Fungal loop transfer of N depends on biocrust constituents and N form

2

Zachary T. Aanderud¹, Trevor B. Smart¹, Nan Wu², Alexander S. Taylor¹,

4 Yuanming Zhang², Jayne Belnap³

- ¹Department of Plant and Wildlife Sciences, Brigham Young University, Provo, Utah
 84602, USA
- 8 ²Xinjiang Institute of Ecology and Geography, Key Laboratory of Biogeography and Bioresource in Arid Land, Chinese Academy of Sciences, Urumqi 830011, China
- ³US Geological Survey, Southwest Biological Science Center, 2290 S. Resource
 Blvd., Moab, UT 84532, USA
- 12

Running head: dark septate fungi translocate ammonium in biocrusts

14

Key words: ammonium, Ascomycota, Colorado Plateau, dark septate endophyte,

16 fungal loop, Indian ricegrass, Pleosporales

Correspondence to: Zachary T. Aanderud (zachary_aanderud@byu.edu)

- 20 Abstract. Besides performing multiple ecosystem services individually and collectively, biocrust constituents may also create biological networks connecting
- 22 spatially and temporally distinct processes. In the fungal loop hypothesis, rainfall variability allows fungi to act as conduits and reservoirs, translocating resources
- between soils and host plants. To evaluate the extent biocrust species composition andN form influence loops, we created a minor, localized rainfall event containing
- ¹⁵NH₄⁺ and ¹⁵NO₃⁻ and measured the resulting δ^{15} N in the surrounding dry cyanobacteria- and lichen-dominated crusts and grass, *Achnatherum hymenoides*,
- 28 after twenty-four hours. We also estimated the biomass of fungal constituents using quantitative PCR and characterized fungal communities by sequencing the 18S rRNA
- 30 gene. We found evidence for the initiation of fungal loops in cyanobacteriadominated crusts where ¹⁵N, from ¹⁵NH₄⁺, moved 40 mm h⁻¹ in biocrust soils with the
- 32 δ^{15} N of crusts decreasing as the radial distance from the water addition increased (linear mixed effects model (LMEM): R² = 0.67, F_{2,12} = 11, *P* = 0.002). In
- 34 cyanobacteria crusts, δ^{15} N, from ¹⁵NH₄⁺, was diluted as Ascomycota biomass increased (LMEM: R² = 0.63, F_{2,8} = 6.8, *P* = 0.02), Ascomycota accounted for 82%
- 36 (± 2.8) of all fungal sequences, and one order, Pleosporales, comprised 66% (± 6.9) of Ascomycota. The seeming lack of loops in moss-dominated crusts may stem from the
- 38 relatively large moss biomass effectively absorbing and holding N from our minor wet deposition event. The substantial movement of ¹⁵NH₄⁺ may indicate a fungal
- 40 preference for the reduced N form during amino acid transformation and translocation. We found a marginally significant enrichment of δ^{15} N in *A. hymenoides*

- 42 leaves but only in cyanobacteria biocrusts translocating ¹⁵N, offering evidence of links between biocrust constituents and higher plants. Our results suggest that minor
- 44 rainfall events may initiate fungal loops potentially allowing constituents, like dark septate Pleosporales, to rapidly translocate N from NH₄⁺ over NO₃⁻ through biocrust
- 46 networks.

48 **1** Introduction

Fungi may act as conduits for biological networks connecting belowground

- 50 ecosystem processes to plants. Soil fungi contribute to all aspects of litter decomposition through the generation of a myriad of extracellular enzymes (Osono
- 52 2007, Schneider et al. 2012); altered trophic dynamics, decomposer species diversity, and nutrient turnover rates (Hattenschwiler et al. 2005); and by forming multiple
- 54 types of endophyte-plant symbioses (Johnson et al. 1997, Saikkonen et al. 2004).Endophytic fungi, in particular, form hubs connecting spatially and temporally
- 56 distinct microbial-mediated soil processes and plants. For example, the pervasive distribution of mycorrhizae in mesic systems allows common mycorrhizal networks
- 58 to deliver essential resources, which promotes or hinders seedling growth depending on the network species composition (van der Heijden and Horton 2009), and
- 60 facilitates the one-way transfer of multiple forms of N and P between two plant species linked by arbuscular mycorrhizae and ectomycorrhizae (He et al. 2003,
- 62 Walder et al. 2012). In xeric systems, endophytic fungi are also implicated in moving resources within biological networks in a theory known as the fungal loop hypothesis.
- 64 The hypothesis states that fungi, supported by biotrophic C from plants and terrestrial cyanobacteria, act as intermediate reservoirs transforming and translocating resources
- between soils and plants (Collins et al. 2008, 2014). Perhaps the most notableexample of a fungal loop, albeit from a limited number of studies, occurred in fungal-
- 15 dominated cyanobacteria biocrusts from a Chihuahuan Desert grassland. Specifically, 15 NO₃⁻ applied to a root-free biocrust rapidly moved into the perennial grass,

- 70 *Bouteloua* sp., up to 1 m away within 24–hours (Green et al. 2008). Furthermore, ¹³C-labeled glutamic acid applied to leaf surfaces of *Bouteloua* was found in biocrusts.
- 72 Despite the intriguing evidence, many aspects of this burgeoning hypothesis remain to be validated (Collins et al. 2014).

Biocrust composition and soil moisture availability interactions may dictate the

- 76 movement of resources in fungal loops. Desert fungal-plant interactions occur across spatially discontinuous patches of vegetation interspersed by patches of soils
- 78 colonized by biocrusts (Belnap et al. 2005). Fungi participating in loops are necessarily associated with a mosaic of other biocrust organisms (i.e., cyanobacteria,
- 80 green algae, lichens, mosses, and other bacteria). The metabolic activity of biocrust constituents participating in fungal loops, including plants, are moisture-dependent
- 82 and regulated by the magnitude and seasonality of episodic rainfall events. A pulsereserve paradigm (Collins et al. 2008) may explain biological activities where minor
- 84 rainfall pulses stimulate microorganisms, generating reserves of resources to be exploited during subsequent rainfall events (Huxman et al. 2004, Welter et al. 2005).
- 86 In such loops, minor rainfall events may stimulate N₂ fixation by free or lichenassociated cyanobacteria (Belnap et al. 2003), N mineralization by bacteria and fungi
- (Cable and Huxman 2004, Yahdjian and Sala 2010), and nitrification and possibly
 denitrification (Wang et al. 2014), potentially altering levels of NH₄⁺ or NO₃⁻. Fungal
- 90 species, including fungal endophytes, may compete with mosses, lichen, cyanobacteria, and other bacteria for newly released N. Once sequestered, the N may

⁷⁴

- be transformed into amino acids and transported within mycelium (Jin et al. 2012,Behie and Bidochka 2014). Larger rainfall events may then activate plants, allowing
- 94 the host plant to receive N from the fungi and transfer photosynthate to the fungal endophyte. If fungal endophytes are poor competitors for newly released N,
- 96 preferentially sequester one inorganic N form over another, or more efficiently transform and transport NH_4^+ rather than NO_3^- , biocrust constituents and N form may

98 influence the translocation of N in fungal loops.

