



1 **A new procedure for processing eddy-covariance data to better**
2 **quantify atmosphere-aquatic ecosystem CO₂ exchanges**

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10

11 **Abstract.** The capture of carbon by aquatic ecosystems and its sequestration in
12 sediments has been studied as a potential method for mitigating the adverse effects of
13 climate change. However, the evaluation of in situ atmospheric CO₂ fluxes is
14 challenging because of the difficulty in making continuous measurements over areas
15 and for periods of time that are environmentally relevant. The eddy covariance (EC)
16 method is the most promising approach to address this concern with the measurement of
17 atmospheric CO₂ fluxes. However, methods to process the data obtained from EC
18 measurements are still being developed, and the estimated air-water CO₂ fluxes have
19 large uncertainties and differ from those obtained using conventional methods. In this
20 study, we improved the post-processing procedure for the EC method to reduce the
21 uncertainty in the measured air-water CO₂ fluxes. Our new procedure efficiently
22 removes erroneous fluxes using a combination of filtering methods based on the
23 received signal strength indicator of the EC sensor, the normalized standard deviation of



24 atmospheric CO₂ and water vapor concentrations, and a high-pass filter. Our procedure
25 is easier to apply to EC measurements than existing correction methods. The improved
26 EC fluxes did not always agree with those obtained by using conventional methods
27 (e.g., the bulk formula method), but this difference was attributable to the difference of
28 measurement heights and the effect on the measured fluxes of the physical and
29 biological properties of the water surface (e.g., the presence of vegetation on the water
30 surface and the temperature gradient in the overlying atmospheric layer). Because the
31 measurement height and the spatiotemporal scales of the flux measurement depend on
32 the applied method, it is essential to select the appropriate method for studies related to
33 CO₂ fluxes and to the determination of ecosystem-atmospheric CO₂ interactions and the
34 role of aquatic ecosystems in mitigating the adverse effects of climate change.

35 **1 Introduction**

36 Aquatic environments are considered critical to the mitigation of adverse climate
37 change effects because of their ability to store atmospheric CO₂. Previous studies have
38 estimated that the ocean absorbs approximately one-fourth of the CO₂ emitted by
39 anthropogenic activities (IPCC, 2013). However, the effect of shallow aquatic
40 ecosystems on atmospheric CO₂ remains a controversial topic. Several previous studies
41 have concluded that shallow aquatic ecosystems are sources of atmospheric CO₂ after
42 taking account of carbon inputs from land (e.g. Gazeau et al., 2005; Borges et al., 2006;
43 Chen et al., 2013). In contrast, some autotrophic, shallow aquatic ecosystems have been
44 reported to be net sinks for atmospheric CO₂ (e.g. Schindler et al., 1997; Tokoro et al.,
45 2014).

46 In situ measurements of atmospheric CO₂ fluxes are necessary for precise analysis



47 of carbon cycling in aquatic environments. CO₂ fluxes in aquatic environments are
48 difficult to determine because of the variability of several factors, including
49 concentrations of CO₂ in the water and air and the physical characteristics of the
50 atmosphere and water surface. Several methods have been proposed for measuring in
51 situ CO₂ fluxes. Because each of these methods works best at a different combination of
52 spatial and temporal scales and is associated with different costs and technical
53 difficulties, a variety of methods have been applied to different aquatic environments
54 (e.g. oceans, estuaries, and lakes) to assess rates of aquatic carbon cycling.

55 Methods of estimating air-water CO₂ fluxes can be assigned to one of two
56 categories: (1) indirect estimations based on CO₂ concentration gradients just below the
57 water surface (Lewis and Whitman, 1924) or from the renewal rate of a very small body
58 of water (Danckwerts, 1951), and (2) direct estimations. With either of the indirect
59 methods, the CO₂ flux is calculated from the product of the difference in the CO₂
60 fugacity (fCO₂) between air and water, the CO₂ solubility, and a physically regulated
61 parameter called the transfer velocity. Because the transfer velocity cannot be estimated
62 directly, empirical and hydrodynamic models for estimating transfer velocity have been
63 proposed (Garbe et al., 2013).

64 At the present time, the empirical model is primarily used for evaluating aquatic
65 CO₂ fluxes because of the difficulty in applying the hydrodynamic model. In the
66 empirical model, the regulating factor for transfer velocity has been identified from
67 several direct CO₂ measurements by using tracers such as ¹⁴C and SF₆ (e.g. Broecker
68 and Peng 1982; Ho et al., 2014) or water-tank experiments (e.g. Komori et al., 1993).
69 Based on these results, several empirical equations have been formulated. The wind
70 speed above the water surface is a metric of one regulating factor (e.g. Liss and



71 Merlivat, 1986; Wanninkhof, 1992; Ho et al., 2006). In the case of shallow systems,
72 water velocity fields and depths have been used to estimate the gas transfer velocity
73 (O’Conner and Dobbins, 1958; Borges et al., 2004). In this study, the method using such
74 empirical model for estimating aquatic CO₂ flux is defined as the Bulk Formula method
75 (BF method).

76 However, application of the BF method is limited because of its poor temporal and
77 spatial coverage. Moreover, in previous studies, air-water CO₂ fluxes have been
78 determined mostly as snapshots that did not account for diurnal changes or annual
79 cycles, the result being considerable uncertainty and bias (Kuwae et al., 2016). In
80 brackish environments in particular, temporal variability of water fCO₂ is significant,
81 whereas the carbonate buffer effect is weak, and the fluctuations of fCO₂ become very
82 large (Zeebe and Wolf-Gladrow, 2001). Use of BF methods to carry out a
83 comprehensive analysis of dynamic carbon cycling in aquatic environments with large
84 spatial and temporal variability would therefore be very costly and require much effort.

85 Another method for evaluating the air-water CO₂ fluxes is direct measurement of in
86 situ fluxes, which involves use of a chamber floating on the water surface (e.g.
87 Frankignoulle, 1988; Tokoro et al., 2008) and eddy covariance devices (vide infra). The
88 floating chamber method is used to determine the air-water CO₂ flux from continuous
89 measurements of CO₂ concentrations in the air inside a hollow, box-shaped device
90 floating on the water surface. Although this method is the easiest of the direct methods
91 to use in shallow coastal waters because of its relative simplicity, like the BF method, it
92 is poorly suited for obtaining long-term measurements over wide areas.

