

# CO<sub>2</sub> partial pressure and CO<sub>2</sub> emission along the lower Red River (Vietnam)

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**Abstract.** The Red River (Vietnam) is a **representative** example of a South-East Asian river system, strongly affected by climate and human activities. This study aims to quantify the spatial and seasonal variability of **CO<sub>2</sub> partial pressure and CO<sub>2</sub> emissions** of the lower Red River system. Water quality monitoring and **riverine pCO<sub>2</sub> measurements** were carried out for 24h at five stations distributed along the lower Red River system during the dry and the **wet** seasons. **The riverine pCO<sub>2</sub> was supersaturated relative to the atmospheric equilibrium (400 ppm), averaging about 1589 ± 43 ppm, resulting in a water–air CO<sub>2</sub> flux of 530.3 ± 16.9 mmol m<sup>-2</sup> d<sup>-1</sup> for the lower Red River. pCO<sub>2</sub> and CO<sub>2</sub> outgassing rates were characterized by significant spatial variations along this system, with the highest values measured at Hoa Binh station, located downstream of the Hoa Binh Dam, on the Da River. Seasonal pCO<sub>2</sub> and CO<sub>2</sub> outgassing rate variations were also observed, with higher values measured during the wet season at almost all sites. The higher river discharges, enhanced external inputs of organic matters from watersheds and direct inputs of CO<sub>2</sub> from soils or wetland were responsible for higher pCO<sub>2</sub> and CO<sub>2</sub> outgassing rates. The difference of pCO<sub>2</sub> between the day time and the night time was not significant, suggesting weak photosynthesis processes in the water column of the Red River due to its high sediment load.**

**Keywords:** carbon, human activities, natural condition, pCO<sub>2</sub>, Red River, Vietnam

## 35 **1 Introduction**

36 Natural hydrological processes and biogeochemistry of many rivers in the world have suffered from the  
37 influences of climate change and human activities in their drainage basins. Riverine carbon fluxes and  
38 outgassing are important parts of the carbon exchange among terrestrial, oceanic and atmospheric  
39 environment. Rivers and streams not only transfer various forms of carbon (dissolved and particulate)  
40 to oceans, but also evade a significant amount of carbon to the atmosphere (Battin et al., 2009; Richey  
41 et al., 2002). Due to CO<sub>2</sub> evasion, the flux of carbon that leaves the terrestrial biosphere through global  
42 fluvial network was suggested to be twice larger than the amount that ultimately reaches the coastal  
43 ocean (Bauer et al., 2013; Regnier et al., 2013). Raymond et al. (2013) estimated a global evasion rate  
44 of 2.1 Pg C yr<sup>-1</sup> from inland waters, and that global hot spots in stream and rivers which occupy only  
45 20 % of the global land surface represented 70 % of the emission. They emphasised that further studies  
46 are needed for identifying the mechanisms controlling CO<sub>2</sub> evasion at a global scale.

47 Riverine carbon concentrations and CO<sub>2</sub> outgassing from rivers are impacted by both natural  
48 and human factors (Liu et al., 2016; Liu et al., 2017). Recently, spatial and temporal dynamics of pCO<sub>2</sub>  
49 and CO<sub>2</sub> outgassing of Asian rivers are attracting the attention of scientists. Studies of pCO<sub>2</sub> and CO<sub>2</sub>  
50 outgassing from the large Southeast Asian rivers are crucial to quantify geochemical cycles accurately  
51 in the context of global changes because the river water discharge, suspended solids and  
52 biogeochemical cycles of these rivers have been altered dramatically over the past decades as a result  
53 of reservoir impoundment, land use, population, and climate changes (Walling and Fang, 2003; 2006;  
54 Lu, 2004). Solid sediment loads not only directly contribute to increase the organic carbon content, but  
55 also affect chemical weathering and hence carbon consumption and possible pCO<sub>2</sub> (Ran et al., 2015b).  
56 Some studies emphasized that data concerning CO<sub>2</sub> outgassing of Southeast Asian rivers is a high  
57 priority in order to improve the global evasion rate from inland waters (Raymond et al., 2013;  
58 Lauerward et al., 2015).

59 The Red River, with a basin area of 156,450 km<sup>2</sup>, is a typical East Asia river that is strongly  
60 affected by climate and human activities. Previous studies reported the hydrology and suspended  
61 sediment load associated to some elements loads (N, P, C) of the Red River (Dang et al., 2010; Lu et  
62 al., 2015; Le et al., 2015). Recently, the transfer of organic carbon of the Red River to ocean has been  
63 studied (Dang et al., 2010; Le et al., 2017). However, there is a lack of data concerning CO<sub>2</sub> outgassing  
64 and carbon budget of the lower Red River (Trinh et al., 2012, Nguyen et al., 2018).

65 Consequently, the objectives of this study were: i) to investigate spatial and temporal  
66 (seasonal and diurnal) variations of CO<sub>2</sub> partial pressure (pCO<sub>2</sub>) and CO<sub>2</sub> fluxes at the water-air  
67 surface of the lower Red River; and ii) to identify some of the factors that may control pCO<sub>2</sub> and CO<sub>2</sub>  
68 outgassing rates in this system. To our knowledge, our study introduced the first measurement and  
69 estimation of CO<sub>2</sub> evasion from the lower Red River.

## 70 **2 Methods**

### 71 **2.1 Study sites**

72 The five stations were studied along the lower Red River (Vietnam): Yen Bai station (at the outlet of  
73 the Thao River); Hoa Binh station (after Son La and Hoa Binh reservoirs, at the outlet of the Da River);  
74 Vu Quang (at the outlet of the Lo River); Hanoi and Ba Lat stations (in the main course of the Red  
75 River downstream). The three stations Yen Bai, Vu Quang and Hoa Binh are representative for water  
76 quality of the three main tributaries (Thao, Da and Lo) of the upstream Red River, whereas the Hanoi  
77 station is representative for the main course Red River after confluence of three main tributaries. Only  
78 the Ba Lat station, which is located at the Red River mouth (about 13 km from the sea) is influenced by  
79 seawater intrusion (Fig 1). A more detailed description of the river characteristics of the Thao, Da, Lo  
80 and the main branch of the Red River can be found in Le et al. (2007).

81 The climate in the Red River basin is tropical East Asia monsoon type, and is controlled by  
82 the North East monsoon in winter and South West monsoon in summer. It is, thus, characterized by  
83 two distinct seasons: rainy and dry seasons. The rainy season lasts from May to October and cumulates  
84 85 – 90 % of the total annual rainfall in the Red River catchment, whereas the dry season covers the  
85 period from November to next April. The monsoon climate weather results in a hydrologic regime  
86 characterized by large runoffs during the wet season and low runoffs during the dry season (Table 1).

87 A series of dams-reservoirs were impounded in both Chinese and Vietnamese territories of the  
88 Red River upstream part (Le et al., 2017). In the Da River, two large dams Hoa Binh and Son La were  
89 constructed in the river main course, whereas in the Lo River, two large dams Thac Ba and Tuyen  
90 Quang were constructed in its tributaries.

91 The upstream part of the Red River in the Chinese part is dominated by mountain areas, which  
92 are tectonically active and unstable, and this, combined with intense rainfall, causes high erosion rates  
93 (Fullen et al., 1998), whereas in the Vietnamese part, soils are mostly (70 %) grey and alluvial soils (Le  
94 et al., 2017). The Delta is located in a very flat and low land, with an elevation ranging from 0.4 to 12  
95 m above sea level (Nguyen Ngoc Sinh et al., 1995). Previous studies showed the difference of lithology  
96 in the three upstream tributaries: Paleozoic sedimentary rocks (55.5%), Mesozoic silicic rocks (18.0%)  
97 and Mesozoic carbonated rocks (16.7%) dominate in the Thao basin, whereas Paleozoic sedimentary  
98 rocks (85.3%) and Mesozoic carbonated rocks (14.7%) cover the Da river basin, and the Lo is  
99 composed of Mesozoic silicic rocks (21.5%) and Paleozoic sedimentary rocks (72.7%). The delta area  
100 is totally covered by alluvial deposits (100%) (Moon et al., 2007; Le et al., 2007).

101 Land use was quite different in the three upstream river basins Thao, Da and Lo: industrial  
102 crops dominate (58 %) in the Lo basin, forests (70 %) in the Da basin, and paddy rice fields (66 %) in  
103 the delta area. The Thao basin is characterized by a larger diversity of land use including forest, paddy  
104 rice fields, and industrial crops (85 %) (Le et al., 2015).

105 Population density varied from the upstream to the downstream part of the Red River basin.  
106 The delta area, where the Hanoi and Ba Lat stations are located, is characterized by high population  
107 density ( $> 1,000$  inhabitants  $\text{km}^{-2}$ ). In the upstream part, where the Yen Bai station (in Thao River),  
108 Hoa Binh station (in Da River) and Vu Quang station (in Lo River) situate, population density was  
109 much lower, about 100 inhabitants  $\text{km}^{-2}$  (Le et al., 2015).

## 110 2.2 Sampling procedures and analysis

111 Sampling campaigns were conducted in September (the rainy season) and November (the dry season)  
112 2014 at the five gauging stations: Yen Bai, Hoa Binh, Vu Quang, Hanoi and Ba Lat.

113 Physico-chemical parameters were automatically recorded every minute during 24h for each  
114 sampling campaign: pH, turbidity, salinity, chlorophyll *a* by a YSI6920 multi-parameters probe (YSI,  
115 USA); temperature and dissolved oxygen (DO) by a HOBO sensor (USA). These sensors have been  
116 calibrated with suitable standard solutions before each measurement campaign: pH electrode  
117 (YSI6920) was calibrated using standard solutions (pH = 4.01 and pH = 6.88, Merck) and the pH  
118 precision and accuracy was  $\pm 0.01$ ; DO electrode was calibrated using the saturated  $\text{Na}_2\text{S}_2\text{O}_3$  solution  
119 (Japan) and the DO accuracy was 0.1.

