

We would like to thank the editor for his constructive criticism and assistance during the discussion process.

**Associate editor Decision: Publish subject to minor revisions (Editor review)**

Comments on the revised manuscript:

**1) There are still issues regarding gross versus net. In your reply you emphasize that your approach considers gross emissions and you do not want to provide a definition for net emissions: “Since we do not work with net emissions, we prefer to concentrate on defining gross emissions, and skipping other definitions to avoid confusions.” However, in SI section 2 you write about “our net AFOLU balance”. E.g on p. 27 “Our net AFOLU balance included sequestration processes from soils under crops, and from paddy rice, both through organic matter fixation (dSOC)”. This may give the impression that all GHG sources and sinks with respect to the atmosphere are included in your assessment of a “net AFOLU balance”. This is apparently not the case, e.g. neglect of regrowth after fire, of the delay of carbon emissions from product pool or neglect of forest sinks and neglect of area with forest gain in the Harris et al data. Thus, the term “net” appears potentially misleading and should be omitted.**

We thank the editor for spotting this error in the Supplementary Information. We have fully removed these paragraphs and doubled check for further inconsistencies. We originally included dSOC sequestration processes from soils under crops and paddy rice in our analysis of AFOLU hotspots, but the contribution of dSOC is so small that we decided to ease the methodological description and work on gross assessments that did not include removals by forests nor removals by soils.

**2) On a similar note, the caption and headings in table 8 refers simply to “emissions”. It is important that you clarify that these are “gross emissions” in the table and headings.**

Agreed. Reviewed and changed.

**3) Referee 2 asked in his first review for “Also a direct comparison of the CO<sub>2</sub> component with the GCP would enhance the credibility of this study.” Such a comparison with GCP and also with results from the Biogeosciences Special Issue: REgional Carbon Cycle Assessment and Processes (RECCAP) is still missing.**

Please see comment 5. We have intentionally omitted this point because our associated BGS manuscript bg-2016-244: ‘Multi-gas and multi-source comparison of six global land use emission databases and AFOLU estimates in the Fifth Assessment Report’, includes an entire section comparing the differences between AFOLU and carbon assessments..

**4) It is necessary that your findings are described in an overall context in the conclusion section (see my previous comment #5). The manuscript and the supplementary information deal with many different fluxes and complex issues. Thus, it is important for this type of work to put your findings into an overall perspective for the general reader.**

Much more structured conclusions are provided. They include a description of general results and a detailed reflexion of the differences between gross and net assessments -including the suggestions of reviewer 2-, and the crucial importance of methodological assumptions (lines 533-560)

**5) Arguably, the influence on AFOLU on the atmospheric carbon balance appears very relevant. It is thus important to point out that gross emissions as determined in this study are not to be confused with the overall net land-to-atmosphere flux due to human land use. This difference arises as important components of the overall terrestrial and atmospheric carbon balance such as legacy effects and changes in litter and soil organic matter are not included in this work. IPCC AR5 WGI Table 6.3 indicates net emissions to the atmosphere from land use change for Central and South America, Africa and Tropical Asia of about 2.5 PgCO<sub>2</sub>/yr (0.69 PgC/yr ) for the 2000s and (and global net emissions of 1.1 PgC/yr in table 6.1 of WGI AR5). Your table 2 indicates gross emissions of CO<sub>2</sub>**

**of 11 PgCO<sub>2</sub>/yr for a similar period (2000 to 2005). Such differences are relevant and should be spelled out clearly and in a quantitative way in your conclusion section.**

Yes, the editor is raising an interesting point that we have developed in more detail in our associated manuscript: <http://www.biogeosciences-discuss.net/bg-2016-244/> where we compare the differences between the carbon and the AFOLU research communities. We were part of the carbon community meeting held in Paris this year 2016, as well as the land use UCLA meeting held at WRI's headquarters in DC, so both communities are well aware of carbon-AFOLU problems and their implications on the AFOLU results under the IPCC AR5. The original manuscript 2016-244 also included comments on inconsistencies on gross emissions from published data in Baccini et al. 2012, and the Figure 11.8 from WG3, IPCC's AR5. We decided not to include all issues and not to mention them in the hotspots, since they were too large and complex. Please note that Skee Houghton is part of our manuscript <http://www.biogeosciences-discuss.net/bg-2016-244/> as a testimony that these issues are being discussed and far more clarity, transparency and documentation is needed for the AFOLU estimates of AR6.

Some pointed differences between carbon and AFOLU communities, as exposed in manuscript 2016-244, include:

1. the CO<sub>2</sub>-only approach of the carbon community versus the multi-gas approach of the AFOLU community (CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O), particularly important for fire emissions
2. land uses in WG1 do not include agriculture. In our hotspots research we have clarified that 'AFOLU' and 'land uses' are synonyms.
3. Non IPCC compliance of the widely used bookkeeping model of Houghton that leads to large differences in forest net emissions between the bookkeeping model and the reports under the UNFCCC, please see Federici et al. (2016), and certainly between our forest gross emissions, and his net emissions that include more than forests. For IPCC versus bookkeeping model approach differences please see:

S. Federici, G. Grassi, N. Harris, D. Lee, T. Neeff, J. Penman, M.J. Sanz, M. Wolosin (2016). GHG fluxes from forests: An assessment of national reporting and independent science in the context of the Paris Agreement (Working Paper). Retrieved from Climate and Land Use Alliance website: <http://www.climateandlandusealliance.org/reports/ghg-fluxes-from-forests/>

And Grassi and Dentener (2015)

Grassi, G., Dentener, F.: Quantifying the contribution of the Land Use Sector to the Paris Climate Agreement. The LULUCF sector within the Intended Nationally Determined Contributions. EUR 27561. JRC Science for Policy Report. Ispra, Italy, 2015. Available at: <http://publications.jrc.ec.europa.eu/repository/bitstream/JRC98451/jrc%20lulucf-indc%20report.pdf>

4. Due to the above mentioned points, the IPCC AR5 has issues in Figure 11.2 and Tables 11.1 and in the general AFOLU budget.

For all these reasons, the comparison that the editor suggests will be omitted. We originally considered contrasting our 8.0 PgCO<sub>2</sub>e.yr<sup>-1</sup> gross AFOLU emissions (2000-2005) against the ca. 10 PgCO<sub>2</sub>e.yr<sup>-1</sup> net AFOLU emission from Latin America, Middle-East Africa and Asia in 2010 (or the >10 PgCO<sub>2</sub>e.yr<sup>-1</sup> in 2000-2005) as reported in Figure 11.2 which is not carbon based for FOLU, but CO<sub>2</sub>e based for AFOLU. But this comparison would need far too many explanations,

since the net budget is higher than our gross budget, partly because the spatial area is probably not the same (China is most likely considered in Asia in Figure 11.2)

**6) The term “gross” should be included in the title.**

Done

**7) SI: section 3, p33 typo: “Table 6 isplays“**

Thanks, corrected

**8) SI p33 : “Table 8 shows a brief overview of the included variables” Please check table number (I think it should be table 7)**

Thank you. We have corrected it. Table and figure numbers have been reviewed, and page indexation updated.

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## Responses to Reviewer 2\_version2

We would like to thank Dr. Benjamin Stockers for his constructive criticism on our manuscript and his useful comments. They have clearly benefitted the clarity of our manuscript.

1. **In my first review, I argued that the authors make inconsistent use of data representing in some instances net emissions and gross emissions in others. This argument was based on my own misunderstanding of the Gross Forest Cover Loss method by Hansen et al., 2010, underlying the deforestation data by Harris et al. (2012) and used in the present study. Therefore, I would like to suggest to regard my comments 4,5,8 and 9 as obsolete. I duly apologize for the unnecessary rebuttal concerning these issues. In that sense, I consider the author’s choice for using gross emissions from different sources to be consistent.**

Thanks. Clarifying gross vs net emissions was important and useful. Misunderstandings are now solved and together with Reviewer 3 comments, and the extra editor suggestions, we hope the topic is now better framed and easier to understand by the readers.

2. **The method used here is generally appropriate for providing spatial data on where largest C losses occur without accounting for subsequent recovery of C stocks. However, I would like to flag that neglecting recovery is an important point especially for the wood harvesting and fire emissions components and that there is an issue that accounting only for gross emissions may lead to implausible conclusions. For example, areas with a high but sustainable wood harvesting regime may appear as “emission hotspots” although the use of (sustainably harvested) fuel wood may replace fossil fuels and thereby contribute to climate mitigation. I’m raising this as a conceptual issue and encourage authors to argue that in reality this is not greatly affecting the conclusions drawn here. In a similar sense, areas where fire is a natural component of ecosystem functioning and is \*not\* leading to a long-term decline in C storage, may unwillingly appear as “emission hotspots”. My concern is that this could be the case in Africa, where fire emissions included here are a quantitatively important component of total emissions. I recommend to include a discussion of these points.**

The reviewer’s optic is still under net assessments, and under a longer term perspective. Thus, under a gross assessment, such as ours, there is no issue whatsoever to affirm that sustainable wood harvesting and fires not leading to long-term declines in C storage are hotspots of emissions. It just needs to be clear that they are hotspots of gross emissions. Moreover, even if we had included sinks in this study, the time period under consideration (2000-2005) would not have compensated for the carbon losses of

sustainable harvesting or sustainable fires. Therefore, the problem is not only a debate between gross and net, but also on time scales, and on the assumptions that affect the emission balance: legacy, instantaneous emissions after disturbance, transboundary effects, and assumptions that affect mitigation balances such as substitution effects.

We have included a detailed description of these points in the conclusions, in lines 533-560.

**3. I appreciate the clarifications offered in the response regarding double counting of peat fires and wood harvesting/deforestation. Could some of this more detailed information be included in the main text?**

We have included some of our more detailed explanations on peat fire-deforestation double counting in lines 175-177, and more contextual information on wood harvesting in lines 188-192.

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### Responses to reviewer 3

We would like to thank Dr. Houghton for his useful comments and for his contribution to help clarify the definitions of gross *versus* net emissions.

***This paper pulls together several different spatial data sets that help map the gross emissions of greenhouse gases (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) from tropical agriculture and use/management of tropical forests (tropical AFOLU). One objective is to show the current hotspots of gross emissions.***

**1. *The authors include emissions from fires in their analyses, arguing that 90% of tropical [forest] fires are the result of human activity. There are clearly interactions between direct human effects and indirect (e.g., climate) effects. In SE Asia, for example, the most effective burning, in terms of clearing forests, is during dry periods associated with El Nino. The fires are intentional, but they are enhanced by El Nino events. My suspicion is the fires are becoming more common in tropical forests, in part, because of the opening of forests as a result of human activity (e.g., agriculture, harvests) and, in part, from increasing frequency and intensity of droughts.***

Yes, we fully agree with the reviewer that there is a climatic component in the enhanced fire emissions of anomalous years (ENSO and AMO episodes). We have clarified this in lines 298-302. We disagree, though, that climate-driven fires should be considered wildfires without human influence and, therefore, left unassessed in emission inventories.

Two comments to sustain our inclusion:

1. *Degradation versus deforestation fires*: it is important to include both deforestation and degradation fires for a complete gross emission assessment. We understand that this decision is much more complicated in net fire assessments, because any emission coming from a degradation fire should then include its removals through recovery. Lack of data on post-disturbance carbon recovery is behind the general assumption of carbon neutrality of degradation fires in the tropics, by the carbon community, and the reason to exclude most degradation fires. Under a gross assessment, this differs.