93 The eddy covariance (EC) method, which is commonly used to determine mass and
94 heat fluxes in terrestrial environments, has recently been used to estimate air-water



95 fluxes of greenhouse gases (e.g. Lee et al., 2004). The determination of the EC CO₂ flux
96 is based on the micro-meteorological behavior of atmospheric eddy diffusion and is
97 calculated from the covariance of atmospheric CO₂ concentrations and vertical wind
98 speeds measured at high frequency (more than 10 Hz). Because EC measurements can
99 be performed automatically and represent the flux over a large area, the EC method can
100 be used to obtain a detailed analysis of CO₂ fluxes.

101 Despite the promise of EC measurements, application of the EC method in aquatic
102 environments remains challenging (Tsukamoto et al., 2004; Rutgersson and Smedman,
103 2010; Vesala, 2012; Blomquist et al., 2013; Kondo et al., 2014; Ikawa and Oechel,
104 2014; Landwehr et al., 2014). The difficulty of making aquatic EC measurements is that
105 the air-water CO₂ flux is small compared with the air-land CO₂ flux (Vesala, 2012;
106 Landwehr et al., 2014). The main technical problem with the EC method is cross
107 sensitivity, which reflects the interference between the atmospheric CO₂ and H₂O
108 measurements caused by spectrometric error (Kohsiek et al., 2000; Prytherch et al.,
109 2010; Kondo et al., 2014; Landwehr et al., 2014). A procedure based on the relationship
110 between atmospheric CO₂ concentration and relative humidity called the PKT
111 correction has been proposed to correct for the effects of cross sensitivity (Prytherch et
112 al., 2010). However, the PKT correction is not always effective. Past studies have
113 revealed that cross sensitivity is reduced only after application of certain operational
114 procedures such as cleaning the optical lens of the sensor (Ikawa and Oechel, 2014;
115 Kondo et al., 2014) and drying the sample gas (Landwehr et al., 2014).

116 There are several problems in addition to cross sensitivity in using EC
117 measurements in aquatic environments. The uncertainty of EC measurements has been
118 attributed to the spatial and temporal heterogeneity of water (Mørk et al., 2014). The EC



119 flux is calculated as the average within a measurement area called the “footprint”, which
120 can range from several hundred meters to several kilometers windward from the
121 measurement point (Schuepp et al., 1990). Therefore, EC fluxes at heterogeneous water
122 sites are different from the fluxes determined by methods that estimate the CO₂ flux in
123 an area of only several square meters (e.g., the BF method and floating chamber
124 method). The EC flux is an average flux over a certain time interval (approximately
125 several tens of minutes) (Lee et al., 2004), whereas the BF method estimates the flux at
126 the time of sampling. Thus, EC fluxes estimated at sites where fluxes are temporally
127 variable also differ from fluxes obtained using other methods. Furthermore, the inflow
128 of terrestrial air into the measurement site can generate uncertainty in the flux
129 measurement because the atmospheric CO₂ concentration over terrestrial vegetation
130 may differ significantly from the concentration over water. The inflow of terrestrial air
131 can cause unnatural temporal changes in the atmospheric CO₂ concentration and spatial
132 heterogeneity at the measurement site. It is therefore necessary to account for the
133 characteristics of the aquatic environment and apply appropriate post-processing (PP)
134 procedures (Leinweber et al., 2009) to avoid large uncertainties or biases in EC flux
135 calculations.

136 In this study, we developed a PP procedure for EC aquatic measurements. This PP
137 procedure involves the exclusion of erroneous data and correction of unnatural changes
138 in the atmospheric CO₂ using a series of data-filtering steps. The new process is based
139 on the idea that cross sensitivity and environmental heterogeneity during flux
140 measurement cause spikes, drifts, offsets, and long-term variation in the CO₂ and H₂O
141 raw data. We compared the results calculated with our new PP procedure to those
142 obtained using conventional EC PP procedures along with BF flux data as an example



143 of a frequently used method. We then discuss the differences between the conventional
144 BF and EC fluxes with respect to atmospheric and environmental characteristics.

145 **2 Methods**

146 **2.1 Field measurements**

147 Continuous EC measurement data were used for the evaluation of the PP procedure and
148 the analysis of atmospheric-aquatic ecosystem CO₂ exchange. The data were collected
149 from a brackish lagoon in Japan (the Furen Lagoon, Fig. 1) from 28 May to 21 October
150 2014, during which time the water surface was not frozen. Most of the study area (57.4
151 km²) is covered by seagrass meadows (mainly *Zostera marina*). The water is shallow
152 (1–2 m), except in a channel that connects the eastern and western basins of the lagoon
153 (depth = approximately 5 m). Freshwater flows into the western basin through several
154 rivers that run through the surrounding grass farms, and seawater is exchanged through
155 the lagoon mouth, which opens to the Okhotsk Sea. A previous study has found that the
156 air-water CO₂ flux in the lagoon is affected by changes of salinity caused by the inflow
157 of river water and tides as well as by changes of dissolved inorganic carbon resulting
158 from biological processes such as photosynthesis (Tokoro et al., 2014). The
159 measurement platform was built at the same site used in the study of Tokoro et al.
160 (2014) (N43° 19.775', E145° 15.463'); the effects of photosynthesis and changes in
161 salinity are most notable at this location in the lagoon (Tokoro et al., 2014).

162 The EC devices used in this study were as follows. Atmospheric CO₂ concentrations
163 and water vapor were measured with an open-path sensor (LI-7500A, LI-COR, USA).
164 The three-dimensional (3D) wind velocity, air temperature, and atmospheric pressure
165 were measured with a 3D acoustic Doppler velocimeter (CSAT-3, Campbell Scientific,



166 USA). The data were logged and managed by a SMARTFlux system (LI-COR). The
167 open-path sensor and the wind velocimeter were attached to the platform approximately
168 3.0–5.5 m above the water surface (the height varied with the tide). The sampling rate
169 for all data was 10 Hz, and the fluxes (CO₂, water vapor, and heat) were calculated as
170 averages over 30-min intervals. Batteries and solar panels were attached to the platform
171 as a power source. Battery replacement, data collection, and device maintenance were
172 performed approximately every two weeks. Water temperature and salinity were
173 measured continuously with a conductivity-temperature sensor (Compact-CT, Alec,
174 Japan).

