Authors comment

Again, a great thanks to the associate editor and the two referees for their work throughout the review process that greatly improved the manuscript.

Some confusion around the atmospheric nutrient forcing lead to a mistake in the previous version of the manuscript. The atmospheric nutrient input has no specific separation between organic and inorganic nutrients and all is considered bioavailable. A comment on this and a new reference have been added to the new manuscript version.

Causes of simulated long-term changes in phytoplankton biomass in the Baltic Proper: A wavelet analysis

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Abstract. The co-variation of key variables with simulated phytoplankton biomass in the Baltic proper has been examined using wavelet analysis and results of a long-term simulation for 1850-2008 with a high-resolution, coupled physical-biogeochemical circulation model for the Baltic Sea. By focusing on inter-annual variations it is possible to track effects acting on decadal time scales such as temperature increase due to climate change as well as changes in nutrient input. The strongest inter-annual coherence indicates that variations in phytoplankton biomass are determined by changes in concentrations of the limiting nutrient. However, after 1950 high nutrient concentrations created a less nutrient limited regime and the coherence was reduced. Furthermore, the inter-annual coherence of mixed layer nitrate with riverine input of nitrate is much larger than the coherence between mixed layer phosphate and phosphate loads. This indicates a greater relative importance of the vertical flux of phosphate from the deep layer into the mixed layer. In addition, shifts in nutrient patterns give rise to changes in phytoplankton nutrient limitation. The modelled pattern shifts from purely phosphate limited to a seasonally varying regime. The results further indicate some effect of inter-annual temperature increase on cyanobacteria and flagellates. Changes in mixed layer depth affect mainly diatoms due to a their high sinking velocity while inter-annual coherence between irradiance and phytoplankton biomass is not found.

1   Introduction

The Baltic Sea is a semi-enclosed brackish water body separated from the North Sea and Kattegat through the Danish Straits. It stretches from about 54° to 66° N and the limited water exchange with the ocean in the south gives rise to a large meridional salinity gradient. The circulation is estuarine with a salty deep-water inflow from the ocean and a fresher surface outflow. The Baltic Sea comprises a number of sub-basins connected by sills further restricting the circulation.

The limited water exchange and the long residence time of water have consequences for the biology and the biogeochemistry. The Baltic Sea is naturally prone to eutrophication and organic matter degradation leads to low deep water oxygen concentrations in between deep water renewal events. In turn, this leads to complex nutrient cycling with different processes acting in oxygenized vs low oxygen environments.

23   The Baltic Sea has experienced extensive anthropogenic pressure over the last century. After 1950, intensive use of agricul-
24   tural fertilizer greatly enhanced the nutrient loads. This led to an expansion of hypoxic bottoms (Carstensen et al., 2014), in turn
25   affecting the cycling of nutrients through the system. Anoxic sediments have lower phosphorus retention capacity resulting in
26   increased deep water phosphate concentrations. Thereby, the flux of phosphate to the surface intensified even though the exter-
27   nal loads decreased after 1980 in response to improved sewage treatment. Furthermore, as the anoxic area increased, the area of
28   interface between oxic and anoxic zones where denitrification occurs also increased. This resulted in a loss of nitrogen. Vahtera
29   et al. (2007) described these processes as generating a "vicious circle" where decreased concentrations of Dissolved Inorganic
30   Nitrogen (DIN) together with increased phosphate enhanced the relative importance of nitrogen fixation by cyanobacteria.

31   The importance of this coupling between oxygen and nutrients have been examined in models. Gustafsson et al. (2012)
32   confirmed, using the model BALTSEM, that internal nutrient recycling has increased due to the reduced phosphate retention
33   capacity, resulting in a self sustained eutrophication where enhanced sedimentary out-flux of nutrients together with increased
34   nitrogen fixation outweigh external load reductions.

35   Satellite monitoring has made it possible to observe changes in several physical and ecological surface variables during the
36   past three decades. Significant changes in seasonality have been observed, such as an earlier start of the phytoplankton growth
37   season and timing of chlorophyll maxima (Kahru et al., 2016).

38   Shifts in nutrient composition and deep water properties remain difficult to evaluate using observations. Even though the
39   Baltic Sea has a dense observational record from ships, stations and satellites, the longest nutrient records comprise station
40   data from the early 1970 (HELCOM, 2012). For longer time periods the use of a model is required.

41   In this paper we construct a thorough analysis of the co-variation of phytoplankton biomass with key variables that have
42   been affected by anthropogenic change over the 20th century. Using the biogeochemical model SCOBI (Eilola et al., 2009;
43   Almroth-Rosell et al., 2011) coupled to the 3d circulation model RCO (Meier et al., 2003) we scrutinize the effect of nutrient
44   loads, nutrient concentration, temperature, irradiance and mixed layer depth on the modelled phytoplankton community.

45   The gap-free dataset provided by the model allows us to decompose the variables in time-frequency space using the wavelet
46   transform. Two variables may than be compared using wavelet coherence (e.g., Torrence and Compo, 1998; Grinsted et al.,
47   2004).

48   We have chosen to use a model run spanning the period 1850 to 2009. Thereby, we capture conditions relatively unaf-
49   fected by anthropogenic forcing as well as current conditions of eutrophication and climate change. Furthermore, we limit our
50   investigation to the Baltic proper so as to capture relatively homogeneous conditions with regards to the biology.

51   **2   Methods**

52   **2.1   Model**

53   We have used a run from the model RCO-SCOBI spanning 1850-2009. RCO (Rossby Centre Ocean model) is a three-
54   dimensional regional ocean circulation model (Meier et al., 2003). It is a z-coordinate model with a free surface and an

55　open boundary in the northern Kattegat. The version used here has a horizontal resolution of 2nm with 83 depth levels at 3m

56　intervals.

57　　The biogeochemical interactions are solved by the Swedish Coastal and Ocean Biogeochemical model (SCOBI) (Eilola

58　et al., 2009; Almroth-Rosell et al., 2011). The model contains the nutrients phosphate, nitrate and ammonia as well as the

59　plankton functional types representing diatoms, flagellates and others (will be referred to as flagellates from here on) and

60　cyanobacteria. Furthermore, the model contains nitrogen and phosphorus in one active homogeneous benthic layer.

61　　The model equations can be found in Eilola et al. (2009). Since we are exploring the effect of different variables on the

62　growth of phytoplankton we will, for clarity, repeat some of them here.

63　　The phytoplankton biomass is described in terms of chlorophyll and with a constant C:Chl ratio. The model thus does not

64　take into account seasonal changes in C:Chl as was found by Jakobsen and Markager (2016).

65　　The net growth of phytoplankton (PHY) is described by the following expression,

66　$$\mathrm{GROWTH_{PHY}} \quad = \quad \mathrm{ANOX \cdot LTLIM \cdot NUTLIM_{PHY} \cdot GMAX_{PHY} \cdot C_{PHY}}. \tag{1}$$

67　Subscript PHY indicates the plankton functional type (diatoms, flagellates or cyanobacteria) and C_{PHY} is the plankton

68　biomass. ANOX is a logarithmic expression that approaches zero as the oxygen concentration becomes small.