The assumption of carbon neutrality of degradation fires, and positioning about it is debated in detail in our associated manuscript <http://www.biogeosciences-discuss.net/bg-2016-244/>

2. *Fire and force majeure*: it is important to fully assess and report gross emissions from all fire events, independently of their climatic circumstances, as exposed in the IPCC (2003) guidance:

- ✓ Under the IPCC, countries must report all emissions, even those occurring under *force majeure*, but can then argue for a reduction under their accounting goals (see Federici et al. (2016)'s report at: [http://www.climateandlandusealliance.org/wp-content/uploads/2016/06/GHG\\_Fluxes\\_From\\_Forests\\_Working\\_Paper.pdf](http://www.climateandlandusealliance.org/wp-content/uploads/2016/06/GHG_Fluxes_From_Forests_Working_Paper.pdf))

**Completeness of reporting:** (...)The impact of natural disturbances often averages out across long time periods. Annually they may cause large variations in the anthropogenic GHG balance of a country. IPCC guidelines are that countries include emissions and subsequent removals associated with natural disturbances on managed land in their national GHGI reporting. However, when assessing performance relative to a national target (accounting), countries may wish under certain circumstances to exclude a portion of emissions and removals associated with disturbances on the basis that the magnitude of disturbance events may overcome the capacity of humans to take them under control and limit their impact. The Paris Agreement has not yet established guidance on how countries may account for natural disturbances in the context of NDCs although rules have been established under the Kyoto Protocol.

**Source:** Federici et al. (2016)

S. Federici, G. Grassi, N. Harris, D. Lee, T. Neeff, J. Penman, M.J. Sanz, M. Wolosin (2016). GHG fluxes from forests: An assessment of national reporting and independent science in the context of the Paris Agreement. Working Paper. Climate and Land Use Alliance. San Francisco. US. Available at:<http://www.climateandlandusealliance.org/reports/ghg-fluxes-from-forests/>

**2. *The definitions of net and gross emissions are important, and I offer some suggestions for clarifying the wording on four pages of the introduction and methods.***

The reviewer's comments have all been included in our final manuscript, with the exception of his suggestion to substitute 'indirect effects' by 'lateral fluxes', which we preferred to keep as it was, to preserve coherence with Ogle et al. data, USEPA (2013)'s report, and the IPCC WG3 Chapter 11 documents.

1 | **Hotspots of tropical land use emissions: a gross assessment of patterns, uncertainties, and**  
2 | **leading emission sources for the period 2000-2005**

3  
4 | **Short title:** AFOLU greenhouse gas emissions hotspots

5  
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31 **Keywords:** AFOLU, mitigation, greenhouse gases, CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, Land use emissions, tropics.

32

### 33 **Abstract**

34 According to the latest report of the Intergovernmental Panel on Climate Change (IPCC),  
35 emissions must be cut by 41-72% below 2010 levels by 2050 for a likely chance of containing the  
36 global mean temperature increase to 2 °C. The AFOLU sector (Agriculture, Forestry and Other  
37 Land Use) roughly contributes with a quarter (~ 10 -12 PgCO<sub>2</sub>e.yr<sup>-1</sup>) of the net anthropogenic GHG  
38 emissions mainly from deforestation, fire, wood harvesting, and agricultural emissions including  
39 croplands, paddy rice and livestock. In spite of the importance of this sector, it is unclear where are  
40 the regions with hotspots of AFOLU emissions, and how uncertain these emissions are. Here we  
41 present a novel spatially comparable dataset containing annual mean estimates of gross AFOLU  
42 emissions (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O), associated uncertainties, and leading emission sources, in a spatially  
43 disaggregated manner (0.5°), for the tropics, for the period 2000-2005. Our data highlight: i) the  
44 existence of AFOLU emissions hotspots on all continents, with particular importance of evergreen  
45 rainforest deforestation in Central and South America, fire in dry forests in Africa, and both  
46 peatland emissions and agriculture in Asia; ii) a predominant contribution of forests and CO<sub>2</sub> to the  
47 total AFOLU emissions (69%) and to their uncertainties (98%), iii) higher gross fluxes from forests  
48 coincide with higher uncertainties, making agricultural hotspots appealing for effective mitigation  
49 action, and iv) a lower contribution of non-CO<sub>2</sub> agricultural emissions to the total gross emissions  
50 (ca. 25%) with livestock (15.5%) and rice (7%) leading the emissions. Gross AFOLU tropical  
51 emissions 8.0 (5.5-12.2) were in the range of other databases 8.4 and 8.0 PgCO<sub>2</sub>e.yr<sup>-1</sup> (FAOSTAT  
52 and EDGAR respectively), but we offer a spatially detailed benchmark for monitoring progress on

53 reducing emissions from the land sector in the tropics. The location of the AFOLU hotspots of  
54 emissions and data on their associated uncertainties, will assist national policy makers, investors  
55 and other decision-makers who seek to understand the mitigation potential of the AFOLU sector.

56

## 57 **1. INTRODUCTION**

58 Currently unabated CO<sub>2</sub>e emissions need effective mitigation action (UNEP, 2015). Emissions  
59 modelling suggests that to maintain the global mean temperature increase on track with the 2°C  
60 target and to remain close to the 450 ppm of CO<sub>2</sub>e by 2100, global greenhouse gas (GHG)  
61 emissions must be cut in a range of 41-72% below the 2010 levels by 2050, and global emissions  
62 levels must be reduced to zero (a balance between sources and sinks) by 2070 and below zero  
63 through removal processes after that (IPCC, 2014; Anderson, 2015; UNEP 2015). To reach these  
64 ambitious goals, it is imperative to identify regions where the mitigation of key emission sectors  
65 may be most promising in terms of reducing fluxes, reducing emission trends, and/or maximizing  
66 returns on mitigation investments. From all the sectors contributing to the total anthropogenic GHG  
67 emissions, the Agriculture, Forestry and Other Land Use (AFOLU) sector participates with roughly  
68 one quarter (10-12 PgCO<sub>2</sub>e.yr<sup>-1</sup>) of the total emissions (49 PgCO<sub>2</sub>e.yr<sup>-1</sup>) (IPCC, 2014). Optimistic  
69 estimates suggest that the AFOLU sector -here used as synonym of land use sector- could  
70 contribute 20 to 60% of the total cumulative abatement to 2030 through land-related mitigation  
71 including bioenergy (Smith et al., 2014). However, it is unclear where are the regions with the  
72 largest AFOLU emissions (hotspots of emissions), and how large their associated uncertainties  
73 are.

74

75 Modelling efforts by the carbon community have long offered useful data but their focus is rather  
76 global and CO<sub>2</sub>-oriented, which omits other land use gases such as CH<sub>4</sub> and N<sub>2</sub>O (Schulze et al.,  
77 2009; Houghton et al., 2012; LeQuéré et al., 2012; Canadell et al., 2014; Tian et al., 2016).  
78 Currently, the most used AFOLU data belong to two global multi-gas (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) CO<sub>2</sub>e  
79 databases: FAOSTAT and EDGAR) (Smith et al., 2014; Tubiello et al., 2015). While they offer very  
80 valuable data, they suffer from several shortcomings: they do not provide uncertainties or  
81 uncertainties are not provided at the spatial scale at which emissions are offered; they suffer from



82 untransparent documentation (e.g. EDGAR) or data are offered at inappropriate spatial scales to  
83 effectively navigate mitigation implementation (e.g. country level in FAOSTAT). Thus, unlike  
84 aggregated estimates, spatially explicit data favour targeted mitigation action and implementation  
85 by identifying where are the areas within a country that hold the largest emissions, and what are  
86 the key emission sources to address in these areas (e.g. deforestation, degradation, livestock,  
87 cropland soils, paddy rice). Spatially explicit assessments of AFOLU emissions and their  
88 associated uncertainties would assist national policy makers, investors and other decision-makers  
89 who seek to understand the mitigation potential of the AFOLU sector, and which areas to prioritize  
90 This potential is here defined as the maximum mitigation reduction that could be achieved without  
91 technical or economic considerations. Better understanding of the AFOLU mitigation potentials will  
92 also be important under the Paris Agreement (PA) since the fulfilment of the 2°C target is  
93 dependent on the mitigation ambition presented by countries in their Nationally Determined  
94 Contributions (NDCs). To safeguard this ambition a stock-take process has been defined, by  
95 which countries are required to update their NDCs every five years, starting from 2020, and to  
96 enhance their mitigation commitments from previous submissions (Bodle et al., 2016). It is  
97 therefore imperative to improve our understanding of where and how much could countries  
98 enhance their AFOLU ambition from what they have currently reported.

99

100 Mitigation action can be directed to reducing emissions by the sources, or to increasing the  
101 absorptions by the sinks, or to both. While gross and net emissions are equally important, they  
102 offer different information (Richter and Houghton, 2011; Houghton et al., 2012). Net land use  
103 emissions represent the sum of emissions by sources and removals by sinks. Land use sinks refer  
104 to any process that stores GHGs (e.g. forest growth, forest regrowth after disturbances, organic  
105 matter stored in soils, etc) (see Richter and Houghton, 2011, for further details). Countries report  
106 their emissions and their reduction targets based on net AFOLU balances (IPCC, 2006; Iversen et  
107 al., 2014; Smith et al., 2014). Gross assessments offer separate data on emissions by sources  
108 (gross emissions) and removals by sinks (gross removals), and are useful for designing mitigation  
109 implementation because they offer direct information on the sources and sinks that may be acted  
110 upon through policies and measures to enhance and promote mitigation. However, lack of ground

111 data makes the assessment of gross sinks much more difficult than the assessment of gross  
112 sources (Lewis et al., 2009; Houghton et al., 2012; Grace et al., 2014; Brienen et al., 2015) with a  
113 particular gap on disturbed standing forests (Poorter et al., 2016).

114

115 For these reasons, we present here an assessment of AFOLU gross emissions only, for the tropics  
116 and subtropics. We exclude sinks (e.g. regrowth of cleared forests or burned areas, and soil  
117 carbon storage). We offer spatially explicit (0.5°) multi-gas (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) CO<sub>2</sub>e gross emission  
118 data that help identifying the hotspots of land use emissions in the tropics and subtropics, and  
119 associated uncertainties for 2000-2005. Our method uses a consistent approach to overcome  
120 problems of different definitions, methods, and input data present in other approaches (e.g.  
121 nationally reported data), allowing data comparability. It is a top-down approach based on  
122 published spatially explicit GHG datasets for the key sources of emissions in the AFOLU sector as  
123 identified in the Fifth Assessment Report of the IPCC (AR5) (Smith et al. 2014): deforestation, fire,  
124 wood harvesting, crop soil emissions, paddy rice emission, enteric fermentation and manure  
125 management. We address three questions at the landscape, tropical, and continental scales: 1.  
126 Where are the hotspots of tropical AFOLU emissions and how uncertain are they? 2. What are the  
127 main GHGs emissions behind these hotspots?, 3. What are the emission sources behind these  
128 hotspots? 4. How do our gross AFOLU emissions relate to other AFOLU datasets such as  
129 FAOSTAT or EDGAR?

130

## 131 **2. MATERIAL AND METHODS**

132 Our study area covers the tropics and the subtropics, including the more temperate regions of  
133 South America (33° N to 54° S, 161° E to 117° W). It expands over a diversity of ecosystems that  
134 range from dry woodlands and dry forests such as the African Miombo and South American  
135 Chaco, to rainforests and moist forests such as evergreen broadleaved rainforests or montane  
136 cloud forests. The years considered by our datasets varied, yet we selected the period 2000-2005  
137 as the common temporal range for all the datasets. The exception was the rice emissions dataset,  
138 that took 2010 as its baseline (See Table S2 in supplementary). This time period represents a  
139 useful historical baseline against which countries can contrast the evolution of their AFOLU gross

140 emission performances. We consider the pixel size ( $0.5^\circ$ ) appropriate for landscape research, and  
141 useful to visualize emissions hotspots. More detailed information about each data source and a  
142 descriptive summary is available in the SOM (Table S2).

143

## 144 **2.1 Datasets**

145 Deforestation (Harris et al., 2012): Deforestation refers to gross emissions, associated to the area  
146 of forest cover loss (above ground and below ground biomass) that is removed due to human or  
147 natural disturbances, at 18.5 km of spatial resolution and aggregated in a 5-year period (2000-  
148 2005). Deforestation areas are based on MODIS data at 18.5 km resolution, while carbon loss  
149 derives from Saatchi et al., (2012) carbon map, at 1 km resolution. The disparate spatial resolution  
150 of these two maps is solved by a randomization procedure (Harris et al. 2012). Information of  
151 uncertainties is expressed as 5<sup>th</sup> and 95<sup>th</sup> percentiles, estimated through Monte Carlo simulations  
152 and showed non-Gaussian distributions. Harris et al data defines the spatial and temporal extent of  
153 our tropical AFOLU analysis.