175 2.2 Calculation of fluxes using the conventional PP procedure (PP1)

176 The air-water CO₂ flux (F) was calculated every 30-min using the following equation:

177

$$178 \quad F = \overline{\rho_c' w'} \cdot F_1 + \mu \frac{\rho_c}{\rho_d} \overline{\rho_v' w'} \cdot F_1 + \rho_c \left(1 + \mu \frac{\rho_v}{\rho_d} \right) \frac{\overline{T_a' w'}}{T_a} \cdot F_2, \quad (1)$$

179

180 where the coefficients F_1 and F_2 are correction terms based on the transfer functions that
181 correct for the frequency attenuation of the air-sea CO₂ flux caused by the response time
182 of the sensor, path-length averaging, sensor separation, signal processing, and flux-
183 averaging time (Massman, 2000). The first term on the right-hand side of Eq. (1) is the
184 product of F_1 and the uncorrected air-sea CO₂ flux calculated as the covariance of the
185 CO₂ density ρ_c and the vertical wind speed w (the bar and the prime indicate the mean
186 and the deviation from the mean, respectively). The second and third terms are the
187 Webb-Pearman-Leuning correction of latent heat and sensible heat, respectively (Webb



188 et al., 1980). The other variables in Eq. (1) are defined as follows: ρ_d = dry air density;
189 ρ_v = water vapor density; T_a = air temperature; and μ = ratio of the molar weight of dry
190 air to that of water vapor. The wind speed was corrected by using a double rotation to
191 make the average vertical wind speed zero during the 30-min time interval (Lee et al.,
192 2004). The footprint (measurement area) depends on several factors, including the
193 measurement height, wind speed, atmospheric stability, and measurement site roughness
194 (10^{-4} cm) (Schuepp et al., 1990). This footprint ranged from several hundred square
195 meters to several square kilometers on the windward side of the measurement site.

196 The deviation of each parameter in Eq. (1) from the 30-min average was calculated by
197 subtracting the 30-min average from the instantaneous data after deleting obviously
198 erroneous data (e.g., negative values of CO₂ or water vapor concentration). Other
199 corrections to the raw data included coordinate rotation of the 3D wind component
200 (double rotation; Lee et al., 2004), time lag of the measurement due to the separation of
201 the CO₂ sensor and the wind velocimeter (covariance maximization; Lee et al., 2000),
202 exclusion of wind data contaminated by the wind velocimeter flame, and correction of
203 the measurement noise (statistical tests of Vickers and Mahrt, 1997) based on the default
204 settings of the data management software (EddyPro 5.1.1, LI-COR).

205 **2.3 Calculation of flux using our new PP procedure (PP2)**

206 After calculating the EC flux using the conventional PP procedure (PP1) described in
207 Sect. 2.2, we recalculated the EC flux using our new procedure (PP2; Fig. 2). The PP2
208 procedure is mainly based on filtering and excluding erroneous data rather than on the
209 PKT method of data correction. The PP2 procedure is also focused on aquatic
210 environments in which the spatial and temporal variations in atmospheric CO₂ are large.



211 The PP2 procedure combines a series of filtering methods based on the received signal
212 strength indicator (RSSI) of the EC sensor, the normalized standard deviations (nSDs)
213 of the atmospheric CO₂ and water vapor concentrations, and high-pass (HP) filtering.

214 The RSSI is one of the data concurrently obtained from the CO₂ sensor of the EC
215 measurement instrumentation and indicates the available signal strength of the sensor.
216 This parameter has been used for evaluating measurement validation of objective gases
217 such as methane. In this study, we used the RSSI to filter the CO₂ data because both
218 CO₂ and methane absorb infrared radiation. First, data in the 30-min time series were
219 excluded if their RSSI was low. The RSSI threshold for exclusion was set to 90 %,
220 because this percentage was sufficiently high to identify valid CO₂ concentrations and
221 allowed most of the measurement data (~81 %) to be retained. This 90 % threshold was
222 also recommended as the quality criterion for the LI-7700 methane analyzer (LI-COR),
223 which has been used to measure methane concentrations.

224 ([http://www.licor.com/env/newslines/2012/04/overcoming-the-challenges-of-open-path-](http://www.licor.com/env/newslines/2012/04/overcoming-the-challenges-of-open-path-methane-measurements-with-the-li-7700/)
225 [methane-measurements-with-the-li-7700/](http://www.licor.com/env/newslines/2012/04/overcoming-the-challenges-of-open-path-methane-measurements-with-the-li-7700/)).

226 Second, the criteria for excluding erroneous fluxes were identified. Erroneous fluxes
227 were identified based on unnatural spikes, jumps, and shifts in the data. Outliers were
228 excluded based on three statistical parameters: (1) the normalized standard deviation,
229 nSD (i.e. the SD over a 30-min period divided by the average value for the entire
230 measurement period after RSSI filtering; (2) skewness; and (3) the absolute value of
231 kurtosis. Each of these threshold values was determined after the top four outliers for
232 each parameter were excluded; the raw data (CO₂ concentration, wind speed) associated
233 with the top four outliers were assumed to be erroneous changes that would not happen
234 naturally.



235 Finally, HP filtering was applied to the calculation of the deviations from the mean
236 of each parameters in Eq. (1). This procedure corrected relatively long-term (several
237 minutes to 30 minutes) variations in CO₂ or water vapor concentrations that were
238 independent of eddy fluctuations and were caused by the temporal and spatial
239 heterogeneity of the atmospheric mass. HP filtering is often applied to measurements in
240 a complex environment; however, incorrect application of HP filtering results in
241 underestimation of fluxes (Lee et al., 2004). HP filtering was applied by using an
242 exponential moving average as follows:

243

$$244 \quad \begin{aligned} x_i' &= A(x_i - x_{i-1}) + Ax_{i-1}' \\ A &= \exp\left(-\frac{1}{\tau \cdot f}\right) \end{aligned} \quad (2)$$

245

246 where x_i and x_i' are an instantaneous datum and deviation from the mean at time i ,
247 respectively. The parameter τ is the time constant of the exponential moving average,
248 which was determined to be 150 s in a previous study (McMillan, 1988), and f is the
249 sampling frequency (10 Hz). HP filtering was applied to all of the measured data (i.e.,
250 3D wind velocity, air temperature, CO₂ and water vapor concentrations, and
251 atmospheric pressure).

