A perturbed biogeochemistry model ensemble evaluated against in situ and satellite observations

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Abstract. The dynamics of biogeochemical models are determined by the mathematical equations used to describe the main biological processes. Earlier studies have shown that small changes in the model formulation may lead to major changes in system dynamics, a property known as structural sensitivity. We assessed the impact of structural sensitivity in a biogeochemical model of intermediate complexity by modelling the chlorophyll and nitrogen concentrations at five different oceanographic stations spanning three different regimes: oligotrophic, coastal, and the abyssal plain, over a 10-year timescale. A 1-D MEDUSA ensemble was used with each ensemble member having a combination of tuned function parameterizations that describe the key biogeochemical processes, namely nutrient uptake, zooplankton grazing, and plankton mortalities. The impact is quantified using phytoplankton phenology (initiation, bloom time, peak height, duration, and termination of phytoplankton blooms) and statistical measures such as RMSE, mean, and range for chlorophyll and nutrients. The spread of the ensemble as a measure of uncertainty is assessed against observations using the Normalised RMSE Ratio (NRR). We found that even small perturbations in model structure can produce large ensemble spreads. The range of 10-year mean surface chlorophyll concentration in the ensemble is between 0.14-3.69 mg m^{-3} at coastal stations, 0.43-1.11 mg m^{-3} on the abyssal plain, and 0.004-0.16 mg m^{-3} at the oligotrophic stations. Changing mortality and grazing functions have the largest impact on chlorophyll concentrations. The in situ measurements of bloom timings, duration, and terminations lie mostly within the ensemble range. The RMSEs between in situ observations and the ensemble mean and median are mostly reduced compared to the default model output. The NRRs for monthly variability suggest that the ensemble spread is generally narrow (NRR 1.21-1.39 for nitrogen and 1.19-1.39 for chlorophyll profiles, 1.07-1.40 for surface chlorophyll, and 1.01-1.40 for depth integrated chlorophyll). Among the five stations, the most reliable ensembles are obtained for the oligotrophic station ALOHA (for the surface and integrated chlorophyll and bloom peak height), for coastal station L4 (for inter-annual mean), and for the abyssal plain station PAP (for bloom peak height). Overall our study provides a novel way to generate a realistic ensemble of a biogeochemical model by perturbing the model equations/parameterizations, which will be helpful for the probabilistic predictions.

1 Introduction

Major changes in ocean biogeochemistry have been driven by anthropogenic activities, leading to ocean acidification, eutrophication, and increased levels of dissolved inorganic carbon (Gehlen et al., 2015; Bopp et al., 2013; Doney, 2010). To understand how the ocean ecosystem responds to these changes, marine biogeochemical models have been developed. The majority of

5 these models focus on the lower trophic food-webs and explicitly represent dissolved nutrients, phytoplankton and zooplankton (NPZ). These models are then coupled with physical general circulation models to address and predict the impact of climate change in the ocean ecosystems (Doney et al., 2012; Yool et al., 2013; Butenschön et al., 2016).

Marine biogeochemical model development began with simple NPZ models, and has become steadily more complex with increasing computing power and knowledge of ocean biogeochemistry (Anderson, 2005; Anderson et al., 2015). NPZ models

10 consist of three compartments: nutrients as the primary resource, phytoplankton as the primary producers, and zooplankton as herbivores or grazers. Such models have been used to investigate the range of possible ecosystem behaviours before coupling them to a physical model (Franks, 2002) and seeking to represent observations at particular sites (Fasham et al., 1990; Robinson et al., 1993). More advanced biogeochemical models represent more processes and feedbacks compared to the NPZ models (Raick et al., 2006), covering much more of the lower-trophic food web (Anderson, 2005). Inclusion of cell size representations

15 (Berelson, 2002; Le Quèrè et al., 2005), different phytoplankton functional types, such as calcifiers and dimethyl sulphide producers (Le Quèrè et al., 2005), and the addition of important micronutrients, such as iron to permit phytoplankton growth limitation (Yool et al., 2011, 2013), are now part of many biogeochemical models. Moreover, in order to investigate the effect of global climate change and anthropogenic activities in the ocean, marine biogeochemical models are now being embedded into earth system models. For example, the Model of Ecosystem Dynamics, nutrient Utilisation, Sequestration, and Acidification

20 (MEDUSA) (Yool et al., 2011, 2013) is the chosen biogeochemical component for the UK Earth System Model, as it has high spatial correlation with patterns of pCO_2, DIC, and alkalinity (Cox and Kwiatkowski, 2013; Kwiatkowski et al., 2014).

Despite becoming more complex (Anderson, 2005), the basic interactions among nutrients, phytoplankton, and zooplankton are still at the heart of all marine biogeochemical models. These interactions are governed by four primary processes, represented in the simplest NPZ models: nutrient uptake, grazing by zooplankton, phytoplankton and zooplankton mortality. These

25 processes are functions of the state concentrations and can be parameterized by different functional forms along with adjustable parameters. Biogeochemical models therefore have different sources of uncertainty, such as the physical input (Sinha et al., 2010; Doney, 1999; Hemmings and Challenor, 2012), parameters (Oschlies and Schartau, 2005; Friedrichs et al., 2006, 2007), and the model structure associated with how the ecosystem is represented, either by the number of model compartments and linkages (Friedrichs et al., 2007; Kriest et al., 2010; Ward et al., 2013), or its mathematical formulations (Anderson et al.,

30 2010; Flora et al., 2011; Adamson and Morozov, 2013; Aldebert et al., 2016). Sensitivity analyses show that small changes in the structural process formulation often produce larger changes in system dynamics, compared to varying parameter values alone (Wood and Thomas, 1999; Fussmann and Blasius, 2005; Levin and Lubchenco, 2008; Flora et al., 2011; Adamson and Morozov, 2013; Aldebert et al., 2016), a result known as structural sensitivity (Wood and Thomas, 1999; Flora et al., 2011; Adamson and Morozov, 2013). A study by Aldebert et al. (2016) shows that parameter values are weakly correlated to

food-web dynamics compared to the model formulations, as equilibrium dynamics are determined by the choice of functional forms.

Structural sensitivity may be less significant in models built on well-tested mechanisms as in the physical sciences, however in biogeochemical models the process functional terms are all gross simplifications. This is even more problematic if the processes are poorly understood so that justification for any specific representation is weak (Adamson and Morozov, 2013). Often it is difficult to implement the functional relations that are observed in the laboratory into a large-scale ecosystem with heterogeneous populations (Englund and Leonardsson, 2008). It is known from studies of simple predator-prey models that similarly shaped equations often lead to completely different stability and oscillatory model dynamics (Fussmann and Blasius, 2005; Roy and Chattopadhyay, 2007). Moreover, a specific functional form may not capture all details of the biological processes, for example, the Michaelis-Menten type function for grazing, commonly known as the 'Holling Type II', fails to correctly describe what happens to grazers' movements when satiation has been reached (Flynn and Mitra, 2016). These discrepancies from simple interaction models suggest that complex biogeochemical models need to be tested by altering their default functional forms (Anderson and Mitra, 2010; Anderson et al., 2010).