69　　LTLIM expresses the phytoplankton light limitation and NUTLIM describes the nutrient limitation. Nutrient limitation

70　follows Michaelis-Menten kinetics where constant Redfield ratios are assumed in nutrient uptake. NUTLIM is further described

71　in Sects. 2.1.1 and 2.1.2. GMAX is temperature dependent and describes the maximum phytoplankton growth rate.

72　　Diatoms and flagellates have different half-saturation constants, maximum growth rate, temperature dependence and sinking

73　rate. Flagellates are more sensitive to changes in temperature than diatoms. Furthermore, the sinking rate of diatoms is five

74　times larger than that for flagellates.

75　　The difference between cyanobacteria and the other phytoplankton is more pronounced. Cyanobacteria can grow either

76　according to Eq. (1) or using nitrogen fixation. The rate of nitrogen fixation is a function of phosphate concentration, N:P

77　ratio and temperature. Both nitrogen fixation and GROWTH of cyanobacteria is zero if the salinity is above 10. Furthermore,

78　cyanobacteria is the most temperature sensitive of the phytoplankton groups and no sinking is assumed.

79　　Other processes important for our results involves chemical reactions occurring in the water column or in the sediment.

80　Denitrification occurs both in the water column and the benthic layer and constitutes a sink for nitrate in case of anoxia.

81　Nitrification transforms ammonium into nitrate as long as oxygen is present. Phosphorus is adsorbed to the sediment and

82　the benthic release capacity of phosphate is a function of the oxygen concentration. The phosphorus release capacity is also

83　dependent on salinity whereby higher salinity leads to lower retention of phosphate in the benthic layer.

84　**2.1.1　Nutrient limitation**

85　Estimating nutrient limitation in nature is difficult. Usually this is done, either by comparing nutrient ratios to Redfield in, e.g.,

86　the surface water or external supply or through nutrient enrichment experiments (Granéli et al., 1990).

87   The implementation of nutrient limitation most commonly used is that the primary production is directly limited by the
88   nutrient concentration in the ambient water and that the internal nutrient ratios in the phytoplankton are constant, i.e., in accor-
89   dance with a Redfield-Monod model (Redfield, 1958). However, cell-quota type models (Droop, 1973) are being increasingly
90   implemented and the use of constant internal nutrient ratios are becoming more and more questioned (Flynn, 2010; Fransner
91   et al., 2018).

92   In our model, nutrient limitation is expressed assuming constant Redfield ratios and phytoplankton growth is limited by
93   either nitrogen or phosphate. The degree of nutrient limitation is described by

$$94 \quad \mathrm{NUTLIM_{PHY}} \quad = \quad \min(\mathrm{NLIM_{PHY}}, \mathrm{PLIM_{PHY}}) \tag{2}$$

95   where $\mathrm{NLIM_{PHY}}$ and $\mathrm{PLIM_{PHY}}$ are the nitrogen and phosphate limitation respectively. $\mathrm{NLIM_{PHY}}$ is defined as

$$96 \quad \mathrm{NLIM_{PHY}} \quad = \begin{cases} \mathrm{NO_3LIM_{PHY}} + \mathrm{NH_4LIM_{PHY}}, & \text{if } \mathrm{NO_3LIM_{PHY}} + \mathrm{NH_4LIM_{PHY}} < 1 \\ 1, & \text{otherwise,} \end{cases} \tag{3}$$

97   where

$$98 \quad \mathrm{NO_3LIM_{PHY}} \quad = \quad \frac{\mathrm{NO3}}{\mathrm{KNO3_{PHY}} + \mathrm{NO3}} \cdot \exp(-\phi_{\mathrm{PHY}} \cdot \mathrm{NH_4}), \tag{4}$$

$$99 \quad \mathrm{NH_4LIM_{PHY}} \quad = \quad \frac{\mathrm{NH4}}{\mathrm{KNH4_{PHY}} + \mathrm{NH4}}, \tag{5}$$

100   where NO3 and NH4 are the concentrations of nitrate and ammonium and $\mathrm{KNO3_{PHY}}$ and $\mathrm{KNH4_{PHY}}$ are the half-saturation
101   constants for nitrate and ammonium uptake, respectively. The exponent in (4) accounts for inhibition of nitrate uptake in the
102   presence of ammonium (e.g., Dortch, 1990; Parker, 1993).

103    $\mathrm{PLIM_{PHY}}$ is modelled as

$$104 \quad \mathrm{PLIM_{PHY}} \quad = \quad \frac{\mathrm{PO4}}{\mathrm{KPO4_{PHY}} + \mathrm{PO4}}. \tag{6}$$

105   The constant $\mathrm{KPO4_{PHY}}$ is the half saturation constants for phosphate.

106   Nutrient limitation, NUTLIM, is thus described by a number between 0 and 1 where 1 is no limitation. Since NUTLIM is
107   calculated as the minimum of NLIM and PLIM, NLIM larger than PLIM will temporally cause P limitation of phytoplankton
108   growth rate. Hence, a different formulation e.g. of NLIM might change a models sensitivity to the limiting nutrient. Further
109   experiments on this issue are out of the scope of the present paper and left for future studies.

110   NUTLIM for our model run has been calculated off-line from the monthly means according to Eq. (2).

111   **2.1.2   Effect of physical parameters**

112   Changes in cloud-cover affect the incoming solar radiation and thereby phytoplankton growth. The effect of light is given by
113   the LTLIM term of Eq. (1) which accounts for photo-inhibition.

114   The mixed layer depth has been defined as the depth where a density difference of 0.125 kg m^{-3} from the surface occurs in
115   accordance with what was previously done by e.g., Eilola et al. (2013). The density was calculated from modelled temperature
116   and salinity using the algorithms from Jackett et al. (2006).

2.2   Study area

118   The Baltic Sea contains several different sub-basins with different characteristics in salinity and nutrient loads. In this study
119   we focus on the Baltic proper as defined in Fig. 1. In order to reduce heterogeneity we exclude areas shallower than 20m and
120   put our focus away from the coasts.
121   We have chosen to use a basin averaged approach in order to remove local variability and gain a better understanding of
122   the system. All variables have thus been horizontally averaged over the study area. Furthermore, we have also averaged all
123   variables over the mixed layer and from the mixed layer down to a depth of 150m.