252 **2.4 Measurement of fluxes using the BF method**

253 BF flux measurements were performed during the daytime on 29 May, 15 July, and 21
254 September 2014 for comparison with the EC measurements. In the BF method, flux is
255 calculated as follows:

256



$$257 \quad F = kS(fCO_{2water} - fCO_{2air}), \quad (3)$$

258

259 where k is the transfer velocity, which was calculated using one or another of the
260 following three empirical equations. The first was the equation of Wanninkhof (1992)
261 (W92). Because this equation was constructed using tracer methodology under oceanic
262 conditions, it might be inappropriate to our measurements because of the differences in
263 fetch and water depth. However, this equation is very commonly used in a variety of
264 oceanographic CO₂ flux studies, including coastal measurements. We therefore used it
265 for comparison. The second equation was that of Borges et al. (2004) (B04), which has
266 been applied in an estuarine study that involved use of the floating chamber method.
267 This equation makes use of the current velocity in addition to the wind speed. The third
268 equation was that of Mørk et al., (2014) (M14), which was formulated to characterize
269 the transfer velocity in a fjord for use with the EC method. Thus, the second and third
270 equations can be used to characterize coastal gas transfer velocity. The wind speed for
271 the gas transfer velocity was measured by the EC device and normalized to a height of
272 10 m from the water surface using a logarithm law (Kondo, 2000). The current speed
273 and depth were measured by the sensors attached to the platform (Compact-TC and
274 Compact-EM, Alec, Japan). The final CO₂ flux was calculated as the weighted average
275 of the fluxes measured within the EC footprint (Schuepp et al., 1990).

276 The parameter S is the dissolution coefficient of CO₂, which was estimated from the
277 water temperature and salinity (Weiss, 1974). The parameters fCO_{2water} and fCO_{2air} are
278 the fugacity of water and atmospheric CO₂, respectively. Water temperature and salinity
279 were measured with a handheld conductivity-temperature-depth sensor (ACTD-DF, JFE
280 Advantech, Japan). The water samples used to determine fCO_{2water} were collected just



281 below the water surface (up to 20 cm below the water surface) to measure the
282 concentration of CO₂ where direct gas exchange with air occurs. The sampling was
283 performed within the EC footprint to minimize the effect of spatial heterogeneity when
284 comparing the BF and EC fluxes. The sampling points were determined from the wind
285 direction and the distance from the platform measured using a hand-held GPS unit
286 (Venture HC, Garmin, USA; see Table S1). The water *f*CO₂ was determined from the
287 total alkalinity and the dissolved inorganic carbon content of the water sample using a
288 batch-type carbonate measurement system (ATT-05, Kimoto electrics, Japan) and the
289 CO₂SYS program (Pierrot et al., 2006).

290

291 **3 Results**

292 **3.1 PP1 data**

293 During the measurement period, 4464 flux data points corresponding to 2232 hours
294 were obtained; 1971 of those data points (44 %) were excluded as erroneous data after
295 PP1 application. The mean and SD of the EC CO₂ fluxes were -1.93 and $52.4 \mu\text{mol m}^{-2}$
296 s^{-1} , respectively. Figure 3(a) shows the retained CO₂ flux data.

297 The largest positive CO₂ flux (release to atmosphere) was $156.51 \mu\text{mol m}^{-2} \text{s}^{-1}$ at
298 2:00 on 23 June (day 56). The largest negative CO₂ flux (uptake of atmospheric CO₂)
299 was $-217.93 \mu\text{mol m}^{-2} \text{s}^{-1}$ at 22:00 on 4 October (day 129). These fluxes were more
300 than three orders of magnitude larger than the magnitude of the average of the measured
301 EC fluxes. Figure 4 shows the instantaneous atmospheric CO₂ concentration, water
302 vapor concentration, and the cumulative covariance between CO₂ and vertical wind
303 speed during the times when the CO₂ fluxes were most positive or most negative.



304 Spikes and discontinuities were observed in the atmospheric CO₂ and water vapor
305 concentrations, despite the prior de-spiking process applied by the data management
306 system. The cumulative covariance indicated that the covariance at certain periods (0–5
307 min) contributed significantly to the total cumulative covariance.

308 **3.2 PP2 data**

309 The EC CO₂ flux data subjected to PP2 (RSSI, nSD, and HP filtering) are shown in Fig.
310 3(b). Of the 2493 total data points remaining after PP1, approximately 234 (9 %) were
311 excluded by RSSI filtering. Subsequent nSD filtering removed 426 additional data
312 points (17 %); approximately 73 % of the measurement data remained after this
313 filtering. The mean and SD of EC CO₂ flux after PP2 were -0.54 and $2.2 \mu\text{mol m}^{-2} \text{s}^{-1}$,
314 respectively.

315 The nSD threshold was calculate using the averages of the CO₂ and water vapor
316 concentrations for the entire measurement period after RSSI filtering (CO₂: 16.02 mmol
317 m^{-3} , vapor: $548.10 \text{ mmol m}^{-3}$). The nSD threshold value obtained from the average
318 values was 0.050. Other thresholds were 0.48 for skewness and 3.1 for the absolute
319 value of kurtosis. Among these comparisons, the nSD was determined to provide the
320 best threshold for the measurement data because the number of data remaining after
321 filtering was the largest for the nSD (nSD: 1661 [74 %], skewness: 422 [19 %], and the
322 absolute value of kurtosis: 445 [20 %]). The nSD filtering resulted in the exclusion of
323 137 (6 %) of the CO₂ data points and 569 (25 %) of the water vapor data points. Figure
324 5 shows the three parameters for the EC CO₂ flux data.

325 HP filtering decreased the absolute values of some CO₂ fluxes (Fig. 6). Most of the
326 data were not changed very much by HP filtering; however, HP filtering decreased the



327 absolute values of some fluxes. Figure 7 shows an example of the results of filtering
328 atmospheric CO₂ concentrations as well as the cumulative covariance of the
329 atmospheric CO₂ concentrations and vertical wind speed (measured at 8:00 on August
330 21 [day 84]). These data were not excluded by the RSSI and nSD filtering (RSSI =
331 100 %, nSD of CO₂ = 9.84×10^{-3} , nSD of H₂O = 2.07×10^{-2}). The concentration of
332 atmospheric CO₂ showed a trend over the 30-min time interval, the indication being that
333 the block average could not extract appropriate eddy movements from the time-series
334 data. The cumulative covariance of the block averaging (BA) was unusually large; after
335 HP filtering, the values were more reasonable because filtering successfully excluded
336 deviations caused by the variation for about 10-min of atmospheric CO₂ concentrations.

