

1 Temperature and UV light affect the activity of marine cell- 2 free enzymes

3

4 Blair Thomson¹, Christopher David Hepburn¹, Miles Lamare¹, Federico Baltar^{1,2}

5 ¹Department of Marine Science, University of Otago, New Zealand

6 ²NIWA/University of Otago Research Centre for Oceanography, Dunedin, New Zealand

7 *Correspondence to:* F. Baltar (federico.baltar@otago.ac.nz)

8

9 **Abstract**

10 Microbial extracellular enzymatic activity (EEA) is the rate-limiting step in the degradation of
11 organic matter in the oceans. These extracellular enzymes exist in two forms, cell-bound
12 which are attached to the microbial cell wall, and cell-free which are completely free of the
13 cell. Contrary to previous understanding, cell-free extracellular enzymes make up a
14 substantial proportion of the total marine EEA. Little is known about these abundant cell-free
15 enzymes, including what factors control their activity once they are away from their sites
16 (cells). Experiments were run to assess how cell-free enzymes (excluding microbes) respond
17 to ultraviolet radiation (UVR) and temperature manipulations, previously suggested as
18 potential control factors for these enzymes. The experiments were done with New Zealand
19 coastal waters and the enzymes studied were alkaline phosphatase [APase], β -glucosidase,
20 [BGase], and leucine aminopeptidase, [LAPase]. Environmentally relevant UVR (i.e., *in situ*
21 UVR levels measured at our site) irradiances reduced cell-free enzyme activities up to 87%
22 when compared to controls, likely a consequence of photodegradation. This effect of UVR on
23 cell-free enzymes differed depending on the UVR fraction. Ambient levels of UV radiation
24 (KJ) were shown to reduce the activity of cell-free enzymes, for the first time. Elevated
25 temperatures (15°C) increased the activity of cell-free enzymes up to 53% when compared to
26 controls (10°C), likely by enhancing the catalytic activity of the enzymes. Our results suggest
27 the importance of both UVR and temperature as control mechanisms for cell-free enzymes.
28 Given the projected warming ocean environment and the variable UVR light regime, it is
29 possible there could be major changes in the activity of cell-free EEA and their contribution
30 to organic matter remineralization in the future.

31

32 **1 Introduction**

33 Heterotrophic microbes are ubiquitous in the marine environment, recycling most of the
34 organic matter available in the oceans. The discovery of the microbial loop made clear that
35 heterotrophic microbes are one of the most important nutrient vectors in marine food webs
36 (Azam et al., 1983;Azam and Cho, 1987). According to the size-reactivity model, microbes
37 selectively prefer high molecular weight dissolved organic matter (HMWDOM) due to its
38 superior nutritional value (Amon and Benner, 1996;Benner and Amon, 2015). The main
39 obstacle for use of HMWDOM by microbes is that these compounds are generally too large to
40 be transported across microbial cell membranes. Enzymatic hydrolysis outside of the cell is
41 required to break HMWDOM down to smaller size fractions (<600 Daltons) before uptake
42 can occur (Weiss et al., 1991). Thus, microbial extracellular enzymatic activity (EEA) is the
43 process that initiates the microbial loop (Arnosti, 2011;Hoppe et al., 2002), and is recognised
44 as the rate limiting step in the degradation of organic matter in the oceans (Hoppe, 1991). This
45 key role has led to extracellular enzymes being referred to as “gatekeepers of the carbon
46 cycle” (Arnosti, 2011).

47 There are two forms of EEA; cell-bound, which are attached to the outside of the microbial
48 cell wall or reside in the periplasmic space, and; cell-free, which are completely free of the
49 cell, suspended in the water column. Cell-free enzymes can come from a variety of sources in
50 the marine environment including the sloppy grazing behaviour of protists (Bochdansky et al.,
51 1995;Hoppe, 1991), microbial starvation (Chróst, 1991), the lysis of cells by viruses (Kamer
52 and Rassoulzadegan, 1995) and the direct release by microbes in response to the detection of
53 appropriate substrates (Alderkamp et al., 2007). Up until recently, research on extracellular
54 enzymes has been mostly on cell-bound enzymes, as they were considered to be the only
55 abundant form (Hoppe, 1983;Hoppe et al., 2002). This lead to a view that cell-bound
56 extracellular enzymes were the only form of ecological significance (Chróst and Rai,
57 1993;Rego et al., 1985). However, studies have now shown that the second form, cell-free
58 extracellular enzymes, can make up a substantial proportion of the total extracellular enzyme
59 pool (Baltar et al., 2010;Baltar et al., 2016;Baltar et al., 2013;Allison et al., 2012;Duhamel et
60 al., 2010;Kamer and Rassoulzadegan, 1995;Li et al., 1998). This has been a major conceptual
61 shift for research in marine enzymatic activity, generating new research questions about what

62 controls cell-free enzymes in the marine environment and how they function (Arnosti,
63 2011;Arnosti et al., 2014;Baltar et al., 2010;Baltar et al., 2016).

64 One of the many consequences of this discovery is that cell-free enzymes can be decoupled
65 temporally and/or spatially from the microbial community that produces them (Arnosti,
66 2011;Baltar et al., 2016;Baltar et al., 2010), since cell-free enzymes have long residence times
67 after they are released lasting up to several weeks (Baltar et al., 2013;Steen and Arnosti,
68 2011). The activity of cell-free enzymes away from their sites (cells) can condition
69 macromolecular DOC and organic surfaces for subsequent microbial growth. This action at a
70 distance complicates discerning links between producing microbes and their enzymes
71 expression, as cell-free enzymes have the potential to contribute to the availability of nutrients
72 at a great distance from the releasing cell (Arnosti, 2011;Baltar et al., 2016;Baltar et al.,
73 2010). It has been suggested that the history of the water mass may be more informative in
74 understanding current cell-free enzyme activities than the *in situ* microbial community present
75 at the time of sampling (Baltar et al., 2016;Baltar et al., 2010;Kamer and Rassoulzadegan,
76 1995;Arnosti, 2011;Baltar et al., 2013).