2.3   Forcing

125   The study use reconstructed (1850-2008) atmospheric, hydrological and nutrient load forcing and daily sea levels at the lateral
126   boundary as described by Gustafsson et al. (2012) and Meier et al. (2012). Monthly mean river flows were merged from
127   reconstructions by Hansson et al. (2011) and Meier and Kauker (2003) and hydrological model data from Graham (1999),
128   respectively. For further details about the physical model set-up used in the present study the reader is referred to Meier et al.
129   (2017) and references therein.
130   The nutrient input from rivers and point sources between 1970 and 2006 were compiled from the Baltic Environmental
131   and HELCOM databases (Savchuk et al., 2012). Estimates of pre-industrial loads for 1900 were based on data from Savchuk
132   et al. (2008). The nutrient loads were linearly interpolated between selected reference years in the period between 1900 and
133   1970. Atmospheric loads were estimated in a similar manner in accordance with Ruoho-Airola et al. (2012). Riverine nutrient
134   loads contain both organic and inorganic phosphorus and nitrogen~~, respectively, while atmospheric loads contain only organic~~
135   ~~nitrogen~~. Bioavailable fractions of 100% ~~for phosphorus~~ and 30% ~~for nitrogen~~ respectively were assumed for ~~river loads while~~
136   ~~a 20% fraction were assumed for atmospheric N loads (Savchuk et al., 2012)~~ phosphorus and nitrogen (Savchuk et al., 2012) .
137   No similar separation between organic and inorganic atmospheric nutrient loads was made (Meier et al., 2018a) .
138   The upper panel of Fig. 2 shows the input of Dissolved Inorganic Phosphorus (DIP) and DIN to the Baltic proper as defined
139   in Fig. 1. The lower panel shows the corresponding simulated mixed layer concentrations. The loads have been calculated from
140   the runoff and annual mean nutrient concentrations (Eilola et al., 2011). Thus the seasonal cycle in river loads is determined
141   by the runoff. After a spin-up simulation for 1850-1902 utilizing the reconstructed forcing as described above, the calculated
142   physical and biogeochemical variables at the end of the spin-up simulation were used as initial condition for 1850. We have
143   used riverine DIN and DIP loads for our analysis. The use of total bioavailable nutrient loads instead does not change the
144   results.
145   The open boundary conditions in the northern Kattegat were based on climatological (1980-2000) seasonal mean nutrient
146   concentrations (Eilola et al., 2009). Similar to Gustafsson et al. (2012) a linear decrease of nutrient concentrations back in

147 time was added assuming that climatological concentrations in 1900 amounted to 85% of present day concentrations (Savchuk
148 et al., 2008). The bioavailable fraction of organic phosphorus at the boundary was assumed to be 100% in accordance with
149 the organic phosphorus supply from land runoff. Organic nitrogen was implicitly added through the Redfield ratio (nitrogen to
150 phosphorus) of detritus in the model (Eilola et al., 2009).

2.4 Evaluation

152 The specific model set-up used here have been shown to agree well with observations for salinity, temperature and nu-
153 trients ~~(Meier et al., 2018b; Eilola et al., 2014)~~ (Meier et al., 2018a, b; Eilola et al., 2014). The different phytoplankton func-
154 tional types have not been previously validated against observations.

155 The phytoplankton functional groups in the simulations and respective observations from BY15 (see Fig.1) are shown in Fig.
156 3 and 4. Phytoplankton biomass from field observations has been estimated through the conversion of biovolumes into carbon
157 in accordance with Menden-Deuer and Lessard (2000). Phytoplankton biomass for the model simulation was estimated from
158 chlorophyll (Chl) assuming a C:Chl ratio of 50. This ratio is in the middle of the salinity dependent range found by Rakko and
159 Seppälä (2014).

160 The time-series display significant inter-annual variability in both model and observations. This variability is also visible as
161 large standard deviations in the modelled and observed monthly means in Figs. 3 and 4.

162 Fig. 4 shows an autumn diatom bloom in the observations while the model generates an autumn flagellate bloom. In addition,
163 the model partly overestimates the diatom spring blooms. In 2006 and 2007, this is a result of too high simulated winter nutrient
164 concentrations at BY15. The relationship between modelled N and P also differ from reality which introduces errors in the
165 distribution of plankton functional types. This may, in part, explain the overestimation of diatoms and the underestimation of
166 flagellates during the first two years in Fig. 3.

167 Similar to comparable models, the simulated cyanobacteria bloom occurs approximately two months too late compared to
168 observations (Hense and Beckmann, 2010). It is also notable that the cyanobacteria displays strong blooms the first four years
169 in both model and observations but that the observations show diminished blooms during the rest of the period where the
170 simulated biomass is still high. There is currently ongoing work of including a cyanobacteria life cycle model and early work
171 shows some improvements. There is also an influence on the sampling frequency on this comparison. While we have model
172 data every other day, the measurements are only done approximately once a month. and will therefore almost certainly miss
173 peak concentration more often than the model values. Differences in the real Chl:C ratio from our fixed value of 50 will also
174 introduce significant errors.

175 The estimated carbon content from observations are potentially affected by patchiness during in-situ sampling and uncer-
176 tainties related to the calculation of biovolumes and transformation to carbon units.

2.5 The wavelet transform and wavelet coherence

178 The wavelet transform and its application have been described in several studies (e.g., Lau and Weng, 1995; Torrence and
179 Compo, 1998; Carey et al., 2016; Grinsted et al., 2004). Below we provide, therefore, only a brief overview of the method.

The continuous wavelet transform provides a method to decompose a signal into time-frequency space. In that it is similar to the windowed Fourier transform where the signal is decomposed within a fixed time-frequency window which is then slided along the time-series. However, the fixed width of the window leads to an underestimation of low frequencies. In comparison, the wavelet transform utilizes wavelets with a variable time-frequency window. Wavelets can have many different shapes and the choice is not arbitrary. We have chosen the commonly used Morlet wavelet providing good time and frequency localization (Grinsted et al., 2004).

In time-series with clear periodic patterns affected by environmental variables such as population dynamics and ecology the benefits of this approach are significant (Cazelles et al., 2008). In recent years, several studies have highlighted the usefulness of wavelet analyses in plankton research (Winder and Cloern, 2010; Carey et al., 2016). The focus has been the increased availability of long observational data sets making it possible to use the wavelet transform to investigate changes in seasonality. Carey et al. (2016) discussed how the wavelet transform can be used to track inter-annual changes in phytoplankton biomass and applied it to a 16-year time series of phytoplankton in Lake Mendota, USA. In doing so they were able to identify periods when the annual periodicity was less pronounced. They discussed the benefit of this technique in scrutinizing changes to the seasonal succession due to changes in external drivers. Winder and Cloern (2010) applied the technique to time-series of chlorophyll-a from marine and freshwater localities and discussed the annual and seasonal periodicities.

Wavelet coherence further expands the usefulness of the wavelet approach by allowing calculation of the time resolved coherence between two time-series (Grinsted et al., 2004; Cazelles et al., 2008). In this way, it is possible to identify transient periods of correlation over different periodicities. The result is given as coherency as a function of time and period as well as a phase lag between the two time-series.

The disadvantage of wavelet transform analysis is that it requires long datasets without gaps, while on the temporal scale of climate change such observations on plankton dynamics are lacking. Hence, for our purpose only a model based approach is feasible.

Schimanke and Meier (2016) used wavelet coherency on a multi-centennial model run to evaluate the correlation of different forcing variables with the Baltic Sea salinity. Here we analyse the coherence between modelled phytoplankton biomass and a few key modelled and forcing variables.

For all wavelet calculations we use the Matlab wavelet package described in Grinsted et al. (2004), which is freely available at http://www.glaciology.net/wavelet-coherence.

3 Results and discussion

We will begin in Sect. 3.1 by presenting the model results on phytoplankton biomass. In Section 3.2 we will present the nutrients and their coherence with the phytoplankton biomass. Coherence between riverine loads and mixed layer nutrients will be discussed in Sect. 3.3. Section 3.4 examines the coherence of phytoplankton with temperature and irradiance. Finally, the coherence between mixed layer depth and phytoplankton biomass is considered in Sect. 3.5. All results shown are monthly means.

3.1 Phytoplankton biomass

Fig. 5 shows the time-series of phytoplankton biomass (a) together with the corresponding wavelet spectrum (b).