A few studies have investigated the effects of biogeochemical process formulations, e.g. Yool et al. (2011) has demonstrated in an intermediately complex model, that linear density-dependent mortality produces the biggest difference in diatoms, with concentrations at mid latitudes being twice as high, compared with sigmoidal, quadratic, or hyperbolic forms. The choice of zooplankton grazing equations affect phytoplankton concentration dramatically in a model with five plankton types, Plank-TOM5.2 (Le Quèrè et al., 2005). The Holling type II grazing function produces 30% less total surface phytoplankton concentration compared to the sigmoidal (Holling type III) function, in the North Atlantic and North Pacific (Anderson et al., 2010). However Anderson et al. (2015) shows that when two similarly-shaped photosynthesis-irradiance curves (Smith and the exponential) function, were used in an NPZ-detritus model, the concentration of chlorophyll during the spring bloom was only slightly higher (0.2 mg m^{-3}) for the exponential function, with little difference in phytoplankton dynamics (Anderson et al., 2015).

Since the individual compartments of models interact with one another, any biological perturbation is likely to affect the whole ecosystem dynamics. In climate modelling, perturbed physics ensembles have been developed to investigate multiple parameter uncertainty (Murphy et al., 2007; Tinker et al., 2016), and multiple parametrization (functional) uncertainties (Subramanian and Palmer, 2017). Inspired by these studies, here we attempt to generate a perturbed biogeochemical ensemble where model equations are varied by embedding different functional forms to describe the core processes, similar to the multi-parameterization ensembles in physical models. We implement this framework in the MEDUSA model (Yool et al., 2011, 2013), which is a lower trophic level model with two phytoplankton functional types, distinguished as large diatoms and small non-diatoms, two zooplankton types represented by mesozooplankton and microzooplankton, and three nutrients: silicic acid, iron, and inorganic nitrogen. Nitrogen is the primary currency of the model, similar to NPZ models, but MEDUSA allows phytoplankton to have different C:N ratios and Si:N ratios for diatoms. Diatoms utilise the silicic acid and can only be grazed by mesozooplankton. MEDUSA also includes an iron submodel developed by Parekh et al. (2005) based on Dutkiewicz et al.

(2005), in which iron is separated into "free" iron and iron bound to organic ligands. Iron is removed by scavenging and added to the ocean by aeolian deposition.

We assess of the uncertainty arising from the MEDUSA model's equations from ensemble outputs generated using possible functional form combinations within the NPZ compartments. For simplicity we use a 1-D version of MEDUSA-1.0 model (Yool et al., 2011; Hemmings et al., 2015), and produce results for five oceanographic stations covering abyssal plain, oligotrophic, and coastal regimes. Apart from the model outputs on concentration of nutrients and chlorophyll, we also examine the emergent properties using phytoplankton phenology metrics. The performance of the ensemble mean, median, and the default MEDUSA run are compared with monthly and inter-annual values from in situ observations at those stations. We assessed the spread of the ensemble using the Normalised RMSE Ratio (NRR) which assesses the likelihood of the observations fitting the ensemble range. Section 2 describes the equations used and how the ensemble is run. The assessment of the uncertainty in terms of chlorophyll concentrations, phytoplankton phenology, and comparisons with the observations are described in section 3, and are further discussed in section 4.

2 Method

To explore structural uncertainty we first make the functional forms representing key processes more similar to each other by tuning the shape-defining parameters. For example, for Holling type II and Holling type III, we fix the maximum rates of each process, and implement a non-linear least squares method to optimize the half saturation coefficients so that the overall shapes are similar. This approach is used for nutrient uptake (4 functional forms), phytoplankton mortality (4 functional forms), and zooplankton mortality (4 functional forms), as in the subsections below. Table 1 shows the equations and parameter values.

2.1 Nutrient uptake

Alongside light, nutrient concentration limits the growth of phytoplankton. In MEDUSA the standard hyperbolic monod, hereafter U_1, function is the default. The growth of cells monotonically increases with ambient nutrient concentration, and halts when nutrients become scarce. If nutrient concentrations are high, the rate of uptake saturates. Other mathematical functions show similar properties including (i) Sigmoidal (Fennel and Neumann, 2014) U_2, (ii) the exponential (Ivlev, 1961), U_3, and (iv) trigonometric functions (Jassby and Platt, 1976), U_4. All these functions include a shape-defining parameter, k, which for monod and sigmoidal can be interpreted as a half saturation constant, and a maximum uptake rate, V_{p^T}, which is a function of temperature (Eppley, 1972): $V_{p^T} = V_p 1.066^T$, where V_p is the maximum growth rate when temperature, T, is at 0°Celsius. The uptake function of different phytoplankton types and nutrients use similar functions but different parameter values for k, summarised in Table 1, obtained by minimising the sum squared difference with U_1. The nutrient uptake functions after optimization are shown in Fig. 1(a). The difference in shape of the optimised functional forms are more obvious at low nutrient concentrations.

2.2 Zooplankton grazing

In MEDUSA, both phytoplankton and zooplankton are grouped into "small" and "large" categories. The small zooplankton, represented by the microzooplankton, graze on small phytoplankton, non-diatoms, and detritus, with the more nutrient rich, higher quality, non-diatoms preferred over detritus. Larger zooplankton, represented by mesozooplankton have a broader range of prey, including both microzooplankton and diatoms, which are high quality food sources. When describing multiple grazing functions, the zooplankton grazing rate is often defined using either the Holling type II hereafter, G_2 or Holling type III hereafter, G_1, function with maximum grazing rate g_m, and a weighted preference on the different food sources p_n (Fasham et al., 1990). The preference parameter changes through the year as a function of the food ratio. G_2 and G_1 grazing on prey Pa are described in Table1. In MEDUSA, the default multiple grazing parameterisation is based on the sigmoid Holling type III (Ryabchenko et al., 1997) function. Apart from the weighted preference, both of these functions include a half saturation constant k_x, where x is the zooplankton type.

These functions both become constant at a maximum grazing rate. At low zooplankton concentrations the sigmoidal response has lower grazing rates than the hyperbolic, and therefore, the sigmoidal curve has a more rapid increase in predation rate before becoming saturated (Edwards and Yool, 2000), shown on Fig. 1(b). Preferences for food types are kept the same as MEDUSA's default parameters, with terms summarized in Table 1.

2.3 Plankton mortality

MEDUSA has both density independent and density dependent mortality rates for all the phytoplankton and zooplankton types. Density-independent loss is modelled by a linear function representing plankton metabolic loss which was kept unchanged. Density-dependent loss includes processes such as higher-trophic grazing and disease. In MEDUSA these processes are modelled using the hyperbolic function of plankton concentration (Fasham et al., 1993). Alternative functions can describe the density-dependent mortality, and we use the combinations of hyperbolic (ρ_1, ξ_1), linear (ρ_2, ξ_2), quadratic (ρ_3, ξ_3), and sigmoidal (ρ_4, ξ_4) functions to describe the phytoplankton (ρ) and zooplankton (ξ) mortalities (equations and abbreviations are shown on Table 1). Similar to grazing and nutrient uptake, the functional forms have different maximum rates for each plankton type. These maximum rates are made the same for all the different functions.