77 There is only a limited number published investigations into the dynamics of cell-free
78 enzymes (Baltar et al., 2010;Baltar et al., 2013;Baltar et al., 2016;Kim et al., 2007;Steen and
79 Arnosti, 2011;Li et al., 1998;Kamer and Rassoulzadegan, 1995;Duhamel et al., 2010). These
80 papers provide good evidence of the importance of cell-free enzymes in the marine
81 environment, but the controls for cell-free enzymes (once separated from the microbial cell)
82 are poorly understood (Arnosti, 2011). Steen and Arnosti (2011) tested the effect of
83 ultraviolet radiation (UVR) on cell-free enzymes directly, finding that a reduction in cell-free
84 enzyme activity only at artificially high UVR doses (i.e., UV-B intensity 5–10 times higher
85 than *in situ*), with natural illumination showing no significant effects of photodegradation.
86 One recent study by Baltar et al. (2016) in the Baltic Sea revealed strong correlations between
87 seasonal temperature change and the proportion of cell-free to total EEA, suggesting seawater
88 temperature and/or solar radiation as the most obvious abiotic mechanisms for the control of
89 cell-free enzymatic activity. However, that was a field study of coastal waters, which includes
90 the whole microbial community and many potential interactions and effects that can co-occur
91 (e.g. production/consumption of free enzymes by microbes, variation in substrate
92 concentration, etc.). Thus, to better understand the factors affecting marine free EEA we need
93 to test the effect of environmental factors on free EEA under controlled conditions.

94 Here we isolated the free extracellular enzymes from a coastal site and specifically studied the
95 effects of temperature and UVR on the activity of three cell-free extracellular enzyme groups;
96 alkaline phosphatase (APase), an enzyme used to acquire phosphorus from organic molecules;
97 β -glucosidase (BGase), a glycolytic enzyme that targets carbohydrates groups, and; leucine
98 aminopeptidase (LAPase), an enzyme associated with the degradation of proteins. UVR
99 treatments were hypothesised to reduce the activity of cell-free enzymes when compared to
100 dark controls by photodegradation, with a ‘high UVR dose’ treatments (including the entire
101 UV-B spectrum (280 to 320 nm) were hypothesized to have a stronger degradative effect on
102 cell-free enzymes than ‘low UVR dose’ treatments (which only include a fraction of the UV-
103 B spectrum, 280 to 305 nm). This was based on the reported effects of UV-B on microbes and
104 their metabolic rates including the total EEA (Herndl et al., 1993; Santos et al., 2012; Müller-
105 Niklas et al., 1995; Demers, 2001). Compared to ambient temperatures (10°C), cell-free
106 enzymes exposed to high temperatures (15°C) were hypothesised to be more active, and
107 *viceversa*, due to the general relationship between temperature and catalytic activity in
108 enzymes (Daniel and Danson, 2010, 2013). Experiments carried out here are the first to
109 directly test temperature effects on cell-free enzymes alone, and to directly test the effect of
110 UVR on cell-free enzymes in the Southern Hemisphere and under *in situ* measured
111 environmental-relevant UV-irradiances.

112

113 **2 Materials and methods**

114 **2.1 Study site, sampling and experiments preparation**

115 The experiments were conducted at the University of Otago’s Portobello Marine Laboratory,
116 situated on the Otago Harbour, Dunedin, New Zealand (45.8281° S, 170.6399° E). Otago
117 Harbour is a tidal inlet which has an area of 46 km², consisting of two basins and with
118 extensive sediment flats (Grove and Probert, 1999; Heath, 1975). The laboratory is based on
119 the outer Otago harbour, which has waters similar in composition to coastal seawater, owing
120 to the rapid residence times for its waters exchanging with the open sea (Rainer, 1981; Grove
121 and Probert, 1999). Samples were taken from the second meter of the water column off the
122 marine laboratory’s wharf that extends into a deep tidal channel. All sampling and laboratory
123 equipment used was prior decontaminated using triplicate rinses of 18 M Ω ·cm high purity
124 water (Milli-QTM) water before and after soaking in 10% hydrochloric acid for >6 hours and

125 oven dried at 60°C. To separate the cell-free extracellular enzymes from the total
126 extracellular enzyme pool and the microbial community, samples were triple filtered through
127 low protein binding 0.22µm Acrodisc filters following published methods (Kim et al.,
128 2007;Baltar et al., 2010). 50 ml glass vials were filled with the 0.22 µm-filtered seawater for
129 use in experiments. Bacterial abundance was determined after both experiments by preserving
130 samples in glutaraldehyde and processing using SYBR Green nucleic acid stain with a BD
131 Accuri C6 flow cytometer (BD biosciences, USA). This was to ensure that no significant
132 bacterial growth occurred after filtering or during the incubation. Bacterial abundance was
133 reduced to less than 1% of the pre-filtered total and remained so during the 36-hour
134 incubations.

