivan et al., 2016; Vargas et al., 1998) led to the common conclusion that Fe oxides exert a suppressive effect

on methanogenesis. Ongoing CH₄ production in the Fe oxide-rich limnic deposits, as deduced from the isotopic

composition of pore water CH₄ (Fig. 4) could then indicate limited organoclastic Fe reduction in these sediments. However, there is increasing evidence that (semi)conductive crystalline Fe oxides (e.g. hematite and magnetite)

However, there is increasing evidence that (semi)conductive crystalline Fe oxides (e.g. hematite and magnetite) can, in fact, stimulate concurrent methanogenesis and organoclastic Fe reduction through direct interspecies electron transfer (DIET), by serving as electron conduits among syntrophic CH₄-producing organisms at rates that are substantially higher than those for interspecies electron transfer by H₂ (Cruz Viggi et al., 2014; Kato et al., 2012; Li et al., 2014; Zhou et al., 2014; Zhuang et al., 2015). The inhibitory effect of Fe reduction on methanogenesis thus appears to be lower for crystalline Fe oxides such as hematite and magnetite, which are less bioavailable to Fereducing organisms than poorly crystalline (amorphous) Fe oxides (e.g. ferrihydrite and lepidocrocite) (Lovley, 1991; Qu et al., 2004; Zhuang et al., 2015). These findings indicate that the crystallinity and conductivity of Fe oxides may play a key role in determining whether methanogenesis is stimulated or suppressed in Fe oxide-rich environments. In addition, the presence of methanogens that are able to rapidly switch between methanogenesis and reduction of Fe oxides could also result in a reactivation of less reactive Fe oxides that were not reduced during initial organoclastic Fe reduction in the deep methanogenic zone as suggested by Sivan et al. (2016). Thus, the deep limnic sediments may be characterized by a complex interplay of concurrent methanogenesis, Fe oxide reduction and methanotrophy, i.e. AOM.

4.2.3 Fe-AOM

- Our model results indicate that Fe-AOM could also be a possible mechanism explaining the buildup of pore water
- 526 Fe²⁺ below the sulfidization front. Previous studies have shown that in systems where production and oxidation of
- 527 CH₄ take place concurrently, methanogenesis might conceal the isotopic signature of AOM (Egger et al., 2015b;

- 528 Seifert et al., 2006; Whiticar, 1999). Thus, unlike SO₄-AOM, Fe-dependent AOM likely only has little effect on the
- 529 isotopic composition of pore water CH₄ due to the removal of small amounts of CH₄ in sediments with ongoing
- methanogenesis. This might explain why pore water CH₄ does not show enrichment in both heavy isotopes below the
- sulfidization front as would be expected if Fe-AOM would occur, but rather indicates antipathetic changes, i.e.
- depletion in ¹³C-CH₄ and enrichment in D-CH₄, usually attributed to CH₄ production from carbonate reduction
- 533 (Chanton et al., 2005; Whiticar, 1999).
- Model derived rates for Fe-AOM of ~ 0.04 pmol cm⁻³ d⁻¹ (Fig. 6) are significantly lower than potential Fe-AOM
- rates of ~ 4 nmol cm⁻³ d⁻¹ estimated from laboratory incubation studies (Egger et al., 2015b; Segarra et al., 2013;
- Sivan et al., 2011) with brackish and limnic sediment samples. This large deviation is likely due to an overestimation
- 537 of Fe-AOM rates derived from stimulated microbial communities under laboratory conditions using freshly
- synthesized and thus easily bioavailable Fe oxides when compared to in-situ conditions.
- In the upper 800 cm of sediment, Fe-AOM accounts for < 1 % of total CH₄ oxidation, with the remaining > 99 %
- attributed to SO₄-AOM (Table 6; see also Supplementary Fig. S2). However, while high rates of SO₄-AOM are
- mainly restricted to the SMTZ, Fe-AOM might occur over a deep methanogenic zone, reaching far down into the
- sediment. To accurately assess the contribution of Fe-AOM to the total CH₄ consumption in Black Sea sediments,
- additional knowledge about the vertical expansion of the Fe oxide-rich limnic sediments deposited during the Blake
- Sea "Lake" phase would be required.

4.3 Impact of S-Fe-CH₄ dynamics on sedimentary P diagenesis

- Degradation of organic matter and the subsequent release of HPO₄²⁻ to the pore water during early diagenesis
- 547 typically results in a sink-switching from organic P to authigenic P-bearing phases such as Ca phosphates (Filippelli,
- 548 1997; Ruttenberg and Berner, 1993; Slomp et al., 1996b), Mn-Ca carbonates (Jilbert and Slomp, 2013; Mort et al.,
- 2010; Suess, 1979) or reduced Fe phosphates (Burns, 1997; Jilbert and Slomp, 2013; Martens et al., 1978; März et
- al., 2008). Reductive dissolution of Fe oxides by dissolved sulfide and the following liberation of HPO₄²⁻ may also
- contribute to the buildup of pore water HPO₄² (Burns, 1997; Egger et al., 2015a; März et al., 2008; Schulz et al.,
- 552 1994). Thus, the downward sulfidization ultimately results in the accumulation of dissolved HPO₄²⁻ in the pore water
- as the sulfidization front moves downward into the limnic deposits (Fig. 7).
- The pore water profile of HPO₄²⁻ (Fig. 3) indicates the presence of a sink for HPO₄²⁻ below the sulfidization front
- and, to a lesser extent, in the sulfidic sediments around the SMTZ, likely unrelated to Ca-P authigenesis (Fig. 5).
- Such a sink for HPO₄²⁻ below sulfidic sediments has been observed previously (Burns, 1997; Egger et al., 2015a;
- März et al., 2008; Schulz et al., 1994; Slomp et al., 2013) and shown to be most likely the result of vivianite
- formation (Egger et al., 2015a; Hsu et al., 2014; März et al., 2008). Abundant dissolved Fe²⁺ and a peak in Fe-
- associated P below the sulfidization front observed in this study (Fig. 3 and Fig. 5) suggest that vivianite authigenesis
- might also be occurring in the limnic deposits below the sulfidzation front in Black Sea sediments.
- Assuming that vivianite formation represents the only sink for pore water HPO₄²⁻ results in a good fit between the
- modeled and measured pore water profile of HPO₄²⁻ below the sulfidization front (Fig. 3). Modeled vivianite
- formation accounts for up to 70 % of total Fe-associated P directly below the sulfidization front. However, the model

underestimates the sharp peak in Fe-associated P directly below the sulfidization front, suggesting that modeled vivianite formation likely underestimates the actual contribution of vivianite in these sediments. In the limnic deposits not yet impacted by the downward sulfidization, modeled vivianite accounts for $\sim 20-30$ % of total Fe-associated P. From this, we estimate that vivianite may be responsible for > 20 % of total P burial directly below the sulfidization front and for ~ 10 % of total P burial in the deep limnic deposits at depth.

Running the model without Fe-AOM and thus without Fe reduction at depth results in modeled pore water HPO_4^{2-} concentrations of up to $\sim 350~\mu M$ at depth in the sediment (Fig. 8). This suggests that Fe-AOM can promote conditions that allow sequestration of a significant proportion of P as vivianite in the limnic deposits below the sulfidization front. Consistent with earlier findings, Fe-AOM likely only accounts for a small fraction of total CH_4 oxidation, but may substantially impact the biogeochemical cycling of sedimentary P (Egger et al., 2015a, 2015b; Rooze et al., 2016).

The deviation between the modeled and measured profiles of HPO₄²⁻ and Fe-associated P in the upper 300 cm of sediment (Fig. 3 and Fig. 5) could indicate apatite authigenesis (Dijkstra et al., 2014) or the formation of vivianite in microenvironments as previously suggested for sulfidic sediments (Dijkstra et al., 2014; Jilbert and Slomp, 2013). For example, *Deltaproteobacteria*, known to be involved in SO₄-AOM, have been shown to accumulate Fe- and Prich inclusions in their cells (Milucka et al., 2012). They may therefore provide a potential explanation for the occurrence of Fe-associated P in sulfidic sediments (Dijkstra et al., 2014; Jilbert and Slomp, 2013). However, such microenvironments are not captured in our model.

In the diagenetic model, vivianite undergoes dissolution if sulfide is present in the pore waters (Table 3). Sulfide-induced vivianite dissolution significantly improved the model fit to the measured HPO₄² and sulfide data. With the downward migration of dissolved sulfide, modeled vivianite becomes increasingly enriched below the sulfidization front (Fig. 7). Thus, similar to the sulfidization front, a downward diffusive vivianite front may exist in sedimentary systems experiencing downward sulfidzation.

In summary, the enhanced downward sulfidization driven by SO₄-AOM leads to dissolution of Fe oxide-bound P in the lake deposits. Below the sulfidization front, downward diffusing HPO₄²⁻ is bound again in authigenic vivianite due to high concentrations of dissolved Fe²⁺ at depth in the sediment generated by ongoing Fe oxide reduction. As a result, trends in total P with depth are significantly altered, showing an accumulation in total P below the sulfidization front unrelated to changes in organic matter deposition and enhanced sedimentary P burial during deposition.