The wavelet power (variance) of the decomposed signal (in colour) is displayed as a function of time (x-axis) and period (y-axis). The black curves in Fig. 5(b) show the 95% confidence level relative to red noise.

Averaging over time generates the global power spectrum displayed in Fig. 5 (c). The wavelet spectrum clearly reveals two main periodicities - the annual and the semi-annual representing the spring and autumn blooms. Further, the power of both periodicities increases markedly after 1950.

Kahru et al. (2016) found a shift in chlorophyll maxima from the diatom dominated spring bloom to the cyanobacteria summer bloom. A similar pattern emerges from our model run as can be seen in Fig. 6. The figure shows the month of maximum biomass of the different phytoplankton species as well as the month of maximum chlorophyll (diatoms+flagellates+cyanobacteria). After 1998 the results display five years where the month of maximum chlorophyll corresponds to the month of maximum cyanobacteria biomass in August or September.

3.2 Nutrients and nutrient limitation

Increased nutrient loads have caused an increase in primary production and thereby also the deep water respiration, resulting in a 10-fold increase in hypoxic area since the beginning of the 20th century (Carstensen et al., 2014).

This has led to a change in nutrient availability and dynamics as anoxia leads to a release in sedimentary phosphate (e.g., Conley et al., 2002; Savchuk, 2010, 2018; Vahtera et al., 2007). A clear relationship between hypoxia and total basin averaged phosphate was first shown by Conley et al. (2002) (and later expanded by Savchuk (2010)) on observational data from the Baltic proper.

The effect of hypoxia on DIN is less straight forward. Expanding hypoxia increases the boundary area between anoxic and oxic water where denitrification occurs resulting in a loss of nitrate. Furthermore, hypoxia causes a reduction in nitrification leading to a further reduction in nitrate. Vahtera et al. (2007) found a negative relationship between basin averaged DIN and hypoxic area in observations from the Baltic Sea.

The changing nutrient patterns for our model run are shown in Fig. 7. In conjunction with the increased anoxic volume we find a clear increase in ammonium and a decrease in nitrate. This is due to a decrease in nitrification and an increase in denitrification. The phosphate concentration increases from the mid 20th century through the rest of the model run as a combined effect of the accumulated terrestrial inputs and hypoxic sedimentary release.

The effect of nutrients on the primary production is in the model controlled by the term NUTLIM, or degree of nutrient limitation, in Eq. (1). NUTLIM can be viewed as a measure of the nutrient composition that linearly affects the phytoplankton growth in the model. We examine this term in as well as below the mixed layer as changes in the concentration of nutrients in the deep water will affect also nutrient concentrations in the mixed layer.

The evolutions of NUTLIM in the mixed layer and deep water for diatoms and flagellates are shown in Fig. 8. Mixed layer values of NUTLIM increase over the 20th century indicating less nutrient limiting conditions.

246   Nitrogen has been shown to most often limit the growth in the Baltic proper, while phosphate is limiting in the northern
247 basins (Granéli et al., 1990; Tamminen and Andersen, 2007). In pre-industrial conditions, N/P ratios indicate a lesser degree
248 of nitrogen limitation and a higher degree of phosphate limitation for the central Baltic Sea (Schernewski and Neumann, 2004;
249 Savchuk et al., 2008; Gustafsson et al., 2012).

250   The mixed layer limitation patterns as estimated from NUTLIM and N/P ratios are shown in Fig. 9. Until 1980, the N/P
251 ratios display a pattern of limitation shifting between nitrogen and phosphate where after persistent N limitation develops. This
252 weaker N limitation during the first part of the run is consistent with the studies of pre-industrial conditions (Schernewski and
253 Neumann, 2004; Savchuk et al., 2008; Gustafsson et al., 2012).

254   Using NUTLIM, the results instead show phosphate limitation for both diatoms and flagellates for the earlier part of the run.
255 After 1980, a different seasonal pattern appears with phosphate still limiting during winter while nitrogen becomes limiting
256 after the spring bloom. Even though the limitation pattern as calculated with NUTLIM differs from what was found using N/P
257 ratios, the overall pattern of increasing degree of N limitation is evident in NUTLIM as well.

258   The changing nutrient limitation patterns affect phytoplankton growth. We analyse the wavelet coherencies of phytoplankton
259 biomass with mixed layer phosphate and DIN (the sum of nitrate and ammonium) in Figs. 10 and 11.

260   As the strongest nutrient limited group, diatoms show persistent inter-annual coherence with phosphate during the first,
261 consistently phosphate limited part of the run (Fig. 10, see also Fig. 9).

262   Since nitrogen limitation as calculated with NUTLIM mostly occurs after 1980 and after the spring bloom (Fig. 9), and thus
263 only affects the much smaller diatom and flagellate autumn blooms, little coherence between phytoplankton and nitrogen can
264 be observed on inter-annual time-scales (Fig. 11).

265   To scrutinize the shift in deep water nutrient composition and the coherence with phytoplankton, we calculate the wavelet
266 coherence between below mixed layer NUTLIM and the diatom and flagellate biomass. The result is shown in Fig. 12. After
267 1980 the phase arrows within the annual coherence period change direction. For diatoms, the phase shifts from NUTLIM
268 preceding diatoms by three months to diatoms preceding NUTLIM by the same amount. Flagellates display a similar shift.

269   The month of maximum NUTLIM shown in Fig. 13, indicates the month when the nutrient composition is most beneficial
270 for phytoplankton growth. The figure shows a clear shift occurring after 1980. Below the mixed layer, NUTLIM changes its
271 maxima from December and January to July, August and September for both diatoms and flagellates while a slight shift from
272 February to March occurs in mixed layer NUTLIM for diatoms. Mixed layer NUTLIM for flagellates displays no clear shift.
273 The shift in NUTLIM is a result of the increase in phosphate and ammonium occurring in conjunction with the increase in
274 anoxic volume shown in Fig. 7. The change in timing is probably due to reduced sedimentary phosphate retention and reduced
275 nitrification after the spring bloom.

3.3   Nutrient loads

277 We here analyse how changes in nutrient loads affect changes in the mixed layer nutrient concentrations.

278   The wavelet coherence between mixed layer nutrients and riverine input is shown in Fig. 14. The phosphate concentration
279 shows little coherence on periodicities longer than one year but DIN displays strong inter-annual coherence. The phase-arrows

280   indicate a phase-lag of about minus 45° on all inter-annual periodicities. For an 8 year period this means that a change in

281   riverine input precedes changes in mixed layer DIN by about 1 yr.

282   To further investigate the lack of inter-annual coherence between riverine phosphate loads and the mixed layer phosphate

283   concentration, the wavelet coherence between mixed layer salinity and nutrients are examined and displayed in Fig. 15. Mixed

284   layer salinity is affected by freshwater input from land, water exchange with adjacent basins, precipitation, evaporation and

285   mixing with deeper layers. For periodicities spanning 1 to 16 yrs, the coherence spectrum reveals higher coherence between

286   mixed layer salinity and phosphate (top) than between salinity and DIN (bottom).