135 **2.2 UVR experiments**

136 To determine *in situ* UVR irradiance and environmentally appropriate treatments for
137 experiments, the attenuation of UVR was measured through the upper 2 m of the water
138 column on site using a LI-COR LI1800UW spectroradiometer (LI-COR biosciences, USA.
139 The spectroradiometer was factory calibrated using NIST traceable standards. Once this was
140 determined, artificial lighting was installed in a controlled temperature room, set to the
141 ambient seawater temperature (10°C). The lighting consisted of two FS20 UV-R lamps
142 (General Electric, Schenectady NY, USA) and a full spectrum Vita-Lite 72 (Duro-Test,
143 Philadelphia, PA, USA) lamp, suspended above the samples. These lights were height
144 adjusted to yield an irradiance of 3.03 W m⁻² s⁻¹ UVR, approximating UVR irradiances
145 measured in the field at 2 m depth (3.5 W m⁻² s⁻¹). Schott WG and GG long pass filters (15
146 cm X 15 cm) with nominal cutoffs (50% T) in the UVB (280 nm, 305 nm) were placed above
147 the filtered cell-free enzyme seawater samples contained in glass vials, with either a ‘high
148 dose’ (<280nm, 3.03 W m⁻² s⁻¹, 130.8 kJ) or a ‘low dose’ (<305nm, 0.42 W m⁻², 18.1 kJ) of
149 UVR. All light was blocked except that which passed directly through the long pass filters
150 onto the open glass vials, to avoid any effect of the glass on the UVR dose. Controls were
151 kept without light by wrapping the glass vials containing the filtered cell-free enzyme
152 seawater samples in several layers of aluminium foil, and were placed in the same controlled
153 temperature room. Readings of enzyme activity rates were taken of three replicates of each
154 treatment at 12 and 36 hours. UVR was not applied directly to the plate incubations, as it can
155 affect the fluorogenic substrate analogues used in the assays. Temperature inside the vials was
156 also monitored to ascertain that the samples were constantly kept at the desired temperature.

157 **2.3 Temperature experiments**

158 For the temperature experiments we utilised a large graded heat block system (see Lamare et
159 al. (2014) for design specifications). This heat block allowed for up to 15 replicate samples to
160 be exposed to constant temperature treatments over time. The heat blocks were tested five
161 times a day for three days in advance with blank samples to ensure the heat blocks were
162 calibrated accurately; the variation in temperature was within 0.5°C of the target temperatures
163 (i.e., 5, 10, and 15°C) in all measurements. These temperatures were selected because 5 to 15
164 °C is the annual range of temperature in the sampling site, and 10°C was the *in situ*
165 temperature at the time of sampling (unpublished data). All treatments were kept in the dark
166 by wrapping the glass vials containing the filtered cell-free enzyme seawater samples in
167 several layers of aluminium foil. Readings of enzyme activity rates were taken of three
168 replicates of each treatment at 6, 12, 24 and 36 hours. When incubating these samples, each
169 was put into a separate incubator which was set to the respective treatment temperature so to
170 avoid confounding the temperature treatments.

171 **2.4 Extracellular enzymatic activities assays**

172 We used the method for assessing extracellular enzymatic activity rates based on the
173 hydrolysis of fluorogenic substrate analogues developed by Hoppe (1983). The fluorogenic
174 substrates: 4-methylcoumarinyl-7-amide (MCA)-L-leucine-7-amido-4-methylcoumarin, 4-
175 methylumbelliferyl (MUF)-β-D-glucoside and MUF-phosphate were used to assess the
176 leucine aminopeptidase, β-glucosidase and alkaline phosphatase activities, respectively.
177 Substrate concentrations of 100μM were used for each enzyme based on pre-established
178 kinetics, tested in the lab. Although differences in UVR or temperature might affect the
179 kinetic parameters, we decided to use the same concentration for all the enzymes (which was
180 saturating at the *in situ* conditions) to allow for a better comparison and reduce confounding
181 factors. 96-well falcon microplates were filled with six replicates of each of the three
182 fluorogenic substrates (10μl) and seawater (290μl) to make up 300μl reactions. Plates were
183 read in a Spectramax M2 spectrofluorometer (Molecular Devices, USA), with excitation and
184 emission wavelengths of 365 and 445nm, both before, and after 3 hour incubations. All
185 incubations were performed in the dark with UVR incubations set to the *in situ* seawater
186 temperature, and temperature incubations set to each respective treatment temperature. Six
187 samples without substrate addition served as blanks in each plate to determine the background

188 fluorescence of the samples, which were used to correct the activity rates in the plate readings
189 before and after incubation.

190 **2.5 Statistical analyses**

191 In all analyses, parametric assumptions were first checked using the Shapiro-Wilk test for
192 normality and the Levene's test for equal variance. Where appropriate, data was Log-
193 transformed to meet normality assumptions prior to analysis. Both experiments use two-way
194 ANOVAs with an interaction term, with post hoc Tukey HSD tests run to assess the
195 individual significant effects between treatments. All analyses were run in the R software
196 environment (R Development Core Team, Austria).