5. Conclusions

In the Black Sea, the shift from a freshwater lake to a marine system and subsequent downward diffusion of marine SO_4^{2-} into the CH₄-bearing lake sediments results in a multitude of diagenetic reactions around the SMTZ (Fig. 9). The diagenetic model developed in this study shows that SO_4 -AOM within the SMTZ significantly enhances the downward diffusive flux of sulfide into the deep limnic deposits, forming a distinct diagenetic sulfidization front around 300 cm depth in the sediment. Our results indicate that without this additional source of dissolved sulfide in the SMTZ, the current sulfidization front would be located around a depth of 150 cm. During the downward

- sulfidization, Fe oxides, Fe carbonates and vivianite are converted to Fe sulfide phases, leading to an enrichment in solid phase S contents and the release of HPO_4^{2-} to the pore water. Our results further support the hypothesis that part of the downward migrating sulfide is re-oxidized to SO_4^{2-} upon reaction with ferric Fe minerals, fueling a cryptic S cycle and thus stimulating slow rates (~ 1-100 pmol cm⁻³ d⁻¹) of SO_4 -AOM in the SO_4^{2-} -depleted limnic deposits below the SMTZ (Holmkvist et al., 2011a, 2011b; Knab et al., 2009; Leloup et al., 2007).
- We propose that besides organoclastic Fe oxide reduction and reactivation of less reactive Fe oxides by methanogens, AOM coupled to the reduction of Fe oxides may also be a possible mechanism explaining the high concentrations of Fe²⁺ in the pore water below the sulfidization front. The buildup of dissolved Fe²⁺ at depth creates conditions that allow sequestration of the downward diffusing HPO₄²⁻ as authigenic vivianite, resulting in an accumulation of total P in these sediments.
- The diagenetic processes described here reveal that AOM may strongly overprint burial records of Fe, S and P in depositional marine systems subject to changes in organic matter loading or water column salinity such as coastal environments (Egger et al., 2015a; Rooze et al., 2016), deep-sea fan sediments (März et al., 2008; Schulz et al., 1994) and many high-latitude seas (Holmkvist et al., 2014; Treude et al., 2014). Interpreting these diagenetic patterns as primary sedimentary signals may lead to incorrect reconstructions of environmental conditions during sediment deposition.

Acknowledgements

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We thank the captain, crew and shipboard party of the PHOXY cruise aboard R/V Pelagia to the Black Sea in June 617 618 2013 and G. J. Reichart. We also thank NIOZ Marine Research Facilities for their support and K. Bakker and S. 619 Ossebaar for their contribution to the pore water analysis. D. van de Meent, T. Claessen, T. Zalm, A. van Dijk, E. 620 Dekker and G. Megens are acknowledged for technical and analytical assistance in Utrecht and M. Hagens for her 621 support with the modelling. We further thank C. van der Veen for the methane isotope analysis. This research was 622 funded by ERC Starting Grant 278364, NWO Open Competition Grant 822.01013 and NWO-Vici Grant 865.13.005 623 (to C. P. Slomp). In addition, P. Kraal would like to acknowledge NWO Veni grant 863.14.014. This work was 624 carried out under the program of the Netherlands Earth System Science Centre (NESSC), financially supported by 625 the Ministry of Education, Culture and Science (OCW). Orit Sivan, Wei-Li Hong and an anonymous reviewer are 626 gratefully acknowledged for their insightful comments and suggestions that improved the quality of the manuscript.

References

627

- 629 Achtnich, C., Bak, F. and Conrad, R.: Competition for electron donors among nitrate reducers, ferric iron reducers,
- sulfate reducers, and methanogens in anoxic paddy soil, Biol. Fertil. Soils, 19, 65–72, doi:10.1007/BF00336349,
- 631 1995.
- 632 Alperin, M. J., Reeburgh, W. S. and Whiticar, M. J.: Carbon and hydrogen isotope fractionation resulting from
- anaerobic methane oxidation, Glob. Biogeochem. Cycl., 2(3), 279–288, doi:10.1029/GB002i003p00279, 1988.

- Amos, R. T., Bekins, B. A., Cozzarelli, I. M., Voytek, M. A., Kirshtein, J. D., Jones, E. J. P. and Blowes, D. W.:
- Evidence for iron-mediated anaerobic methane oxidation in a crude oil-contaminated aquifer, Geobiology, 10(6),
- 636 506–17, doi:10.1111/j.1472-4669.2012.00341.x, 2012.
- Arthur, M. A. and Dean, W. E.: Organic-matter production and evolution of anoxia in the Holocene Black Sea,
- 638 Paleoceanography, 13(4), 395–411, doi:10.1029/98PA01161, 1998.
- 639 Beal, E. J., House, C. H. and Orphan, V. J.: Manganese- and iron-dependent marine methane oxidation, Science (80-.
- 640)., 325, 184–187, doi:10.1126/science.1169984, 2009.
- 641 Berg, P., Rysgaard, S. and Thamdrup, B.: Dynamic modeling of early diagenesis and nutrient cycling. A case study
- in an Arctic marine sediment, Am. J. Sci., 303(10), 905–955, doi:10.2475/ajs.303.10.905, 2003.
- Berner, R. A.: Iron sulfides in Pleistocene deep Black Sea sediments and their paleo-oceanographic significance, in
- The Black Sea Geology, Chemistry and Biology, edited by E. T. Degens and D. A. Ross, pp. 524-531, Plenum
- 645 Press., 1974.
- Berner, R. A.: Early Diagenesis: A Theoretical Approach, Princeton University Press., 1980.
- 647 Bodegom, P. M., Scholten, J. C. M. and Stams, A. J. M.: Direct inhibition of methanogenesis by ferric iron, FEMS
- 648 Microbiol. Ecol., 49(2), 261–268, doi:10.1016/j.femsec.2004.03.017, 2004.
- Bond, D. R. and Lovley, D. R.: Reduction of Fe(III) oxide by methanogens in the presence and absence of
- 650 extracellular quinones, Environ. Microbiol., 4(2), 115–24, doi:10.1046/j.1462-2920.2002.00279.x, 2002.
- Boudreau, B. P.: The diffusive tortuosity of fine-grained unlithified sediments, Geochim. Cosmochim. Acta, 60(16),
- 652 3139–3142, doi:10.1016/0016-7037(96)00158-5, 1996.
- 653 Boudreau, B. P.: Diagenetic models and their implementation. Modelling transport and reactions in aquatic
- sediments, Springer., 1997.
- 655 Brass, M. and Röckmann, T.: Continuous-flow isotope ratio mass spectrometry method for carbon and hydrogen
- isotope measurements on atmospheric methane, Atmos. Meas. Tech., 3(6), 1707–1721, doi:10.5194/amt-3-1707-
- 657 2010, 2010.
- 658 Burns, S. J.: Early diagenesis in Amazon fan sediments, in Proceeding of the Ocean Drilling Program, Scientific
- 659 Results, vol. 155, edited by R. D. Flood, D. J. W. Piper, A. Klaus, and L. C. Peterson, pp. 497–504., 1997.
- Burton, E. D., Sullivan, L. A., Bush, R. T., Johnston, S. G. and Keene, A. F.: A simple and inexpensive chromium-
- 661 reducible sulfur method for acid-sulfate soils, Appl. Geochemistry, 23(9), 2759–2766,
- doi:10.1016/j.apgeochem.2008.07.007, 2008.
- 663 Capet, A., Beckers, J.-M. and Grégoire, M.: Drivers, mechanisms and long-term variability of seasonal hypoxia on
- the Black Sea northwestern shelf is there any recovery after eutrophication?, Biogeosciences, 10, 3943–3962,
- doi:10.5194/bg-10-3943-2013, 2013.
- 666 Chanton, J., Chaser, L., Glasser, P. and Siegel, D.: Carbon and hydrogen isotopic effects in microbial methane from
- 667 terrestrial environments, in Stable isotopes and biosphere-atmosphere interactions, edited by L. Flanagan, pp. 85–

- 668 105, Elsevier-Academic Press., Amsterdam., 2005.
- 669 Chanton, J. P.: The effect of gas transport on the isotope signature of methane in wetlands, Org. Geochem., 36, 753–
- 670 768, doi:10.1016/j.orggeochem.2004.10.007, 2005.
- 671 Claff, S. R., Sullivan, L. A., Burton, E. D. and Bush, R. T.: A sequential extraction procedure for acid sulfate soils:
- Partitioning of iron, Geoderma, 155(3-4), 224–230, doi:10.1016/j.geoderma.2009.12.002, 2010.
- 673 Crowe, S. A., Katsev, S., Leslie, K., Sturm, A., Magen, C., Nomosatryo, S., Pack, M. A., Kessler, J. D., Reeburgh,
- W. S., Roberts, J. A., González, L., Douglas Haffner, G., Mucci, A., Sundby, B. and Fowle, D. A.: The methane
- 675 cycle in ferruginous Lake Matano, Geobiology, 9(1), 61–78, doi:10.1111/j.1472-4669.2010.00257.x, 2011.
- 676 Cruz Viggi, C., Rossetti, S., Fazi, S., Paiano, P., Majone, M. and Aulenta, F.: Magnetite particles triggering a faster
- and more robust syntrophic pathway of methanogenic propionate degradation, Environ. Sci. Technol., 48(13), 7536–
- 678 7543, doi:10.1021/es5016789, 2014.
- 679 Degens, E. T. and Ross, D. A.: The Black Sea geology, chemistry, and biology: American Association of
- Petroleum Geologists Memoir 20., 1974.
- 681 Dijkstra, N., Kraal, P., Kuypers, M. M. M., Schnetger, B. and Slomp, C. P.: Are iron-phosphate minerals a sink for
- phosphorus in anoxic Black Sea sediments?, PLoS One, 9(7), doi:10.1371/journal.pone.0101139, 2014.
- 683 Eckert, S., Brumsack, H.-J., Severmann, S., Schnetger, B., März, C. and Frollje, H.: Establishment of euxinic
- conditions in the Holocene Black Sea, Geology, 41(4), 431–434, 2013.
- 685 Egger, M., Jilbert, T., Behrends, T., Rivard, C. and Slomp, C. P.: Vivianite is a major sink for phosphorus in
- methanogenic coastal surface sediments, Geochim. Cosmochim. Acta, 169, 217–235, doi:10.1016/j.gca.2015.09.012,
- 687 2015a.
- Egger, M., Rasigraf, O., Sapart, C. J., Jilbert, T., Jetten, M. S. M., Röckmann, T., van der Veen, C., Bândă, N.,
- 689 Kartal, B., Ettwig, K. F. and Slomp, C. P.: Iron-mediated anaerobic oxidation of methane in brackish coastal
- 690 sediments, Environ. Sci. Technol., 49(1), 277–283, doi:10.1021/es503663z, 2015b.
- Ettwig, K. F., Butler, M. K., Le Paslier, D., Pelletier, E., Mangenot, S., Kuypers, M. M. M., Schreiber, F., Dutilh, B.
- 692 E., Zedelius, J., de Beer, D., Gloerich, J., Wessels, H. J. C. T., van Alen, T., Luesken, F., Wu, M. L., van de Pas-
- 693 Schoonen, K. T., Op den Camp, H. J. M., Janssen-Megens, E. M., Francoijs, K.-J., Stunnenberg, H., Weissenbach, J.,
- 694 Jetten, M. S. M. and Strous, M.: Nitrite-driven anaerobic methane oxidation by oxygenic bacteria, Nature, 464, 543–
- 695 548, doi:10.1038/nature08883, 2010.
- 696 Filippelli, G. M.: Controls on phosphorus concentration and accumulation in oceanic sediments, Mar. Geol., 139(1–
- 697 4), 231–240, 1997.
- 698 Froelich, P. N., Klinkhammer, G. P., Bender, M. L., Luedtke, N. A., Heath, G. R., Cullen, D., Dauphin, P.,
- 699 Hammond, D., Hartman, B. and Maynard, V.: Early oxidation of organic matter in pelagic sediments of the eastern
- equatorial Atlantic: suboxic diagenesis, Geochim. Cosmochim. Acta, 43(7), 1075-1090, doi:10.1016/0016-
- 701 7037(79)90095-4, 1979.
- 702 Grasshoff, K.: A simultaneous multiple channel system for nutrient analysis in seawater with analog and digital data