Of the four different mortality functions, linear and quadratic functions are most different in shape, as shown on Fig. 1(b). Using the linear term is similar to a change in the value of maximum mortality rate, μ. To make the linear function similar to the sigmoidal and hyperbolic functions, the maximum grazing rate is set so that the total loss integrated over the range of prey density (calculated as the area below the function representing the total loss) is similar to that for the hyperbolic curve. The quadratic term, instead of asymptoting, continues to grow with plankton abundance. In order to keep this similar to other forms, after reaching a certain concentration the function is switched to linear, so that the rate plateaus at high abundance. For sigmoidal mortality, the default μ are not changed but the half-saturation constant, k_M is optimised. The optimised mortality functions are shown in Fig. 1(c). A distinctive feature of these functional forms after optimisation is that the quadratic mortality

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rate remains low until phytoplankton concentration reaches 10 mmol N m^{-3}, and the linear function shows consistently high plankton mortality (Fig. 1(c)).

2.4 Model Parameters

Apart from sinking rate, maximum growth, and grazing rates, parameters not listed in Table 1 are kept at their default values (Yool et al. (2011) shown on table 1-4). From a previous 3-D MEDUSA run, in the oligotrophic regions show a low 'background' chlorophyll concentration (Yool et al., 2011) so to raise this concentration a higher maximum growth rate and lower grazing rate has been used. The maximum uptake rate, V_p, as 0.8 day^{-1}, similar to that in the HadOCC model (Palmer and Totterdell, 2001). For zooplankton grazing, similar to NPZ models (Fasham et al., 1990; Fasham, 1995; Anderson et al., 2015) we use 1 day^{-1} as the maximum grazing rate, g_m. MEDUSA also contains both slow and fast detritus sinking factors. It is assumed that the latter sinks rapidly relative to the model time-step, and remineralisation of the detrital nitrogen and silicon is done implicitly. In the default model 3 m day^{-1} is used for the slow sinking detritus, however over long runs we found this leads to downward loss of nutrients from the euphotic zone. Earlier studies have used lower detrital sinking rates (Steele and Henderson, 1981; Fasham et al., 1990; Lacroix and Gregoire, 2002; Raick et al., 2006), between 0 to 1.25 m day^{-1} and other study have suggested to use 0 m day^{-1} (Ward et al., 2013). We chose a lower sinking rate of 0.1 m day^{-1} to prevent depletion of state variables particularly at the shallower stations.

2.5 Running the Model and Generating the Ensemble

MEDUSA is run in the Marine Model Optimization Testbed (MarMOT-1.1) (Hemmings and Challenor, 2012; Hemmings et al., 2015), a site-based mechanistic emulator, where simulations are run in 1-D. MarMOT was developed to investigate the effect of sensitivity in plankton model simulations, especially in regard to parameter and environmental inputs (Hemmings and Challenor, 2012). Despite some uncertainties associated with the differences in physical forcing, fluxes, and initial values of biogeochemical properties, using 1-D simulations to approximate 3-D model behaviour for calibrating models based on specific sites has improved the 3-D models' predictive skill (Oschlies and Garçon, 1999; Oschlies and Schartau, 2005; Kane et al., 2011; McDonald et al., 2012). The 1-D MEDUSA is run at five oceanographic stations: PAP, ALOHA, BATS, Cariaco, and L4 shown in Fig. 2. These are chosen as they represent different oceanographic regimes: abyssal plain (PAP), oligotrophic (ALOHA, BATS), and coastal (Cariaco, L4).

At each oceanographic station, all combinations of the optimized functional forms (as described in subsection 2.1, 2.2, and 2.3), are then embedded into the 1-D MEDUSA code. The same process function is always used for both diatoms and non-diatoms, or mesozooplankton and microzooplankton. The ensemble model at each station is initialized using in situ measurements of chlorophyll, nitrogen, silicic acid, and iron, and the ensemble is run over 10 years starting from January 1998. This provides a total number of 128 combinations, arising from 4 types of nutrient uptake, 4 phytoplankton mortalities, 2 types of zooplankton grazing, and 4 zooplankton mortalities.

2.5.1 Physical input

Physical input files consist of gridded values of vertical velocity (m day^{-1}), vertical diffusion coefficient (m^2 day^{-1}), and temperature ($^{\circ}$C), which are applied at each depth level. Additionally, time series of downwelling solar radiation (W m^{-2}) and mixed layer depth (m) are also used as input. These are obtained from the 5-day mean output of the Nucleus for European

5 Modelling of the Ocean (NEMO) model, using the Met Office Forecast Ocean Assimilation Model (FOAM), which controls the physical parameters and therefore the biogeochemical tracers every 5 days. The FOAM-NEMO system assimilates *in situ* satellite SST, sea-level anomaly, sea-ice concentration, temperature, and salinity profile data, in order to make the physical system more realistic (Storkey et al., 2010). However, assimilating physical data directly into a coupled physical-biogeochemical model often does not improve the simulation of the ecosystem. For example when assimilation is used in the 3-D HadOCC

10 model it overestimates the nutrient concentrations due to spurious vertical velocities (Ford et al., 2012; Ourmières et al., 2009).

To avoid overestimating surface nutrients the vertical velocities from the FOAM system were capped at the 90th and 10th quantiles, and the 10-year mean of the vertical velocity is also removed. This means that there is no time mean vertical velocity. These adjustments gave a better long-term vertical structure to the nutrient and other distributions. Since input data on the vertical diffusivity was not stored in FOAM, we used values from NEMO ORCA025-N102 output from January 1998-December

15 2001 and from ORCA0083-N01 from January 2002-December 2007, both obtained from the CEDA Group workspace web (http://gws-access.ceda.ac.uk/public/nemo/#_top). These physical inputs are 5-day averaged and are available at 75 depth levels (from 0.5 to 6000 m) for NEMO-FOAM and ORCA0083-N01, and 63 depth level (spanning from 6 to 5800 m) for NEMO ORCA024-N102. The level thickness increases exponentially as the depth goes deeper. Our 1D model uses these same 63 depth levels vertical resolution.

2.5.2 Biogeochemical input and validation data

The 1D MEDUSA ensemble is run at five oceanographic stations: Porcupine Abyssal Plain Sustained Observatory (PAP-SO, hereafter, PAP), A long time Oligotrophic Habitat Assessment (ALOHA), Bermuda Atlantic Time Series (BATS), Cariaco, and L4. The input for the biogeochemical environment are the initial conditions for the 11 primary tracers (state variables) including; dissolved organic nitrogen (DIN), non-diatom, diatom, silicon in diatom, silica, detritus, microzooplankton, meso-

25 zooplankton, non-diatom chlorophyll, diatom chlorophyll, and iron (mmol m^{-3}), along with the model parameter values. Initial conditions for chlorophyll, silicate, iron, and nitrogen concentrations and are taken from the in situ data at the five oceanographic stations. We did not use spin up runs when initialising, as discussed in the supplementary material section S1. Location coordinate, data source, and maximum depth are summarised in Table 2 and the stations locations are shown in Fig. 2. After initialization, in situ data from these stations are used to validate the model results. For station PAP, we also use SeaWIFS-

30 derived chlorophyll-a data with 9 km spatial resolution and 8-day averages provided by GlobColor (http://hermes.acri.fr/) for validating the surface chlorophyll.