197

198 **3 Results and Discussion**

199 **3.1 UVR experiments revealed photodegradation of cell-free enzymatic activities at** 200 **environmentally relevant levels**

201 The proportion of cell-free EEA in the seawater at the time of sampling was 93.3%, 96.7 and
202 21.1% for APase, BGase and LAPase respectively (Table 1). UVR overall significantly
203 decreased cell-free APase when compared to dark controls ($p < 0.001$, $F_{2,12} = 15.85$, two-way
204 ANOVA) (Fig. 1a). Individual significant effects between treatments in APase were seen as a
205 significant decrease in activity in the low-dose treatment relative to the dark control at 12 h
206 ($p < 0.05$, Tukey HSD), and between the dark control and both the high and low UV-dose
207 treatment at the 36-hour sampling point ($p < 0.05$, Tukey HSD). BGase cell-free activity was
208 not significantly affected by UVR ($p = 0.53$, $F_{2,12} = 0.67$, two-way ANOVA). UVR had a
209 significant overall effect on LAPase, decreasing the cell-free activity when compared to dark
210 controls ($p < 0.01$, $F_{2,12} = 40.994$, two-way ANOVA) (Fig. 1c). Individual significant effects
211 were seen in LAPase, showing after 12 h a significant decrease in activity between the low
212 and high at 12 h ($p < 0.01$, Tukey HSD), and after 36 h a gradual decrease from high to low
213 dose ($p < 0.05$, Tukey HSD), and dark control to both low and high dose ($p < 0.001$, Tukey
214 HSD). Changes observed in the controls of all the enzymes from 12 to 36 h were not
215 statistically significant ($p > 0.05$, Tukey HSD).

216 Apart from the possibility that UVR treatments may have influenced the composition of the
217 seawater substrate itself, these experiments revealed a significant reduction in cell-free

218 extracellular enzymatic activity for both APase and LAPase in response to UVR, consistent
219 with the predicted photodegradation; which was not evident for BGase. This was the first time
220 that UVR has been demonstrated to reduce cell-free enzymatic activities at environmentally
221 relevant intensities. The only previous study (Steen and Arnosti, 2011) did show a reduction
222 in the cell-free extracellular enzymatic activity of APase and LAPase but only at artificially
223 high UVR intensities where UV-B was 5–10 times more intense from artificial lamps in the
224 lab than outdoors. Interestingly, they could not show significant UVR effects on BGase at
225 any treatment level, which is consistent with the present study.

226 Both APase and LAPase showed the strongest effect of UVR at the 36-hour sampling point,
227 suggesting a UV-B dose-dependent response. LAPase also showed a gradual decrease in the
228 effect between the low and high UVR treatments, which suggests the increase in UV-B
229 irradiances also enhanced the degree of photodegradation. UV-B has been demonstrated to be
230 a highly active part of the spectrum for degrading DNA in general (Sinha and Häder,
231 2002;Dahms and Lee, 2010), which is not only included in cells but is also an abundant
232 component of the dissolved (extracellular) seawater fraction (Paul et al., 1987;Paul and
233 David, 1989). Specific effects of UV-B on total extracellular enzymatic activities have been
234 previously reported (Herndl et al., 1993;Santos et al., 2012;Demers, 2001;Müller-Niklas et al.,
235 1995). However, it is important to distinguish these previous studies from the cell-free
236 enzyme experiments performed here. Those previous studies tested the response of the entire
237 microbial community, for total extracellular enzymatic activity, based on the assumption that
238 UVR affects the organism (source of enzymes) directly. What is shown in this study is that
239 UVR affects cell-free exclusively without the need to impact the source organism. The effects
240 of UVR were different among the enzymes assessed, which may be of importance as some
241 enzymes could be more impacted by UVR than others. For example, in this study, APase and
242 LAPase were more affected by UVR than BGase, which could change the spectrum of
243 extracellular enzyme activity in the surface of the ocean. The resulting higher BGase relative
244 to APase or LAPase, could potentially condition macromolecular DOC composition by
245 hydrolysing relatively less proteins than carbohydrates in response to UV. In turn, it is
246 conceivable that any change in the enzyme spectrum due to variability in UVR light could
247 cause a loss of productivity (e.g. due to a decrease in the inorganic P made available through
248 APase activities), as the nutrients made available by extracellular enzymes may not be in
249 suitable ratios for the effective growth of microbes (Arnosti et al., 2014;Häder et al., 2007).

250 **3.2 Temperature experiments revealed enhanced catalytic activity of cell-free enzymes**

251 The proportion of cell-free EEA in the original seawater at the time of sampling was 99.9%,
252 85.8 and 30.0% for APase, BGase and LAPase respectively (Table 1). Temperature
253 significantly increased cell-free APase at the high temperature of 15°C when compared to the
254 ambient control of 10°C ($p < 0.01$, $F_{2,24} = 11.57$, two-way ANOVA) (Fig. 2a). APase activity
255 was significant increased, after 6 h, in the high relative to the low temperature ($p < 0.001$,
256 Tukey HSD), after 12 h between low and high temperature ($p < 0.001$, Tukey HSD), and
257 control and high treatments ($p < 0.05$, Tukey HSD). Cell-free BGase showed a similar pattern
258 of increased activity in response to higher temperature but it was not significant (Fig 2b). This
259 lack of significant differences in cell-free BGase in response to temperature could be due to a
260 relatively high variability in EEA among the high temperature (15°C) treatments. LAPase
261 significantly decreased in the low temperature treatment (5°C), relative to the ambient control
262 ($p < 0.01$, $F_{2,24} = 13.97$, two-way ANOVA) (Fig 2c). LAPase cell-free activity significantly
263 increased between the low and high temperature treatments at the 6h and 12h time points
264 ($p < 0.05$, Tukey HSD). The temperature effect was dependent on time, finding significant
265 effects after 6 and 12h, but not later for any of the studied enzymes.