154

155 Fire (Van der Werf et al., 2010): Fire emissions ( $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$ ) were obtained from the Global  
156 Fire Emission Database (GFED) (www1) at  $0.5^\circ$  resolution, based on the CASA model which  
157 includes four carbon pools (above and below ground biomass, litter and coarse woody debris).  
158 Only carbon from organic soils was included. Original data were of global coverage for the period  
159 1997-2013. We extracted a subset for the tropics and 2000-2005. Annual uncertainties for different  
160 regions are expressed in Van der Werf et al. (2010) as the 5th, 25th, 50th, 75th, and 95th  
161 percentiles of 2000 Monte Carlo runs.  $1\sigma$  uncertainties (expressed as percentage of the 50th  
162 percentile) were also given, and considered Gaussian distributions. To move to pixel ( $0.5^\circ$ )  
163 uncertainties, we assigned the regional  $1\sigma$  to all the pixels within each region, for each gas. Total  
164 fire emissions ( $\text{CO}_2\text{e}$ ) per pixel were the sum of the annual means. The uncertainties of the  
165 different gases ( $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{CO}_2$ ) were assumed independent and estimated by square rooting  
166 the sum of their variances. Fire emissions are partitioned into six classes (savannah, agriculture,  
167 woodlands, forests, peatlands and deforestation), which helped us remove  $\text{CO}_2$  emissions from  
168 savannahs and agriculture since the burning of these non-woody land uses is assumed carbon

169 neutral (e.g. biomass burned gets recovered by biomass growth in the next growing season)  
170 (IPCC, 2006). CH<sub>4</sub> and N<sub>2</sub>O emissions were, however, retained. We also removed deforestation  
171 fires, to avoid double counting with deforestation emissions from Harris *et al.* (2012). Some  
172 overlapping of deforestation and soil peat burning might however occur where peat fires and  
173 deforestation fires show similar fire recurrences and might be wrongly labelled (Van der Werf *et al.*,  
174 2010). Some peat fires might, therefore, respond to deforestation fires and cause some double  
175 counting with Harris deforestation emissions. This would only affect Indonesia since it is the only  
176 country that counts on spatially explicit peatland maps (Van der Werf et al. 2010), and would  
177 therefore represent a small bias.

178

179 Wood harvesting (Poulter et al., 2015): Wood harvesting is a 1° global gridded data set, generated  
180 in the frame of the GEOCARBON project. It uses National Forest Inventory data and the FAO  
181 Forest Resources Assessment (FRA). Aboveground biomass data were downscaled using a forest  
182 mask from the Global Land Cover (GLC) 2000, and assuming that wood harvest was distributed  
183 evenly. The original data were produced at the resolution of the GLC2000 (approx. 1X1 km) and  
184 finally aggregated to the 1° scale. Wood Harvesting data consisted of five layers: 1. Round wood  
185 Forest Area in hectares for each cell, 2. Fuelwood forest area in hectares for each cell, 3. Round  
186 wood (industrial) harvest volume in m<sup>3</sup>, 4. Fuelwood harvest volume in m<sup>3</sup>, 5. Total harvest volume  
187 (round wood + fuelwood) in m<sup>3</sup>. We chose fuel and industrial round wood harvest (m<sup>3</sup>) as our  
188 harvest data. Wood harvest is a gross flux since no regrowth is considered. We assumed  
189 instantaneous emissions assigned to the place of removal, without considering lags in decay, nor  
190 the fate of the harvested product (i.e., slash, paper, furniture, construction), ~~nor possible~~  
191 ~~substitution effects (e.g. energy production using wood biomass instead of fossil fuels).~~ nor the  
192 possible substitution effects (e.g. energy production using wood biomass instead of fossil fuels).

193 We therefore acknowledge that the instantaneous flux from wood harvest would be lower if these  
194 effects had been considered. Emissions were transformed from m<sup>3</sup> to MgCO<sub>2</sub>.yr<sup>-1</sup> using an  
195 emission factor of 0.25 (Mg C/m<sup>3</sup>) (Grace et al., 2014), and a C to CO<sub>2</sub> factor shown in Table 1.  
196 The resolution of this layer was larger than our grid so wood estimates were equally distributed  
197 among our 0.5° grid cells. Because wood harvesting relied on official data reported by countries to

198 FAO, the authors assumed that harvesting emissions only derive from forests remaining forests  
199 (legal logging), and assigned these emissions to forested areas only. Figure S3 in SOM shows  
200 different spatial locations for deforestation and wood harvesting emissions. However, this  
201 assumption might be wrong and some unprecise amount of double counting may occur.  
202 Uncertainties were not estimated in the original harvest emission data. Therefore, and based on  
203 the authors' expert opinion, we chose a 20 percent uncertainty value, per pixel.

204

205 Cropland soils (USEPA 2013): Cropland emissions ( $\text{N}_2\text{O}$  and soil dSOC) (changes in soil organic  
206 carbon) were produced by Ogle et al., for the Environmental Protection Agency MAC-Report  
207 (USEPA, 2013), at  $0.5^\circ$  resolution, for time periods 2000-2030 with five-year increments, based on  
208 the DAYCENT ecosystem model (Ogle et al., 2007). For our AFOLU analysis we used the annual  
209 mean emission data for the period 2000-2005. The original units ( $\text{g N}_2\text{O-N.m}^{-2}.\text{y}^{-1}$  and  $\text{gC.m}^{-2}.\text{5y}^{-1}$ )  
210 were transformed to  $\text{CO}_2\text{e.y}^{-1}.\text{grid cell}^{-1}$  (Table 1). The original dataset included direct and indirect  
211 emissions from mineral-based cropland soil processes: synthetic and organic fertilization, residue  
212 N, mineralization and fixation). To be consistent with other data sets we did not include indirect  
213 emissions (e.g.  $\text{NO}_3^-$  leaching, N runoff in overland water flow). Emissions estimated by the  
214 DAYCENT include soil and litter pools and modelled six major crop types only (maize, wheat,  
215 barley, sorghum, soybean and millet) excluding other important tropical crops (sugar, coffee,  
216 cacao, cotton, tobacco, etc). As a result, the cropland area simulated by DAYCENT was about  
217 61% of the global non-rice cropland areas reported by FAOSTAT, which resulted in lower cropland  
218 emissions when compared to other databases (e.g. FAOSTAT and EDGAR). Moreover, due to the  
219 known poor performance of the DAYCENT model over organic soils, cropland emissions over  
220 drained histosols were not part of the estimated emissions. Uncertainties were offered per pixel  
221 ( $0.5^\circ$ ) as standard deviations per dSOC and  $\text{N}_2\text{O}$  separately. Final  $\text{CO}_2\text{e}$  uncertainties per pixel  
222 were propagated as independent data using the squared root of the summed variances. To  
223 complement the emission gap from the organic cultivated soils, we used a Tier 1 approach that  
224 relied on the location of the tropical areas of histosols (ISRIC's global soil database), the location of  
225 cropland areas per crop types (Monfreda et al., 2008) and a Tier 1 annual emission factor for  
226 cultivated organic soils ( $20 \text{ MgC.ha}^{-1} \text{ yr}^{-1}$ ) derived from the IPCC (IPCC 2006) (Supplementary).

227

228 Paddy Rice (USEPA 2013): We used data by Li et al., from the USEPA's MAC Report (2013).  
229 Emissions were estimated by the Denitrification-Decomposition (DNDC) model, which simulates  
230 production, crop yields, greenhouse gas fluxes (CH<sub>4</sub>, N<sub>2</sub>O) and organic soil carbon (dSOC) of  
231 global paddy rice, at 0.5° resolution under "business-as-usual" (BAU) condition and various  
232 mitigation strategies as explained in Li et al., (2001, 2006). This model includes soil, litter, above  
233 and below ground biomass as main carbon pools. Model outputs were reported for 2010 as the  
234 baseline, and used 22 years of replications to account for climate variability. The original units  
235 (KgC.ha<sup>-1</sup>.yr<sup>-1</sup> for dSOC and CH<sub>4</sub> and KgN. ha<sup>-1</sup>.yr<sup>-1</sup> for N<sub>2</sub>O) were re-projected to equal-area  
236 values, and transformed to CO<sub>2</sub>e (Table 1). Emissions were estimated using the MSF (Most  
237 Sensitive Factor) method which relies on an envelope approach and estimates maximum and  
238 minimum emissions based on extreme soil properties. No mean values were offered. The  
239 distribution of the data were known to be right skewed, and through the authors' expert judgement  
240 a log-normal approach was considered to be the best –although not perfect- fit, from where to  
241 estimate the mean (50<sup>th</sup> percentile), max and min (10<sup>th</sup> and 90<sup>th</sup> percentile) for each cell.

242

243 Livestock (Herrero et al., 2013): Livestock emission data includes enteric fermentation (CH<sub>4</sub>) and  
244 manure management (N<sub>2</sub>O, CH<sub>4</sub>) for the year 2000, for twenty-eight regions, eight livestock  
245 production systems, four animal species (cattle, small ruminants, pigs, and poultry), and three  
246 livestock products (milk, meat, and eggs), at 0.1°cell resolution. The CO<sub>2</sub>e of enteric fermentation  
247 and manure management were then summed to obtain a total emission value of livestock per grid  
248 cell. Since no spatially explicit uncertainty data were provided, and based on the authors' expert  
249 judgement, we applied a 20% value for livestock emissions per cell and per gas. Per cell livestock  
250 GHG uncertainties were estimated by square rooting the sum of their variances.

251

## 252 **Other AFOLU databases**

253 FAOSTAT database: covers agriculture, forestry and other land uses and their associated  
254 emissions of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O, following IPCC 2006 Guidelines at Tier 1 (Tubiello et al., 2014).  
255 Emissions are estimated for nearly 200 countries, for the reference period 1961–2012 (agriculture)

256 and 1990–2012 (FOLU), based on activity data submitted to and collated by FAO (www1).  
257 FAOSTAT includes estimates of emissions from biomass fires, peatland drainage and fires, based  
258 on geo-spatial information, as well as on forest carbon stock changes (both emissions and  
259 absorptions) based on national-level FAO Forest Resources Assessment data (FRA, 2010). FOLU  
260 carbon balances in FAOSTAT are emissions from afforestation, reforestation, degradation,  
261 regrowth, and harvest activities. The FAOSTAT emission estimates are based on annual FAO  
262 emissions updates for AFOLU (Tubiello et al., 2014).

263

264 *EDGAR database*: The Emissions Database for Global Atmospheric Research (EDGAR) provides  
265 global GHG emissions from multiple gases (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) at 0.1° and country levels. It covers all  
266 IPCC sectors (energy, industry, waste management, and AFOLU), mainly applying IPCC 2006  
267 guidelines for emission estimations (EDGAR, 2012). We selected EDGAR's 4.2 Fast Track 2010  
268 (FT 2010) data (www2). Emissions cover the period 2000-2010 in an annual basis, at the country  
269 level and are offered as Gg of gas. No uncertainties are provided. Transformation to CO<sub>2</sub>e used  
270 AR4 100year-Global Warming Potential values to be consistent with other datasets. Metadata can  
271 be found at EDGAR (2012), although further transparency and more complete documentation are  
272 required for this database.

