



1 **Exploring the contributions of vegetation and dune size to early dune**
2 **building using unmanned aerial vehicle (UAV)-imaging**

3 Short running head: Dune size and vegetation

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21 Abstract

22 Dune development along highly dynamic land-sea boundaries is the results of interaction between
23 vegetation and dune size with sedimentation and erosion processes. Disentangling the contribution of
24 vegetation characteristics from that of dune size would improve predictions of dune development under
25 a changing climate, but has proven difficult due to scarcity of spatially continuous monitoring data.

26 This study explored the contributions of vegetation and dune size to dune development for locations
27 differing in shelter from the sea. We monitored a natural dune field of 8 hectares, along the coast of the
28 island Texel, the Netherlands, for one year using an Unmanned Aerial Vehicle (UAV) with camera.
29 After constructing a Digital Surface Model and orthomosaic we derived for each dune 1) vegetation
30 characteristics (species composition, vegetation density, and maximum vegetation height), 2) dune size
31 (dune volume, area, and maximum height), 3) degree of shelter (proximity to other dunes and the
32 sheltering by the foredune). Changes in dune volume over summer and winter were related to
33 vegetation, dune size and degree of shelter.

34 We found that a positive change in dune volume (dune growth) was linearly related to dune volume
35 over summer but not over winter. Big dunes accumulated more sand than small dunes due to their larger
36 surface area. Exposed dunes increased more in volume than sheltered dunes over summer, while the
37 opposite occurred over winter. Vegetation characteristics did not significantly affect dune growth in
38 summer, but did significantly affect dune growth in winter. Over winter, dunes dominated by
39 *Ammophila arenaria*, a grass species with high vegetation density throughout the year, increased more
40 in volume than dunes dominated by *Elytrigia juncea*, a grass species with lower vegetation density. The
41 effect of species was irrespective of dune size or distance to the sea.



42 Our results show that dune growth in summer is mainly determined by dune size, whereas in winter
43 dune growth was determined by vegetation. In our study area the growth of exposed dunes was likely
44 restricted by storm erosion, whereas growth of sheltered dunes was restricted by sand supply. Our
45 results can be used to improve models predicting coastal dune development.

46 **Key words:** Biogeomorphology, embryo dunes, *Ammophila arenaria*, *Elytrigia juncea*, beach-dune
47 interaction, landform morphology, the Netherlands



48 **1. Introduction**

49 Coastal dunes occur along the sandy shores of most continents (Martínez and Psuty, 2008), and are
50 important to protect these coasts against flooding, provide areas for recreation, store drinking water and
51 shelter unique biodiversity (Everard et al., 2010). Coastal dunes and their services are threatened by
52 climate-induced sea-level rise (Carter, 1991; Feagin et al., 2005; Keijsers et al., 2016). However, dunes
53 also provide self-adapting systems of coastal protection, since the threat by sea-level rise can be
54 mitigated by the development of new dunes. Despite the obvious importance of dunes, we know
55 surprisingly little about the factors that determine early dune development. Understanding these factors
56 is essential for predicting dune development, and for safeguarding their services.

57 Dune development is the result of an interaction between vegetation and aeolian processes and
58 starts above the high-water line by the establishment of dune-building plant species (Maun, 2009). Once
59 vegetation establishes on the bare beach, it forms a roughness element that facilitates local sand
60 deposition and reduces erosion, forming a small dune within discrete clumps of vegetation (Hesp,
61 2002). At the lee side of these small clumps of vegetation a shadow dune develops by sand deposition,
62 this shadow dune has a ridge parallel to the wind direction (Hesp, 1981). Vegetation and shadow dune
63 together are known as embryo dunes, or incipient dunes (Hesp, 2002; Hesp and Smyth, 2017). Their
64 further development strongly depends on the balance between summer accumulation of sand and
65 vegetation growth and winter erosion of sand and loss of vegetation (Montreuil et al., 2013). Summer
66 growth and winter erosion depend on weather conditions, such as wind speed, precipitation and storm
67 intensity (Montreuil et al., 2013; van Puijenbroek et al., 2017). As a result, net dune growth can differ



68 from year to year. Over time the smaller vegetated dunes can develop into an established foredune that
69 forms the first line of coastal defense against flooding.

70 Most research on coastal dune growth and erosion have focussed on processes and factors that
71 influence the supply of sand to the dunes and the effect of storm intensity on dune erosion (Anthony,
72 2013; Haerens et al., 2012; Houser et al., 2008; Keijsers et al., 2014; Saye et al., 2005; de Vries et al.,
73 2012). However, how coastal dune growth and erosion rates are influenced by the individual dune
74 characteristics, such as dune size, vegetation and degree of sheltering are less well studied. Dune size
75 affects the wind flow pattern, thus affecting sand deposition (Walker and Nickling, 2002) for example
76 increasing height or length of the shadow dune (Hesp, 1981; Hesp and Smyth, 2017). Dune size also
77 influences storm erosion: Claudino-Sales (2008) found that foredunes with a higher volume were less
78 sensitive to erosion. Whether the latter also applies to embryo dunes, is unknown. Differences in
79 vegetation density between plant species are known to modify sand deposition, storm erosion, dune
80 morphology, and growth (Charbonneau et al., 2017; Hacker et al., 2012; Seabloom et al., 2013;
81 Zarnetske et al., 2012). Sheltering by other dunes can decrease the sand supply but can also reduce
82 erosion by waves (Arens, 1996; Montreuil et al., 2013). Although dune size, vegetation and sheltering
83 are known to be important for individual dune development, the relative contributions of these factors
84 are unknown.

85 In this study, we explored the contribution of vegetation and dune size to dune development.
86 Using an unmanned aerial vehicle (UAV) with camera we monitored a natural dune field for one year.
87 From the aerial images we constructed a digital terrain model (DTM) and an orthomosaic. From the



88 DTM and orthomosaic we extracted detailed data on dune size (dune area, volume and maximum
89 height), vegetation characteristics and the degree of sheltering. We related dune size and vegetation to
90 changes in dune volume over a summer (April - August) and winter period (November - April). We
91 expected dune growth to be a function of dune size and vegetation density, dune growth being the
92 largest for big dunes with high vegetation density. We also expected that the effect of sheltering on
93 dune growth would depend on season: exposed dunes growing faster in summer, but slower in winter.

94 2. Methods

95 2.1 Study area

96 We monitored 8 hectares (200 m x 400 m) of a natural dune field with a large range of dune sizes at ‘the
97 Hors’, the southern tip of the barrier island at Texel, the Netherlands, coordinates: 52°59’43.70”N,
98 4°43’47.53”E (Fig. 1). The Hors is a wide dissipative beach with a high degree of hydrodynamic
99 reworking of the sand, which results in a high transport potential and opportunity for dunes to develop.
100 In the last 20 years many vegetated dunes have developed on the beach. At this area permanent dunes
101 are formed by plant species *Ammophila arenaria*, *Elytrigia juncea* or a mixture of both species. These
102 three types of vegetated dunes occurred at similar distances from the sea, making this area ideal for
103 testing the effects of dune size and species composition on dune growth. *A. arenaria* and *E. juncea*
104 differ in their vegetation characteristics: *A. arenaria* grows in dense patches, whereas *E. juncea* has a
105 more sparse growth form. This difference in growth form probably also results into a different dune
106 morphology: *A. arenaria* forms higher ‘hummocky’ shaped dunes, whereas *E. juncea* builds broader
107 and lower dunes (Bakker, 1976; Hacker et al., 2012). The monitoring area is bisected by a low,



108 continuous foredune ridge that runs parallel to the shore. The dunes that occur at the seaward side of
109 this foredune are more exposed to the sea, while the dunes occurring at the landward side of the
110 foredune are more sheltered from the sea, enabling us to explore whether the effects of dune size and
111 vegetation are modified by the degree of shelter. The foredune in our monitoring area has a maximum
112 height of 7m NAP (NAP refers to Amsterdam Ordnance Date, which refers to mean sea level near
113 Amsterdam).

