

# Dissolved carbon biogeochemistry and export in mangrove-dominated rivers of the Florida Everglades

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15 **Abstract.** The Shark and Harney Rivers, located on the southwest coast of Florida, USA, originate in the freshwater,  
karstic marshes of the Everglades and flow through the largest contiguous mangrove forest in North America. In  
November 2010 and 2011, dissolved carbon source-sink dynamics were examined in these rivers during SF<sub>6</sub> tracer  
release experiments. Approximately 80% of the total dissolved carbon flux from all sources (i.e., freshwater  
wetlands, mangrove, carbonate dissolution, and marine input) out of the Shark and Harney Rivers during these  
20 experiments was as inorganic carbon, either via air-water CO<sub>2</sub> exchange or longitudinal flux of inorganic carbon to  
the coastal ocean. Of the total mangrove-derived dissolved inorganic carbon (DIC) exported from the forests into  
these rivers, between 42 and 48% was emitted to the atmosphere, with the remaining discharged to the coastal  
ocean. Dissolved organic carbon (DOC) represented ca. 10% of the total mangrove-derived dissolved carbon export  
from the forests. The sum of mangrove-derived DIC and DOC export to these rivers was estimated to be at least  
25 18.9 to 24.5 mmol m<sup>-2</sup> d<sup>-1</sup>, a rate lower than other independent estimates from Shark River and from other mangrove  
forests. Results from these experiments also suggest that in this region, mangrove contribution to the estuarine flux  
of dissolved carbon to the ocean is less than 10%.

## 1 Introduction

In many tropical and sub-tropical regions, mangrove forests are a typical feature surrounding estuaries  
30 (Twilley et al., 1992; Bouillon et al., 2008a). Mangroves are thought to play an important role in tropical and  
subtropical coastal biogeochemical cycling and the global coastal carbon budget, due to their high productivity and  
rapid cycling of organic and inorganic carbon (Twilley et al., 1992; Jennerjahn and Ittekkot, 2002; Dittmar et al.,  
2006). However, there remain uncertainties regarding the fate of mangrove-fixed carbon and the amount of carbon  
exported to the coastal waters from these ecosystems (Bouillon et al., 2008a; Bouillon et al., 2008b; Kristensen et  
35 al., 2008).

Bouillon et al. (2008a) showed that over 50% of the carbon fixed by mangroves through photosynthesis  
could not be accounted for by growth in biomass, accumulation in soils, and export of organic carbon, and suggested  
that a large fraction of this missing organic carbon may be mineralized to dissolved inorganic carbon (DIC) and  
either lost to the atmosphere or exported to the surrounding waters. In fact, several studies have shown that the  
40 lateral advective transport of interstitial waters through tidal pumping represents a major carbon export pathway  
from mangroves into adjacent waters, both for DIC (Koné and Borges, 2008; Miyajima et al., 2009; Maher et al.,  
2013) and dissolved organic carbon (DOC) (Dittmar and Lara, 2001; Bouillon et al., 2007c). However, to date,  
lateral mangrove-derived aquatic carbon fluxes (as a proportion of overall forest carbon mass balance) have only  
been estimated for short time periods and over limited spatial (e.g., plot) scales (e.g., Troxler et al., 2015). These  
45 studies also typically do not determine the fate of mangrove-derived carbon once it is exported from the forest  
through tidal pumping and drainage. Additional measurements of the magnitude and fate of mangrove carbon export  
at the basin scale are needed to help quantify connections between inter-tidal, estuarine and coastal ocean carbon  
cycles.

Basin-scale rates of lateral advective dissolved carbon export from a tidal mangrove forest are difficult to  
50 measure, because these fluxes are heterogeneous over space and time due to large variability in inundation patterns,  
forest structure, topography, and soil hydraulic properties. However, mangrove-derived dissolved carbon fluxes may  
be estimated in some systems using information on the spatial distribution of carbon-related measurements in  
adjacent waters. For example, the carbon balance of tidal riverine systems adjacent to mangrove forests should  
integrate the spatial and temporal variability of these lateral fluxes.

55           The objective in the study is to quantify dissolved carbon source-sink dynamics in a subtropical estuary dominated by two tidal rivers, the Shark and Harney Rivers in Everglades National Park, Florida, USA. These rivers are centrally-located within the largest contiguous mangrove forest in North America and they discharge to the Gulf of Mexico. The total dissolved carbon inventories and fluxes in these rivers are determined using a series of discrete and continuous measurements of carbon-related parameters along a salinity gradient, and the mangrove contribution  
60 ~~separated~~ using measurements of stable isotopic composition of dissolved organic and inorganic carbon. The results are then scaled by the area of mangrove forest that surrounds these rivers to express dissolved carbon fluxes on an aerial basis for comparison to independent measurements of dissolved carbon fluxes from this forest.

## 2    **Methods**

### 2.1   **Study site**

65           The tidal-dominated Shark and Harney Rivers (river and estuary are used interchangeably in this contribution) are surrounded by mangrove forests and located on the southwest coast of Florida (Fig. 1), within Everglades National Park. The subtropical climate in southern Florida is characterized by a May to October wet season, when approximately 60% of the annual precipitation occurs (Southeast Regional Climate Center, <http://www.sercc.com>). The Shark and Harney Rivers together discharge approximately 50% of the flow from the  
70 Shark River Slough (SRS), the primary drainage feature of Everglades National Park, to the Gulf of Mexico (GOM) (Levesque, 2004). Seasonal variation of the water discharge from SRS mostly follows the precipitation patterns (Saha et al., 2012), and influences the transport of nutrients to the mangrove ecotone (Rivera-Monroy et al., 2011). The Shark and Harney Rivers are each approximately 15 km long, and connected in Tarpon Bay (Fig. 1). The mean depths of Tarpon Bay, Shark River, and Harney River at mid tide are  $1.4 \pm 0.3$ ,  $2.8 \pm 0.4$ , and  $2.6 \pm 0.4$  m (Ho et al.,  
75 2014), respectively, and the surface areas are  $1.48 \times 10^6$ ,  $2.54 \times 10^6$ , and  $2.75 \times 10^6$  m<sup>2</sup>, respectively. The inter-tidal zones bordering the Shark and Harney Rivers are dominated by *Rhizophora mangle* (red mangrove), *Avicennia germinans* (black mangrove), *Laguncularia racemose* (white mangrove), and *Conocarpus erectus* (buttonwood). Semi-diurnal tides in this region inundate the forest as often as twice a day. River discharge to the GOM is primarily influenced by tides, wind, and freshwater inflow from SRS (Levesque, 2004).

80           Discharges are determined by the US Geological Survey at stations near the midpoints of Shark River (USGS 252230081021300 Shark River) and Harney River (USGS 252551081050900 Harney River) (Fig. 1).

Discharges are generally lower during March-May than the rest of the year. Hourly mean residual discharge values (i.e., filtered for tides) from March to May of the 5-year period from 2007 to 2011 ranged from -21.9 to 24.1 m<sup>3</sup> s<sup>-1</sup>, with a mean of 0 m<sup>3</sup> s<sup>-1</sup> for Shark River, and ranged from -28.9 to 38.5 m<sup>3</sup> s<sup>-1</sup>, with a mean of 4.4 m<sup>3</sup> s<sup>-1</sup> for Harney River. Positive values indicate flow towards the GOM. For the rest of the year (i.e., June to February), these values ranged from -46.2 to 89.2 m<sup>3</sup> s<sup>-1</sup>, with a mean of 8.8 m<sup>3</sup> s<sup>-1</sup> for Shark River, and -41.6 to 75.0 m<sup>3</sup> s<sup>-1</sup>, with a mean of 11.3 m<sup>3</sup> s<sup>-1</sup> for Harney River.

## 2.2 Shark River Tracer Release Experiments

Two field studies were conducted as part of the Shark River Tracer Release Experiment (SharkTREx 1: 20 to 25 November 2010; SharkTREx 2: 10 to 15 November 2011; (Ho et al., 2014)). The mean residual discharges for Shark River were 6.9 (hourly range: -2 to 19.9) and 4.9 (hourly range: -18.9 to 34.8) m<sup>3</sup> s<sup>-1</sup>, during SharkTREx 1 and 2, respectively, and those for Harney River were 6.0 (hourly range: -1.6 to 22.8) and 1.9 (hourly range: -17.3 to 30.6) m<sup>3</sup> s<sup>-1</sup>, during SharkTREx 1 and 2, respectively (U.S. Geological Survey, 2016).

During both campaigns, an inert tracer (sulfur hexafluoride; SF<sub>6</sub>) was injected in the river near the point where the rivers diverge just downstream of Tarpon Bay (25.4092, -81.0083) to determine the rates of longitudinal dispersion, and the water residence time. Each day, longitudinal surveys were made along the Shark and Harney Rivers from Tarpon Bay to the GOM, and included continuous underway measurements of temperature, salinity, SF<sub>6</sub>, dissolved O<sub>2</sub> (DO; μmol kg<sup>-1</sup>); and partial pressure of CO<sub>2</sub> (pCO<sub>2</sub>; μatm), and discrete measurements of total alkalinity (TAlk; μmol kg<sup>-1</sup>), dissolved inorganic carbon (DIC; μmol kg<sup>-1</sup>), dissolved organic carbon (DOC; μmol kg<sup>-1</sup>), stable carbon isotopic composition of DIC and DOC (δ<sup>13</sup>C<sub>DIC</sub> and δ<sup>13</sup>C<sub>DOC</sub>, respectively; ‰).

## 2.3 Discrete measurements

During SharkTREx 1, three to five surface water samples were collected daily in the Shark River with a 5-L Niskin bottle at ~0.5 m below the surface for the analysis of TAlk, DOC, δ<sup>13</sup>C<sub>DIC</sub>, and δ<sup>13</sup>C<sub>DOC</sub>. At each sampling site, vertical profiles of temperature, salinity, and DO were recorded using a conductivity, temperature, and depth sonde (Sea-Bird SBE 19plus V2) equipped with a Clark-type polarographic O<sub>2</sub> sensor (SBE 43). These profiles showed that the water column was vertically well mixed. No discrete samples were collected in the Harney River during SharkTREx 1. During SharkTREx 2, discrete samples for DIC, TAlk, DOC, δ<sup>13</sup>C<sub>DIC</sub>, and δ<sup>13</sup>C<sub>DOC</sub> were collected daily at 20 stations distributed within the Shark and Harney Rivers (Fig. 1).

### 2.3.1 Total alkalinity and dissolved inorganic carbon

110 During SharkTREx 1, samples for TAlk were collected in 250 mL HDPE bottles after passing through a 0.45  $\mu\text{m}$  filter. They were stored on ice for transport to the laboratory at FIU, where TAlk was determined at room temperature using an automated titrator (Brinkman Titrino 751) with 0.1 N HCl to a pH of 2. TAlk was calculated from the volume of acid added at the inflection point closest to a pH of 4, and reported as  $\text{meq L}^{-1} \text{HCO}_3^-$  since the original pH of the water samples was near neutral. The precision of the measurements was  $\pm 2\%$  from replicate analysis ( $n = 5$ ) with an accuracy of  $\pm 2\%$  as determined by analysis of certified reference material (Dickson, 2010). 115 DIC and pH were computed from TAlk and  $\text{pCO}_2$  using the dissociation constants of (Millero, 2010) for estuarine waters.

During SharkTREx 2, samples for TAlk and DIC were collected in 550 mL borosilicate glass bottles, poisoned with  $\text{HgCl}_2$ , and sealed with hydrocarbon grease (Apiezon M). The samples were stored at room 120 temperature in the dark for travel to the laboratory at NOAA/AOML. Samples for TAlk were measured in an open thermostated cell (25  $^\circ\text{C}$ ) with an automated titrator (Metrohm 765 Dosimat) connected to a pH glass-reference electrode system (Orion), using 0.2 M HCl as a titrant, and determined from the equivalence point of the titration curve using a non-linear least-squares fit. For DIC analysis, water samples were first acidified to convert all the carbonate species to  $\text{CO}_2$  in a DIC analyzer (Apollo SciTech), and then measured with a NDIR detector (LI-COR 125 LI-7000). Calibrations for DIC and TAlk were performed using certified reference material (Dickson, 2010). The analytical uncertainty of the DIC and TAlk measurements based on replicate samples are 0.1 and 0.2%, respectively.

The measured TAlk and  $\text{pCO}_2$  from SharkTREx 2 were used to calculate DIC using CO2SYS (Pierrot et al., 2006) and the dissociation constants of (Millero, 2010), and the results were  $1.3 \pm 1.1\%$  (range: -2.4 to +4.4%) higher than the measured DIC, possibly indicating a slight contribution (ca. 1%) to TAlk from organic or particulate 130 material, as the samples were not filtered.

### 2.3.2 Dissolved organic carbon

The samples analyzed for DOC were filtered with pre-combusted 0.7  $\mu\text{m}$  GF/F filters and collected in pre-cleaned, acid-washed, brown high-density polyethylene bottles (HDPE; Nalgene). Containers were rinsed three times before sample collection, transported on ice to the FIU SERC Nutrient Analysis Lab, and stored in a 135 refrigerator until analyses within three weeks of collection. DOC was measured using the high-temperature catalytic combustion method on a total organic carbon analyzer (Shimadzu TOC-V).