120 In parallel of in-situ measurement, river water samples were hourly collected for analysis of  
121 other water quality variables (TSS, DOC, POC, and total alkalinity) during 24h. A known volume of  
122 well-mixed sample was filtered immediately by vacuum filtration through pre-combusted (at  $450^\circ\text{C}$  for  
123 6 h) glass fiber filters (Whatman GF/F, 47 mm diameter). The filters were then kept in a freezer ( $-20$   
124  $^\circ\text{C}$ ) until analysis of TSS and POC. For the measurement of TSS, each filter was dried for 1h at  $105^\circ\text{C}$   
125 and then weighed. Taking into account the filtered volume, the increase in weight of the filter  
126 represented the total TSS per unit volume ( $\text{mg L}^{-1}$ ).

127 POC concentrations were estimated on the same filters. Filters were then weighed before and  
128 after calcination at  $550^\circ\text{C}$  for 4 hours. The difference in weight before and after calcination was  
129 multiplied by 0.4 to provide an estimation of the POC content (Servais et al., 1995).

130 A volume of 30 ml sub-sample of filtrate was acidified with  $35\ \mu\text{l}$  85 %  $\text{H}_3\text{PO}_4$  acid and then  
131 stored at  $4^\circ\text{C}$  in amber glass bottles until measurement of the DOC concentrations using a TOC- $V_E$   
132 (Shimadzu, Japan). The samples, standards and blank measurements were measured in triplicate and  
133 the analytical error was below 3 %.

134 Total alkalinity of the hourly samples was immediately determined on non-filtered water  
135 samples (30 ml water sample) in situ by titration method with 0.01M HCl (APHA, 1995). For each  
136 sample, triplicates were titrated and the analytical error was below 3 %.

## 137 2.3 Hydrological data collection

138 Daily and hourly data of river water discharges in 2014 at the 5 hydrological stations studied were  
139 collected from the Vietnam Ministry of Natural Resources and Environment (MONRE, 2014). The  
140 daily data were collected for all days in 2014 (Figure SM1), whereas hourly data were obtained for the  
141 exact dates of field measurements at the 5 sites (Table 1). The mean annual river flows in 2014 of the  
142 Thao, Da, Lo Rivers and in the main axe of the Red River at the Hanoi and Ba Lat stations were:  $527 \pm$   
143  $515$ ;  $1369 \pm 833$ ;  $1302 \pm 517$ ;  $1867 \pm 1089$ ;  $615 \pm 293\ \text{m}^3\ \text{s}^{-1}$ , respectively. Higher values of river  
144 discharges were observed in wet season (May to October) than in dry season (January-April;  
145 November-December) at all sites (Table 1).

146 Water velocity at the 5 sites varied from  $0.3\ \text{m s}^{-1}$  at Vu Quang site in the dry season to  $1.0\ \text{m}$   
147  $\text{s}^{-1}$  at Hoa Binh and Yen Bai sites in the wet season. The mean water depth varied with the highest

148 values recorded at the Vu Quang Site in the rainy season and the lowest at Ba Lat estuary in both the  
149 rainy and the dry seasons (Table 1).

## 150 **2.4 pCO<sub>2</sub> determination**

151 pCO<sub>2</sub> in the water column was measured using an equilibrator connected to a portable infrared gas  
152 analyser (IRGA), and also calculated using T<sub>alk</sub> and pH measured in-situ.

### 153 **2.4.1 Measured pCO<sub>2</sub>**

154 An equilibrator was used to determine the pCO<sub>2</sub> in water equilibrated with the air. The equilibrator was  
155 designed, as described in Frankignoulle et al. (2001), as follow: a vertical plastic tube (height: 73 cm,  
156 diameter: 9 cm), which is filled up with about 250 glass marbles (diameter = 1.5 cm) in order to  
157 increase the surface exchange between water and air. The river water (water inlet) through a submerged  
158 pump at 20 cm below the river surface water comes into the equilibrator from the top of the tube. The  
159 water inlet can be regulated by a flow controller installed under the tygon tubing, which joins the water  
160 inlet with the pump. A closed air circuit ensures circulation through the equilibrator (from the bottom  
161 to the top), a water trap, a particle filter, a flow regulator and a portable infrared gas analyser (IRGA)  
162 (Licor 820, Licor<sup>®</sup>, USA), which was calibrated before each sampling campaign using a series of  
163 standards concentrations of 0, 551 and 2756 ppm CO<sub>2</sub> (Air Liquide<sup>®</sup>). The IRGA was connected to a  
164 computer interface, which allows recording the pCO<sub>2</sub> every second. Values were recorded during 24 h  
165 continuously. The accuracy is <3% of reading.

### 166 **2.4.2 Calculated pCO<sub>2</sub>**

167 DIC content may be calculated from the sum of total dissolved inorganic carbon in water including  
168 HCO<sub>3</sub><sup>-</sup>, CO<sub>3</sub><sup>2-</sup>, H<sub>2</sub>CO<sub>3</sub> and CO<sub>2</sub>, or can be calculated from a combination of any two of the following  
169 measured parameters total alkalinity, pH, or partial pressure of CO<sub>2</sub> (pCO<sub>2</sub>) (Park, 1969). In this study,  
170 DIC contents were calculated from the sum of including HCO<sub>3</sub><sup>-</sup>, CO<sub>3</sub><sup>2-</sup>, H<sub>2</sub>CO<sub>3</sub> and CO<sub>2</sub> contents,  
171 which were given by the calculation from the CO<sub>2</sub>-SYS EXCEL Macro Software (version 2.0) based  
172 on the total alkalinity contents and pH values measured in-situ as described above (Sect. 2.3).

173

## 174 **2.5 CO<sub>2</sub> fluxes determination**

175 The water-air CO<sub>2</sub> fluxes from the equilibrator measurement at each site were calculated by  
176 the formula proposed by Raymond and Cole (2011) as followings:

$$177 F_{\text{Equi}} = k_{600} * \alpha * (p\text{CO}_2 \text{ water} - p\text{CO}_2 \text{ air}) \quad \text{Eq. (1)}$$

178 Where F is the CO<sub>2</sub> flux from water (μmol m<sup>-2</sup> s<sup>-1</sup>) and converted in mmol m<sup>-2</sup> d<sup>-1</sup>;

179 k<sub>600</sub>, was gas transfer velocity of CO<sub>2</sub> or piston velocity (cm h<sup>-1</sup>). Some studies indicate that k<sub>600</sub> values  
180 are closely related to flow velocity and channel gradient for rivers (Alin et al., 2011). In this study, k<sub>600</sub>  
181 was calculated using the equation from Raymond et al. (2012) based on stream velocity (V, in m s<sup>-1</sup>),  
182 slope (S, unitless), depth (D, in meters) and discharge (Q, in m<sup>3</sup> s<sup>-1</sup>), as follow:

183 
$$k_{600} = 4725 \pm 445 \times (V \times S)^{0.86 \pm 0.016} \times Q^{-0.14 \pm 0.012} \times D^{0.66 \pm 0.029} \text{ Eq. (2)}$$

184  $\alpha$  is the solubility coefficient of CO<sub>2</sub> for given temperature and salinity (Weiss, 1974) (mol L<sup>-1</sup> atm<sup>-1</sup>).  
185 In this case,  $\alpha = 0.034 \text{ mol L}^{-1} \text{ atm}^{-1}$ . In this study, salinity variations were low, except for the Ba Lat  
186 station. Temperature did not change a lot. We checked the influence of different  $\alpha$  values in the dry ( $\alpha$   
187 =  $3.941 \cdot 10^{-2} \text{ mol L}^{-1} \text{ atm}^{-1}$  at 24 °C) and the wet season ( $\alpha = 3.138 \cdot 10^{-2} \text{ mol L}^{-1} \text{ atm}^{-1}$  at 27 °C) at the 5  
188 sites and compared with the constant  $\alpha$  value of  $0.034 \text{ mol L}^{-1} \text{ atm}^{-1}$ .

189 Both  $p\text{CO}_2$  in the water determined from Equilibrator measurement (in ppm) and from  
190 CO<sub>2</sub>\_SYS calculation (in atm) were converted in  $\mu\text{mol}$  when calculating the flux of CO<sub>2</sub> outgassing.

191

## 192 **2.5. Statistical analysis**

193 To detect the correlation between environmental variables and  $p\text{CO}_2$ , statistical software R version  
194 3.3.2 (R Core Team, 2016) was applied to calculate the Pearson correlation coefficients. Some  
195 environmental variables were evaluated by “cor” to compare the correlation and selected representative  
196 variables. PCA analysis was then used for identifying representative variables that could relate to the  
197 dynamic of  $p\text{CO}_2$ .

198 Student t-test was used to test the difference of variables values between the two different  
199 times (the wet and the dry) and (the day and the night), whereas ANOVA was used to test the  
200 difference of variables within stations on the measured mean variables. Probabilities (p) were  
201 determined and a p value of < 0.05 was considered to be significant.

202

## 203 **3. Results**

### 204 **3.1. Physical and chemical variables of the lower Red River**

205 Water temperature varied from 23.3 to 29.4 °C, and the mean value in rainy period (27.4 °C) was higher  
206 than the one in dry period (24.5 °C) at almost all stations, except at the Hoa Binh site where the water  
207 temperatures did not show seasonal variations, remaining around 26.3 - 26.5 °C. Among the five  
208 hydrological stations, the higher water temperatures were recorded at the Hanoi and Ba Lat stations,  
209 ranging from 28 to 29 °C in the wet period, whereas they were close to 23 °C during the dry period.  
210 Temperatures at the Yen Bai and Vu Quang stations were approximately 26 °C in the wet period and 24  
211 °C in the dry period (Table 2). No clear difference for water temperature during the day and at the night  
212 time was observed at the 5 sites in both the rainy and the dry season (p < 0.05).

213 pH values were slightly different between the two periods, being higher in the dry season than  
214 in the wet season at all the sites (Table 2) (p < 0.05). pH ranged from 7.7 to 8.2 with an average of 8.1  
215 for all the sites. The lowest pH values were measured at the Hoa Binh in both periods (< 8), whereas  
216 they ranged between 8.0 and 8.4 at the other sites.