114 Summer conditions during our study period were similar to previous years, while winter
115 conditions were calmer than usual (Supplementary S1). The precipitation during the growing season
116 was 276 mm, and the average temperature in June and July was 16 °C. The most common wind
117 direction was South to South-West. The most common wind speed in summer was 4 - 5 m s⁻¹, and the
118 maximum wind speed was 13 m s⁻¹. In winter the wind speed was higher compared to summer, the
119 most common wind speed was 5 – 6 m s⁻¹ and the maximum wind speed was 19 m s⁻¹. We registered
120 one storm during the study period. This storm, however, could be classified as relatively weak. The
121 highest water level was 211 cm NAP; compared to 248 cm NAP and 254 cm NAP from previous years.
122 The storm, which was the first of the season, occurred after the beginning of our mapping campaign.

123 **2.2 Data collection**

124 Three UAV flights in November (2015), April (2015) and August (2016) were carried out with a rotary
125 octocopter UAV system (Aerialtronics Altura Pro AT8 v1) and camera equipment of WageningenUR
126 *Unmanned Aerial Remote Sensing Facility*. The octocopter was equipped with a Canon EOS 700D
127 single-lens reflex camera with a 28mm f/2.8 Voigtländer Color Scopar SL-II N objective. The camera



128 sensor was modified to give a false colour output. The red channel of the camera had been converted to
129 be sensitive in the near-infrared, with centre point around 720nm. The blue channel of the camera had
130 been extended to also cover the UV region of the spectrum. The green channel was left with almost
131 original response. The false colour modification enabled the calculation of a modified Normalised
132 Difference Vegetation Index (NDVI), a commonly used measure for vitality and/or cover of the
133 vegetation (Carlson and Ripley, 1997). Aerial images were acquired by auto-piloted flights at an
134 altitude of 80 m at $4 - 5 \text{ m s}^{-1}$ velocity. The camera was set to take one image per second. The auto-
135 piloted flights enabled us to have the same flight paths for each of the three mapping campaigns. The
136 flight paths ensured that images had a minimum of 85% forward and 65% side-way overlap. Four
137 flights of 10 minutes were needed to cover the study area, yielding up to 900 RAW false colour images
138 per mapping campaign. Five ground control points were permanently placed in the flight area and
139 measured with a RTK-DGPS Trimble R6 Model 3 (TSC3) to calibrate our images with coordinates.
140 During our mapping campaign, a Spectralon reference panel was measured with our camera
141 immediately before take-off and after landing.

142 **2.3 Radiometric calibration**

143 In order to compare the images over the time, they were calibrated and converted from RAW to 16 bit
144 tiff format. First, we ensured that each individual pixel within an image was comparable, by converting
145 the RAW digital number into radiance units using a pixel-wise dark current and flat field calibration.
146 Second, each radiance image was calibrated to a reflectance factor image in order to correct for changes
147 in incident irradiance on different flight days. This calibration was done by using a Spectralon panel



148 with a known reflectance factor. The radiometric calibration is described in more detail by Suomalainen
149 et al. (2014).

150 The images were subsequently converted into NDVI images. Usage of the standard NDVI was
151 not possible due to lack of red channel in the false color modified camera. Thus we used a custom
152 NDVI equation (Eq. 1), which was recommended by the company that modified the sensor. On their
153 website (MaxMax.com) this equation was shown to be just as effective for green vegetation as the
154 traditional NDVI formula ($R^2 = 0.77$) where the red band is taken as the absorption channel.

$$155 \quad 1) \quad NDVI = \frac{(NIR + G) - (2B)}{(NIR+G) + (2B)}$$

156 Where NIR, G, and B are the near-infrared, green and blue bands of the false colour image respectively.
157 For photogrammetric reconstruction, the NDVI image layer was stacked with the original green and
158 blue bands to form a three-color image.

159 **2.4 Photogrammetric reconstruction**

160 The large overlap between the consecutive images was necessary for photogrammetric software to
161 successfully process the aerial images into a 3D point cloud. The 3D point cloud was generated using
162 Agisoft Photoscan Professional (v. 1.2.6), using the Structure-from-Motion (SfM) and Multi-View
163 Stereo (MVS) algorithms (Fonstad et al., 2013; Westoby et al., 2012). The correlated 3D points are
164 georeferenced to match the ground control points, and contain pixel intensity values of the input
165 imagery. From this 3D point cloud we interpolated a 5 cm pixel size digital surface model (DSM) and a
166 1 cm pixel size orthomosaic image. The vertical error distribution of a DSM produced by UAV



167 photogrammetry is expected to be equivalent to airborne LIDAR data and terrestrial laser scanning
168 (Hugenholtz et al., 2013; Mancini et al., 2013). The DSM included also vegetation, which resulted in a
169 vertical error in dune height in areas where vegetation is present. We removed the vegetation from the
170 point cloud by identifying and removing the vegetation points. Vegetation points were removed by
171 distinguishing vegetation from sand using k-means clustering of the 3-D point cloud with NDVI using
172 the Hartigan and Wong (1979) algorithm in R (R Core Team, 2016). The holes in the point cloud that
173 arose by removing the vegetation were filled by using LAStools (the tool Blast2dem) (Isenburg, 2016),
174 which resulted in a Digital Terrain Model (DTM) without vegetation.

175 **2.5 Defining dunes**

176 To be able to relate dune growth to characteristics of an individual dune, we first had to define
177 individual dunes from the DTM. We followed a step-wise procedure for each of our mapping
178 campaigns (November, April, and August) using ArcGIS 10.3 (ESRI, 2016) that resulted into different
179 polygons in which each individual dunes expanded or decreased in volume over the study period. Dune
180 size and growth were later calculated using the same polygons for each measurement campaign through
181 time (see next section). To define the polygons we used the step-wise procedure described below: 1) we
182 constructed a baseline raster by calculating the average elevation in a circle of 5m radius around each
183 pixel in the DTM. A higher or lower radius resulted in either a too low or too high baseline. 2) We then
184 qualified pixels of the DTM as dunes, if they were 5 cm above a baseline raster, or had a slope of 15° or
185 higher. From these selected ‘dune’ pixels we created dune polygons. 3) Dune polygons of consecutive
186 campaigns were overlaid to construct the largest dune-covered area during the study period. 4) Each



187 polygon was visually checked for minimum size and presence of vegetation: dunes consisting of only
188 one clump of vegetation (0.4 m^2 or smaller) and dunes with no vegetation were discarded.

189 2.6 Variables

190 For each dune and for each mapping campaign we extracted dune volume (m^3), max height (m) and
191 horizontal area (m^2) from the dune polygons (see previous section) in the DTM. We calculated the
192 absolute change in dune volume by subtracting the current dune volume from the volume of the
193 previous mapping campaign, and we corrected for the number of weeks between the mapping
194 campaigns. To explore relationships irrespective of dunes size, we also calculated the relative change in
195 dune volume per week as $(V_t/V_{t-1})/\text{week}$. Where V_t is the dune volume and V_{t-1} the dune volume of the
196 previous mapping campaign.