### 2.3.3 Stable carbon isotopic composition

Samples for  $\delta^{13}\text{C}_{\text{DIC}}$  were collected in 40 ml glass bottles after passing the sample through a GF/F filter, and then poisoning with  $\text{HgCl}_2$ . In the laboratory at RSMAS, vials with 0.5 ml 103%  $\text{H}_3\text{PO}_4$  were flushed for 60 s with He. Approximately 2 ml of sample were then injected into the vial, and after sonification the accumulated  $\text{CO}_2$  was analyzed by a gas chromatograph (GC) coupled to an isotope ratio mass spectrometer (GC-IRMS; Thermo Delta V). The  $\delta^{13}\text{C}$  was calibrated using two standards of  $\text{NH}_4\text{HCO}_3$  with differing  $\delta^{13}\text{C}$  values dissolved in  $\text{H}_2\text{O}$  whose isotopic compositions had been previously calibrated relative to NBS-19 using conventional dual inlet mass spectrometry (Finnigan-MAT 251). The  $\delta^{13}\text{C}$  values are reported relative to the Vienna Pee Dee Belemnite (VPDB) standard.

Samples for  $\delta^{13}\text{C}_{\text{DOC}}$  were collected in 60 ml brown HDPE bottles and stored on ice until returned to the lab at FIU.  $\delta^{13}\text{C}_{\text{DOC}}$  samples were filtered with GF/F (0.7  $\mu\text{m}$ ) filter, and then stored in pre-cleaned 40 ml bottles until analyses. Measurements for  $\delta^{13}\text{C}_{\text{DOC}}$  were made using a total organic carbon (TOC) analyzer (Aurora 1030W, OI Analytical) coupled to a cavity ring-down spectroscopy system (CRDS; G1111-i, Picarro) following the approach of Ya et al. (2015). DIC was removed by adding phosphoric acid and sparging with  $\text{N}_2$  gas. 1.5 ml of sample was chemically oxidized to  $\text{CO}_2$  at a temperature of 98°C in the presence of sodium persulfate ( $\text{Na}_2\text{S}_2\text{O}_8$ ). The  $\text{CO}_2$  generated was detected by non-dispersive infrared absorption (NDIR) for determination of DOC. The  $\text{CO}_2$  was collected in a gas-tight bag and then pulsed into the CRDS for the  $\delta^{13}\text{C}$  measurement. Analytical precision based on replicated standards ranged from  $\pm 0.15$  to  $\pm 1.52$  ‰ for this study.

### 2.4 Underway measurements

Surface water was continuously pumped from an intake located near the bow of the boat at a water depth of approximately 1 m during tracer recovery operations. Water temperature and salinity were continuously recorded using a thermosalinograph (SBE 45 MicroTSG). During SharkTREx 1, DO was measured underway with a membrane covered galvanic sensor (WTW Cellox 325) calibrated with saturated air. During SharkTREx 2, DO was measured using an oxygen optode (Aanderaa 3835) calibrated against Winkler titration.

Underway measurements of atmospheric and waterside  $\text{pCO}_2$  were made. Waterside  $\text{pCO}_2$  were obtained with a showerhead type equilibrator coupled to a non-dispersive infrared (NDIR) analyzer (LI-COR 840A). Measurements of underway  $\text{SF}_6$  were made with an automated  $\text{SF}_6$  analysis system (Ho et al., 2002). Both of these measurements have been described in detail in Ho et al. (2014)

### 2.5 Inventories of DIC, DOC and DO

The inventories of DIC, DOC and DO were calculated in the same way that  $\text{SF}_6$  inventories were determined in Ho et al. (2014). The river was divided into 100-m longitudinal sections, and the measured concentrations, corrected for tidal movement to slack before ebb for each day, were assigned to each section  $i$  and

then summed over the entire length of the river. For example, to calculate the inventory of DIC, denoted

170  $\sum[\text{DIC}]_{\text{observed}}$  (mol):

$$\sum[\text{DIC}]_{\text{observed}} = \sum_{i=1}^n [\text{DIC}]_i \times V_i, \quad (1)$$

where  $[\text{DIC}]_i$  is the mean concentration ( $\text{mol L}^{-1}$ ) in section  $i$ ,  $V_i$  is the volume of the river (L) in section  $i$  at mid-tide, and  $n$  is the number of sections in each river ( $n = 273$  for Shark River and Tarpon Bay;  $n = 152$  for Harney River). DOC and DO inventories were also calculated using Eq. (1), by substituting  $[\text{DOC}]_i$  or  $[\text{DO}]_i$  for  $[\text{DIC}]_i$

175 accordingly. The inventories of DIC and DOC were separated into contributions from estuarine and non-estuarine sources, first by determining inventories for DIC assuming conservative mixing between the freshwater and marine end members and then subtracting these inventories from the total observed inventories. The estuarine DIC inventory,  $\sum[\text{DIC}]_{\text{estuary}}$ , representing the DIC from all estuarine sources, was calculated as follows:

$$\sum[\text{DIC}]_{\text{estuary}} = \sum[\text{DIC}]_{\text{observed}} - \sum[\text{DIC}]_{\text{conserv}} + \sum[\text{DIC}]_{\text{gasex}}, \quad (2)$$

180 where  $\sum[\text{DIC}]_{\text{conserv}}$  is the inventory of DIC assuming conservative mixing between freshwater and marine end members (i.e., from non-estuarine sources), and  $\sum[\text{DIC}]_{\text{gasex}}$  is the inventory of DIC lost to air-water gas exchange from the estuary, due to  $\text{pCO}_2$  in the water being above solubility equilibrium with the atmosphere (see section 2.6). The freshwater and marine end-members were assigned ~~to the~~ values measured at the lowest (Tarpon Bay) and highest salinities, respectively.

185 The total  $\text{O}_2$  deficit in Shark River during the experiments was determined by examining the difference in  $\text{O}_2$  inventories for conservative mixing and actual measurements, correcting for  $\text{O}_2$  influx due to gas exchange using a formulation similar to Eq. (2) above (i.e.,  $\sum[\text{DO}]_{\text{deficit}} = \sum[\text{DO}]_{\text{conserv}} - \sum[\text{DO}]_{\text{observed}} + \sum[\text{DO}]_{\text{gasex}}$ ).

## 2.6 Air-water $\text{O}_2$ and $\text{CO}_2$ fluxes

190 To enable comparison between different gases and different aquatic environments, it is customary to normalize gas transfer velocities to a Schmidt number ( $Sc$ ; kinematic viscosity of water divided by diffusion coefficient of gas in water) of 600,  $k(600)$ , corresponding to that of  $\text{CO}_2$  in freshwater at  $20^\circ\text{C}$ .  $k(600)$  for SharkTREx 1 and 2, determined from the parameterization proposed in Ho et al. (2016), were  $3.5 \pm 1.0$  and  $4.2 \pm 1.8 \text{ cm h}^{-1}$ , respectively. To determine  $k$  for  $\text{O}_2$  and  $\text{CO}_2$  at the temperature and salinity measured in the rivers, the following equation was used, assuming a  $Sc^{-1/2}$  scaling (Jähne et al., 1987):

$$195 \quad k_{\text{O}_2} = k(600) \left( \frac{Sc_{\text{O}_2}}{600} \right)^{-1/2}, \quad (3)$$



where  $k$  and  $Sc$  of  $\text{CO}_2$  could be substituted in Eq. (3) for  $\text{O}_2$ , and  $Sc$  for  $\text{O}_2$  and  $\text{CO}_2$  were calculated as a function of temperature and salinity using data compiled by Wanninkhof (2014).

Air-water  $\text{O}_2$  fluxes ( $F_{\text{O}_2}$ ;  $\text{mmol m}^{-2} \text{d}^{-1}$ ) were calculated as follows:

$$F_{\text{O}_2} = k_{\text{O}_2} (\text{O}_{2_{\text{equil}}} - \text{O}_2), \quad (4)$$

200 where  $k_{\text{O}_2}$  ( $\text{cm h}^{-1}$ ) is the gas transfer velocity for  $\text{O}_2$ ,  $\text{O}_{2_{\text{equil}}}$  ( $\text{mmol m}^{-3}$ ) is the equilibrium concentration of  $\text{O}_2$  in the water at a given temperature and salinity (Garcia and Gordon, 1992), and  $\text{O}_2$  is the measured oxygen concentration in the water.

Similarly, air-water  $\text{CO}_2$  fluxes ( $F_{\text{CO}_2}$ ;  $\text{mmol m}^{-2} \text{d}^{-1}$ ), which were used to determine changes in DIC due to gas exchange, were calculated as follows:

$$205 \quad F_{\text{CO}_2} = k_{\text{CO}_2} K_0 \Delta p\text{CO}_2, \quad (5)$$

where  $k_{\text{CO}_2}$  ( $\text{cm h}^{-1}$ ) is the gas transfer velocity for  $\text{CO}_2$ ,  $K_0$  ( $\text{mol atm}^{-1} \text{m}^{-3}$ ) is the aqueous-phase solubility of  $\text{CO}_2$  (Weiss, 1974), and  $\Delta p\text{CO}_2$  ( $\mu\text{atm}$ ) is the difference between the measured  $p\text{CO}_2$  in air equilibrated with water and atmospheric  $p\text{CO}_2$ .

As with the inventories,  $F_{\text{CO}_2}$  were separated into estuarine and non-estuarine contributions. Because of the  
210 non-linearity in the relationship between  $p\text{CO}_2$  and other carbonate system parameters, the  $p\text{CO}_2$  in the river expected from conservative mixing was calculated by assuming conservative mixing for DIC and TALK, and then calculating  $p\text{CO}_2$  using CO2SYS (Pierrot et al., 2006), with the dissociation constants of Millero (2010). Then, the non-estuarine  $F_{\text{CO}_2}$  was calculated as above with Eq. (5), and the  $F_{\text{CO}_2}$  attributed to estuarine sources was determined as the difference between total and non-estuarine  $F_{\text{CO}_2}$ .

## 215 **2.7 Estuarine and mangrove contributions to DIC**

DIC in the Shark and Harney Rivers may originate from several sources in addition to input from the freshwater marsh upstream and the coastal ocean, including: 1) mangrove root respiration; 2) organic matter mineralization in sediments or in river water; 3) dissolution of  $\text{CaCO}_3$  in sediments or in river water; and 4) groundwater discharge.

Groundwater in this region is likely to contain DIC from  $\text{CaCO}_3$  dissolution that occurs when saltwater intrudes into  
220 the karst aquifer that underlies this region (Price et al., 2006). In this setting, the combination of #1 and #2 represents the mangrove source of DIC ( $[\text{DIC}]_{\text{mangrove}}$ ), and the combination of #3 and #4 represents the  $\text{CaCO}_3$  dissolution source ( $[\text{DIC}]_{\text{dissolution}}$ ) to estuarine [DIC]:

$$[\text{DIC}]_{\text{estuary}} = [\text{DIC}]_{\text{observed}} - [\text{DIC}]_{\text{conserv}} + [\text{DIC}]_{\text{gasex}} = [\text{DIC}]_{\text{mangrove}} + [\text{DIC}]_{\text{dissolution}} \quad (6)$$

where  $[\text{DIC}]_{\text{observed}}$  is the observed DIC concentration,  $[\text{DIC}]_{\text{conserv}}$  is the DIC concentration expected by conservative mixing of the two end-members, and  $[\text{DIC}]_{\text{gasex}}$  is the correction for change in  $[\text{DIC}]_{\text{observed}}$  due to loss through air-water gas exchange as the water transits through the estuary.  $[\text{DIC}]_{\text{gasex}}$  was determined from  $F_{\text{CO}_2}$  and the residence time of water during each experiment (Ho et al., 2016).

Measurements of  $\delta^{13}\text{C}_{\text{DIC}}$  and estuarine DIC/Talk ratios were used to determine the mangrove sources to estuarine DIC. Fixation of  $\text{CO}_2$  through photosynthesis is neglected in both models as these rivers are characterized by low chlorophyll-*a* concentration and low phytoplankton biomass (Boyer et al., 1997). During SharkTREx 1 and 2, there was a negligible difference between  $\text{pCO}_2$  measured during the day and night.

### 2.7.1 Determining mangrove contribution from $\delta^{13}\text{C}_{\text{DIC}}$

Processes 1 through 4 listed above influence  $\delta^{13}\text{C}_{\text{DIC}}$  in the estuary differently due to the differences in the  $\delta^{13}\text{C}$  values originating from respiration of mangrove-derived organic matter, and **CaCO<sub>3</sub> dissolution**. The isotopic fractionation during respiration of organic matter is small, and the  $\delta^{13}\text{C}_{\text{DIC}}$  values produced via this pathway should be equivalent to the  $\delta^{13}\text{C}$  of the organic matter respired (DeNiro and Epstein, 1978). Also, the isotopic fractionation during dissolution/re-precipitation of  $\text{CaCO}_3$  is thought to be negligible (Salomons and Mook, 1986).

The expected  $\delta^{13}\text{C}$  values of DIC in the rivers as a result of conservative mixing ( $\delta^{13}\text{C}_{\text{conserv}}$ ) of the marine and freshwater end-members of the Shark and Harney Rivers were calculated as follows (Mook and Tan, 1991):

$$\delta^{13}\text{C}_{\text{conserv}} = \frac{S([\text{DIC}]_F \delta^{13}\text{C}_F - [\text{DIC}]_M \delta^{13}\text{C}_M) + S_F[\text{DIC}]_M \delta^{13}\text{C}_M - S_M[\text{DIC}]_F \delta^{13}\text{C}_F}{S([\text{DIC}]_F - [\text{DIC}]_M) + S_F[\text{DIC}]_M - S_M[\text{DIC}]_F}, \quad (7)$$

where  $[\text{DIC}]$  is the observed DIC concentration,  $S$  is the measured salinity, and  $M$  and  $F$  subscripts refer to the marine and freshwater end-members, respectively.