217 The percentage of dissolved oxygen (% DO) varied from 50.5 % to 70.7 % with an average  
218 value of 64.3 % (Table 2). The mean values were the highest for the Yen Bai station (70.1 %) in the  
219 wet period, and 69.5 % for the Ba Lat station in the dry period. The lowest values were observed at the  
220 Hoa Binh station in both periods (55.0 % in the wet period, 51.4% in the dry period) (Table 2). DO  
221 showed the seasonal and spatial variations but no clear day-night difference was observed ( $p < 0.05$ ).

222 Salinity values at the four upstream sites were under the detection limit both in the rainy and  
223 dry seasons, but in the estuary downstream river at the Ba Lat station, values up to 8.75 were measured  
224 during the dry season (Table 2). Conductivity followed the same trend as salinity, and was close to  
225  $0.2 \pm 0.0 \text{ mS cm}^{-1}$  for the 4 upstream sites, and reached up to  $6.6 \pm 3.4 \text{ mS cm}^{-1}$  at Ba Lat (Table 2).

226 Total alkalinity ranged from  $84.3 \pm 1.9$  to  $152.9 \pm 6.6 \text{ mg L}^{-1}$ , with higher values measured in  
227 the dry season than in the rainy season ( $p < 0.05$ ), except at Vu Quang station. The difference of total  
228 alkalinity was spatially recorded but no clear variation appeared between values in day and night times  
229 at 5 sites ( $p < 0.05$ ).

230 Chlorophyll *a* was quite low during the two sampling campaigns, ranging from 0.23 to 2.77  
231  $\mu\text{g L}^{-1}$ , with an average of  $1.61 \mu\text{g L}^{-1}$ . Higher values in the rainy season than in the dry season were  
232 observed but no clear day-night difference was observed at almost sites ( $p < 0.05$ ). From Yen Bai to Ba  
233 Lat, Chl-*a* concentrations in the main axe (at Yen Bai and Hanoi stations) were higher than in the two  
234 tributaries Da and Lo (Table 2), even under the higher values of turbidity.

### 235 3.2. Carbon concentrations of the lower Red River

236 During the two sampling campaigns, DOC concentrations ranged from 0.5 to  $4.6 \text{ mgC L}^{-1}$ , averaging  
237  $1.5 \text{ mgC L}^{-1}$ . Higher values were observed during the rainy season ( $2.0 \text{ mgC L}^{-1}$  vs.  $1.5 \text{ mgC L}^{-1}$  during  
238 the dry season), and the highest value was recorded at Hanoi station (Table 2). POC concentrations  
239 varied from 0.4 to  $4.6 \text{ mgC L}^{-1}$ . Among the 5 sites, POC concentrations in the main reach of the Red  
240 River (Yen Bai, Hanoi and Ba Lat sites) were higher than in the two tributaries Da and Lo, where dams  
241 were constructed. Spatial and seasonal variations of DOC and POC were observed but no clear  
242 difference in day – night time was found ( $p < 0.05$ ) (Table 2, table SM1).

243 DIC concentrations at the five sites fluctuated between 16.7 and  $32.9 \text{ mgC L}^{-1}$ , averaging  $23.8$   
244  $\text{mgC L}^{-1}$ . Lower values were measured in the rainy season ( $22.3 \text{ mgC L}^{-1}$ ) than in the dry season ( $25.3$   
245  $\text{mgC L}^{-1}$ ) and the difference of DIC was noted for the 5 sites ( $p < 0.05$ ) (Table 2).

246

### 247 3.3. Comparisons of the $p\text{CO}_2$ results obtained by the two methods

248  $p\text{CO}_2$  along the lower Red River (Vietnam) in the dry and the wet seasons were determined by two  
249 methods: i) direct measurements using an equilibrator connected to an IRGA, ii) calculated from pH  
250 and alkalinity using the  $\text{CO}_2\text{-SYS}^{\text{®}}$  software. The direct  $p\text{CO}_2$  measurements gave slightly higher  
251 values than the calculated ones (Table 2), but the values of two methods were similar and presented the

252 same trend of spatial and seasonal variations ( $R^2 = 0.77$ , Fig. 2; Table 2). Lower values of the  
253 calculated  $p\text{CO}_2$  in this study may be caused by the analytical errors in pH or under-estimation of total  
254 alkalinity. Similarly, the  $\text{CO}_2$  outgassing rates which were calculated from measured  $p\text{CO}_2$  from  
255 equilibrator were higher than the ones derived from the calculated  $p\text{CO}_2$  from  $\text{CO}_2$ -SYS, however they  
256 are in the same orders and have similar variation trends (Table 3, Fig 2).

257 Below, we use the results of  $p\text{CO}_2$  (and  $f\text{CO}_2$ ) from direct measurements to discuss the spatial and  
258 seasonal variations of  $p\text{CO}_2$  (and  $f\text{CO}_2$ ) of the lower Red River.

### 259 **3.4. Relations between $p\text{CO}_2$ and water chemistry variables**

260 The riverine water  $p\text{CO}_2$  was supersaturated with  $\text{CO}_2$  in contrast to the atmospheric equilibrium (400  
261 ppm), averaging  $1,589 \pm 43$  ppm for all sites observed. In general, the results did not show a clear  
262 variation in  $p\text{CO}_2$  between the day and night, except higher values at the night time at the Ba Lat site  
263 and higher value in daytime at Vu Quang in dry season ( $p < 0.05$ ) (Table 2). This leads to the same  
264 trends of  $\text{CO}_2$  outgassing rates: no clear difference between daytime ( $548.9 \pm 17.9 \text{ mmol m}^{-2} \text{ day}^{-1}$ ) and  
265 night time ( $551.8 \pm 15.9 \text{ mmol m}^{-2} \text{ day}^{-1}$ ) ( $p < 0.05$ ) (Table 3, Fig. 3).

266  $p\text{CO}_2$  values fluctuated from 694 ppm (at Yen Bai) in the dry season to 3,887 ppm (at Hoa  
267 Binh) in the wet season in 2014. The mean values were the highest for the Hoa Binh station in both  
268 seasons whereas the lowest one was observed at the Yen Bai site. Spatial variations of both  $p\text{CO}_2$  and  
269  $f\text{CO}_2$  flux for all 5 sites were observed ( $p < 0.05$ ). Higher values of  $p\text{CO}_2$  in the wet season than in the  
270 dry season were observed at almost all the sites ( $p < 0.05$ ) (Table 2).

271  $\text{CO}_2$  outgassing rates of the 5 stations of the lower Red River showed seasonal and spatial  
272 variations ( $p < 0.05$ ). The highest value was recorded at Hoa Binh site in both the rainy and the dry  
273 season, averaging  $1447.5 \pm 27.4 \text{ mmol m}^{-2} \text{ d}^{-1}$  and the lowest value was observed at Ba Lat site,  
274 averaging  $54.6 \pm 6.5 \text{ mmol m}^{-2} \text{ d}^{-1}$ .  $\text{CO}_2$  outgassing rates were higher in the wet season than in the dry  
275 season at all sites (Table 3, Fig. 3).

276 PCA and Pearson correlation coefficient were performed to analyze the relationships between  
277 nine environmental variables and  $p\text{CO}_2$  at the five sampling stations of the lower Red River in the wet  
278 season (September 2014) and the dry season (November 2014). The PCA of the seasonal data for five  
279 sampling stations presented a clear separation between two periods (Fig. 4a). The rainy period was  
280 characterized by the factors of flow, temperature,  $p\text{CO}_2$ , POC, DOC, turbidity and Chl a. The dry  
281 season is mainly governed by the factors of DIC, salinity, and conductivity. The spatial differences  
282 appeared for almost variables in both wet and dry seasons (Fig. 4b). Among the five stations, the Hoa  
283 Binh station is characterized by the  $p\text{CO}_2$  and  $\text{CO}_2$  flux, whereas the Ba Lat station had strong  
284 influences by salinity and conductivity. The other stations showed the combination of different factors.  
285 The Hoa Binh station has highest flows that correlate with the  $p\text{CO}_2$  and  $\text{CO}_2$  flux.

286 The Pearson correlation coefficient showed a strong negative correlation between  $p\text{CO}_2$  and  
287 pH and oxygen saturation (%) ( $r \sim -0.8$  for both). A low positive correlation between  $p\text{CO}_2$  and DIC



288 and DOC was found ( $r \sim 0.15$ ) (Table 4). However, the  $p\text{CO}_2$  is positively correlated with the flow of  
289 the river ( $r = 0.3$ ). Consequently, we included that the  $p\text{CO}_2$  are the results of a combination of multiple  
290 parameters, rather than a single one, such as the flow of river, season (including precipitation and  
291 temperature), dam construction, population density, geomorphological characteristics of the catchment.

## 292 **4. Discussion**

293

### 294 **4.1 Temporal variations of $p\text{CO}_2$ and $\text{CO}_2$ fluxes of the lower Red River**

295 Different explanations were given for the day-night variation of  $p\text{CO}_2$  and  $\text{CO}_2$  flux for aquatic  
296 ecosystems in the world. Previous studies indicated that water temperature could alter the riverine  
297  $p\text{CO}_2$  value because  $\text{CO}_2$  solubility decreases with the temperature increase during the day (Parkin and  
298 Kaspar, 2003). This effect was observed for some rivers in the world (Guasch et al., 1998; Dornblaser  
299 and Striegl, 2013; Peter et al., 2014). Other study revealed that photosynthesis of phytoplankton may  
300 have a strong influence on circadian variation of  $p\text{CO}_2$  or  $\text{CO}_2$  outgassing, since this process consumes  
301  $\text{CO}_2$  during the day (Linn and Doran, 1984).

302         Concerning the lower Red River, water temperature did not show clear variation between the  
303 day and the night. In addition, low Chl-a concentrations were measured, from 0.5 to 3.1  $\mu\text{g L}^{-1}$ ,  
304 probably as a result of the high turbidity limiting light penetration in the water column. Thus,  
305 phytoplankton activity had a low influence on C dynamic in the lower Red River system.  
306 Consequently, there are no clear variations of  $p\text{CO}_2$  and  $\text{CO}_2$  fluxes between the day and the night time  
307 at the different stations along the lower Red River.