287   The coherence that does exist between salinity and DIN on periodicities longer than one year is anti-phase i.e. low salinity

288   here coheres with high DIN concentrations. This indicates that high runoff is connected to high nitrogen loads and high DIN

289   concentrations in the mixed layer. It is also possible that low salinity in the mixed layer indicate periods with deep mixing and

290   better oxygen conditions in and below the halocline (Stigebrandt and Gustafsson, 2007). This could reduce the denitrification

291   during these periods and thus result in higher mixed layer DIN concentrations.

292   In contrast, the stronger inter-annual in-phase coherence between salinity and phosphate suggests that the reason for the

293   coherence might be a greater importance of phosphorus release from the sediments that eventually reaches the mixed layer

294   through mixing with deeper layers (cf. Eilola et al., 2014).

295   Riverine nutrient loads show little inter-annual coherence with phytoplankton biomass (not shown) other than on a 16 yr

296   period which probably reflects the overall pattern of simultaneous increase in riverine loads and phytoplankton biomass over

297   the second half of the 20th century.

298   **3.4   Temperature and irradiance**

299   The mixed layer temperature in the Baltic proper has increased over the 20th century. To analyse the effect of temperature on

300   the phytoplankton biomass, the wavelet coherence between temperature and phytoplankton have been plotted in Fig. 16. The

301   results suggest that the temperature increase after 1990 might have had an effect on cyanobacteria and flagellates. It is also

302   noticeable that the temperature increase observed between 1900 and 1940 probably had an effect on cyanobacteria. This is also

303   in agreement with the model formulation where cyanobacteria are the most sensitive to temperature followed by flagellates.

304   Light impacts primary production through the term LTLIM in Eq. (1). However, irradiance display very little variation on

305   any other periodicity than the annual as can be observed in a wavelet power spectrum (not shown). Therefore there exists

306   almost no coherence between phytoplankton and irradiance apart from the seasonal signal.

307   **3.5   Mixed layer depth**

308   We calculate the coherence between mixed layer depth and diatoms, flagellates and cyanobacteria in Fig. 17.

309   Apart from the annual cycle there is a strong coherence between mixed layer depth and diatoms, and to some extent flag-

310   ellates, on shorter periodicities as well. That is, the diatom biomass residing in the mixed layer seems to covary quite well

311   on periodicities equal to or shorter than one year. The model value for diatom sinking rate is five times higher than that for

312   flagellates while cyanobacteria is assumed to have no sinking rate. In a shallow mixed layer the diatom biomass decreases

313 faster than in a deep mixed layer because of the large sinking rate. Furthermore, in a deeper mixed layer stronger turbulence

314 counteract the sinking. In the wavelet coherence spectrum we thus see in-phase short term coherence.

4 Conclusions

316 With a focus on simulated inter-annual variations, the wavelet coherence of the mixed layer phytoplankton biomass with key

317 variables affecting the primary production has been examined for the Baltic proper.

318 The simulated chlorophyll concentration maximum shifted from spring to late summer at the end of the 20th century in

319 agreement with Kahru et al. (2016).

320 The phytoplankton group most strongly limited by nutrients in the model is diatoms. The connection between phytoplankton

321 biomass and nutrients is reflected in the strong inter-annual coherence between diatoms and phosphate as well as NUTLIM

322 before 1940. After 1940, NUTLIM and the biomass of individual phytoplankton groups increased to such an extent that inter-

323 annual variations are small compared to the seasonal signal. Similarly, flagellates, which are less limited by nutrients than

324 diatoms, show much smaller inter-annual coherence with phosphate even before 1940. NUTLIM for this group is high enough

325 that small long-term variations do not reflect strongly in the results.

326 Very little inter-annual coherence is observed also between phytoplankton and DIN. Using the models definition of nutrient

327 limitation, the spring bloom is phosphate limited throughout the run except for a few years after 1990 where diatoms are limited

328 by nitrogen. Calculating instead limitation as given by mixed layer N/P ratios generates a pattern in line with previous estimates

329 (Schernewski and Neumann, 2004; Savchuk et al., 2008; Gustafsson et al., 2012). The more prevalent phosphate limitation in

330 the model is thus not a manifestation of incorrect N/P ratios. Rather, it reflects a difference between the NUTLIM concept and

331 N/P ratios. NUTLIM is basically an efficiency, mapping a 3d space made up of PO_4^{3-}, NO_3^- and NH_4^+ concentrations onto a

332 value between 0 and 1. Limitations from N/P ratios meanwhile, are a 2d mapping from PO_4^{3-} and DIN to a boolean variable.

333 We found strong coherence between riverine input of DIN and mixed layer DIN but not a similar relationship between

334 riverine phosphate input and the corresponding mixed layer concentration. As mixed layer salinity displayed in-phase inter-

335 annual coherence with phosphate and only weak anti-phase coherence with DIN we hypothesise that this is due to a greater

336 importance of the flux of phosphate from lower layers.

337 The mixed layer temperature in the Baltic proper has increased during the 20th century. We found some response of this

338 mainly from the most temperature sensitive phytoplankton group cyanobacteria during periods of large inter-annual tempera-

339 ture increases. Flagellates, being more temperature sensitive than diatoms, seems to display a coherence with the temperature

340 increase occurring after 1980.

341 Variations in mixed layer depth affects mainly diatoms as these have a high sinking velocity. In-phase coherence between

342 diatoms and mixed layer depth on periodicities shorter than one year indicates that large seasonal changes in the mixed layer

343 depth significantly affects the mixed layer diatom biomass, while smaller inter-annual variations are of little consequence.

344 Finally, inter-annual variations in irradiance have little effect on phytoplankton biomass accumulation.

5 Data availability

The model data on which the results in the present study are based on are stored and available from the Swedish Meteorological and Hydrological Institute. Please send your request to ocean.data@smhi.se.

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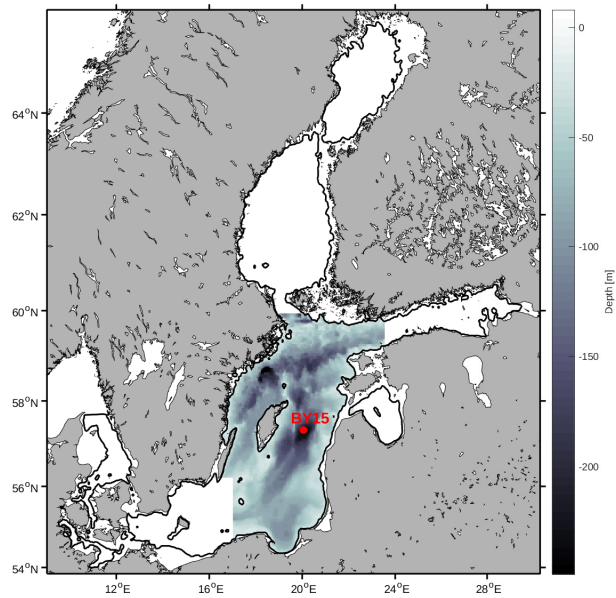
Figure 1. Study area. The grey scale represents depth in m. The red dot represents the monitoring station BY15
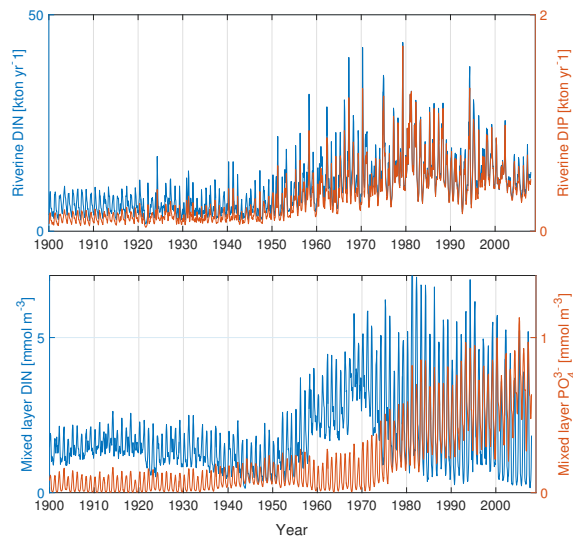