An estimate of the maximum contribution of  $[\text{DIC}]_{\text{mangrove}}$  and  $[\text{DIC}]_{\text{dissolution}}$  to  $[\text{DIC}]_{\text{estuary}}$  can be obtained by solving Equations 6 and 8:

$$\delta^{13}\text{C}_{\text{DIC}} \times [\text{DIC}]_{\text{observed}} = \delta^{13}\text{C}_{\text{conserv}} \times [\text{DIC}]_{\text{conserv}} + \delta^{13}\text{C}_{\text{mangrove}} \times [\text{DIC}]_{\text{mangrove}} + \delta^{13}\text{C}_{\text{dissolution}} \times [\text{DIC}]_{\text{dissolution}} - \left( \delta^{13}\text{C}_{\text{DIC}} - \epsilon^{13}\text{C}_{\text{DIC}-\text{CO}_2} \right) \times [\text{DIC}]_{\text{gasex}}, \quad (8)$$

where the  $\delta^{13}\text{C}_{\text{conserv}}$  value is the DIC isotopic composition expected for conservative mixing (Mook and Tan, 1991),  $\delta^{13}\text{C}_{\text{mangrove}}$  is the isotopic composition for mangrove-derived material (-30‰; Mancera-Pineda et al., 2009), the

250  $\delta^{13}\text{C}_{dissolution}$  value is the  $\delta^{13}\text{C}$  composition of calcite ( $\sim 1\text{‰}$ ), and  $\epsilon^{13}\text{C}_{\text{DIC}-\text{CO}_2}$  is the equilibrium isotope fractionation between DIC and  $\text{CO}_2$  gas ( $\sim 8\text{‰}$ ; Zhang et al., 1995).

### 2.7.2 Determining mangrove contribution from TAlk/DIC

An independent approach to separate the mangrove contribution from  $\text{CaCO}_3$  dissolution is to use the covariation of  $[\text{DIC}]_{estuary}$  and  $[\text{TAlk}]_{estuary}$  as an indicator of the biogeochemical processes affecting DIC dynamics (Borges et al., 2003; Bouillon et al., 2007c), as these processes have different effects on DIC and TAlk. Assuming that  $[\text{TAlk}]_{estuary}$  is mainly produced by the dissolution of  $\text{CaCO}_3$ ,  $[\text{DIC}]_{dissolution}$  can be determined as  $0.5 \times [\text{TAlk}]_{estuary}$ , and then  $[\text{DIC}]_{mangrove}$  can be calculated from Eq. (6). However, since sulfate reduction, a primary mineralization pathway in mangrove sediments, may also contribute to  $[\text{TAlk}]$  (Alongi, 1998; Alongi et al., 2005) this calculation represents an upper bound estimate for  $[\text{DIC}]_{dissolution}$  and a lower bound estimate for  $[\text{DIC}]_{mangrove}$ .

### 2.8 Determining mangrove contribution to DOC

260 In the Shark and Harney Rivers, dissolved organic matter may be derived from upstream freshwater wetland species such as periphyton and sawgrass, from seagrass communities and marine phytoplankton, or from mangrove vegetation inside the estuary (Jaffe et al., 2001). The estuarine contributions to DOC ( $[\text{DOC}]_{estuary}$ ) in the rivers was determined in the same way as for DIC above using Eq. (6), by substituting DOC for DIC accordingly, without the correction for gas exchange:

$$265 \quad [\text{DOC}]_{estuary} = [\text{DOC}]_{observed} - [\text{DOC}]_{conserv}, \quad (9)$$

where  $[\text{DOC}]_{observed}$  is the observed DOC concentration,  $[\text{DOC}]_{conserv}$  is the DOC concentration expected from conservative mixing of the two end-members.

Then, measurements of  $\delta^{13}\text{C}_{\text{DOC}}$  were made to ascertain the mangrove source of DOC in the river, in order to determine the proportion of  $[\text{DOC}]_{estuary}$  that is of mangrove origin. The expected  $\delta^{13}\text{C}$  values of DOC as a result of conservative mixing ( $\delta^{13}\text{C}_{conserv}$ ) were calculated using Eq. (7), substituting DOC for DIC. Assuming that  $[\text{DOC}]_{estuary}$  was entirely mangrove-derived,  $[\text{DOC}]_{mangrove}$  should equal:

$$[\text{DOC}]_{mangrove} = \frac{[\text{DOC}]_{observed}\delta^{13}\text{C}_{\text{DOC}} + [\text{DOC}]_{conserv}\delta^{13}\text{C}_{conserv}}{\delta^{13}\text{C}_{mangrove}}, \quad (10)$$

where  $\delta^{13}\text{C}_{mangrove}$  is the isotopic composition for mangrove-derived material ( $\sim -30\text{‰}$ ).

## 2.9 Longitudinal dispersion

275 The longitudinal SF<sub>6</sub> distribution was corrected for tidal movement to slack water before ebb for each day using a method described in Ho et al. (2002). The absolute magnitudes of the average daily corrections were 2.0 and 2.7 km for SharkTREx 1 and 2, respectively, with a range for individual measurements of 0 to 5.8 km and 0 to 7.3 km for SharkTREx 1 and 2, respectively. Longitudinal dispersion coefficient  $K_x$  (m<sup>2</sup> s<sup>-1</sup>) was calculated from the change of moment of the longitudinal SF<sub>6</sub> distribution over time as follows (Fischer et al., 1979; Rutherford, 1994):

$$280 \quad K_x = \frac{1}{2} \left( \frac{d\sigma_x^2}{dt} \right), \quad (11)$$

where  $\sigma_x^2$  is the second moment of the longitudinal SF<sub>6</sub> distribution for each day.

## 2.10 Longitudinal fluxes to the Gulf of Mexico

The longitudinal fluxes of DIC and DOC from Shark and Harney Rivers to the Gulf of Mexico were calculated using the averaged DIC or DOC inventories, and the residence time of water ( $\tau$ ; d), which was determined from the decrease in the inventory of SF<sub>6</sub> after correcting for air-water gas exchange (Ho et al., 2016). For example, the longitudinal DIC flux ( $F_{\text{DIC}}$ ; mol d<sup>-1</sup>) can be calculated as follows:

$$F_{\text{DIC}} = \frac{\Sigma[\text{DIC}]_{\text{observed}}}{\tau}. \quad (12)$$

Equation (12) can be used to calculate the fluxes of any other dissolved or suspended substance in the river by substituting its inventory in place of DIC. As with the inventory calculations, longitudinal fluxes were separated into estuarine and non-estuarine contributions.

The advantage of this method to calculate longitudinal flux in a tidal river over a method that uses net discharge and constituent concentration is that the effect of tidal flushing is implicitly accounted for by the residence time, and therefore there is ~~not a~~ need to explicitly define the fraction of river water in the return flow during each flood tide.

## 295 3 Results and Discussion

### 3.1 Distribution patterns and carbon inventories

During SharkTREx 1, the salinity along the longitudinal transects ranged from 1.2 to 27.1, and the mean ( $\pm$  s.d.) water temperature was  $23.4 \pm 0.2$  °C. During SharkTREx 2, salinity ranged from 0.6 to 27.1, and water temperatures averaged  $22.7 \pm 0.9$  °C.

300 Both pCO<sub>2</sub> and DO showed large spatial variability within the Shark and Harney Rivers during SharkTREx 1 and 2 (Fig. 2). Measured pCO<sub>2</sub> values were well above atmospheric equilibrium along the entire salinity range,

with values ranging from ca. 1000 to 6200  $\mu\text{atm}$ . Maximum  $\text{pCO}_2$  values were observed at intermediate salinities, decreasing towards both end-members, while DO showed the opposite pattern, with saturations ranging from 36 to 113%.

305 The patterns of TALK and DIC along the salinity gradient followed the same trend as  $\text{pCO}_2$  and were clearly non-conservative (Fig. 3a-f). TALK varied between ca. 3400 and 5000  $\mu\text{mol kg}^{-1}$  during SharkTREx 1 and between ca. 3000 and 3900  $\mu\text{mol kg}^{-1}$  during SharkTREx 2. DIC ranged from ca. 3400 to 5100  $\mu\text{mol kg}^{-1}$  during SharkTREx 1, and ca. 2800 to 4000  $\mu\text{mol kg}^{-1}$  during SharkTREx 2.  $\delta^{13}\text{C}_{\text{DIC}}$  values ranged from -10.3 to -6.6 ‰ and from -11.4 to -5.8 ‰ during SharkTREx 1 and 2, respectively. Higher DIC, TALK and  $\text{pCO}_2$  coincided with lower  $\text{O}_2$  saturation, 310 more depleted  $\delta^{13}\text{C}_{\text{DIC}}$ , and lower pH values (Fig. 3g-i), indicative of mineralization of mangrove-derived organic matter within the estuary. A negative correlation was observed between  $\delta^{13}\text{C}_{\text{DIC}}$  and both DIC and  $\text{pCO}_2$  values, demonstrating that the source of estuarine DIC was depleted in  $^{13}\text{C}$ .

During SharkTREx 1, the DOC concentrations in the freshwater end member were higher than SharkTREx 2 (Fig. 4). For both experiments, DOC concentrations followed a non-conservative pattern (see also Cawley et al., 315 2013), but this trend was less apparent during SharkTREx 1 compared to SharkTREx 2 (Fig. 4).

The inventories of DIC, DOC, DO, TALK, and  $\text{pCO}_2$  were relatively constant in the Shark and Harney Rivers, indicating quasi steady state conditions during SharkTREx 1 and 2. Under these conditions, carbon inputs and exports are balanced, and fluxes and concentrations may be examined interchangeably.  $K_x$  during the experiments ( $16.4 \pm 4.7$  and  $77.3 \pm 6.5 \text{ m}^2 \text{ s}^{-1}$  for Shark River during SharkTREx 1 and 2, respectively, and  $136.1 \pm$  320  $16.5 \text{ m}^2 \text{ s}^{-1}$  for Harney River during SharkTREx 2) were relatively large, and suggest that any perturbations (such as export of DIC from mangroves) would be quickly mixed thoroughly in the estuary.

In the following, for brevity, fluxes and inventories are summarized as ranges, which cover the two rivers and two experiments so they reflect both temporal and spatial variability. The individual values are given in Tables 1 and 2.

325 DIC was the dominant form of dissolved carbon in both rivers and accounted for 79 to 82% of the total dissolved carbon in the rivers. The contribution of DOC to the total carbon pool varied between 18 and 21% (Table 1).

### 3.2 Air-water CO<sub>2</sub> fluxes

As shown by Ho et al. (2014), pCO<sub>2</sub> observed during SharkTREx 1 and 2 fall in the upper range of those reported in other estuarine (Borges, 2005) and mangrove-dominated systems (Bouillon et al., 2003; Bouillon et al., 2007a; Bouillon et al., 2007b; Koné and Borges, 2008; Call et al., 2015). The mean air-water CO<sub>2</sub> fluxes in Shark River for SharkTREx 1 and 2 were 105 ± 9 and 99 ± 6 mmol m<sup>-2</sup> d<sup>-1</sup>, (Ho et al., 2016). The analysis is taken further here by including data from Harney River. The mean air-water CO<sub>2</sub> fluxes in Harney River were 150 ± 8 and 114 ± 21 mmol m<sup>-2</sup> d<sup>-1</sup> for SharkTREx 1 and 2, respectively.

Borges et al. (2003) summarized all available pCO<sub>2</sub> data from mangrove surrounding waters, and calculated CO<sub>2</sub> fluxes to the atmosphere that averaged 50 mmol m<sup>-2</sup> d<sup>-1</sup> (with a range of 4.6 to 113.5 mmol m<sup>-2</sup> d<sup>-1</sup>), and Bouillon et al. (2008a) estimated a global CO<sub>2</sub> flux from mangroves of ca. 60 ± 45 mmol m<sup>-2</sup> d<sup>-1</sup>. One reason that the fluxes from SharkTREx 1 and 2 are on the upper end of those estimates may be that the Shark and Harney Rivers receive a large input of DIC from the freshwater marsh upstream (Table 1), causing higher pCO<sub>2</sub> in the estuary compared to the global average.

Scaling the air-water CO<sub>2</sub> fluxes by the area of open water in the Shark and Harney Rivers, where Tarpon Bay is included with Shark River, suggests that the total carbon emissions to the atmosphere through air-water gas exchange in Shark River was 4.2 ± 0.4 × 10<sup>5</sup> and 4.0 ± 0.2 × 10<sup>5</sup> mol d<sup>-1</sup> during SharkTREx 1 and 2, respectively, and were 4.1 ± 0.2 × 10<sup>5</sup> and 3.1 ± 0.6 × 10<sup>5</sup> mol d<sup>-1</sup> from the Harney River during SharkTREx 1 and 2, respectively (Fig. 5), which is remarkably consistent, both spatially and temporally.

These fluxes were incorporated into the DIC mass balance of the Shark and Harney Rivers (Eq. 2) by calculating the total CO<sub>2</sub> degassed over the residence time of water in the rivers. Given the mean air-water CO<sub>2</sub> fluxes (Table 2), the total CO<sub>2</sub> degassed in the Shark River represents approximately 13 and 21% of  $\sum[\text{DIC}]_{\text{observed}}$  during SharkTREx 1 and 2, respectively, and the CO<sub>2</sub> degassed from the Harney River during SharkTREx 2 represents 20% of  $\sum[\text{DIC}]_{\text{observed}}$ , indicating that air-water CO<sub>2</sub> exchange removes a non-negligible fraction of the inorganic carbon in these rivers. Exclusion of  $\sum[\text{DIC}]_{\text{gasex}}$  from the mass balance in Eq. (2) would lead to an underestimation of  $\sum[\text{DIC}]_{\text{estuary}}$  of between 33 and 44%.