At these stations, the DIN consists of ammonia, nitrate, and nitrite, however at oligotrophic stations like ALOHA the ammonium is below the detection limit (Hawaii Ocean Time Series), and therefore DIN only consists of nitrate and nitrite. At PAP we

use the initial condition from one of MarMOT's test stations, located at 50°N, 20°W (Hemmings et al., 2015), since the nitrate data were only collected between 30-400 m. At station L4, chlorophyll and nitrogen data were collected from the surface from 1999-2008. Therefore the initial concentrations for chlorophyll and nitrogen are the same at every depth (total chlorophyll = 0.27 mg m^{-3}, nitrogen = 6 mmol m^{-3}). Other inputs that are not available at the websites mentioned above, such as microzoo-
plankton, mesozooplankton, and detritus were taken from the nearest test stations. In the oligotrophic stations, 75 % of total chlorophyll was allocated initially to the non-diatom phytoplankton since these dominated the water column (Villareal et al., 2012). At the other stations half of the total chlorophyll goes into the diatoms.

For validation of the model, we consider the total chlorophyll-a concentration, instead of separating diatoms and non-diatoms. The model is simulated at 37 depth levels, spanning from 6-1200 m to minimise computational cost, apart from
station L4, with maximum depth is 50 m, and Cariaco, where the maximum depth for the physical input is available down to 500 m. At the lowest level, vertical velocity and diffusion are set to zero and this level is a sink for detritus. Additionally, apart from the physical input files a time series for soluble iron flux from dust deposition is applied, but this is constant using the average value from (Mahowald et al., 2009).

2.6   Model Metrics

We use statistical metrics including correlation coefficient, root-mean squared error (RMSE), bias, ensemble range, and 10-year mean depth profiles of nitrogen and chlorophyll and integrated chlorophyll. For surface chlorophyll, apart from the metrics above we use the mean chlorophyll abundance each year in order to see inter-annual variability, and monthly abundance for to the seasonal variations. A similar approach is applied to nitrogen, however we use the integrated nitrogen over 200 m (integrated nitrogen / depth) to calculate the inter-annual mean and monthly abundance. These statistical metrics are compared with in situ
data. We also consider the phenological aspects of the phytoplankton spring bloom, which are useful ecological indicators for detecting natural and anthropogenic impacts on the pelagic ecosystem (Platt and Sathyendranath, 2008). We consider seven phenology indicators as metrics, including an initiation time where the chlorophyll concentration exceeds a certain threshold, at half the concentration of the bloom peak. When the bloom concentration starts to diminish, we derived a termination time, where bloom concentration falls below the same threshold. The number of days when chlorophyll concentration is higher
than the threshold is the bloom duration. The concentration at the bloom peak and the date it takes place, are also included as indicators. We also note the amplitude of the bloom, which is half of the peak height minus the minimum chlorophyll concentration. These indicators are derived using the method described in appendix A, and applied to all ensemble outputs for each year.

In an ensemble forecast system, an ensemble with good reliability is the one that is statistically consistent with the observa-
tions, such that the observation is statistically indistinguishable from the ensemble members. In order to assess the value of the ensemble probability distribution we must assess the consistency of the ensemble spread as well as the ensemble mean error (Moradkhani and Meskele, 2010). A simple method is discussed by Anderson (2001) which takes the ratio R_a of RMSE of the ensemble mean and the mean RMSE of all the ensemble members which has the expectation value $E[R_a] = \sqrt{\frac{(n+1)}{2n}}$, where n is the number of ensemble members. This is called the Normalised RMSE Ratio (NRR= $R_a/E[R_a]$) where the desirable

ensemble spread is expected to have NRR=1. If the NRR >1 then the spread is too small, and NRR <1 indicates that the ensemble spread is too large. We may expect different NRR values for different metrics and also for variability on different timescales, such as monthly or inter-annual data. This method has previously been used to set the number of ensemble members in data assimilation (Moradkhani et al., 2006; Roy et al., 2012).

3 Results

First the ensemble range and mean are compared with the observational fields, followed by the error statistics for the ensemble mean/median, the default run, and the ensemble range in order to assess whether the ensemble spans the observational data. Then variability are compared with the in situ data, followed by the NRR to assess the ensemble spread, and phytoplankton bloom phenology.

3.1 Abyssal Plain

In station PAP, in situ nitrate was only measured from mid 2002 to mid 2004 with a maximum depth of 400 m and chlorophyll from mid 2003 to mid 2005 with maximum depth of 200 m, as in Table 2. Surface chlorophyll is derived from SeaWIFs (8-day averaged) and is available for the full 10-year time series (see supplementary Fig. S5).

Distinct seasonality is simulated by the ensemble mean. High nitrate concentrations at the surface occur during winter (December-April) and decline in summer. From the seasonal mean nitrogen profile in Fig. 7(a), the ensemble inter-quartile range shows later bloom peaks compared to in situ, and mean nitrogen concentrations are high during months when both the ensemble and in situ nitrogen decline (between May to June), and show an earlier spike of nitrogen in July instead of September, and therefore underestimating the increase of nitrogen between October to December. These shared errors make the ensemble spread (NRR=1.25) still too narrow for the phenological metrics.

Chlorophyll concentration starts to decline at a depth of ~ 50 m, which also corresponds to the decline in the chlorophyll inter-quartile (between 25th and 75th percentile) range shown on Fig. 3(b). Chlorophyll also shows seasonality, similar to that of nitrogen. The in situ data show high concentrations of chlorophyll in the top 70 m during May-June, coinciding with the shallowing mixed layer depth. In the model this occurs earlier in spring (between end of April to May), and slightly deeper, to 100 m, as summarised in Fig. 3(a) and (c).

Chlorophyll and nitrogen profile 10-year means are within the ensemble range, with ensemble spread on the narrow side, with NRR= 1.20 and 1.25 for chlorophyll and nitrogen respectively. For chlorophyll and nitrogen profiles, the ensemble median shows the highest correlation and lowest RMSE and bias, compared to the default or the ensemble mean. High RMSEs in nitrogen occur from ensemble members that contain the $U_4 G_2$ combinations, as shown in Fig. 13(j), which also correspond to high nitrogen mean (< 9 mmol m^{-3}). High chlorophyll profile RMSEs (>0.62) are produced from members that combine G_1 with $\rho_1 \xi_2$, $\rho_3 \xi_3$, and $\rho_1 \xi_4$, and this coincides with high chlorophyll mean (> 0.7 mg m^{-3}). Surface chlorophyll 10-year mean and RMSEs (>0.8) are notably high when combining U_1 with $\rho_1 \xi_2$, $\rho_3 \xi_3$, and $\rho_1 \xi_4$, as summarised in Fig. 12(e) and (j).