197 We manually identified the species composition on each dune from the orthomosaic. Species
198 identification was verified in the field for a random subset of 100 dunes (23%). To this end we created 2
199 transects from the southwest border to the northeast border of the area. For these transects we
200 determined the species on each dune in the field in May 2016. We compared the presence of species in
201 the field with the orthomosaic, and adjusted the species composition if necessary. In our dataset, dunes
202 have either *A. arenaria*, *E. juncea* vegetation, or a mixture of both species. A dune was defined as
203 covered by a mixture of both species, when it had distinct vegetation patches of both species present.
204 For each dune and mapping campaign we also extracted the vegetation density and the maximum plant
205 height. To assess vegetation density we first distinguished vegetated pixels from non-vegetated pixels
206 based on the orthomosaic using k-means classification of the NDVI using the MacQueen (1967)



207 algorithm. Hereafter, the vegetation density (NDVI/cm² dune) was calculated by summing the NDVI
208 values of all vegetated pixels within the dune polygon and then dividing this summed NDVI by the total
209 number of cm² pixels within the dune polygon. The maximum plant height was calculated by
210 subtracting the DTM (with vegetation) from the DSM (without vegetation).

211 Sheltering can affect the sand supply and storm erosion. We used two methods to define the
212 extent of sheltering. Firstly, we distinguished whether a dune was seaward or landward from the
213 foredune. Secondly we determined how much a dune was clustered with other dunes. We extracted the
214 degree of clustering for each dune by calculating the mean height from the DTM in a 25 m radius
215 around the dune. All data extraction from the DSM, DTM and orthomosaic were done in R (R Core
216 Team, 2016).

217

218 **2.7 Statistical analysis**

219 First we explored if dune area, volume, maximum dune height, clustering (mean height in a 25m radius
220 around the dune), vegetation density and maximum plant height depended on species composition using
221 August 2016 data. As the number of dunes per species composition was unequal, we used an ANOVA
222 type III SS, to compensate for the unequal sample size (Fox and Weisberg, 2011) and then used a Tukey
223 HSD test (Hothorn et al., 2008) to determine significant differences between the dunes with different
224 species compositions.



225 Secondly, we tested how absolute changes in dune volume over winter (November – April) and
226 summer (April – August) periods related to the dune volume at the beginning of the period at location
227 with different degree in sheltering with a linear regression model.

228 Thirdly, we analysed how the relative changes in dune volume over winter and summer
229 depended on dune size and vegetation characteristics in separated linear mixed models (Pinheiro et al.,
230 2009). To correct for spatial autocorrelation and species distribution we ran this analyses on a subset of
231 236 (54%) dunes. To this end we first explored the degree of spatial autocorrelation in our dataset by
232 creating a variogram. To account for the spatial autocorrelation of 25 m in our dataset we imposed a 50
233 m x 50 m grid over our study area; all dunes that were located within a grid cell (referred to as block)
234 were assumed to show spatial autocorrelation to some extent. This spatial autocorrelation was corrected
235 for in our statistical model by including block as a random intercept. We had 10 blocks seaward from
236 the foredune and 11 blocks landward from the foredune (Fig. 2), in which all species combinations
237 occurred (*A. arenaria* dunes, *E. juncea* dunes and *A. arenaria* + *E. juncea* dunes). By only including
238 dunes that were located within a block in the analysis, our selection was biased towards smaller dunes,
239 since larger dunes often fell within multiple blocks. We do expect that the effect of vegetation is more
240 apparent for these smaller dune compared to larger dunes. To better distinguish between effects of
241 species compositions and vegetation structure we used two different models. The effect of species
242 composition was tested in a model with dune volume, maximum dune height, clustering and species,
243 whereas the effect of vegetation structure was tested in a model with dune volume, maximum dune
244 height, dune clustering, vegetation density and maximum plant height as explanatory variables. Within
245 each model we used the initial conditions for the explanatory variables, with initial conditions being the



246 values at the start of each measurement campaign. We included all two-way interactions. We selected
247 the best model by using Akaike information criterion (AIC). As we were mainly interested in the
248 importance of the explanatory variables relative to each other, we calculated the standardised estimates
249 for all the models by scaling the explanatory data.

250 The normality and homogeneity of the variance of the data was visually checked. All statistical
251 analyses were conducted in R (R Core Team, 2016). In the results we use statistic notation to show the
252 results of the ANOVA and linear regression models. We mention the F- value (ANOVA) or t-value
253 (linear regression), which indicates the difference of the explanatory variable to the variation in the data.
254 The p-value indicates the probability that the null-hypothesis is correct, we used a p-value of 0.05 as a
255 cut off to reject the null-hypothesis. The number in subscript indicates the degrees of freedom.

256 **3. Results**

257 **3.1 Dune characteristics**

258 Within the 8 hectare dune field we distinguished 434 polygons that were covered with dunes for at least
259 one moment during our mapping campaigns. Most of the dunes were *E. juncea* dunes (50.23%),
260 followed by *A. arenaria* dunes (28.11%) and mixed dunes (22.66%) in August 2016. Species
261 composition of the dunes changed along a gradient from sea to land. Close to the sea dunes were
262 vegetated by *E. juncea*, while, further from the sea, dunes were also vegetated by *A. arenaria* alone, or
263 in a mix with *E. juncea* (Fig. 2). Landward of the foredune dunes were also vegetated by *E. juncea*, *A.*
264 *arenaria* alone, or a mix of both species. The foredune bisecting our study area was mainly vegetated
265 with *A. arenaria*.



266 In August 2016 dune area, volume and maximum height differed significantly between dunes
267 differing in species composition (volume: $F_{2,428}=3.05$, $p=0.048$; max. height: $F_{2,428}=59.6$, $p < 0.001$), but
268 did not differ between dunes contrasting in shelter. Dunes with a mix of *E. juncea* and *A. arenaria* had
269 overall the highest volume and maximum height, whereas *E. juncea* dunes had the lowest volume and
270 height. *A. arenaria* dunes had the largest range in dune volume (Fig. 3A, B, C). For *E. juncea* dunes
271 seaward from the foredune the distance between dunes was higher compared to *A. arenaria* dunes and
272 dunes with both species ($F_{2,428}=52.5$, $p<0.001$), the distance between dunes landward from the foredune
273 was overall smaller than dunes seaward from the foredune (Fig. 3D, $F_{1,428}=70.2$, $p<0.001$). The dune
274 volume did not significantly differ between dunes seaward and landward from the foredune (volume:
275 $F_{1,428}=0.76$, $p=0.39$), but the dune height above NAP was significantly higher for dunes landward from
276 the foredune ($F_{1,428}=15.9$, $p<0.001$).

277 For the statistical model with relative change in dune volume as response variable, we had to
278 correct for species distribution and spatial autocorrelation. We created a grid, with blocks of 50 m x 50
279 m, and we selected dunes that fell within a block. In total, we selected 236 dunes, which consisted of
280 41.95% of *E. juncea* dunes, 36.02% of *A. arenaria* dunes, and 22.03% of dunes with both species.
281 These subset of dunes had an overall lower dunes size compared to all the dunes, but had overall similar
282 dune morphology and vegetation characteristics (Supplementary data S2).

283 # Figure 3 approximately here #

284 Vegetation characteristics depended on the plant species dominating the dunes and on the degree
285 of shelter. *E. juncea* dunes had significantly the lowest vegetation density, *A. arenaria* dunes the highest



286 and dunes which consisted of both species had an intermediate vegetation density (Fig. 3E, $F_{2,428}=49.30$,
287 $p<0.001$). Similar to vegetation density, *E. juncea* dunes also had the lowest maximum plant height,
288 whereas *A. arenaria* and dunes consisting of both species had the highest maximum plant height (Fig.
289 3F, $F_{2,428}=42.70$, $p<0.001$). Dunes landward from the foredune had significantly higher vegetation
290 densities compared to seaward dunes ($F_{1,428}=45.7$, $p<0.001$). There was no significant difference in
291 maximum plant height between dunes seaward and landward from the foredune ($F_{1,428}=0.41$, $p=0.52$).

292 3.2 Change in dune number and volume

293 The number of dunes within the measurement area changed over time, with dune numbers declining
294 over winter and increasing during summer. The degree of dynamics depended on season, species and
295 degree of sheltering.