Figure 2. The top panel shows riverine DIN (blue) and DIP (red) loads to the Baltic proper as defined in Fig. 1. The bottom panel shows mixed layer DIN (blue) and mixed layer phosphate (red) averaged over the study area.
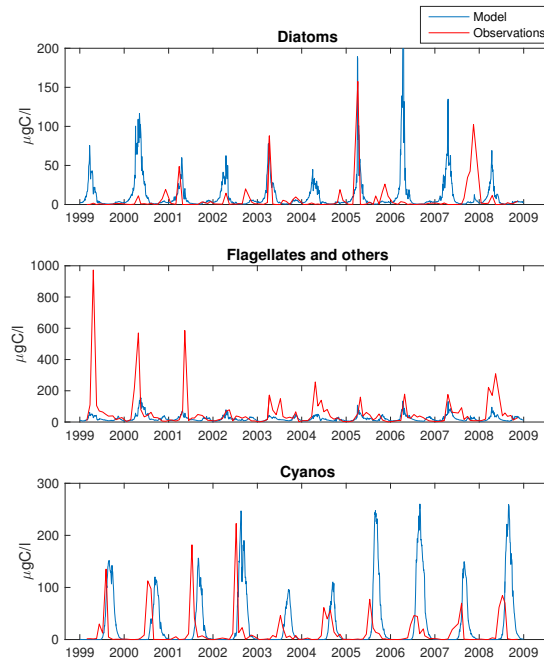
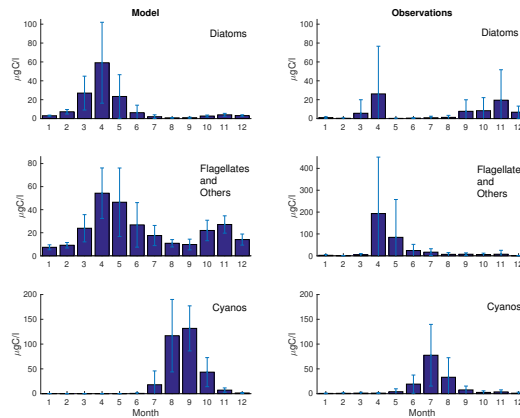**Figure 3.** Simulated (blue) and observed (red) biomass of diatoms (top), flagellates and others (middle) and cyanobacteria (bottom) at BY15.