When compared to satellite-derived chlorophyll-a, the surface chlorophyll at this station shows higher correlations and lower RMSEs than in other regions, especially with the ensemble median, compared to the default run or the ensemble mean. In years when satellite-derived chlorophyll is not within the ensemble range this is due to the ensemble overestimating the satellite-derived chlorophyll (supplementary material Fig. S5), giving a slightly narrow ensemble spread (NRR=1.29). There is also inter-annual decline in satellite-derived chlorophyll, (r = -0.21, $p < 0.05$). Six ensemble members capture this decline in surface chlorophyll, although with weaker correlations ($r = -0.14$ (± 0.06), $p < 0.05$). In certain years (1998, 1999, and 2001) the observed chlorophyll are not within the ensemble range, Fig. 4(a), making the ensemble spread too narrow, with NRR of 1.26.

The range of surface chlorophyll annual mean is 0.7 mg m^{-3}. If we only allow one process function at a time to change the ensemble range reduces to 0.58 mg m^{-3}, covering 84% of the all ensemble members. If the original MEDUSA parameters are used, the interannual chlorophyll fits the ensemble slightly better, but the nitrogen fit gets worse. The results from using MEDUSA parameters and in situ nitrogen concentration as initial condition can be found in the supplementary material, section S2 and S3. For monthly data (Fig. 5(a)), the ensemble shows only slightly earlier peak chlorophyll concentrations in May compared to in situ, which occur in either May or June. However, since the ensemble mean and median overestimate the satellite-derived chlorophyll during months of high chlorophyll (> 0.5 mg m^{-3}), during peak chlorophyll in May the satellite-derived chlorophyll is outside the ensemble range.

3.2 Oligotrophic Ocean

In oligotrophic regions nutrients are scarce at the surface but may be plentiful at deeper depths (Dave and Lozier, 2010; Lipschultz, 2001). Figure 9(d) and (e), show that the ensemble range decreases as the depth increases, with high ensemble range found at depths between 3-50 m. At ALOHA all ensemble members have similar distributions for in situ nitrogen, Fig. 9(e), with nitrogen levels > 1.0 mmol m^{-3} only found $\sim$ 150 m depth. However, for inter-annual means, Fig. 4(b), the ensemble inter-quartile range mostly overestimates the in situ observations. There is also an increasing trend of in situ nitrogen ($r = 0.69$, $p < 0.03$), which is not captured by the ensemble, Fig. 6(b), leading to an NRR of 1.38. This overestimation is also observed in the seasonal mean, Fig. 7(b), and in situ data is rarely within the ensemble range.

At BATS, the nitrogen concentration in the top 200 m is clearly overestimated, Fig. 9(k), with nitrogen levels > 1.0 mmol m^{-3} at $\sim$ 10 m (with some members occasionally showing such concentrations at 3 m). Consequently, the overall mean nitrogen concentration is overestimated, as indicated by the positive bias in the ensemble mean, in Table 3. Similar to ALOHA, nitrogen inter-annual and seasonal means are overestimated, summarised in Fig.6(c) and 7(c), respectively. This results in narrow ensemble spread (NRR= 1.40). As at ALOHA, an increasing trend in nitrogen is observed ($r = 0.67$, $p < 0.03$), but only 28.9% of the ensemble results, which uses G_2 as its grazing function shows similar trend.

Another feature of the oligotrophic ocean is a deep chlorophyll maximum (DCM) that occurs below the mixed layer (Fennel and Boss, 2003). In Fig. 9(b) and 9(h), high chlorophyll concentrations are simulated by the ensemble mean between 70-90 m in BATS and up to 150 m in ALOHA. A DCM occurs when lower chlorophyll is detected at the surface, which roughly matches with the in situ profiles at ALOHA (see Fig. 9(c) and Letelier et al. (2004)) and BATS (Fig. 9(i)) although the depth of

the ensemble DCM at both stations is 10-20 m shallower than in situ. The high subsurface chlorophyll coincides with a higher ensemble range, with the range decreasing with depth. However neither BATS nor ALOHA show the continuous DCM seen in the in situ profiles, Fig. 9(c) and 9(i).

The majority of ensemble members underestimate in situ 10-year mean chlorophyll profile concentrations, especially at BATS where all ensemble members show positive bias for both surface and integrated chlorophyll profiles. Figure 11(a) and (b) show that ensemble members with G_2, $\rho_2\xi_2$, $\rho_2\xi_3$, $\rho_2\xi_4$, $\rho_3\xi_1$, and U_3 produce lowest chlorophyll concentrations (< 0.08 mg m^{-3}), which then coincide with high RMSE, shown in Fig. 11(e) and (f). The magnitude and range of RMSEs at BATS are highest, and the default run has lower RMSEs than ensemble mean and median, summarised in Table 3. Ensemble members that use U_1G_1 and U_4G_1 show highest profile 10-year mean concentrations of chlorophyll at both stations, although when paired with $\rho_3\xi_3$ and $\rho_1\xi_4$, the RMSEs increase. High nitrogen concentrations are almost always observed when U_3 and U_2 were used in these oligotrophic regions, summarised in Fig. 13(a) and (b).

As for profiles, surface chlorophyll at ALOHA (supplementary Fig. S2) has lower RMSEs and higher 10-year mean concentrations compared to BATS, summarised in Table 3. Ensemble members with low surface chlorophyll concentrations are the same as those with low chlorophyll profiles, although high surface chlorophyll RMSEs also coincide with high surface concentrations, summarised in Fig. 12(a) and (f). The low RMSEs for surface chlorophyll at ALOHA are also reflected in the NRR (1.07), and the ensemble almost always encompasses the in situ observations, (see supplementary material Fig. S2). During summer months (June-September), most ensemble members still underestimate the in situ monthly means, summarised in Fig. 5(b), and peak values are not always covered by the ensemble.

Figure 4(b) and (c), show no distinct inter-annual variability at either ALOHA or BATS, and in most years the default run is closer to in situ. Model output at BATS have a lower 10-year mean surface chlorophyll than in situ data and most ensemble members underestimate the surface inter-annual means, making the ensemble spread appear narrow both in the 10-year and inter-annual means, shown on Table 3 and 4. Changing the functional forms one at a time produces an ensemble range of 0.11 and NRR =1.39, slightly wider than the whole ensemble, summarised in Fig. 8. At BATS in 2004, a high in situ chlorophyll mean was recorded (0.65 mg m^{-3}) and none of the ensemble members captured this high mean, see the supplementary material Fig. S1 and 4(c), which therefore narrows the NRR value for annual mean at BATS. At ALOHA the range for inter-annual means is too wide, with NRR= 0.84. This is broader than the overall time series mean, as the in situ inter-annual means are mostly within the 75th quartile, making the mean RMSE of the ensemble higher than the RMSE of the ensemble mean (0.043 and 0.025 respectively). However, when only one process is perturbed at a time, the NRR becomes narrow (1.17), and the in situ is no longer within the inter-quartile range, as shown on Fig. 8 and summarised in Table 4. In 1999, the ensemble mean and median is higher than the in situ, which is even clearer in the inter-annual mean of the primary production.

Unlike at PAP, there is no distinct seasonality in chlorophyll. At ALOHA during months of low chlorophyll (< 0.1 mg m^{-3}, occurring in July-October), most ensemble members underestimate surface chlorophyll. At BATS, in situ concentrations are clearly underestimated during the same months, summarised in Fig. 5. At ALOHA, between December to May, when in situ chlorophyll seasonal means are > 0.1 mg m^{-3}, the in situ data are within the inter-quartile range, but at BATS this only happens when in situ chlorophyll means decrease from 0.20 to 0.08 mg m^{-3} in May.