- record, Contrib. 2338, Woods Hole, 1969.
- Happell, J. D., Chanton, J. P. and Showers, W. J.: Methane transfer across the water-air interface in stagnant wooded
- swamps of Florida: Evaluation of mass-transfer coefficients and isotopic fractionation, 40(2), 290–298, 1995.
- Henkel, S., Mogollón, J. M., Nöthen, K., Franke, C., Bogus, K., Robin, E., Bahr, A., Blumenberg, M., Pape, T.,
- 707 Seifert, R., März, C., de Lange, G. J. and Kasten, S.: Diagenetic barium cycling in Black Sea sediments A case
- study for anoxic marine environments, Geochim. Cosmochim. Acta, 88, 88–105, doi:10.1016/j.gca.2012.04.021,
- 709 2012.
- 710 Hindmarsh, A. C.: ODEPACK, a systematized collection of ODE solvers, in IMACS Transactions on Scientific
- Computation, Vol. 1, edited by R. S. Stepleman and Others, pp. 55–64, North-Holland, Amsterdam., 1983.
- Holler, T., Wegener, G., Knittel, K., Boetius, A., Brunner, B., Kuypers, M. M. M. and Widdel, F.: Substantial (13)
- 713 C/(12) C and D/H fractionation during anaerobic oxidation of methane by marine consortia enriched in vitro,
- 714 Environ. Microbiol. Rep., 1(5), 370–6, doi:10.1111/j.1758-2229.2009.00074.x, 2009.
- Holler, T., Wegener, G., Niemann, H., Ferdelman, T. G., Boetius, A., Kristiansen, T. Z., Molina, H., Pandey, A.,
- Werner, J. K., Juluri, K. R., Xu, Y., Glenn, D., Parang, K. and Snyder, S. H.: Carbon and sulfur back flux during
- anaerobic microbial oxidation of methane and coupled sulfate reduction, Proc. Natl. Acad. Sci., 109(51), 21170-
- 718 21170, doi:10.1073/pnas.1218683109, 2012.
- 719 Holmkvist, L., Ferdelman, T. G. and Jørgensen, B. B.: A cryptic sulfur cycle driven by iron in the methane zone of
- 720 marine sediment (Aarhus Bay, Denmark), Geochim. Cosmochim. Acta, 75(12), 3581-3599,
- 721 doi:10.1016/j.gca.2011.03.033, 2011a.
- 722 Holmkvist, L., Kamyshny, A., Brüchert, V., Ferdelman, T. G. and Jørgensen, B. B.: Sulfidization of lacustrine
- 723 glacial clay upon Holocene marine transgression (Arkona Basin, Baltic Sea), Geochim. Cosmochim. Acta, 142, 75–
- 724 94, doi:10.1016/j.gca.2014.07.030, 2014.
- 725 Holmkvist, L., Kamyshny, A., Vogt, C., Vamvakopoulos, K., Ferdelman, T. G. and Jørgensen, B. B.: Sulfate
- reduction below the sulfate-methane transition in Black Sea sediments, Deep Sea Res. Part I Oceanogr. Res. Pap.,
- 727 58(5), 493–504, doi:10.1016/j.dsr.2011.02.009, 2011b.
- 728 Hsu, T.-W., Jiang, W.-T. and Wang, Y.: Authigenesis of vivianite as influenced by methane-induced sulfidization in
- 729 cold-seep sediments off southwestern Taiwan, J. Asian Earth Sci., 89, 88–97, doi:10.1016/j.jseaes.2014.03.027,
- 730 2014.
- 731 Ingall, E. D., Bustin, R. M. and Van Cappellen, P.: Influence of water column anoxia on the burial and preservation
- of carbon and phosphorus in marine shales, Geochim. Cosmochim. Acta, 57(2), 303–316, 1993.
- Jilbert, T. and Slomp, C. P.: Iron and manganese shuttles control the formation of authigenic phosphorus minerals in
- the euxinic basins of the Baltic Sea, Geochim. Cosmochim. Acta, 107, 155–169, doi:10.1016/j.gca.2013.01.005,
- 735 2013.
- 736 Jørgensen, B. B.: A comparison of methods for the quantification of bacterial sulfate reduction in coastal marine
- 737 sediments. II. Calculations from mathematical models, Geomicrobiol. J., 1, 29-51,

- 738 doi:10.1080/01490457809377721, 1978.
- 739 Jørgensen, B. B., Böttcher, M. E., Lüschen, H., Neretin, L. N. and Volkov, I. I.: Anaerobic methane oxidation and a
- deep H2S sink generate isotopically heavy sulfides in Black Sea sediments, Geochim. Cosmochim. Acta, 68(9),
- 741 2095–2118, doi:10.1016/j.gca.2003.07.017, 2004.
- 742 Jørgensen, B. B., Weber, A. and Zopfi, J.: Sulfate reduction and anaerobic methane oxidation in Black Sea
- 743 sediments, Deep Sea Res. Part I Oceanogr. Res. Pap., 48, 2097–2120, doi:10.1016/S0967-0637(01)00007-3, 2001.
- 744 Kato, S., Hashimoto, K. and Watanabe, K.: Methanogenesis facilitated by electric syntrophy via (semi)conductive
- 745 iron-oxide minerals, Environ. Microbiol., 14(7), 1646–1654, doi:10.1111/j.1462-2920.2011.02611.x, 2012.
- Kemp, W. M., Testa, J. M., Conley, D. J., Gilbert, D. and Hagy, J. D.: Temporal responses of coastal hypoxia to
- 747 nutrient loading and physical controls, Biogeosciences, 6(12), 2985–3008, doi:10.5194/bg-6-2985-2009, 2009.
- Knab, N. J., Cragg, B. A., Hornibrook, E. R. C., Holmkvist, L., Pancost, R. D., Borowski, C., Parkes, R. J. and
- 749 Jørgensen, B. B.: Regulation of anaerobic methane oxidation in sediments of the Black Sea, Biogeosciences, (6),
- 750 1505–1518, doi:10.5194/bg-6-1505-2009, 2009.
- 751 Knittel, K. and Boetius, A.: Anaerobic oxidation of methane: progress with an unknown process, Annu. Rev.
- 752 Microbiol., 63, 311–334, doi:10.1146/annurev.micro.61.080706.093130, 2009.
- 753 Kraal, P. and Slomp, C. P.: Rapid and extensive alteration of phosphorus speciation during oxic storage of wet
- 754 sediment samples, PLoS One, 9(5), e96859, 2014.
- 755 Kraal, P., Slomp, C. P., Forster, A. and Kuypers, M. M. M.: Pyrite oxidation during sample storage determines
- 756 phosphorus fractionation in carbonate-poor anoxic sediments, Geochim. Cosmochim. Acta, 73(11), 3277–3290,
- 757 doi:10.1016/j.gca.2009.02.026, 2009.
- 758 Leloup, J., Fossing, H., Kohls, K., Holmkvist, L., Borowski, C. and Jørgensen, B. B.: Sulfate-reducing bacteria in
- 759 marine sediment (Aarhus Bay, Denmark): abundance and diversity related to geochemical zonation, Environ.
- 760 Microbiol., 11(5), 1278–1291, doi:10.1111/j.1462-2920.2008.01855.x, 2009.
- Leloup, J., Loy, A., Knab, N. J., Borowski, C., Wagner, M. and Jørgensen, B. B.: Diversity and abundance of sulfate-
- 762 reducing microorganisms in the sulfate and methane zones of a marine sediment, Black Sea, Environ. Microbiol.,
- 763 9(1), 131–142, doi:10.1111/j.1462-2920.2006.01122.x, 2007.
- 764 Li, H., Chang, J., Liu, P., Fu, L., Ding, D. and Lu, Y.: Direct interspecies electron transfer accelerates syntrophic
- oxidation of butyrate in paddy soil enrichments, Environ. Microbiol., 17, 1–45, doi:10.1111/1462-2920.12576, 2014.
- Liu, D., Wang, H., Dong, H., Qiu, X., Dong, X. and Cravotta, C. A.: Mineral transformations associated with
- 767 goethite reduction by Methanosarcina barkeri, Chem. Geol., 288(1-2), 53-60, doi:10.1016/j.chemgeo.2011.06.013,
- 768 2011.
- Lovley, D. R.: Dissimilatory Fe(III) and Mn(IV) reduction, Microbiol. Rev., 55(2), 259–287, 1991.
- 770 Lovley, D. R. and Phillips, E. J. P.: Competitive mechanisms for inhibition of sulfate reduction and methane
- production in the zone of ferric iron reduction in sediments, Appl. Environ. Microbiol., 53(11), 1987.