296 3.2.1 Summer

297 Of the 434 dunes present in August 2016, 22.36% appeared over summer (April – August). Most of
298 these new dunes (65.93%) were *E. juncea* dunes, 31.87% were *A. arenaria* dunes and only 2.20% were
299 mixed dunes. Most (73.63%) new dunes developed seaward from the foredune and were quite small in
300 size with a volume of $2.72 \pm 0.29 \text{ m}^3$ (mean \pm SE). We assumed that most of these dunes established
301 over the growing season, as the orthomosaic showed a large amount of wrack line material (plant
302 material, woody debris, rope etc.) in their polygon in November and April. However we cannot exclude
303 that part of the large increase in the smaller *E. juncea* over summer is a result of their poor recognition
304 in November and April.

305



306 Over summer, most dunes increased in dune volume, including the foredune which increased
307 over summer with 0.28% per week, reaching a volume of 64,444 m³ in August. Only 4.15% of the
308 dunes showed a small decrease in the volume with a mean of -0.041 ± 0.014 m³/week. Changes in dune
309 volume were positively related to dune volume (Fig. 4A, $t\text{-value}_{430} = 57.20$, $p < 0.001$) and were higher
310 for dunes seaward of the foredune compared to dunes landward of the foredune, resulting in a
311 significant effect of shelter ($t\text{-value}_{430} = -41.70$, $p < 0.001$).

312 # Figure 4 approximately here #

313 Over summer, the relative change in dune volume was mainly influenced by sheltering, were the
314 relative change in dune volume was higher seaward of the foredune compared to landward of the
315 foredune (Fig. 5A). We found no significant difference in relative change in dune volume between
316 dunes with different species composition (Fig. 5A, Table 1). In our statistical model plant height had a
317 statistically significant effect on the relative dune growth. However, when tested in a single linear
318 mixed model with block as random intercept, plant height had a R^2 of 0.0038, thus hardly explaining
319 any variation in relative dune growth (Table 2). Several dune size variables were significant, but the
320 individual variation explained by dune volume, and dune height was very low, their R^2 ranging between
321 0.05 – 0.0033. The significant interactions between variables were mostly caused by the slight
322 correlations between the explanatory variables. The clustering of dunes did not significantly affect the
323 relative dune growth. We tested whether the effect of clustering was masked by the use of blocks as
324 random intercept, since the amount of clustering was different between the blocks. We re-analysed the
325 data without the blocks as random factor and found again no effect of clustering on the relative growth
326 rate of dunes.



327 # Table 1 & 2 approximately here #

328 3.2.2 Winter

329 Over winter (November – April) 7.85% of the 344 dunes disappeared, of which 40.74% were *E. juncea*
330 dunes, 55.56% were *A. arenaria* dunes and 3.70% dunes with both species. These dunes disappeared
331 both seaward (40.74%) and landward (59.26%) from the foredune and were overall quite small with an
332 average volume of $2.23 \pm 0.19 \text{ m}^3$.

333 Over winter dunes still increased in volume, the large foredune even increased with 0.22% per
334 week. However on average the changes in dune volume was less positive than over summer, 21.20% of
335 the dunes decreased -0.061 ± 0.015 (SE) m^3/week in volume, particularly seaward of the foredune.
336 25.00% of these decreased dunes were covered with *A. arenaria*, 50.00% with *E. juncea* and 25.00%
337 with both species. The absolute change in dune volume between November and April was positively
338 related to dune volume in November (Fig. 4B, $t\text{-value}_{430}=2.1$, $p=0.033$), but was only significant for
339 dunes landward of the foredune. Dunes seaward of the foredune showed no relationship between
340 absolute change in dune volume and the dune volume in November ($t\text{-value}_{430}=16.37$, $p<0.001$).

341 The relative change in dune volume was influenced by species composition and degree of shelter
342 (Table 1). Dunes with *E. juncea* increased relatively less in volume than *A. arenaria* dunes (Fig. 5B);
343 this effect was only significant for dunes seaward of the foredune. We found no significant relationship
344 between relative change in dune volume and vegetation density or maximum plant height (Table 2).
345 There was a significant interaction between vegetation density and sheltering by the foredune, which
346 could be related to the higher vegetation density at the dunes landward of the foredune. Dune volume,



347 the position relative to the foredune, had a significant negative effects on the relative change in dune
348 volume, whereas clustering had a positive significant effect, but the relationship was very weak (R^2
349 between 0.002 – 0.05).

350 3.3 Net dune growth

351 Net absolute dune growth per week over the whole observation period November – August was higher
352 at the seaward side of the foredune than at the sheltered landward side (slope seaward dunes: 0.37%,
353 slope landward dunes: 0.25%, dune volume*position from foredune: $t\text{-value}_{430} = -11.7$, $p < 0.001$).

354 Seaward dunes also had a slightly higher relative change in dune volume over November to August
355 compared to the landward dunes (seaward dunes: 0.27 ± 0.00009 (m^3/m^3)/week (means \pm SE), landward
356 dunes: 0.026 ± 0.0001 (m^3/m^3)/week, $F\text{-value}_{1,230} = 18.51$, $p < 0.001$).

357 4. Discussion

358 The aim of this study was to explore the contributions of vegetation and dune size to dune development
359 and how these effects are modified by the degree of shelter. We expected dune growth to be a function
360 of dune size and vegetation density, dune growth being the largest for big dunes with high vegetation
361 density. We also expected that the effect of sheltering on dune growth would depend on season:
362 exposed dunes growing faster in summer, but slower in winter. Our results show that the contribution of
363 vegetation and dune size depended on season and degree of shelter. In summer dune volume change
364 was explained by dune size and to a lesser extent by dune height, while species composition, vegetation
365 height or density had no effect. In winter dune volume change was explained by vegetation and dune
366 size, depending on the degree of shelter. Exposed dunes with sparsely growing *E. juncea* grew less in



367 volume than exposed dunes with densely growing *A. arenaria*. In contrast, growth of sheltered dunes
368 was a function of dune volume. These findings are the first to show the relative contribution of
369 vegetation and dune size for dune development over a winter and summer season, and these results can
370 be used for modelling coastal dune development.

371 **4.1 Dune size**

372 **4.1.1 Summer growth**

373 We found a positive linear relationship between dune volume and the absolute change in dune volume
374 over summer. It is known that dunes affect sedimentation by changing the wind flow patterns (Dong et
375 al., 2004; Li et al., 2008). Previous studies have found that with increased dune volume the area where
376 the wind speed is reduced increases, which result in higher sedimentation rates (Hesp, 1981; Hesp and
377 Smyth, 2017). The linear relationship between dune volume and dune growth found in our study
378 indicates that different dune sizes have similar effect on the wind flow pattern per unit of area. This
379 result has also been found in a modelling study by Walmsey and Howard (1985), who found that
380 different sized desert barchan dunes experienced similar disruptions of wind flow patterns, suggesting
381 similar relative rates of deposition and erosion.

382 Our study focussed on a relatively small size range. It is likely that the linear relationship
383 between dune volume change and dune size will saturate when dunes continue to grow and processes
384 other than wind speed reduction become important. The latter is supported by the volume change of the
385 low foredune bisecting our study area. Over summer the large foredune increased 0.28% per week in
386 volume, which is much lower than the overall increase of 0.81% per week of the dune seaward of



387 foredune. Therefore, we expect that there is a critical dune size at which the relationship between dune
388 volume and absolute dune growth is no longer linear. However, what exactly the critical dune size is, is
389 difficult to predict, it probably depends on multiple factors such as available sediment supply and
390 vegetation growth. The wind flow patterns are not only influenced by dune volume, but also by
391 maximum dune height (Walker and Nickling, 2002). In our study we found a significant, albeit weak
392 effect of the maximum dune height on the relative growth, suggesting differences in height did not have
393 a large effect on the wind flow pattern and the subsequent deposition of sand.