337 3.3 BF data

338 The measured BF fluxes showed spatial and seasonal variations (see Table S1). The
339 means and SDs obtained by using the three gas transfer velocity equations to analyze
340 the weighted averages of the BF CO₂ flux data were $0.44 \pm 0.33 \mu\text{mol m}^{-2} \text{s}^{-1}$ ($n = 15$)
341 on 29 May (day 1), $2.10 \pm 1.58 \mu\text{mol m}^{-2} \text{s}^{-1}$ ($n = 14$) on 15 July (day 48), and $-0.11 \pm$
342 $0.13 \mu\text{mol m}^{-2} \text{s}^{-1}$ ($n = 9$) on 21 September (day 115). The differences in the number of
343 data reflect problems during the measurements, such as inadequate water depth and
344 problems with the water-depth sensor. Except for the measurements on 15 July, the
345 difference between the results obtained with the three gas transfer velocity equations
346 were not significant.

347 4 Discussion

348 Our new filtering method, PP2, successfully excluded erroneous outliers. The SD was



349 decreased by a factor of 24 ($52.4 \mu\text{mol m}^{-2} \text{s}^{-1}$ in PP1 to $2.2 \mu\text{mol m}^{-2} \text{s}^{-1}$ in PP2). The
350 atmospheric CO_2 uptake rate calculated via PP1 measurements ($-1.93 \mu\text{mol m}^{-2} \text{s}^{-1}$)
351 was reduced in magnitude by 72 % after PP2 to $-0.54 \mu\text{mol m}^{-2} \text{s}^{-1}$. This uptake rate is
352 more consistent with the range of atmospheric CO_2 uptake rates reported in previous
353 studies (e.g. Borges et al., 2005; Chen et al., 2013; Laruelle et al., 2013). The most
354 negative CO_2 flux ever reported was $-1.08 \mu\text{mol m}^{-2} \text{s}^{-1}$ during spring in the Baltic Sea
355 (Chen et al., 2013).

356 The exclusion of erroneous outliers by PP2 also contributed to the time-series
357 analysis. The power spectra of the EC CO_2 fluxes after PP1 evidenced large, noise-like
358 fluctuations at high frequencies (Fig. 8), and thus any suggestion of peaks in the time
359 series was obscured. After PP2, however, the noise-like fluctuations were smaller, and
360 two peaks associated with semi-diurnal (~ 12.5 h) and diurnal (~ 24 h) time intervals
361 were apparent. The $f\text{CO}_2$ variations in the lagoon, which are among the parameters that
362 regulate air-water CO_2 fluxes, have been confirmed to be related to mixing of lagoon
363 water with freshwater coming from rivers and with biological processes such as
364 photosynthesis (Tokoro et al., 2014). Given that the former and latter phenomenon are
365 caused by the semi-diurnal tidal cycle and diel changes of irradiance, respectively, the
366 peaks in the power spectra are consistent with the results of Tokoro et al. (2014). This
367 consistency is a good demonstration of the utility of the PP2.

368 The EC data that were much different from the BF results were excluded by the PP2
369 (Fig. 9). The remaining EC fluxes in May and September seemed to agree well with the
370 BF fluxes. We believe that one of the reasons that the results were comparable was the
371 improvement in the accuracy of the EC fluxes with the use of PP2. Another reason was
372 our strategy of making BF measurements at multiple points within the EC footprint to



373 filter out the noise associated with the spatial heterogeneity of the BF fluxes. In July, the
374 BF fluxes estimated with the W92 seemed to be the most consistent with the EC fluxes.
375 This similarity indicates that the effect of currents in the B04 and the effect of early-
376 breaking waves in the M14 were overestimated at our measurement site.

377 However, the EC fluxes estimated with PP2 did not always agree with the BF
378 fluxes. Because the $fCO_{2\text{water}}$ is theoretically never negative, a theoretical maximum
379 negative BF flux can be calculated by arbitrarily setting fCO_2 equal to zero. The
380 maximum negative flux calculated in this way (with M14) was $-6.16 \mu\text{mol m}^{-2} \text{s}^{-1}$ at
381 15:00 on 30 May (day 2), when the maximum wind speed was recorded (11.9 m/s).
382 Forty-seven EC flux data points (3 % of all data) indicated even lower fluxes.
383 Moreover, the mean of the EC fluxes with PP2 ($-0.54 \mu\text{mol m}^{-2} \text{s}^{-1}$) was more negative
384 or almost the same as the mean of the theoretical BF fluxes ($-0.26 \mu\text{mol m}^{-2} \text{s}^{-1}$ in, $-$
385 $0.55 \mu\text{mol m}^{-2} \text{s}^{-1}$, and $-0.43 \mu\text{mol m}^{-2} \text{s}^{-1}$ estimated with W92, B04, and M14,
386 respectively). Because actual BF fluxes include less negative as well as positive fluxes,
387 these EC fluxes cannot be explained by only the BF fluxes.

388 Similar inconsistencies between air-water CO_2 fluxes calculated with the EC method
389 and other conventional methods have been reported in several studies (e.g., Tsukamoto
390 et al., 2004; Rutgersson et al., 2010). In the case of coastal measurements, water-side
391 convection due to the vertical temperature difference inside water has been postulated to
392 enhance the gas transfer velocity (Rutgersson et al., 2010). However, such an
393 enhancement was not previously observed with direct flux measurement using a floating
394 chamber at our site (Tokoro et al., 2014). Because of the very shallow water depth (less
395 than 2 m) at our site, we suspect that water-side convection was weak and was not the
396 main reason for the inconsistency of the fluxes.



397 Assuming that the EC fluxes obtained with PP2 are valid, the discrepancy between
398 the EC and BF fluxes was also postulated to reflect the limitations of the BF method
399 and/or the difference of the measurement height of the BF and EC methods; the former
400 and the latter are at the water surface and the height of the EC devices, respectively. As
401 for the BF method limitations, seagrass leaves, which reached the water surface during
402 low tide at the study site, might have affected the physical and chemical conditions at
403 the water surface (Watanabe and Kuwae, 2015). In BF theory, the CO₂ flux is caused by
404 the CO₂ concentration gradient just below the water surface. The BF method should
405 therefore not be applied when seagrass is present on the water surface. A previous study
406 that investigated the radiocarbon isotopic signatures of seagrass at the study site
407 indicated that of the total CO₂ assimilated by the seagrass, 0–40 % (mean = 17 %)
408 originated from the atmosphere and the rest from the water (Watanabe and Kuwae,
409 2015). The implication is that there is direct uptake of atmospheric CO₂ (rather than
410 uptake through the water column) by seagrass when seagrass leaves are on the water
411 surface. Atmospheric CO₂ is therefore directly taken up within a thin film of water over
412 the seagrass leaves, but this seagrass-driven CO₂ flux is not included in the BF flux
413 calculations.