### 3.3 Mangrove contribution to DIC inventory

The highest DIC concentrations were correlated with low DO (Fig. 2) and characterized by  $^{13}\text{C}$ -depletion (Fig. 3j, k, l). Observations of elevated DIC and  $\text{pCO}_2$  in the middle of the estuary, coupled with  $\delta^{13}\text{C}_{\text{DIC}}$  and  $\text{O}_2$  depletion may indicate the importance, noted by other authors, of lateral transport of pore water from the peat-based mangrove forest into the river via tidal pumping (Bouillon et al., 2008a; Maher et al., 2013). However, as demonstrated below, the observed DIC and  $\delta^{13}\text{C}_{\text{DIC}}$  distributions in these rivers cannot be explained solely by mineralization of mangrove-derived organic carbon.

#### 3.3.1 Evidence from $\delta^{13}\text{C}_{\text{DIC}}$

The distributions of DIC and  $\delta^{13}\text{C}_{\text{DIC}}$  cannot be explained solely by the addition of mangrove-derived DIC and air-water gas exchange. Solving Eq. (8) for  $\delta^{13}\text{C}_{\text{DIC}}$ , assuming that  $[\text{DIC}]_{\text{dissolution}}$  is negligible and that the only source of DIC in the rivers is of mangrove origin, would result in  $\delta^{13}\text{C}$  values significantly lower than those observed. The low pH in interstitial waters of mangrove sediments due to organic matter mineralization processes may be favorable to  $\text{CaCO}_3$  dissolution in mangrove sediments, and this process could have an effect on estuarine  $\delta^{13}\text{C}_{\text{DIC}}$ . Groundwater discharge could also influence DIC and  $\delta^{13}\text{C}_{\text{DIC}}$ . Inputs of DIC derived from  $\text{CaCO}_3$  dissolution from either of these sources may explain the differences in observed  $\delta^{13}\text{C}_{\text{DIC}}$  and those expected if  $[\text{DIC}]_{\text{estuary}}$  was entirely of mangrove origin.

Other recent studies have observed similar differences in pore water  $\delta^{13}\text{C}_{\text{DIC}}$  values and the values expected from the carbon inputs derived from organic matter decomposition and  $\text{CaCO}_3$  dissolution (Walter et al., 2007). In these studies, the differences were attributed to isotopic exchange during  $\text{CaCO}_3$  dissolution and re-precipitation, and if these processes were active in the substrate or mangrove sediments surrounding the Shark and Harney Rivers, they would affect observed  $\delta^{13}\text{C}_{\text{DIC}}$  values without affecting  $[\text{DIC}]_{\text{estuary}}$ .

Solving Equations 6 and 8, the mineralization of mangrove-derived organic matter is estimated to account for ca.  $60 \pm 6\%$  of  $\Sigma[\text{DIC}]_{\text{estuary}}$  (Table 3), with the remainder originating from the dissolution of  $\text{CaCO}_3$ . This estimate is sensitive to the end member value chosen for  $\delta^{13}\text{C}_{\text{mangroves}}$  and  $\delta^{13}\text{C}_{\text{dissolution}}$ . For instance, if  $\delta^{13}\text{C}_{\text{mangroves}}$  were  $-29\%$  instead of  $-30\%$ , the mangrove contribution would increase to  $62\%$ .

### 3.3.2 Evidence from DIC and TALK

In the Shark and Harney Rivers, the high correlation ( $r^2 = 0.99$ ; Fig. 6) between  $[\text{DIC}]_{\text{estuary}}$  and  $[\text{TALK}]_{\text{estuary}}$  indicates the same processes control the inputs of DIC and TALK to these rivers. By examining the covariation of  $[\text{DIC}]_{\text{estuary}}$  and  $[\text{TALK}]_{\text{estuary}}$ , mangroves were found to contribute a minimum of  $70 \pm 3\%$  of  $\Sigma[\text{DIC}]_{\text{estuary}}$  (Table 3), with the remainder due to the dissolution of  $\text{CaCO}_3$ . These estimates are in reasonable agreement with those based on the carbon isotopic mass balance. One reason why the  $\delta^{13}\text{C}_{\text{DIC}}$ -based estimates might be lower is that the potential of  $^{13}\text{C}$  isotopic exchange during  $\text{CaCO}_3$  dissolution and re-precipitation is not being considered.

The  $[\text{TALK}]_{\text{estuary}}$  vs.  $[\text{DIC}]_{\text{estuary}}$  ratios were 0.84 and 0.92 for Shark River during SharkTREx 1 and 2, and 0.90 for the Harney River during SharkTREx 2 (Fig. 6). Given the estimated contribution of  $\text{CaCO}_3$  dissolution to  $\Sigma[\text{DIC}]_{\text{estuary}}$  of ca. 30%, sulfate reduction and aerobic respiration were estimated to contribute 29 to 34% and 36 to 41%, respectively.

### 3.3.3 Evidence from DO

The deficit of  $\text{O}_2$  in Shark River was found to be  $2.7 \pm 0.7 \times 10^6$  and  $3.7 \pm 0.3 \times 10^6$  mol during SharkTREx 1 and 2, respectively. Assuming a stoichiometric ratio of ca. 1.1 for  $\text{O}_2$  to  $\text{CO}_2$  during degradation/remineralization of terrestrial organic matter (Severinghaus, 1995; Keeling and Manning, 2014), the maximum contribution of aerobic respiration to the DIC added to the estuary was estimated to be 57 to 69%. However,  $\text{O}_2$  may also be consumed during oxidation of reduced products from anaerobic metabolism, such as  $\text{H}_2\text{S}$ ,  $\text{Mn}^{2+}$  or  $\text{Fe}^{2+}$ , with similar stoichiometry as aerobic respiration. Hence, the numbers derived above represent an upper limit for aerobic respiration, and if there were complete re-oxidation of metabolites from anaerobic respiration, the  $\text{O}_2$  deficit would represent total mineralization of terrestrial organic matter instead of just aerobic respiration. The mangrove contributions estimated from  $\delta^{13}\text{C}_{\text{DIC}}$  (section 3.3.1) and TALK/DIC (section 3.3.2) are consistent with this analysis of the  $\text{O}_2$  deficit, which indicates that a minimum of 57-69% of  $\Sigma[\text{DIC}]_{\text{estuary}}$  derived from the mineralization of organic matter.

## 3.4 Mangrove contributions to DOC inventory

During both experiments, the  $\delta^{13}\text{C}_{\text{DOC}}$  was highly depleted, indicative of contribution from higher plants, including mangroves. During SharkTREx 1, the lowest observed  $\delta^{13}\text{C}_{\text{DOC}}$  value ( $-33.8\text{‰}$ ) was in the mid-estuary (i.e., from salinity of ca. 10 to 20) (Fig. 4d), and it was lower than mangrove-derived material, indicative of a highly



405 reworked organic matter source, and perhaps preferential degradation of enriched compounds had resulted in further  $^{13}\text{C}$  depletion (Hayes, 1993). The overall  $\delta^{13}\text{C}_{\text{DOC}}$  depletion was less during SharkTREx 2, and the overall distribution was indicative of a stronger marine influence and/or mixing (Fig. 4e, f). The marine end member had a more enriched  $\delta^{13}\text{C}_{\text{DOC}}$ , perhaps suggesting also a greater contribution of seagrass and/or marine phytoplankton derived organic matter to the marine DOC pool (Anderson and Fourqurean, 2003). These observations are consistent  
410 with the greater longitudinal dispersion observed during SharkTREx 2 compared to SharkTREx 1.

The calculations of mangrove contribution using  $\delta^{13}\text{C}_{\text{DOC}}$  mass balances (Eq. 10) are in agreement with the above estimates, and suggest that the majority of  $[\text{DOC}]_{\text{estuary}}$ , but only a small percentage of the total DOC inventory, was derived from mangroves (7 and 5% in the Shark River during SharkTREx 1 and 2, and 7% in the Harney River during SharkTREx 2).

### 415 3.5 Longitudinal fluxes to the Gulf of Mexico and comparison with previous studies

~~Residence times of~~ Shark River (including Tarpon Bay) for SharkTREx 1 and 2 were,  $5.8 \pm 0.4$  and  $8.1 \pm 1.1$  days, respectively (Ho et al., 2016), and that of Harney River was  $4.7 \pm 0.7$  days for SharkTREx 2. The resulting longitudinal DIC fluxes to the Gulf of Mexico ( $15.8$  to  $33.6 \times 10^5 \text{ mol d}^{-1}$ ) were significantly larger than the longitudinal DOC fluxes ( $3.3$  to  $7.5 \times 10^5 \text{ mol d}^{-1}$ ) at salinity of ca. 27 (Fig. 5; Table 2).

420 There are no previously published DIC inventories or fluxes for the Shark and Harney Rivers, so comparison with previous studies is focused on the DOC results. The DOC flux from the Shark River to the coastal ocean in SharkTREx 1 ( $7.5 \pm 0.2 \times 10^5 \text{ mol d}^{-1}$ ) is in very good agreement to that estimated by Bergamaschi et al. (2011) in an experiment conducted in the Shark River from 20-30 September 2010 ( $7.6 \pm 0.5 \times 10^5 \text{ mol d}^{-1}$ ). However, the net discharge during the Bergamaschi et al. (2011) study was higher than SharkTREx 1 ( $9.1 \pm 7.1$  vs.  
425  $6.9 \pm 5.3 \text{ m}^3 \text{ s}^{-1}$ ), which would lead to a shorter residence time of 4.6 days using a relationship presented in Ho et al. (2016). Using the DOC concentration data presented in Bergamaschi et al. (2011) yields an inventory that is ca. 3% higher than the DOC inventory in Shark River during SharkTREx 1. Calculations using the shorter residence time and higher DOC inventory yields a DOC flux of  $9.7 \pm 0.2 \times 10^5 \text{ mol d}^{-1}$ , which is ca. 30% higher than the estimates of Bergamaschi et al. (2011). The difference between both methods may be due to the fact that Bergamaschi et al.  
430 (2011) estimated the longitudinal flux at a location ca. 9.5 km from the mouth of the Shark River. Because the DOC concentration gradients in the middle of the estuary are higher than at the mouth, this may overestimate the flux.

The longitudinal flux of mangrove-derived DOC from Shark River during SharkTREx 1 ( $0.3 \pm 0.2 \times 10^5$  mol d<sup>-1</sup>; Table 2) is in rough agreement with the estimate of Cawley et al. (2013) during the same period ( $0.2 \times 10^5$  mol d<sup>-1</sup>), but the value for Harney River ( $0.6 \pm 0.6 \times 10^5$  mol d<sup>-1</sup>) is lower than their estimate ( $1.6 \times 10^5$  mol d<sup>-1</sup>).

435 Mangroves contributed 4% to 6% of the total longitudinal DOC flux in the Shark River and 7% in the Harney River during SharkTREx 2 (Tables 1 and 4). These calculations are in agreement with those by Cawley et al. (2013), who estimated a mangrove contribution to DOC flux of  $3 \pm 10\%$  for Shark River and  $21 \pm 8\%$  for the Harney River during November 2010, the same time period as SharkTREx 1. DOC measurements were not made in Harney River as part of SharkTREx 1. However, using the November 2010 DOC data from Harney River collected  
440 by Cawley et al. (2013) for inventory calculations, a mangrove contribution of 19% to the total DOC longitudinal flux to the Gulf of Mexico was obtained.

### 3.6 Distribution of carbon fluxes

During SharkTREx 1 and 2,  $\Sigma[\text{DIC}]_{\text{estuary}}$  made up 20-28% of the total DIC in the rivers, and  $\Sigma[\text{DOC}]_{\text{estuary}}$  made up only 4 to 7% of the total DOC in the rivers. Mangroves are estimated to contribute 13 to  
445 19% to the total DIC inventory. In all cases, the mangrove contribution to the DIC inventory is a factor of 3 greater than the mangrove contribution to the DOC inventory (Table 1). During SharkTREx 1 and 2, the inventory of mangrove-derived DIC exceeded that of DOC by a factor of 15 to 17, which supports the idea that a large fraction of the carbon exported by mangroves to surrounding water is as DIC (Bouillon et al., 2008a), but is considerably larger than the estimates of ca. 3 to 10 compiled by Bouillon et al. (2008a) for mangroves at 5 sites in Asia and Africa.

450 The total dissolved carbon fluxes from all sources (i.e., freshwater wetland, mangrove, carbonate dissolution, and marine input) out of the Shark and Harney Rivers during SharkTREx 1 and 2 are dominated by inorganic carbon (82-83%; see Tables 2 and 4), either via air-water CO<sub>2</sub> exchange or longitudinal flux of DIC to the coastal ocean (Fig. 5). The remaining 17-18% of the export is as DOC. This proportioning is remarkably similar between SharkTREx 1 and 2, and between the Shark and Harney Rivers (Table 1). The estuarine contribution to  
455 these fluxes is relatively small (generally <15%), with the exception of air-water CO<sub>2</sub> flux, where the estuary contribution was 49 to 63% (Table 4).

In this study, the particulate organic carbon (POC) flux was not examined. However, He et al. (2014) estimated the mangrove-derived POC flux in Shark River by taking the total volume discharge from the five major

460 rivers along the southwest coast of Everglades National Park from 2004 to 2008, and assuming that Shark River contributed 14% to the mean annual discharge. They then multiplied this discharge by the average POM concentration ( $5.20 \pm 0.614 \text{ mg L}^{-1}$ ) in the middle of the estuary to yield an annual POM flux from Shark River. Based on analysis of organic matter biomarkers, He et al. (2014) estimated that mangrove-derived POM was 70–90% of the total POM pool in the Shark River. Using this contribution and further assuming that 58% of POM weight is POC (Howard, 1965), they estimated a POC flux of  $1.0 \text{ to } 2.2 \times 10^4 \text{ mol d}^{-1}$ . Because this estimate was 465 based on biomarker and POM data from the mid-estuary, where the POM concentration and the mangrove contribution to POM are both likely to be much higher than either toward the freshwater end member or the marine end member, it is likely an overestimate of the mangrove derived POC flux. Nevertheless, the mangrove-derived POC flux determined by He et al. (2014) is still only a small fraction (3 to 7%) of the mangrove-derived dissolved carbon fluxes in Shark River during SharkTREx 1 and 2.