At ALOHA the ensemble mean and median produce smaller RMSEs for both chlorophyll and nitrogen although in the depth profiles compared to BATS. Bias in the default run is still smaller than for the ensemble mean and median. However for surface and integrated chlorophyll the ensemble mean and median have lower bias than the default. At BATS low RMSEs and bias with high correlation coefficient are produced by the default run compared to ensemble mean and median. At both stations, integrated chlorophyll ensemble mean and median have smaller RMSEs and a better correlation coefficient, compared to the default run. At ALOHA, NRR for the integrated chlorophyll is closer to 1 than for the surface and chlorophyll profiles. However the default run in oligotrophic regions generally produces higher chlorophyll and lower nitrogen concentrations compared to the ensemble mean and median. This also matches better with in situ data as the correlation coefficient, r is higher. Using U_1G_1 is seen to give rise to higher chlorophyll concentrations.

3.3 Coastal

In the coastal stations, in situ observations show strong seasonality, shown on Fig. 14(c), (f), (g), and (h). This has not been captured by the ensemble mean in station Cariaco, shown in Fig. 14(a), with chlorophyll concentration almost constant above 0.7 mg m^{-3} in the upper 30 m and the surface (supplementary Fig. S3), shown in Fig. 5(d). Since the range of seasonal chlorophyll is wide, apart from in August and November, in situ concentrations are within the ensemble range. A decline of chlorophyll has been recorded at Cariaco from 2004 (Taylor et al., 2012), and this is captured by the ensemble mean, median, and default ($r = $ -0.72, $p < 0.05$, $r = $ -0.66, $p < 0.05$, and $r = $-0.35, $p < 0.05$ respectively). Similar to chlorophyll, nitrogen from the ensemble also shows no seasonality, see Fig. 7(d). Nonetheless, in situ concentrations are still mostly within the ensemble range, apart from November, where in situ nitrogen decreases to 5.32 mg m^{-3}. However in 2001, and between 2005-2006, inter-annual downwelling of nutrients is well reproduced, Fig. 14(d). Despite the lack of seasonality, annual means of chlorophyll and nitrogen, are mostly within the ensemble range, Fig. 6(f) and 4(f), with the NRR 0.78 and 1.15 for chlorophyll and nitrogen respectively.

At station L4 the in situ and ensemble means both show seasonality of nitrogen with high concentration (> 8 mmol m^{-3}) during November to February, and close to zero (> 0.1 mmol m^{-3}), during summer months, consistent with the observation from Smyth et al. (2010). The interquartile range often overestimates the surface nitrogen concentrations, especially during the sharp decline in April, May, September, and October, shown on Fig. 7(e). For inter-annual means, in situ nitrogen concentrations are mostly within the ensemble range, apart from 2001 and 2002, but this results in NRR of 1.29. For chlorophyll, Fig. 14(g), sharp peaks are observed during spring (March-April) and fall (September), coinciding with the sharp decline of nitrogen in spring. However this has not been represented in the model, where only one peak is simulated between May-June, summarised in Fig. 5(e). If only diatom chlorophyll concentration is shown, the two bloom events are clearer, especially in the default run (see supplementary material Fig. S4). The ensemble mostly overestimates the in situ during non-bloom periods (in situ range = 0.09-2 mg m^{-3}, ensemble range= 0.28-3.13 mg m^{-3}), so that the surface chlorophyll is not fully captured by the ensemble, and therefore NRR value is 1.21, indicating a too narrow spread.

At both stations, for the inter-annual mean, the ensemble range always includes the observations (Fig. 4(c) and (d)), the NRR values are given in Table 3. Weak positive correlations of surface chlorophyll for the ensemble mean is seen at both stations, and

these are better, and with smaller RMSEs than the default run, apart from nitrogen at L4. Similar to the oligotrophic stations, the integrated chlorophyll shows better correlation with in situ, compared to both surface and chlorophyll profiles. At L4, the ensemble mean shows high RMSE for surface nitrogen, but low RMSE for surface chlorophyll, summarised in Table 3.

Although from Table 3, in situ surface chlorophyll concentrations are slightly overestimated by the ensemble mean, most other ensemble outputs are underestimated at Cariaco, except for ensemble members that use the combinations $\rho_2\xi_3$, $\rho_1\xi_2$, $\rho_3\xi_3$, and $\rho_1\xi_4$. This in turn makes the ensemble spread appear too narrow in the NRR. Unlike the oligotrophic regions, these high chlorophyll concentrations coincide with higher RMSE (> 1.7). Higher nitrogen concentrations (>1.2 mmol m^{-3}) with high RMSEs (> 1.5) are also associated with the same ensemble members, summarised in Fig 13(c). These same ensemble members however show relatively low nitrogen concentration (>5 mmol m^{-3}) at station L4. The chlorophyll mean at L4 shows that high concentrations (> 0.2 mg m^{-3}) are produced when the model uses $\rho_2\xi_3$,$\rho_1\xi_2$, $\rho_3\xi_3$, and $\rho_1\xi_4$ combinations. These coincide with high RMSEs, especially in members which pair U_2 and $\rho_2\xi_3$, $\rho_1\xi_2$, $\rho_3\xi_3$, and $\rho_1\xi_4$. High nitrogen concentrations (> 9 mmol m^{-3}) are produced by U_4G_2, with correspondingly high RMSE.

Surface chlorophyll at these coastal stations also has a higher relative range than other stations, with L4 higher than Cariaco, see Table 3. The wider spread for annual means compared to the monthly data over 10-years have been observed at all the stations, including the coastal stations. Despite having lower range than L4 in terms of surface 10-year means, for the inter-annual means (Fig 4(d) and (g)), the NRR value for Cariaco is too small (0.78), indicating the ensemble spread is wider than necessary, which is also observed for inter-annual primary production, shown in Fig. 10(b) and (c). However, if processes are perturbed one at a time, the NRR is closer to the ideal ensemble range (NRR= 0.90). On the other hand, at L4 the in situ annual mean is reliable (NRR=1.001), since the in situ chlorophyll is close to the ensemble median (see Fig. 4(e)). However if the ensemble is reduced by only perturbing one process at a time the NRR increases to 1.36, and the in situ data is no longer within the ensemble mean, shown on Fig. 8, despite the range still covering 86% of the full ensemble (2.14).

3.4 Phytoplankton Phenology

At most stations, the phenology metrics are covered by the ensemble range. There are differences in the timing of phenological events between the ensemble mean, median, and default run, ranging from a couple of days to a couple of weeks, as shown in Table 4. The timing of initiation, bloom peak, and termination show wide interquartile ranges for all stations and can lie between $\sim$20 and 100 days earlier than the in situ timing, apart from stations PAP and ALOHA, see Fig. 15(b). At stations PAP and ALOHA the inter-quartile range is at least $\sim$40 days too early. However, the ensemble mean and median at station L4 and Cariaco are later than in situ timings. For initiation both stations are two months late but are stil within the ensemble range. In terms of the timing of the bloom peak and termination, these are up to 3 months and 120 day late respectively.