- 100 The fungal endophytes most likely involved in the loop hypothesis are dark septate fungi. Few arbuscular mycorrhizal fungi are found in biocrusts (Porras-Alfaro et al.
- 102 2011) or as endophytes in desert plants (Titus et al. 2002), due to mycorrhizae being relatively sensitive to dry soil conditions (Aguilera et al. 2016). In contrast, the
- 104 majority of biocrust fungi are Ascomycota, with the Pleosporales being widespread and dominant (Bates et al. 2012, Porras-Alfaro et al. 2011). Pleosporales, along with
- 106 other Ascomycota fungal orders, contain dark septate endophytes (Jumpponen and Trappe 1998). Dark septate endophytes are thermal- and drought-tolerant fungi due to
- 108 melanin-rich cell walls conferring protection from UV and drought stress (Gostincar et al. 2010). Taken together, the prevalence of dark septate fungi in desert systems,
- along with their ability to maintain metabolic activity under low water potentials(Barrow 2003), makes these endophytes excellent candidates to translocate resources
- 112 in loops (Green et al. 2008).

- 114 Minor rainfall events may allow fungi to act as conduits and reservoirs for N. To investigate the potential for biocrust constituents and N form to influence the
- 116 movement of N through the putative fungal loops, we created minor, localized rainfall events and measured δ^{15} N, from ¹⁵N-NH₄⁺ and ¹⁵N-NO₃⁻, within the surrounding dry
- 118 cyanobacteria- and moss-dominated crusts, and grass, *Achnatherum hymenoides* (Indian ricegrass). In tandem with ¹⁵N analyses, we estimated the biomass of two
- 120 major division of fungi (Ascomycota and Basidiomycota) and bacteria, and characterized fungal communities by sequencing the 18S rRNA gene to identify
- 122 potential links between fungal taxa and ¹⁵N movement.

124 **2** Materials and methods

2.1 Site description

- We conducted our study in two cold desert ecosystems of the Colorado Plateau, UT.One site was near Castle Valley (40°05'27.43"N-112°18'18.24"W) and the other was
- adjacent to the US Geological Survey (USGS) Southwest Biological Science CenterResearch Station in Moab, UT (40°05'27.43"N-112°18'18.24"W). Rugose crusts
- 130 consisting of 22% moss *Syntrichia caninervis*, 5-7% cyano-lichens *Collema tenax* and *Collema coccophorum* and 55% cyanobacteria cover the Castle Valley site
- 132 (Darrouzet-Nardi et al. 2015), while smooth, light algal crusts dominated by one cyanobacterium, *Microcoleus vaginatus*, cover the USGS site. Across both sites,
- 134 vegetation is dominated by perennial grass *Achnatherum hymenoides* (Roem & Schult) and the native perennial shrub *Atriplex confertifolia* (Torr. & Frém). Mean

- 136 annual temperature and precipitation at Castle Valley is 13°C and 269 mm, while the USGS site is slightly warmer (MAT=13.8°C) and drier (MAP=189 mm; based on
- 138 1981-2010 data; WRCC 2017). Both soils are Aridisols with Castle Valley classified as a sandy loam, calcareous Rizno series (Darrouzet-Nardi et al. 2015) and USGS as a

140 Bluechief series sandy loam.

142 **2.2** Simulated rainfall events and ¹⁵N form applications

We simulated minor, localized rainfall events and measured $\delta^{15}N$, from ^{15}N -NH₄⁺ and

- ¹⁵N-NO₃⁻ rainfall events containing two isotopically-labeled, inorganic N forms and tracked the movement of the label through our moss-dominated (Castle Valley) and
- 146 cyanobacteria-dominated biocrusts (USGS Station), and *A. hymenoides*. First, we randomly selected six circular plots per site with a radius of 1.0 m and at least 10 m
- apart from each other. Three plots were assigned to be labeled with $K^{15}NO_3$ (99 at.%) and the other three plots to be labeled with ($^{15}NH_4$)₂SO₄ (99 at.%). Second, we
- randomly selected five biocrust patches and five *A. hymenoides* along eight axes (e.g., N, NE, E, SE, S, SW, W, and NW) radiating from the center of each circular plot and
- 152 measured the radial distance to biocrusts or grasses. Third, we simulated a 2.5 mm rainfall event by spraying 3 mL of deionized water solution and either isotopic label
- 154 (0.30 mg ¹⁵N) onto a 5 cm diameter circle in the center of the circular plots (2 biocrust types \times 3 circular plots locations \times 2 N forms $\times \approx$ 10 samples [4 5 biocrusts or 4 7
- 156 *A. hymenoides* depending on grass density in the circular plot] = 137). For the isotopic applications, either 2.60 g of 99 at.% $K^{15}NO_3$ or 1.70 g of 99 at.% $(^{15}NH_4)_2SO_4$ was

- 158 dissolved in 18 mL of deionized water to create a 1.43 M or 0.72 M solution, respectively. The ¹⁵N additions wetted the sandy loams (bulk density ≈ 1.5 g cm⁻³) to
- 160 a depth of 1 cm and added approximately equal NH_4^+ (10 µg N g soil⁻¹) and double NO_3^- concentrations to surface soils (Sperry et al. 2006). All additions were
- 162 completed midday in April, as *A. hymenoides* were starting to set seed.

164 **2.3 Sample collection and ¹⁵N analyses**

Biocrust and foliage samples were collected twenty-four hours after the simulated

- 166 rainfall event containing our different inorganic ¹⁵N forms. Biocrusts were removed as three subsamples from each biocrust location with a soil corer (2 cm diameter \times 5
- 168 cm length) to a depth of 2 mm. Crust distances away from the tracer application ranged from 22 - 97 cm. The composited soil samples were kept cold (5°C) in the
- 170 field, split in the lab, and a portion of the soil was frozen (-20°C) until we performed fungal and bacterial DNA analyses. We randomly selected five leaves from
- 172 *Achnatherum*, which ranged in distance anywhere from 29 120 cm away from the tracer application and in volume from 0.002 0.048 m³. The leaves and remaining
- 174 soils (sieved 2 mm) were air-dried, ground in a reciprocating tissue homogenizer, and analyzed for ¹⁵N using a PDZ Europa ANCA-GSL elemental analyzer, interfaced
- 176 with a PDZ Europa 20–20 isotope ratio mass spectrometer (Sercon Ltd., Cheshire, UK) at the University of California Davis Stable Isotope Facility
- 178 (http://stableisotopefacility.ucdavis.edu). We expressed the resulting isotope ratios in δ notation as parts per thousand (‰) where:

180
$$\delta^{15} \mathrm{N} = (R_{\mathrm{sample}} / R_{\mathrm{standard}}) \times 1000$$

R is the molar ratio of the heavier to the lighter isotope $({}^{15}N/{}^{14}N)$ for the standard or

- 182 sample. To track the movement of inorganic N forms through our two biocrust types (moss-lichen-dominated and cyanobacteria dominated biocrust) and into grasses, we
- analyzed the relationships between δ^{15} N present in crust and leaf tissue to the distance of the crust and *Achnatherum* by site by creating linear mixed effects models for the
- 186 biocrust type and N form combinations. In all models, distances of biocrusts or grasses was nested within individual circular plots, and the three plots were treated as
- a random factor to control for the impact of plot differences on our response variable.All models were created with the *lm* function in R (R Development Core Team 2017).