394 The positive linear relationship between dune volume and dune growth was modified by
395 sheltering; dunes landward of the foredune increased 0.60% per week less in volume than dunes
396 seaward of the foredune. This reduction in dune growth rate is likely the result of decreased sand supply
397 landward of the foredune; presumably a large amount of the sand was captured by the foredune as was
398 also observed for other foredunes (Arens, 1996). In our study the decrease in sand transport was less
399 sharp as observed by Arens (1996), however the difference in foredune sink strength between the
400 foredune in our study and those measured in Arens (1996) could be related to its smaller size, its
401 relatively low height and/or its sparse vegetation cover of 29% (Keijsers et al., 2015). Clustering of
402 dunes did not have any significant effect on the relative growth rate, which suggests that these smaller
403 dunes do not significantly reduce the sand supply to the landward situated dunes.

404 **4.1.2 Winter**

405 In winter dune size was only a good predictor for growth of the dunes occurring landward of the
406 foredune. For these sheltered dunes, growth again followed a linear relationship with dune volume. The



407 absence of a relationship between dune size and dune growth for the exposed dunes occurring seaward
408 from the foredune, suggests that dune erosion is less dependent on dune size than dune growth. Dune
409 erosion has mainly been attributed to wave run-up during storms (Haerens et al., 2012; Vellinga, 1982).
410 Therefore, it seems reasonable to assume that the degree of erosion depends on whether the dune can be
411 reached by high energy waves. Large dunes that are reached by high water levels can erode
412 substantially, whereas small dunes can have no erosion if they are protected by other dunes from the
413 high water.

414 Interestingly, the sheltered dunes had a slightly higher dune growth in winter compared to
415 summer. This increase in dune growth can perhaps be explained by more frequent and/or intensive
416 aeolian transport events during winter resulting into higher sand supply to the sheltered dunes.

417 **4.2 Vegetation characteristics**

418 Vegetation characteristics were a poor predictor of dune volume change over the summer period, but
419 were a significant predictor for dune volume change over winter. Over summer dune growth did not
420 differ between dunes covered by different dune building plant species when corrected for dune size.
421 Similarly we did not find a clear effect of vegetation density and plant height on dune growth. This
422 results contrast with other studies that report a significant difference in the ability of species to trap sand
423 mediated by differences in shoot density and cover (Keijsers et al., 2015; Zarnetske et al., 2012).
424 Perhaps the discrepancy with our study can be explained by the differences in spatial scale used
425 between studies. We studied dune volume change at the scale of a dune, whereas the other studies
426 focussed on the scale of the vegetation patch, where species specific effects are probably more



427 pronounced than at the scale of the whole dune. Also Al-Awadhi and Al-Dousari (2013) found that the
428 effects of vegetation on dune growth are scale dependent for coastal dunes. They found that the linear
429 relationship between shrub vegetation characteristics and dune morphology levels off for bigger dunes.
430 In our statistical models we selected the smaller dunes, which was a consequence of only selecting
431 dunes that were located within one block. However even for these smaller dunes vegetation had no
432 significant effect on relative dune growth.

433 Over winter *E. juncea* dunes had a significantly lower relative growth rate than *A. arenaria*
434 dunes, presumably because of their higher sensitivity to erosion. This species-effect might be related to
435 the sparser growth form of *E. juncea* in comparison to *A. arenaria* as dense vegetation has been found
436 to reduce the amount of dune erosion, by more effective wave attenuation (Charbonneau et al., 2017;
437 Koch et al., 2009; Silva et al., 2016). However, the effect of vegetation density was not significant in
438 our model suggesting that the species effect might be due to other species differences, such as
439 differences in rooting pattern. Another explanation is that the vegetation density measurement did not
440 reflect the real vegetation density, *E. juncea* was difficult to detect due to the low NDVI values. The
441 species effect was only significant for dunes situated at the exposed, seaward side of the foredune where
442 erosion by water likely occurred during the single storm covered by our study period. Despite being
443 statistically significant, the differences in relative growth rate between exposed *A. arenaria* and *E.*
444 *juncea* dunes was not very large. Nevertheless the species effect might become more pronounced with
445 higher erosion pressure during more stormy winters (Charbonneau et al., 2017).



446 Interestingly, our species did show differences in dunes size. On average, *A. arenaria* dunes
447 were higher than *E. juncea* dunes, that were broader (Bakker, 1976; Zarnetske et al., 2012). This
448 difference in dune morphology suggests a higher sand catching efficiency of *A. arenaria* that might be
449 masked by using dune volume, mean height or dune area as explanatory variables. We explored whether
450 there is an effect of species composition on the change in maximum dune height over summer, but
451 found no consistent effect. Perhaps the difference in dune morphology could be a result of differences in
452 erosion between the dune types over winter.

453 **4.3 Implication for dune development**

454 **4.3.1 Net dune growth**

455 Exposed dunes had an overall higher net growth compared to sheltered dunes, indicating that summer
456 growth offset winter erosion in our study period which was characterised by an average summer and
457 calm winter. This balance might have been different if winter conditions had been more severe.

458 During winter, storms determine the erosion of dunes seaward of the foredune. Multiple low
459 intensity storms can lead to more erosion than one high intensity storm (Dissanayake et al., 2015;
460 Ferreira, 2006; van Puijenbroek et al., 2017). Whether exposed dunes have a higher net dune growth
461 compared to dunes landward from the foredune depends mainly on the storm intensity and frequency. A
462 single high intensity storm can erode all the sand that exposed dunes have accumulated over a whole
463 summer, and in such case sheltered dunes could have a higher growth rate than the exposed dunes. The
464 exact relative growth rate over summer depends on the number of aeolian transport events. Linking the



465 number of aeolian transport event to the relative growth rate over summer would be a worthwhile
466 avenue for future research.

467 Sand supply and storm intensity are also affected by local conditions as beach morphology. A
468 minimum beach width is needed to reach maximum aeolian transport, the fetch length (Delgado-
469 Fernandez, 2010; Dong et al., 2004; Shao and Raupach, 1992). Our study site had a wide beach (0.9 km
470 wide), and we assume that the maximum aeolian transport was reached. The net growth of our foredune
471 was approximately 30 m^3 per m foredune parallel to the sea for a period of 10 months. This growth rate
472 does also occur at other places along the Dutch coast, but is not very common (Keijsers et al., 2014).
473 Storm intensity is also influenced by beach morphology. The presence of intertidal bars and a wide
474 beach can reduce the storm intensity by wave attenuation (Anthony, 2013; Ruggiero et al., 2004).
475 Therefore we can assume that the net dune growth we found in our study will depend on the beach
476 morphology. On smaller beaches we expect the net dune growth to be lower compared to wider
477 beaches, due to the lower sand supply by reduced fetch length and higher storm erosion of dune (van
478 Puijenbroek et al., 2017)

479 **4.4.2 Vegetation**

480 For coastal dune development vegetation is essential, however the species-composition of the
481 vegetation seems less important than we assumed: species did not seem to affect dune growth over the
482 summer, but did affect dune growth over winter.

483 We did find differences in dune morphology between the species, which indicates an effect of
484 species composition on dune morphology. However, this difference in dune morphology is probably



485 also caused by vegetation succession. In Western Europe, the primary succession of coastal dunes is
486 generally assumed to start with *E. juncea*. Only after a fresh water lens has developed in the *E. juncea*
487 dune, *A. arenaria* will establish (Westhoff et al., 1970). Over time *A. arenaria* will outcompete *E.*
488 *juncea*. This assumed succession pathway matches part of the spatial patterns that we found in our study
489 site and explains why dunes with only *E. juncea* are relatively small. Over time these small dunes merge
490 together after which *A. arenaria* is assumed to establish. However, we found that *A. arenaria* has a
491 large range in dune volume suggesting that, contrary to current assumptions, *A. arenaria* can also
492 establish on the bare beach without *E. juncea*, as long as the soil salinity is not too high.