308         Regarding seasonal variations, some authors suggested that higher water temperatures in the  
309 wet season in tropical regions were responsible for increased  $p\text{CO}_2$  and higher  $\text{CO}_2$  emissions to the  
310 atmosphere (Hope et al., 2004; Li et al., 2012). Dessert et al., (2003) suggested that higher temperature  
311 should also induce higher weathering rates, leading to higher DIC export. Increase in temperature  
312 decreases  $\text{CO}_2$  solubility but increase OM decomposition processes, which produce  $\text{CO}_2$ . These  
313 processes may partly explain the higher  $p\text{CO}_2$  of the lower Red River during the hot and rainy season.  
314 However, direct relationship between temperature and  $p\text{CO}_2$  was not evidenced during the rainy  
315 season, probably because riverine inputs were the dominant factor driving  $p\text{CO}_2$ . Conversely during the  
316 dry season,  $p\text{CO}_2$  clearly increased with temperature, suggesting that metabolic rate controlled  $p\text{CO}_2$   
317 when adjacent soils inputs are limited (Fig. 5).

318         Another important factor that impacted  $p\text{CO}_2$  seasonal variations in the lower Red River was  
319 the river discharge. Indeed, during the monsoon season, the Red River discharges were about 2 to 3  
320 times higher at all the sites ( $p < 0.05$ ) (Table 1). Higher  $p\text{CO}_2$  and  $\text{CO}_2$  flux values in wet season were  
321 observed at almost all sites ( $p < 0.05$ ).  $\text{CO}_2$  flux varied from  $54.6 \pm 6.5 \text{ mmol m}^{-2} \text{ d}^{-1}$  (at Ba Lat) in the  
322 dry season to  $1447.5 \pm 27.4 \text{ mmol m}^{-2} \text{ d}^{-1}$  (at Hoa Binh) in the wet season. The higher  $p\text{CO}_2$  and  $\text{CO}_2$   
323 flux values observed during the wet season may reflect the influence of soil organic matter inputs to the  
324 riverine water column, evidenced by the higher values of DOC and POC in the rainy seasons measured  
325 in our study ( $p < 0.05$ ). In tropical regions, the wet season usually experienced higher  $p\text{CO}_2$  than the dry

326 season because the intense rainfall induced higher OM inputs into the river (Richey et al., 2002) or in  
327 addition inputs of CO<sub>2</sub> from wetlands. This process was observed in some subtropical rivers: the  
328 Longchuan River (Li et al., 2012) and the Xijiang River (Yao et al., 2007), with pCO<sub>2</sub> values  
329 increasing significantly when baseflow and interflow increased, and flushed significant amount of  
330 carbon into the streams. Other example which could be mentioned is the case of the Godavari River in  
331 Indonesia, where eextreme high value of pCO<sub>2</sub>, up to ~30,000 ppm, were measured probably due to  
332 significant organic carbon decomposition during peak discharge period. This was in contrast with the  
333 very low values measured (<500 ppm) during the dry season for this river (Sharma et al., 2011).

334 To conclude, no clear day-night time variation of both pCO<sub>2</sub> and fCO<sub>2</sub> at 5 sites of the lower  
335 Red River in 2014 was found but our results showed the clear seasonal variation of both pCO<sub>2</sub> and  
336 fCO<sub>2</sub>. The spatial variation of pCO<sub>2</sub> and fCO<sub>2</sub> at 5 sites of the lower Red River under the natural and  
337 anthropogenic factors will be discussed below.

338

## 339 **4.2 Spatial variations of pCO<sub>2</sub> and fCO<sub>2</sub> outgassing**

### 340 **4.2.1 Influence of geomorphological characteristics**

341 The upstream of the Red River is located in mountainous areas, where chemical and  
342 mechanical erosion are among the world highest (500 mm per 1,000 years) (Meybeck et al., 1989),  
343 which may participate in the elevated pCO<sub>2</sub> values measured. The geologic substratum of the upstream  
344 Red River is dominated by consolidated paleozoic sedimentary rocks, with variable contributions of  
345 mesozoic silicic or carbonate rocks. During rainfall events, the erosion of these rocks may increase  
346 pCO<sub>2</sub> in the tributary river waters of the Thao, Da and Lo, which were supersaturated with CO<sub>2</sub> in air  
347 from about 2 to 14 times. Our results showed that the pCO<sub>2</sub> mean value of the lower Red River (1589 ±  
348 43 ppm) was close to the ones of some Asian rivers such as the downstream Mekong River: 703 – 1597  
349 ppm (Alin et al., 2011); the Longchuan River: 2101 – 2601 ppm (Li et al., 2012); the Changjiang  
350 River: 1,297 ± 901 ppm (Wang et al., 2007); the Yellow River: 2,811 ± 1,986 ppm (Ran et al., 2015a)  
351 (Table 5). However, very high pCO<sub>2</sub> value, up to 11,000 ppm, was also observed for some other Asian  
352 rivers, like the Xijiang River (Yao et al., 2007).

353 CO<sub>2</sub> emissions from the lower Red River varied in a high range, from 54.6 ± 6.5 to 1447.5 ±  
354 27.4 mmol m<sup>-2</sup> d<sup>-1</sup>, averaging 550.3 ± 16.9 mmol m<sup>-2</sup> d<sup>-1</sup>. They were close to the values of some large  
355 Asian rivers as the Yellow River (856±409 mmol m<sup>-2</sup> d<sup>-1</sup>) (Ran et al., 2015b) and the Xijiang river (357  
356 mmol m<sup>-2</sup> d<sup>-1</sup>) (Yao et al., 2007) or some rivers in South America reported by Rasera et al. (2013) such  
357 as the Negro, the Solimoes, the Caxiuana rivers (855 ± 294, 518 ± 17, 778 ± 17 mmol m<sup>-2</sup> d<sup>-1</sup>,  
358 respectively) (Table 5). Thus, the high alkalinity, pCO<sub>2</sub> and fCO<sub>2</sub> in the Red River in this study can be  
359 partly explained by wide distribution of carbonate-silicate rocks in the upper Red River drainage area,  
360 especially during high water discharge as observed for other Asian rivers.

361

### 362 **4.2.2 Influence of hydrological characteristics**

363 Spatially, pCO<sub>2</sub> differences between the three upstream tributaries and the main downstream axe of the  
364 Red River are suggested to be partially related to different hydrological characteristics and  
365 management of the three sub-basins and delta area, as observed in other systems (Yao et al., 2007; Li et

366 al., 2012). Our results showed that within the 3 upstream sites studied, the highest  $p\text{CO}_2$  values were  
367 always measured in the Da River at Hoa Binh site, where river discharges were the highest ( $2,189 \pm 39$   
368  $\text{m}^3 \text{s}^{-1}$  in the wet season and  $868 \pm 319 \text{m}^3 \text{s}^{-1}$  in the dry season) ( $p < 0.05$ ), whereas the lowest  $p\text{CO}_2$   
369 were measured at the Yen Bai station of the Thao River, where river discharges were the lowest ( $840 \pm$   
370  $68 \text{m}^3 \text{s}^{-1}$  in the wet season and  $260 \pm 18 \text{m}^3 \text{s}^{-1}$  in the dry season) (Table 1 and Table 2). Figure 4  
371 showed the clear difference of  $p\text{CO}_2$ ,  $\text{CO}_2$  flux and river discharges in the rainy and the dry seasons for  
372 the lower Red River.

373         Regarding the Ba Lat site, which is situated in the Red River estuary and thus in a very low  
374 and flat land,  $p\text{CO}_2$  values were lower than in Hanoi. It is interesting to observe that the river water  
375 discharge at Hanoi site ( $3,296 \pm 86$  and  $1,915 \pm 149 \text{m}^3 \text{s}^{-1}$ ) was about 3 times higher than the one at Ba  
376 Lat ( $1,269 \pm 93$  and  $453 \pm 31 \text{m}^3 \text{s}^{-1}$ ) in both wet and dry seasons respectively (Table 1), whereas higher  
377  $p\text{CO}_2$  values were measured during the dry season in Hanoi than in Ba Lat (1,150 and 800 ppm,  
378 respectively), but during the rainy season the values were close, i.e. around 1,450 ppm. We think that  
379 dilution by seawater may lead to a reduction of riverine surface water  $p\text{CO}_2$ , especially in the dry  
380 season when the river flow was lower ( $p < 0.05$ ). The higher salinity values measured at Ba Lat site in  
381 the dry season (3.6) than in the wet season (0.2) may confirm our suggestion that tidal action  
382 influenced at Ba Lat site in the Red River estuary. This result is consistent with previous observations  
383 in the Changjiang River estuary (Chen et al., 2008; Bai et al., 2015).

384

#### 385 *4.2.3 Influence of land-use on $p\text{CO}_2$ and $\text{CO}_2$ emissions*

386 Land cover in the river basin may play a considerable role in controlling riverine  $p\text{CO}_2$ . As known,  
387 severe erosion due to sparse vegetation cover may enhance chemical weathering by increasing the  
388 exposure surface of fresh minerals to atmosphere (Millot et al., 2002). Li and Bush (2015)  
389 demonstrated that deforestation and agricultural expansion in the Mekong River basin accelerated  
390 chemical and physical weathering rates, leading to changes in riverine carbon fluxes. In fact, very high  
391  $p\text{CO}_2$  values (up to 30,008 ppm) were observed in the Godavari estuary due to large-scale erosion and  
392 deforestation in the catchment area, which accelerated the export of organic carbon into the river,  
393 especially in wet season (Sharma et al., 2011). In contrast, recently in another study, lower  $p\text{CO}_2$  in one  
394 sub-basin (HR) than the average of the Yellow River basin was observed because of the development  
395 of an alpine meadow ecosystem reducing soil erosion (Ran et al., 2015a). For the Red River, highest  
396 riverine  $p\text{CO}_2$  and  $\text{CO}_2$  flux values were observed in the Da at the Hoa Binh site despite this sub-basin  
397 is dominated by forest as its land cover. This may suggest other factors (reservoir impoundment and  
398 geological characteristics) which may strongly control  $p\text{CO}_2$  and  $\text{CO}_2$  flux of the Da River.