273

## 274 **2.2 Methods**

### 275 *Hotspots dataset*

276 Our AFOLU assessment is based on several assumptions: we focus on human-induced gross  
277 emissions only, excluding sinks. We exclude emissions and sinks from unmanaged land (e.g. CH<sub>4</sub>  
278 or N<sub>2</sub>O emissions from unmanaged natural wetlands). We focused on direct gross emissions  
279 excluding indirect emissions whenever possible (e.g. indirect emissions from nitrate leaching and  
280 surface runoff from croplands). Delayed fluxes (legacies) are important (e.g. underestimations of  
281 up to 62% of the total emissions when recent legacy fluxes are excluded) (Houghton et al., 2012)  
282 but are frequently omitted in GHG analyses that derive from remote sensing, such as our  
283 deforestation emissions from Harris et al., (2012). Wood harvesting emissions also excluded  
284 legacy fluxes. Therefore, no forest regrowth of cleared, burned, or disturbed forests are included in

285 our AFOLU 2000-2005 assessment. Other important components of the overall terrestrial and  
286 carbon balance such as changes in litter, coarse woody debris and soil carbon, are also not part of  
287 the emissions from deforestation and wood harvesting, since these pools were not considered in  
288 the original datasets (see Table S2, SOM). For the other land uses, fire, agricultural soils, and  
289 paddy rice, their emission models (e.g. CASA, DAYCENT and DCDN) included temporal spin-ups  
290 to guarantee the stability of the emissions for their temporal scales under analysis. Certain legacies  
291 have, therefore, been considered (please see references for further understanding of these  
292 models). In the case of fires, since 90 percent of tropical fires are the result of human activity  
293 (Roman-Cuesta et al., 2003; Van der Werf et al., 2010), we assumed all emissions to be human-  
294 induced, independently of whether they are climate-driven. This might have resulted in some  
295 overestimation of fire emissions in drier ecosystems where lightning may start the fires (e.g.  
296 African woodlands). However, since we have excluded deforestation fires (to avoid double  
297 counting with deforestation), and we have also excluded savanna and agricultural fires (under the  
298 assumption of carbon neutrality), we are quite certain that our gross fire emissions for 2000-2005  
299 are rather conservative. As requested by the IPCC (Federici et al., 2016), we have included all  
300 biomass burning emissions without considerations of climate extremes, even though we  
301 acknowledge the role of increased frequencies and intensities of droughts, and their interaction  
302 with fire in human-disturbed landscapes in the tropics (Brando et al., 2014). We assumed  
303 instantaneous emissions of all carbon that is lost from the land after human action (Tier 1, IPCC  
304 2006) (e.g. fire, deforestation and wood harvesting), with no transboundary considerations (e.g. the  
305 emissions are assigned wherever the disturbance takes place, particularly important for the  
306 Harvested Wood Products). Life-cycle substitution effects are neither considered for harvested  
307 wood (Peters et al., 2012).

308

309 Figure 1 describes the steps followed to produce our spatially explicit layers of gross AFOLU  
310 emissions and uncertainties. We first assessed all possible emissions, land uses and human  
311 activities under the framework of the IPCC 2006 AFOLU guidelines. We then selected the key  
312 AFOLU emissions sources as identified in the IPCC Fifth Assessment Report (AR5) (Smith et al.,  
313 2014). There were seven key emission sources, three within the forest sector: deforestation, fire,



314 and wood harvesting (these last two were considered as forest degradation), and four within  
315 agriculture: cropland soils, paddy rice, enteric fermentation and manure management (aggregated  
316 as livestock). We chose 100-year global warming potentials as provided in the Fourth Assessment  
317 Report (AR4) (IPCC, 2007) (Table 1) because all emission datasets were prior to the launching of  
318 AR5. We have preserved their choice to be consistent with their published estimates and with  
319 emissions that could not be reproduced. To promote the spatial assessment we produced an  
320 empty grid with cells of  $0.5^{\circ} \times 0.5^{\circ}$  in a World Geographical reference System (WGS-84, lat-lon). To  
321 correct for the unaccounted Earth distortions that come with a geographical system we used equal  
322 area re-projected values whenever we needed area-weighted estimates of the emissions. This grid  
323 was then populated with the seven emission sources, unit transformed and quality controlled and  
324 assessed (see Supplementary). We used Monte Carlo simulations to aggregate the gross AFOLU  
325 emissions and their uncertainties and produced four final estimates, per cell: mean annual AFOLU  
326 emissions (50<sup>th</sup> percentiles) ( $\text{CO}_2\text{e.y}^{-1}$ ), associated variance, and 5<sup>th</sup> and 95<sup>th</sup> confidence intervals.  
327 Data were then aggregated to continental, and tropical scales. When aggregating uncertainties at  
328 the pixel level we assumed emission sources to be mutually uncorrelated. However, when the  
329 aggregation of the uncertainties included a change of spatial support (e.g. pixel to continental, or  
330 pixel to tropical) we assumed data complete dependence, which offered a conservative (worst-  
331 case) scenario approach for the final aggregated uncertainties (see supplementary for further  
332 information). To understand which emission sources (e.g. deforestation, degradation, livestock,  
333 paddy rice, etc) contributed the most to the final uncertainties at the continental scale, we used the  
334 variance data produced per pixel and aggregated them using the dependence assumption  
335 expressed above. The attribution of the uncertainty was then estimated as percentages of the final  
336 aggregated variance, for each emission source.

337

### 338 *Database comparison*

339 We contrast our hotspots of gross AFOLU emissions against the FAOSTAT and EDGAR  
340 databases. We run the comparisons at the country level, and produce the estimates selecting the  
341 same countries, years, emission sources, assumptions (e.g. carbon neutrality of grasslands and  
342 agricultural waste), and rules (e.g. only direct emissions) to guarantee comparability.

343

### 344 **3. RESULTS AND DISCUSSION**

#### 345 **3.1 AFOLU hotspots of emissions and uncertainties**

346 Tropical AFOLU hotspots were located on all continents but spatially concentrated in a few areas  
347 only, with 25% of the tropical area responsible for 70% of the tropical AFOLU emissions (Figure  
348 2a). Gross fluxes reached values of up to  $90 \text{ MgCO}_2\text{e ha}^{-1} \text{ yr}^{-1}$  in the hotspots, with Brazil, India,  
349 Indonesia, Democratic Republic of Congo, Angola, Central African Republic, Mozambique,  
350 Zambia, Malaysia, Sudan and Bangladesh, as the major contributors of tropical AFOLU emissions.  
351 Hotspots were mainly led by forest emissions, both by deforestation and degradation, with large  
352 hotspots over humid rainforests in the arch of deforestation in Brazil, and over the Miombo-Mopane  
353 dry forests of Africa. Deforestation and peatland fires were important in Indonesia, combined with  
354 agricultural hotspots from livestock and paddy rice. Agricultural emissions contributed less to the  
355 hotspots, with Asia, particularly India and Bangladesh, as main emitting regions with different  
356 relative contributions from livestock, paddy rice and cropland emissions (Figure 3). Southern Brazil,  
357 northern Argentina and southeastern Paraguay also showed agricultural hotspots, mainly related to  
358 livestock. Main GHGs followed these patterns, with  $\text{CO}_2$  dominating the emissions from forest  
359 activities, turning this gas into the main target for mitigation action.  $\text{CH}_4$  dominated rice and  
360 livestock emissions, while  $\text{N}_2\text{O}$  explained high cropland emissions (Figure 4).

361

362 Emissions uncertainties were the highest for the hotspot regions, reaching values of up to 30% of  
363 the mean AFOLU emissions (Figure 2b), which is lower than the reported uncertainties for global  
364 AFOLU values (e.g. 50 percent) (Smith et al., 2014). The coincidence of high AFOLU emissions  
365 and high uncertainties is not surprising since the emissions from the hotspots were led by forests,  
366 and forests host the largest emission uncertainties, in particular humid tropical forests undergoing  
367 deforestation, such as Brazil and Indonesia, and forests with high fire emissions such as dry  
368 Miombo ecosystems in Africa or peatlands in Asia. Deforestation has long been identified as a  
369 main source of emissions uncertainties in the tropics due to the combined effect of uncertain areas  
370 and uncertain carbon densities (Houghton 2010, Baccini et al., 2012, Houghton et al., 2012). High  
371 uncertainties in fire emissions relate to biomass, burned soil depths, and combustion

372 completeness, which are the most uncertain components of Van der Werf et al.(2010)'s fire  
373 emission model. Consequently, equatorial Asia and the African continent were the regions with the  
374 largest fire uncertainties of the globe (Van der Werf et al., 2010) (Fig S5 in Supplementary).

375

376 Areas with high gross emissions that also host high uncertainties (e.g. forests) complicate the  
377 effectiveness of the mitigation action. Thus, while these areas have higher mitigation potentials  
378 (e.g. high emissions that could potentially be reduced) their uncertainties affect the reliability of  
379 their emissions estimates and, therefore, the effectiveness to implement actions to stabilize  
380 atmospheric GHGs (Grassi et al., 2008). For this reason, from a climate mitigation perspective and  
381 without economic nor technical considerations, optimal mitigation scenarios would rather focus on  
382 areas with large gross fluxes and low(er) uncertainties. These areas would include agricultural  
383 hotspots (croplands, paddy rice and livestock) without much contribution from forest emissions  
384 such as parts of India, Southeastern Brazil, Northern Argentina, and Central and Southern Africa  
385 (southern DRC, Zambia, Angola) (Figure 5). Carter et al. (2015) identified that agricultural  
386 intensification and the use of available non-forest land offer opportunities for agricultural mitigation  
387 of up to 1 PgCO<sub>2</sub>e. This value coincides with sectorial analyses of mitigation targets for 2030 that  
388 would keep agricultural emissions in line with the 2 degree target (Wollenberg et al. 2016).  
389 However, food security and economic development in countries with agro-businesses make  
390 supply-based agricultural mitigation challenging (Smith et al., 2008; 2013). Moreover, as discussed  
391 in Wollenberg et al. (2016) more transformative technical and policy options will be needed to help  
392 agriculture achieve this 1 PgCO<sub>2</sub>e target. Mitigation in the agricultural sector is further complicate  
393 by being technically more complex and more expensive than forest mitigation (USEPA, 2013,  
394 Smith et al., 2014). For these reasons, and in spite of their higher uncertainties, forests still remain  
395 high in the mitigation agenda, as recently seen in the Paris Agreement, associated COP decisions,  
396 and the New York Declaration on Forests.

397

### 398 **3.2 Tropical AFOLU emissions**

399 AFOLU data from the AR5 (e.g. Figure 11.2 in Smith et al., 2014) show how the tropics have  
400 contributed with ≥70% of the global AFOLU emissions in the last decades, making this region the

401 right place to search for hotspots of land use emissions. Our aggregated gross AFOLU emissions  
402 estimates of 8.0 (5.5-12.2) PgCO<sub>2</sub>e.yr<sup>-1</sup> were in the range of other gross estimates for the same  
403 region and time period: 8.4, and 8.0 PgCO<sub>2</sub>e.yr<sup>-1</sup> for FAOSTAT and EDGAR respectively (Table 2).  
404 In spite of this good agreement, databases disagreed on the relative contribution of the leading  
405 emissions sources (Figure 6). Forests emissions showed the largest divergences, particularly  
406 forest degradation (fire and wood harvesting emissions). This outcome was expected since forest  
407 emissions were responsible for ≥70% of the tropical gross AFOLU emissions in all the databases  
408 (Table 2). Gross degradation emissions –rather than deforestation- led the forest emissions in our  
409 AFOLU gross emissions (39% vs 36% of the tropical emissions, respectively) (Table 2), with a  
410 degradation to deforestation emission ratio of 108%, reinforcing the great importance of reducing  
411 degradation for effective mitigation. Ratios above 100% for gross degradation vs deforestation had  
412 already been reported by Houghton et al. (2012) and Federici et al. (2015). Lower ratios have been  
413 observed in smaller areas (e.g. 40% Amazon, 47% Peruvian Amazon) (Asner et al., 2010;  
414 Berenguer et al., 2014) or when the ratio focuses on net fluxes of degradation (e.g. 25-35% of the  
415 net LULCC flux, if wood harvesting and shifting cultivation were not considered, and an extra 11%  
416 over the net LUCC flux when excluding peatland fire emissions in Southeast Asia alone)  
417 (Houghton et al., 2012).