493 At our study site only two dune building species occur, however there are many different dune-
494 building species. It could very well be that other dune building species do have a significant effect on
495 the dune growth over summer. For further research it would be interesting to study if these results are
496 similar in another dune system with different plant species.

497 **4.4.4 Application**

498 To our knowledge, we are the first to report on the relationship between dune volume and dune growth
499 in the field. The linear relationship that we found in our studies can be incorporated in mathematical
500 models that predict dune development. Furthermore, our research shows that for predicting dune growth
501 species identity does not matter during the summer, however it does matter during the winter. This
502 indicates that for dune building models, species identity is especially important when winter survival of
503 dunes is modelled. Furthermore, for the construction of an artificial dune it appears to be crucial to plant
504 the more storm resistant species.



505 Despite the presence of smaller dunes seaward of the foredune, the foredune showed a large
506 increase in volume compared to similar foredunes along the Dutch coast. This indicates that sand supply
507 to the foredune was not seriously hampered by the presence of the small vegetated dunes, while the
508 smaller dunes seaward of the foredune likely added to the protection of the foredune against storm
509 erosion. For coastal management it could be beneficial for foredune growth to have embryo dunes
510 seaward of the foredune given a high sand supply.

511 **5. Conclusions**

512 The purpose of this study was to explore the contribution of vegetation and dune size on dune
513 development at locations differing in shelter from the sea. Our results show that 1) the contribution of
514 vegetation and dune size depend on season and degree of shelter. 2) Vegetation does not affect dune
515 growth over summer, but does affect dune growth during winter, particularly at exposed sites. 3) During
516 early dune development, dune growth is linearly related to dune volume, whereas dune volume does not
517 seem to matter for dune erosion. 4) Sheltering by a foredune reduces both sand supply and dune
518 erosion; the net effect of shelter on dune growth therefore likely depends on beach morphology and
519 weather conditions. These results can be incorporated in models predicting dune development and can
520 be used by managers to determine coastal safety.

521

522 **Acknowledgements** We thank Ministry of Defence and Staatsbosbeheer to allow UAV flights in their
523 nature area. We thank the technology foundation STW (grant number STW 12689 S4) for funding the
524 NatureCoast project, which made this research possible.



525 **Competing interests** The authors declare that they have no conflict of interest.

526 **Data availability** Data will be made available through the 4TU.Center for Research Data

527 (<http://researchdata.4tu.nl/home/>)

528 **Author contributions** MvP, CN and JS performed UAV flights and image calibration. MvP, CN and
529 JL analysed the data. MvP, CN, JS, AdG, MR, FB and JL provided guidance on the scope and design of
530 the project, and contributed to the writing of the manuscript.

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668 **Supporting information**

669 Additional supporting information can be found in the online version of this article:

670 **Supplement S1** Weather conditions in our study site for 2013 - 2016

671 **Supplement S2** Dune morphology of selected dunes



672 **Table 1.** Statistical models for the relative change in dune volume between April – August (summer)
 673 and November – April (winter). In this model we tested the effect of species, dune size, and degree of
 674 sheltering. The data was analysed with a general linear mixed model with blocks as random intercept.
 675 The standardized estimates and level of significance are shown for the models. Model selection was
 676 performed with AIC (Akaike information criterion) as selection criteria. Marginal R^2 is the variation
 677 explained by the fixed factors, whereas the conditional R^2 is the variation explained by the fixed and
 678 random factors.

<i>Model with species</i>	<i>Dependent variable:</i>			
	Relative change in dune volume			
	<i>Summer</i>		<i>Winter</i>	
	Full model	Model selection	Full model	Model selection
Main effects				
Intercept	1.18** *	1.17***	0.92** *	0.94***
<i>E. juncea</i>	-0.02		0.005	-0.02**
Mix	0.02		0.02	-0.003
Dune volume	6.10	8.27***	-6.0*	-3.43**
Clustering	-0.22	-0.18	0.22	0.23
Max. dune height	-0.25	-0.31*	0.15	0.087
Sheltering by foredunes	0.29*	0.31**	-0.31**	-0.31**
Interaction effects				
<i>E. juncea</i> * Dune volume	0.90		1.90	
Mix * Dune volume	-0.11		1.41	
<i>E. juncea</i> * clustering	0.11		0.04	
Mix * clustering	0.01		-0.006	
<i>E. juncea</i> * max. dune height	-0.08		-0.09	
Mix * max. dune height	-0.02		-0.033	
<i>E. juncea</i> * Shel. by foredune	-0.05		0.03	
Mix * Shel. by foredune	-0.02		0.001	
Dune volume * clustering	-4.64*	-5.65**	4.44**	4.10**
Dune volume * max. dune height	-1.16	-2.01*	0.62	
Dune volume * Shel. by foredune	1.85	2.00*	-1.11	-1.31*
Clustering * max. dune height	0.31	0.34*	-0.29	-0.27*
Clustering * Shel. by foredune	-0.12	-0.17*	0.12	0.13
Max. dune height * Shel. by foredune	-0.20*	-0.18*	0.19**	0.19**
Marginal R^2	0.31	0.31	0.25	0.23
Conditional R^2	0.34	0.33	0.39	0.39
Observations	236	236	236	236



Akaike Inf. Crit.	-632.60	-685.45	-673.10	-709.11
Bayesian Inf. Crit.	-555.08	-641.04	-595.57	-661.35

Note:

* $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$

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681 **Table 2.** Statistical models for the relative change in dune volume between April – August (summer)
 682 and November – April (winter). In this model we tested the effect of vegetation characteristics, dune
 683 size and degree of sheltering. The data was analysed with a general linear mixed model with blocks as
 684 random intercept. The standardized estimates and significance values are shown for the models. Model
 685 selection was performed with AIC as selection criteria. Marginal R^2 is the variation explained by the
 686 fixed factors, whereas the conditional R^2 is the variation explained by the fixed and random factors.

<i>Model with vegetation characteristics</i>	<i>Dependent variable:</i>			
	Relative change in dune volume			
	<i>Summer</i>		<i>Winter</i>	
	Full model	Model selection	Full model	Model selection
Main effects				
Intercept	1.24***	1.24***	0.90***	0.81***
Vegetation density	-0.003		-0.05	-0.03
Max. plant height	0.15	0.14**	0.04	
Dune volume	8.65***	6.62***	-2.72	-3.67**
Clustering	-0.21	-0.23	0.29	0.40**
Max. dune height	-0.44*	-0.41**	0.07	0.17
Sheltering by foredune	0.26*	0.29*	-0.28*	-0.25**
Veg. density * max. plant height	-0.01		0.001	
Veg. density * dune volume	0.83		0.92	
Veg. density * clustering	-0.03		0.078	0.06
Veg. density * max. dune height	0.04		-0.03	
Veg. density * Shel. by foredune	-0.005		-0.03	-0.04**
Max. plant height * dune volume	-0.58		-0.19	
Max. plant height * Clustering	0.02		-0.06	
Max. plant height * max. dune height	-0.11	-0.10**	0.04	
Max. plant height * Shel. by foredune	0.004		-0.01	
Dune volume * clustering	-6.37**	-6.30***	4.51**	4.65***
Dune volume * max. dune height	-1.54		-1.11	
Dune volume * Shel. by foredune	1.63	1.95*	-2.23*	-1.82**
Clustering * max. dune height	0.40*	0.41**	-0.32	-0.42**
Clustering * Shel. by foredune	-0.15	-0.17*	0.05	



Max. dune height * Shel. by foredune	-0.16	-0.16*	0.28**	0.31***
Marginal R ²	0.33	0.31	0.24	0.21
Conditional R ²	0.37	0.35	0.42	0.40
Observations	236	236	236	236
Akaike Inf. Crit.	-622.85	-674.05	-656.46	-704.97
Bayesian Inf. Crit.	-542.07	-626.28	-575.68	-657.20

Note:

* p<0.05; ** p<0.01; *** p<0.001

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689 Figure captions

690 **Fig. 1** Overview of the Hors on Texel, the Netherlands. The white lines show the flight path for the four
691 different flights. The points show the position of the ground control markers. The white polygon is the
692 monitoring area, which is 200 m x 400 m.