414 The fact that the CO₂ flux was larger than the BF flux may have been partly caused
415 by the temperature and CO₂ gradients in the atmospheric layer between the EC
416 measurement height and the water surface. Vertical gradients in air temperature
417 frequently occur because of the large difference in temperature between the atmosphere
418 and water. Indeed, the difference between the temperature at the EC measurement
419 height and the water temperature (ΔT) ranged from +8 °C to –10 °C during the
420 measurement period (Fig. 10). The atmospheric CO₂ concentration was inversely



421 proportional to air temperature, mainly because of the ideal gas law. The inflow of air
422 masses with different temperatures and CO₂ concentrations from other regions
423 surrounding the lagoon (e.g., terrestrial environments and open sea) could also have
424 contributed to this inverse relationship. We therefore believe that the ΔT caused a
425 vertical gradient in atmospheric CO₂ concentration, the result being a vertical CO₂ flux
426 in accord with Fick's law. Indeed, the absolute values of the EC CO₂ fluxes were large
427 when ΔT was large (Fig. 11b). For example, a large negative EC flux that was more than
428 10 times the maximum negative BF flux was observed when the ΔT was negative and
429 large in magnitude. This phenomenon frequently occurs when a cold air mass overlies a
430 warm air mass (Fig. 11a). The opposite phenomenon is observed when a warm air mass
431 overlies a cold air mass (Fig. 11b). The positive temperature gradient did not result in a
432 large EC flux similar to the flux associated with the negative temperature gradient
433 because the positive temperature gradient produced stable stratification and thereby
434 prevented vertical eddy movement and CO₂ flux, a phenomenon observed in the case of
435 terrestrial EC fluxes at night (Aubinet and Feigenwinter, 2010). Nevertheless, a
436 temperature gradient-driven vertical CO₂ flux may explain the discrepancy between the
437 EC and BF fluxes.

438 In summary, we attribute the discrepancy between the EC and BF fluxes to the
439 following four factors: (1) major technical uncertainties in both methods; (2) differences
440 in measurement location and in the temporal scales of the measurements; (3) limitations
441 of the BF method related to the presence of vegetation on the water surface; and (4) the
442 gradient of atmospheric conditions between the height of the EC measurements and just
443 above the water surface (Table 1). The latter two factors may cause the EC CO₂ flux to
444 be larger than the BF flux in aquatic systems that have large amounts of vegetation or



445 are located where temperature gradients form in the atmospheric boundary layer.
446 Determination of the contribution of aquatic ecosystems to mitigating the adverse
447 effects of climate change will require consideration of all processes related to
448 atmosphere-aquatic ecosystem exchange. For this purpose, the EC CO₂ flux should be a
449 more robust indicator than the BF flux, which includes only processes related to air-
450 water exchanges. Improving the EC method is therefore essential for a re-evaluation of
451 atmosphere-aquatic ecosystem CO₂ gas exchanges and comprehensive analyses of the
452 contributions of aquatic environments to mitigating the adverse effects of climate
453 change.

454

455 *Competing interests.* The authors declare that they have no conflict of interests.

456

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463 Climate Change).

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588 **Table 1.** Summary of the differences in fluxes calculated by the EC and BF methods

	EC	BF
Major sources of uncertainty	- Cross sensitivity - Long-term variation (minutes) of CO ₂ and water vapor concentrations in air	- Wind-dependent formula - Heterogeneity of measurement site
Measurement location and scales	- CO ₂ flux at EC devices - 100 m to km	- CO ₂ flux at the air-water surface - < 100 m
Vegetation on the water surface	- effect included	- effect not included
Atmospheric gradient between the measurement height and water surface	- effect included	- effect not included

589

590



591 **Figure captions**

592 **Figure 1.** Location of the measurement site (Furen Lagoon, Hokkaido, Japan). The
593 lagoon is shallow (1–2 m). Shading indicates seagrass meadows. Eddy covariance (EC)
594 measurements were performed on the platform in 2014.

595

596 **Figure 2.** Post-processing (PP) methods for EC flux calculation. The EC CO₂ flux was
597 calculated by using the conventional PP method (PP1) in EddyPro. The corrections
598 involved in PP1 have been described in previous publications (e.g., Lee et al., 2004).
599 Detrending was performed by using block averaging (BA). Our new PP method (PP2)
600 included three data-filtering steps based on the received signal strength indication of the
601 CO₂ sensor and the standard deviation of the CO₂ and water vapor concentrations
602 divided by the corresponding average standard deviation during the measurement period
603 (nSD). Detrending in PP2 was performed by using high-pass filtering (Massman, 2000).

604

605 **Figure 3.** EC CO₂ fluxes with (a) PP1 (mean: $-1.93 \mu\text{mol m}^{-2} \text{s}^{-1}$, SD: $52.4 \mu\text{mol m}^{-2}$
606 s^{-1} , $n = 2502$) and with (b) PP2 (mean: $-0.54 \mu\text{mol m}^{-2} \text{s}^{-1}$, SD: $2.2 \mu\text{mol m}^{-2} \text{s}^{-1}$, $n =$
607 $1,833$). Several data points in panel (a) are off the scale and not shown for comparison
608 with (b), in which all data are shown. The red line in (b) shows the theoretical maximum
609 negative flux estimated from the BF method ($-2.96 \mu\text{mol m}^{-2} \text{s}^{-1}$).

610

611 **Figure 4.** Instantaneous atmospheric CO₂ concentration (a), water vapor (atmospheric
612 H₂O) concentration (b), and cumulative covariance of atmospheric CO₂ concentration
613 and vertical wind speed calculated with PP1 when the CO₂ fluxes (c) showed the largest



614 positive value ($156.5 \mu\text{mol m}^{-2} \text{s}^{-1}$; blue) and the largest negative value ($-217.9 \mu\text{mol}$
615 $\text{m}^{-2} \text{s}^{-1}$; red). Note that the covariance was not equal to the CO_2 flux because there was
616 no Webb-Pearman-Leuning correction.

617

618 **Figure 5.** Comparison between the effects of the three filtering parameters and the CO_2
619 flux after the RSSI filtering procedure. The red circles indicate the top four outliers for
620 each parameter, which were determined as erroneous (not natural) fluxes. Each
621 threshold (broken red line) was determined so as to remove these data. (a) Normalized
622 standard deviation ($\text{nSD} = \text{standard deviation over 30-min divided by the average}$
623 during the entire measurement period; threshold = 0.05; 74 % of data retained). (b)
624 Skewness (threshold = 0.48; 19 % of data retained). (c) Absolute value of kurtosis
625 (threshold = 3.1; 20 % of data retained).