266 The relationship found between temperature and cell-free activity is consistent with the
267 general pattern of increased catalytic activity of enzymes in relation to temperature (Daniel
268 and Danson 2013). The positive relationship between temperature and the activity of cell-free
269 enzymes observed in this study is contrary to the negative relationship between temperature
270 and the proportion of cell-free relative to total EEA measured in a seasonal field study in the
271 Baltic Sea (Baltar et al., 2016). However, it is important to take into consideration the fact that
272 the study by Baltar et al. (2016) took place over a much longer temporal scale (1.5 years) and
273 included the whole microbial community; whereas in this study different factors were teased
274 apart by focusing only on the cell-free enzymes. This is supported by Baltar et al. (2016)
275 where the proportion of cell-free relative to total EEA was significantly negatively correlated
276 to prokaryotic heterotrophic production, suggesting that the low temperature preserves the
277 constitutive activity of the cell-free enzymes better (than warm temperature) due to a
278 reduction in the metabolism of heterotrophic microbes that would reduce the
279 consumption/degradation of dissolved enzymes. The exclusion of heterotrophic microbes
280 from our samples precluded this effect (i.e., heterotrophic degradation/consumption of free
281 enzymes) of temperature from occurring, and allowed us to tease apart the effect directly on

282 the cell-free enzymatic activities. This also highlights the importance of scales when dealing
283 with microbial oceanographic processes.

284

285 **Conclusions**

286 Overall, temperature and UVR were both demonstrated as potential control mechanisms for
287 the activity of marine cell-free enzymes, providing a baseline for future research. This is the
288 first report revealing the effects of photodegradation of cell-free enzymes at environmentally
289 relevant levels of UVR, and the effects of enhanced temperature on the catalytic activity of
290 marine cell-free enzymes. Environmentally relevant UVR had a significant photodegradative
291 effect that might be enzyme-specific (affecting APase and LAPase but not BGase), with the
292 potential to alter not only the rates of cell-free EEA but also the spectrum of enzyme
293 expression in the seawater. Alteration of the cell-free EEA spectrum from UVR variability,
294 could have ecological and biogeochemical implications like the conditioning of
295 macromolecular DOM (i.e., affecting DOM composition by hydrolysing some DOM
296 compounds more relative to others), and the change of the elemental ratio of some nutrients
297 (e.g., affecting the availability of inorganic P due to a change in APase activity), with
298 implications for productivity and nutrient cycling. Additionally, given the variable UVR light
299 regime spatially and temporally (i.e. the 150% increase in UV-B in polar regions during
300 spring-time ozone depletion, Smith et al., 1992) and the documented anthropogenic changes
301 in ocean temperature (Chen et al., 2007), it is probable that the activity of cell-free EEA and
302 their contribution to organic matter remineralization might be affected in the future, if not
303 already.

304

305 **Acknowledgements**

306 We would like to thank the team of technicians out at Portobello Marine Laboratory, most
307 notably, Linda Groenewegen and Reuben Pooley. This research was supported by a
308 University of Otago Research Grant and a Rutherford Discovery Fellowship (Royal Society
309 of New Zealand) to FB. We would like to acknowledge the support and insightful comments
310 of the reviewers, which clearly helped improve the overall merit of the manuscript. The
311 authors declare that they have no conflict of interest.

312 **References**

- 313 Alderkamp, A. C., van Rijssel, M., and Bolhuis, H.: Characterization of marine bacteria and
314 the activity of their enzyme systems involved in degradation of the algal storage glucan
315 laminarin, *FEMS Microbiol Ecol*, 59, 108-117, 10.1111/j.1574-6941.2006.00219.x, 2007.
- 316 Allison, S. D., Chao, Y., Farrara, J. D., Hatosy, S., and Martiny, A.: Fine-scale temporal
317 variation in marine extracellular enzymes of coastal southern California, *Frontiers in*
318 *Microbiology*, 3, 10.3389/fmicb.2012.00301, 2012.
- 319 Amon, R. M. W., and Benner, R.: Bacterial utilization of different size classes of dissolved
320 organic matter, *Limnology and Oceanography*, 41, 41-51, 10.4319/lo.1996.41.1.0041, 1996.
- 321 Arnosti, C., Bell, C., Moorhead, D., Sinsabaugh, R., Steen, A., Stromberger, M., Wallenstein,
322 M., and Weintraub, M.: Extracellular enzymes in terrestrial, freshwater, and marine
323 environments: perspectives on system variability and common research needs,
324 *Biogeochemistry*, 117, 5-21, 10.1007/s10533-013-9906-5, 2014.
- 325 Azam, F., Fenchel, T., Field, J., Gray, J., Meyer-Reil, L., and Thingstad, F.: The Ecological
326 Role of Water-Column Microbes in the Sea, *Marine Ecology Progress Series*, 10, 257-263,
327 10.3354/meps010257, 1983.
- 328 Azam, F., and Cho, B.: Bacterial utilization of organic matter in the sea, *Symposia of the*
329 *Society for General Microbiology*(Cambridge). 1987., 1987,
- 330 Baltar, F., Arístegui, J., Gasol, J. M., Sintes, E., van Aken, H. M., and Herndl, G. J.: High
331 dissolved extracellular enzymatic activity in the deep Central Atlantic Ocean, *Aquatic*
332 *Microbial Ecology*, 58, 287-302, 10.3354/ame01377, 2010.
- 333 Baltar, F., Arístegui, J., Gasol, J., Yokokawa, T., and Herndl, G.: Bacterial Versus Archaeal
334 Origin of Extracellular Enzymatic Activity in the Northeast Atlantic Deep Waters, *Microbial*
335 *Ecology*, 65, 277-288, 10.1007/s00248-012-0126-7, 2013.
- 336 Baltar, F., Legrand, C., and Pinhassi, J.: Cell-free extracellular enzymatic activity is linked to
337 seasonal temperature changes: a case study in the Baltic Sea, *Biogeosciences*, 13, 2815-2821,
338 2016.
- 339 Bochdansky, A. B., Puskaric, S., and Herndl, G.: Influence of zooplankton grazing on free
340 dissolved enzymes in the sea, *Marine Ecology Progress Series*, 121, 53-63,
341 10.3354/meps121053, 1995.
- 342 Chen, Z., Marquis, M., Averyt, K. B., Tignor, M., and Miller, H.: *Climate change 2007: the*
343 *physical science basis. Contribution of Working Group I to the Fourth Assessment Report of*
344 *the Intergovernmental Panel on Climate Change*, Cambridge: Cambridge University, 2007.
- 345 Chróst, R., and Rai, H.: Ectoenzyme activity and bacterial secondary production in nutrient-
346 impoverished and nutrient-enriched freshwater mesocosms, *Microbial Ecology*, 25, 131-150,
347 10.1007/BF00177191, 1993.
- 348 Dahms, H.-U., and Lee, J.-S.: UV radiation in marine ectotherms: Molecular effects and
349 responses, *Aquatic Toxicology*, 97, 3-14, 10.1016/j.aquatox.2009.12.002, 2010.
- 350 Daniel, R. M., and Danson, M. J.: A new understanding of how temperature affects the
351 catalytic activity of enzymes, *Trends in Biochemical Sciences*, 35, 584-591,
352 10.1016/j.tibs.2010.05.001, 2010.