190

2.4 Biomass estimations of major fungal components

- 192 To investigate the potential for fungi to translocate our ¹⁵N forms, we estimated the biomass of two major divisions of fungi (Ascomycota and Basidiomycota) and
- 194 bacteria in biocrusts using quantitative PCR. From the frozen biocrust samples, we extracted genomic DNA using a DNeasy PowerLyzer PowerSoil Kit (Qiagen, MD,
- 196 USA) and quantified the gene copy numbers of Ascomycota and Basidiomycota on a Mastercycler EP Realplex qPCR (Eppendorf, Hamburg, Germany) with SYBR Green.
- 198 We amplified division-specific regions of the internal transcribed spacer (ITS) with primer pair ITS5 (forward) and ITS4A (reverse) for Ascomycota (Larena et al. 1999)
- 200 and ITS4B (forward) and 5.8sr (reverse) for Basidiomycota (Fierer et al. 2005). We selected the universal bacterial 16S rRNA primer set EUB338, forward, and Eub518,

(1)

- 202 reverse, to estimate the biomass of bacteria (Aanderud et al. 2013). In 12.5 μl reactions, using KAPA2G Robust PCR Kits (KAPA Biosystems, Wilmington, MA,
- 204 USA), we amplified targeted genes using the following thermocycler condition: an initial denaturation step at 94°C for 3 min followed by 35 cycles of denaturation at
- 206 94°C for 45 s, annealing at either at 55 °C (Ascomycota), 64°C (Basidiomycota), or 60°C (bacteria) for 30 s, and extension at 72°C for 90 sec. We generated qPCR
- 208 standards for Basidiomycota, Ascomycota, and bacteria from biocrusts using the TOPO TA Cloning® Kit (ThermoFisher Scientific, MA, USA) as outlined by
- Anderud et al. (2013). The coefficients of determination (R^2) for our assays ranged from 0.90 to 0.99, and amplification efficiencies fell between 0.99 and 1.92. We
- analyzed the relationships between biocrust $\delta^{15}N$ and the gene copy number of Ascomycota, Basidiomycota, and bacteria by creating linear mixed effects models for
- 214 each biocrust type and N form combination (2 biocrust types \times 3 circular plots locations \times 2 N forms $\times \approx 1 - 5$ biocrusts = 48). DNA from eight moss-dominated and
- 216 four cyano-dominated biocrusts were difficult to amplify, especially one moss biocrust circular plot reducing the number of replicants from five to one. We used the
- 218 *lm* function in R and treated plot as a random factor and gene copy number nested within plot. We further tested for differences in our biomass estimates between the
- 220 crust types using multiple t-tests and a Benjamini-Hochberg correction to control for the false discovery rate associated with multiple comparisons (Benjamini and
- 222 Hochberg 1995).

224 **2.5 Biocrust fungal communities**

To identify the fungal taxa participating in N translocation, we characterized fungal

- communities in biocrusts using bar-coded sequencing. We PCR amplified the V9 region of the 18S rRNA gene using a universal eukaryote primer set, 1391F and
- EukBr, with a unique 12-bp Golay barcode fused to EukBr (Amaral-Zettler et al.2009, Hamady et al. 2008). Thermocycler parameters were similar to qPCR analyses
- 230 and consisted of a denaturation step at 94°C for 3 min, followed by 35 cycles of denaturation at 94°C for 45 s, an annealing step at 57°C for 60 s, elongation at 72°C
- for 90 s, and a final extension at 72°C for 10 min. We then purified and pooled PCR amplicon libraries to near equimolar concentrations using SequalPrep[™]
- 234 Normalization Plate Kits (Invitrogen, Carlsbad, CA, USA) and quantified the amplicon libraries by real-time qPCR using a KAPA Library Quantification Kit (Kapa
- Biosystems, Wilmington, MA, USA). All samples were sequenced at the BrighamYoung University DNA Sequencing Center (http://dnasc.byu.edu/) using the Illumina
- HiSeq 2500 platform (Illumina Biotechnology, San Diego, CA, USA), generating 2 ×
 250 paired-end reads. Illumina sequence reads were analyzed within QIIME (v.
- 240 1.9.1), an open-source software pipeline suitable for microbial community analysis(Caporaso et al. 2010). We removed barcodes and primers with a custom, in-house
- 242 script previous to joining paired-end reads by using fastq-join under default parameters (Aronesty 2011). Joined reads were then de-multiplexed and checked for
- chimeras (Haas et al. 2011). We then clustered the de-multiplexed reads into OTUs, applying a similarity threshold of 97%, using QIIME's default OTU clustering tool-

- 246 uclust (Edgar et al. 2011). Taxonomies of representative OTUs were assigned using uclust and the 18S rRNA gene SILVA 128 database which was clustered into OTUs
- 248 at 97% similarity (Quast et al. 2013). To evaluate if biocrust type supported similar fungal composition, we calculated the relative recovery of 27 fungal orders, including
- 250 dark septate lineages. We tested for differences between biocrust types using t-tests and a Benjamini-Hochberg correction (2 biocrust types × 6 randomly selected

biocrusts = 12).

254 **3 Results**

3.1 Translocation of ¹⁵NH₄⁺ in cyanobacteria biocrusts

The movement of ¹⁵N was clearly apparent in cyanobacteria-dominated biocrusts
following the addition of ¹⁵NH₄⁺, with δ¹⁵N decreasing as the radial distance from the
central application point of ¹⁵NH₄⁺ increased. In the overall model:

260
$$\delta^{15}$$
N biocrust = 51 – 0.71 (distance) + 0.06 (distance × plot) (eq. 1)

- 262 ($R^2 = 0.67$, $F_{2,12} = 11$, P = 0.002, figure 1A), distance was significant (P = 0.003) and possessed a negative slope. Surrounding the tracer application, $\delta^{15}N$ was enriched
- 264 upwards of 40‰ more than 20 cm away and continued to be enriched to approximately 10‰ almost 100 cm away with enrichment ranging from 9.1 – 49‰.
- 266 There was no such relationship between $\delta^{15}N$ and the radial distance from any other crust type or N form as demonstrated by the mixed effect models (cyanobacteria

- 268 biocrust ¹⁵NO₃⁻ addition: $R^2 = 0.17$, $F_{2,12} = 1.2$, P = 0.32, figure 1B; moss biocrust ¹⁵NH₄⁺ addition: $R^2 = 0.01$, $F_{2,12} = 0.07$, P = 0.94, figure 1A; and moss biocrust
- ¹⁵NO₃⁻ addition: $R^2 = 0.16$, $F_{2,12} = 1.2$, P = 0.33, figure 1B). The location of the three circular plots was not a significant factor in any linear mixed effects models for

biocrusts.

