

1 **Retrogressive thaw slumps temper dissolved organic carbon delivery to streams of the Peel Plateau,**

2 **NWT, Canada**

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4 Cara A. Bulger¹, Suzanne E. Tank¹, and Steven V. Kokelj²

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6 ¹Department of Biological Sciences, University of Alberta, Edmonton, AB, T6G 2E9, Canada

7 ²Northwest Territories Geological Survey, Government of the Northwest Territories, Yellowknife, NT,

8 X1A 1K3, Canada

9 *Correspondence to:* Cara A. Bulger (cara.bulger@gmail.com)

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11

12 **Abstract**

13 In Siberia and Alaska, permafrost thaw has been associated with significant increases in the delivery of
14 dissolved organic carbon (DOC) to recipient stream ecosystems. Here, we examine the effect of
15 retrogressive thaw slumps (RTS) on DOC concentration and transport, using data from eight RTS features
16 on the Peel Plateau, NT, Canada. Like extensive regions of northwestern Canada, the Peel Plateau is
17 comprised of thick, ice-rich tills that were deposited at the margins of the Laurentide Ice Sheet. RTS
18 features are now widespread in this region, with headwall exposures up to 30 m high, and total
19 disturbed areas often exceeding 30 ha. We find that intensive slumping on the Peel Plateau is universally
20 associated with decreasing DOC concentrations downstream of slumps, even though the composition of
21 slump-derived dissolved organic matter (DOM; assessed using specific UV absorbance and slope ratios)
22 is similar to permafrost-derived DOM from other regions. Comparisons of upstream and downstream
23 DOC flux relative to fluxes of total suspended solids suggest that the substantial fine-grained sediments
24 released by RTS features may sequester DOC. Runoff obtained directly from slump rillwater, above entry
25 into recipient streams, indicates that the deepest RTS features, which thaw the greatest extent of
26 buried, Pleistocene-aged glacial tills, release low concentration DOC when compared to paired
27 upstream, un-disturbed locations, while shallower features, with exposures that are more limited to a
28 relict Holocene active layer, have within-slump DOC concentrations more similar to upstream sites.
29 Finally, fine-scale work at a single RTS site indicates that temperature and precipitation serve as primary
30 environmental controls on above-slump and below-slump DOC flux, but that the relationship between
31 climatic parameters and DOC flux is complex for these dynamic thermokarst features. These results
32 demonstrate that we should expect clear variation in thermokarst-associated DOC mobilization across
33 Arctic regions, but that within-region variation in thermokarst intensity and landscape composition is
34 also important for determining the biogeochemical response. Geological and climate legacy shape the
35 physical and chemical composition of permafrost, and thermokarst potential. As such, these factors
36 must be considered in predictions of land-to-water carbon mobilization in a warming Arctic.

37 **1. Introduction**

38 Anthropogenic climate change is significantly affecting the Arctic cryosphere (IPCC, 2014).
39 Temperature increases in circumpolar regions are predicted to be at least 40 % greater than the global
40 mean, while precipitation is also expected to increase significantly in most locations (IPCC, 2014). The
41 resulting degradation of permafrost is forecast to have wide-ranging effects, because thawing has the
42 potential to greatly alter the physical, chemical, and biological functioning of landscapes (Frey and
43 McClelland, 2009; Khvorostyanov et al., 2008a, 2008b; Kokelj et al., 2017b; Schuur et al., 2008, 2013). In
44 particular, permafrost acts as a long term storage medium for solutes and sediments, and as a barrier to
45 the participation of permafrost-sequestered constituents within active biogeochemical cycles (Frey and
46 McClelland 209; Vonk et al. 2015b). Consequently, permafrost thaw can enhance linkages between
47 terrestrial and aquatic systems, via increased transport of terrestrial compounds from land to water
48 (Kokelj et al. 2013; Tanski et al., 2016; Vonk et al., 2015b). Given that circumpolar stores of permafrost
49 carbon are estimated to be almost double that of the atmospheric carbon pool (Hugelius et al., 2014),
50 there is great potential for large increases in carbon mobilization as a result of permafrost thaw (Schuur
51 et al., 2015). Within this context, the mobilization of dissolved organic carbon (DOC) from previously
52 frozen soils is of particular interest, because DOC acts as the primary substrate for the microbially-
53 mediated mineralization of organic carbon to carbon dioxide (Battin et al., 2008), and serves as the
54 primary vehicle for the delivery of terrestrial carbon to the Arctic Ocean (Dittmar and Kattner, 2003;
55 Holmes et al., 2012; Spencer et al., 2015). As a result, the implications of thaw-mediated DOC
56 mobilization may range from effects on the permafrost-carbon feedback, to the ecological and
57 biogeochemical functioning of streams, rivers, and the nearshore ocean (e.g. Fritz et al. 2017; Tank et
58 al., 2012b; Vonk et al., 2015b).

59 Permafrost thaw can manifest in many different forms, ranging from an increase in active layer
60 thickness and terrain subsidence, to thermokarst features that significantly reconfigure the physical
61 structure of landscapes (Kokelj and Jorgenson, 2013). Of these, thermokarst has the potential to rapidly

62 expose significant quantities of previously-frozen soils to biological and chemical processing (Abbott et
63 al., 2014, 2015; Malone et al., 2013; Tanski et al. 2017). One of the most conspicuous manifestations of
64 thermokarst is the retrogressive thaw slump (RTS; Fig. 1), which develops as a result of mass wasting in
65 ice-rich glacial deposits across northwestern Canada, Alaska, and western Siberia (Kokelj et al., 2017b),
66 and in Yedoma regions of Alaska and Siberia (Murton et al., 2017). Thaw slumps are widespread
67 throughout glaciated terrain in the western Canadian Arctic (Kokelj et al., 2017b; Lantuit et al. 2012),
68 including on the Peel Plateau (Lacelle et al., 2015). These dynamic landforms develop via the ablation of
69 an ice-rich headwall and are particularly efficient at thawing thick zones of ice-rich permafrost and
70 translocating large volumes of sediment from slopes to downstream environments (see Fig. 1). RTS
71 features remain active for decades (Lantuit et al. 2012). They typically stabilize following sediment
72 accumulation at the base of the headwall (Kokelj et al., 2015), but can reactivate causing thaw within
73 the scar zone and upslope expansion of the disturbance (Kokelj et al., 2013; Lantuit and Pollard, 2008).
74 During periods of activity, thawed materials accumulate as a saturated slurry in the slump scar zone (see
75 Fig. 1b) and are transported downslope by mass flow processes, which are accelerated by meltwater-
76 and rainfall-induced saturation (Kokelj et al. 2015). Surface runoff can also remove solutes and
77 suspended sediment from the thawed substrate to downstream environments. Although variation in
78 temperature, precipitation and solar radiation have been correlated with development rates and growth
79 of RTS features (Kokelj et al., 2009, 2013, 2015; Lacelle et al., 2010; Lewkowicz, 1986, 1987), we know
80 little about how these and other environmental drivers might control permafrost-DOC dynamics at the
81 individual-slump to small watershed scale.

82 On the Peel Plateau, an individual thaw slump can impact tens of hectares of terrain, displace
83 hundreds of thousands of cubic meters of sediments, and significantly alter surface water sediment and
84 solute loads (Kokelj et al., 2013; Malone et al., 2013), and thus downstream ecosystems (Chin et al.,
85 2016; Malone et al., 2013). The magnitude of these disturbances and their cumulative impacts is great
86 enough to alter solute loads in the Peel River (70,000 km² watershed area; Kokelj et al., 2013), even

87 though only a small portion of that river's total catchment area (<1%) is influenced by thermokarst
88 (Kokelj et al., 2017b; Segal et al., 2016). This contrasts with many other thaw-affected regions, where
89 increases in solute loads following permafrost disturbance can be transient and have little overall effect
90 on annual solute fluxes (e.g., in High Arctic regions affected by active layer detachments; Lafrenière &
91 Lamoureux, 2013). In addition, permafrost thaw on the Peel Plateau is notable in that it exposes vast
92 quantities of mineral-rich glacial till, which is overlain by a relatively shallow layer of slightly more
93 organic-rich soils (Duk-Rodkin and Hughes, 1992; Kokelj et al. 2017a). Although this till-associated, RTS-
94 susceptible landscape type is found across the Laurentide and Barents-Kara glacial margins of Canada,
95 Alaska, and Siberia (Kokelj et al. 2017b), it contrasts with regions of Alaska and eastern Siberia that are
96 either Yedoma-rich or were covered by patchy or thin drift during the late Pleistocene, and have been a
97 focus for study of permafrost-DOC interactions to date (Abbott et al., 2014, 2015; Drake et al., 2015;
98 Mann et al., 2012; Vonk et al., 2013b).

99 Thermokarst has been documented to enhance DOC concentrations in recipient aquatic
100 ecosystems in several Arctic regions (Frey and McClelland, 2009; Tank et al., 2012a; Vonk et al., 2013a;
101 Vonk and Gustafsson, 2013). In Alaska, streams affected by thaw slumps have higher DOC
102 concentrations than un-affected systems across various terrain types (2-3 fold increase; Abbot et al.,
103 2014), while in eastern Siberia the DOC concentration in runoff from thawing Yedoma is considerably
104 greater than concentrations in recipient river systems (~30-fold elevation; Spencer et al. 2015).
105 However, multiple factors, including variable carbon content in permafrost soils (Hugelis et al. 2014) and
106 variation in ground ice type and volume (Fritz et al. 2015) may affect DOC release from permafrost. In
107 regions where thermokarst transports fine-grained sediments to aquatic systems, sorption processes
108 may also be important, because dissolved organic matter (DOM) can readily sorb to mineral soils (e.g.,
109 Kothawala et al. 2009). Sorption to mineral sediments can cause DOM to be rapidly removed from
110 solution in stream systems (Kaiser and Guggenberger, 2000; Kothawala et al. 2009; McDowell, 1985),
111 while enabling the downstream transport and continued sequestration of organic carbon (Hedges et al.,

112 1997). This process may be particularly important for regulating DOC dynamics in glacial margin
113 landscapes, where a predisposition to thaw slumping results in an abundance of thermokarst-related
114 slope disturbances which mobilize fine-grained glacial sediment stores to downstream systems (Kokelj
115 et al., 2017a, 2017b; Lantuit et al. 2012; Rampton, 1988). Despite this, we know little about the
116 downstream consequences of permafrost thaw for carbon biogeochemistry in till-dominated glacial
117 landscapes, which are emerging as some of the most geomorphically dynamic permafrost environments
118 in the circumpolar Arctic.