- 772 Lovley, D. R., Phillips, E. J. P. and Lonergan, D. J.: Hydrogen and formate oxidation coupled to dissimilatory
- reduction of iron or manganese by Alteromonas putrefaciens, Appl. Environ. Microbiol., 55(3), 700–706, 1989.
- Martens, C. S., Albert, D. B. and Alperin, M. J.: Stable isotope tracing of anaerobic methane oxidation in the gassy
- 775 sediments of Eckernforde Bay, German Baltic Sea, Am. J. Sci., 299, 589–610, doi:10.2475/ajs.299.7-9.589, 1999.
- Martens, C. S., Berner, R. A. and Rosenfeld, J. K.: Interstitial water chemistry of anoxic Long Island Sound
- sediments. 2. Nutrient regeneration and phosphate removal, Limnol. Oceanogr., 23(4), 605–617, 1978.
- März, C., Hoffmann, J., Bleil, U., de Lange, G. J. and Kasten, S.: Diagenetic changes of magnetic and geochemical
- signals by anaerobic methane oxidation in sediments of the Zambezi deep-sea fan (SW Indian Ocean), Mar. Geol.,
- 780 255(3-4), 118–130, doi:10.1016/j.margeo.2008.05.013, 2008.
- Meysman, F. J. R., Boudreau, B. P. and Middelburg, J. J.: Modeling reactive transport in sediments subject to
- 782 bioturbation and compaction, Geochim. Cosmochim. Acta, 69(14), 3601–3617, doi:10.1016/j.gca.2005.01.004,
- 783 2005.
- Milucka, J., Ferdelman, T. G., Polerecky, L., Franzke, D., Wegener, G., Schmid, M., Lieberwirth, I., Wagner, M.,
- Widdel, F. and Kuypers, M. M. M.: Zero-valent sulphur is a key intermediate in marine methane oxidation, Nature,
- 786 491(7425), 541–6, doi:10.1038/nature11656, 2012.
- 787 Mogollón, J. M., Dale, A. W., Jensen, J. B., Schlüter, M. and Regnier, P.: A method for the calculation of anaerobic
- 788 oxidation of methane rates across regional scales: An example from the Belt Seas and The Sound (North Sea-Baltic
- 789 Sea transition), Geo-Marine Lett., 33(4), 299–310, doi:10.1007/s00367-013-0329-z, 2013.
- 790 Moodley, L., Middelburg, J. J., Herman, P. M. J., Soetaert, K. and de Lange, G. J.: Oxygenation and organic-matter
- 791 preservation in marine sediments: Direct experimental evidence from ancient organic carbon–rich deposits, Geology,
- 792 33(11), 889, doi:10.1130/G21731.1, 2005.
- 793 Mort, H. P., Slomp, C. P., Gustafsson, B. G. and Andersen, T. J.: Phosphorus recycling and burial in Baltic Sea
- 794 sediments with contrasting redox conditions, Geochim. Cosmochim. Acta, 74(4), 1350–1362,
- 795 doi:10.1016/j.gca.2009.11.016, 2010.
- Murphy, J. and Riley, J. P.: A modified single solution method for the determination of phosphate in natural waters,
- 797 Anal. Chim. Acta, 27, 31–36, doi:10.1016/S0003-2670(00)88444-5, 1962.
- 798 Neretin, L. N., Böttcher, M. E., Jørgensen, B. B., Volkov, I. I., Lüschen, H. and Hilgenfeldt, K.: Pyritization
- 799 processes and greigite formation in the advancing sulfidization front in the Upper Pleistocene sediments of the Black
- 800 Sea, Geochim. Cosmochim. Acta, 68(9), 2081–2093, doi:10.1016/S0016-7037(00)00450-2, 2004.
- 801 Petzoldt, L. R.: Automatic selection of methods for solving stiff and nonstiff systems of ordinary differential
- 802 equations, SIAM J. Sci. Stat. Comput., 4, 136–148, doi:10.1137/0904010, 1983.
- 803 Poulton, S. and Canfield, D.: Development of a sequential extraction procedure for iron: implications for iron
- 804 partitioning in continentally derived particulates, Chem. Geol., 214(3-4), 209-221,
- 805 doi:10.1016/j.chemgeo.2004.09.003, 2005.
- 806 Qu, D., Ratering, S. and Schnell, S.: Microbial reduction of weakly crystalline iron (III) oxides and suppression of

- 807 methanogenesis in paddy soil, Bull. Environ. Contam. Toxicol., 72(6), 1172–1181, doi:10.1007/s00128-004-0367-3,
- 808 2004.
- Raghoebarsing, A. A., Pol, A., van de Pas-Schoonen, K. T., Smolders, A. J. P., Ettwig, K. F., Rijpstra, W. I. C.,
- Schouten, S., Damsté, J. S. S., Op den Camp, H. J. M., Jetten, M. S. M. and Strous, M.: A microbial consortium
- couples anaerobic methane oxidation to denitrification, Nature, 440(7086), 918–921, doi:10.1038/nature04617, 2006.
- Raiswell, R. and Canfield, D. E.: The iron biogeochemical cycle past and present, Geochemical Perspect., 1(1),
- 813 doi:10.7185/geochempersp.1.1, 2012.
- Rayleigh, J. W. S.: Theoretical considerations respecting the separation of gases by diffusion and similar processes,
- 815 Philos. Mag., (42), 493–499, 1896.
- 816 Reeburgh, W.: Oceanic methane biogeochemistry, Am. Chem. Soc., 107(2), 486–513, doi:10.1021/cr050362v, 2007.
- Reed, D. C., Slomp, C. P. and Gustafsson, B. G.: Sedimentary phosphorus dynamics and the evolution of bottom-
- water hypoxia: A coupled benthic-pelagic model of a coastal system, Limnol. Oceanogr., 56(3), 1075–1092,
- 819 doi:10.4319/lo.2011.56.3.1075, 2011a.
- 820 Reed, D. C., Slomp, C. P. and de Lange, G. J.: A quantitative reconstruction of organic matter and nutrient
- diagenesis in Mediterranean Sea sediments over the Holocene, Geochim. Cosmochim. Acta, 75(19), 5540-5558,
- 822 doi:10.1016/j.gca.2011.07.002, 2011b.
- 823 Regnier, P., Dale, A. W., Arndt, S., LaRowe, D. E., Mogollón, J. and Van Cappellen, P.: Quantitative analysis of
- anaerobic oxidation of methane (AOM) in marine sediments: A modeling perspective, Earth-Science Rev., 106(1-2),
- 825 105–130, doi:10.1016/j.earscirev.2011.01.002, 2011.
- 826 Reiche, M., Torburg, G. and Küsel, K.: Competition of Fe(III) reduction and methanogenesis in an acidic fen, FEMS
- 827 Microbiol. Ecol., 65(1), 88–101, doi:10.1111/j.1574-6941.2008.00523.x, 2008.
- Rickard, D. and Luther, G. W.: Kinetics of pyrite formation by the H2S oxidation of iron (II) monosulfide in aqueous
- 829 solutions between 25 and 125°C: The mechanism, Geochim. Cosmochim. Acta, 61(1), 135–147, doi:10.1016/S0016-
- 830 7037(96)00322-5, 1997.
- 831 Riedinger, N., Formolo, M. J., Lyons, T. W., Henkel, S., Beck, A. and Kasten, S.: An inorganic geochemical
- 832 argument for coupled anaerobic oxidation of methane and iron reduction in marine sediments, Geobiology,
- 833 doi:10.1111/gbi.12077, 2014.
- 834 Rooze, J., Egger, M., Tsandev, I. and Slomp, C. P.: Iron-dependent anaerobic oxidation of methane in coastal surface
- sediments: potential controls and impact, Limnol. Oceanogr., doi:10.1002/lno.10275, 2016.
- 836 Ruttenberg, K. C.: Development of a sequential extraction method for different forms of phosphorus in marine
- 837 sediments, Limnol. Oceanogr., 37(7), 1460–1482, 1992.
- 838 Ruttenberg, K. C. and Berner, R. A.: Authigenic apatite formation and burial in sediments from non-upwelling,
- 839 continental margin environments, Geochim. Cosmochim. Acta, 57(5), 991–1007, doi:10.1016/0016-7037(93)90035-
- 840 U, 1993.

- Van Santvoort, P. J. M., De Lange, G. J., Thomson, J., Colley, S., Meysman, F. J. R. and Slomp, C. P.: Oxidation
- and origin of organic matter in surficial Eastern Mediterranean hemipelagic sediments, Aquat. Geochemistry, 8(3),
- 843 153–175, doi:10.1023/A:1024271706896, 2002.
- Sapart, C. J., van der Veen, C., Vigano, I., Brass, M., van de Wal, R. S. W., Bock, M., Fischer, H., Sowers, T.,
- 845 Buizert, C., Sperlich, P., Blunier, T., Behrens, M., Schmitt, J., Seth, B. and Röckmann, T.: Simultaneous stable
- isotope analysis of methane and nitrous oxide on ice core samples, Atmos. Meas. Tech., 4(12), 2607–2618,
- 847 doi:10.5194/amt-4-2607-2011, 2011.
- 848 Scheller, S., Yu, H., Chadwick, G. L., McGlynn, S. E. and Orphan, V. J.: Artificial electron acceptors decouple
- 849 archaeal methane oxidation from sulfate reduction, Science (80-.)., 351(6274), 703-707,
- 850 doi:10.1126/science.aad7154, 2016.
- 851 Schulz, H. D., Dahmke, A., Schinzel, U., Wallmann, K. and Zabel, M.: Early diagenetic processes, fluxes, and
- reaction rates in sediments of the South Atlantic, Geochim. Cosmochim. Acta, 58(9), 2041–2060, doi:10.1016/0016-
- 853 7037(94)90284-4, 1994.
- 854 Segarra, K. E. A., Comerford, C., Slaughter, J. and Joye, S. B.: Impact of electron acceptor availability on the
- anaerobic oxidation of methane in coastal freshwater and brackish wetland sediments, Geochim. Cosmochim. Acta,
- 856 115, 15–30, doi:10.1016/j.gca.2013.03.029, 2013.
- 857 Seifert, R., Nauhaus, K., Blumenberg, M., Krüger, M. and Michaelis, W.: Methane dynamics in a microbial
- community of the Black Sea traced by stable carbon isotopes in vitro, Org. Geochem., 37(10), 1411–1419,
- 859 doi:10.1016/j.orggeochem.2006.03.007, 2006.
- 860 Sivan, O., Adler, M., Pearson, A., Gelman, F., Bar-Or, I., John, S. G. and Eckert, W.: Geochemical evidence for
- 861 iron-mediated anaerobic oxidation of methane, Limnol. Oceanogr., 56(4), 1536–1544,
- 862 doi:10.4319/lo.2011.56.4.1536, 2011.
- Sivan, O., Antler, G., Turchyn, A. V., Marlow, J. J. and Orphan, V. J.: Iron oxides stimulate sulfate-driven anaerobic
- methane oxidation in seeps, Proc. Natl. Acad. Sci., 1–9, doi:10.1073/pnas.1412269111, 2014.
- 865 Sivan, O., Shusta, S. and Valentine, D. L.: Methanogens rapidly transition from methane production to iron
- 866 reduction, Geobiology, doi:10.1111/gbi.12172, 2016.
- 867 Slomp, Van der Gaast, S. J. and Van Raaphorst, W.: Phosphorus binding by poorly crystalline iron oxides in North
- 868 Sea sediments, Mar. Chem., 52(1), 55–73, doi:10.1016/0304-4203(95)00078-X, 1996a.
- 869 Slomp, C. P., Epping, E. H. G., Helder, W. and Raaphorst, W. Van: A key role for iron-bound phosphorus in
- authigenic apatite formation in North Atlantic continental platform sediments, J. Mar. Res., 54(6), 1179–1205,
- 871 doi:10.1357/0022240963213745, 1996b.
- 872 Slomp, C. P., Mort, H. P., Jilbert, T., Reed, D. C., Gustafsson, B. G. and Wolthers, M.: Coupled dynamics of iron
- and phosphorus in sediments of an oligotrophic coastal basin and the impact of anaerobic oxidation of methane.
- PLoS One, 8(4), e62386, doi:10.1371/journal.pone.0062386, 2013.
- 875 Soetaert, K., Herman, P. M. J. and Middelburg, J. J.: A model of early diagenetic processes from the shelf to abyssal