418

419 In our hotspots analyses, fire led forest degradation in the tropics with almost a quarter (24.6%) of  
420 the gross AFOLU emissions. Since we had excluded deforestation fires, most of our fire emissions  
421 relate to woodlands and forest degradation. Their exclusion or incomplete inclusion would  
422 therefore result in large emission omissions in gross AFOLU assessments, and their management  
423 are key for reducing tropical emissions. Fire degradation emissions are recurrently omitted in  
424 global AFOLU assessments under the assumption of carbon neutrality of the affected burned  
425 areas (e.g. whatever carbon is emitted through fire will be fixed again by regrowth and recovery).  
426 (Houghton et al., 2012; Le Quéré et al., 2012; Canadell et al., 2014; Smith et al. 2014). This  
427 assumption does not consider current evidence of non-steady states after fire due to climatic  
428 pressures, humanized landscapes (fragmented, multi-disturbed), and increased frequencies of  
429 fires (Cochrane et al., 1999; Roman-Cuesta et al., 2014; Alencar et al., 2011, 2015; Brando et al.,

430 2014; Oliveras et al., 2014; Pütz et al., 2014). Halted successional pathways and vegetation shifts  
431 represent a large, yet unknown proportion of the burned forest ecosystems in the tropics that  
432 require further research. Recent estimates suggest that post-fire carbon recoveries in the Amazon  
433 are leading to degradation emissions in the order of  $46 \pm 29.9 \text{ MgC} \cdot \text{ha}^{-1}$  (Balch et al., under review).  
434 In spite of our uncomplete knowledge of forest post-disturbance recovery pathways (e.g. poorly  
435 understood processes, omitted emissions, missing pools, unconsidered GHGs), the importance of  
436 the forest sector for mitigation action is evidenced by the large amount of countries explicitly  
437 mentioning it in their INDCs (60%) (Grassi et al., 2015). Moreover, countries count on financial  
438 support to minimize their forest emissions and enhance their sinks, at national scale, through the  
439 REDD+ mechanism, which has now become part of the Paris Agreement (Climate Focus, 2015).

440

441 In the agricultural sector, our emissions reached estimates of  $1.9 (1.5-2.5) \text{ PgCO}_2\text{e} \cdot \text{yr}^{-1}$ , in the  
442 range of the other databases ( $2.5, 2.1 \text{ PgCO}_2\text{e} \cdot \text{yr}^{-1}$  for FAOSTAT and EDGAR respectively). These  
443 values, represent a relatively small part of the AFOLU emissions in the tropics (25-30%) but an  
444 attribution of the forest emissions to their drivers would highlight back the importance of agriculture  
445 as the main engine behind tropical forest loss and emissions. Thus, for the period 1980-2000,  
446 83% of the agricultural expansion in the tropics was at the expense of forests (Gibbs et al., 2010),  
447 calling for integrated mitigation programmes that simultaneously include forestry and agriculture  
448 (Carter et al., 2015). Our agricultural estimates represented only ca. half of the agricultural  
449 emissions reported globally for 2000-2009 ( $5-6 \text{ PgCO}_2\text{e} \cdot \text{yr}^{-1}$ ) (Smith et al., 2014; Tubiello et al.,  
450 2015). This highlights the major role of agriculture in non-tropical countries and emergent  
451 economies like China, although agricultural emissions are rising faster in developing countries than  
452 in developed ones (Smith et al., 2014). The agricultural sector is the largest contributor to global  
453 anthropogenic non- $\text{CO}_2$  GHGs, accounting for 56 % of emissions in 2005 (USEPA, 2013). Enteric  
454 fermentation and agricultural soils are globally the main sources of agricultural emissions (Smith et  
455 al., 2014), and show a strong rising trend since the 70's (1% per decade) due to increases in  
456 animal heads and the use of synthetic fertilizers (Tubiello et al., 2015). Areas with growing  
457 emission trends are attractive for land-based mitigation action and countries are engaging in  
458 agricultural mitigation in their INDCs through climate smart initiatives (Richards et al., 2015).

459 However, more transformative technical and policy options and higher level of financial support will  
460 be needed for further achievements in this sector (Wollenberg et al., in press). The most prominent  
461 agricultural mitigation practices include improved cropland and grazing land management,  
462 restoration of degraded lands, and cultivated organic soils. Lower, but still significant mitigation  
463 potential is provided by water and rice management, livestock management and manure  
464 management, set-aside, land use change and agroforestry (Smith et al., 2008).

465  
466 In terms of gases, CO<sub>2</sub> led the AFOLU emissions in the tropics with ca.70% of the tropical  
467 emissions 5.5 (3.3-9.5) PgCO<sub>2</sub>e.yr<sup>-1</sup> (Table 2, Figure 4). The remaining non-CO<sub>2</sub> contribution (30%)  
468 was mainly led by CH<sub>4</sub> 1.5 (1.1-1.9) PgCO<sub>2</sub>e.yr<sup>-1</sup>, due to livestock and rice. Non-CO<sub>2</sub> emissions  
469 from biomass burning (N<sub>2</sub>O and CH<sub>4</sub>), represented 15-34% of the CO<sub>2</sub> emissions in the tropics  
470 (Table 2). These values reinforce the need to run multi-gas assessments (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) for the  
471 AFOLU sector in order to gain a more coherent understanding on how the land affects the  
472 atmospheric composition and forces the climate. Thus, while temperature rise by around the end of  
473 this century will relate to the total emissions of long-lived greenhouse gases between 2000 and  
474 2100 (e.g. CO<sub>2</sub>) (Anderson 2012) recent research concludes that cumulative warming capacity of  
475 concurrent biogenic CH<sub>4</sub> and N<sub>2</sub>O emissions is about a factor of 2 larger than the cooling effect  
476 resulting from the global land CO<sub>2</sub> uptake in the 2000s (Tian et al., 2016). This results in a net  
477 positive cumulative impact of the three GHGs on the planetary energy budget, which calls for  
478 shorter-term mitigation initiatives (Tian et al., 2016).

479  
480 At the aggregated tropical scale, uncertainties were higher (up to 50% of the mean emissions) than  
481 at the landscape scale (0.5°) (30%), in line with other reports (Smith et al., 2014; Tubiello et al.,  
482 2015). The spatial scale of the emission assessments influences, therefore, the final uncertainties  
483 due to assumptions about the spatial correlation of the errors. Several authors have suggested the  
484 importance of working at more detailed spatial scales (e.g.30m) to reduce the uncertainties,  
485 particularly of forest emissions, by having more accurate data on forest area changes and carbon  
486 densities (Houghton, 2005; Grassi et al., 2008; Asner et al., 2010; Baccini et al., 2012; Houghton et  
487 al., 2012).

489 To better understand the uncertainty role of the different emission sources at the tropical  
490 aggregated scale, we ran a partitioning of the tropical uncertainty. We found a disproportional  
491 contribution of deforestation to the tropical uncertainty budget (92.5%) (Table 2), which agreed with  
492 the results from other researchers (Morton et al., 2011) but left the remaining emission sources  
493 with a surprisingly modest contribution to the final uncertainty (7.5%). As it was the case for the  
494 hotspots, untangling the relative contribution of the emission sources to the tropical uncertainty  
495 budget brings in trade-offs between prioritizing mitigation action on sources that are large emitters  
496 but are highly uncertain (e.g. deforestation is responsible for 36% of the emissions but carries  
497 almost all the tropical emission uncertainty (92.5%) (Table 2) or choosing emitters that contribute  
498 less to the total emissions but are more certain (e.g livestock contributed less to the tropical  
499 emissions (15%) but had a very small part on the uncertainty budget (1.4%) (Table 2) (Figure 5).

500

### 501 **3.3 Continental AFOLU emissions**

502 Continents contributed similarly to the tropical AFOLU gross emissions: 2.7 (1.8-4.5), 2.8 (1.9-4.0),  
503 2.5 (1.7-3.8) PgCO<sub>2</sub>e.yr<sup>-1</sup>, for Central and South (CS) America, Africa, and Asia, respectively  
504 (Table 2). Area-weighted emissions would, however, turn Asia into the largest continental source  
505 with a mean of 3.2 MgCO<sub>2</sub>e.ha<sup>-1</sup>. yr<sup>-1</sup> followed by Africa and CS America with 1.3 and 1.35  
506 MgCO<sub>2</sub>e.ha<sup>-1</sup>. yr<sup>-1</sup>, each. The leading sources for the continental emissions disagreed among  
507 databases but our hotspot research suggested that African emissions were dominated by fire over  
508 dry forests (52.6% of the African emissions, Table 2) which corroborates its description as “the fire  
509 continent” (Figure 7) (Mbow, 2014). Any effective mitigation action will therefore need to consider  
510 fire management, particularly in Miombo dry forests and Sudano-Sahelian woodlands, which are  
511 the most affected forests and the largest contributors to emissions hotspots. Contrastingly, Central  
512 and South America were mainly led by deforestation (60% of the continental emissions) and forest  
513 degradation (20%), mostly affecting humid forests. And while deforestation in Brazilian rainforests  
514 has since reduced, it has increased in dry forest in the region (e.g. the Chaco region in Argentina,  
515 Paraguay, and Bolivia) (Hansen et al., 2013), and forest degradation has also increased (Brando et  
516 al., 2014; Federici et al., 2015). The Asian emissions were the most diverse and were similarly led

517 by different sources: i) paddy rice (Asia is the world's largest rice-producing region and is  
518 responsible for over 80% of the total CH<sub>4</sub> emissions) (USEPA 2013); ii) livestock activities  
519 (Tubiello et al., 2014); iii) deforestation (Hansen et al., 2013), and iv) fire over peatlands,  
520 particularly in Indonesia (Van der Werf et al., 2010; Gaveau et al., 2014) (Table 2). Moreover, the  
521 Asian continent has the peculiarity of emitting almost half of the tropical non-CO<sub>2</sub> emissions (47%,  
522 Table 2), as observed by other authors (USEPA, 2013) and still has positively growing emission  
523 trends (Tubiello et al., 2014). Effective mitigation action on non-CO<sub>2</sub> emissions is therefore key for  
524 Asian and global mitigation.

525

526 The partitioning of the tropical AFOLU uncertainty at continental level showed that CS America  
527 contributed with half of the variance (48%, Table 2), which was expected since the emissions of  
528 this continent are led by the most uncertain source (deforestation). Africa and Asia contributed  
529 similarly to the rest of the uncertainty (27.3% and 24.7% respectively). Based on the uncertainty of  
530 the emissions, mitigation investments in CS America, would be, therefore, less effective than  
531 investing in Africa and Asia, particularly out of the forests.

532

#### 533 4. CONCLUSIONS

534 Our dataset offers novel landscape scale information on the spatial distribution of hotspots of  
535 AFOLU gross emissions and their uncertainties, disaggregated by gases and by leading emission  
536 sources. This AFOLU analysis can be useful as a benchmark against which countries can assess  
537 their progress on reducing their AFOLU gross emissions, in a comparable and comprehensive  
538 manner across the tropics. Moreover, gross assessments offer useful insights on potential drivers  
539 behind the emission sources, which can then lead to policies and measures to reduce these  
540 emissions, through appropriate mitigation actions. Aggregated gross emissions at the country level  
541 are offered in the Supplementary Material. Our data highlight: i) the existence of AFOLU emissions  
542 hotspots on all continents, with particular importance of evergreen rainforest deforestation in  
543 Central and South America, fire in dry forests in Africa, and both peatland emissions and  
544 agriculture in Asia; ii) a predominant contribution of forests and CO<sub>2</sub> to the total AFOLU emissions  
545 (69%) and to their uncertainties (98%), iii) higher gross fluxes from forests coincide with higher



546 uncertainties, making agricultural hotspots appealing for effective mitigation action, however iv)  
547 agricultural non-CO<sub>2</sub> emissions are much lower (ca. 25% of the total gross emissions in the tropics  
548 for 2000-2005) than forests, with livestock (15.5%) and rice (7%) leading the emissions. Gross  
549 AFOLU tropical emissions 8.0 (5.5-12.2) were in the range of other databases 8.4 and 8.0  
550 PgCO<sub>2</sub>e.yr<sup>-1</sup> (FAOSTAT and EDGAR respectively).