119 The objective of this study was to quantify how RTS features affect the concentration and
120 composition of DOC across a series of slump-affected streams on the Peel Plateau, and to examine how
121 observed variation in slump morphometry affects DOC dynamics in downstream environments. We
122 further investigated how short-term variation in precipitation, temperature, and solar radiation affect
123 DOC delivery from land to water, using measurements of DOC flux above and below a single RTS feature.
124 We targeted the thermokarst-sensitive Peel Plateau for this work, which is characteristic of till-rich,
125 glacial margin landscapes throughout Canada, Alaska, and Siberia (Kokelj et al. 2017b). By comparing our
126 results to those from elsewhere, we highlight how broad variation in permafrost soil composition,
127 permafrost genesis, and Quaternary history may drive variation in land-freshwater DOC dynamics across
128 divergent regions of the warming circumpolar Arctic.

129

130

131 **2 Study Site**

132 *2.1 General study site description*

133 Our study was conducted on the Peel Plateau, situated in the eastern foothills of the Richardson
134 Mountains, NWT, Canada, in the zone of continuous permafrost (Fig. 1a). The fluvially-incised Plateau
135 ranges in elevation from 100 to 650 masl. The region was covered by the Laurentide Ice Sheet (LIS) for a
136 brief period (a maximum of 2,000-3,000 years) 18,500 cal yr BP (Lacelle et al., 2013). The bedrock of the

137 region is Lower Cretaceous marine shale from the Arctic River formation (Norris, 1984) and siltstone
138 overlain by Late Pleistocene glacial, glacio-fluvial and glacio-lacustrine sediments (Duk-Rodkin and
139 Hughes, 1992). These Pleistocene deposits host ice-rich permafrost, overlain by a shallow and commonly
140 organic-rich active layer. Radiocarbon dating in the region has placed the age of relict ground ice in the
141 late Pleistocene ($18,100 \pm 60$ ¹⁴Cyr BP; Lacelle et al., 2013). Upper layers of permafrost thawed during
142 the early Holocene and host younger, Holocene-aged organic materials (Lacelle et al., 2013). These are
143 distinguished from deeper Pleistocene-aged permafrost by a thaw unconformity (Burn 1997; Fig. 1),
144 which developed when warmer climate during the early Holocene prompted the thawing of near-
145 surface permafrost. The regional increase in active layer thickness integrated organic matter into the
146 thawed soils and enabled the leaching of soluble ions (see Fig. 1c-d). Climate cooling and permafrost
147 aggradation have archived this notable stratigraphic variation in geochemistry, organic matter content,
148 and cryostructure (Burn 1997; Fritz et al. 2012; Kokelj et al., 2002; Lacelle et al., 2014; Murton and
149 French, 1994).

150 Ice-marginal glacial landscapes such as the Peel Plateau host thick layers of ice-rich
151 sediments, and thus have a predisposed sensitivity to climate-driven thaw slump activity (Kokelj et al.,
152 2017). On the Peel Plateau, slumping is largely constrained by the maximum extent of the LIS, because
153 the thick layers of ice-rich permafrost necessary for RTS activity are typically not present beyond the
154 glacial limits (Lacelle et al., 2015). Fluvial incision provides the topographic gradient necessary for thaw
155 slump development and RTS features are common; ranging in size from numerous small features, to
156 those greater than 20 ha, which are rare (<5% prevalence; Lacelle et al., 2015). The recent intensification
157 of slumping on the Peel Plateau is driven in part by increasing air temperatures and summer rainfall
158 intensity (Kokelj et al., 2015). This intensification is also increasing the thaw of the deepest layer of ice-
159 rich, organic-poor, Pleistocene-aged glacial tills that underlie this region. The pattern of abundant
160 thaw slump development across ice-marginal glaciated permafrost landscapes extends from the Peel
161 Plateau across the western Canadian Arctic, and persists at continental scales (Kokelj et al., 2017b).

162

163 *2.2 Regional climate*

164 The regional climate is typical of the subarctic with long, cold winters and short, cool summers.

165 Mean annual air temperature (1981-2010) at the Fort McPherson weather station (Fig. 1a) is -7.3 °C

166 with average summer (June-August) temperatures of 13.3 °C (Environment Canada, 2015). A warming

167 trend of 0.77 °C per decade since 1970 has been recorded; however these increases are most apparent

168 in the winter months (Burn and Kokelj, 2009). Our sample period spanned the thaw months of July and

169 August; average 1981-2010 temperatures for those months, recorded at Fort McPherson, are 15.2 and

170 11.8 °C, respectively, similar to temperatures at Fort McPherson during 2014 (15.6 and 11.6 °C), but

171 slightly higher than 2014 averages observed at a recently established meteorological station on the Peel

172 Plateau (Fig. 1a; 13.2 °C in July and 9.5 °C in August). Annual cumulative rainfall (1981-2010) at Fort

173 McPherson averages 145.9 mm, with July and August having the highest rainfall levels at 46.4 and 39.1

174 mm (Environment Canada, 2015). In 2014, rainfall for July and August was 71 and 121 mm at Fort

175 McPherson, and 128.7 and 170.7 mm on the Peel Plateau. This continues the trend for this region of

176 increasingly wet summers with numerous extreme rainfall events (Kokelj et al., 2015).

177

178 **3 Methods**

179 *3.1 Slump site selection*

180 Eight RTS features were selected from across the study region, using aerial surveys and previous

181 knowledge of the region (Fig. 1; Fig. S1; Table 1). Selected slumps were characterized by a debris tongue

182 that connected the slump to the valley bottom and directly impacted a stream system. Sampling at each

183 slump occurred at three discrete locations: upstream, within-slump, and downstream of slump influence

184 (Fig. 1b). Upstream sites were trunk streams that connected with the slump flow path further

185 downstream, and were un-affected by any major geomorphic disturbance and thus representative of an

186 undisturbed, pristine environment. Within-slump sampling occurred at points of channelized slump

187 runoff within the scar zone or upper debris tongue. Downstream sampling locations were below the
188 confluence of the sampled upstream flow and all within-slump runoff paths, and were chosen to be
189 representative of slump impact on aquatic ecosystems across the Peel Plateau landscape. In one
190 instance (Slump HD, August 17), a fluidized flow event between sampling events saturated the scar zone
191 and obliterated within-slump channelized surface flow. As a result, the within-slump sample taken at
192 this site was not representative of typical channelized slump runoff that characterized all other slump
193 sampling conditions, and has been discarded from all analyses.

194 A general classification of the slumps is difficult as these features are influenced by a diverse
195 range of geomorphic processes that vary in intensity over time (Table 1; Fig. S1). Three of the slumps
196 (FM4, FM2, FM3) are classified as 'mega slumps', characterized by areas greater than 5 ha, a headwall
197 greater than 4 m in height, and a debris tongue that connects the slope to the valley below (Kokelj et al.,
198 2013, 2015). Of these, FM4 possesses a headwall approximately 20 m in height, but was largely
199 stabilized in 2014 (Fig. S1). FM2 is among the largest active slumps in the region, with a headwall 25-30
200 m high and visible as a much smaller feature in air photos since 1944 (Lacelle et al. 2015). Slump FM3,
201 which was the focus for 'environmental controls' work (further described below), covers an area of
202 approximately 10 ha with a headwall of approximately 10 m height and a debris tongue that extends
203 nearly 600 m down valley (Table 1). Headwall retreat rate at FM3 over a 20 year period has been
204 calculated at 12.5 m yr⁻¹ (Lacelle et al., 2015). FM2 and FM3 geochemistry and geomorphology were
205 previously described by Malone et al. (2013). SD is the smallest and youngest slump that we studied,
206 and was initiated when diversion of a small creek caused lateral bank erosion. In 2014, the SD headwall
207 was 2-4 m high with no defined debris tongue and a scar zone extending approximately 20m upslope.
208 The remaining slump sites (HA, HB, HC, HD) were all well-developed active RTS features with headwalls
209 similar to, or smaller than, FM3, but with smaller debris tongues (Table 1). With the exception of SD,
210 slump headwalls exposed permafrost well below a thaw unconformity, indicating that Pleistocene-aged,
211 unweathered glacial materials were being thawed (Lacelle et al., 2013).

212

213 3.2 Field sampling and data collection

214 3.2.1 The effect of slumping on DOC and stream water chemistry

215 The majority of our sampling was conducted during the summer of 2014. Of the eight slumps
216 that were sampled, three were accessed from the Dempster Highway three times over the sampling
217 season, one (FM3; see also Sect. 3.2.2) was accessed twice from the highway, and four were accessed
218 twice via helicopter (Table 1). At each of the upstream, downstream, and within-slump sampling
219 locations, specific conductivity, pH, and temperature were recorded using a YSI Pro Plus multi-
220 parameter meter. Water samples were collected from directly below the stream surface into 1 L acid
221 washed HDPE bottles and allowed to sit in chilled, dark conditions for 24 h to enable the considerable
222 sediments in these samples to partially settle out of suspension. Sample water was then filtered with
223 pre-combusted (475 °C, 4 hours) Whatman GF/F filters (0.7 µm pore size). Filtered sample water was
224 transferred into 40 mL acid washed, pre-combusted glass bottles for DOC analysis, or 60 mL acid washed
225 HDPE bottles for the analysis of absorbance and major ions. DOC samples were acidified with
226 hydrochloric acid (1 µL mL⁻¹), following Vonk et al. (2015b). The GF/F filters were retained for analysis of
227 total suspended solids (TSS). Samples for stable water isotopes were collected directly from streams into
228 acid washed 40 mL HDPE bottles with no headspace and sealed. During summer 2016, samples were
229 additionally collected from a subset of slump locations (FM2, FM3, FM4 and SD) for the ¹⁴C signature of
230 DOC at upstream and within-slump sites. DO¹⁴C samples were collected in acid-washed polycarbonate
231 bottles, allowed to settle for 24 h, and filtered using pre-combusted Whatman GF/F filters into pre-
232 combusted glass media bottles with phenolic screw caps and butyl septa. All samples were refrigerated
233 until analysis. Absorbance samples were analyzed within 1 week of collection, cation samples within 4
234 months of collection, and DOC (including ¹⁴C) samples within 1-2 months of collection. Samples for Fe
235 and δ¹⁸O were analyzed within 6 months of collection.