- depths, Geochim. Cosmochim. Acta, 60(6), 1019–1040, doi:10.1016/0016-7037(96)00013-0, 1996.
- 877 Soetaert, K. and Meysman, F.: Reactive transport in aquatic ecosystems: Rapid model prototyping in the open source
- software R, Environ. Model. Softw., 32, 49–60, doi:10.1016/j.envsoft.2011.08.011, 2012.
- Soetaert, K., Petzoldt, T. and Meysman, F. J. R.: marelac: Tools for Aquatic Sciences. R Package Version 2.1.3,
- 880 2010.
- 881 Soulet, G., Delaygue, G., Vallet-Coulomb, C., Böttcher, M. E., Sonzogni, C., Lericolais, G. and Bard, E.: Glacial
- hydrologic conditions in the Black Sea reconstructed using geochemical pore water profiles, Earth Planet. Sci. Lett.,
- 883 296(1-2), 57–66, doi:10.1016/j.epsl.2010.04.045, 2010.
- 884 Soulet, G., Ménot, G., Lericolais, G. and Bard, E.: A revised calendar age for the last reconnection of the Black Sea
- to the global ocean, Quat. Sci. Rev., 30(9-10), 1019–1026, doi:10.1016/j.quascirev.2011.03.001, 2011.
- Stoll, M. H. C., Bakker, K., Nobbe, G. H. and Haese, R. R.: Continous-Flow Analysis of Dissolved Inorganic Carbon
- 887 Content in Seawater, Anal. Chem., 73(17), 4111–4116, doi:10.1021/ac010303r, 2001.
- 888 Strickland, J. D. and Parsons, T. R.: A Practical Handbook of Seawater Analysis, Bulletin 1., Fisheries Research
- 889 Board of Canada, Ottawa, Canada., 1972.
- 890 Suess, E.: Mineral phases formed in anoxic sediments by microbial decomposition of organic matter, Geochim.
- 891 Cosmochim. Acta, 43, 339–352, doi:10.1016/0016-7037(79)90199-6, 1979.
- 892 Thamdrup, B., Roselló-Mora, R. and Amann, R.: Microbial manganese and sulfate reduction in Black Sea shelf
- 893 sediments, Appl. Environ. Microbiol., 66(2251), 2288–2297, doi:10.1128/AEM.66.7.2888-2897.2000, 2000.
- Treude, T., Krause, S., Maltby, J., Dale, A. W., Coffin, R. and Hamdan, L. J.: Sulfate reduction and methane
- oxidation activity below the sulfate-methane transition zone in Alaskan Beaufort Sea continental margin sediments:
- 896 Implications for deep sulfur cycling, Geochim. Cosmochim. Acta, 144, 217–237, doi:10.1016/j.gca.2014.08.018,
- 897 2014.
- 898 Vargas, M., Kashefi, K., Blunt-Harris, E. L. and Lovley, D. R.: Microbial evidence for Fe(III) reduction on early
- 899 Earth, Nature, 395, 65–67, doi:10.1038/25720, 1998.
- 900 Wang, Y. and Van Cappellen, P.: A multicomponent reactive transport model of early diagenesis: Application to
- 901 redox cycling in coastal marine sediments, Geochim. Cosmochim. Acta, 60(16), 2993–3014, doi:10.1016/0016-
- 902 7037(96)00140-8, 1996.
- Wankel, S. D., Adams, M. M., Johnston, D. T., Hansel, C. M., Joye, S. B. and Girguis, P. R.: Anaerobic methane
- oxidation in metalliferous hydrothermal sediments: influence on carbon flux and decoupling from sulfate reduction,
- 905 Environ. Microbiol., 14(10), 2726–40, doi:10.1111/j.1462-2920.2012.02825.x, 2012.
- Whiticar, M. J.: Carbon and hydrogen isotope systematics of bacterial formation and oxidation of methane, Chem.
- 907 Geol., 161(1-3), 291–314, doi:10.1016/S0009-2541(99)00092-3, 1999.
- 908 Yoshinaga, M. Y., Holler, T., Goldhammer, T., Wegener, G., Pohlman, J. W., Brunner, B., Kuypers, M. M. M.,
- 909 Hinrichs, K. and Elvert, M.: Carbon isotope equilibration during sulphate-limited anaerobic oxidation of methane,

- 910 Nat. Geosci., 7, 190–194, doi:10.1038/NGEO2069, 2014.
- 211 Zhou, S., Xu, J., Yang, G. and Zhuang, L.: Methanogenesis affected by the co-occurrence of iron(III) oxides and
- 912 humic substances, FEMS Microbiol. Ecol., 88(1), 107–120, doi:10.1111/1574-6941.12274, 2014.
- 213 Zhuang, L., Xu, J., Tang, J. and Zhou, S.: Effect of ferrihydrite biomineralization on methanogenesis in an anaerobic
- 914 incubation from paddy soil, J. Geophys. Res. Biogeosciences, (120), 876–886, doi:10.1002/2014JG002893, 2015.

917 Tables

918

Table 1. Overview of the sequential P, Fe and S fractionation methods used in this study.

Step and code	Extractant, extraction time	Target phase		
P fractionation (1	modified from Ruttenberg (1992); done for site 4 (MC &	& GC) and site 5 (MC & GC))		
1 P _{exch}	1 M MgCl ₂ , pH 8, 0.5 h	Exchangeable P		
2ª P _{Fe}	25 g L ⁻¹ Na dithionite, pH 7.5, 8 h	Fe-associated P		
3ª Pauthi Ca-P	Na acetate buffer, pH 4, 6 h	P in authigenic and biogenic Ca-P minerals and CaCC		
4 P _{detr}	1 M HCl, 24 h	Detrital P		
5 Porg	Ashing at 550 °C (2h), then 1 M HCl, 24 h	Organic P		
Fe fractionation	(after Poulton and Canfield (2005); done for site 4 (MC	& GC) and site 5 (MC))		
1 Fe _{carb}	1 M Na acetate, pH 4.5, 24 h	Carbonate-associated Fe		
2 Fe _{ox1}	1 M hydroxylamine-HCl, 24 h	Amorphous Fe oxides (ferrihydrite)		
3 Fe _{ox2}	50 g L ⁻¹ Na dithionite, pH 4.8, 2 h	Crystalline Fe oxides (goethite, hematite)		
4 Fe _{mag}	0.2 M ammonium oxalate/ 0.17 M oxalic acid, 2 h	Recalcitrant Fe oxides (mostly magnetite)		
Fe fractionation	(modified from Claff et al. (2010); done for site 4 (MC &	& GC) and site 5 (MC))		
1 Fe(II) _{HCl}	1 M HCl, 4 h	Labile Fe (carbonates, poorly ordered sulfides)		
2 Fe(III) _{HCl}	1 M HCl, 4 h	Labile Fe (easily reducible oxides)		
3 Fe(III) _{CDB}	50 g L ⁻¹ Na dithionite, pH 4.8, 4 h	Crystalline Fe oxides		
4 Fe _{pyrite}	Concentrated HNO ₃ , 2 h	Pyrite (FeS ₂)		
S fractionation (a	after Burton et al. (2008); done for site 4 (MC) and site 5	5 (MC & GC))		
1 AVS	6 M HCl, 24 h	S in Fe monosulfides (FeS)		
2 CRS	Acidic chromous chloride solution, 48 h	S in pyrite (FeS ₂)		

^{920 =} gravity core.

919

Table 2. Chemical species included in the diagenetic model.