3.2 Potential movement of ${}^{15}NH_4^+$ **into grass leaves in cyanobacteria biocrusts** None of the label appeared to reach *A. hymenoides* leaves in either of the two types of

- 276 biocrusts. However, in cyanobacteria-dominated biocrusts where ¹⁵N from ¹⁵NH₄⁺ was translocated, the mixed effects model relating δ^{15} N from ¹⁵NH₄⁺ into *A*.
- 278 *hymenoides* leaves was significant ($R^2 = 0.37$, $F_{2,17} = 5.1$, P = 0.02):

280
$$\delta^{15}$$
N leaves = 3.1 – 0.03 (distance) + 0.01 (distance × plot) (eq. 2)

- and the radial distance was marginally significant (P = 0.08, data not shown). The δ^{15} N in leaves declined the further away the grass was from the ¹⁵N application point,
- with leaf enrichment between 1.0 6.7%. In the other three models, the R², F values, and *P*-values of all other mixed effects models ranged from 0.06 0.10, 0.46 0.67,
- and 0.54 0.64 respectively (data not shown). The $\delta^{15}N$ found in *A. hymenoides* leaves was 2.6‰ ±0.45 (means ±SEM) in cyanobacteria biocrusts exposed to $^{15}NO_3^{-1}$
- and $0.85\% \pm 0.47$ in moss biocrusts.

3.3 ¹⁵NH₄⁺ movement in cyanobacteria biocrusts related to Ascomycota
 The biocrust that translocated N also exhibited a robust relationship between the

- 292 proxy for Ascomycota biomass and biocrust δ¹⁵N. In cyanobacteria biocrusts, the greater the gene copy number of Ascomycota the lower the δ¹⁵N from ¹⁵NH₄⁺. In the
 294 overall model:
- 296 $\delta^{15}N = 280 37$ (gene copy number Ascomycota) + 1.1 (distance × plot) (eq. 3)
- 298 ($R^2 = 0.63$, $F_{2,8} = 6.8$, P = 0.02, figure 2A), the proxy for Ascomycota biomass was significant (P = 0.01) and negatively related to $\delta^{15}N$. In cyanobacteria crusts exposed
- 300 to ¹⁵NO₃⁻, and in the moss-dominated crusts, however, there was no such relationship between δ^{15} N from ¹⁵NO₃⁻ and Ascomycota gene copy number (cyanobacteria crust
- 302 model: $R^2 = 0.08$, $F_{2,12} = 0.6$, P = 0.58, figure 2B; moss crust model: $R^2 = 0.16$, $F_{2,12} = 1.2$, P = 0.33, figure 2B) or in moss crusts exposed to ¹⁵NH₄⁺ additions ($R^2 = 0.01$,
- 304 $F_{2,12} = 0.07$, P = 0.94, figure 2A). Basidiomycota and bacteria biomass in both crust types was not related to either N form (data not shown). The biomass estimates of all
- 306 measured biocrust components were consistently higher in moss- than cyanobacteriadominated crusts. Basidiomycota biomass was $1.5 \times 10^9 \pm 5.5 \times 10^8$ in cyanobacteria and
- 308 $5.8 \times 10^9 \pm 7.2 \times 10^8$ in moss biocrusts (t-test, t = 4.5, P < 0.0001, df = 1, data not shown). Ascomycota biomass was $2.6 \times 10^7 \pm 4.5 \times 10^6$ in cyanobacteria and 1.1×10^8
- 310 $\pm 2.4 \times 10^7$ in moss biocrusts (t-test, t = 3.3, *P* = 0.003, *df* = 1). Bacterial biomass was at least two orders of magnitude lower in cyanobacteria biocrusts

312 (cyanobacteria= $5.5 \times 10^6 \pm 8.9 \times 10^5$ and moss crusts= $2.7 \times 10^7 \pm 4.8 \times 10^6$, (t-test, t = 4.5, P < 0.0001, df = 1).

314

3.4 Dark septate fungi as major components of biocrusts

- 316 Four of the nine fungal orders, known to contain dark septate endophyte members, were present in both biocrust types, with the Pleosporales and Pezizales being
- 318 dominant taxa. In biocrusts, fungi comprised much of eukaryotic community (cyanobacteria = $30\% \pm 4.7$ and moss = $33\% \pm 4.0$), Ascomycota was the most
- 320 common fungal division (cyanobacteria = $82\% \pm 2.8$ and moss = $87\% \pm 2.9$), and orders with known dark septate members accounted for at least 67% of the
- 322 Ascomycota (cyanobacteria = $83\% \pm 4.8$ and moss = $67\% \pm 8.6$, figure 3). In cyanobacteria biocrusts, Pleosporales accounted for $66\% (\pm 6.9)$ of all dark septates
- and the recovery of this taxa was twice as high in cyanobacteria- than mossdominated crusts (t-test, t = 03.0, P = 0.01, df = 1). Even though the relative
- 326 abundance of Pleosporales differed, the number of gene copies of Pleosporales was similar between the two biocrusts (cyanobacteria = $1.7 \times 10^7 \pm 6.3 \times 10^6$ and
- 328 moss= $2.9 \times 10^7 \pm 1.3 \times 10^7$, t-test, t = 0.99, P = 0.35, df = 1) as determined by an extrapolation of qPCR values in conjunction with percent recovery of taxa for
- 330 Ascomycota. The *Pezizales* comprised a relatively large percentage of the biocrust community in moss-dominated biocrust with a recovery of 15% (\pm 8.3) and 28% (\pm
- 9.0) in cyanobacteria- and moss-dominated crusts, respectively (t-test, t = 1.1, P =

0.32, df = 1). Eukaryotic community data was based on the recovery of 1,232,312 334 quality sequences and 5,176 unique OTUs.