236

237 3.2.2 Environmental controls on DOC flux

238 To explore how environmental variables control the flux of DOC from RTS-affected streams, we
239 visited slump FM3 an additional 17 times beyond the sampling described above. This intensively-studied
240 site was chosen to be representative of active Peel Plateau slumps that are eroding Holocene- to
241 Pleistocene-aged sediments. During each visit, we measured discharge at the upstream and downstream
242 locations to calculate DOC flux, and collected upstream and downstream DOC concentration samples.
243 Downstream discharge was measured using an OTT C2 current meter at three locations across the small
244 stream and at 40 % depth. Due to the shallow, low flow conditions at the upstream site, upstream
245 discharge was measured using the cross sectional method (Ward and Robinson, 2000). In both cases,
246 discharge was calculated as the product of velocity and stream cross-sectional area. Local daily climate
247 data were obtained from an automated meteorological station established in 2010 by the Government
248 of the Northwest Territories (Kokelj et al. 2015). The station is located within 2 km of slump FM3 (Fig.
249 1a) and is instrumented for the measurement of air temperature, rainfall, and net radiation.

250

251 3.3 Laboratory analyses

252 3.3.1 Major ions, dissolved organic carbon, $\delta^{18}\text{O}$ and DO^{14}C

253 Cation concentrations (Ca^{2+} , Mg^{2+} , Na^+) were analyzed on a Perkin Elmer Analyst 200 Atomic
254 Absorption Spectrometer at York University. A subset of collected samples were analyzed for total
255 dissolved Fe at the University of Alberta on an Inductively Coupled Plasma - Optical Emission
256 Spectrometer (Thermo Scientific ICP6300), to allow for the correction of our Specific UV Absorbance
257 results (see below). DOC samples were analyzed on a Shimadzu TOC-V analyzer; DOC was calculated as
258 the mean of the best 3 of 5 injections with a coefficient of variation of <2%; the precision of a 10 mg L⁻¹
259 caffeine standard across all sample runs was 0.32 mg L⁻¹. A Picarro liquid water isotope analyzer was
260 used to measure $\delta^{18}\text{O}$ at the University of Alberta, following filtration (0.45 μm cellulose acetate,
261 Sartorius) into 2 mL autosampler vials (National Scientific), without headspace. The precision of our

262 $\delta^{18}\text{O}$ analysis is $\pm 0.2\%$. The radiocarbon signature of DOC was measured following extraction and
263 purification at the A.E. Lalonde AMS facility (University of Toronto) using a 3MV tandem accelerator
264 mass spectrometer (High Voltage Engineering) following established methodologies (Lang et al., 2016;
265 Palstra and Meijer, 2014; Zhou et al., 2015), and is reported with an error estimate of 1σ .

266

267 3.3.2 Total suspended solids

268 Samples for TSS were filtered in the field for later analysis, ensuring that there was enough
269 sediment on the pre-combusted (475 °C, 4 hours) and pre-weighed GF/F filters. Filters were stored
270 frozen, dried at 60 °C for 8 hours, placed in a desiccator overnight and promptly weighed. TSS was
271 calculated as the difference in filter weight before and after sediment loading, divided by volume
272 filtered.

273

274 3.3.3 Dissolved organic matter spectral characteristics

275 DOM composition was assessed using absorbance-based metrics. A 5 cm quartz cuvette was
276 used to obtain UV-visible spectra data from 250-750 nm, using a Genesys 10 UV-Vis spectrophotometer.
277 A baseline correction was applied to eliminate any minor interference from particles $< 0.7\ \mu\text{m}$ (Green
278 and Blough 1994). Specific UV absorbance at 254 nm (SUVA_{254}), which is correlated with DOM
279 aromaticity (Weishaar and Aiken, 2003), was calculated by dividing the decadal absorbance at 254 nm
280 (m^{-1}) by the DOC concentration (mg L^{-1}). SUVA_{254} values were corrected for Fe interference following
281 Poulin et al. (2014) using maximum Fe concentrations from laboratory analyses or as reported in Malone
282 et al. (2013). Spectral slopes between 275 and 295 nm, and 350 and 400 nm ($S_{275-295}$, $S_{350-400}$) were
283 calculated following Helms et al. (2008), and are reported as positive values to adhere to mathematical
284 conventions. Slope ratios (S_R), which correlate with DOM molecular weight (Helms et al., 2008), were
285 calculated as the ratio of $S_{275-295}$ to $S_{350-400}$.

286

287 *3.4 Statistical analyses and calculations*

288 Statistical analyses were completed in R version 3.1.3 (R Core Team, 2015) using packages ‘nlme’
289 (Pinheiro et al., 2015), ‘lme4’ (Bates and Kuznetsov, 2015), ‘lmerTest’ (Kuznetsov and Hothorn, 2002), ‘lmSupport’ (Curtin, 2015), ‘car’ (Fox and
290 Weisberg, 2011), and ‘zoo’ (Zeileis and Grothendieck, 2005). The effect of slumping on stream chemistry
291 and optical characteristics was assessed using linear mixed effects models in the ‘nlme’ package of R. For
292 each parameter, analyses were split into two separate models that included data for upstream and
293 downstream chemistry, and upstream and within-slump chemistry. We used this approach to separately
294 assess the effects of slumping downstream of slump systems, and to compare the composition of slump
295 runoff to nearby, pristine environments. For each analysis, we included slump location (see Table 1) as a
296 random effect, and considered models that either nested Julian date within the random effect of slump
297 location, or allowed Julian date to occur as a fixed effect. The best model was chosen using the Akaike
298 Information Criterion (AIC), and best-fit models were refit with a variance structure to ensure that
299 model assumptions were met. The variance structures varIdent (for within-slump site and slump
300 location) and varFixed (for Julian date) were used together (using varComb) and in isolation for this
301 purpose (Zuur et al., 2009). AIC values for the weighted and un-weighted models were again compared
302 to choose a final model of best fit for each analysis.

303 We used the high-frequency data from slump FM3 to assess how environmental conditions
304 (rainfall, temperature, solar radiation) and TSS affect DOC delivery to slump-affected streams. To do
305 this, we conducted multiple linear regressions, using AIC values to determine models of best fit
306 (Burnham and Anderson, 2002). To enable a specific assessment of environmental controls on
307 downstream DOC flux, upstream DOC flux was separated out into a distinct regression analysis, because
308 upstream DOC flux was strongly correlated with flux downstream, and therefore overwhelmed all
309 environmental variables in the downstream model. Models were tested for serial correlation using the
310 auto-correlation function, and models with variance inflation factors greater than 10 or significant
311 Durbin Watson test results (indicative of correlated variables; Durbin & Watson, 1950; Hair et al., 1995)

312 were discarded. Residuals were examined to ensure the model was a good fit for the data (Zuur et al.,
313 2009). We considered both time-of-sampling (0 h) and past (48, 72, and 120 h) environmental conditions
314 in our analyses. Because cumulative values for environmental variables (i.e. accumulated rainfall in the
315 previous 48, 72 and 120 h) showed a strong positive correlation to one another, we used temporally
316 shifted data (i.e. rainfall 48, 72 and 120 h prior to the DOC flux measurement) in the final model. Similar
317 models were also constructed to examine the effects of environmental drivers on DOC concentration.
318 Differences in paired upstream-downstream measures of DOC flux and concentration at slump FM3
319 were also assessed using a Wilcoxon Signed Rank Test, a non-parametric analog to the paired-t test.

320 Following our finding of decreasing DOC concentrations downstream of slumps (see Sect. 4.1
321 and 5.1) we used data from slump FM3, where we have upstream, downstream, and within-slump DOC
322 concentration measurements, and upstream and downstream discharge measurements, to calculate a
323 mass balance for DOC across the three sampling locations. These data – available for all three locations
324 on two dates during the summer of 2014 – were used to calculate DOC flux at upstream and
325 downstream sites as $\text{flux}_{\text{DOCdown}} = [\text{DOC}]_{\text{down}} \cdot \text{discharge}_{\text{down}}$ or $\text{flux}_{\text{DOCup}} = [\text{DOC}]_{\text{up}} \cdot \text{discharge}_{\text{up}}$, and at
326 within-slump sites as $\text{flux}_{\text{DOCwithin}} = [\text{DOC}]_{\text{within}} \cdot (\text{discharge}_{\text{down}} - \text{discharge}_{\text{up}})$. We calculate a similar mass
327 balance for TSS, which we use as a rough tracer for the inflow of slump runoff over the < 1 km span
328 between upstream and downstream locations at this site.

329

330 **4. Results**

331 *4.1 DOC concentration across slump sites*

332 While DOC concentrations ranged broadly across pristine streams on the Peel Plateau (Fig. 2;
333 from 5.4 to 26.1 mg L⁻¹ at upstream, pristine sites), concentrations consistently declined downstream of
334 slumps, when compared to paired, upstream locations ($p < 0.001$; Fig. 2; Table 2). Although this effect
335 was modest (typically less than 20 %; Fig. 2), it occurred reliably across all slump sites. In contrast,
336 comparisons of upstream and within-slump sites showed no consistent trend in DOC concentration,

337 when evaluated across all slump locations ($p=0.153$; Fig. 2; Table 2). Instead, the effects of slumping on
338 the DOC concentration of slump runoff varied by site. At the largest, most well-developed slump
339 complexes (FM4, FM2, and FM3), where debris tongues are extensive and thaw extends well into the
340 deepest layer of Pleistocene-aged glacial materials, DOC concentrations tended to be lower in slump
341 runoff than at the paired upstream sites (Fig. 2). At more modestly-sized slumps (HB, HC, and HD),
342 where modern and relict Holocene active layers comprise a greater proportion of thawed materials,
343 within-slump DOC concentrations tended to be higher than values upstream (Fig. 2). Within each site,
344 DOC concentrations were relatively consistent across the 2-3 sampling periods (Fig. 2).