693 **Fig. 2** Overview of the monitoring area. A) The elevation is shown with the Digital Terrain Model (m
694 NAP), the green pixel indicates grass cover and the polygons indicate the dunes. B) The colour
695 indicates the species present on the dune and the squares the blocks. The foredune in the middle of the
696 monitoring area is excluded from the statistical analysis. Some dunes that were cut-off by the edge of
697 the DTM, we discarded these dunes.

698 **Fig. 3** Different dune characteristics for dunes in August with *A. arenaria*, *E. juncea* and a mix of both
699 species separated for dunes seaward and landward of the foredune: A) Dune area, B) Maximum dune
700 height, C) Dune volume, D) Clustering: mean height around a 25m radius around the dune, E)
701 Vegetation density, F) Plant height. The letters denote the significant difference between the bars.
702 Seaward of the foredune there were 41 *A. arenaria* dunes, 198 *E. juncea* dunes, and 53 dunes with both
703 species, landward of the foredune there were 81 *A. arenaria* dunes, 23 *E. juncea* dunes, and 41 dunes
704 with both species. NAP refers to Amsterdam Ordnance Date, which refers to mean sea level near
705 Amsterdam

706 **Fig. 4** The relationship between dune volume (m^3) and the absolute change in dune volume (m^3 /week)
707 for: A) summer (April – August); B) winter (November – April). The data is shown for dunes seaward
708 and landward of the foredune. The black line shows the regression prediction, the grey dashed line the
709 95% confidence interval. The formulas are the result of a linear regression model.

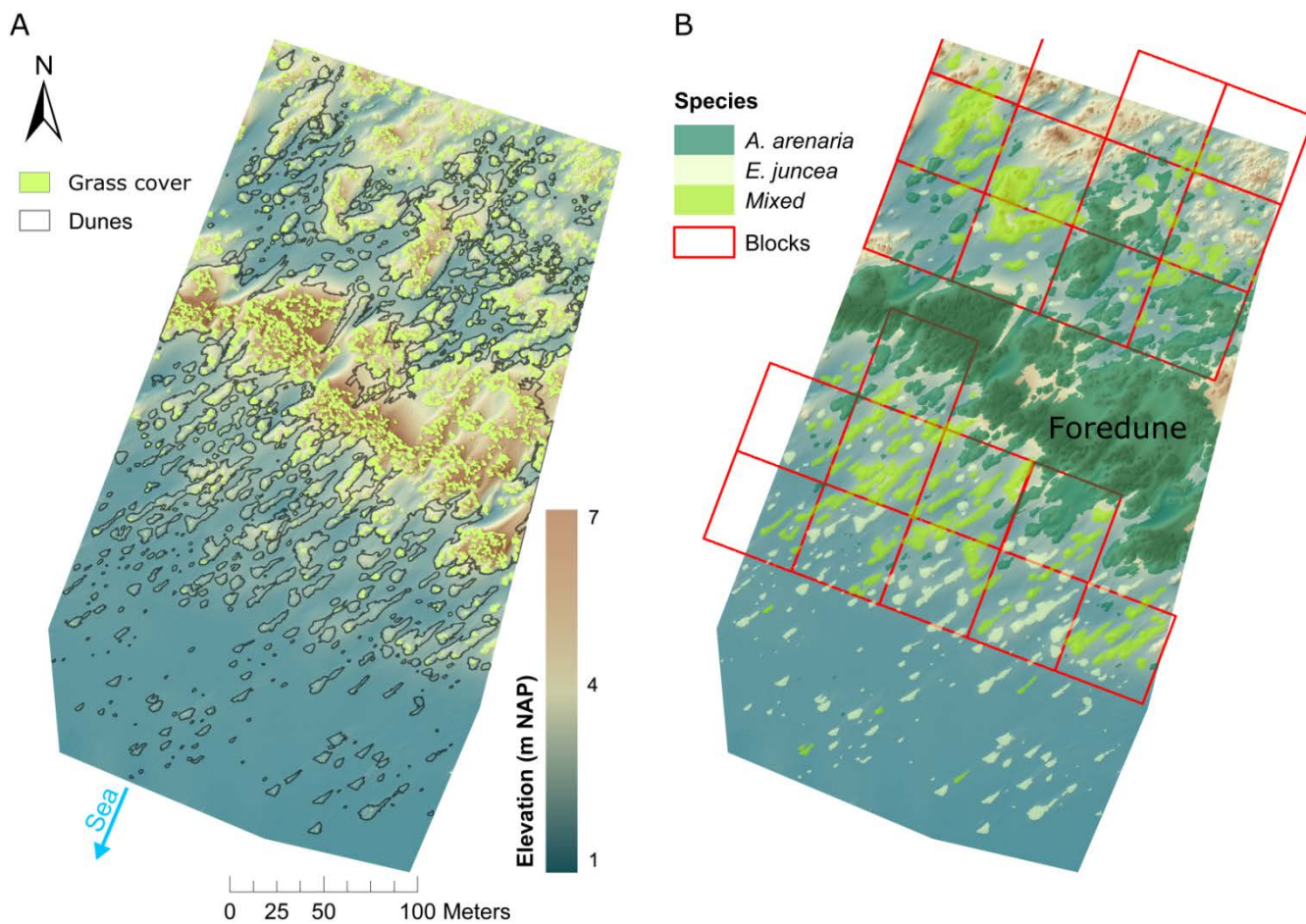
710 **Fig. 5** Relative change in dune volume (m^3/m^3)/week for dunes with *A. arenaria*, *E. juncea* and a mix of
711 both species and separated for dunes seaward and landward of the foredune for: A) summer, April –
712 August; B) winter, November – April. The letters denote the significant difference between the bars.
713 Seaward of the foredune there were 28 *A. arenaria* dunes, 77 *E. juncea* dunes, and 28 dunes with both
714 species, landward from the foredune there were 57 *A. arenaria* dunes, 22 *E. juncea* dunes, and 25 dunes
715 with both species.