336 4 Discussion

In biological networks, the magnitude and direction of resource transfer in fungi is

- 338 predominantly thought to be influenced by the physiological source-sink gradients created by individual plants (Fellbaum et al. 2014) or between plants (Weremijewicz
- 340 et al. 2016). However, fungi may be more than just passive conduits by exerting control over resources due to their own sink-source resource needs (Simard and
- 342 Durall 2004). Our findings suggest that a minor, localized rainfall event may initiate fungal loops by allowing the rapid translocation of N in biocrusts at a rate of 40 mm
- h^{-1} in twenty-four hours. The movement of N was only apparent in the cyanobacteriadominated crusts, where δ^{15} N decreased as the distance from the addition of the
- 15 NH₄⁺ label increased. Further, the presence of Ascomycota was related to the amount of biocrust δ^{15} N from 15 NH₄⁺, with the isotope being diluted in soils with
- 348 higher levels of Ascomycota biomass. Eighty-three percent of the Ascomycota were from four fungal orders containing known dark septate endophytes and 66% of these
- 350 taxa were from one order, the Pleosporales. Taken together, our results suggest that fungal loops are potentially structured by fungal constituents, especially Pleosporales,
- 352 translocating N from NH_4^+ over NO_3^- .

354 4.1 Fungal loops only in cyanobacteria-dominated crusts

Our findings suggest that fungal loops do occur in cyanobacteria-dominated biocrusts.

- 356 Biocrust components, predominantly cyanobacteria, are known to fix and secrete up to 50% of their newly fixed C and 88% of N into surrounding soils within minutes to
- 358 days of fixation, depending on precipitation characteristics (Belnap et al. 2003). Thus, active crust constituents may exude resources for other biocrust constituents, such as
- fungi, to acquire and initiate fungal loops. Our results are similar to Green et al.
 (2008) whose previous work identified loops in cyanobacteria biocrusts moving ¹⁵N
- 362 from isotopically labeled nitrate in the Chihuahuan Desert. Although we only found movement of ¹⁵N as labeled ammonium, our label moved at a comparable rate to
- 364 Green et al. (2008; 40 mm h^{-1} versus 44 mm h^{-1} , respectively). The lack of ¹⁵N movement from labeled nitrate in our study may stem from the short time period that
- 366 we employed to measure translocation. When Green et al. (2008) evaluated ¹⁵N movement at two time points, 24 hours and four days after tracer application, they
- 368 found that biocrust enrichment consistently increased over time. Therefore, if we had continued to assess biocrusts isotopic signatures after 24 hours, we potentially could
- are a swell.

372 **4.2** Addressing the lack of loops in moss-dominated crusts

The seeming lack of loops in moss-dominated crusts may stem from the relatively

- 374 large moss biomass of *S. caninervis* effectively absorbing and holding N from our minor wet deposition event. We measured elevated levels of biomass of all biocrust
- 376 constituents (i.e., mosses, fungi, and bacteria) in moss-dominated relative to

cyanobacteria-dominated biocrusts. If most biocrust constituents were N limited,

- 378 including mosses, then the ¹⁵N was possibly retained in place to meet N deficiencies before being disseminated to other crust constituents outside the application zone. The
- 380 ability of mosses to scavenge atmospheric deposited N is well recognized in other systems (Liu et al. 2013, Fritz et al. 2014). Most mosses acquire N from either
- 382 wet/dry atmospheric deposition (Yaunming et al. 2016) or as biogenic sources from cyanobacterial associations on their leaves (Rousk et al. 2013). Mosses do alter N
- 384 dynamics in arid environments. When *S. caninervis* was lost from our system, a dramatic increase in nitrification rates were observed (Reed et al. 2012). The higher
- 386 levels of nitrification were most likely supported by the decomposition of dead moss biomass and subsequent release of new NH₄⁺; however, after the moss mortality,
- inorganic N pooled as NO₃⁻, in the remaining cyanobacteria-dominated biocrust.
 Thus, the presence or absence of mosses has the potential to alter N pools and fluxes
- in biocrusts.

The lack of ¹⁵N movement in moss-dominated crusts may reside in the nature of

- 392 our minor rainfall event. Our moss, *S. caninervis*, became photosynthetically active following the 2 mm rainfall event, changing in color from brown to green, but only in
- 394 the discrete biocrust patches that we watered. Mosses, including *S. caninervis*, are stimulated by minor rainfall events (Wu et al. 2014), with events as small as 1 mm
- 396 activating moss photosynthesis (Coe et al. 2012). Our rainfall event was intended to wet a small circle of biocrust to a depth of 1 cm. However, the additional
- 398 aboveground biomass of mosses and the rugose topography of moss-dominated crusts

relative to the smooth cyanobacteria-dominated crusts may have limited the depth our

- 400 minor rainfall event penetrated the soil and, in turn, activated other biocrust components. Also, water from our event might have evaporated more quickly from
- 402 the mossy biocrust surface, limiting the activity time of all constituents involved. To more conclusively determine the potential for fungal loops to exist in moss-dominated
- 404 biocrusts, more information is needed to determine the importance of effective rainfall size in initiating fungal loops.
- 406 N may move differently in moss-dominated biocrusts compared to other biocrust types. If desert mosses have fungal associations, then fungi have the potential to move
- 408 sequestered N to the mosses in a new kind of loop. Rhizoids and thalli of bryophytes may contain fungal associations (Pressel et al. 2010). Dark septate endophytes
- 410 colonize mosses (Day and Currah 2011). Taken together, the exchange ofphotosynthate and N may occur in a tighter, more localized loop between mosses and

412 dark septate fungi, even following a minor rainfall event.

414 **4.3 Loops may preferentially move NH4⁺ over NO3⁻**

We found that NH4⁺, but not NO3⁻, was rapidly translocated within cyanobacterial

- 416 crusts. The enrichment of δ^{15} N from 15 NH₄⁺ in cyanobacteria biocrusts was related to our proxy for Ascomycota biomass and potentially dark septate fungi due to their
- 418 dominance in our sequencing effort. We explain the negative relationship between Ascomycota gene copy number and $\delta^{15}N$ signal as a simple dilution—the higher our
- 420 biomass estimates of Ascomycota, the more spread in the ¹⁵N signal. Although the

physiology of dark septate fungi remains relatively unexplored, if our desert fungi are

- 422 like most other fungi, then the preferential movement of NH_4^+ is understandable. Generally, fungi prefer NH_4^+ over NO_3^- (Eltrop and Marschner 1996), as NH_4^+ is
- 424 readily acquired by fungi and assimilated into amino acids. After NH₄⁺ uptake and assimilation via the glutamate synthase or GS/GOGAT cycle, N is incorporated into
- 426 arginine through the urea cycle (Jin et al. 2012) due to the direct assimilation of NH_4^+ into the GS/GOGAT pathway (Courtly et al., 2015). Thus, NH_4^+ is most likely
- 428 transformed into arginine and moved within mycelium by amino acid transporters (Govindarajulu et al. 2005, Garcia et al. 2016). There is some evidence for the
- 430 importance of amino acid transporters in fungal endophytes, as arbuscularcolonization leads to an increase in the uptake of arginine and multiple other amino
- 432 acids (e.g., phenylalanine, lysine, asparagine, histidine, methionine, tryptophan, and cysteine) by their host plants (Whiteside et al. 2012). However, to verify the
- 434 movement of N through Ascomycota and the role of biomass in translocation, a more direct approach is needed. For example, quantum dots (fluorescent nanoscale
- 436 semiconductors) have tracked the flow of organically derived N into arbuscular mycorrhizae and into *Poa annua* in less than 24 hours (Whiteside et al. 2009). The N
- 438 form NO_3^- did move into our cyanobacteria crusts but not nearly to the extent reported by Green et al. (2008). Besides fungal preferences, other factors may play a role in the
- 440 uptake of N, such as the increase in mobility of NO_3^- in soils, differences in soil cation exchange capacity due to clay content, or fungi capitalizing on the more
- 442 abundant N form specific to a soil. Unfortunately, based on our design, we were

unable to distinguish the form of N captured or translocated by biocrust constituents.