### 470 3.7 Mangrove contributing area and estuary carbon balance

One of the challenges of relating the results reported here to other studies is to scale the results to a mangrove contributing area, and thereby relate the findings to mangrove forest carbon balance, ~~typically~~ expressed on an aerial basis. Estimates of forest carbon export derived here are compared with other investigations in this estuary. The entire area of mangroves surrounding the Shark and Harney Rivers region is ca.  $111 \text{ km}^2$ , and the water 475 area is ca.  $17.5 \text{ km}^2$  (Ho et al., 2014). Scaling the forest area by the water area of Shark River ( $2.5 \text{ km}^2$ ) yields an associated forest area of  $15.9 \text{ km}^2$ . The forest area associated with Harney River ( $2.8 \text{ km}^2$ ) is  $17.4 \text{ km}^2$ .

Using the total forest area associated with Shark River to scale estimates of total export of mangrove-derived carbon (the combination of longitudinal fluxes and air-water gas exchange) suggests an average dissolved carbon lateral export rate from the forest of  $18.9 \text{ to } 24.5 \text{ mmol m}^{-2} \text{ d}^{-1}$ , including both DIC and DOC. However, 480 since ~~it is unknown what~~ fraction of the total forest area associated with these rivers exported dissolved carbon through tidal pumping (a function of tidal height and duration), this is considered to be a minimum estimate. Average water levels at high tide during SharkTREx 1 and 2 at the USGS Shark River station were 88% and 95% of maximum wet season water levels reported at this site over the period from November 2007 to December 2012 (U.S. Geological Survey, 2016), and 12 inundation events occurred during both SharkTREx 1 and 2. Water levels in 485 the main river channel at the USGS Shark River station were above an estimate of the average minimum ground surface elevation derived from nearby groundwater monitoring wells in the estuary (sites SH3 and SH4;

<http://sofia.usgs.gov/eden/stationlist.php>) for 21% and 28% of the time during the SharkTREx 1 and 2 experimental periods, respectively. These values indicate the export of dissolved carbon from flooded portions of the forest during the discontinuous inundation periods should be significantly greater than the dissolved carbon lateral export rate derived above in order to produce the observed inventories of mangrove-derived dissolved carbon in the main channel.

Bergamaschi et al. (2011) proposed an annual total DOC export from the forest surrounding Shark River of  $15.1 \pm 1.1 \text{ mol m}^{-2} \text{ y}^{-1}$  and describe their method of calculating contributing area using a model based on the relationship between discharge volume and changes in water levels during tidal cycles. They do not provide a contributing area, but this can be calculated from their results. They determined longitudinal DOC fluxes of  $7.6 \pm 0.5 \times 10^5$  and  $1.3 \pm 0.02 \times 10^5 \text{ mol d}^{-1}$  for the wet and dry seasons, respectively, and assumed that they are entirely of mangrove origin. Given the lengths of the wet and dry seasons, this would yield a mean annual DOC flux of  $3.9 \pm 0.2 \times 10^5 \text{ mol d}^{-1}$ , and  $9.4 \pm 0.7 \text{ km}^2$  of mangrove forest contributing to carbon fluxes thru tidal flushing in this segment of Shark River. However, data from SharkTREx 1 and 2 indicate that ca. 5% of the total longitudinal DOC fluxes were of mangrove origin, with an average mangrove-derived DIC to DOC flux ratio of 10.5. Using this information, the Bergamaschi et al. (2011) results were recalculated to yield a wet season dissolved carbon lateral export rate of  $46.5 \pm 4.4 \text{ mmol m}^{-2} \text{ d}^{-1}$  (as DIC and DOC) from the forest.

Another method of estimating forest lateral carbon export utilizes the difference between measurements of net ecosystem-atmosphere  $\text{CO}_2$  exchange (NEE) above the mangrove forest surrounding Shark River ( $267 \pm 15 \text{ mmol m}^{-2} \text{ y}^{-1}$  in 2004; (Barr et al., 2012) and corresponding measures of net ecosystem carbon balance (NECB;  $227 \pm 14 \text{ mmol m}^{-2} \text{ d}^{-1}$ ). NECB in 2004 can be estimated as the sum of carbon in litter fall ( $104 \pm 8 \text{ mmol m}^{-2} \text{ d}^{-1}$ ), wood production ( $44 \pm 3 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) (Castañeda-Moya et al., 2013), root growth ( $47 \pm 11 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) (Castañeda-Moya et al., 2011) and soil carbon accumulation ( $31.7 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) (Breithaupt et al., 2014) measured at the same location (FCE LTER site SRS6) in this forest. The difference between NEE and NECB ( $40 \pm 17 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) provides an estimate of the annual rate of forest carbon export to Shark River on a daily basis (Chapin et al., 2006).

The rate of mangrove-derived carbon exported to estuarine waters is likely to vary over space and time, as a result of factors that include tidal cycles, phenology, and forest and soil structural characteristics. For example, Bergamaschi et al. (2011) found that DOC fluxes were 6 times higher during the wet season (September) than the dry season (April), whereas Cawley et al. (2013) found that the DOC fluxes were 4 and 10 times higher during the

515 wet vs. dry season (November vs. March) in the Shark and Harney Rivers, respectively. Barr et al. (2013) showed  
that forest respiration rates derived from NEE data are greater during the wet than dry seasons. Higher respiration  
rates combined with increased inundation during the wet compared to dry seasons suggest that wet season DIC  
export will also be greater than dry season values. For these reasons, the annual carbon export rates derived from the  
difference between NECB and NEE are expect to underestimate wet season values. If annual lateral carbon export  
520 rates are considered as equivalent to a time-weighted sum of dry season (7 months) and wet season (5 months)  
values (after Bergamaschi et al. 2011), and wet season export is assumed to be, for example, 5 times greater than dry  
season values, the seasonal export rates (15 and 75 mmol m<sup>-2</sup> d<sup>-1</sup> for dry and wet seasons, respectively) that  
correspond with the difference between annual NECB and NEE can be calculated.

The discrepancies between the estimates of carbon export rates derived here, and those derived from  
525 Bergamaschi et al. (2011) and the difference between NEE and NECB point out the need for additional studies to  
reduce the uncertainty in the relationships between riverine carbon fluxes, forest carbon export, and estimates of  
contributing areas. For example, Bergamaschi et al. (2011) conducted an Eulerian study at a single location in the  
middle of the estuary, where the mangrove influence might be higher than the Lagrangian study conducted during  
SharkTREx 1 and 2, which covered the entire estuary. Also, the estimate of forest carbon export based on the  
530 difference between NEE and NECB is from a single location along Shark River (at FCE LTER site SRS6), and may  
not be representative of the entire forest. Furthermore, forest lateral carbon export rates and contributing areas  
should be considered dynamic, varying over semi-diurnal time scales with the extent and duration of inundation  
during individual tidal cycles. The correct interpretation of a single, static value for contributing area such as derived  
above is therefore uncertain, since the tracer-based results represent an integration of carbon sources and sinks  
535 calculated over the water residence time and expressed on daily time scales. To improve understanding of how  
mangrove forest carbon balance and export influence riverine carbon inventories and fluxes to the Gulf of Mexico in  
this system, wet and dry season measurements over multiple years, information on the relationships between forest  
structure, productivity and lateral carbon export rates, and independent estimates of forest inundation area in relation  
to tidal height are needed.

## 540 4 Conclusions

The SharkTREx 1 and 2 studies are the first to provide estimates of longitudinal DIC export, air-water CO<sub>2</sub> fluxes, and mangrove-derived DIC inputs for the Shark and Harney Rivers. The results show that air-water CO<sub>2</sub> exchange and longitudinal DIC fluxes account for ca. 90% of the mangrove-derived dissolved carbon export out of the Shark and Harney Rivers, with the remainder being exported as dissolved organic carbon.

545 The mangrove contribution to the total longitudinal flux was 6.5 to 8.9% for DIC and 4 to 18% for DOC. A lower bound estimate of the dissolved carbon export (DIC and DOC) from the forest surrounding Shark River during the wet season was 18.9 to 24.5 mmol m<sup>-2</sup> d<sup>-1</sup> with 15.9 km<sup>2</sup> of mangrove contributing area. Other independent estimates of lateral export from this mangrove forest are somewhat higher by comparison. However, mangrove forest carbon export rates on an aerial basis are expected to vary with the spatial and temporal scales over  
550 which they are calculated, and depend on factors such as tidal inundation frequency, distance from the riverbank and the coast, and forest and soil characteristics.

Future experiments should investigate the contribution of DIC from groundwater to the rivers, by making measurements of  $\delta^{13}\text{C}_{\text{DIC}}$  of groundwater, Sr and Ca concentrations in the river to quantify CaCO<sub>3</sub> dissolution and to separate carbonate alkalinity from TAlk, radon to quantify groundwater discharge,  $^{14}\text{C}_{\text{DIC}}$  to separate input of DIC  
555 from remineralization of organic matter from dissolution of CaCO<sub>3</sub>. Experiments should also examine the seasonal variability in the carbon dynamics and export, by conducting process-based studies like SharkTREx during both wet and dry seasons. Also, time series measurement of current velocities, wind speeds, pCO<sub>2</sub> and pH (to calculate DIC), DO, chromophoric dissolved organic matter (CDOM, as a proxy for DOC), and radon will also allow the temporal variability of the sources and sinks of DIC in these rivers to be examined.

## 560 Author contribution

D. Ho, S. Ferron, and V. Engel conceived and executed the experiment, interpreted the data, and prepared the manuscript with input from the other authors. W. Anderson measured the samples for  $\delta^{13}\text{C}$  of dissolved organic carbon, P. Swart measured the samples for  $\delta^{13}\text{C}$  of dissolved inorganic carbon, R. Price measured the total alkalinity samples for SharkTREx 1, and L. Barbero measured the total alkalinity and dissolved inorganic carbon samples for  
565 SharkTREx 2.

## Data availability

The pCO<sub>2</sub> data collected during SharkTREx 1 and 2 are available from the SOCAT database <[www.socat.info](http://www.socat.info)>.

The other data may be obtained by contacting the corresponding author.

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580 imply endorsement by the U.S. Government.

## References

- Alongi, D. M.: Coastal Ecosystem Processes, CRC Marine Science Series, edited by: Kennish, M. J., and Lutz, P. L., CRC press, Boca Raton, 448 pp., 1998.
- Alongi, D. M., Pfitzner, J., Trott, L. A., Tirendi, F., Dixon, P., and Klumpp, D. W.: Rapid sediment accumulation and microbial mineralization in forests of the mangrove *Kandelia candel* in the Jiulongjiang Estuary, China, *Estuar. Coast. Shelf Sci.*, 63, 605-618, 10.1016/j.ecss.2005.01.004, 2005.
- Anderson, W. T., and Fourqurean, J. W.: Intra- and interannual variability in seagrass carbon and nitrogen stable isotopes from south Florida, a preliminary study, *Organic Geochemistry*, 34, 185-194, 10.1016/S0146-6380(02)00161-4, 2003.
- 585 Barr, J. G., Engel, V., Smith, T. J., and Fuentes, J. D.: Hurricane disturbance and recovery of energy balance, CO<sub>2</sub> fluxes and canopy structure in a mangrove forest of the Florida Everglades, *Agric. For. Meteorol.*, 153, 54-66, 10.1016/j.agrformet.2011.07.022, 2012.
- Barr, J. G., Fuentes, J. D., DeLonge, M. S., O'Halloran, T. L., Barr, D., and Zieman, J. C.: Summertime influences of tidal energy advection on the surface energy balance in a mangrove forest, *Biogeosciences*, 10, 501-511, 595 Doi 10.5194/Bg-10-501-2013, 2013.
- Bergamaschi, B. A., Krabbenhoft, D. P., Aiken, G. R., Patino, E., Rumbold, D. G., and Orem, W. H.: Tidally Driven Export of Dissolved Organic Carbon, Total Mercury, and Methylmercury from a Mangrove-Dominated Estuary, *Environ. Sci. Technol.*, 46, 1371-1378, 10.1021/es2029137, 2011.
- Borges, A. V., Djenidi, S., Lacroix, G., Théate, J., Delille, B., and Frankignoulle, M.: Atmospheric CO<sub>2</sub> flux from mangrove surrounding waters, *Geophys. Res. Lett.*, 30, 1558, 10.1029/2003GL017143, 2003.
- 600 Borges, A. V.: Do we have enough pieces of the jigsaw to integrate CO<sub>2</sub> fluxes in the coastal ocean?, *Estuaries*, 28, 3-27, 10.1007/BF02732750, 2005.
- Bouillon, S., Frankignoulle, M., Dehairs, F., Velimirov, B., Eiler, A., Abril, G., Etcheber, H., and Borges, A. V.: Inorganic and organic carbon biogeochemistry in the Gautami Godovari estuary (Andhra Pradesh, India) during pre-monsoon: The local impacts of extensive mangrove forests, *Glob. Biogeochem. Cycle*, 17, 605 10.1029/2002GB002026, 2003.
- Bouillon, S., Dehairs, F., Schiettecatte, L.-S., and Borges, A. V.: Biogeochemistry of the Tana estuary and delta (northern Kenya), *Limnol. Oceanogr.*, 52, 46-59, 10.4319/lo.2007.52.1.0046, 2007a.