BATS has the largest range of phenological timings, especially in termination time. For initiation, the in situ timing is within the interquartile range and only three days earlier than the ensemble median. However, at ALOHA the initiation time shows more inter-annual variability (supplementary Fig. S6) and may occur in June, August and October, as well as in December and January. This causes the mean observed initiation time to end up in May. From Fig. 15(a), the ensemble run shows a mean initiation time between late January and April instead and so the observations fall outside the ensemble range. Due to

this variable initiation, although peak bloom time is within the full ensemble range at ALOHA, the timing is outside the 75th and 25th percentile range, making the ensemble spread too narrow (NRR=1.35). The peak chlorophyll at ALOHA shown in Fig. 5(b), where high (> 0.1 mg m^{-3}) chlorophyll monthly means are recorded in June, August, and September as well as December and January, yet the ensemble mean and median show highest concentrations only in January and February, also placing the bloom timing outside the inter-quartile range, see Fig. 15(b). At BATS the earliest initiation in the ensemble is mid January, but the earliest in situ initiation occurs in February. Therefore, peak bloom time from the ensemble at BATS are usually later than in situ. However, ensemble estimates of bloom peaks for 30°N, where BATS is located, agree with a study by Racault et al. (2012), who identify early April as the peak time. Although the range of peak bloom time in BATS is very high (174 days), the NRR suggests that this range is still narrow (1.17), because most ensemble members produce blooms between April and May, and the in situ timing occurs on 29 March and so is still outside the interquartile range. The large full range is caused by some ensemble members blooming much later. However, since the in situ timing is earlier, it is not within most of the ensemble range, so the overall ensemble spread appears narrow.

Both coastal stations show in situ initiation typically happens in mid-March, which is usually within the ensemble range, which spans 100 days (between the end of February and late June). The ensemble means show later initiation, with the 75th and 25th spanning mid April to end of May for Cariaco, and between early and mid May for L4. This later timing is also clear in peak bloom times, shown on Fig. 15(b). Figure 5(e) shows the in situ bloom at L4 is one to two months overestimated by the ensemble. Cariaco is the only station with peak bloom time, duration, and termination outside the ensemble range, due to the lack of chlorophyll seasonality, as noted in section 3.3, also resulting in higher NRR values.

Initiation timing is captured best at station PAP, with the ensemble median's initiation averaging only eight days earlier than for the satellite-derived chlorophyll, resulting in NRR=1.14 for initiation, closer to one compared to other stations. A typical North Atlantic bloom happens during spring (Raymont, 1980), however most blooms at PAP occur in late May-early June, as shown in Table 4 and Fig. 5. Later blooms are recorded from satellite-derived chlorophyll-a in 2005, three months later than the average and much later than the ensemble mean and median, although the bloom timing is still within the ensemble range, although the NRR value is 1.31. At L4, also in the North Atlantic, the spring bloom is in April, but most ensemble members show later initiation and peak bloom time, mostly in June. Due to this delay the NRR values at L4 indicate that the ensemble range is too narrow, although still within the full ensemble range. Ensemble mean and median at PAP show good agreement with in situ termination date, and although other station termination times are also within the ensemble range, most are later than the inter-quartile range. However, at ALOHA, located at 22°N, the ensemble median for termination at the end of August agrees with the observations from Racault et al. (2012).

Compared to running only the default MEDUSA, where only a single bloom peak is produced, the ensemble range mostly encompasses the in situ peak amplitudes, shown on Fig. 15(c). Only at BATS are in situ peak height and amplitude outside the ensemble range. This is expected since most of the ensemble members underestimate in situ chlorophyll. At Cariaco, in situ peak heights are within the ensemble range, but observed peaks are higher (mean= 3.5 mg m^{-3}, maximum peak= 7.7 mg m^{-3}), and the ensemble reaches less than half of the in situ peak (mean= 1.2 mg m^{-3}, maximum height= 5 mg m^{-3}). This underestimate of the peak and bloom amplitude, results in NRR of 1.40 and 1.39 respectively. Ensemble members with higher

peak and amplitudes are also those with higher chlorophyll biases. Despite the narrow ensemble range, at L4 chlorophyll peaks are within the 75th and 25th range box, and amplitudes are within the full spread. ALOHA and PAP have reliable ensemble spreads according to their NRR values for peak height (see Table 4).

Similar to peak heights, the bloom durations at most stations are within the ensemble range, apart from station Cariaco, which shows the narrowest ensemble spread according to its NRR. The duration at Cariaco is overestimated because the peak is very wide (up to 143 bloom days). This, along with the late initiation of the bloom, results in a three month late termination. At ALOHA, duration is outside the 75th and 25th quartile box, since the peak is also much broader compared to in situ blooms. This results in too narrow ensemble mean according to the NRR. The opposite is true at BATS where in situ peaks are generally broader, and the ensemble members with lower chlorophyll concentration showing narrower peaks, and a greater range in bloom durations, which consequently lowers the NRR value.

4   Summary and Discussion

In this paper we have investigated structural sensitivity, associated with the mathematical formulation of the processes in an intermediately complex biogeochemical model by generating its ensemble outputs of chlorophyll and nitrogen and comparing them with a single default run, and with in situ observations at five oceanographic stations. The ensemble consists of 128 ensemble members, each with different process function combinations. In order to maintain phenomenological similarity, these functions are calibrated using non-linear least squares, while keeping the maximum process rates fixed. We have chosen nutrient uptake, zooplankton grazing, and plankton mortalities to vary, as these are the core processes of every marine biogeochemical model, from the simplest to the most complex. Through this approach, we provide a perturbed biology ensemble conditioned upon process structural uncertainties. Applying structural sensitivity in the 1-D framework has also allowed a large parameter space of concurrent variations to be explored for several different oceanographic regions, and with minimal computational cost. From these assessments, we find that small perturbations in model structure can produce a wide range of results regarding chlorophyll and nutrient concentration as well as phytoplankton phenology.

Our findings reveal that in all regions, the Holling Type II (G_2) grazing function lowers the chlorophyll concentrations, which have also been observed by Anderson et al. (2010) at low concentrations. The nutrients respond in the opposite direction with enhanced nitrogen concentrations. This is expected as at low concentrations, using the G_2 function would graze more phytoplankton, as shown on Fig. 1(b). Pairing G_2 with the linear (ρ_2) mortality of phytoplankton, which constantly removed the phytoplankton regardless of the phytoplankton concentrations, will reduce the chlorophyll concentration even further; but the opposite will happen when G_2 is paired with linear zooplankton mortality. Yool et al. (2011) has similarly shown that using a linear mortality causes the biggest changes in phytoplankton concentrations compared to quadratic and sigmoidal. In contrast, the default phytoplankton (ρ_1) and sigmoidal zooplankton mortality (ξ_4) produce the highest chlorophyll concentrations in all regions, similar to the experiment from Yool et al. (2011). In terms of nutrient uptake, the exponential (U_3) and sigmoidal (U_2) show inefficient uptake, as they produce low chlorophyll and especially high nitrogen concentrations, as shown on Fig. 12(a), (b), 13(a), and (b), especially in the oligotrophic region. Even though the functional forms have been optimised, the most

deviations are observed when nitrogen is < 1 mmol N m^{-3} shown in Fig. 1(a). This makes the U_3 and U_2, which uptake less nitrogen in low nitrogen concentrations, produce high nitrogen and low chlorophyll. However, the effect is not as noticeable compared to using G_1 or G_2.