- 444 More information is needed to identify the importance of N form and the movement of organic N within fungal loops.
- 446

4.4 Dark septate and Pleosporales as conduits

- 448 Our results support the idea that Pleosporales are the most likely conduits for N. Four of the nine fungal orders we identified contained known dark septate endophytes with
- one order, Pleosporales, being the most abundant. The Pleosporales accounted for66% of the Ascomycota taxa in cyanobacteria crusts. Based on the relationship
- between $\delta^{15}N$ and Ascomycota biomass, the overwhelming abundance of Pleosporales, and the universal occurrence of Ascomycota in biocrusts, the
- 454 Pleosporales assumedly play a role in fungal loops. We are not the first to reach this conclusion. Green et al. (2008) also identified Pleosporales as being the primary
- 456 candidate involved in fungal loops. In their semi-arid grassland, Pleosporales were the most common taxa on *Bouteloua* roots, in the rhizosphere, and in biocrusts. We found
- 458 799 operational taxonomic units, based on 97% similarity, with all of the identifiable sequences, belonging to three genera: *Leptosphaeria* (1.6% of Pleosporales
- 460 sequences), *Morosphaeria* (3.8% of Pleosporales sequences), and *Ophiosphaerella* (8.1% of Pleosporales sequences; data not shown). *Leptosphaeria* and
- 462 *Ophiosphaerella* may be pathogenic endophytes on grass species (Martin et al. 2001, Yuan et al. 2017), but may also be beneficial by delaying and reducing the symptoms
- 464 of other fungal pathogens (Yuan et al. 2017). However, 86% of our Pleosporales taxa

were unidentifiable based on our target genetic region and are potentially novel,

- 466 suggesting that much remains unknown about dark septates in deserts. Further research is needed to address the theory of Pleosporales conduits in biocrusts and the
- 468 ecological importance of the dark septate endophytes in desert systems.

470 **4.5 Little N translocation to grass**

Due to the discrete nature of our minor, localized rainfall event, we were not surprised

472 that little of the label entered the leaves of *A. hymenoides*. In the fungal loop hypothesis, a larger rainfall event triggers the plant to become a sink for the N

- 474 building up in fungi over previous minor rainfall events. We conducted our experiment absent of a larger rainfall event and our 2.5 mm rainfall event was applied
- 476 over a 5 cm diameter circle of soil in early summer. When a similar precipitation event size (2 mm) was applied across a much larger area (4 x 4 m² plot) on Colorado
- 478 Plateau soils during spring or summer, the predawn water potential of *A. hymenoides* was similar one day after and one day prior to watering (Schwinning et al. 2003).
- 480 Thus, our minor rainfall event most likely failed to alter the water status of the grass or cause grasses to become a sink for N. However, we did find some evidence of ¹⁵N
- 482 in *A. hymenoides* leaves following our ${}^{15}NH_4^+$ application in cyanobacteria-dominated crusts. The pattern of the enrichment, a dilution with increasing distance from the
- 484 application point, was similar to the pattern found in biocrusts, but the most enriched leaves (6.7‰) were less than the least enriched crust (9.1‰). Even though the
- 486 enrichment was minor, the presence of ¹⁵N in *A. hymenoides* provides some evidence

of biocrust constituents being linked to plants and the translocation of N occurring

- 488 even in the absence of a larger rainfall event to enhance grass activity. In April, at the time of the experiment, *A. hymenoides* was photosynthetically active. If we had added
- 490 more label or evaluated the isotope signature of roots, we may have detected more ¹⁵N label in grass tissue.

492

5. Conclusion

- 494 Cyanobacterial biocrusts are potentially interconnected in extensive biological networks. In light of the absence of N movement in moss-dominated crusts, mosses
- 496 may hinder fungal loops. The abundant moss biomass in some biocrust may absorb and retain the applied N label. Due to the dominance of dark septate endophytes in
- 498 biocrusts, fungi may act as conduits within the network. Our results add to the indirect evidence of fungal loops, but more information is needed to quantify the
- 500 environmental conditions and biocrust constituents controlling the magnitude and directionality of the translocation of N to vascular plants.

502

Acknowledgements The portion of the research conducted by Dr. Wu and Dr.

- 504 Yuanming was funded by National Natural Science Foundation of China (grant # 41571256 and 41771299). Dr. Belnap thanks the Ecosystems Program of U.S.
- 506 Geological Survey. Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

Author contributions ZTA and JB designed the study. ZTA, TBS, NW, AST, and JB

- 510 conducted the experiments. ZTA, TBS, NW, AST, YZ, and JB analyzed and interpreted the data. ZTA, TBS, NW, AST, YZ, and JB helped write and review the
- 512 manuscript. ZTA agrees to be accountable for all aspects of the work in ensuring that questions related to the accuracy or integrity of any part of the work are appropriately
- 514 investigated and resolved.
- 516 **Conflict of interest** The authors declare no conflict of interest.

520 Figure legend

Figure 1 Cyanobacteria biocrusts facilitated the translocation of N from ¹⁵NH₄⁺

- 522 in twenty-four hours. Based on linear mixed effects models, δ^{15} N, from 15 NH₄⁺, (R² = 0.67, F_{2,12} = 11, *P* = 0.002, A) decreased as the radial distance from the isotopic
- 524 application increased. There was no such relationship between $δ^{15}N$ and the radial distance from any other crust type or N form (A, B) Values are $δ^{15}N$ (‰) from two
- 526 biocrust types, cyanobacteria- and moss-dominated crusts, across circular plots (radius of 1.0 m) with a central application (5 cm diameter circle) of $^{15}\text{NH}_4^+$ or $^{15}\text{NO}_3^-$
- 528 in a simulated, minor rainfall event.