399         In addition, agricultural soil in the river basin may have significant impact on riverine  $p\text{CO}_2$ .  
400 A study concerning the long-term variation of  $p\text{CO}_2$  of the Yellow River showed that high pH in  
401 irrigation water caused the increase in riverine pH, leading to further reducing  $p\text{CO}_2$  (Ran et al.,  
402 2015a). This agrees with our results for  $p\text{CO}_2$  in the Red River when considering the decrease of the  
403  $p\text{CO}_2$  values especially in dry season from upstream Delta area at Hanoi ( $1,139 \pm 29$  ppm) to the  
404 estuary at Ba Lat (mean value of  $816 \pm 69$  ppm) where rice field and irrigation channels are very dense.

405

#### 4.2.4 Influence of dams on $p\text{CO}_2$ and $\text{CO}_2$ emission

Previously, reservoirs were suggested to decrease riverine  $p\text{CO}_2$  due to increased residence times and autotrophic production (Wang et al., 2007). However, Lauerward et al., (2015) found a low negative correlation between them. Abril et al., (2005) noted that intense mineralization of organic matter (OM) originating from the reservoir was possibly a significant source for  $p\text{CO}_2$  value in downstream river. In addition, the influence of the dam on the gas transfer velocity and then  $\text{CO}_2$  outgassing flux in the river downstream of the dam was also demonstrated in the study of the Sinnamary River (Guérin et al., 2007). In the present study, in the upstream part,  $p\text{CO}_2$  ranged from 964 ppm (at Yen Bai) to 3,830 ppm (at Hoa Binh), being highest at the Hoa Binh site where the lowest pH values were measured. Higher  $k_{600}$  values (from 63 to 68  $\text{cm h}^{-1}$ ) were also observed at the Hoa Binh and Vu Quang sites. Noted that the Hoa Binh site is situated downstream a series of reservoirs, which have been constructed in both Chinese and Vietnamese parts including two large dams Hoa Binh (in 1989) and Son La (in 2010). The Vu Quang site is located in the downstream of a series of reservoirs, including two important Thac Ba (in 1970) and Tuyen Quang (in 2010). Previous studies emphasized that these dams have impacted water and sediment discharges downstream (Ha and Vu 2012; Ngo et al. 2014; Lu et al. 2015) with significant sediment deposition being observed in the reservoirs (Dang et al. 2010; Vinh et al. 2014; Lu et al. 2015). Thus, the higher  $p\text{CO}_2$  measured at these sites (average value of  $3129 \pm 32$  ppm) may reflect the increased decomposition of OM and/or the water perturbation due to dam construction, especially for the Da River. The impact of dams on downstream  $p\text{CO}_2$  may be less for the Lo and the Thao Rivers (average values of  $1395 \pm 63$  ppm and  $993 \pm 14$  ppm, respectively), where less numbers and less size (only small and medium) of dams/reservoirs were built up in their upstream parts. Thus, the high  $p\text{CO}_2$  measured at these stations may reflect the increased decomposition of OM and/or the water perturbation due to the large dam construction.

429

#### 4.2.5 Influence of population density on $p\text{CO}_2$ and $\text{CO}_2$ emission

Previous studies demonstrated very high value of  $p\text{CO}_2$  in river estuaries as a result of different human activities. For instance,  $p\text{CO}_2$  up to  $\sim 25,000$  ppm was measured in the Rhine estuary (Kempe, 1982) or up to  $\sim 15,200$  ppm in the Scheldt estuaries due to high discharge of pollutants (Borges and Frankignoulle, 2002).

Concerning the Red River, from the upstream to the downstream part of the main axe,  $p\text{CO}_2$  together with  $\text{CO}_2$  outgassing flux slightly increased from Yen Bai ( $993 \pm 14$  ppm and  $364.9 \pm 10.3$   $\text{mmol m}^{-2} \text{d}^{-1}$  respectively) to Hanoi ( $1,275 \pm 17$  ppm and  $304 \pm 7.3$   $\text{mmol m}^{-2} \text{d}^{-1}$ ), whatever the season. However, it is worth to note that the Hanoi station was located within the city itself and at this station, the river has not yet received the wastewater discharge of the whole city. Consequently, the Hanoi station in this study may not reflect the influence of whole city, with probably lower  $\text{O}_2$  and higher  $p\text{CO}_2$  levels as observed for other urban rivers in the Red River Delta (Trinh et al., 2007; 2009; 2012).

Consequently, our results revealed that  $p\text{CO}_2$  and  $\text{CO}_2$  flux along the lower Red River were spatially different which reflect the influence of both anthropogenic activities (dam, urban effluents), and natural characteristics (rainfall-river discharge, temperature and geology) in the watershed.

446

## 447 **5. Conclusions**

448 This work presented the spatial and seasonal variability of  $p\text{CO}_2$  along the lower Red River system.  
449 The riverine water was supersaturated with  $\text{CO}_2$  in contrast to the atmospheric equilibrium (400 ppm),  
450 with  $p\text{CO}_2$  values averaging about  $1589 \pm 43$  ppm, resulting thus in a water–air  $\text{CO}_2$  flux of  $550.3 \pm 16$   
451  $\text{mmol m}^{-2} \text{d}^{-1}$  from the lower Red River system. The  $p\text{CO}_2$  from the water surface of the lower Red  
452 River network was characterized by significant spatial variations, being the highest at the Hoa Binh  
453 dam downstream and in the main axis at Hanoi station. The highest value obtained at Hoa Binh site may  
454 reflect the important impact of a series of large dams (Son La, Hoa Binh) and geomorphological  
455 characteristics in the Da River, but also the high water discharge, whereas the high  $p\text{CO}_2$  value in  
456 Hanoi may partly reflect the influence of population density through the release of organic carbon into  
457 the river. The monsoon season resulted in an increased amount of OM inputs from adjacent soil led to  
458 higher  $p\text{CO}_2$  and  $\text{CO}_2$  flux values. Consequently, this study evidenced that  $p\text{CO}_2$  and  $\text{CO}_2$  flux along the  
459 lower Red River were controlled by both anthropogenic activities (dam, urban effluents), and natural  
460 characteristics (rainfall-river discharge, temperature and geology) in the watershed. Long-term  
461 variations of  $p\text{CO}_2$  and  $\text{CO}_2$  outgassing flux of the Red River may be performed in a future research  
462 effort to contribute to studies of regional and global carbon emission under natural and anthropogenic  
463 impacts.

## 464 **Author contribution**

465 Le TPQ, Marchand C and Ho TC designed the experiments. Le TPQ, Ho TC and Vu DA carried the in-  
466 situ experiments. Le TPQ, Le ND contribute to data treatment and calculations. Le TPQ and Marchand  
467 C prepared the manuscript with the contributions from all co-authors.

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697 Table 5.  $p\text{CO}_2$  (mean value and standard deviation) from some World Rivers.

**Table 1.** Average values (and standard deviation) of river water discharge at five studied sites of the Red River in 2014 (MONRE, 2014; Le, 2008)

Studied sites	Altitude (m a.s.l.)	Latitude	Mean water velocity, m/s		Mean water depth, m		Mean slope	Average daily water discharge in 2014, m <sup>3</sup> s <sup>-1</sup>	Water discharge, m <sup>3</sup> s <sup>-1</sup>			
			Wet	Dry	Wet	Dry			Wet season		Dry season	
									Mean value in wet season in 2014 (May – Oct)	On the date of measurement Sept 2014	Mean value in wet season in 2014 (May – Oct)	On the date of measurement in Nov 2014
Yen Bai	56	21°42'00.0"N 104°52'00.0"E	1.0	0.9	4.8	3.7	0.0012	527 ± 515	788 ± 459	840 ± 68	262 ± 530	260 ± 18
Hoa Binh	23	20°54'00.0"N 105°21'00.0"E	0.5	0.4	11.5	9.9	0.0015	1,369 ± 833	1,907 ± 451	2,189 ± 39	825 ± 515	868 ± 319
Vu Quang	25	21°33'00"N 105°16'00"E	0.5	0.3	12.3	9.2	0.0018	1,302 ± 517	1,618 ± 378	2,240 ± 88	982 ± 284	725 ± 11
Hanoi	5	21°02'00"N 105°52'00"E	1.0	0.9	4.7	3.9	0.0012	1,867 ± 1089	2,598 ± 780	3,296 ± 86	1,127 ± 490	1,915 ± 149
Ba Lat	0	20°18'07.6"N 106°32'25.4"E	0.4	0.3	2.1	1.4	0.0003	615 ± 293	824 ± 200	1,269 ± 93	403 ± 96	453 ± 31

**Table 2.** Average values in day and night times of the different physico-chemical variables (average value and standard deviation) at 5 sites in wet and dry seasons in 2014.