345

346 *4.2 Bulk chemistry of pristine waters and slump runoff*

347 To better understand how the input of slump runoff affects downstream DOC, we examined
348 concentrations of major ions, conductivity and TSS as 'tracers' of slump activity, because these
349 constituents have previously been shown to be significantly affected by slumping in this region (Kokelj et
350 al., 2005, 2013; Malone et al., 2013; Thompson et al., 2008). Major ion (Ca^{2+} , Mg^{2+} , Na^+) concentrations
351 in slump runoff were considerably greater than in pristine streams (a 2.7 to 11.7-fold increase; Fig. 3b-d;
352 Table 2). These patterns were similar, though muted, at slump-affected downstream sites, where major
353 ion concentrations were 1.5 to 3.5-fold greater than at pristine sites (Fig. 3b-d; Table 2). Mean
354 conductivity also increased significantly as a result of slumping ($p < 0.001$; Table 2): within-slump sites
355 had conductivity values that were 9.2-fold greater than upstream sites, while downstream values were
356 an average of 2.6 times greater than those upstream (Fig. 3e). Finally, TSS was also significantly elevated
357 at slump-affected sites ($p < 0.001$; Table 2) with concentrations being more than two orders of
358 magnitude greater within slumps, and more than one order of magnitude greater downstream, when
359 compared to upstream sites (Fig. 3a). The effect of slump runoff on downstream chemistry is also
360 reflected in DOC: ion, and DOC: TSS ratios, which decreased markedly between upstream and
361 downstream locations. For example, molar ratios of ($\text{Ca}^{2+} + \text{Mg}^{2+}$): DOC averaged 0.78 ± 0.37 (mean \pm

362 standard error) upstream of slumps, but 2.07 ± 0.45 downstream, while average gram-weight ratios of
363 TSS: DOC were 32 ± 12 upstream, but 1454 ± 332 at downstream locations.

364

365 *4.3 Spectral and isotopic characteristics*

366 SUVA₂₅₄, which is positively correlated with DOM aromaticity (Weishaar and Aiken, 2003), was
367 significantly lower within slumps, and downstream of slumps, than in upstream, pristine, environments
368 ($p < 0.001$; Fig. 4; Table 2). Mean within-slump SUVA₂₅₄ was less than half of that observed for pristine
369 waters (Fig. 4), while downstream values declined by approximately 20 %. In accordance with the
370 SUVA₂₅₄ results, $S_{275-295}$, $S_{350-400}$, and S_R were all significantly greater within slumps when compared to
371 upstream sites ($p < 0.001$; Fig. 4; Table 2), indicating lower DOM molecular weight within slumps (Helms
372 et al., 2008). Differences in slope parameters between upstream and downstream locations were muted
373 relative to the within-slump: upstream comparisons (Fig. 4), with $S_{275-295}$ ($p = 0.011$) and S_R ($p < 0.001$)
374 increasing significantly, but more modestly, downstream of slumps, and $S_{350-400}$ declining slightly
375 ($p = 0.001$; Fig. 4; Table 2).

376 Upstream $\delta^{18}\text{O}$ averaged $-20.1 \text{‰} \pm 0.12$, which corresponds to a modern active-layer pore
377 water $\delta^{18}\text{O}$ signature for this region (Lacelle et al., 2013; Fig. 5). Within-slump $\delta^{18}\text{O}$ was discernibly
378 depleted when compared to upstream locations, with mean values of $-22.7 \text{‰} \pm 0.72$, which falls
379 between previously-identified regional endmembers for Pleistocene-aged ground ice ($18,100 \pm 60 \text{ }^{14}\text{C yr}$
380 BP) and the modern active layer (Lacelle et al., 2013; Fig. 5). Within-slump $\delta^{18}\text{O}$ was also much more
381 variable between RTS features than upstream and downstream $\delta^{18}\text{O}$ values. Similar to upstream sites,
382 downstream $\delta^{18}\text{O}$ clustered near the modern active layer $\delta^{18}\text{O}$ endmember, but with a small depletion
383 that was consistent with a contribution from slump inflow ($-20.7 \text{‰} \pm 0.21$).

384 To further investigate the effect of water source on DOM composition, we examined the
385 relationship between SUVA₂₅₄ and $\delta^{18}\text{O}$. More depleted samples taken from within-slump sites had
386 clearly depressed SUVA₂₅₄ values when compared to samples with more enriched $\delta^{18}\text{O}$ (Fig. 5). Of the

387 large, most well-developed slumps that were identified in Sect. 4.1, two (FM2 and FM3), in addition to
388 site HB, had $\delta^{18}\text{O}$ values that were more depleted than the Holocene-aged icy diamicton values reported
389 in Lacelle et al. (2013), suggesting some contribution of runoff from older, Pleistocene-aged permafrost
390 (Fig. 5). It is likely that the $\delta^{18}\text{O}$ signal at the relatively stable mega-slump site (FM4) was somewhat
391 diluted by the 7.2 mm of rainfall that fell in the 48 hours preceding our sample. Although sites FM3 and
392 SD received 12.4 and 3.5 mm of rain, respectively, in the 48 hours prior to sampling, these are both
393 much more active slump sites, and thus less prone to dilution of the slump outwash signature. There
394 was no significant rainfall immediately preceding sampling at any other sites.

395 The radiocarbon signature of DOC from upstream and within-slump locations at sites FM4, FM2,
396 FM3, and SD largely mirrors the $\delta^{18}\text{O}$ results. DOC from sites upstream of slump disturbance was
397 approximately modern in origin (ranging from 217 ± 24 ^{14}C yr BP to modern in age; Table 3). In contrast,
398 within-slump waters from site FM2 and FM3 were early Holocene-aged (9592 ± 64 , and 8167 ± 39 ^{14}C yr
399 BP, respectively; Table 3). Slump runoff from site SD was older than at upstream sites, but younger than
400 for the larger slumps, described above (1157 ± 23 ^{14}C yr BP; Table 3).

401

402 *4.4 Patterns and environmental drivers of DOC flux*

403 Similar to our findings for the distributed sampling scheme (Fig. 2), downstream DOC
404 concentration was consistently lower than concentrations upstream, across the 19 paired
405 measurements taken at the intensively studied site FM3 ($p < 0.001$, $N = 19$, $W = 0$; Wilcoxon Signed Rank
406 Test; mean decline of 2.5 ± 0.2 mg L^{-1} , compared to a mean upstream concentration of 13.6 ± 0.5 mg L^{-1}). To explore environmental drivers of DOC movement within this landscape, however, we focus on
407 DOC flux, which allows a direct assessment of slump-mediated DOC addition to this system.
408 Downstream DOC flux (mg s^{-1}) tended to be slightly greater than upstream flux on most, but not all,
409 sampling occasions (Fig. 6). As a result, paired comparisons indicate no statistical difference between
410 upstream and downstream DOC flux at this site (Wilcoxon signed rank test; $p = 0.096$, $N = 19$, $W = 53$).

412 Because upstream and downstream DOC flux were strongly correlated to one another ($r^2 = 0.94$;
413 $p < 0.0001$), our downstream model was run without upstream DOC flux as a predictor variable. The best-
414 fit multiple linear regression model for downstream DOC flux ($r^2 = 0.84$; $p < 0.01$) retained seven
415 variables, of which two were significant (Table 4). Of these, air temperature (72 h prior to sampling)
416 showed a negative relationship with downstream DOC flux while rainfall (0 h; time of sampling) showed
417 a strong positive relationship (Table 4). The best-fit model for upstream DOC flux ($r^2 = 0.87$; $p < 0.001$)
418 also retained seven variables, of which four were significant ($p < 0.05$; Table 4). Similar to the
419 downstream analysis, air temperature (0 h, 72 h) displayed a negative relationship, and time-of-
420 sampling (0 h) rainfall a strong positive relationship, with DOC flux (Table 4). However, 120 h rainfall
421 showed a negative relationship with DOC flux in this model. Regressions assessing controls on
422 downstream DOC flux relative to upstream flux (i.e., as a ratio, or the difference between the two
423 values) were not significant. Models to explore controls on upstream and downstream DOC
424 concentration were also relatively similar to one another, showing strong, positive relationships
425 between DOC concentration and air temperature, and more modest negative relationships between
426 DOC concentration and net radiation (Table 4).

427

428

429 **5. Discussion**

430 *5.1 Retrogressive thaw slumps and carbon delivery to streams of the Peel Plateau*

431 In both Eastern Siberia (Spencer et al. 2015; Vonk et al., 2013b) and Alaska (Abbott et al., 2014)
432 permafrost slumping has been associated with significant increases in DOC mobilization from terrestrial
433 to aquatic systems. Our data show that this was not the case on the Peel Plateau, where the landscape-
434 induced variation in DOC concentration among pristine stream sites was much greater than the change
435 in stream water DOC as a result of slumping. Across all of our study sites, DOC concentrations
436 consistently declined downstream of slumps when compared to upstream locations, while at an

437 intensively-sampled slump, DOC flux did not differ significantly between upstream and downstream
438 locations. In contrast, comparisons of channelized slump runoff (our within-slump sites) and paired un-
439 affected sites showed no consistent DOC trend. Instead, DOC concentrations in slump runoff were either
440 greater than, or less than, their comparison upstream locations, in a manner that differed depending on
441 slump morphological characteristics such as slump size and headwall height (Fig. 1; see further
442 discussion in Sect. 5.3). The moderate effect of slumping on DOC concentration occurred despite the
443 significant influence of these disturbances on the delivery of many biogeochemical constituents to
444 recipient streams. For example, conductivity was approximately one order of magnitude greater, and
445 TSS two orders of magnitude greater, in slump-derived runoff than at upstream, un-affected sites. This
446 led to substantially increased TSS:DOC and (Ca + Mg):DOC ratios downstream of slumps, when
447 compared to pristine, upstream locations.