551  
552 It is worth remarking that under gross assessments, sustainable wood harvesting and fires not  
553 leading to long-term declines in carbon storage would still be considered as hotspots of emissions  
554 since the subsequent recovery of carbon stocks is not considered. Moreover, gross emission are  
555 not to be confused with the overall net land-to-atmosphere flux due to human land use, because  
556 legacy effects and changes in litter and soil organic matter are not included. Independently of  
557 whether we are working with gross or net emissions, the time scale under analysis (e.g. 5 years in  
558 our study) and the selected methodological assumptions, will strongly influence the final AFOLU  
559 estimates (e.g. exclusion of indirect emissions, exclusion of non-managed land, exclusion of legacy  
560 emissions, consideration of instantaneous emissions, exclusion of transboundary effects) and  
561 mitigation options (e.g. exclusion of substitution effects).

562  
563 Net assessments of AFOLU emissions would be closer to what countries are requested to report  
564 under the UNFCCC but, how different would it be from our gross hotspots results? We argue that,  
565 under our short-term temporal analysis (5 years), the spatial location of our gross AFOLU hotspots  
566 would not differ much, although the absolute emission estimates could be lower. Thus, for  
567 agricultural driven hotspots (crops, livestock, rice), gross and net assessments would result in the  
568 same hotspots of AFOLU emissions because only soil organic storage acts as a sink to  
569 compensate against agricultural non-CO<sub>2</sub> emissions, and soil carbon storage rates are small and  
570 short-lived (Smith et al., 2008). For areas with extended forests, gross and net AFOLU hotspots  
571 can differ. This would not be the case for areas affected by large scale deforestation and/or intense  
572 degradation, because the 5-year time frame of our study would not allow for significant carbon  
573 recovery after disturbance. Most of the differences between gross and net assessments would  
574 then concentrate in two forested areas: i) those undergoing large scale, high density removals (e.g.

575 afforestation/ reforestation processes) such as in China –not included in this research-, Viet Nam  
576 and India, and in ii) forested areas affected by smaller scale disturbances and less intense carbon  
577 processes that do not lead to deforestation, where forest emissions could be compensated by  
578 removals from the standing forests. Main differences between gross and net land use hotspots  
579 assessments in the tropics would then concentrate on areas with extended forests undergoing  
580 small-scale mid-level disturbances: Amazonian areas out of the arch of deforestation, wood  
581 harvesting in the Congo Basin, and lower impact disturbances in Mesoamerica, including Mexico.

582  
583  
584 This study also contributes to the debate on tropical mitigation potentials of agriculture and  
585 forestry. Thus, even if global estimates of agriculture and forestry emissions have roughly similar  
586 mitigation potentials (Smith et al., 2014; Tubiello et al., 2015), economic feasibilities differ. Forests  
587 have two to three-fold greater economic mitigation potentials than agriculture (e.g. 0.2-13 vs 0.3-  
588 4.6 PgCO<sub>2</sub>e.yr<sup>-1</sup> respectively) at prices up to 100 USD/MgCO<sub>2</sub>e (Bajzelj et al., 2014; Havlik et al.,  
589 2014; Smith et al., 2014). This means that for the same price, more emission reductions can be  
590 achieved in the forest sector. These unequal results relate to forests sector being much more  
591 carbon dense, to lower costs per area unit to monitor and implement actions against deforestation  
592 and degradation, but also to concerns about food security and adaptation needs (Smith et al.,  
593 2013; Havlik et al., 2014, Wollenberg et al, 2016). Thus, notwithstanding the importance of  
594 agricultural mitigation, forests are more cost effective alternatives and, although uncertain, their  
595 multiple ecosystem services will keep them high as desirable mitigation targets in the political  
596 arena.

597

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## 808 **Websites**

809 www1: <http://faostat3.fao.org/home/E>

810 www2: <http://edgar.jrc.ec.europa.eu/overview.php?v=42FT2010>

811 www3: [http://edgar.jrc.ec.europa.eu/docs/IEA\\_PARTIII.pdf](http://edgar.jrc.ec.europa.eu/docs/IEA_PARTIII.pdf)

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## 813 **6. CONTRIBUTIONS**

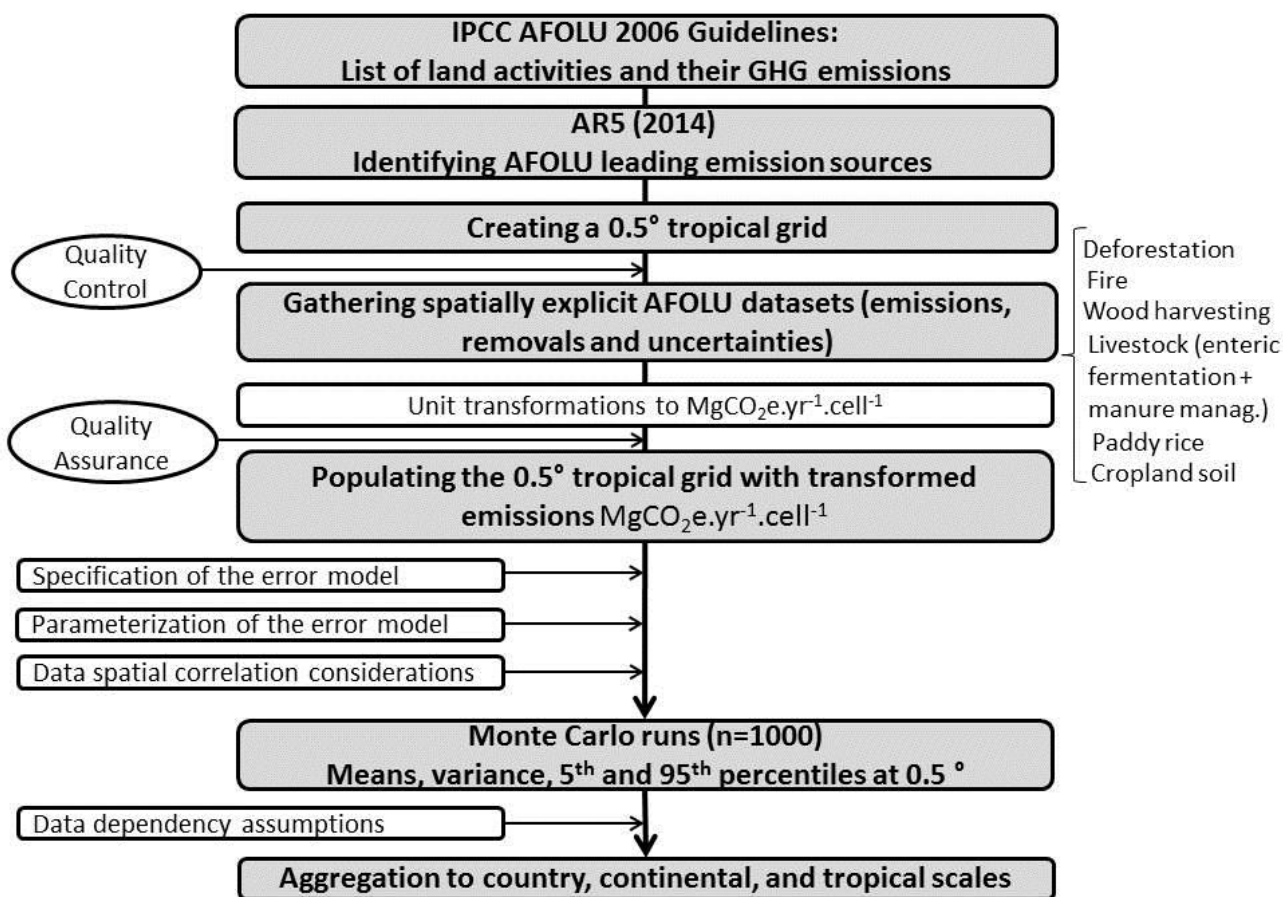
814 RMRC, MR, MH1, KBB, TR, JS, LV designed the study. RMRC, MH2, CM, CL, SO, BP provided  
815 data and ran quality control, quality assessments and uncertainties expert judgements on the data  
816 sets. SdB guided on statistics and developed all the scripts for the Monte Carlo aggregation of the  
817 data. JS assisted with GIS analyses and scripts. RMRC, MR, MH1, KBB, TR, SdB, LV, CM, JS,  
818 MH2, CL, SO, BP, discussed the results and contributed to writing.

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## 820 **7. ACKNOWLEDGEMENTS**

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823 Climate Change, Agriculture, and Food Security (CCAFS). Funding also came from two European  
824 Union FP7 projects: GEOCarbon (283080) and Independent Monitoring of GHG Emissions-N°  
825 CLIMA.A.2/ETU/2014/0008. Partial funds came through CIFOR from the governments of Australia  
826 (Grant Agreement # 46167) and Norway (Grant Agreement #QZA-10/0468). In memoriam: Dr.  
827 Changsheng Li. The authors of this manuscript would like to homage Dr. Li for his life-long  
828 dedication to science.

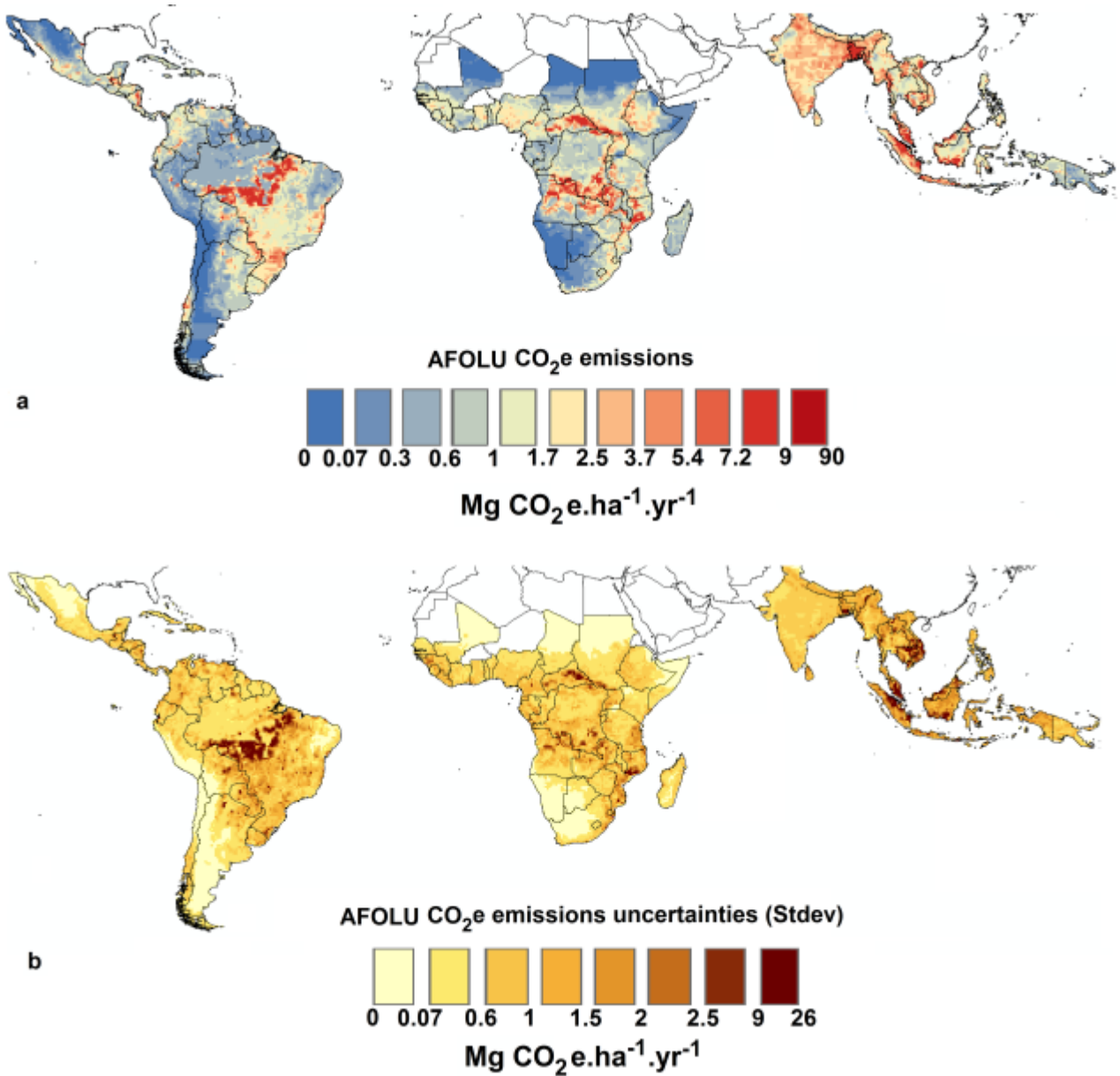
829 Figure legends



830

831 **Figure 1:** Methodological framework used to estimate the aggregated AFOLU emissions (annual means)  
 832 and associated uncertainties (variance, 5<sup>th</sup>, 95<sup>th</sup> percentiles) at 0.5° resolution, for 2000-2005.