- Bouillon, S., Dehairs, F., Velimirov, B., Abril, G., and Borges, A. V.: Dynamics of organic and inorganic carbon  
610 across contiguous mangrove and seagrass systems (Gazy Bay, Kenya), *J. Geophys. Res.*, 112,  
10.1029/2006JG000325, 2007b.
- Bouillon, S., Middelburg, J. J., Dehairs, F., Borges, A. V., Abril, G., Flindt, M. R., Ulomi, S., and Kristensen, E.:  
Importance of intertidal sediment processes and porewater exchange on the water column biogeochemistry  
in a pristine mangrove creek (Ras Dege, Tanzania), *Biogeosciences*, 4, 311-322, 10.5194/bg-4-311-2007,  
615 2007c.
- Bouillon, S., Borges, A. V., Castañeda-Moya, E., Diele, K., Dittmar, T., Duke, N. C., Kristensen, E., Lee, S. Y.,  
Marchand, C., Middelburg, J. J., Rivera-Monroy, V. H., Smith, T. J., and Twilley, R. R.: Mangrove  
production and carbon sinks: A revision of global budget estimates, *Glob. Biogeochem. Cycle*, 22,  
GB2013, 10.1029/2007GB003052, 2008a.
- 620 Bouillon, S., Connolly, R. M., and Lee, S. Y.: Organic matter exchange and cycling in mangrove ecosystems:  
Recent insights from stable isotope studies, *Journal of Sea Research*, 59, 44-58,  
10.1016/j.seares.2007.05.001, 2008b.
- Boyer, J. N., Fourqurean, J. W., and Jones, R. D.: Spatial characterization of water quality in Florida Bay and  
Whitewater Bay by multivariate analyses: Zones of similar influence, *Estuaries*, 20, 743-758,  
625 10.2307/1352248, 1997.
- Breithaupt, J. L., Smoak, J. M., Smith, T. J., and Sanders, C. J.: Temporal variability of carbon and nutrient burial,  
sediment accretion, and mass accumulation over the past century in a carbonate platform mangrove forest  
of the Florida Everglades, *J. Geophys. Res.*, 119, 2032-2048, 10.1002/2014JG002715, 2014.
- Call, M., Maher, D. T., Santos, I. R., Ruiz-Halpern, S., Mangion, P., Sanders, C. J., Erler, D. V., Oakes, J. M.,  
630 Rosentreter, J., Murray, R., and Eyre, B. D.: Spatial and temporal variability of carbon dioxide and methane  
fluxes over semi-diurnal and spring–neap–spring timescales in a mangrove creek, *Geochim. Cosmochim.  
Acta*, 150, 211-225, 10.1016/j.gca.2014.11.023, 2015.
- Castañeda-Moya, E., Twilley, R. R., Rivera-Monroy, V. H., Marx, B. D., Coronado-Molina, C., and Ewe, S. M. L.:  
Patterns of Root Dynamics in Mangrove Forests Along Environmental Gradients in the Florida Coastal  
635 Everglades, USA, *Ecosystems*, 14, 1178-1195, 10.1007/s10021-011-9473-3, 2011.

- Castañeda-Moya, E., Twilley, R. R., and Rivera-Monroy, V. H.: Allocation of biomass and net primary productivity of mangrove forests along environmental gradients in the Florida Coastal Everglades, USA, *Forest Ecol Manag*, 307, 226-241, 10.1016/j.foreco.2013.07.011, 2013.
- 640 Cawley, K. M., Yamashita, Y., Maie, N., and Jaffé, R.: Using Optical Properties to Quantify Fringe Mangrove Inputs to the Dissolved Organic Matter (DOM) Pool in a Subtropical Estuary, *Estuar Coast*, 37, 399-410, 10.1007/s12237-013-9681-5, 2013.
- Chapin, F. S., Woodwell, G. M., Randerson, J. T., Rastetter, E. B., Lovett, G. M., Baldocchi, D. D., Clark, D. A., Harmon, M. E., Schimel, D. S., Valentini, R., Wirth, C., Aber, J. D., Cole, J. J., Goulden, M. L., Harden, J. W., Heimann, M., Howarth, R. W., Matson, P. A., McGuire, A. D., Melillo, J. M., Mooney, H. A., Neff, J. 645 C., Houghton, R. A., Pace, M. L., Ryan, M. G., Running, S. W., Sala, O. E., Schlesinger, W. H., and Schulze, E.-D.: Reconciling Carbon-cycle Concepts, Terminology, and Methods, *Ecosystems*, 9, 1041-1050, 10.1007/s10021-005-0105-7, 2006.
- DeNiro, M. J., and Epstein, S.: Influence of diet on the distribution of carbon isotopes in animals, *Geochim. Cosmochim. Acta*, 42, 495-506, 10.1016/0016-7037(78)90199-0, 1978.
- 650 Dickson, A. G.: Standards for ocean measurements, *Oceanography*, 23, 34-47, 10.5670/oceanog.2010.22, 2010.
- Dittmar, T., and Lara, R. J.: Do mangroves rather than rivers provide nutrients to coastal environments south of the Amazon River? Evidence from long-term flux measurements, *Marine Ecology Progress Series*, 213, 67-77, 2001.
- Dittmar, T., Hertkorn, N., Kattner, G., and Lara, R. J.: Mangroves, a major source of dissolved organic carbon to the 655 oceans, *Glob. Biogeochem. Cycle*, 20, GB1012, 10.1029/2005GB002570, 2006.
- Fischer, H. B., List, E. J., Imberger, J., Koh, R. C. Y., and Brooks, N. H.: *Mixing in inland and coastal waters*, Academic Press, New York, 483 pp., 1979.
- Garcia, H. E., and Gordon, L. I.: Oxygen solubility in seawater: better fitting equations, *Limnol. Oceanogr.*, 37, 1307-1312, 10.4319/lo.1992.37.6.1307, 1992.
- 660 Hayes, J. M.: Factors controlling  $^{13}\text{C}$  contents of sedimentary organic compounds: Principles and evidence, *Mar. Geol.*, 113, 111-125, 1993.

- He, D., Mead, R. N., Belicka, L., Pisani, O., and Jaffé, R.: Assessing source contributions to particulate organic matter in a subtropical estuary: A biomarker approach, *Organic Geochemistry*, 75, 129-139, 10.1016/j.orggeochem.2014.06.012, 2014.
- 665 Ho, D. T., Schlosser, P., and Caplow, T.: Determination of longitudinal dispersion coefficient and net advection in the tidal Hudson River with a large-scale, high resolution SF<sub>6</sub> tracer release experiment, *Environ. Sci. Technol.*, 36, 3234-3241, 10.1021/es015814+, 2002.
- Ho, D. T., Ferrón, S., Engel, V. C., Larsen, L. G., and Barr, J. G.: Air-water gas exchange and CO<sub>2</sub> flux in a mangrove-dominated estuary, *Geophys. Res. Lett.*, 41, 108–113, 10.1002/2013GL058785, 2014.
- 670 Ho, D. T., Coffineau, N., Hickman, B., Chow, N., Koffman, T., and Schlosser, P.: Influence of current velocity and wind speed on air-water gas exchange in a mangrove estuary, *Geophys. Res. Lett.*, 43, 10.1002/2016GL068727, 2016.
- Howard, P. J. A.: The Carbon-Organic Matter Factor in Various Soil Types, *Oikos*, 15, 229-236, 10.2307/3565121, 1965.
- 675 Jaffé, R., Mead, R., Hernandez, M. E., Peralba, M. C., and DiGuida, O. A.: Origin and transport of sedimentary organic matter in two subtropical estuaries: a comparative, biomarker-based study, *Org. Geoch.*, 32, 507-523, 10.1016/S0146-6380(00)00192-3, 2001.
- Jähne, B., Munnich, K. O., Bosinger, R., Dutzi, A., Huber, W., and Libner, P.: On the parameters influencing air-water gas exchange, *J. Geophys. Res.*, 92, 1937-1949, 10.1029/JC092iC02p01937, 1987.
- 680 Jennerjahn, T. C., and Ittekkot, V.: Relevance of mangroves for the production and deposition of organic matter along tropical continental margins, *Naturwissenschaften*, 89, 23-30, 10.1007/s00114-001-0283-x, 2002.
- Keeling, R. F., and Manning, A. C.: Studies of Recent Changes in Atmospheric O<sub>2</sub> Content, in: *Treatise on Geochemistry (Second Edition)*, Second ed., edited by: Holland, H. D., and Turekian, K. K., Elsevier, Oxford, 385-404, 2014.
- 685 Koné, Y. J.-M., and Borges, A. V.: Dissolved inorganic carbon dynamics in the waters surrounding forested mangroves of the Ca Mau Province (Vietnam), *Estuar. Coast. Shelf Sci.*, 77, 409-421, 10.1016/J.Ecss.2007.10.001, 2008.
- Kristensen, E., Bouillon, S., Dittmar, T., and Marchand, C.: Organic carbon dynamics in mangrove ecosystems: A review, *Aquatic Botany*, 89, 201-219, 10.1016/j.aquabot.2007.12.005, 2008.

- 690 Levesque, V. A.: Water Flow and Nutrient Flux from Five Estuarine Rivers along the Southwest Coast of the  
Everglades National Park, Florida, 1997-2001: U.S. Geological Survey Scientific Investigations Report  
2004-5142, U.S. Geological Survey Scientific Investigations Report 2004-5142, 24, 2004.
- Maher, D. T., Santos, I. R., Golsby-Smith, L., Gleeson, J., and Eyre, B. D.: Groundwater-derived dissolved  
inorganic and organic carbon exports from a mangrove tidal creek: The missing mangrove carbon sink?,  
695 *Limnol. Oceanogr.*, 58, 475-488, 10.4319/lo.2013.58.2.0475, 2013.
- Mancera-Pineda, J. E., Twilley, R. R., and Rivera-Monroy, V. H.: Carbon ( $\delta^{13}\text{C}$ ) and nitrogen ( $\delta^{15}\text{N}$ ) isotopic  
discrimination in mangroves in Florida Coastal Everglades as a function of environmental stress,  
*Contributions in Marine Science*, 38, 109-129, 2009.
- Millero, F. J.: Carbonate constants for estuarine waters, *Mar Freshwater Res*, 61, 139-142, 10.1071/Mf09254, 2010.
- 700 Miyajima, T., Tsuboi, Y., Tanaka, Y., and Koike, I.: Export of inorganic carbon from two Southeast Asian  
mangrove forests to adjacent estuaries as estimated by the stable isotope composition of dissolved  
inorganic carbon, *J. Geophys. Res.*, 114, G01024, 10.1029/2008JG000861, 2009.
- Mook, W. G., and Tan, T. C.: Stable carbon isotopes in rivers and estuaries, in: *Biogeochemistry of the major world  
rivers*, SCOPE Report 42, edited by: Degens, E. T., Kempe, S., and Richey, J. E., John Wiley & Sons,  
705 Chichester, 245-264, 1991.
- Pierrot, D., Lewis, E., and Wallace, D. W. R.: MS Excel Program Developed for CO<sub>2</sub> System Calculations.  
ORNL/CDIAC-105a. Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, U.S.  
Department of Energy, Oak Ridge, Tennessee, 10.3334/CDIAC/otg.CO2SYS\_XLS\_CDIAC105a, 2006.
- Price, R. M., Swart, P. K., and Fourqurean, J. W.: Coastal groundwater discharge – an additional source of  
710 phosphorus for the oligotrophic wetlands of the Everglades, *Hydrobiologia*, 569, 23-36, 10.1007/s10750-  
006-0120-5, 2006.
- Rivera-Monroy, V. H., Twilley, R. R., Davis III, S. E., Childers, D. L., Simard, M., Chambers, R., Jaffé, R., Boyer,  
J. N., Rudnick, D. T., Zhang, K., Castañeda-Moya, E., Ewe, S. M. L., Price, R. M., Coronado-Molina, C.,  
Ross, M., Smith III, T. J., Michot, B., Meselhe, E., Nuttle, W., Troxler, T. G., and Noe, G. B.: The role of  
715 the Everglades mangrove ecotone region (EMER) in regulating nutrient cycling and wetland productivity in  
South Florida, *Critical Reviews in Environmental Science and Technology*, 41, 663-669,  
10.1080/10643389.2010.530907, 2011.

- Rutherford, J. C.: River mixing, John Wiley & Sons, New York, 347 pp., 1994.
- 720 Saha, A., Moses, C., Price, R., Engel, V., Smith, T., III, and Anderson, G.: A Hydrological Budget (2002–2008) for  
a Large Subtropical Wetland Ecosystem Indicates Marine Groundwater Discharge Accompanies  
Diminished Freshwater Flow, *Estuar Coast*, 35, 459-474, 10.1007/s12237-011-9454-y, 2012.
- Salomons, W., and Mook, W. G.: Isotope Geochemistry of Carbonates in the Weathering Zone, in: *Handbook of  
Environmental Isotope Geochemistry, Vol 2, The Terrestrial Environment, B*, edited by: Fritz, P., and  
Fontes, J. C., Elsevier, Amsterdam, 239-269, 1986.
- 725 Severinghaus, J. P.: Studies of the terrestrial O<sub>2</sub> and carbon cycle in sand dune gases and in Biosphere 2, Ph.D.  
thesis, Columbia University, New York, 50 pp., 1995.
- Troxler, T. G., Barr, J. G., Fuentes, J. D., Engel, V., Anderson, G., Sanchez, C., Lagomasino, D., Price, R., and  
Davis, S. E.: Component-specific dynamics of riverine mangrove CO<sub>2</sub> efflux in the Florida coastal  
Everglades, *Agric. For. Meteorol.*, 213, 273-282, 10.1016/j.agrformet.2014.12.012, 2015.
- 730 Twilley, R. R., Chen, R. H., and Hargis, T.: Carbon sinks in mangroves and their implications to carbon budget of  
tropical coastal ecosystems, *Water, Air and Soil Pollution*, 64, 265-288, 10.1007/BF00477106, 1992.
- U.S. Geological Survey: National Water Information System data available on the World Wide Web (USGS Water  
Data for the Nation), accessed February 10, 2016, at url <http://waterdata.usgs.gov/nwis/>. 2016.
- 735 Walter, L. M., Ku, T. C. W., Muehlenbachs, K., Patterson, W. P., and Bonnell, L.: Controls on the  $\delta^{13}\text{C}$  of dissolved  
inorganic carbon in marine pore waters: An integrated case study of isotope exchange during  
syndepositional recrystallization of biogenic carbonate sediments (South Florida Platform, USA), *Deep-Sea  
Research II*, 54, 1163-1200, 10.1016/j.dsr2.2007.04.014, 2007.
- Wanninkhof, R.: Relationship between wind speed and gas exchange over the ocean revisited, *Limnol. Oceanogr.  
Methods*, 12, 351-362, 10.4319/lom.2014.12.351, 2014.
- 740 Weiss, R. F.: Carbon dioxide in water and seawater: the solubility of a non-ideal gas, *Mar. Chem.*, 2, 203-215,  
10.1016/0304-4203(74)90015-2, 1974.
- Ya, C., Anderson, W., and Jaffé, R.: Assessing dissolved organic matter dynamics and source strengths in a  
subtropical estuary: Application of stable carbon isotopes and optical properties, *Continental Shelf  
Research*, 92, 98-107, 10.1016/j.csr.2014.10.005, 2015.