448 Decreasing DOC concentrations downstream of slumps, despite increasing concentrations of
449 indicators of slump activity (major ions, TSS) could be driven by several, potentially co-occurring factors.
450 In some locations, decreases may be partially caused by low DOC concentrations in slump outflow (a
451 dilution effect; see slumps FM2, FM3, and FM4 in Fig. 2; further discussed in Sect. 5.3). However, our
452 results suggest that DOC sorption to suspended inorganic sediments could also play a role in regulating
453 DOC dynamics in slump-affected systems. At multiple sites (HB, HC, and HD), DOC concentrations
454 declined downstream of slumps despite a modest elevation in DOC concentration in slump drainage
455 waters (Fig. 2). Thermokarst contributes significant amounts of fine-grained glacial sediment to
456 fluvial systems on the Peel Plateau (Kokelj et al., 2013; silty-clay sediment classification for FM3 in
457 Lacelle et al., 2013). DOC sorption can occur in seconds to minutes in freshwater systems (Qualls and
458 Haines, 1992), with fine-grained materials being particularly conducive to this process (Kothawala et al.,
459 2009). Data from site FM3, where we have upstream and downstream discharge data coupled with DOC
460 and TSS concentrations at upstream, downstream, and within-slump locations on two separate dates,
461 allows possible DOC sorption to be assessed. On these dates, DOC flux declined downstream of the

462 slump (i.e., $\text{flux}_{\text{DOCdown}} < \text{flux}_{\text{DOCup}}$), despite a clear and measurable efflux of DOC from the slump to the
463 receiving stream system ($\text{flux}_{\text{DOCwithin}}$; Fig. 7). This same calculation using TSS as a rough tracer of slump
464 inflow shows the calculated efflux of TSS from this slump ($\text{flux}_{\text{TSSwithin}}$) to be almost identical to the
465 increase in TSS flux downstream of the disturbance (as $\text{flux}_{\text{TSSdown}} - \text{flux}_{\text{TSSup}}$; Fig. 7). Thus, it seems likely
466 that relatively rapid processes, such as sorption to mineral surfaces, are affecting DOC dynamics in
467 thermokarst-affected fluvial systems on the Peel Plateau.

468 Although a similar decrease in DOC concentration with slumping has been found for lakes in this
469 region (Kokelj et al., 2005), our findings contrast with those from other previously-studied areas of the
470 Arctic, where thermokarst leads to an efflux of high-DOC waters from slump features (e.g., Abbott et al.,
471 2014; Vonk et al., 2013a). However, ice-marginal glaciated landscapes are common throughout the
472 western Canadian Arctic, and in many other Arctic regions. The thick, mineral-rich, carbon-poor tills with
473 high ice contents that characterize these landscapes are predisposed to intense thaw slumping and the
474 mobilization of glacial sediments from slope to stream (Kokelj et al., 2017b). As a result, DOC
475 ‘sequestration’ following slumping seems unlikely to be limited to the Peel Plateau. Given the high TSS
476 export and apparent organic carbon sorption to glacial sediments observed with slumping on the
477 Peel Plateau, we expect that substantial organic carbon is mobilized from these disturbances in the
478 particle-attached, rather than dissolved, form (i.e., as particulate organic carbon; POC). Quantifying this
479 POC mobilization and fate once subject to contemporary biogeochemical processing, and the
480 mechanisms that enable DOC sequestration to occur, are key avenues for future research on the Peel
481 Plateau and elsewhere.

482

483 *5.2 The effect of retrogressive thaw slumps on DOM composition*

484 Although DOC concentrations did not increase in RTS-affected streams, absorbance metrics
485 clearly indicate that slump-derived DOM on the Peel Plateau is compositionally different than DOM from
486 upstream locations. Upstream waters had significantly higher SUVA_{254} values than downstream and

487 within-slump sites (Table 2, Fig. 4). Similarly, while the average S_R of Peel Plateau upstream waters (0.74
488 ± 0.005) was within the range of S_R typically associated with fresh, terrestrial DOM (~ 0.70 ; Helms et al.,
489 2008), values were significantly greater within-slump (0.92 ± 0.015) and downstream (0.89 ± 0.009)
490 (Table 2, Fig. 4), indicating decreasing DOM molecular weight as a result of RTS activity. High $SUVA_{254}$
491 values accompanied by low S_R at upstream sites suggest that water flow in undisturbed catchments is
492 restricted to shallow, organic-rich flowpaths through the active layer, with permafrost inhibiting water
493 contributions from deeper, groundwater or mineral-associated sources (Balcarczyk et al., 2009;
494 MacLean et al., 1999; Mann et al., 2012; O'Donnell et al., 2010; Street et al. 2016). In contrast, within-
495 slump and downstream measurements indicate a clear transition in DOM source.

496 The comparatively low $SUVA_{254}$, and high S_R values for downstream and within-slump sites
497 indicate that permafrost-derived carbon on the Peel Plateau is characterized by relatively low molecular
498 weight and aromaticity, and is thus similar in its composition to permafrost carbon from other regions.
499 For example, $SUVA_{254}$ values were low in waters draining active thaw slumps when compared to
500 stabilized and undisturbed sites on the North Slope of Alaska (Abbott et al., 2014), while in Siberia, ^{14}C -
501 depleted DOM from small tributary streams affected by thermokarst had lower $SUVA_{254}$ values
502 compared to younger DOM from the Kolyma River mainstem (Mann et al., 2015; Neff et al., 2006).
503 Although $SUVA_{254}$ values for waters draining Peel Plateau thaw slumps are slightly lower than those
504 reported for Siberian Yedoma disturbances (Mann et al., 2015), the overall similarity of permafrost-
505 derived DOM composition across these various regions is striking, given the regional differences in
506 permafrost origin and landscape history. For example, the DOM released by permafrost thaw on the
507 Peel Plateau is till-associated, and early-Holocene in mean age, while east Siberian Yedoma is composed
508 of loess-derived Pleistocene deposits that sequestered carbon in association with synengetic permafrost
509 aggradation. This suggests that common processes may enable the organic matter contained in
510 permafrost soils to become compositionally similar across diverse Arctic regions. Such compositional
511 similarity also indicates that permafrost-origin DOM from the Peel Plateau – similar to that from other

512 regions (Abbott et al., 2014; Drake et al., 2015) – may be readily degraded by bacteria, despite the
513 divergent origin of this carbon.

514

515 *5.3 The effect of slump morphometry on runoff water biogeochemistry*

516 $\delta^{18}\text{O}$ and DO^{14}C data provide further evidence that intense slumping enables novel sources of
517 water and solutes to be transported to fluvial systems on the Peel Plateau. For most of the RTS features
518 that we studied, the $\delta^{18}\text{O}$ signature of within-slump waters ranged from similar to the ‘icy diamicton’
519 that overlies the early Holocene thaw unconformity, to that for underlying Pleistocene-aged ground ice
520 (Lacelle et al., 2013; Fig. 5). Similarly, DO^{14}C from a subset of sites indicates slump-derived DOC is early
521 Holocene in age for all but the shallowest slump surveyed. This suggests that our slump outflow samples
522 were likely comprised of a mixture of Pleistocene-, Holocene-, and modern-sourced water (see Fig. 1c-
523 e), but that the contribution of these end-members varied across slumps depending on the relative
524 volume of different stratigraphic units being mobilized.

525 The between-site variation in $\delta^{18}\text{O}$ signature (Fig. 5) and relative DOC concentration (Fig. 2b) of
526 slump runoff waters appears to be related to differences in slump morphometry (size, headwall height,
527 and the length and area of the debris tongue; see Table 1 and Fig. 1c-e) across sites. The well-developed,
528 larger slump complexes (FM4, FM2 and FM3) were more likely to have $\delta^{18}\text{O}$ signatures that lie between
529 end-member values for Holocene-aged icy diamicton and Pleistocene-aged ground ice (Fig. 5; although
530 note that dry and stabilized FM4 differs somewhat from this trend). These well-developed slumps also
531 stood out as displaying within-slump DOC concentrations that were lower than at upstream comparison
532 sites (Fig. 2b). The headwall exposure at these largest slumps exposes Pleistocene-aged permafrost to
533 several m depth (see Fig. 1c), while the evacuation of scar zone materials has produced extensive debris
534 tongues up to several km long (Table 1, Figs. 1b, S1e and S1g). This significant exposure of mineral-rich,
535 Pleistocene-aged glacial till contributes solutes from low-carbon mineral soils and low-DOC ground ice
536 (Fritz et al. 2015; Tanskii et al. 2016) to runoff, while entraining fine-grained sediments which provide

537 mineral surface area for possible DOC adsorption. Adsorption may be further enhanced as slump and
538 stream runoff continue to entrain sediments as flows incise the lengthy debris tongue deposits. In
539 contrast, slumps with slightly shallower headwalls (HA, HB, HC, HD; see Fig. 1d), and less well-developed
540 debris tongues (Table 1), appear to elicit a slightly different response than the largest slumps discussed
541 above. At these mid-sized sites, within-slump DOC concentrations were typically higher than those
542 found at upstream comparison sites (Fig. 2b), which may reflect the greater relative inputs from thawing
543 of the Holocene-aged relict active layer, and decreased interaction with debris tongue deposits at these
544 smaller disturbances. Similarly, runoff $\delta^{18}\text{O}$ tends to lie between Holocene and modern end-member
545 values at these sites (though note the more depleted value for HB; Fig. 5), indicating a lower relative
546 contribution of Pleistocene-aged ground ice to slump outflow waters.

547 Finally, the youngest and shallowest slump surveyed (SD), exposed only near-surface permafrost
548 soils for leaching and geochemical transport (Figs. 1e and S1; Table 1), and not the underlying mineral
549 and ice-rich glacial substrates. Accordingly, the effects of slumping on stream chemistry, optical
550 parameters, and isotopes were muted at SD when compared to the larger slumps discussed above.
551 These morphometry-related shifts in the downstream effects of slumping suggest that we should expect
552 non-linearity in the biogeochemical response as RTS features develop over time, particularly if slumping
553 continues to intensify with future warming on the Peel Plateau (e.g., Kokelj et al., 2017b), underscoring
554 the importance of long-term monitoring on the Peel Plateau and elsewhere.