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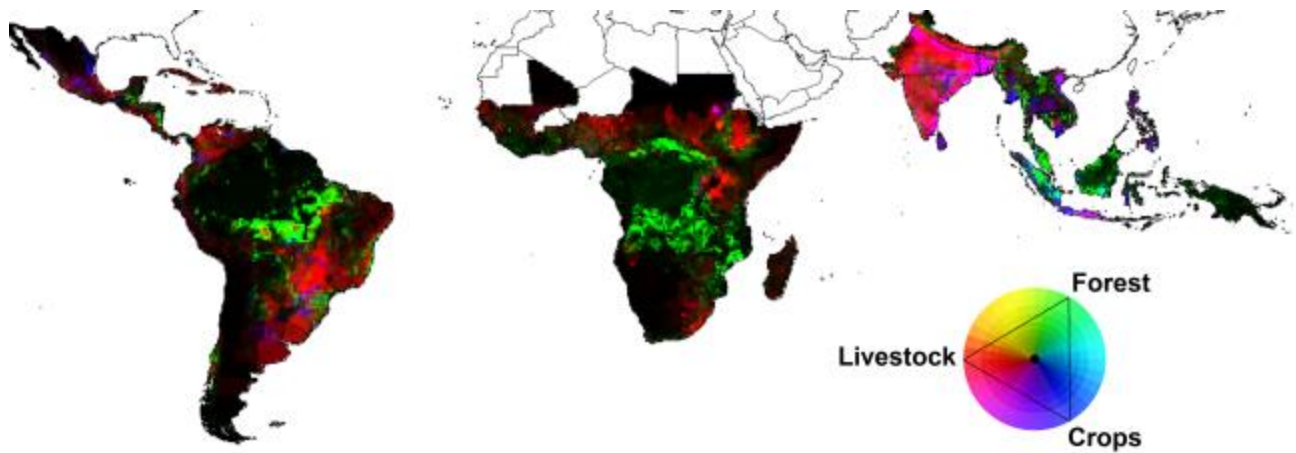
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Figure 2: (a) Hotspots of annual AFOLU emissions (red cells) and (b) associated uncertainties ( $1\sigma$ ) in  $\text{MgCO}_2\text{e.ha}^{-1}.\text{yr}^{-1}$  for the tropical region, for the period 2000-2005, at  $0.5^\circ$  resolution. Emissions are the result of 1000 Monte Carlo simulations for the leading AFOLU emission sources (deforestation, degradation (fire, wood harvesting), soils (crops, paddy rice), livestock (enteric fermentation and manure management)

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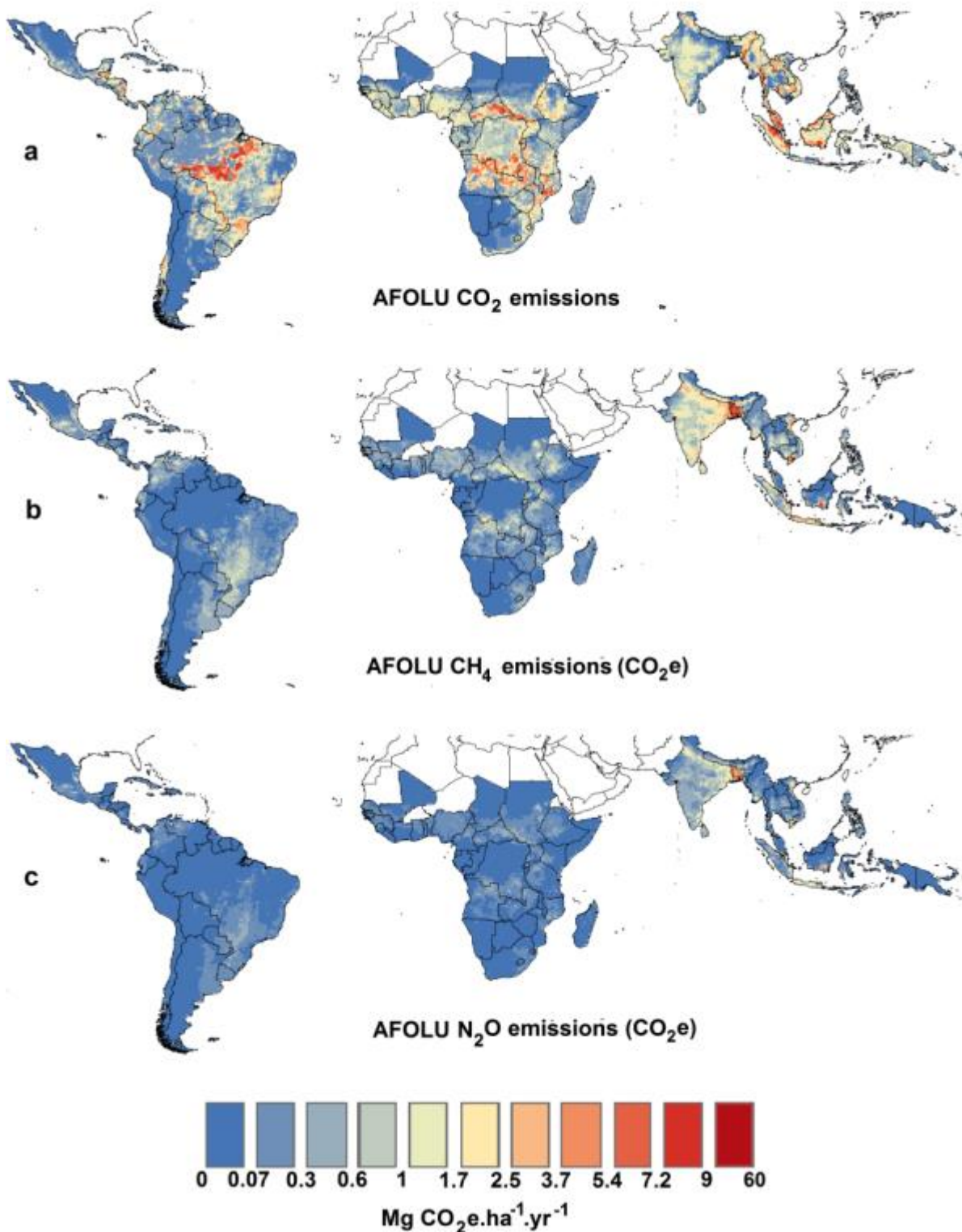
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845 Figure 3: Contribution of the leading emission sources in percent of total emissions (grouped into forests,  
846 crops and livestock) to the per pixel (0.5°), for 2000-2005. Forest emissions include fire, deforestation and  
847 wood harvesting. Crops emissions includes paddy rice, cropland soil and croplands over drained histosols.  
848 Livestock includes enteric fermentation and manure management emissions. This figure is an RGB image  
849 where final colours represent the strength of the emissions for the three sources (e.g, fuchsia colours in Asia  
850 represent equal emissions from livestock (red) and crops (blue). Dark represent areas of low emissions.