- 353 Demers, S.: The Responses of a Natural Bacterioplankton Community to Different Levels of
354 Ultraviolet-B Radiation: A Food Web Perspective, *Microbial Ecology*, 41, 56-68, 2001.
- 355 Duhamel, S., Dyhrman, S. T., and Karl, D. M.: Alkaline phosphatase activity and regulation
356 in the North Pacific Subtropical Gyre, *Limnology and Oceanography*, 55, 1414-1425,
357 10.4319/lo.2010.55.3.1414, 2010.
- 358 Grove, S., and Probert, P. K.: Sediment macrobenthos of upper Otago Harbour, New Zealand,
359 *New Zealand Journal of Marine and Freshwater Research*, 33, 469-480,
360 10.1080/00288330.1999.9516892, 1999.
- 361 Häder, D. p., Kumar, H. D., Smith, R. C., and Worrest, R. C.: Effects of solar UV radiation on
362 aquatic ecosystems and interactions with climate change, *Photochemical & Photobiological
363 Sciences*, 6, 267-285, 10.1039/b700020k, 2007.
- 364 Heath, R. A.: Stability of some New Zealand coastal inlets, *New Zealand Journal of Marine
365 and Freshwater Research*, 9, 449-457, 10.1080/00288330.1975.9515580, 1975.
- 366 Herndl, G. J., Müller-Niklas, G., and Frick, J.: Major role of ultraviolet-B in controlling
367 bacterioplankton growth in the surface layer of the ocean, *Nature*, 361, 717-719, 1993.
- 368 Hoppe, H.-G.: Microbial extracellular enzyme activity: a new key parameter in aquatic
369 ecology, in: *Microbial enzymes in aquatic environments*, Springer, 60-83, 1991.
- 370 Hoppe, H. G.: Significance of exoenzymatic activities in the ecology of brackish water:
371 measurements by means of methylumbelliferyl-substrates, *Marine Ecology Progress Series*,
372 11, 299-308, 10.3354/meps011299, 1983.
- 373 Kamer, M., and Rassoulzadegan, F.: Extracellular enzyme activity: Indications for high short-
374 term variability in a coastal marine ecosystem, *Microbial Ecology*, 30, 143-156,
375 10.1007/BF00172570, 1995.
- 376 Kim, C., Nishimura, Y., and Nagata, T.: High potential activity of alkaline phosphatase in the
377 benthic nepheloid layer of a large mesotrophic lake: implications for phosphorus regeneration
378 in oxygenated hypolimnion, *Aquatic Microbial Ecology*, 49, 303-311, 10.3354/ame01137,
379 2007.
- 380 Lamare, M., Pecorino, D., Hardy, N., Liddy, M., Byrne, M., and Uthicke, S.: The thermal
381 tolerance of crown-of-thorns (*Acanthaster planci*) embryos and bipinnaria larvae:
382 implications for spatial and temporal variation in adult populations, *Coral Reefs*, 33, 207-219,
383 10.1007/s00338-013-1112-3, 2014.
- 384 Li, H., Veldhuis, M., and Post, A.: Alkaline phosphatase activities among planktonic
385 communities in the northern Red Sea, *Marine Ecology Progress Series*, 173, 107-115,
386 10.3354/meps173107, 1998.
- 387 Müller-Niklas, G., Heissenberger, A., Puskaric, S., and Herndl, G.: Ultraviolet-B radiation
388 and bacterial metabolism in coastal waters, *Aquatic Microbial Ecology*, 9, 111-116,
389 10.3354/ame009111, 1995.
- 390 Paul, J. H., Jeffrey, W. H., and DeFlaun, M. F.: Dynamics of extracellular DNA in the marine
391 environment, *Applied and Environmental Microbiology*, 53, 170, 1987.
- 392 Paul, J. H., and David, A. W.: Production of extracellular nucleic acids by genetically altered
393 bacteria in aquatic-environment microcosms. [*Escherichia coli*, *Pseudomonas aeruginosa*,
394 *Pseudomonas cepacia*, *Bradyrhizobium japonicum*], *Applied and Environmental
395 Microbiology*, 55, 1989.