745 Zhang, J., Quay, P. D., and Wilbur, D. O.: Carbon isotope fractionation during gas-water exchange and dissolution of CO<sub>2</sub>, *Geochim. Cosmochim. Acta*, 59, 107-114, 10.1016/0016-7037(95)91550-D, 1995.

**Table 1.** Inventories of DIC and DOC in Shark and Harney Rivers, as well as contributions from estuarine and non-estuarine sources.

		SharkTREx 1				SharkTREx 2							
		Shark River		Shark River		Shark River		Harney River		Harney River		Harney River	
		Inventory (x10 <sup>6</sup> mol)	Percentage of total Inventory <sup>a</sup>	Contribution to mangrove carbon <sup>b</sup>	Contribution to total carbon <sup>c</sup>	Inventory (x 10 <sup>6</sup> mol)	Percentage of total Inventory <sup>a</sup>	Contribution to mangrove carbon <sup>b</sup>	Contribution to total carbon <sup>c</sup>	Inventory (x 10 <sup>6</sup> mol)	Percentage of total Inventory <sup>a</sup>	Contribution to mangrove carbon <sup>b</sup>	Contribution to total carbon <sup>c</sup>
DIC	Observed	19.5 ± 0.9	-			15.2 ± 1.3	-			7.4 ± 0.4	-		
	Gas Exchange	2.5 ± 0.2	11%			3.2 ± 0.1	17%			1.5 ± 0.3	17%		
	Non-estuarine	17.6 ± 0.7	80%	94%	82%	13.6 ± 1.2	74%	93%	79%	6.5 ± 0.4	72%	93%	79%
	Estuarine	4.4 ± 1.1	20%			4.8 ± 1.8	26%			2.5 ± 0.6	28%		
	Mangrove	2.9 ± 0.8	13%			3.1 ± 1.2	17%			1.7 ± 0.4	19%		
DOC	Observed	4.4 ± 0.1				4.1 ± 0.4				1.9 ± 0.1			
	Non-estuarine	4.2 ± 0.1	96%	6%	18%	3.9 ± 0.4	94%	7%	21%	1.8 ± 0.1	93%	7%	21%
	Estuarine <sup>d</sup>	0.2 ± 0.1	4%			0.2 ± 0.6	6%			0.1 ± 0.2	7%		

<sup>a</sup> The DIC inventory is relative to the total DIC (i.e.,  $\sum[\text{DIC}]_{\text{observed}} + \sum[\text{DIC}]_{\text{gaseous}}$ ).

750 <sup>b</sup> Contribution of each form of carbon (i.e., DIC, DOC) to the total mangrove-derived carbon pool.

<sup>c</sup> Contribution of each form of carbon (i.e., DIC, DOC) to the total carbon pool.

<sup>d</sup> Estuarine DOC is assumed to be entirely of mangrove origin.

**Table 2.** Longitudinal DIC and DOC fluxes, and air-water CO<sub>2</sub> fluxes for the Shark and Harney Rivers during SharkTREx 1 and 2.

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	SharkTREx 1		SharkTREx 2	
	Shark River	Harney River	Shark River	Harney River
<i>Longitudinal DIC Fluxes (x 10<sup>5</sup> mol d<sup>-1</sup>)</i>				
Total	33.6 ± 1.6	N/A	18.8 ± 1.6	15.8 ± 0.9
Non-estuarine contribution	30.3 ± 1.1	N/A	16.8 ± 1.5	13.7 ± 0.8
Estuarine Contribution	3.3 ± 1.9	N/A	2.0 ± 2.2	2.1 ± 1.3
Mangrove Contribution	2.2 ± 1.3	N/A	1.3 ± 1.5	1.4 ± 0.8
<i>Air-Water CO<sub>2</sub> Fluxes (x 10<sup>5</sup> mol d<sup>-1</sup>)</i>				
Total	4.2 ± 0.4	4.1 ± 0.2	4.0 ± 0.2	3.1 ± 0.6
Non-estuarine contribution	2.1 ± 0.2	2.0 ± 0.1	1.9 ± 0.1	1.1 ± 0.2
Estuarine Contribution	2.1 ± 0.4	2.1 ± 0.3	2.1 ± 0.2	2.0 ± 0.6
Mangrove Contribution	1.4 ± 0.3	1.4 ± 0.2	1.4 ± 0.1	1.3 ± 0.4
<i>Longitudinal DOC Fluxes (x 10<sup>5</sup> mol d<sup>-1</sup>)<sup>a</sup></i>				
Total	7.5 ± 0.2	3.3 ± 0.4	5.1 ± 0.5	4.2 ± 0.2
Non-estuarine contribution	7.2 ± 0.1	2.6 ± 0.3	4.8 ± 0.5	3.9 ± 0.2
Estuarine Contribution <sup>b</sup>	0.3 ± 0.2	0.6 ± 0.6	0.3 ± 0.7	0.3 ± 0.3

<sup>a</sup> Data for DOC concentration in Harney River during SharkTREx 1 taken from *Cawley et al. (2013)*.

<sup>b</sup> Estuarine contribution to DOC is assumed to be entirely of mangrove origin.



**Table 3.** Mangrove contribution to  $\sum[\text{DIC}]_{\text{estuary}}$  determined from  $\delta^{13}\text{C}_{\text{DIC}}$  mass balance and TAlk/DIC ratios.

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River	Experiment	Methods	
		$\delta^{13}\text{C}_{\text{DIC}}$	TAlk/DIC
Shark River	SharkTREx 1	$60 \pm 6\%$	$70 \pm 3\%$
	SharkTREx 2	$61 \pm 6\%$	$70 \pm 3\%$
Harney River	SharkTREx 1	-	-
	SharkTREx 2	$61 \pm 6\%$	$70 \pm 2\%$

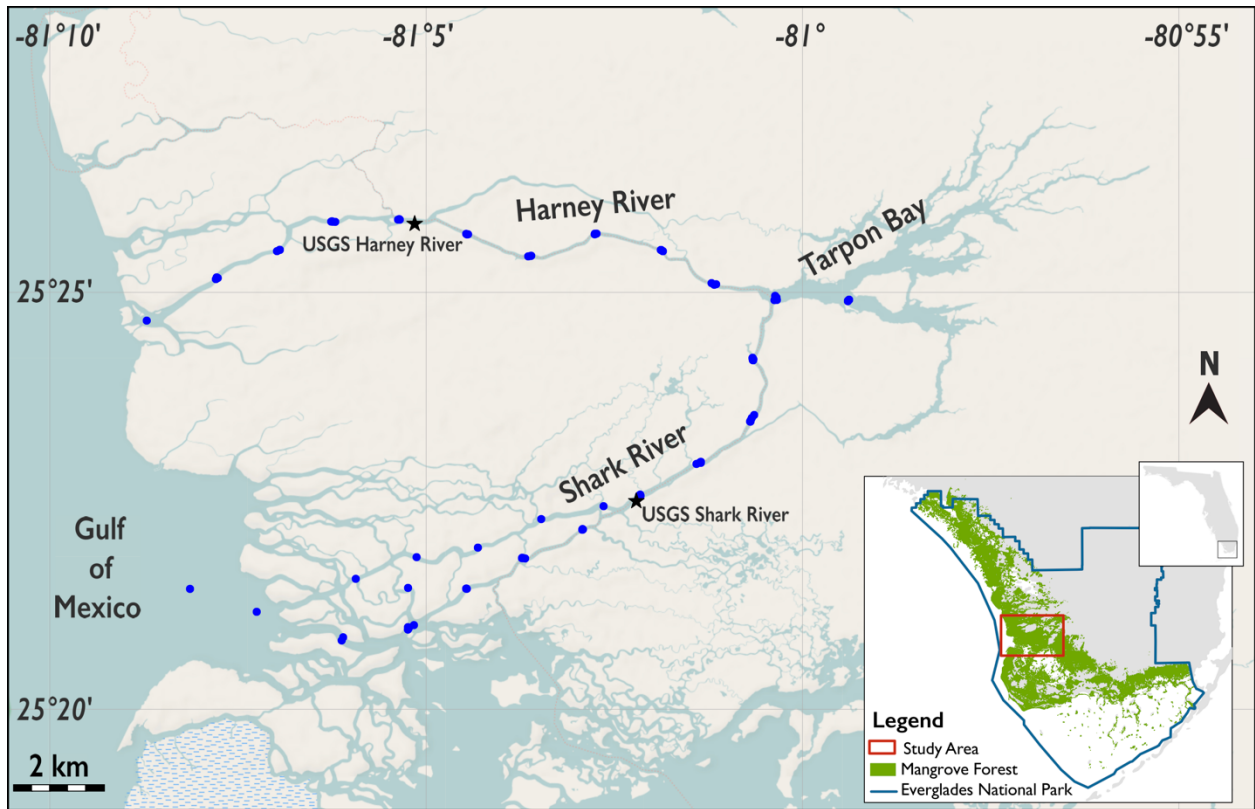
**Table 4.** Distribution of total and mangrove fluxes of DIC and DOC for Shark and Harney Rivers during SharkTREx 1 and 2.

		SharkTREx Experiment #	Estuarine Contribution <sup>a</sup>	Percent of Total Export Flux <sup>b</sup>	Percent of total Mangroves Flux <sup>c</sup>
Longitudinal DIC Flux	Shark River	1	10%	74%	57%
		2	11%	67%	45%
	Harney River	1	-	-	-
		2	13%	68%	48%
Air-Water CO <sub>2</sub> Flux	Shark River	1	49%	9%	35%
		2	52%	14%	45%
	Harney River	1	51%	-	-
		2	63%	14%	43%
All DIC Fluxes	Shark River	1		83%	92%
		2		82%	90%
	Harney River	1		-	-
		2		82%	91%
Longitudinal DOC Flux	Shark River	1	4%	17%	8%
		2	6%	18%	10%
	Harney River	1	19%	-	-
		2	7%	18%	9%

765 <sup>a</sup> Estuarine contribution to the individual fluxes in each river during each experiment

<sup>b</sup> Flux as a percentage of the total dissolved carbon flux (i.e., longitudinal DIC, DOC and air-water CO<sub>2</sub> fluxes)

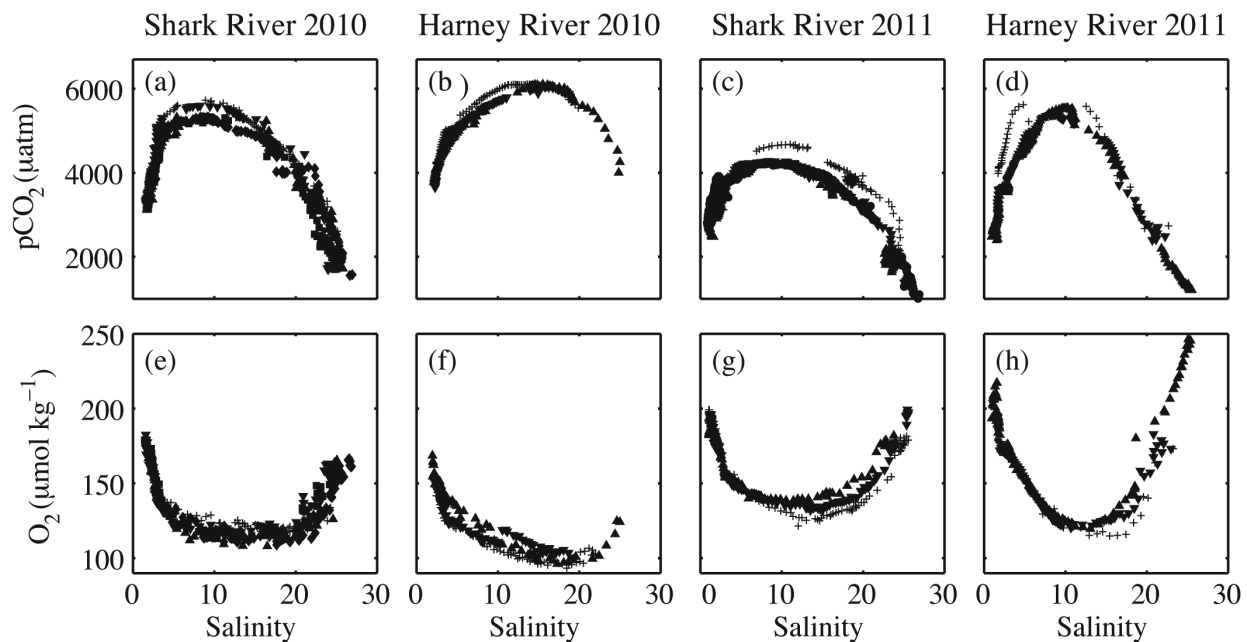
<sup>c</sup> Flux as a percentage of the total mangrove-derived dissolved carbon flux (i.e., longitudinal DIC, DOC and air-water CO<sub>2</sub> fluxes)