626

627 **Figure 6.** Comparison of CO_2 fluxes calculated by the block averaging (BA) method
628 and EC flux with HP filtering. Most of the data lay on or close to the solid line ($y = x$)
629 and were not much changed by HP filtering. However, in the case of the data in the
630 shaded area, HP filtering decreased the absolute value of fluxes by removing the long-
631 term effect of CO_2 change (see Fig. 7).

632

633 **Figure 7.** Examples of the deviation calculations of atmospheric CO_2 concentration (a)
634 and the cumulative covariance of atmospheric CO_2 concentration and vertical wind
635 speed (b).

636

637 **Figure 8.** Power spectra of CO_2 flux with PP1 and PP2. The spectra were normalized



638 using the covariances of CO₂ and vertical wind velocity for the entire measurement
639 period after PP1 and PP2. The shaded areas indicate the frequency of the 24-h diurnal
640 cycle (left) and 12.5-h tidal cycle (right). Average CO₂ flux data during the entire
641 measurement period were used to replace missing CO₂ flux data.

642

643 **Figure 9.** Comparison of BF flux, EC flux with PP1, and EC flux with PP2 in May (a),
644 July (b) and September (c).

645

646 **Figure 10.** (a) Relationship between air temperatures and atmospheric CO₂
647 concentrations. The gray circles indicate the data averaged over 30 min. The open
648 diamonds and error bars indicate the binned averages every 1 °C and 1 SD, respectively.
649 The solid line indicates the slope estimated from the change in air volume assuming that
650 CO₂ behaves as an ideal gas. (b) Relationship and between temperature difference (ΔT)
651 and CO₂ flux. The gray circles indicate the EC flux data calculated every 30 min. The
652 solid diamonds and error bars indicate the binned averages every 1 °C and 1 SD,
653 respectively. Note that for clarity not all data are plotted on both graphs.

654

655 **Figure 11.** Schematic diagram showing the relationships between negative temperature
656 gradient and negative EC CO₂ flux (a) and between positive temperature gradient and
657 positive EC CO₂ flux (b). The atmospheric CO₂ gradient reflects the air temperature
658 gradient and assumes that CO₂ behaves like an ideal gas. The temperature gradient-
659 driven vertical CO₂ flux may explain the discrepancy between the EC and BF fluxes.