396 Rainer, S. F.: Soft-bottom benthic communities in Otago Harbour and Blueskin Bay, New
397 Zealand, Blueskin Bay, New Zealand, Dept. of Scientific and Industrial Research,
398 Wellington], 1981.

399 Rego, J. V., Billen, G., Fontigny, A., and Somville, M.: Free and attached proteolytic activity
400 in water environments, *Mar. Ecol. Prog. Ser.*, 21, 245-249, 1985.

401 Santos, A. L., Oliveira, V., Baptista, I. s., Henriques, I., Gomes, N. C. M., Almeida, A.,
402 Correia, A., and Cunha, A.: Effects of UV-B Radiation on the Structural and Physiological
403 Diversity of Bacterioneuston and Bacterioplankton, *Applied and Environmental*
404 *Microbiology*, 78, 2066, 2012.

405 Smith, R. C., Prézelin, B. B., Baker, K. S., Bidigare, R. R., Boucher, N. P., Coley, T.,
406 Karentz, D., Macintyre, S., Matlick, H. A., Menzies, D., Ondrusek, M., Wan, Z., and Waters,
407 K. J.: Ozone Depletion: Ultraviolet Radiation and Phytoplankton Biology in Antarctic Waters,
408 *Science*, 255, 952-959, 1992.

409 Steen, A. D., and Arnosti, C.: Long lifetimes of β -glucosidase, leucine aminopeptidase, and
410 phosphatase in Arctic seawater, *Marine Chemistry*, 123, 127-132,
411 [10.1016/j.marchem.2010.10.006](https://doi.org/10.1016/j.marchem.2010.10.006), 2011.

412 Weiss, M. S., Abele, U., Weckesser, J., Welte, W., Schiltz, E., and Schulz, G. E.: Molecular
413 Architecture and Electrostatic Properties of a Bacterial Porin, *Science*, 254, 1627-1630, 1991.

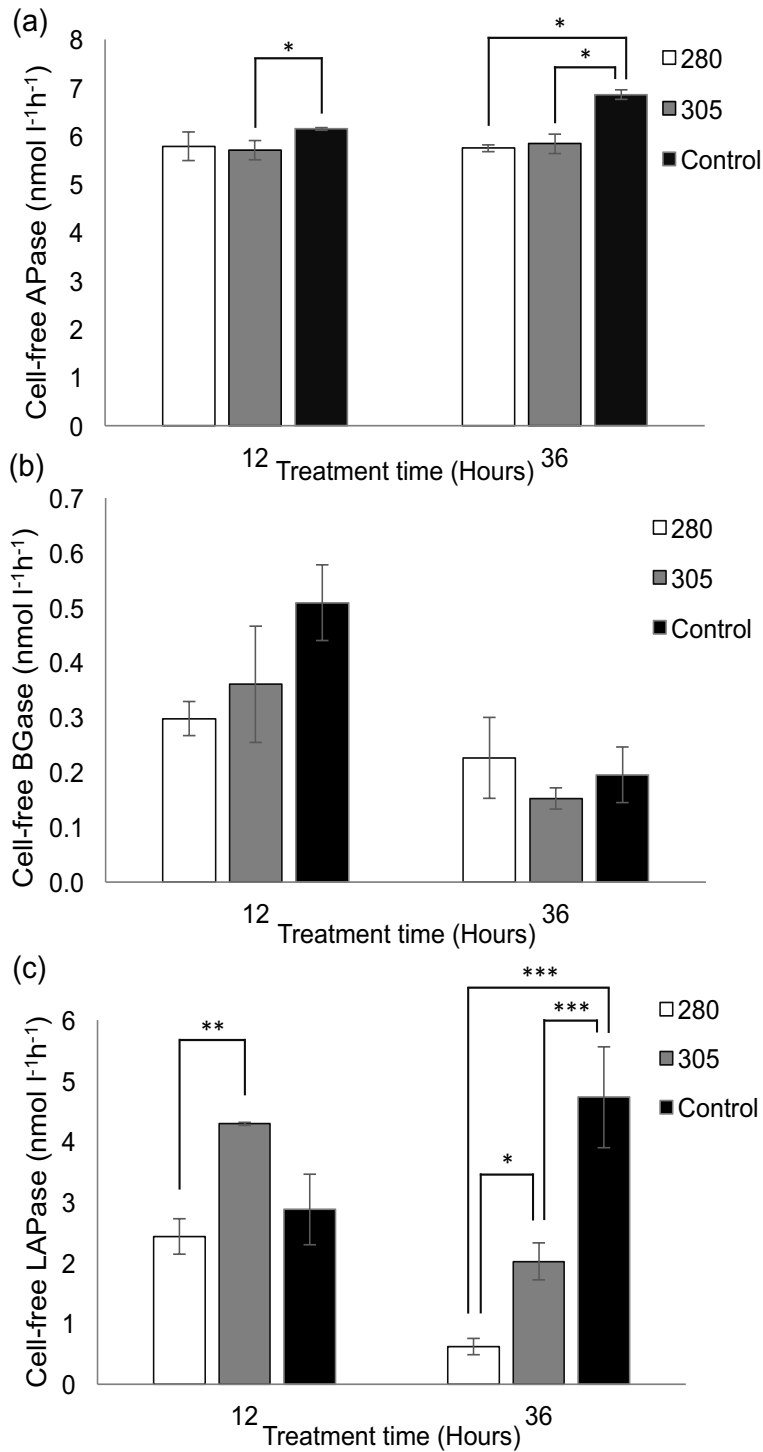
414

415 **Table 1.** Total, cell-free and proportion of cell-free relative to total extracellular enzymatic activity *in*
 416 *situ* for the seawater collected for the UVR and Temperature experiments at the time of sampling.
 417