530 Figure 2 Ascomycota biomass influenced the distance N traveled. In

cyanobacteria crusts, δ^{15} N, from 15 NH₄⁺, was diluted as Ascomycota gene copy

- 532 number increased (R² = 0.63, F_{2,8} = 6.8, *P* = 0.02, A). A similar pattern was not apparent in any other crust type or N form (A, B). Values are δ^{15} N (‰) from biocrusts
- 534 and Ascomycota gene copy numbers, an approximation of biomass, from qPCR of the ITS region with primer pair ITS5 and ITS4A.

536

Figure 3 Pleosporales were the dominant Ascomycota order and contained

- 538 **dark septate species.** Pie chart values are means (*n*=6) of the relative recovery from nine fungal orders, four of which contain dark septate endophytic taxa. Recovery was
- 540 based on OTUs from eukaryotic community libraries of the 18S rRNA gene (97% similarity cutoff).

References cited

544	Aanderud ZT, Jones SE, Schoolmaster DR, Fierer N, Lennon JT. 2013. Sensitivity of
	soil respiration and microbial communities to altered snowfall. Soil Biology &
546	Biochemistry 57:217-227.
	Aguilera LE, Armas C, Cea AP, Gutierrez JR, Meserve PL, Kelt DA. 2016. Rainfall,
548	microhabitat, and small mammals influence the abundance and distribution of soil
	microorganisms in a Chilean semi-arid shrubland. Journal of Arid Environments
550	126:37-46.
	Amaral-Zettler LA, McCliment EA, Ducklow HW, Huse SM. 2009. A method for
552	studying protistan diversity using massively parallel sequencing of V9
	hypervariable regions of small-subunit ribosomal RNA genes. PLOS ONE (art.
554	e6372)
	Aronesty E. 2011. Command-line tools for processing biological sequence data.
556	Expression Analysis. (14 October 2017;
	https://github.com/ExpressionAnalysis/ea-utils)
558	Barrow JR. 2003. Atypical morphology of dark septate fungal root endophytes of
	Bouteloua in arid southwestern USA rangelands. Mycorrhiza 13:239-247.
560	Bates ST, Nash TH, Garcia-Pichel F. 2012. Patterns of diversity for fungal
	assemblages of biological soil crusts from the southwestern United States.
562	Mycologia 104:353-361.
	Behie SW, Bidochka MJ. 2014. Nutrient transfer in plant-fungal symbioses. Trends in
564	Plant Science 19:734-740.

Belnap J, Hawkes CV, Firestone MK. 2003. Boundaries in miniature: Two examples

from soil. BioScience 53:739-749.

Belnap J, Phillips SL, Sherrod SK, Moldenke A. 2005. Soil biota can change after

- 568 exotic plant invasion: does this affect ecosystem processes? Ecology 86:3007-3017.
- 570 Benjamini Y, Hochberg Y. 1995. Controlling the false discovery rate: a practical and powerful approach to multiple testing. Journal of the Royal Statistical Society

572 Series B: (Methodological) 57:289-300.

Cable JM, Huxman TE. 2004. Precipitation pulse size effects on Sonoran Desert soil

574 microbial crusts. Oecologia 141:317-324.

Caporaso JG, et al. 2010. QIIME allows analysis of high-throughput community

576 sequencing data. Nature Methods 7:335-336.

Coe KK, Belnap J, Sparks JP. 2012. Precipitation-driven carbon balance controls

- 578 survivorship of desert biocrust mosses. Ecology 93: 1626-1636Collins SL, Sinsabaugh RL, Crenshaw C, Green L, Porras-Alfaro A, Stursova M,
- 580 Zeglin LH. 2008. Pulse dynamics and microbial processes in aridland ecosystems.Journal of Ecology 96:413-420.
- 582 Collins SL, et al. 2014. A multiscale, hierarchical model of pulse dynamics in aridland ecosystems. Annual Review of Ecology, Evolution, and Systematics 45:397-

584 419.

Darrouzet-Nardi A, Reed SC, Grote EE, Belnap J. 2015. Observations of net soil

- 586 exchange of CO₂ in a dryland show experimental warming increases carbon lossesin biocrust soils. Biogeochemistry 126:363-378.
- 588 Day MJ, Currah RS. 2011. Role of selected dark septate endophyte species and other hyphomycetes as saprobes on moss gametophytes. Botany-Botanique 89:349-359.
- 590 Edgar RC, Haas BJ, Clemente JC, Quince C, Knight R. 2011. UCHIME improves sensitivity and speed of chimera detection. Bioinformatics 27:2194-2200.
- 592 Eltrop L, Marschner H. 1996. Growth and mineral nutrition of non-mycorrhizal and mycorrhizal Norway spruce (Picea abies) seedlings grown in semi-hydroponic

sand culture. New Phytologist 133:469-478.

Fellbaum CR, Mensah JA, Cloos AJ, Strahan GE, Pfeffer PE, Kiers ET, Bucking H.

596 2014. Fungal nutrient allocation in common mycorrhizal networks is regulated by the carbon source strength of individual host plants. New Phytologist 203:646-

598 656.

Fierer N, Jackson JA, Vilgalys R, Jackson RB. 2005. Assessment of soil microbial

600 community structure by use of taxon-specific quantitative PCR assays. Applied and Environmental Microbiology 71:4117-4120.

Fritz C, Lamers LPM, Riaz M, van den Berg LJL, Elzenga T. 2014. Sphagnum mosses-masters of efficient N-uptake while avoiding intoxication. PLOS ONE
 (art. e79991).

Garcia K, Doidy J, Zimmermann SD, Wipf D, Courty PE. 2016. Take a trip through

- 606 the plant and fungal transportome of mycorrhiza. Trends in Plant Science 21:937-950.
- 608 Gostincar C, Grube M, de Hoog S, Zalar P, Gunde-Cimerman N. 2010. Extremotolerance in fungi: evolution on the edge. FEMS Microbiology Ecology

610 71:2**-**11.

Govindarajulu M, Pfeffer PE, Jin HR, Abubaker J, Douds DD, Allen JW, Bucking H,

- 612 Lammers PJ, Shachar-Hill Y. 2005. Nitrogen transfer in the arbuscular mycorrhizal symbiosis. Nature 435:819-823.
- 614 Green LE, Porras-Alfaro A, Sinsabaugh RL. 2008. Translocation of nitrogen and carbon integrates biotic crust and grass production in desert grassland. Journal of
- 616 Ecology 96:1076-1085.

Haas BJ, et al. 2011. Chimeric 16S rRNA sequence formation and detection in Sanger

- and 454-pyrosequenced PCR amplicons. Genome Research 21:494-504.Hamady M, Walker JJ, Harris JK, Gold NJ, Knight R. 2008. Error-correcting
- 620 barcoded primers for pyrosequencing hundreds of samples in multiplex. Nature Methods 5:235-237.
- 622 Hattenschwiler S, Tiunov AV, Scheu S. 2005. Biodiversity and litter decomposition in terrestrial ecosystems. Annual Review of Ecology 36:191-218.
- 624 He XH, Critchley C, Bledsoe C. 2003. Nitrogen transfer within and between plants through common mycorrhizal networks (CMNs). Critical Reviews in Plant
- 626 Sciences 22:531-567.