555

556 *5.4 Environmental controls on DOC flux and concentration*

557 Air temperature and rainfall exerted the strongest control on DOC flux at our intensively studied
558 site, which was chosen to be representative of active Peel Plateau slumps eroding Holocene- to
559 Pleistocene-aged sediments (slump FM3; Fig. 6; Table 4). Upstream of the slump, rainfall was positively
560 correlated, and air temperature negatively correlated, with DOC flux. However, precipitation events
561 were negatively related to temperature (Fig. 6), suggesting that over a single season, precipitation

562 served as the primary environmental control on upstream DOC flux. DOC concentration was relatively
563 constant with upstream discharge ($r=-0.342$, $p=0.151$), indicating that precipitation controlled DOC flux
564 largely as a result of changes in runoff, and that DOC was not source-limited over the time scale of our
565 investigation. However, upstream DOC concentration was positively related to temperature (Table 4),
566 suggesting a link between biological activity and within-soil DOC production (c.f. Pumpanen et al., 2014).
567 These upstream-of-slump results are consistent with work from other undisturbed permafrost and
568 boreal regions, where precipitation and catchment runoff have been shown to control DOC flux in
569 streams (Prokushkin et al., 2005; Pumpanen et al., 2014), and increasing temperature has been shown
570 to increase DOC production in soils (Christ and David, 1996; Neff and Hooper, 2002; Prokushkin et al.,
571 2005; Yanagihara et al., 2000). They are also consistent with the concept that the permafrost barrier
572 forces runoff to travel through the shallow active layer, where high hydraulic conductivity leads to rapid
573 transport of carbon into fluvial systems (O'Donnell et al., 2010; Striegl et al., 2005).

574 Slumping did not significantly affect downstream DOC flux at the intensively studied slump site,
575 when compared to DOC flux upstream (Fig. 6; Sect. 4.4). Although concentration consistently declined
576 downstream at FM3 (Sect. 4.1 and 4.4), downstream DOC flux was either slightly higher, or slightly
577 lower, than upstream flux; a result that seems likely to play out at other, comparable Peel Plateau
578 slumps, given the coherent concentration patterns that we observed across this landscape. Concordant
579 with the lack of slump effect on DOC flux, neither the ratio of (downstream: upstream) or difference
580 between (downstream – upstream) upstream and downstream DOC flux could be explained by any of
581 our environmental variables, while the environmental controls on downstream flux were almost
582 identical to those upstream (Table 4). The lack of clear environmental control on relative downstream:
583 upstream DOC flux occurred despite the fact that precipitation has been shown to be a strong driver of
584 sediment movement from slump features on the Peel Plateau, at time scales similar to those used for
585 this work (Kokelj et al., 2015).

586 Considering the Peel Plateau landscape as a whole, it appears that precipitation serves as a

587 primary, positive control on DOC flux. Thus, this study adds DOC production to the list of changes – such
588 as increasing slump activity and sediment mobilization – that can be expected with the increased
589 precipitation that is affecting this region, and is predicted for many Arctic locations (IPCC, 2014; Kokelj et
590 al., 2015). However, it appears that slumping does not over-ride the landscape-scale control on DOC flux
591 in this system – at least at the scale of this single-season – perhaps because processes like DOC sorption
592 mask the influx of slump-derived DOC (Fig. 6). This result highlights the complexity of the interaction
593 between changing climatic parameters and DOC dynamics on the Peel Plateau, where thaw slumps of
594 increasing size mobilize till, glaciolacustrine, glaciofluvial, and organic deposits, while also draining
595 contemporary active layers across a shrub-tundra to spruce forest upland gradient. DOC dynamics are
596 thus affected by both water and carbon generation across these variable landform types, and by
597 biogeochemical interactions such as mineral adsorption in recipient systems. Future work to tease apart
598 the interactions between changing climatic parameters, slump development, and resultant
599 biogeochemical effects is clearly warranted, with the recognition that environmental controls on slump
600 activity, and thus downstream biogeochemistry, can be expected to show marked regional variation (see
601 for example, work from Eureka Sound; Grom & Pollard 2008).

602

603 **6. Conclusions: Dissolved carbon mobilization across diverse permafrost landscapes**

604 Carbon dynamics in Arctic aquatic systems are influenced by numerous factors, including
605 geology, Quaternary and glacial history, soil composition, vegetation, active layer dynamics, and the
606 nature and intensity of thermokarst. As a result, the effect of permafrost thaw on DOC concentration
607 and flux should – at a fundamental level – vary across broad, regional scales. Our results demonstrate
608 that we can expect marked inter-regional variation in DOC transport to streams in response to
609 permafrost degradation. For example, declines in DOC concentration downstream of slumps on the Peel
610 Plateau clearly differ from what has been found in eastern Siberia and regions of Alaska, where
611 thermokarst releases substantial quantities of DOC (e.g., Spencer et al. 2015), and increases DOC

612 concentrations in downstream systems (Abbott et al. 2015). Efforts that incorporate information
613 concerning the geology and Quaternary history of thawing landscapes, the physical and geochemical
614 composition of permafrost soils, and the nature and intensity of thermokarst processes within
615 landscapes (see, for example, Olefeldt et al. 2016) will considerably increase our ability to predict
616 climate-driven changes in carbon delivery from land to water on a pan-Arctic scale.

617 At finer scales, this work underscores the variability of thermokarst effects within regions, and
618 the local-scale control on this variability. On the Peel Plateau, between-site differences in the
619 biogeochemical effect of thermokarst are related to variation in soil stratigraphy (i.e., the relative depth
620 of the Holocene-aged paleo-active layer) and ever-evolving slump morphometry. Although striking
621 within-region variability in biogeochemical response to thermokarst has been seen elsewhere (e.g.,
622 Watanabe et al., 2011), responses in other regions occur as a result of very different – and region-
623 specific – landscape-level drivers. This landscape-specificity also extends to the non-linear
624 biogeochemical response as thermokarst features develop over time. Changes in downstream
625 biogeochemistry with slump development are very different on the Peel Plateau, for example, than in
626 other regions (e.g., Abbot et al. 2015), while temporal non-linearity can also be expected for other types
627 of permafrost thaw (Kokelj et al. 2002, Vonk et al. 2016) such as increasing active layer thickness
628 (Romanovsky et al. 2010). It seems clear that a tiered approach, targeted within regions to understand
629 local controls on thaw-driven DOC mobilization, and across regions to document the effects of broad-
630 scale variation imposed by geological and climate legacy, is required to understand future
631 biogeochemical functioning of thermokarst-affected landscapes in a warming circumpolar Arctic.

632

633

634 **Data availability:** Data associated with this manuscript have been made available in Tables S1 and S2.

635

636 **Competing interests:** The authors declare that they have no conflict of interest.

637

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937 **Table 1:** Slump characteristics and sampling information for eight retrogressive thaw slumps sampled
 938 during the 2014 field season on the Peel Plateau, NWT, Canada. Characteristics are derived from
 939 published values and field estimations.

Slump location	Sample dates (Julian day) ^a	Latitude	Longitude	Area (ha)	Debris tongue (m) ^b	Headwall height (m)
FM4	202, 210, 223	67 16.679	-135 09.573	8.8	960	16 to 20 ^d
FM2	200, 209, 222	67 15.462	-135 14.216	31.7	1529	25 ^e
FM3	197, 212	67 15.100	-135 16.270	6.1	576	10 ^e
SD	196, 213, 234	67 10.818	-135 43.630	3.3	NA	2 – 4 ^d
HA	190, 229	67 09.057	-135 41.121	5.9	288	6 – 10 ^d
HB	190, 229	67 14.397	-135 49.167	13.6 ^c	257	6 – 10 ^d
HC	190, 229	67 19.652	-135 53.620	10.3, 10.3 ^c	408	6 – 10 ^d
HD	190, 229	67 24.025	-135 20.048	1.8	137	6 – 10 ^d
Weather Station		67 14.756	-135 12.920			

940
 941 ^a Excludes samples for the FM3 ‘environmental controls’ analysis which was conducted on 17 additional
 942 dates; HD, Julian date 229 did not include a within-slump sample.
 943 ^b The length of debris tongue measured from the base of the debris scar, along the valley bottom stream
 944 ^c Site HB is comprised of two smaller slump features that have merged into the scar zone delineated
 945 here; site HC is comprised of 5 separate slump features that have merged into two scar zones, each with
 946 an area of 10.3 ha
 947 ^d Rough estimates by field crews over 2014 and 2015 field seasons
 948 ^e Kokelj et al. 2015
 949
 950
 951

952 **Table 2:** Results of the mixed-effects models used to assess the effects of slumping on stream water
 953 chemistry and optical characteristics. Downstream models incorporated data from downstream and
 954 upstream sites; within-slump models incorporated data from within-slump and upstream sites. Provided
 955 are degrees of freedom (df), t-statistics, and p-values for individual model runs. Further details on the
 956 statistical approach are provided in Section 3.4.

957

	Downstream			Within-slump		
	df	t	p	df	t	p
DOC	20	-12.895	<.0001	30	-1.468	0.153
Na	33	9.662	<.0001	30	7.278	0.000
Ca	33	9.767	<.0001	30	4.782	0.000
Mg	33	6.166	<.0001	30	8.593	0.000
Conductivity	32	43.083	<.0001	30	11.895	0.000
TSS	29	6.692	<.0001	28	2.187	0.037
SUVA	32	-4.460	<.0001	30	-35.052	0.000
S _R	32	5.333	<.0001	31	8.065	0.000
S ₂₇₅	31	2.856	0.008	31	8.159	0.000
S ₃₅₀	32	-2.196	0.036	31	16.665	0.000

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959

960 **Table 3:** Measured fraction modern carbon ($F^{14}C$) and estimated calendar years before present for ^{14}C of
 961 dissolved organic carbon samples collected upstream of, and within drainage waters of, selected slump
 962 sites. Data were collected during the summer of 2016. nc indicates sample not collected. Error
 963 estimates indicate 1σ .