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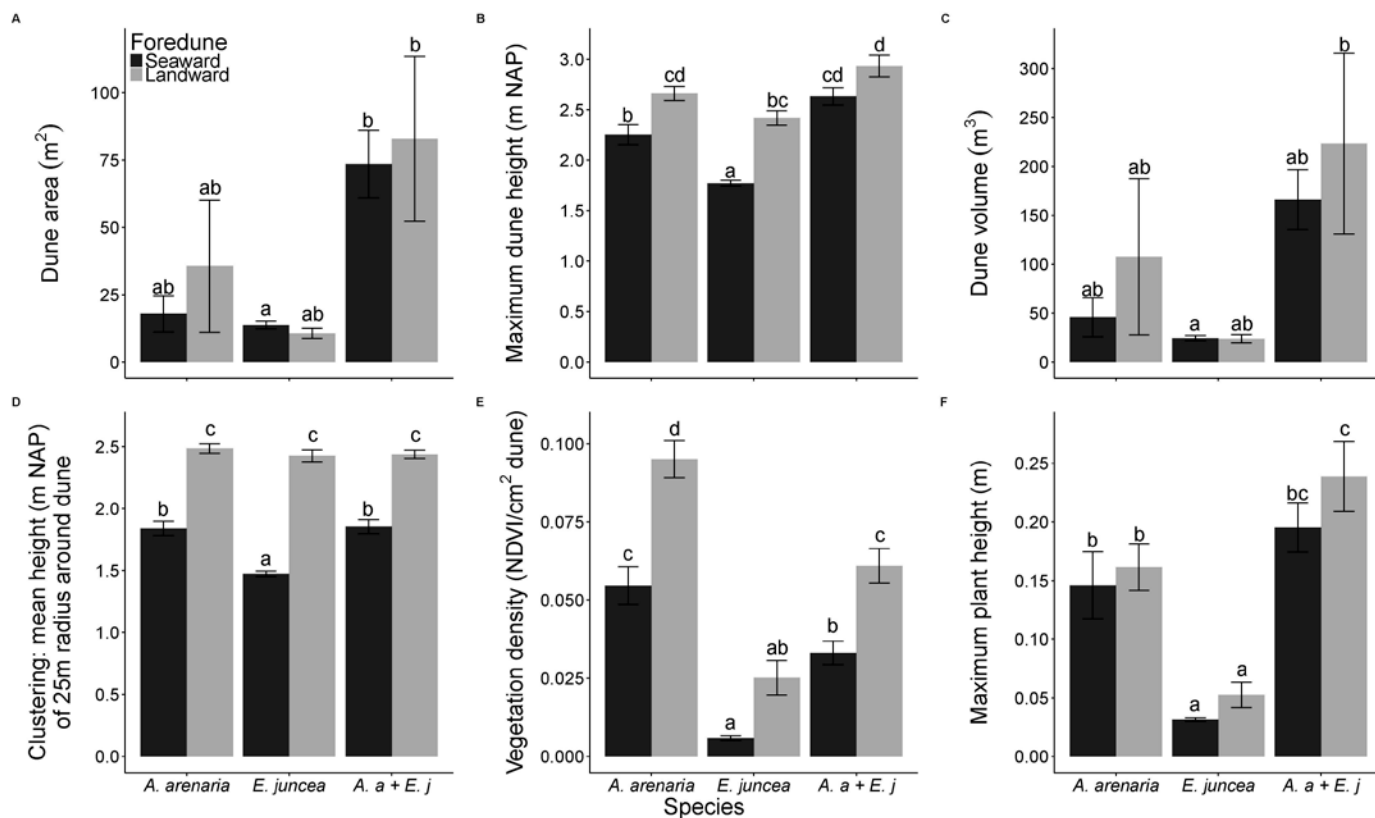






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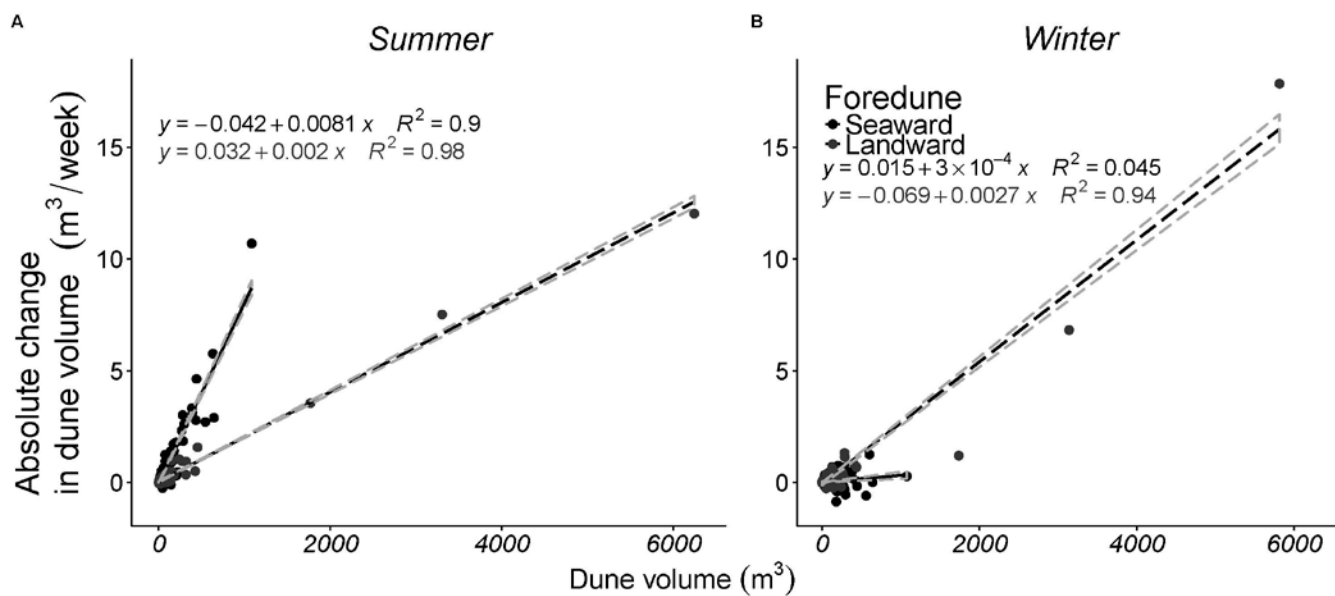


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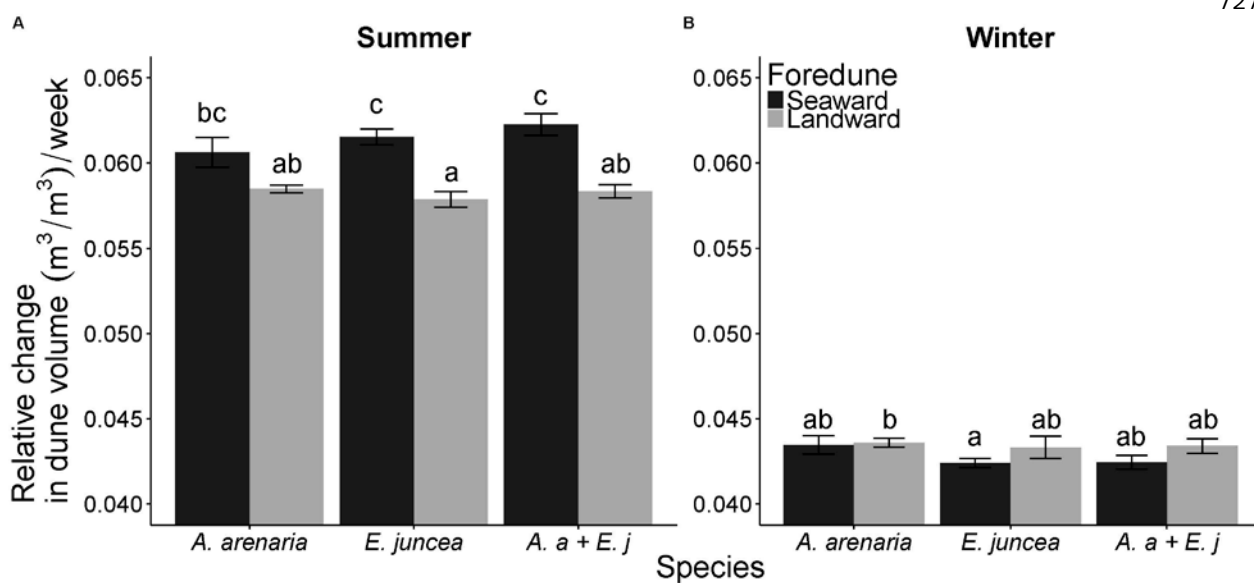




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