Huxman TE, Snyder KA, Tissue D, Leffler AJ, Ogle K, Pockman WT, Sandquist DR,

- 628 Potts DL, Schwinning S. 2004. Precipitation pulses and carbon fluxes in semiarid and arid ecosystems. Oecologia 141:254-268.
- 630 Jin HR, Liu J, Huang XW. 2012. Forms of nitrogen uptake, translocation, and transfer via arbuscular mycorrhizal fungi: A review. Science ChinaLife Sciences 55:474-

632 482.

Johnson NC, Graham JH, Smith FA. 1997. Functioning of mycorrhizal associations

along the mutualism-parasitism continuum. New Phytologist 135:575-586.

Jumpponen A, Trappe JM. 1998. Dark septate endophytes: a review of facultative

- biotrophic root-colonizing fungi. New Phytologist 140:295-310.Larena I, Salazar O, Gonzalez V, Julian MC, Rubio V. 1999. Design of a primer for
- ribosomal DNA internal transcribed spacer with enhanced specificity for ascomycetes. Journal of Biotechnology 75:187-194.
- 640 Liu XY, Koba K, Makabe A, Li XD, Yoh M, Liu CQ. 2013. Ammonium first: natural mosses prefer atmospheric ammonium but vary utilization of dissolved organic
- 642 nitrogen depending on habitat and nitrogen deposition. New Phytologist 199:407-419.
- Martin DL, Bell GE, Baird JH, Taliaferro CM, Tisserat NA, Kuzmic RM, Dobson DD, Anderson JA. 2001. Spring dead spot resistance and quality of seeded
- bermudagrasses under different mowing heights. Crop Science 41:451-456.Osono T. 2007. Ecology of ligninolytic fungi associated with leaf litter
- decomposition. Ecological Research 22:955-974.

Porras-Alfaro A, Herrera J, Natvig DO, Lipinski K, Sinsabaugh RL. 2011. Diversity

- and distribution of soil fungal communities in a semiarid grassland. Mycologia103:10-21.
- 652 Pressel S, Bidartondo MI, Ligrone R, Duckett JG. 2010. Fungal symbioses in bryophytes: new insights in the twenty first century. Phytotaxa 9:238-253.
- Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glockner FO.2013. The SILVA ribosomal RNA gene database project: improved data
- processing and web-based tools. Nucleic Acids Research 41:D590-D596.

R Development Core Team, 2017. R: A language and environment for statistical

- 658 computing. R Foundation for Statistical Computing, Vienna, Austria, URL http://www.R-project.org.
- 660 Reed SC, Coe KK, Sparks JP, Housman DC, Zelikova TJ, Belnap J. 2012. Changes to dryland rainfall result in rapid moss mortality and altered soil fertility. Nature

662 Climate Change 2:752-755.

Rousk K, Jones DL, DeLuca TH. 2013. Moss-cyanobacteria associations as biogenic

- sources of nitrogen in boreal forest ecosystems. Frontiers in Microbiology 4: 150.Saikkonen K, Wali P, Helander M, Faeth SH. 2004. Evolution of endophyte-plant
- 666 symbioses. Trends in Plant Science 9:275-280.Schneider T, Keiblinger KM, Schmid E, Sterflinger-Gleixner K, Ellersdorfer G,
- Roschitzki B, Richter A, Eberl L, Zechmeister-Boltenstern S, Riedel K. 2012.Who is who in litter decomposition? Metaproteomics reveals major microbial
- 670 players and their biogeochemical functions. ISME Journal 6:1749-1762.

Simard SW, Durall DM. 2004. Mycorrhizal networks: a review of their extent,

- 672 function, and importance. Canadian Journal of Botany 82:1140-1165.Sperry LJ, Belnap J, Evans RD. 2006. Bromus tectorum invasion alters nitrogen
- dynamics in an undisturbed arid grassland ecosystem. Ecology 87:603-615.Schwinning S, Starr BL, Ehleringer JR. 2003. Dominant cold desert plants do not
- 676 partition warm season precipitation by event size. Ecosytems Ecology 136:252-260.
- 678 Titus JH, Titus PJ, Nowak RS, Smith SD. 2002. Arbuscular mycorrhizae of MojaveDesert plants. Western North American Naturalist 62:327-334.
- 680 van der Heijden MGA, Horton TR. 2009. Socialism in soil? The importance of mycorrhizal fungal networks for facilitation in natural ecosystems. Journal of
- 682 Ecology 97:1139-1150.

Walder F, Niemann H, Natarajan M, Lehmann MF, Boller T, Wiemken A. 2012.

- 684 Mycorrhizal networks: common goods of plants shared under unequal terms of trade. Plant Physiology 159:789-797.
- 686 Wang C, et al. 2014. Aridity threshold in controlling ecosystem nitrogen cycling in arid and semi-arid grasslands. Nature Communications (art. 4799).
- 688 Welter JR, Fisher SG, Grimm NB. 2005. Nitrogen transport and retention in an arid land watershed: Influence of storm characteristics on terrestrial-aquatic linkages.
- Biogeochemistry 76:421-440.

Weremijewicz J, Sternberg L, Janos DP. 2016. Common mycorrhizal networks

- 692 amplify competition by preferential mineral nutrient allocation to large host plants. New Phytologist 212:461-471.
- 694 Whiteside MD, Treseder KK, Atsatt PR. 2009. The brighter side of soils: quantum dots track organic nitrogen through fungi and plants. Ecology 90:100-108.
- 696 Whiteside MD, Garcia MO, Treseder KK. 2012. Amino Acid Uptake in Arbuscular Mycorrhizal Plants. PLOS ONE (art. e47643).
- 698 Wu N, Zhang YM, Downing A, Aanderud ZT, Tao Y, Williams S. 2014. Rapid adjustment of leaf angle explains how the desert moss, Syntrichia caninervis,
- copes with multiple resource limitations during rehydration. Functional PlantBiology 41:168-177.
- Yahdjian L, Sala OE. 2010. Size of precipitation pulses controls nitrogen
 transformation and losses in an arid Patagonian ecosystem. Ecosystems 13:575-

704 585.

Yuan Y, Feng HJ, Wang LF, Li ZF, Shi YQ, Zhao LH, Feng ZL, Zhu HQ. 2017.

- 706 Potential of endophytic fungi isolated from cotton roots for biological control against verticillium wilt disease. PLOS ONE (art. e0170557):12.
- 708 Zelikova TJ, Housman DC, Grote EE, Neher DA, Belnap J. 2012. Warming and increased precipitation frequency on the Colorado Plateau: implications for
- biological soil crusts and soil processes. Plant and Soil 355:265-282.





Figure 2



Figure 3