Species	Notation	Type
Organic matter ^a	$OM^{\propto,\beta,\gamma}$	Solid
Iron oxides ^a	$Fe(OH)_3^{\alpha,\beta,\gamma}$	Solid
Iron monosulfide	FeS	Solid
Pyrite	FeS_2	Solid
Siderite	$FeCO_3$	Solid
Elemental sulfur	S_0	Solid
Iron oxide-bound phosphorus	$Fe_{ox}P$	Solid
Vivianite	$Fe_3(PO_4)_2$	Solid
Organic phosphorus	P_{org}	Solid
Authigenic (Ca) phosphorus	СаР	Solid
Detrital phosphorus	DetrP	Solid
Chloride	Cl^-	Solute
Oxygen	O_2	Solute
Sulfate	SO_4^{2-}	Solute
Iron	Fe^{2+}	Solute
Hydrogen sulfide ^b	$\sum H_2 S$	Solute
Methane	CH_4	Solute
Ammonium ^b	$\sum NH_4^+$	Solute
Nitrate	NO_3^-	Solute
Phosphate	ΣHPO_4^{2-}	Solute
Dissolved inorganic carbon	DIC	Solute

There are three types of species: reactive (α) , less reactive (β) and refractory (γ)

Primary redox reactions*

926

$$OM^{\infty,\beta} + aO_2 \rightarrow aCO_2 + bNH_4^+ + cH_3PO_4 + aH_2O$$

$$OM^{\infty,\beta} + \frac{4a}{5}NO_3^- + \frac{4a}{5}H^+ \rightarrow aCO_2 + bNH_4^+ + cH_3PO_4 + \frac{2a}{5}N_2 + \frac{7a}{5}H_2O$$

$$R2$$

$$OM^{\infty,\beta} + 4aFe(OH)_3^{\infty} + 4a\chi^{\infty}Fe_{ox}P + 12aH^+ \rightarrow aCO_2 + bNH_4^+ + (c + 4a\chi^{\infty})H_3PO_4 + 4aFe^{2+} + 13aH_2O$$

$$R3$$

$$OM^{\infty,\beta} + \frac{a}{2}SO_4^{2-} + aH^+ \rightarrow aCO_2 + bNH_4^+ + cH_3PO_4 + \frac{a}{2}H_2S + aH_2O$$

$$R4$$

$$OM^{\infty,\beta} \rightarrow \frac{a}{2}CO_2 + bNH_4^+ + cH_3PO_4 + \frac{a}{2}CH_4$$

$$R5$$

$$CO_2 + 4H_2 \rightarrow CH_4 + 2H_2O$$

Secondary redox and other reaction equations[†]

$$\begin{array}{c} 2O_2 + NH_+^4 + 2HCO_3^- \rightarrow NO_3^- + 2CO_2 + 3H_2O \\ O_2 + 4Fe^{2^+} + 8HCO_3^- + 2H_2O + 4\chi^\times H_2PO_4^- \rightarrow 4Fe(OH)_3^\times + 4\chi^\times Fe_{ox}P + 8CO_2 \\ R8 \\ 2O_2 + FeS \rightarrow SO_4^{2^-} + Fe^{2^+} \\ R9 \\ 7O_2 + 2FeS_2 + 2H_2O \rightarrow 4SO_4^{2^-} + 2Fe^{2^+} + 4H^+ \\ R10 \\ 2O_2 + H_2S + 2HCO_3^- \rightarrow SO_4^{2^-} + 2CO_2 + 2H_2O \\ R11 \\ 2O_2 + CH_4 \rightarrow CO_2 + 2H_2O \\ R12 \\ 2Fe(OH)_3^\infty + 2\chi^\times Fe_{ox}P + H_2S + 4CO_2 \rightarrow 2Fe^{2^+} + 2\chi^\times H_2PO_4^- + S_0 + 4HCO_3^- + 2H_2O \\ R13 \\ 2Fe(OH)_3^0 + 2\chi^\beta Fe_{ox}P + H_2S + 4CO_2 \rightarrow 2Fe^{2^+} + 2\chi^\beta H_2PO_4^- + S_0 + 4HCO_3^- + 2H_2O \\ R14 \\ Fe^{2^+} + H_2S \rightarrow FeS + 2H^+ \\ FeS + H_2S \rightarrow FeS_2 + H_2 \\ R16 \\ 4S_0 + 4H_2O \rightarrow 3H_2S + SO_4^{2^-} + 2H^+ \\ R17 \\ FeS + S_0 \rightarrow FeS_2 \\ R18 \\ SO_4^{2^-} + CH_4 + CO_2 \rightarrow 2HCO_3^- + H_2S \\ CH_4 + 8Fe(OH)_3^\infty + 8\chi^\infty Fe_{ox}P + 15H^+ \rightarrow HCO_3^- + 8Fe^{2^+} + 8\chi^\infty FH_2PO_4^- + 21H_2O \\ R20 \\ Fe(OH)_3^\infty + (\chi^\infty - \chi^\beta)Fe_{ox}P \rightarrow Fe(OH)_3^\beta + (\chi^\infty - \chi^\beta)H_2PO_4^- \\ R21 \\ Fe(OH)_3^\infty + (\chi^\beta - \chi^\gamma)Fe_{ox}P \rightarrow Fe(OH)_3^\beta + (\chi^\infty - \chi^\beta)H_2PO_4^- \\ R22 \\ 3Fe^{2^+} + 2HPO_4^{2^-} \rightarrow Fe_3(PO_4)_2 + 2H^+ \\ R23 \\ Fe^{2^+} + CO_3^{2^-} \rightarrow FeCO_3 \\ R24 \\ FeCO_3 + H_2S \rightarrow FeS + HCO_3^- + H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R25 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R26 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R26 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R26 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R26 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R26 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R27 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R28 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R29 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R20 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HPO_4^{2^-} + 4H^+ \\ R20 \\ Fe_3(PO_4)_2 + 3H_2S \rightarrow 3FeS + 2HP$$

927 * Organic matter (OM) is of the form $(CH_2O)_a(NH_4^+)_b(H_3PO_4)_c$, with 'a'=1, 'b' = 1/16 and 'c' = 1/106. Under anoxic bottom 928 water conditions, 'c' reduces to 0.25 to account for the preferential regeneration of P (e.g. Ingall et al. (1993)). † $\chi^{\alpha,\beta,\gamma}$ refers to the P:Fe ratio of $Fe(OH)_3^{\alpha,\beta,\gamma}$ (see Supplementary Table S1). $R6 = CH_4$ production from DIC (i.e. CO_2); R7 = nitrification; R8 =929 Fe(OH)₃ formation; R9 = FeS oxidation; $R10 = \text{FeS}_2$ oxidation; $R11 = \text{H}_2\text{S}$ oxidation; $R12 = \text{aerobic CH}_4$ oxidation; R13 and R14930 931 = Fe(OH)₃ reduction by H₂S; R15= FeS formation; R16 = pyrite formation (H₂S pathway); R17 = S₀ disproportionation; R18 = 932 pyrite formation (polysulfide pathway); $R19 = SO_4$ -AOM; R20 = Fe-AOM; R21 = conversion (i.e. crystallization) from α to β 933 phase; R22 = crystallization from β to γ phase; R23 = vivianite formation; R24 = siderite precipitation; R25 = conversion from 934 siderite to FeS; R26 = vivianite dissolution by dissolved sulfide

Table 4. Reaction equations implemented in the model.

Primary redox reaction equations

$$R_1 = k_{\alpha,\beta} O M^{\alpha,\beta} \left(\frac{[o_2]}{K_{o_2} + [o_2]} \right) \tag{E1}$$

$$\boldsymbol{R}_{2} = \boldsymbol{k}_{\alpha,\beta} \boldsymbol{O} \boldsymbol{M}^{\alpha,\beta} \left(\frac{[NO_{3}^{-}]}{K_{NO_{3}^{-}} + [NO_{3}^{-}]} \right) \left(\frac{K_{O_{2}}}{K_{O_{2}} + [O_{2}]} \right)$$
(E2)

$$R_{3} = k_{\alpha,\beta} O M^{\alpha,\beta} \left(\frac{[Fe(OH)_{3}^{\alpha}]}{K_{Fe(OH)_{3}^{\alpha}+}[Fe(OH)_{3}^{\alpha}]} \right) \left(\frac{K_{NO_{3}^{-}}}{K_{NO_{3}^{-}+}[NO_{3}]} \right) \left(\frac{K_{O_{2}}}{K_{O_{2}^{-}+}[O_{2}]} \right)$$
(E3)

$$\boldsymbol{R_4} = \Psi \boldsymbol{k}_{\alpha,\beta} \boldsymbol{O} \boldsymbol{M}^{\alpha,\beta} \left(\frac{[so_4^{2-}]}{K_{so_4^{2-}} + [so_4^{2-}]} \right) \left(\frac{K_{Fe(OH)_3^{\alpha}}}{K_{Fe(OH)_3^{\alpha}} + [Fe(OH)_3^{\alpha}]} \right) \left(\frac{K_{No_3^{-}}}{K_{No_3^{-}} + [No_3^{-}]} \right) \left(\frac{K_{o_2}}{K_{o_2} + [o_2]} \right)$$
 (E4)

$$R_{5} = \Psi k_{\alpha,\beta} O M^{\alpha,\beta} \left(\frac{K_{SO_{4}^{2}}}{K_{SO_{4}^{2}} + [SO_{4}^{2}]} \right) \left(\frac{K_{Fe(OH)_{3}^{\alpha}}}{K_{Fe(OH)_{3}^{\alpha}} + [Fe(OH)_{3}^{\alpha}]} \right) \left(\frac{K_{NO_{3}^{-}}}{K_{NO_{3}^{-}} + [NO_{3}]} \right) \left(\frac{K_{O_{2}}}{K_{O_{2}} + [O_{2}]} \right)$$
(E5)

$$R_{6} = k_{1}DIC\left(\frac{K_{SO_{4}^{2}-}}{K_{SO_{4}^{2}-}|SO_{4}^{2}-|}\right)\left(\frac{K_{Fe(OH)_{3}^{\alpha}}}{K_{Fe(OH)_{3}^{\alpha}}+|Fe(OH)_{3}^{\alpha}|}\right)\left(\frac{K_{NO_{3}^{-}}}{K_{NO_{3}^{-}}+|NO_{3}^{-}|}\right)\left(\frac{K_{O_{2}}}{K_{O_{2}}+|O_{2}|}\right)$$
(E6)