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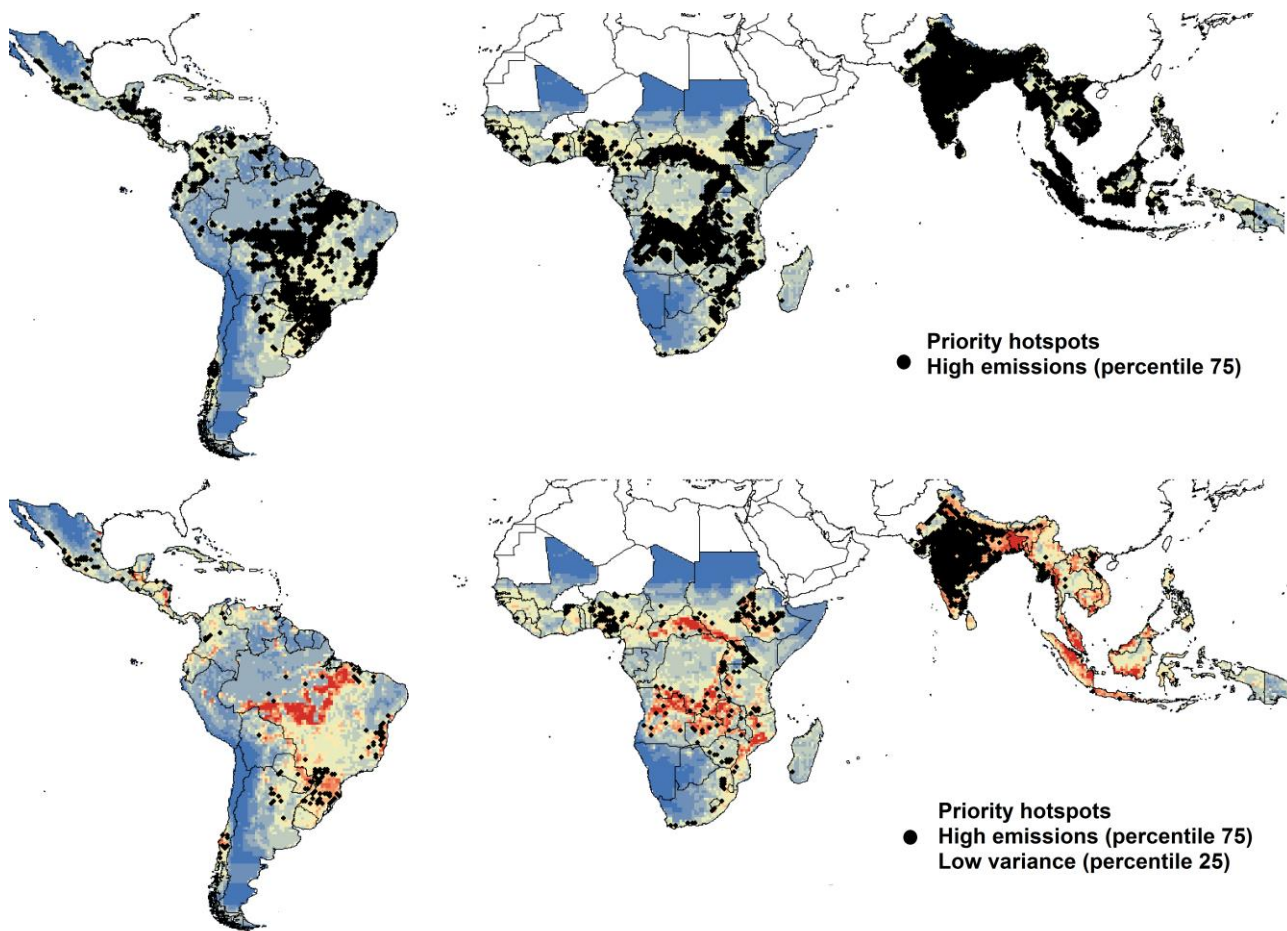
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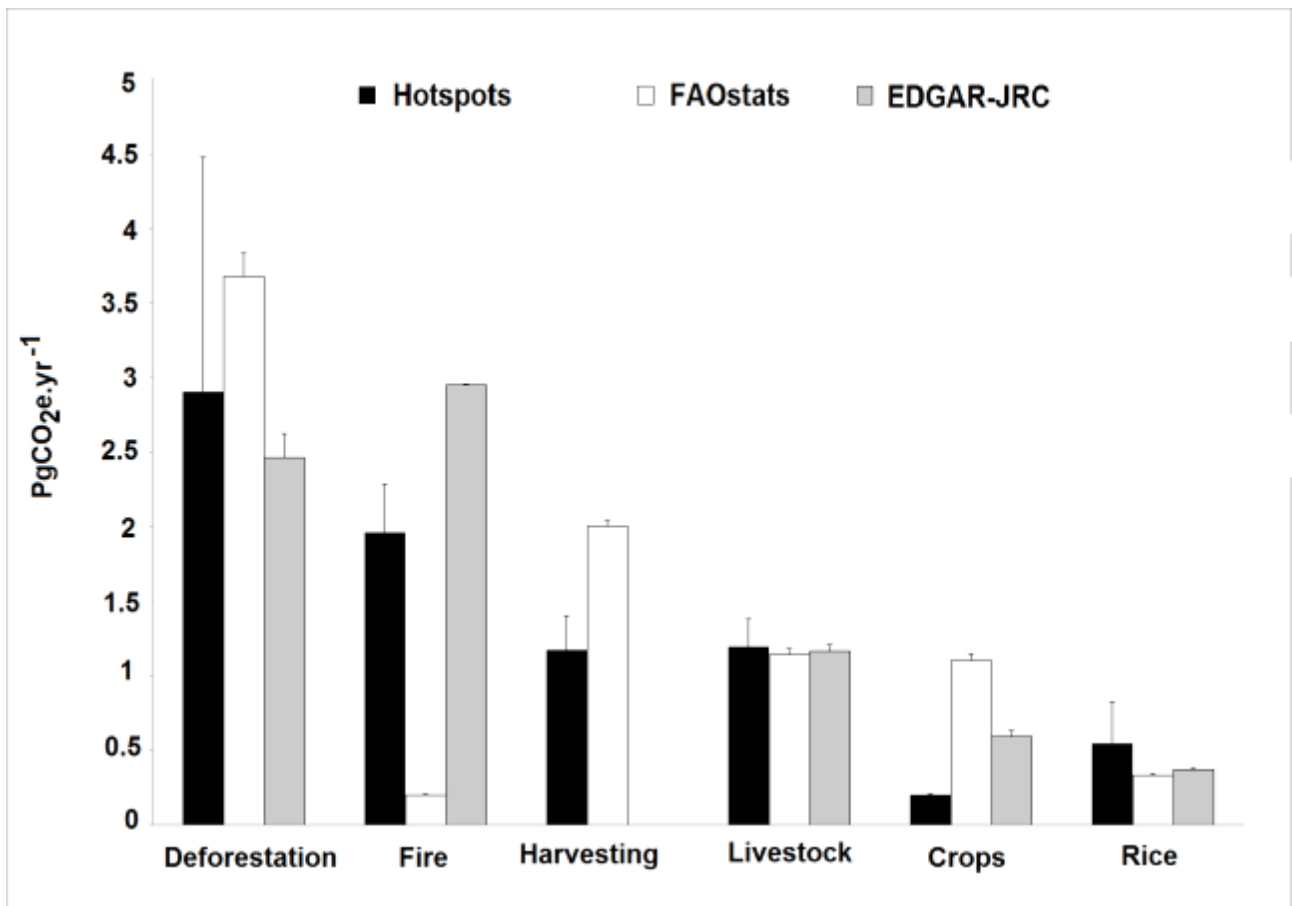
**Figure 4:** Mean annual AFOLU emissions (MgCO<sub>2</sub>e.ha<sup>-1</sup>.yr<sup>-1</sup>), for the period 2000-2005, disaggregated by greenhouse gases: a) CO<sub>2</sub> emissions, which are a proxy of forest emissions, (b) CH<sub>4</sub>, and (c) N<sub>2</sub>O emissions, which are proxies of agricultural emissions. Emissions are the result of 1000 Monte simulations for the considered leading AFOLU emission sources.



860

861 **Figure 5:** Identification of priority regions for mitigation in the AFOLU sector in the tropics, for 2000-2005  
 862 considering mitigation potentials only (higher emissions, percentile 75), and degree of certainty of these  
 863 potentials (low uncertainties, percentile 25 of the variance). Economic feasibility would lead to different  
 864 priority regions.



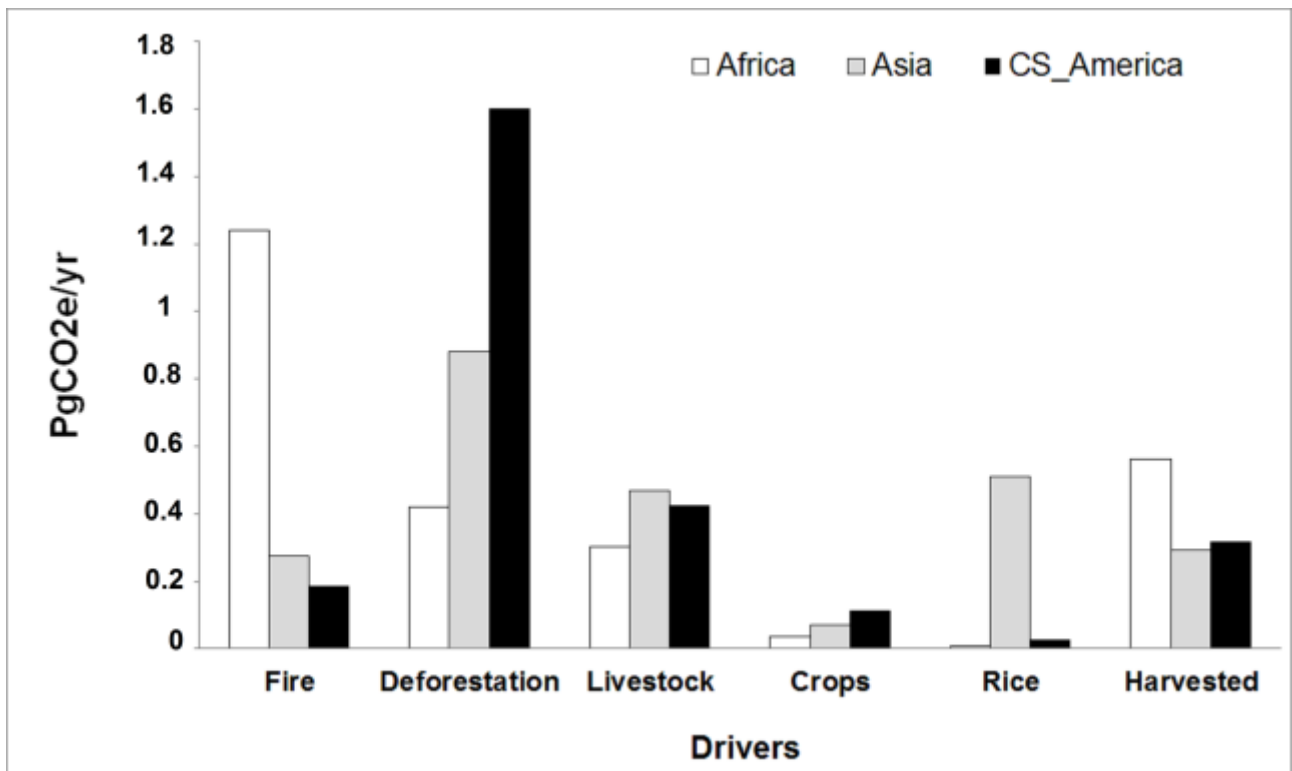


865  
 866 Figure 6: Distribution of mean annual emissions per sources of emissions for the three data bases, in the  
 867 tropics, for 2000-2005. Bars indicate uncertainty estimates for the hotspot data base, and standard  
 868 deviations for FAOstats and EDGAR-JRC estimated from country data variability since no uncertainty data  
 869 exist for these two data bases.

870

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872

873 Figure 7: Continental contribution of the leading emission drivers for our hotspot data base, in the tropics, for  
 874 the period 2000-2005.

875

From units	To units	Molecular weights conversion	Global Warming Potentials 100-yr
kgC (dSOC)	kg CO <sub>2</sub> eq.	kgC * 44 / 12	1
kgC (CH <sub>4</sub> )	kg CO <sub>2</sub> eq.	kgC * 16 / 12	21
kgN (N <sub>2</sub> O)	kg CO <sub>2</sub> eq.	kgN * 44 / 28	310

877 **Table 1:** Data conversions to CO<sub>2</sub>e for different chemical elements (C, N). dSOC is the change in Soil  
878 Organic Carbon. Molecular weights and global warming potentials use values from the Fourth Assessment  
879 Report (IPCC 2007)

880

Gross AFOLU emissions (PgCO <sub>2</sub> e.yr <sup>-1</sup> )						
		CO <sub>2</sub> e	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	
Tropical		8.0 (5.5-12.2)	5.5 (3.3-9.5)	1.5 (1.1-1.9)	1 (0.8-1.2)	
Central & South America		2.7 (1.8-4.5)	2.1 (1.3-3.8)	0.35 (0.25-0.45)	0.25 (0.2-0.3)	
Africa		2.8 (1.9-4.0)	2.1 (1.4-3.2)	0.39 (0.27-0.5)	0.3 (0.22-0.39)	
Asia		2.5 (1.7-3.8)	1.3 (0.7-2.4)	0.74 (0.56-0.95)	0.41 (0.35-0.47)	
Contribution of leading emission sources to the tropical AFOLU gross emissions (%)						
	Deforestation	Fire	Rice	Wood Harvesting	Livestock	Crops
Tropical	36.3	24.6	6.9	14.6	15	2.5
Central & South America	59.6	8.2	1.1	11.9	15.9	3.4
Africa	15.2	52.6	0.3	20.3	11	0.7
Asia	34.8	11.3	20.2	11.5	18.5	3.7
Contribution of leading emission sources to total uncertainty (%)						
	Deforestation	Fire	Rice	Harvesting	Livestock	Crops
Tropical	92.5	4.5	0.2	1.4	1.4	0.0
Central & South America	98.4	0.3	0.0	0.5	0.8	0.0
Africa	69.8	25.5	0.0	3.7	1.1	0.0
Asia	91.4	2.4	2.1	1.1	2.9	0.0
Contribution of different gases to the tropical AFOLU gross emissions (%)						
		CO <sub>2</sub> e	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	
Tropical			69	19	12	
Central & South America		34	78	13	9	
Africa		35	75	14	11	
Asia		31	53	30	17	
Contribution of different gases to total uncertainty (%)						
			CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	
Tropical			98.3	1.3	0.4	
Central & South America		48	99.4	0.5	0.1	
Africa		27.3	98.2	1.1	0.7	
Asia		24.7	95.5	3.9	0.6	

881

882 **Table 2:** i) Contribution of the different greenhouse gases (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O) to continental and tropical  
883 AFOLU annual mean gross emissions for the period 2000-2005 (in parenthesis are the 5<sup>th</sup> and the 95<sup>th</sup>  
884 percentiles of the aggregated AFOLU emissions). ii) Contribution of the different leading emission sources to  
885 the tropical and continental AFOLU gross emissions (expressed as % of emissions). And iii) partitioning of  
886 the AFOLU emissions uncertainties among the leading emission sources and the considered GHG gases  
887 (expressed as % of variance) AFOLU emissions are the result of 1000 Monte Carlo runs for the leading  
888 emission sources.