**Table 2a**

Stations	Temperature °C	pH	TALK mg L <sup>-1</sup>	Salinity	Chl-a µg L <sup>-1</sup>	Turbidity NTU	Conductivity mS cm <sup>-1</sup>	DOC mg L <sup>-1</sup>	POC mg L <sup>-1</sup>	DO %	Measured pCO <sub>2</sub> ppm	Calculated pCO <sub>2</sub> ppm
Wet season												
<i>1-Yen Bai</i>												
Day	26.4±0.1	8.2±0.1	105.1±5.2	0.1±0.0	3.1±0.1	141.6±8.6	0.2±0.0	1.5±0.2	2.1±0.4	69.9±0.2	963.9±9.4	699.1±86.1
Night	26.6±0.0	8.3±0.0	103.8±3.3	0.1±0.0	3.1±0.1	135.4±4.0	0.2±0.0	1.4±0.2	1.9±0.2	70.4±0.1	981.3±7.7	609.1±26.5
<i>2-Vu Quang</i>												
Day	26.8±0.1	8.1±0.0	148.9±6.7	0.1±0.0	1.2±0.2	51.4±7.5	0.2±0.0	1.1±0.3	1.4±0.2	63.6±1.2	1598.7±53.3	1169.2±131.2
Night	27.0±0.1	8.2±0.0	144.9±3.3	0.1±0.0	1.2±0.2	49.9±5.3	0.2±0.0	1.0±0.2	1.4±0.2	63.2±0.7	1583.0±36.6	1058.5±33.0
<i>3-Hoa Binh</i>												
Day	26.5±0.1	7.8±0.0	110.4±3.3	0.1±0.0	0.8±0.3	42.5±4.7	0.2±0.0	1.5±0.4	1.1±0.2	54.9±0.2	3827.1±60.6	2125.9±294.5
Night	26.4±0.0	7.8±0.0	107.8±5.4	0.1±0.0	1.2±0.0	41.0±0.1	0.2±0.0	1.3±0.2	1.1±0.1	55.1±0.6	3830.2±19.1	1747.0±56.8
<i>4-Ha Noi</i>												
Day	28.6±0.2	8.0±0.0	84.3±1.9	0.1±0.0	2.0±0.6	88.9±1.3	0.2±0.0	4.7±0.6	2.0±0.3	64.0±0.4	1412.6±4.0	888.7±108.3
Night	28.6±0.2	8.1±0.0	84.5±1.5	0.1±0.0	2.7±0.1	88.5±2.7	0.2±0.0	4.2±0.9	2.2±0.4	63.5±0.3	1411.0±7.0	768.6±47.3
<i>5-Ba Lat</i>												
Day	28.9±0.4	8.0±0.1	116.4±4.6	0.3±0.3	1.8±0.3	47.7±8.8	0.6±0.6	1.5±0.4	1.1±0.2	66.1±1.4	1499.2±103.2	1312.2±267.3
Night	28.8±0.3	8.1±0.0	114.9±3.5	0.1±0.1	2.5±0.1	81.3±10.0	0.3±0.2	1.7±0.5	1.5±0.2	65.1±1.6	1471.3±118.1	1156.7±77.6

Table 2b

Stations	Temperature °C	pH	TAlk mg L <sup>-1</sup>	Salinity	Chl-a µg L <sup>-1</sup>	Turbidity NTU	Conductivity mS cm <sup>-1</sup>	DOC mg L <sup>-1</sup>	POC mg L <sup>-1</sup>	DO %	Measured pCO <sub>2</sub> ppm	Calculated pCO <sub>2</sub> ppm
Dry season												
<i>1-Yen Bai</i>												
Day	24.1±0.5	8.1±0.1	113.9±7.9	0.1±0.0	1.2±0.3	49.3±7.9	0.2±0.0	1.3±0.3	1.4±0.1	69.3±0.7	995.8±17.5	896.8±320.3
Night	24.2±0.3	8.2±0.0	109.3±2.8	0.1±0.0	1.6±0.2	42.5±4.7	0.2±0.0	1.2±0.2	1.2±0.3	69.1±0.5	1030.6±21.5	655.0±19.5
<i>2-Vu Quang</i>												
Day	24.7±0.2	8.3±0.0	134.9±5.1	0.1±0.0	1.0±0.2	28.1±1.7	0.2±0.0	1.1±0.2	1.1±0.2	67.0±0.9	1235.3±76.2	756.2±81.7
Night	24.8±0.4	8.4±0.0	129.3±2.7	0.1±0.0	1.4±0.2	32.4±3.7	0.2±0.0	1.1±0.2	1.2±0.1	69.0±0.9	1163.3±86.3	604.2±35.0
<i>3-Hoa Binh</i>												
Day	26.3±0.0	7.8±0.0	122.5±6.1	0.1±0.0	0.5±0.1	16.9±0.3	0.2±0.0	0.9±0.2	0.5±0.1	51.5±0.4	2399.3±33.6	2091.7±227.2
Night	26.3±0.0	7.8±0.0	120.6±6.1	0.1±0.0	0.5±0.1	17.1±0.5	0.2±0.0	0.9±0.2	0.5±0.1	51.3±0.2	2458.9±14.0	2003.9±200.7
<i>4-Ha Noi</i>												
Day	23.8±0.1	8.2±0.0	123.5±2.4	0.1±0.0	1.7±0.2	65.2±1.8	0.2±0.0	2.7±0.7	1.5±0.3	66.8±0.4	1141.3±33.5	797.7±95.4
Night	23.8±0.1	8.3±0.0	123.8±1.5	0.1±0.0	1.6±0.1	62.6±0.7	0.2±0.0	2.0±0.7	1.3±0.1	67.1±0.3	1136.0±24.2	726.1±5.9
<i>5-Ba Lat</i>												
Day	23.7±0.1	8.3±0.0	152.9±6.6	3.9±2.4	1.8±0.2	34.1±8.3	6.6±3.4	1.4±0.2	2.0±0.4	70.0±0.5	751.4±49.3	753.6±56.5
Night	23.4±0.1	8.3±0.0	150.3±5.6	3.3±1.6	1.3±0.2	28.8±4.2	5.7±2.6	1.2±0.2	1.9±0.4	68.8±0.6	881.0±88.4	795.1±46.3

**Table 3.**  $k_{600}$  parameterization, and calculated water-air CO<sub>2</sub> fluxes for day time and night time at five hydrological stations of the Red River in dry and wet seasons in 2014.

	Wind speed m s <sup>-1</sup>	$k_{600}$ cm h <sup>-1</sup>	Water-air CO <sub>2</sub> flux, mmol m <sup>-2</sup> d <sup>-1</sup> (with $p\text{CO}_2$ measured from equilibrator)
<b>Wet season</b>			
Yen Bai			
Day	1.1±0.6	66.1±0.55	304.5±3.9
Night	0.5±0.6	66.5±0.09	315.2±4.0
Vu Quang			
Day	0.9±0.7	84.2±0.12	824.0±36.7
Night	0.4±0.6	84.1±0.02	811.8±25.1
Hoa Binh			
Day	1.3±1.1	69.1±0.27	1935.2±38.3
Night	0.2±0.5	69.0±0.03	1931.3±10.0
Hanoi			
Day	1.8±0.7	54.0±0.96	446.6±8.7
Night	1.2±0.6	53.5±0.17	441.6±3.6
Ba Lat			
Day	0.5±0.5	9.1±0.05	81.1±7.8
Night	0.2±0.3	9.1±0.02	79.4±8.6
<b>Dry season</b>			
1-Yen Bai			
Day	1.4±0.9	59.6±0.72	290.5±8.2
Night	0.5±0.8	60.1±0.54	309.3±13.3
2-Vu Quang			
Day	1.3±0.6	52.4±0.11	356.6±33.0
Night	0.7±1.3	52.4±0.15	326.3±37.0
3- Hoa Binh			
Day	1.2±0.8	58.4±2.85	938.4±22.7
Night	0.5±0.5	58.5±1.97	985.0±38.6



4-Hanoi			
Day	2.4±0.5	47.4±0.60	287.0±16.2
Night	1.4±0.5	47.3±0.57	284.2±12.7
5-Ba Lat			
Day	3.1±1.5	8.6±0.07	24.7±3.5
Night	1.3±0.8	8.5±0.03	33.5±6.1

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**Table 4. Relationship between  $p\text{CO}_2$  and  $f\text{CO}_2$  with other water quality variables**

	Temp, °C	DOC, mg.L <sup>-1</sup>	POC, mg.L <sup>-1</sup>	Talk, mg L <sup>-1</sup>	$p\text{CO}_2$ , ppm	pH	DO, mg.L <sup>-1</sup>	Sal, %	Chl_a, µgL <sup>-1</sup>	DO, %	Turb., NTU	Conduct., mScm <sup>-1</sup>	River Flow, m <sup>3</sup> .s <sup>-1</sup>	$f\text{CO}_2$ , mmol.m <sup>-2</sup> .d <sup>-1</sup>
Temp, °C	1													
DOC, mg.L <sup>-1</sup>	0.29	1.00												
POC, mg.L <sup>-1</sup>	0.00	0.36	1.00											
Talk, mg L <sup>-1</sup>	-0.47	-0.52	-0.08	1.00										
$p\text{CO}_2$ , ppm	0.31	-0.14	-0.64	-0.15	1.00									
pH	-0.50	0.01	0.60	0.28	-0.85	1.00								
DO conc, mg.L <sup>-1</sup>	-0.58	0.02	0.62	0.18	-0.83	0.91	1.00							
Sal, %	-0.37	-0.08	0.30	0.50	-0.26	0.31	0.33	1.00						
Chl_a, µgL <sup>-1</sup>	0.30	0.32	0.66	-0.46	-0.48	0.41	0.43	0.02	1.00					
DO, %	-0.33	0.12	0.71	0.05	-0.86	0.89	0.96	0.25	0.60	1.00				
Turb., NTU	0.34	0.33	0.57	-0.56	-0.33	0.25	0.27	-0.23	0.87	0.44	1.00			
Conduct., mScm <sup>-1</sup>	-0.38	-0.08	0.32	0.51	-0.27	0.32	0.34	0.99	0.02	0.25	-0.23	1.00		
River Flow, m <sup>3</sup> .s <sup>-1</sup>	0.53	0.57	0.10	-0.33	0.36	-0.34	-0.44	-0.30	0.07	-0.33	0.18	-0.31	1.00	
$f\text{CO}_2$ , mmol.m <sup>-2</sup> .d <sup>-1</sup>	0.21	-0.16	-0.56	-0.12	0.95	-0.75	-0.75	-0.28	-0.48	-0.79	-0.27	-0.29	0.38	1