Secondary redox and other reaction equations

$$R_7 = k_2[O_2][NH_4^+]$$

$$R_8 = k_3[O_2][Fe^{2+}]$$
(E8)

$$R_9 = k_4[O_2][FeS] \tag{E9}$$

$$R_{10} = k_5[O_2][FeS_2]$$
 (E10)

$$R_{11} = k_6[O_2][\Sigma H_2 S]$$
 (E11)

$$R_{12} = k_7[\mathbf{O}_2][\mathbf{C}\mathbf{H}_4] \tag{E12}$$

$$R_{13} = k_8 [Fe(OH)_3^{\alpha}][\sum H_2 S] \tag{E13}$$

$$R_{14} = k_9 [Fe(OH)_2^{\beta}] [\Sigma H_2 S]$$
 (E14)

$$R_{14} = k_{9}[Fe^{(UH)_{3}}][\Sigma H_{2}S]$$

$$R_{15} = k_{10}[Fe^{2+}][\Sigma H_{2}S]$$
(E15)

$$R_{16} = k_{11}[FeS][\Sigma H_2 S] \tag{E16}$$

$$R_{17} = k_{12}[S_0] \tag{E17}$$

$$R_{18} = k_{13}[FeS][S_0]$$
 (E18)

$$R_{19} = k_{14}[SO_4^{2-}][CH_4]$$
 (E19)

$$R_{20} = k_{15} [Fe(OH)_{3}^{\alpha,\beta}] [CH_{4}]$$
 (E20)

$$R_{21} = k_{16}[Fe(OH)_3^{\alpha}] \tag{E21}$$

$$R_{22} = k_{17} [Fe(OH)_3^{\beta}] \tag{E22}$$

$$R_{23} = k_{18}[Fe^{2+}][HPO_4^{2-}]$$
 (E23)

$$R_{24} = k_{19}[Fe^{2+}][DIC] \tag{E24}$$

$$R_{25} = k_{20}[FeCO_3][\Sigma H_2S]$$
 (E25)

$$R_{26} = k_{21}[Fe_3(PO_4)_2][\Sigma H_2S]$$
 (E26)

Table 5. Reaction parameters used in the diagenetic model.

Parameter	Symbol	Value	Units	Values given in literature	
Decay constant for OM^{α}	k_{α}	0.05	yr ⁻¹	0.05-1.62 ^{a,b}	
Decay constant for OM^{β}	\mathbf{k}_{β}	0.0086	yr ⁻¹	0.0086^{b}	
Limiting concentration of O ₂	K_{O2}	0.02	mM	0.001-0.03 ^c	
Limiting concentration of NO ₃	K_{NO3} -	0.004	mM	0.004 - 0.08^{c}	
Limiting concentration of Fe(OH) ₃	$K_{\text{Fe(OH)3}}$	65	μmol g ⁻¹	65-100 ^c	
Limiting concentration of SO ₄ ² -	K_{SO42}	1.6	mM	1.6 ^c	
Attenuation factor for SO_4^{2-} and methanogenesis	Ψ	0.0042	-	0.00157-0.075 ^{b,d}	
Rate constant for reaction E6	\mathbf{k}_1	0.0011	yr ⁻¹		
Rate constant for reaction E7	k_2	10'000	$mM^{-1} yr^{-1}$	5'000-39'000 ^{c,d}	
Rate constant for reaction E8	k_3	140'000	mM ⁻¹ yr ⁻¹	140'000°	
Rate constant for reaction E9	k_4	300	$mM^{-1} yr^{-1}$	$300^{\rm c}$	
Rate constant for reaction E10	\mathbf{k}_{5}	1	$mM^{-1} yr^{-1}$	1 ^c	
Rate constant for reaction E11	\mathbf{k}_{6}	160	$mM^{-1} yr^{-1}$	$\geq 160^{\rm c}$	
Rate constant for reaction E12	\mathbf{k}_7	10'000'000	$mM^{-1} yr^{-1}$	10'000'000°	
Rate constant for reaction E13	\mathbf{k}_{8}	9.5	$mM^{-1} yr^{-1}$	$\leq 100^{\rm c}$	
Rate constant for reaction E14	k_9	0.95	$mM^{-1} yr^{-1}$	Model constrained	
Rate constant for reaction E15	\mathbf{k}_{10}	150	$mM^{-1} yr^{-1}$	100-14'800 ^{b, d}	
Rate constant for reaction E16	\mathbf{k}_{11}	0.0003	$mM^{-1} yr^{-1}$	3.15 ^e	
Rate constant for reaction E17	\mathbf{k}_{12}	3	yr ⁻¹	3^{f}	
Rate constant for reaction E18	\mathbf{k}_{13}	1	$mM^{-1} yr^{-1}$	$7^{\rm f}$	
Rate constant for reaction E19	\mathbf{k}_{14}	0.14	$mM^{-1} yr^{-1}$	10 ^c	
Rate constant for reaction E20	k_{15}	0.00000016	$mM^{-1} yr^{-1}$	0.0074^{g}	
Rate constant for reaction E21	k_{16}	0.6	yr ⁻¹	0.6^{f}	
Rate constant for reaction E22	k ₁₇	0.000013	yr ⁻¹	Model constrained	
Rate constant for reaction E23	\mathbf{k}_{18}	0.052	$mM^{-1} yr^{-1}$	Model constrained	
Rate constant for reaction E24	k_{19}	0.0027	$mM^{-1} yr^{-1}$	Model constrained	
Rate constant for reaction E25	k_{20}	0.0008	$mM^{-1} yr^{-1}$	Model constrained	
Rate constant for reaction <i>E26</i>	k_{21}	0.0008	mM ⁻¹ yr ⁻¹	Model constrained	

^a Moodley et al. (2005); ^b Reed et al. (2011a); ^c Wang and Van Cappellen (1996); ^d Reed et al. (2011b); ^e Rickard and Luther (1997); ^f Berg et al. (2003); ^g Rooze et al. (2016)

Table 6. Depth-integrated rates of key processes for selected depth intervals in µmol m⁻² d⁻¹.

Process	0 – 90 cm ^a	90 - 300 cm ^b	300 – 800 cm ^c	0 – 800 cm
Organoclastic SO ₄ ²⁻ reduction ^d	68.9	5.3	0.003	74.2
CH ₄ production ^{e,f}	10.21	37.7	91.8	139.8
SO ₄ - AOM	9.4	151.6	1.2	162.2
$Fe - AOM^e$	0	0	1.2	1.2
S ₀ disproportionation	0	0	0.9	0.9

^a Marine deposits; ^b limnic sediments around the SMTZ with dissolved sulfide; ^c non-sulfidic limnic deposits; ^d per mol of SO₄²⁻; ^e per mol of CH₄; ^f sum of CH₄ production from organic matter and from DIC (i.e. CO₂)

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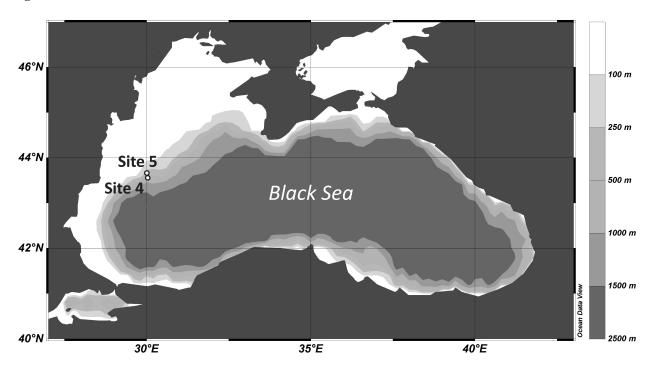


Figure 1. Map showing the locations of site 4 (43°40.6' N, 30°7.5' E; 377 mbss) and site 5 (43°42.6' N, 30°6.1' E; 178 mbss), sampled in June 2013.

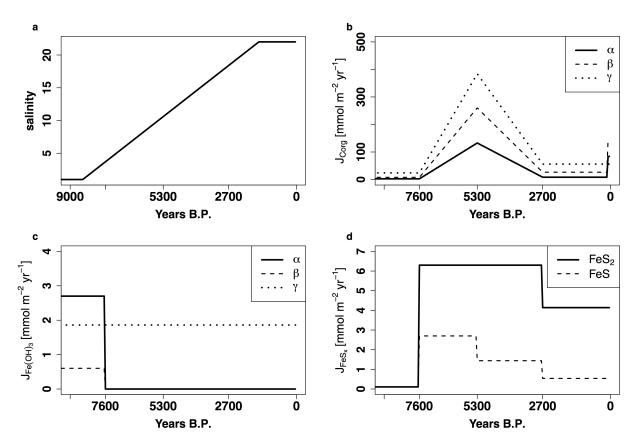


Figure 2. Transient evolution of salinity with a linear increase from 1 to 22 between 8500 and 1500 years B.P. (a), fluxes of organic matter ($J_{C_{0rg}}$; b), Fe oxides ($J_{Fe(OH)_3}$; c) and Fe sulfides (J_{FeS_x} ; d) as implemented in the diagenetic model (site 4).

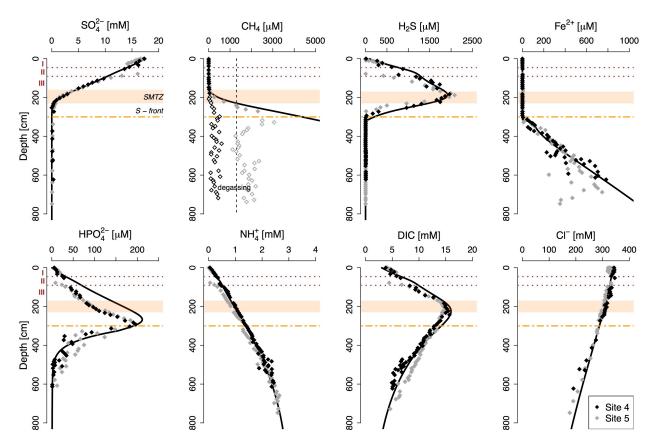


Figure 3. Pore water profiles of key components for site 4 (black diamonds) and site 5 (gray diamonds) and corresponding modeled profiles as calculated with the diagenetic model (black lines). Red dotted lines and roman numbers indicate the transitions between the lithological Unit I (modern coccolith ooze), Unit II (marine sapropel) and Unit III (limnic deposits). The orange bar represents the sulfate-methane transition zone (SMTZ) and the orange dashed line shows the current position of the downward migrating sulfidization front (S-front). The dashed vertical line indicates the CH_4 saturation concentration at atmospheric pressure (Mogollón et al., 2013). The open diamonds indicate CH_4 concentrations that are likely underestimated due to outgassing of CH_4 during coring.