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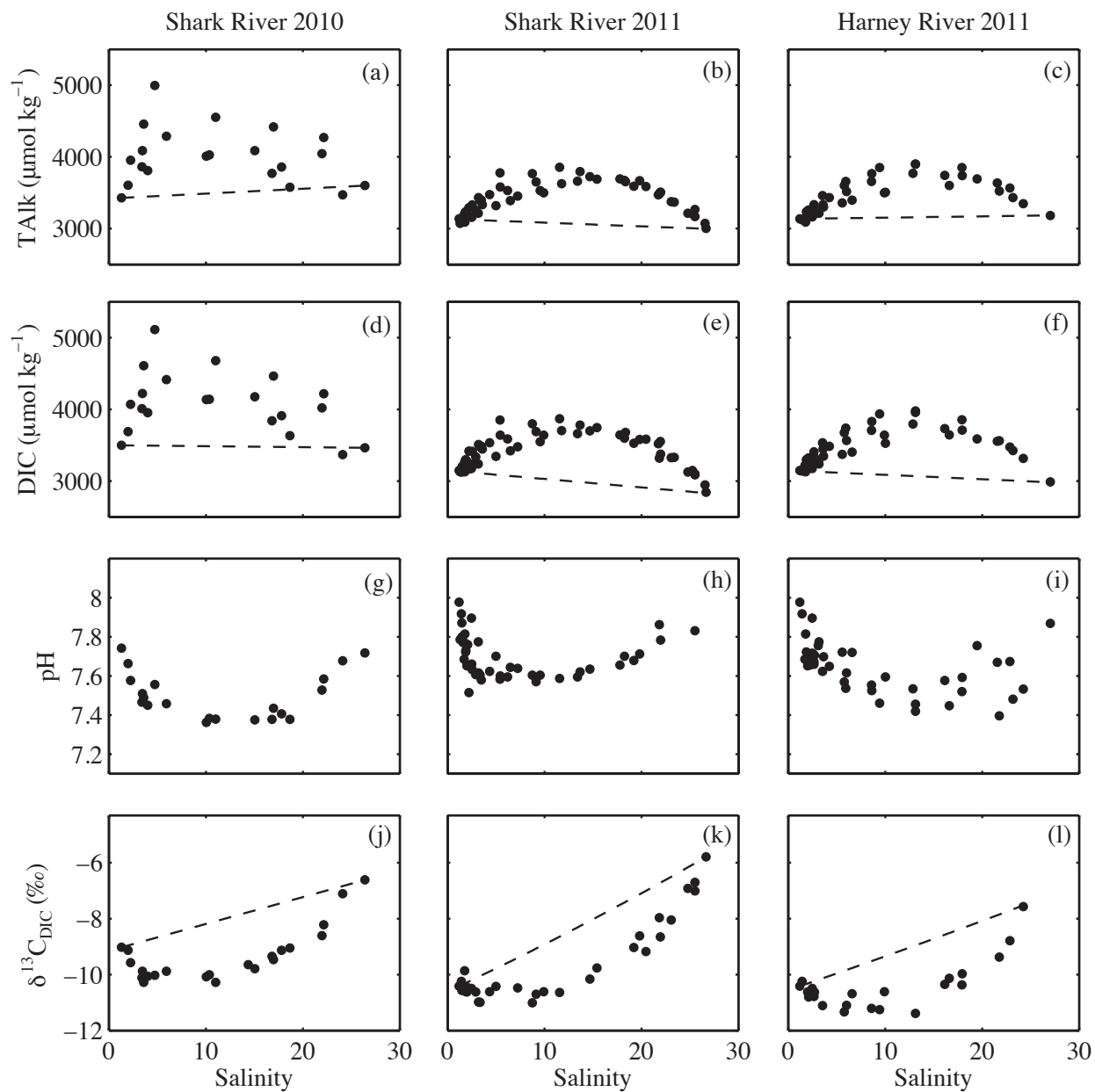
**Figure 1.** Map of the study area near the southern tip of Florida, USA, showing locations of Shark River, Harney River, and Tarpon Bay. The blue circles indicate the locations where discrete samples were taken, and the black stars denote the USGS gaging stations on both rivers. The green areas in the inset are part of the largest contiguous mangrove forest in North America. Indicated in the inset are the boundaries of Everglades National Park.

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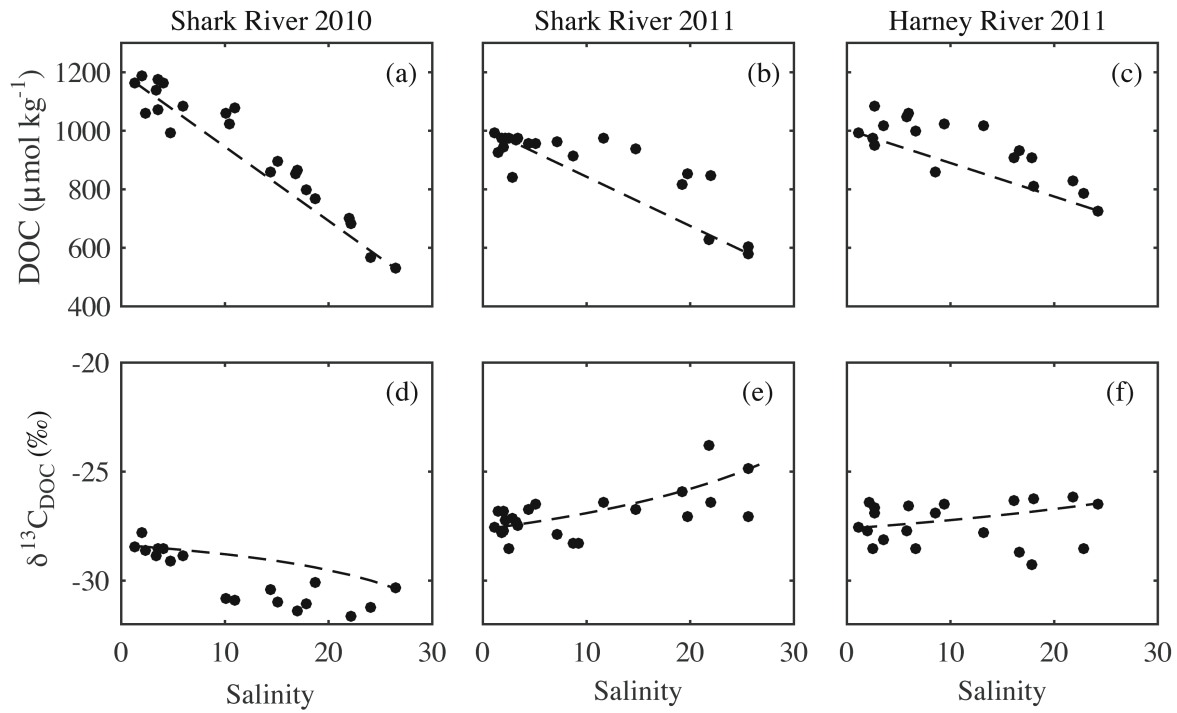


**Figure 2.** Distributions of pCO<sub>2</sub> (a-d) and dissolved O<sub>2</sub> (e-h) along the salinity gradient in the Shark and Harney Rivers during the 2010 (SharkTREx 1) and 2011 (SharkTREx 2) campaigns. Different symbols represent measurements made on different days.

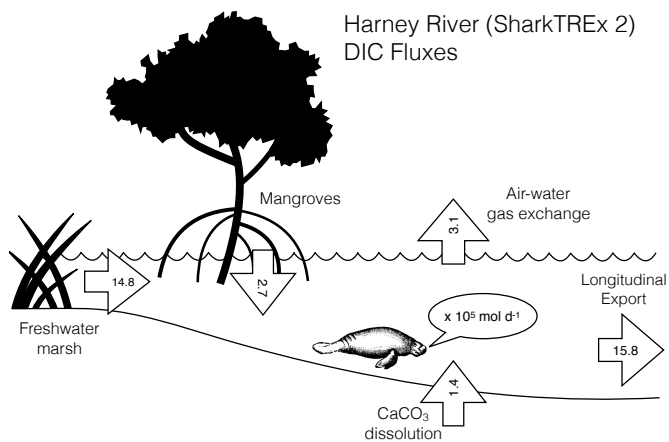
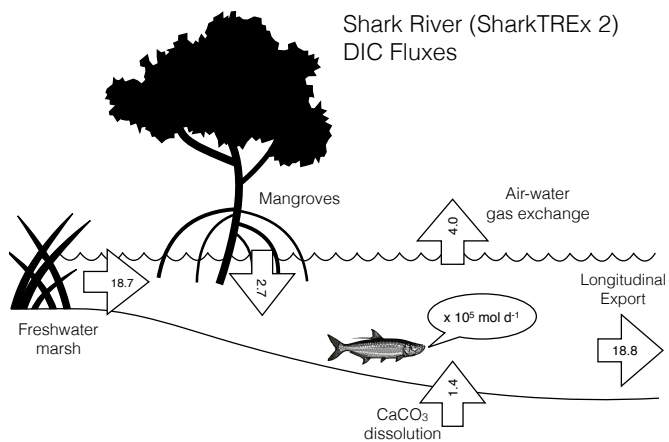
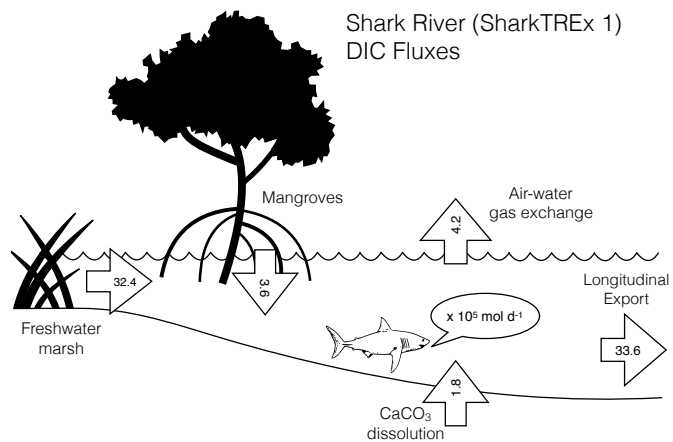
780



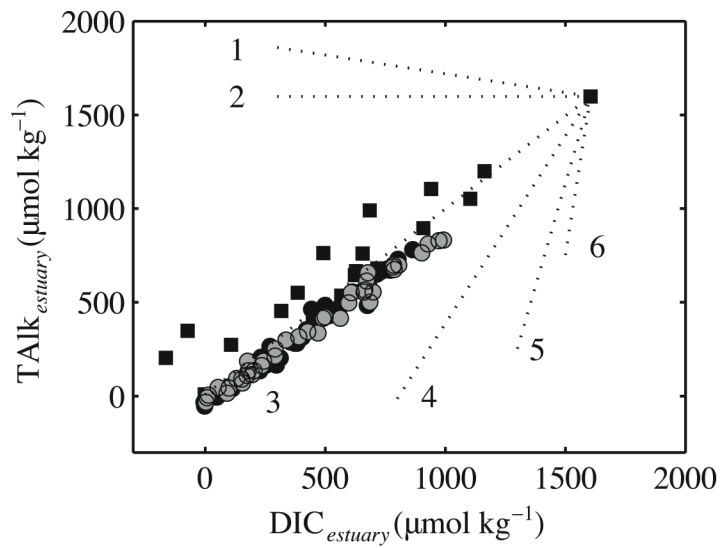
**Figure 3.** Distribution of TAlk (a-c), DIC (d-f), pH (g-i) and  $\delta^{13}\text{C}_{\text{DIC}}$  (j-l) along the salinity gradient in the Shark and Harney Rivers during the 2010 (SharkTREx 1) and 2011(SharkTREx 2) campaigns. During SharkTREx 1, TAlk and pH were measured at FIU, and DIC was calculated using CO2SYS (Pierrot et al., 2006). During SharkTREx 2, DIC and TAlk were measured at NOAA/AOML, and pH was calculated using CO2SYS. The dashed lines indicate the distribution expected for conservative mixing.



**Figure 4.** Distribution of DOC and  $\delta^{13}\text{C}_{\text{DOC}}$  along the salinity gradient in the Shark and Harney Rivers in samples collected during SharkTReX 1 and 2. The dashed lines indicate the distribution expected for conservative mixing.



795 **Figure 5.** Diagrams showing the main DIC fluxes (in  $10^5 \text{ mol d}^{-1}$ ) entering and exiting the Shark and Harney Rivers during SharkTREx 1 and 2. Fluxes from the freshwater marsh were assumed to be fluxes estimated from the conservative DIC curves.



**Figure 6.** (a) Covariation of  $DIC_{estuary}$  and  $TAlk_{estuary}$ . Black squares are samples from the Shark River during SharkTREx 1, and black and gray circles are from the Shark and Harney Rivers, respectively, during SharkTREx 2. Dotted lines represent the theoretical covariation of DIC and TAlk for different biogeochemical processes: 1) aerobic respiration; 2)  $CO_2$  emission, 3) sulfate reduction, 4)  $CaCO_3$  dissolution, 5) manganese reduction, and 6) iron reduction.