964

Site	$F^{14}C$		^{14}C yr BP	
	Upstream	Within-slump	Upstream	Within-slump
FM4	0.9734 ± 0.0029	nc	217 ± 24	nc
FM2	0.9764 ± 0.0032	0.3030 ± 0.0024	192 ± 27	9592 ± 64
FM3	1.0023 ± 0.0030	0.3618 ± 0.0018	modern	8167 ± 39
SD	1.0216 ± 0.0035	0.8659 ± 0.0025	modern	1157 ± 23

965

966

967 **Table 4:** Results of multiple linear regression analyses to assess environmental controls on upstream and downstream DOC flux, and upstream
 968 and downstream DOC concentration. nr indicates variables that were not retained in the best fit regression model; NA indicates variables that
 969 were not run in individual analyses. Significant p-values are indicated with bold text; marginal results ($0.05 < p < 0.10$) are indicated in italics.
 970 Model statistics are as follows: downstream flux $r^2=0.84$, $F_{7,11}=8.25$, $p = 0.001$; upstream flux $r^2=0.87$, $F_{7,11}=10.79$, $p < 0.001$; downstream
 971 concentration $r^2=0.85$, $F_{4,14}=19.57$, $p < 0.001$; upstream concentration $r^2=0.91$, $F_{5,13}=27.05$, $p < 0.001$.

Coefficient	Downstream DOC flux			Upstream DOC flux			Downstream DOC concentration			Upstream DOC concentration		
	Estimate	t	p	Estimate	t	p	Estimate	t	p	Estimate	t	p
Average Air Temperature (°C)												
0 h	-67.08	-1.685	0.120	-115.96	-3.286	0.007	nr	nr	nr	0.165	2.349	0.035
48 h	nr	nr	nr	56.32	1.534	0.153	0.332	6.886	<0.001	0.396	5.510	<0.001
72 h	-95.15	-2.594	0.025	-94.17	-2.717	0.020	nr	nr	nr	nr	nr	nr
120 h	nr	nr	nr	nr	nr	nr	0.134	3.527	0.003	0.203	4.411	<0.001
Rainfall (mm)												
0h	116.13	5.411	<0.001	105.47	6.039	<0.001	<i>-0.066</i>	<i>-1.967</i>	<i>0.069</i>	nr	nr	nr
48h	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr
72h	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr
120h	<i>-23.94</i>	<i>-1.970</i>	<i>0.075</i>	-24.15	-2.529	0.028	nr	nr	nr	nr	nr	nr
Average net radiation (W m⁻²)												
0h	4.96	1.286	0.225	nr	nr	nr	-0.021	-4.043	0.001	-0.021	-3.387	0.005
48h	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr
72h	5.58	1.545	0.151	4.04	1.563	0.146	nr	nr	nr	nr	nr	nr
120h	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr	nr
Total suspended solids (mg L⁻¹)												
Downstream	<i>-0.02</i>	<i>-2.102</i>	<i>0.059</i>	NA	NA	NA	nr	nr	nr	NA	NA	NA
Upstream	NA	NA	NA	-0.32	-1.626	0.132	NA	NA	NA	-0.0006	-1.627	0.128

972

973 **Figure captions:**

974 **Fig. 1:** Location and morphometry of thaw slumps on the Peel Plateau, Northwest Territories, Canada.
975 Panel A depicts the stream networks and location of the eight retrogressive thaw slumps studied. Panel
976 B depicts representative sampling locations at each slump site; FM3 depicted. Panels C-E depict
977 representative thaw-slump headwall stratigraphies. Panel C shows a mega-slump (FM3, the smallest
978 mega-slump, is depicted); panel D shows a moderate-sized slump (HB); panel E shows the smallest
979 slump that was sampled (SD). In panels C and D, the approximate location of the modern active layer (a),
980 early Holocene-aged relict active layer (b), and Pleistocene-aged glacial materials (c) is shown. Photo
981 credit: Scott Zolkos.

982 **Fig. 2:** The effect of retrogressive thaw slumps on stream water dissolved organic carbon (DOC)
983 concentration. Each data point represents the mean and standard error of measurements across all
984 sampling dates, as described in Table 1. The bottom two panels show the ratio of within-slump:
985 upstream, and downstream: upstream DOC concentrations within individual slumps, with points
986 indicating the mean and standard error of this ratio across sample dates.

987 **Fig. 3:** Box and whisker plots to illustrate the effects of retrogressive thaw slump activity on stream
988 geochemistry. Each boxplot includes data from across all slumps and sampling periods, and indicates
989 median values, 25th and 75th percentiles (box extremities), 10th and 90th percentiles (whiskers), and
990 outlier points. U=upstream sites; W=within-slump sites; D=downstream sites.

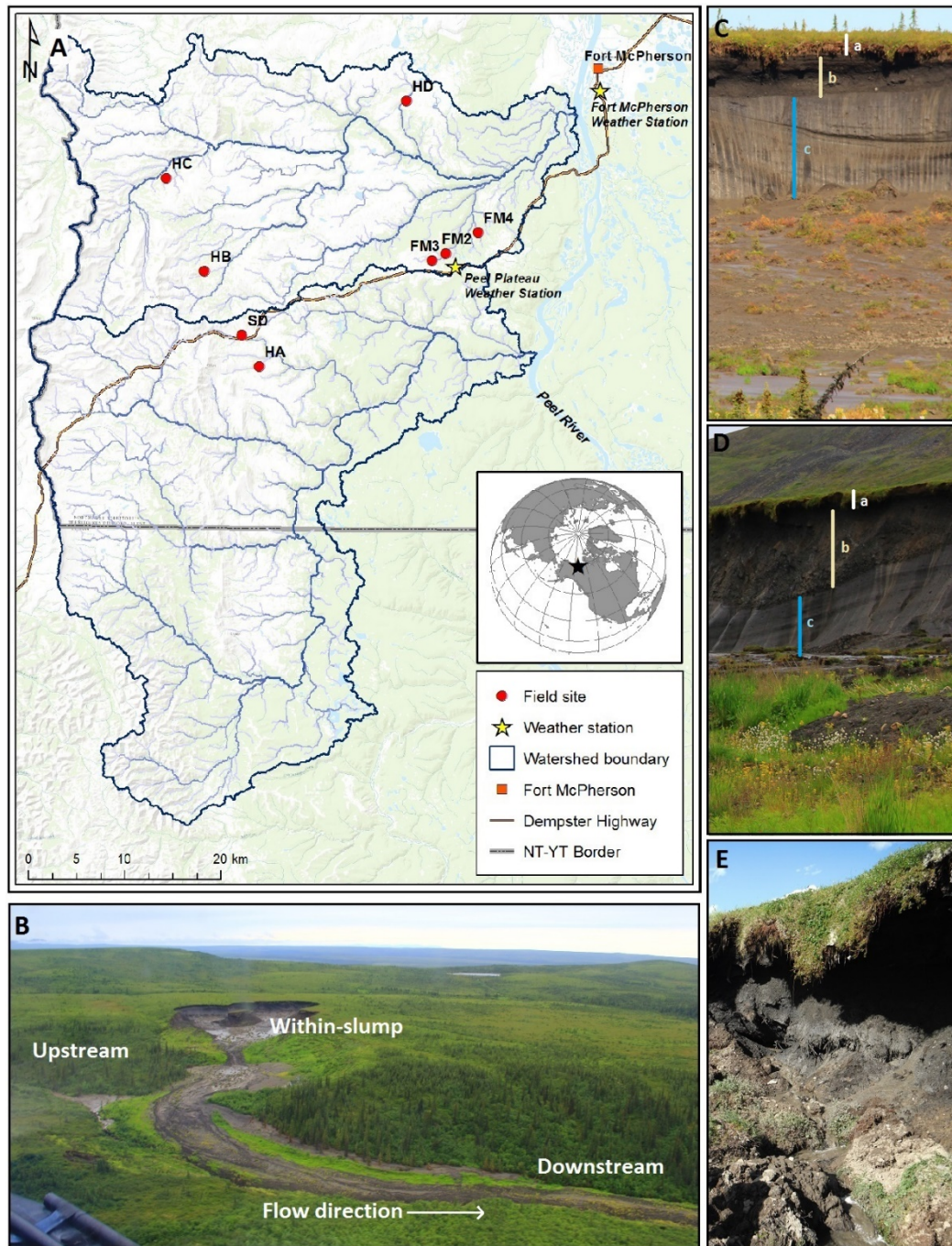
991 **Fig. 4:** The effect of retrogressive thaw slumps on the optical properties of stream water dissolved
992 organic matter. Each data point represents the mean and standard error of measurements across all
993 sampling dates, as described in Table 1. Shown are specific UV absorbance (SUVA₂₅₄), spectral slopes
994 between 275-295 and 350-400 nm ($S_{275-295}$; $S_{350-400}$) and the slope ratio (S_R).

995 **Fig. 5:** Paired oxygen isotopic ($\delta^{18}\text{O}$ ‰) and SUVA₂₅₄ ($\text{L mg C}^{-1} \text{m}^{-1}$) data, to demonstrate the relationship
996 between source water age and dissolved organic matter composition. Reference $\delta^{18}\text{O}$ values are from
997 Lacelle et al. (2013): the modern active layer value is derived from active layer pore water in this region,
998 icy diamicton has been sourced as Holocene in origin, and the $\delta^{18}\text{O}$ value for Pleistocene-aged ground ice
999 is the most positive value for this region.

1000 **Fig. 6:** Environmental conditions (solar radiation, precipitation and mean daily air temperature) and DOC
1001 flux upstream and downstream of slump FM3 across a month-long sample period (July 12-August 12,
1002 2014). Corresponding multiple linear regressions are described in Table 4.