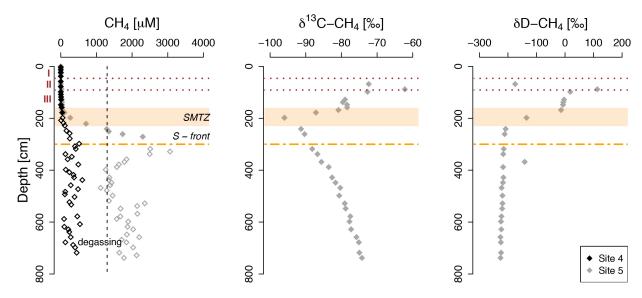


Figure 4. Pore water profiles of CH_4 for site 4 (black diamonds) and 5 (gray diamonds) and corresponding isotopic composition of dissolved CH_4 (available for site 5 only). $\delta^{13}C$ - CH_4 values are given in ‰ vs. VPDB (Vienna Pee Dee Belemnite) and δD - CH_4 values are given in ‰ vs. V-SMOW (Vienna Standard Mean Ocean Water). Red dotted lines and roman numbers indicate the transitions between the lithological Unit I (modern coccolith ooze), Unit II (marine sapropel) and Unit III (limnic deposits). The orange bar represents the sulfate-methane transition zone (SMTZ) and the orange dashed line shows the current position of the downward migrating sulfidization front (S-front). The dashed vertical line indicates the CH_4 saturation concentration at atmospheric pressure (Mogollón et al., 2013). The open diamonds indicate CH_4 concentrations that are likely underestimated due to outgassing of CH_4 during coring.

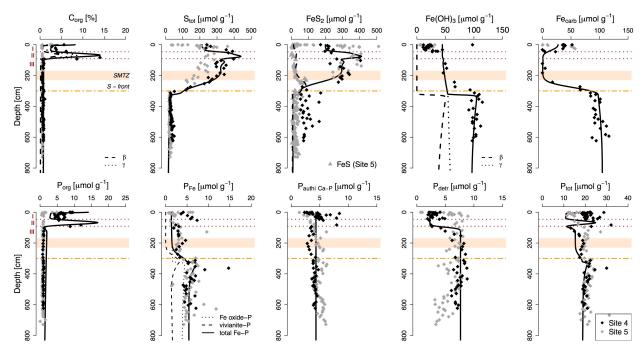


Figure 5. Solid phase sediment profiles for site 4 (black diamonds) and 5 (gray diamonds). Fe oxides represent the sum of amorphous, crystalline and recalcitrant oxides, i.e. Fe_{ox1} , Fe_{Ox2} and Fe_{mag} (Table 1, Supplementary Fig. S4). Fe_{carb} was corrected for apparent AVS dissolution during the Na acetate extraction step (the uncorrected Fe_{carb} data is given in Supplementary Fig. S4). Black lines represent profiles derived from the diagenetic model. Red dotted lines and roman numbers indicate the transitions between the lithological Unit I (modern coccolith ooze), Unit II (marine sapropel) and Unit III (limnic deposits). The orange bar represents the sulfate-methane transition zone (SMTZ) and the orange dashed line shows the current position of the downward migrating sulfidization front (S-front).

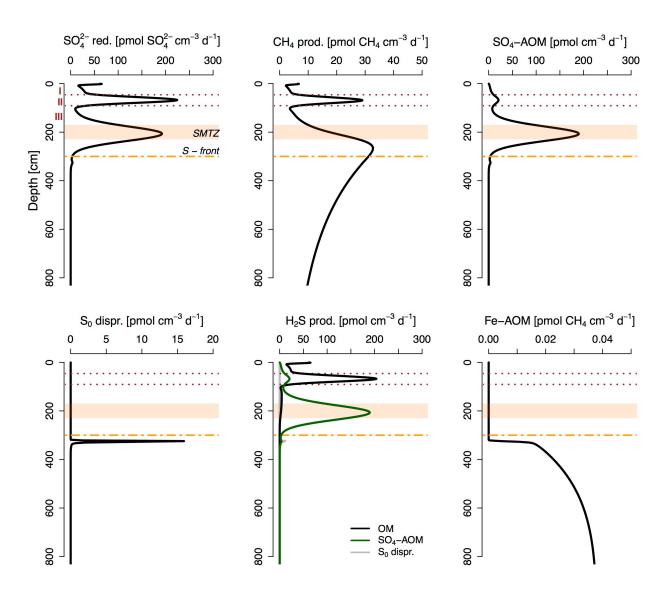


Figure 6. Modeled rates of total SO₄²⁻ reduction, total CH₄ production, SO₄-AOM, S₀ disproportionation, sulfide production and Fe-AOM. Red dotted lines and roman numbers indicate the transitions between the lithological Unit I (modern coccolith ooze), Unit II (marine sapropel) and Unit III (limnic deposits). The orange bar represents the sulfate-methane transition zone (SMTZ) and the orange dashed line shows the current position of the downward migrating sulfidization front (S-front).

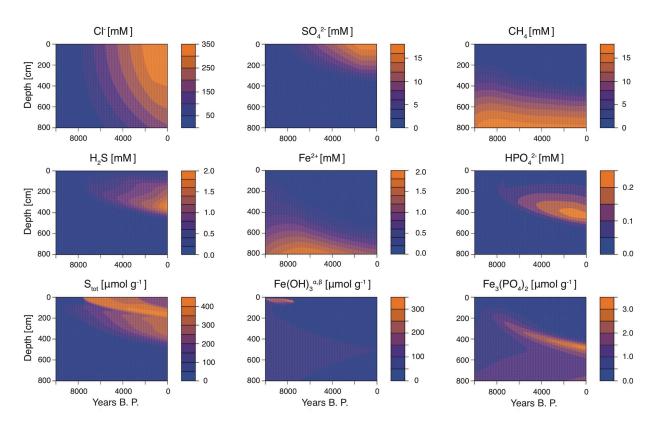


Figure 7. Transient evolution of selected pore water and sediment profiles with depth as calculated for site 4 using the diagenetic model.

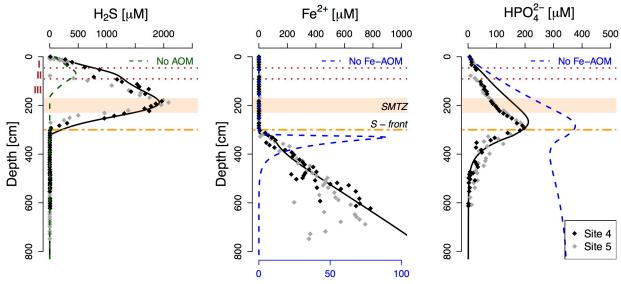


Figure 8. Pore water profiles of dissolved sulfide, Fe²⁺ and HPO₄²⁻. The green dashed line represents the modeled sulfide profile without SO₄-AOM, indicating that the latter significantly enhances the downward sulfidization. Blue dashed lines denote the modeled Fe²⁺ and HPO₄²⁻ profiles without ongoing Fe oxide reduction in the limnic deposits (i.e. no Fe-AOM). Note that concentrations of Fe²⁺ were multiplied 10 times in the model simulation without Fe oxide reduction to better visualize the potential release of Fe²⁺ through a cryptic S cycle (corresponding x axis at bottom). Red dotted lines and roman numbers indicate the transitions between the lithological Unit I (modern coccolith ooze), Unit II (marine sapropel) and Unit III (limnic deposits). The orange bar represents the sulfate-methane transition zone (SMTZ) and the orange dashed line shows the current position of the downward migrating sulfidization front (S-front).

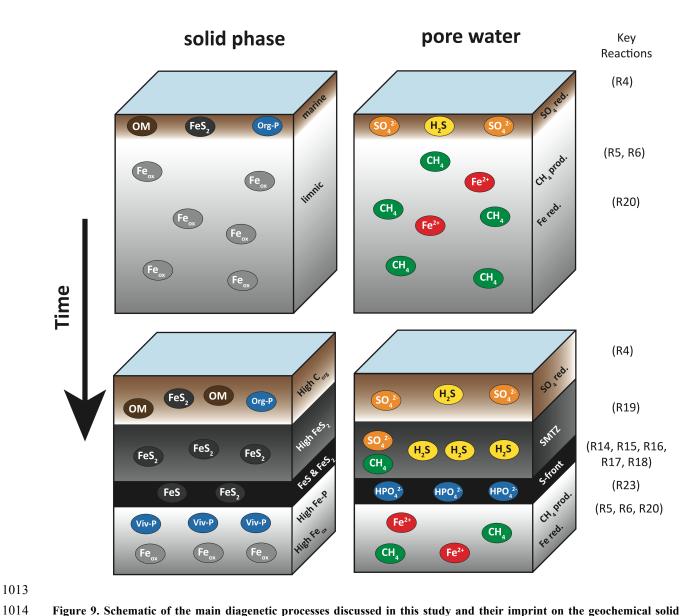


Figure 9. Schematic of the main diagenetic processes discussed in this study and their imprint on the geochemical solid phase (left) and pore water profiles (right). Accumulation of marine sediments with time and the subsequent downward diffusion of SO₄²⁻ into the CH₄-bearing limnic sediment stimulate SO₄-AOM around the sulfate-methane transition zone (SMTZ), thus enhancing the downward sulfidization of the Fe oxide-rich lake deposits. Below the sulfidization front (Sfront), HPO₄²⁻ released during reductive dissolution of Fe oxides is bound again in vivianite, leading to an enrichment in sedimentary P in these sediments. Numbers on the right indicate the key reactions occurring in the corresponding sediment layers as described in Table 3. Note that in this study, Fe-AOM (R20) was assumed as the main source of pore water Fe²⁺ below the S-front to further test the potential impact of Fe-AOM on pore water CH₄. However, based on the geochemical data, we cannot exclude a potential role for organoclastic Fe reduction (R3) and/or reactivation of less reactive Fe oxides by methanogens.