Figure 4. Monthly means of simulated (left) and observed (right) diatoms (top), flagellates and others (middle) and cyanobacteria (bottom) at BY15. Standard deviations are shown as error bars.
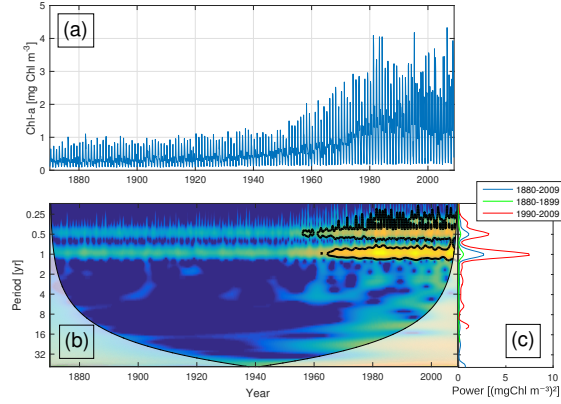
Figure 5. Time-series of phytoplankton biomass (a) together with the corresponding wavelet power spectrum (b) and global wavelet spectrum (c). More yellow means more power. The black curves in (b) represent the 95% confidence level relative to red noise. The white areas in (b) represent the cone of influence in which the results are impacted by edge-effects and are therefore not shown. The different lines in (c) represent the global spectrum 1880-2009 (blue), 1880-1899 (green), 1990-2009 (red).
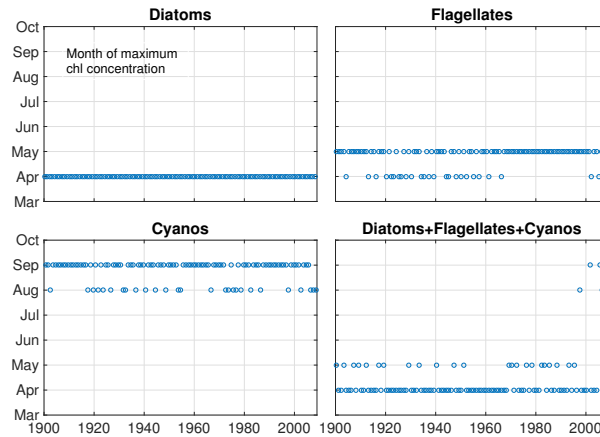
Figure 6. The month of maximum biomass of diatoms, flagellates and cyanobacteria as well as their sum.
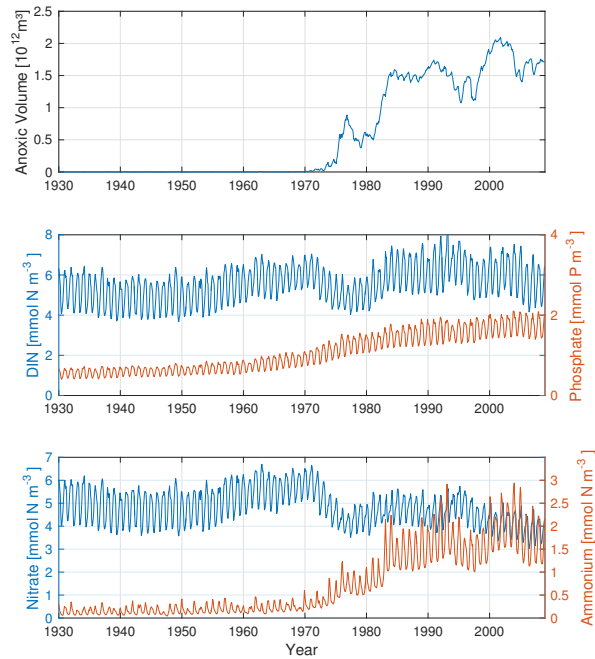
Figure 7. Time-series of anoxic volume (top panel), below mixed layer concentrations of DIN (nitrate + ammonium, blue) and phosphate (red) (middle panel) as well as nitrate (blue) and ammonium (red)(bottom panel). Deep water concentrations where averaged below the mixed layer depth for the Baltic proper.
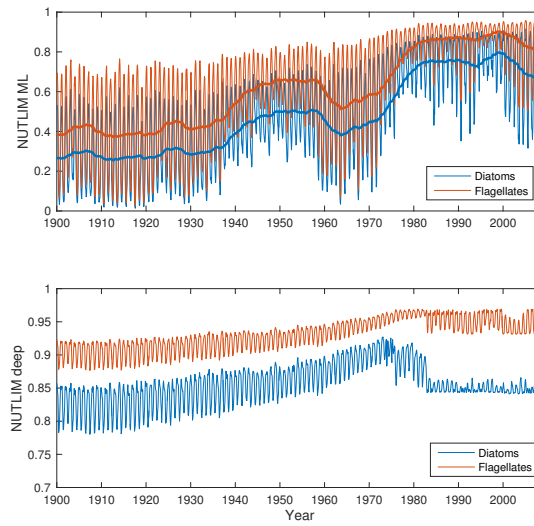
Figure 8. Time-series of nutrient limitation in the mixed layer (top) and below (bottom) for diatoms (blue) and flagellates (red). The thicker lines in the top panel show the 5-year moving average.
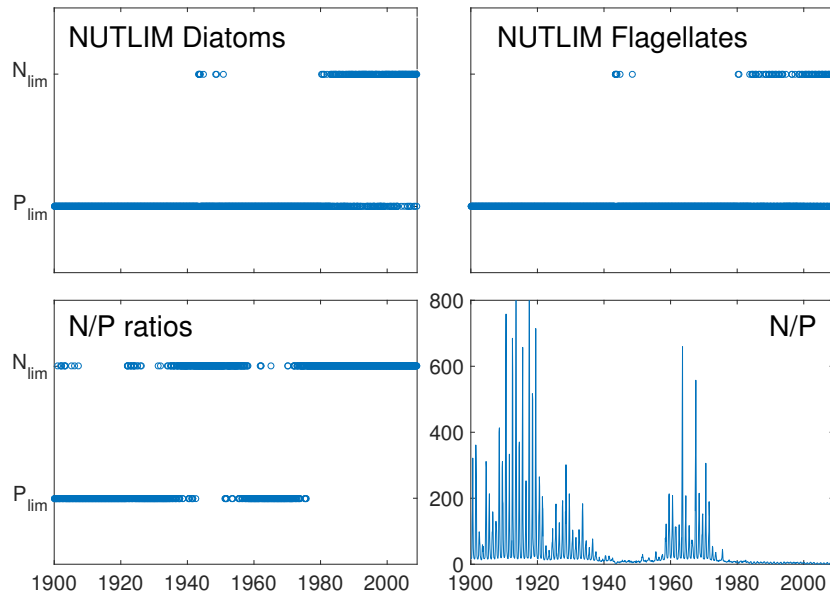
Figure 9. Mixed layer nitrogen or phosphate limitation as function of time for diatoms (upper left) and flagellates (upper right) as calculated through Eq. (2) where N limitation occurs when NLIM<PLIM. The bottom panel shows nutrient limitation as calculated through N/P ratios, where N limitation occurs when N/P<16 (left) and actual DIN/Phosphate (right). Note that simultaneous N and P limitation is not possible although the size of the rings gives this appearance.
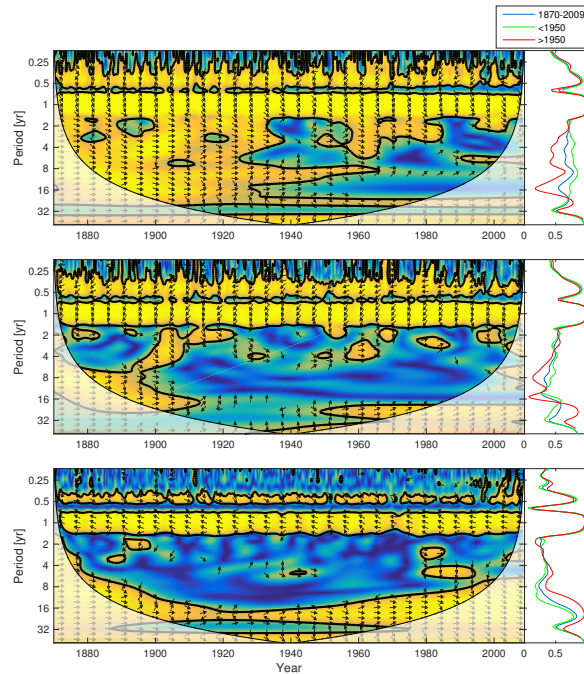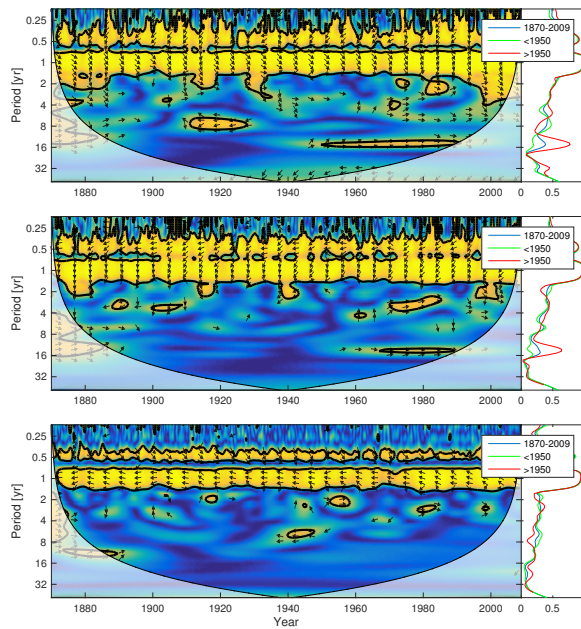
Figure 10. Wavelet coherence between mixed layer phosphate concentration and diatoms (top), flagellates (middle) and cyanobacteria (bottom). More yellow means more coherence. The arrows indicate the phase lag. When pointing to the right the two time-series are in phase and when pointing in the opposite direction anti-phase. Arrows pointing downwards indicate phosphate preceding plankton group by 90 degrees and upwards mean plankton preceding phosphate by the same amount. The right panels show the coherence averaged over the whole period (blue) and before (green) and after (red) 1950.

Figure 11. Wavelet coherence between mixed layer DIN concentration and diatoms (top), flagellates (middle) and cyanobacteria (bottom). Arrows pointing downwards indicate DIN preceding plankton group by 90 degrees and upwards mean plankton preceding DIN by the same amount.
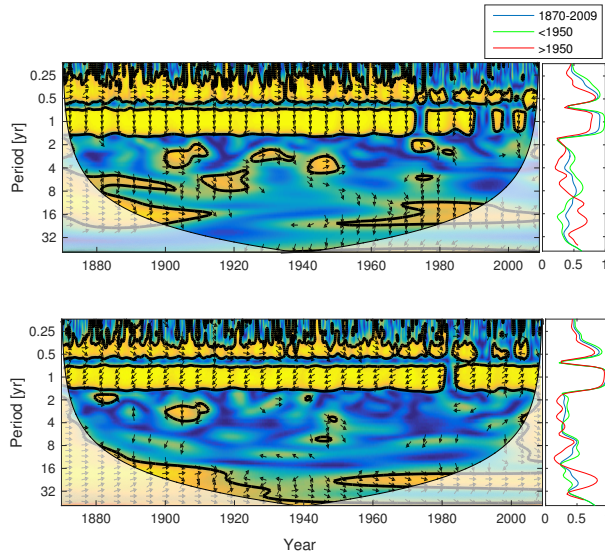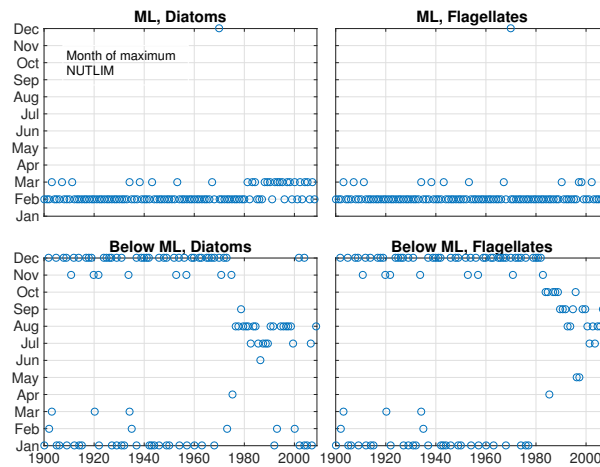
Figure 12. Wavelet coherence between deep water NUTLIM and diatoms (top), flagellates (middle). More yellow means more coherence. Arrows pointing downwards indicate NUTLIM preceding plankton group by 90 degrees and upwards mean plankton preceding NUTLIM by the same amount.