Figure 1

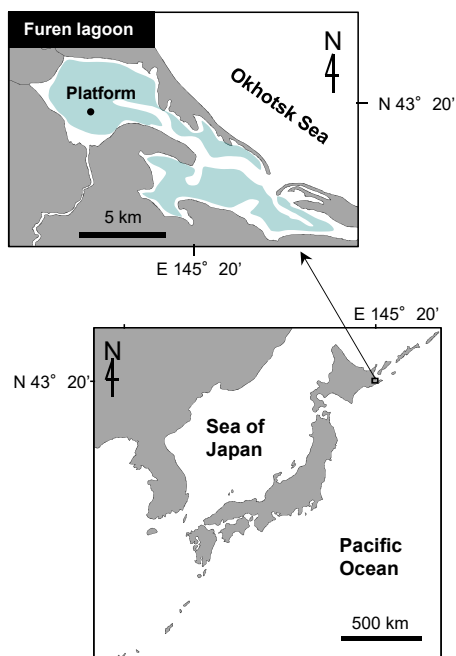




Figure 2

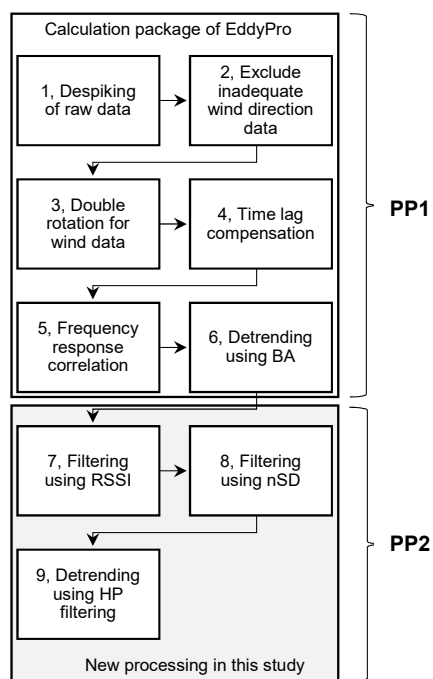




Figure 3

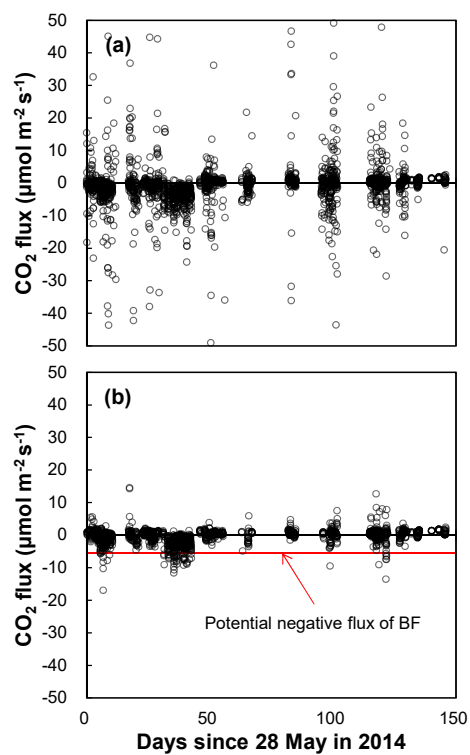




Figure 4

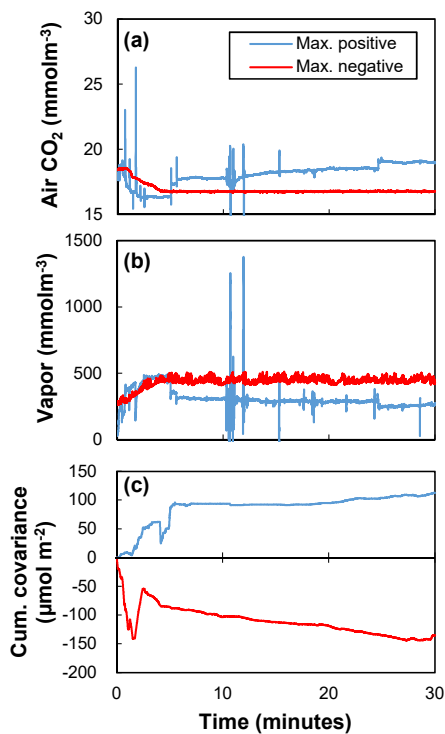




Figure 5

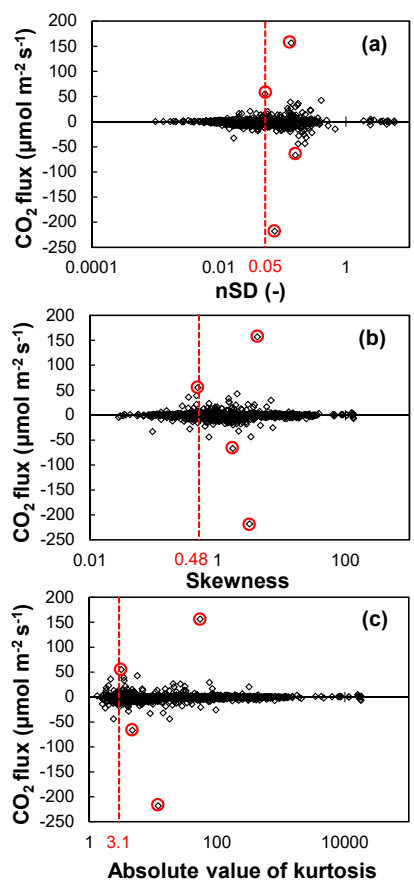




Figure 6

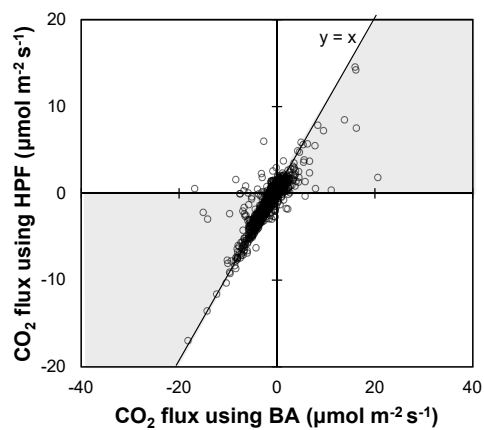




Figure 7

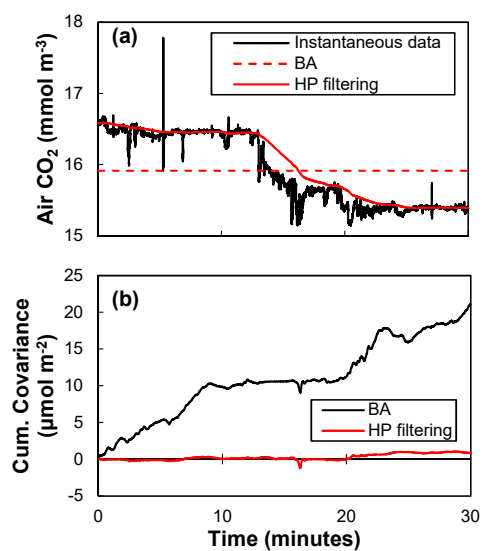




Figure 8

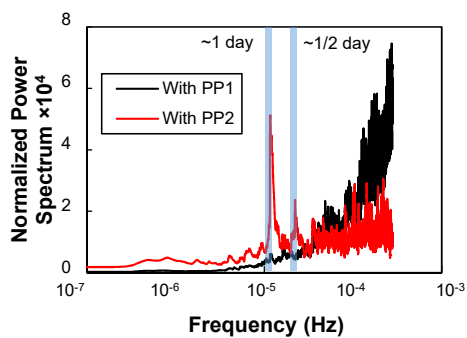




Figure 9

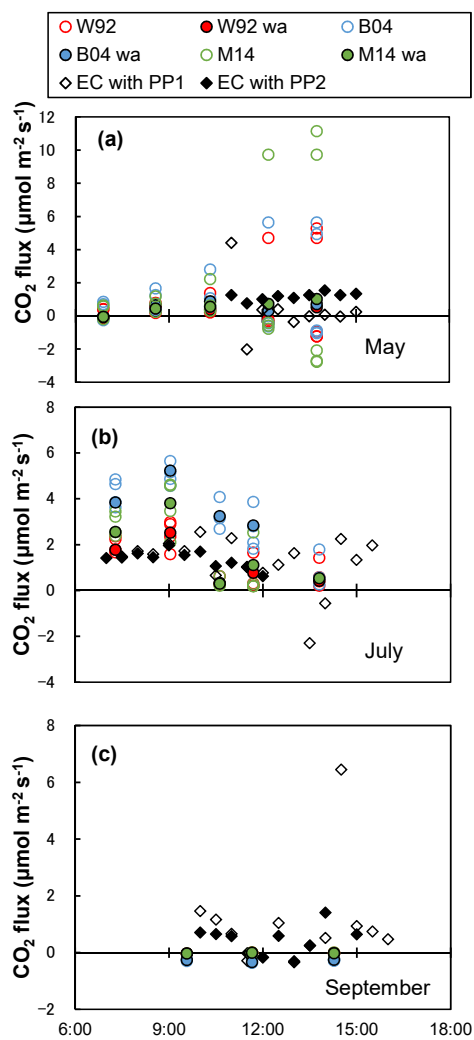




Figure 10

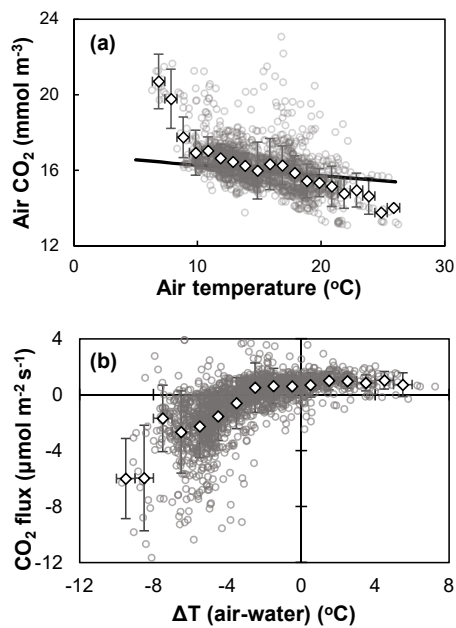




Figure 11