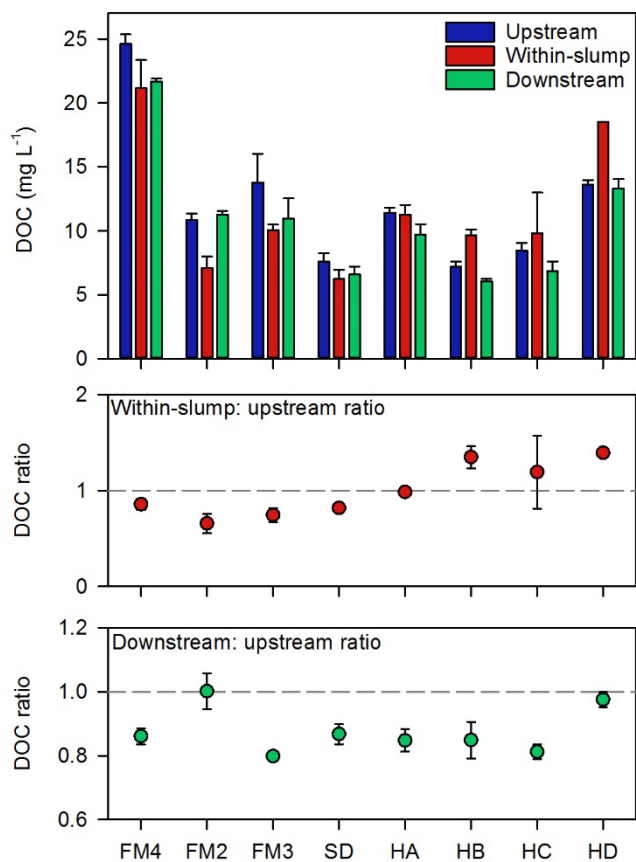
1003 **Fig. 7:** Within-slump fluxes of dissolved organic carbon (DOC) and TSS, compared to the calculated
1004 (downstream - upstream) fluxes for these two constituents.

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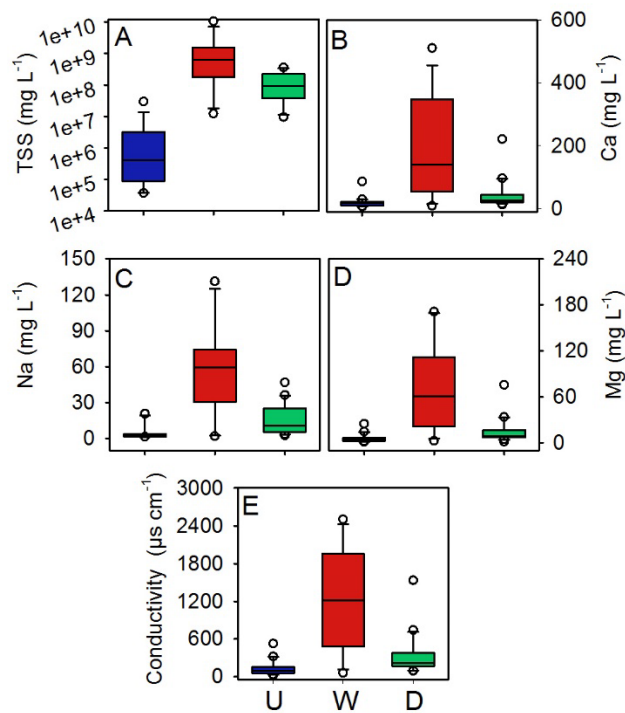
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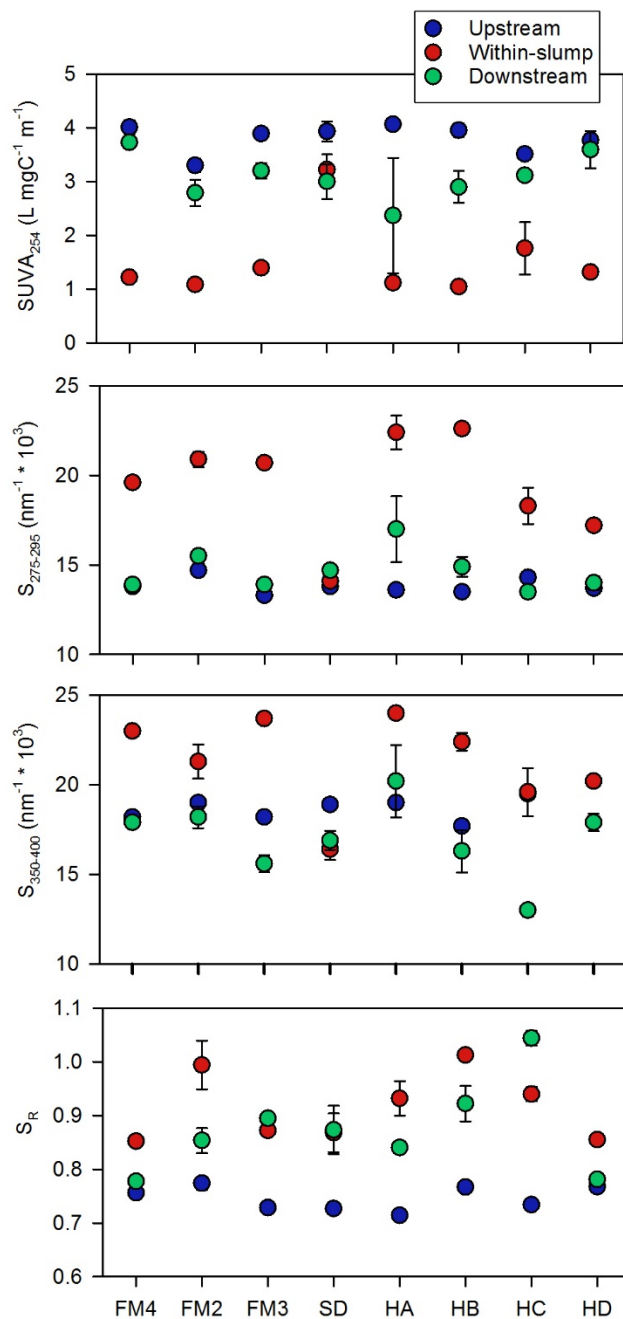
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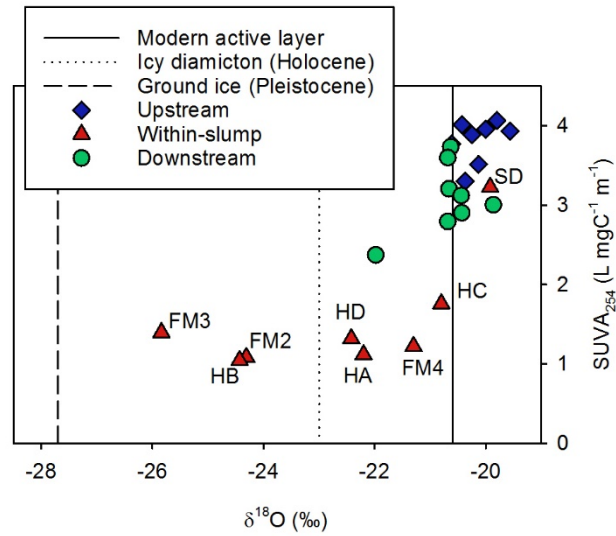
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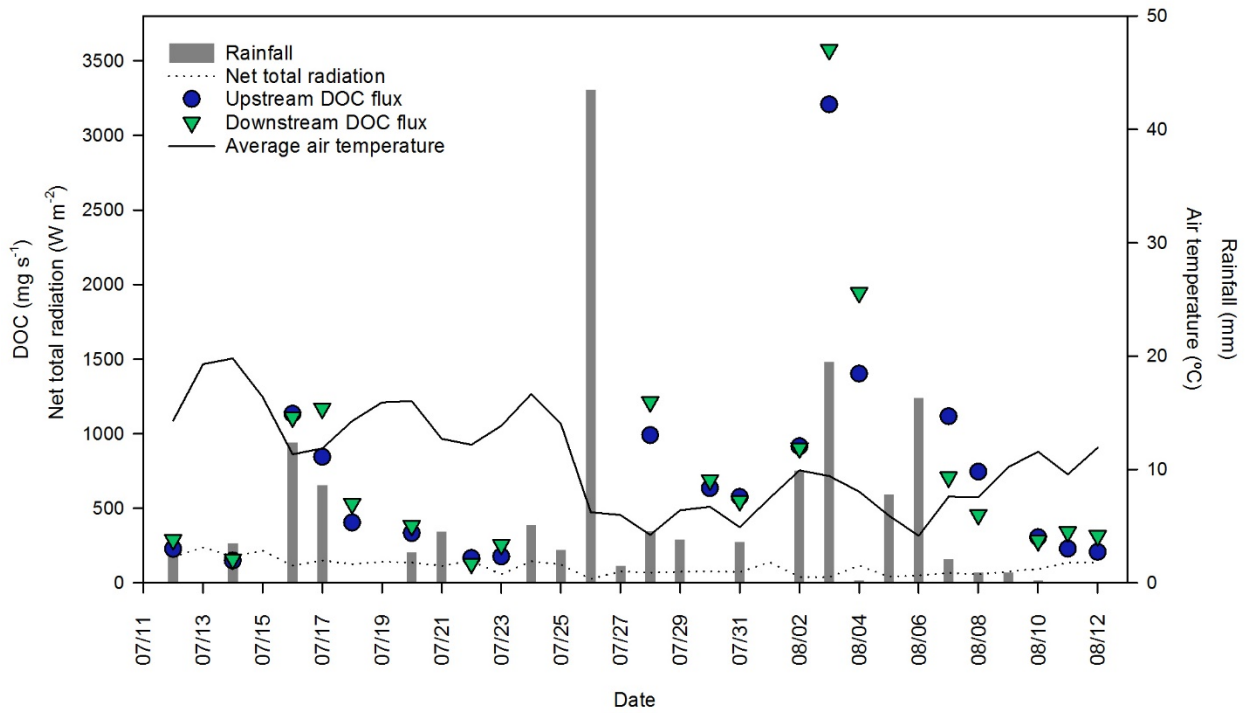
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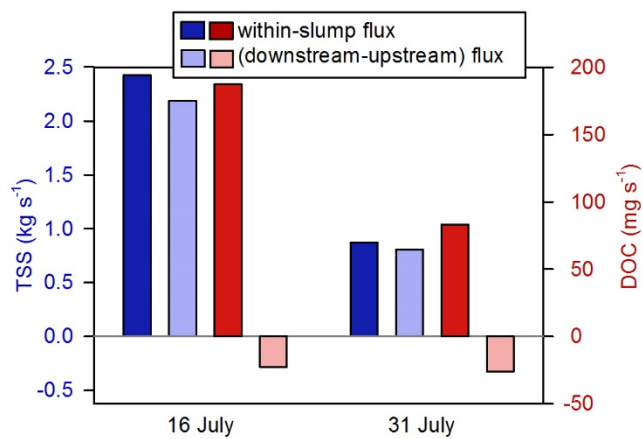
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