**Table 5.**  $p\text{CO}_2$  (average value and standard deviation) of some World Rivers.

River or Tributary	Location	Country	Mean $p\text{CO}_2$ ppm	$F_{\text{CO}_2}$ $\text{mmol m}^{-2}$ $\text{day}^{-1}$	$k600 \pm \text{SD}$ $\text{cm h}^{-1}$	References
Red		Vietnam	$1,589 \pm 43$	$550.3 \pm 16.9$	$50.9 \pm 27$	This study
Mekong	Downstream	Laos and Cambodia	703 – 1597	88.1 - 378.4	12.4 - 44.5	<i>Alin et al., 2011</i>
Tonle Sap	Stung Siem Reap	Cambodia	3,067	139.1	$5.6 \pm 0.9$	<i>Alin et al., 2011</i>
Tonle Sap	Pousat River	Cambodia	1,404	98.5	$10.8 \pm 2.8$	<i>Alin et al., 2011</i>
Musi		Indonesia	$4,317 \pm 928$	$5 \pm 1.1$	$21.8 \pm 4.7$	<i>Wit et al., 2015</i>
Batanghari		Indonesia	$2,401 \pm 18$	$1.8 \pm 0.4$	$21.8 \pm 4.7$	<i>Wit et al., 2015</i>
Indragiri		Indonesia	$5,779 \pm 527$	$9.7 \pm 2.2$	$21.8 \pm 4.7$	<i>Wit et al., 2015</i>
Siak		Indonesia	$8,557 \pm 528$	$8.3 \pm 1.9$	$22.0 \pm 4.7$	<i>Wit et al., 2015</i>
Lupar		Malaysia	$1,274 \pm 148$	$13 \pm 3.0$	$26.5 \pm 9.3$	<i>Wit et al., 2015</i>
Saribas		Malaysia	$1,159 \pm 29$	$14.6 \pm 3.3$	$17.0 \pm 13.6$	<i>Wit et al., 2015</i>
Changjiang		China	$1,297 \pm 901$	143	8 - 15	<i>Wang et al., 2007</i>
Maotiao		China	3741	108	10	<i>Wang et al. 2011</i>
Longchuan		China	2,101	156	8	<i>Li et al., 2012</i>
Yellow		China	$2,811 \pm 1,986$	$856 \pm 409$	$42.1 \pm 16.9$	<i>Ran et al., 2015</i>
Xijiang		China	600 - 7200	160-357	15	<i>Yao et al., 2007</i>
Krishna		India	$17,210 \pm 3501$		nd	<i>Sarma et al.,</i>

Godavari		49,832±1042			2012
Mahanadi		95,884±2235			
Ganges		5,030±100			
Gaderu Creek	India	2,216 ± 864a	56.0 ± 100.9	4 ± 5	<i>Borges et al., 2003</i>
Rhone	France	2,016±944	nd	15	<i>Cole et al., 2001</i>
Hudson	USA	1,125±403	nd	4.1	<i>Raymond et al., 1997</i>
Ottawa	Canada	1,200	80.8	4	<i>Telmer and Veizer, 1999</i>
Amazon		4,351 ± 1900	190 ± 55	9.6 ± 3.8	<i>Richey et al., 2002</i>
Mississippi		100 - 600	270	3.9	<i>Dubois et al., 2010; Lohrenz and Cai, 2006</i>
Nagada Creek	The northern Papua New Guinea coast	799 ± 357	43.6 ± 33.2	8 ± 6	<i>Borges et al., 2003</i>
Negro		3,011±304	534±148	20.3±7.6	
Solimoos		5,685±464	488±83	7.3±2.4	
Arguaia	South America	1,012±309	153±44	8.4±2.4	<i>Rasera et al., 2013</i>
Javaes		1,673±273	108±34	7.2±2.7	
Caxiuana		3,216±496	335±61	11.3±3.5	

Teles Pires			1,419±224	100±27	10.1±1.5	
Cristalino			1,938±201	202±31	15.7±2.2	
	Upper	North America	1,220 ± 9	6	1.25	
Yukon	Middle		1,890 ± 10	62	7.92	<i>Striegl et al., 2007; 2012</i>
	Lower		3,091 ± 17	193	15	
Congo			2019 - 6855	298.6	9.3 – 10.3	<i>Wang et al., 2013</i>

<sup>a</sup> calculated the values for the  $p\text{CO}_2$  Water-Air Gradient (  $p\text{CO}_2$  in ppm)

nd. No data

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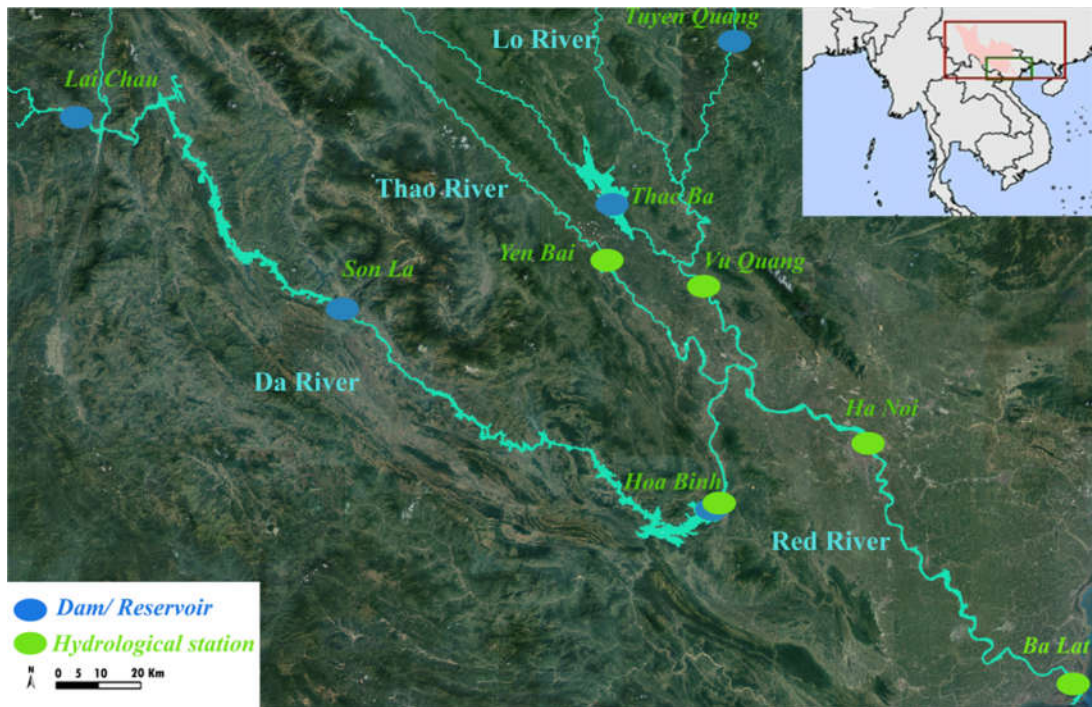
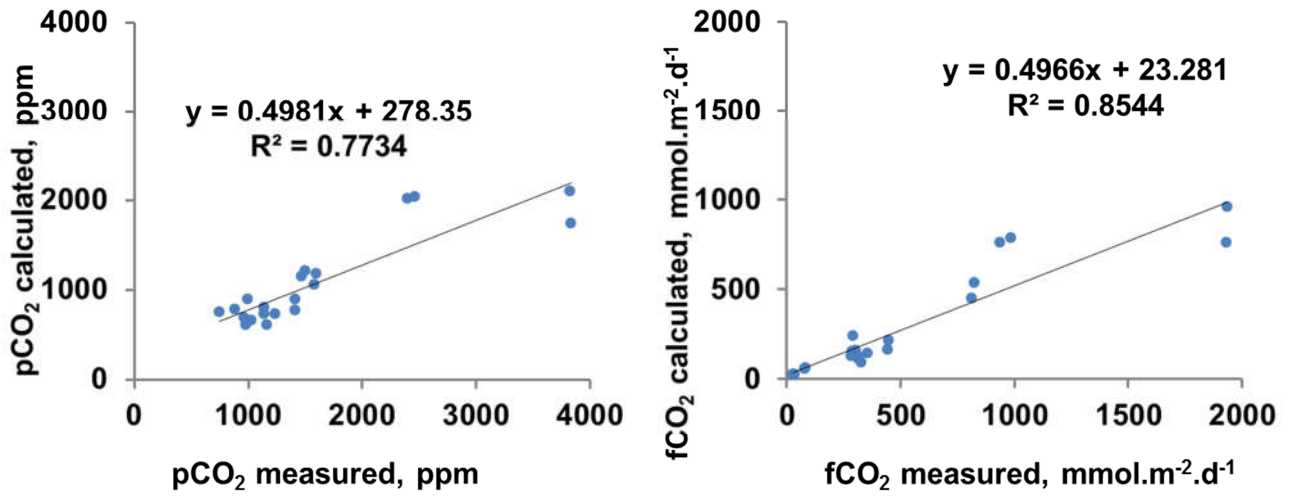