Figure 13. The month of maximum NUTLIM for diatoms (left) and flagellates (right) in the mixed layer (top) and below (bottom).
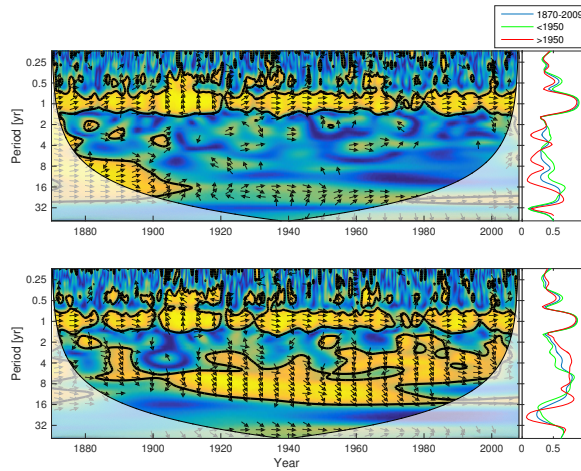
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Figure 14. Wavelet coherence between riverine phosphate and mixed layer phosphate concentration (top) and riverine DIN and mixed layer DIN concentration (bottom). Arrows pointing downwards indicate riverine phosphate/DIN preceding mixed layer phosphate/DIN by 90 degrees and upwards mean mixed layer phosphate/DIN preceding riverine phosphate/DIN by the same amount. The right panels show the averaged coherence for the whole period (blue) and before (green) and after (red) 1950.
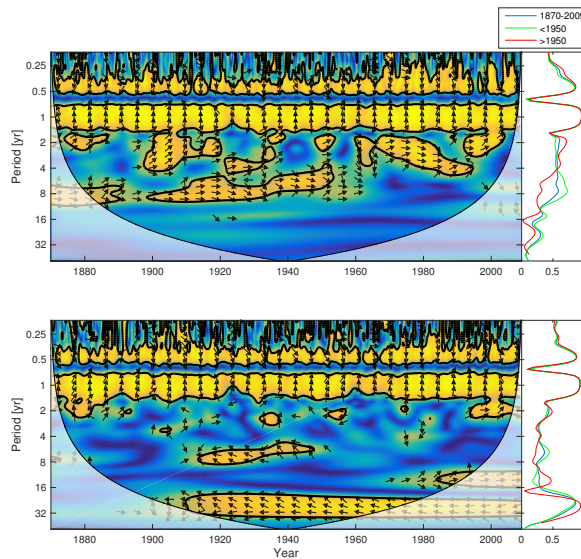

Figure 15. Wavelet coherence between mixed layer salinity and phosphate concentration (top) and mixed layer salinity and DIN (bottom). Arrows pointing downwards indicate salinity preceding mixed layer phosphate/DIN by 90 degrees and upwards mean mixed layer phosphate/DIN preceding salinity by the same amount. The right panels show the averaged coherence spectrum.
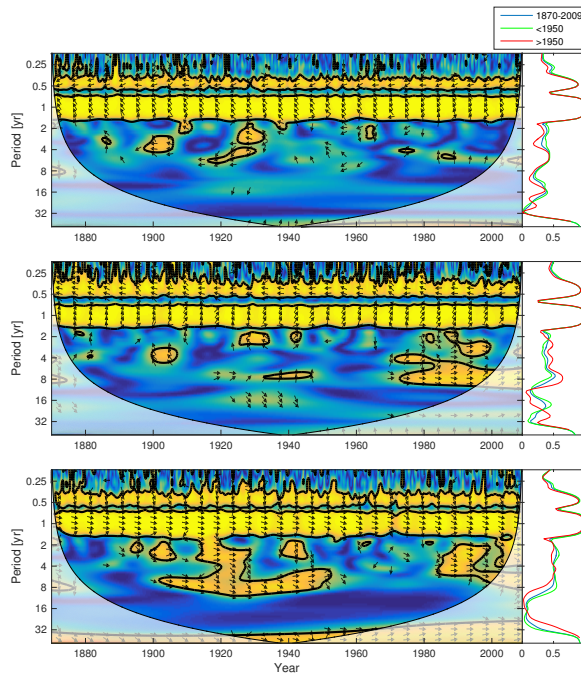
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Figure 16. Wavelet coherence between mixed layer temperature and diatoms (top), flagellates (middle) and cyanobacteria (bottom). Arrows pointing downwards indicate temperature preceding plankton group by 90 degrees and upwards mean plankton preceding temperature by the same amount.
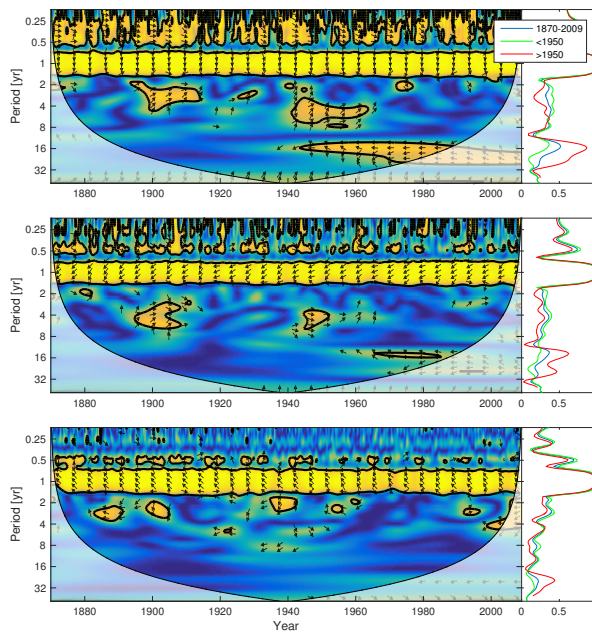
Figure 17. Wavelet coherence between mixed layer depth and diatoms (top), flagellates (middle) and cyanobacteria (bottom). Arrows pointing downwards indicate mixed layer depth preceding plankton group by 90 degrees and upwards mean plankton preceding mixed layer depth by the same amount.