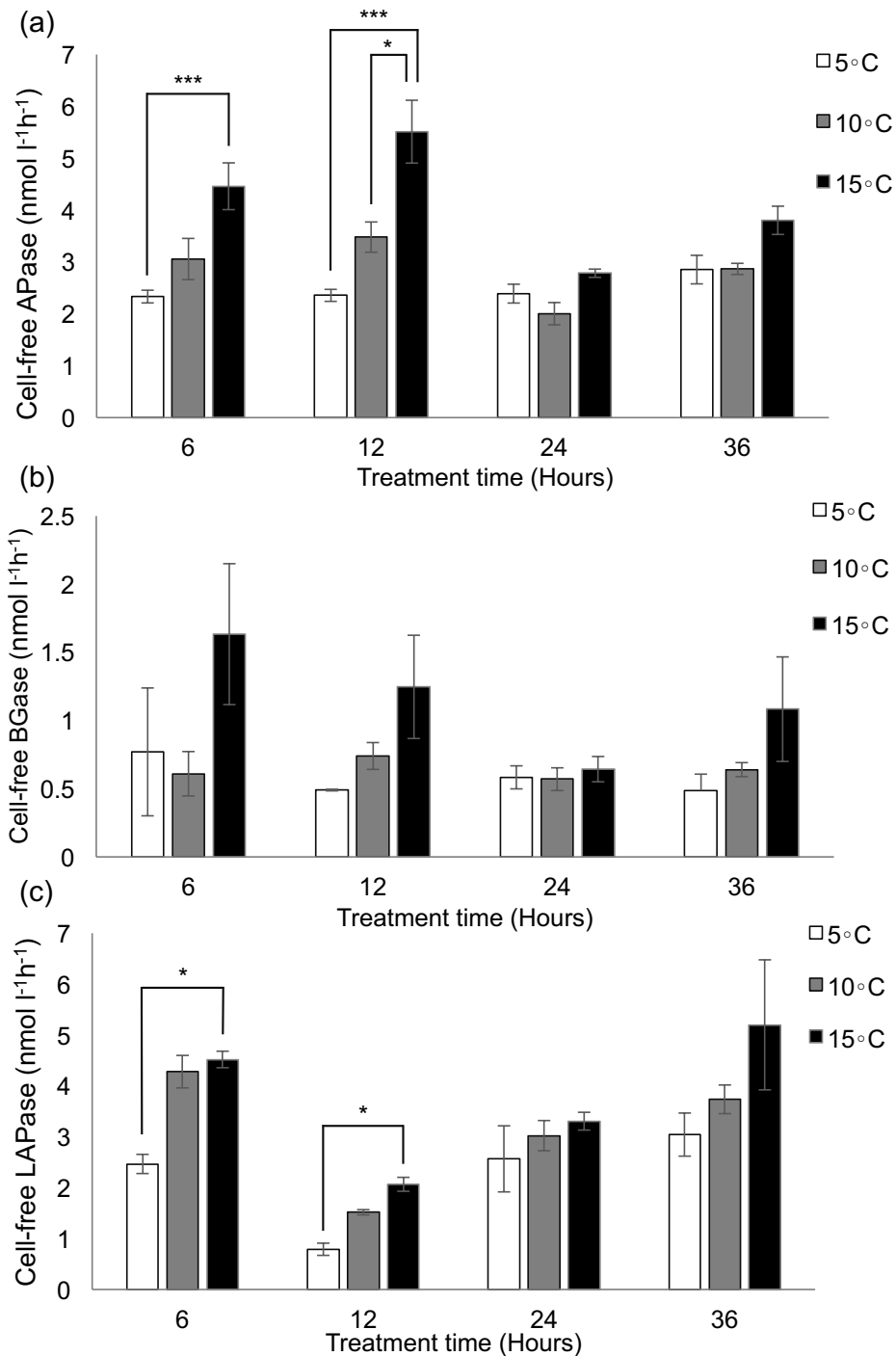
	Total (nmol l⁻¹ h⁻¹)	Cell-free (nmol l⁻¹ h⁻¹)	% Cell-free (%)
UVR experiment			
APase	75.4	70.3	93.3
BGase	2.3	2.2	96.7
LAPase	35.4	7.5	21.1
Temperature experiment			
APase	121.9	121.8	99.9
BGase	3.2	2.7	85.8
LAPase	33.1	9.9	30.0

418

419



433 **Figure 1.** Results from UVR experiments showing the mean (\pm SE) cell-free extracellular enzyme
 434 activity for alkaline phosphatase (a), beta-glucosidase (b), and leucine aminopeptidase (c), under a
 435 high dose (of 280nm and above) and a low dose (of 305nm and above) in comparison to dark controls.
 436 Asterisks above graphs represent individual significant effects between treatments in post hoc Tukey
 437 test (* $<$ 0.05, ** $<$ 0.01, *** $<$ 0.001) (N=3).



438

439 **Figure 2.** Results from temperature modification experiments showing the mean (\pm SE) cell-free
 440 extracellular enzyme activity for alkaline phosphatase (a), beta-glucosidase (b), and leucine
 441 aminopeptidase (c), under a high (15°C) and a low temperature (5°C) treatments in comparison to
 442 ambient controls (10°C). Asterisks above graphs represent individual significant effects between
 443 treatments in post hoc Tukey test (* $<$ 0.05, ** $<$ 0.01, *** $<$ 0.001) (N=3).