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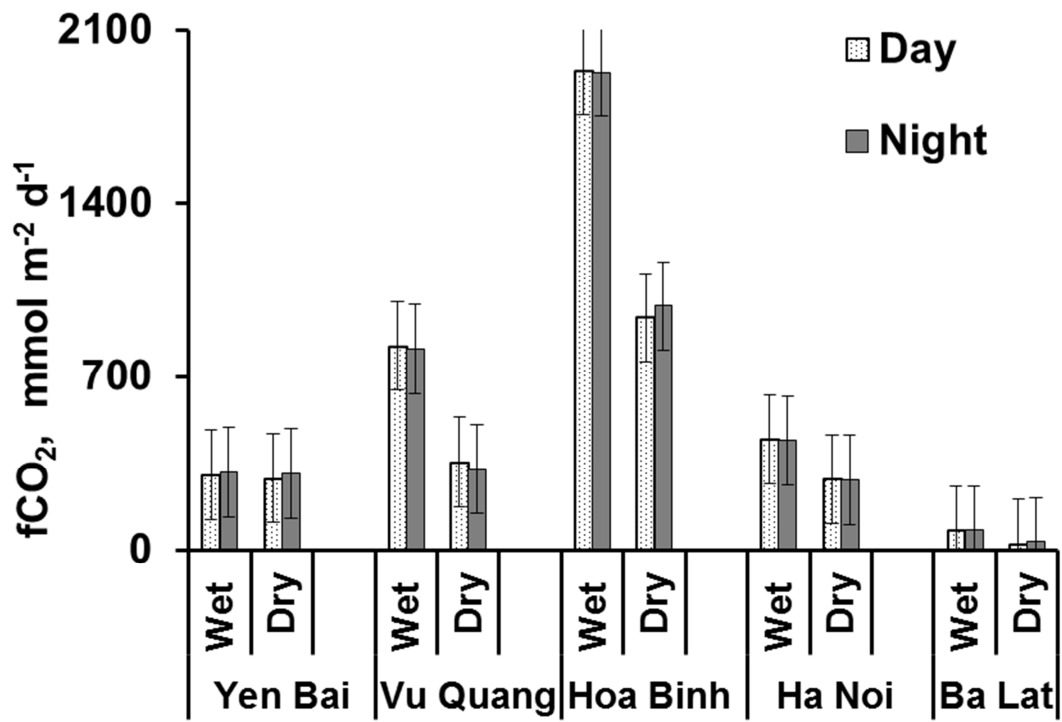
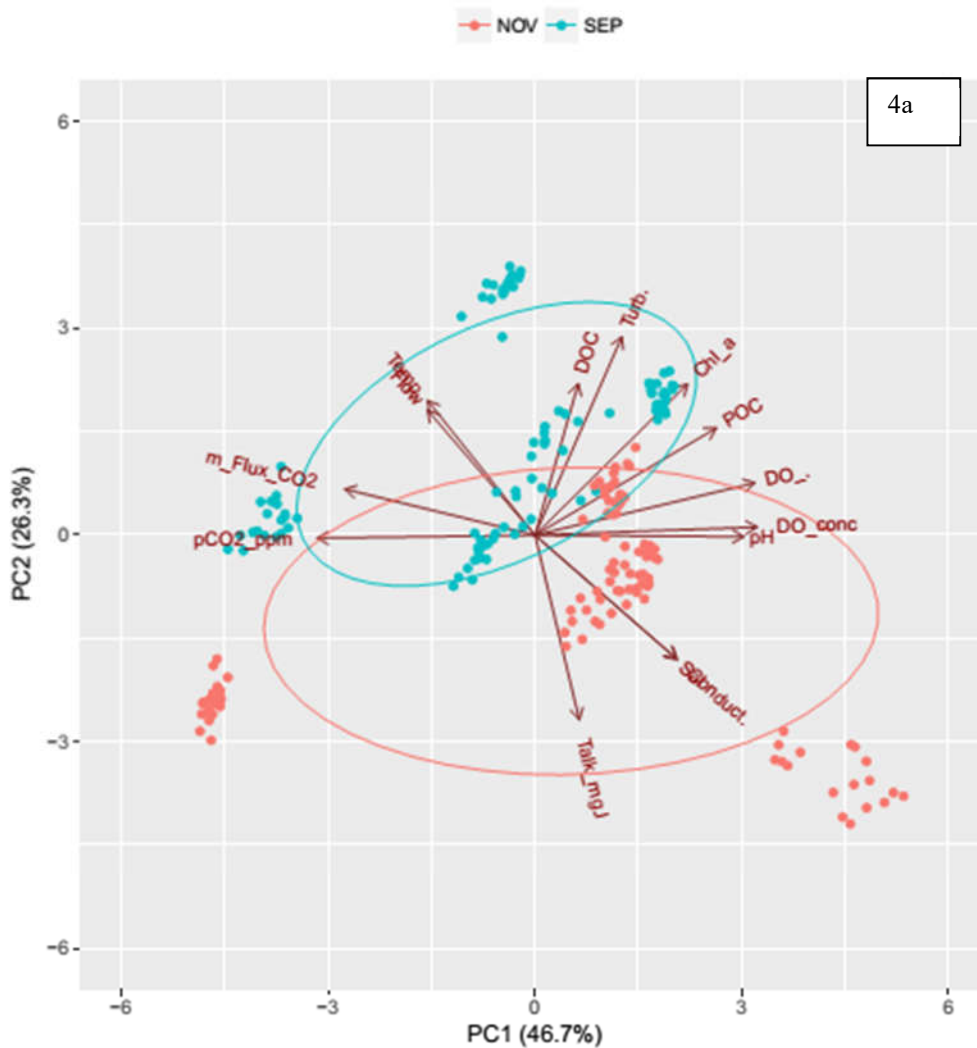


Figure 3 Spatial and seasonal variation of CO<sub>2</sub> out-gassing flux in the Red River system in 2014.



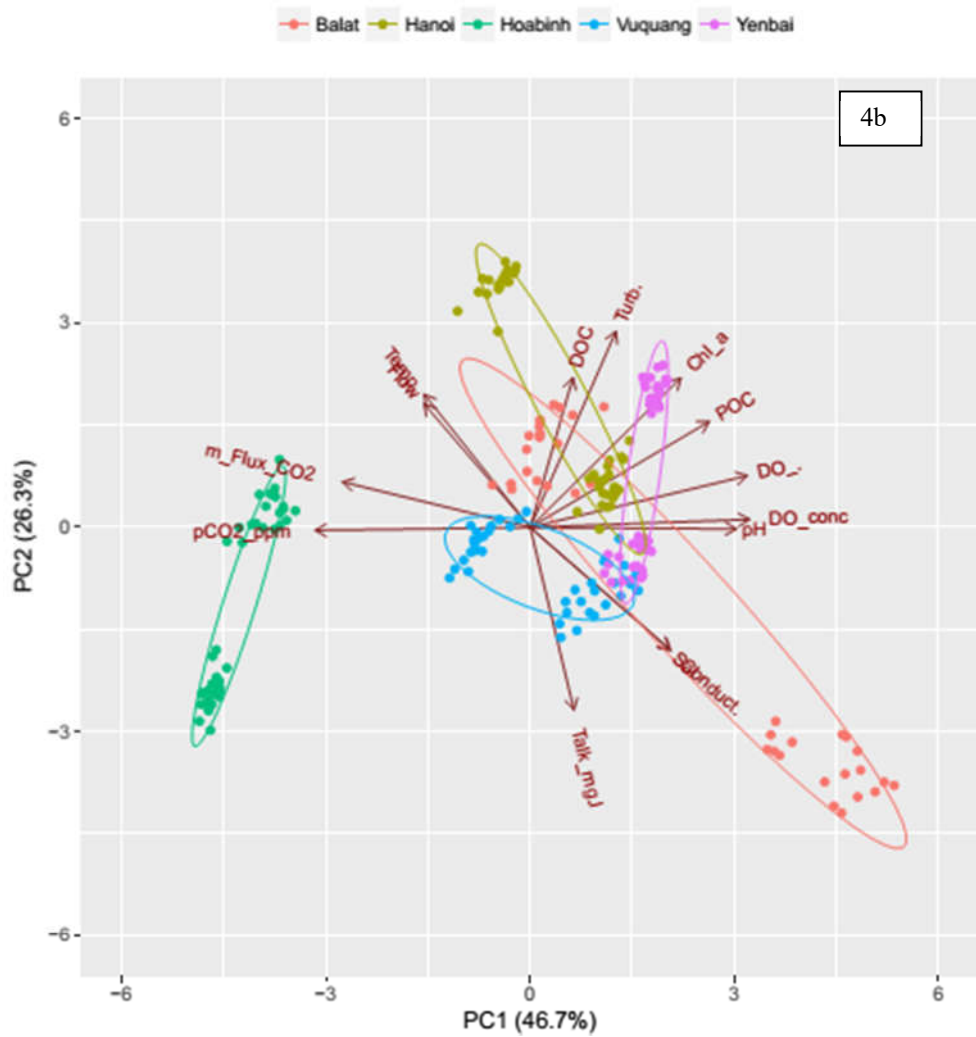
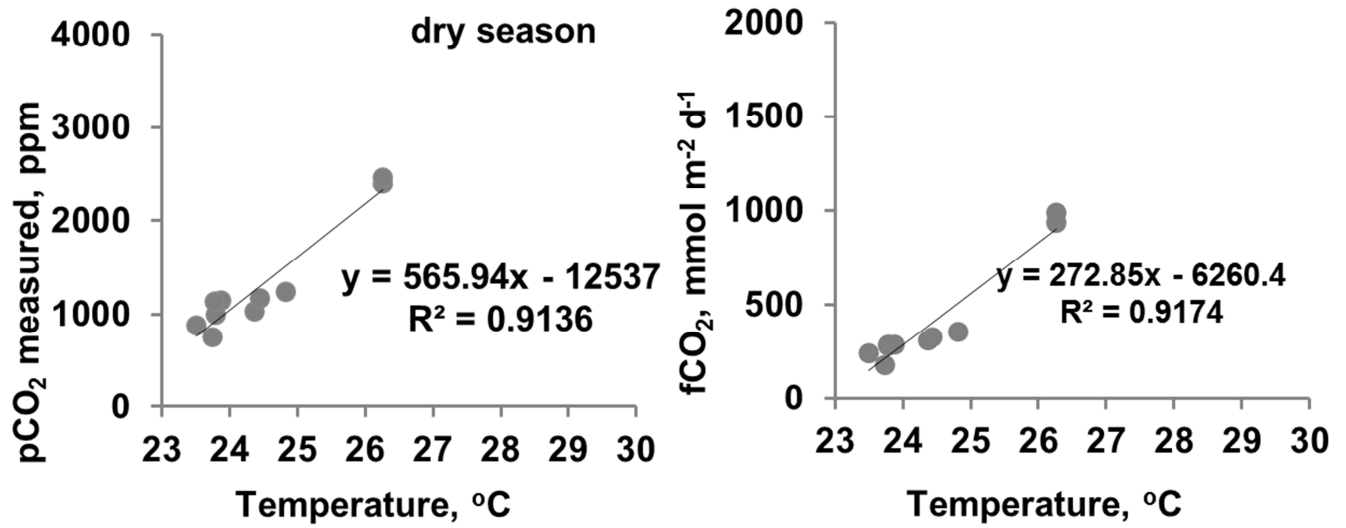


Figure 4.

5 4a: Seasonal variation of different variables at 5 sites of the lower Red River in dry (Nov) and wet (Sept) seasons in 2014

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