These disparities between the ensemble members that are caused by the structural differences therefore garner some range.
5 Stations that have produced high chlorophyll concentrations also have high ensemble range. For example, in station Cariaco where chlorophyll concentration is high, despite the discrepancy between the in situ seasonal pattern and the ensemble, the range still covers the in situ concentrations. The mismatch between the observation and ensemble is mostly caused by the physical dynamics, in which the upwelling of nutrients that feeds the phytoplankton is not well simulated by the assimilated vertical velocity. This emphasises that despite using the ensemble approach, a biogeochemical model is only as good as its
10 physical model (Doney, 1999). Even though there are discrepancies, the chlorophyll profile at Cariaco has an NRR value closest to 1, due to the large range in the ensemble. However, in the annual mean of chlorophyll and primary production (Fig. 4(d) and 10(b)) the large range makes the ensemble spread too large. The ensemble range is considerable even if the equation of only one process is changed at a time, which accounts for at least 80% of the full ensemble range. This has been observed at all of the stations, summarised in Table 4 and Fig. 8, emphasising that perturbing functional forms will produce a large range
15 of model results. In some cases, this reduced range may be statistically more meaningful than the full range. For example, compared with the full ensemble, the reduced ensemble range for Cariaco's annual-mean chlorophyll gives an NRR closer to unity. Therefore, it may be possible through a further study to systematically reduce the number of ensemble members, whilst retaining a realistic ensemble range, which will reduce computational costs.

At most of the stations, the ensemble mean produced lower RMSE compared to the default run, suggesting that the structural
20 ensemble with a wide range of predictions covering the in situ observations, is likely to produce a mean field closer to the observation, than a single-structure model. Even in stations such as BATS, where the in situ chlorophyll is underestimated by most ensemble members, the monthly means (Fig. 5(c)) of the in situ values (during months of high chlorophyll) are within the ensemble range. Similar monthly-mean chlorophyll patterns have also been observed in PAP, ALOHA, Cariaco and L4 (with some exceptions in summer month) (Fig. 5(b)), whereby the in situ chlorophyll in generally within the ensemble range. We
25 further note that, considerable model bias such as lower modelled concentrations of chlorophyll, compared to the in situ data, has been observed for the default 3-D MEDUSA model itself, e.g., in the subtropical gyre (Yool et al., 2011). This may be due to the absence of nitrogen fixers and picoplankton in MEDUSA, which cause the increase of plankton concentration in the summer (White et al., 2015), or due to the fact that phytoplankton uptake equation in MEDUSA does not allow phytoplankton to acclimatise in the oligotrophic region through optimum uptake kinetics (Smith et al., 2009; Yool et al., 2011).
30 Apart from the model's state variables such as chlorophyll and nutrient concentrations, we have looked into the model-derived phytoplankton phenology because of its importance to marine ecosystems e.g. importance of the timing of phytoplankton blooms for the survival of zooplankton and fish larvae (Cushing, 1990). The timing of the blooms has also been shown to control the variability of pCO$_2$ in the sub-polar region (Bennington et al., 2009). Despite having a reliable spread in the annual mean, in terms of phytoplankton phenology, stations such as L4 show some mismatch with the observation. In situ
35 initiation, bloom timing, and duration in L4 are earlier than in most of the ensemble members, although still lying within the

ensemble range and some ensemble mean timings (termination and peak bloom time) are similar to the satellite observations at this latitude (Racault et al., 2012). When in situ chlorophyll is fitted with a smooth curve, the highest peak mostly occurs during spring (March-April). But model metrics, including ensemble mean and median, are noisy, and peaks mostly fall in the summer (May-July). Moreover, at L4, distinct phytoplankton blooms occur twice a year: first in spring and the second in autumn (Smyth et al., 2010). These blooms are sometimes well simulated, e.g. in Fig. 14(g) and 5(d), but are not as distinct as in situ measurements due to the variability of the model. Some of these discrepancies may also be caused by the way zooplankton select their prey in MEDUSA. In a study by Sailley et al. (2014) grazing selection based on total prey concentration would result in rapid nutrient turn-over, which results in a single peak event, but if the selection is based on the stoichiometry of C:N, the nutrients would regenerate slower, and therefore result in two chlorophyll peaks. However, the difference in peak timing does not affect the duration of the blooms, and the in situ duration is well within the ensemble inter-quartile range. More generally, discrepancies in predicting bloom timing by large-scale biogeochemical models are also reported in many studies, e.g., Henson et al. (2017) and Kostadinov et al. (2017). Henson et al. (2017) shows that compared with the satellite data, the 3-D MEDUSA 2.0 (Yool et al., 2013) model estimates spring bloom start date ∼50 days late, and in southern hemisphere, model estimates subtropical bloom start date ∼50 days earlier. By generating an ensemble of 7 CMIP5 models, Kostadinov et al. (2017) highlighted that the difference in bloom timing between the model ensemble and satellite-derived chlorophyll can be more than one month over most of the ocean. This agrees with our study (see, Table 4), as most of our ensemble members have earlier bloom initiation dates, and the difference between the ensemble mean and in situ timing of bloom, e.g. PAP and L4, are more than one month. Additionally, the whole ensemble range produced by this study can help to provide an uncertainty range for the timing of phytoplankton blooms. The ensemble range almost always encompasses the observed annual mean, peak height, and amplitude. Therefore it may be suitable to use the ensemble model in order to forecast these phenological aspects. Further, it may also be possible to improve the accuracy of the ensemble range, by systematically removing certain ensemble members in a future study.

Finally, the unresolved discrepancy between in situ observations and large number of process ensemble results, such as in the oligotrophic stations for nitrogen and L4 for phytoplankton peak timings, emphasise that the inclusion of some missing processes, such as active prey selection, and species would improve the performance of the model (Friedrichs et al., 2007; Kriest et al., 2010; Sailley et al., 2014) and functional forms which describe chemostat experiments, such as the droop function or the active prey selection (Sailley et al., 2014) are not as structurally sensitive as the logistic equations (Aldebert et al., 2018). Additionally, MEDUSA uses logistic functions such as Monod and Holling type III equations to describe its processes and are well known to be structurally sensitive (Aldebert et al., 2018). We did not include equations that allow such selection or species, as in this paper we tried to ensure that all the equations have similar properties to the default MEDUSA, in order to show that perturbing the structure of the model equations would result in different plankton and nutrient dynamics. Therefore, comparing the performance of model complexity and the ensemble method was beyond the scope of this study.

5 Conclusions

Our study highlights that it is important to conduct structural sensitivity analyses in addition to parameter sensitivity analyses and it is crucial to include mathematical functions that can capture sufficient information of the key biogeochemical processes known from experimental studies. However, none of the deterministic functions can capture all details of these processes (Anderson et al., 2010), therefore we have introduced a method whereby instead of having only one default model output, we have an ensemble generating a range of possible outcomes arising from alternative model structures. We have explored the structural sensitivity of the 1-D version of MEDUSA, the ocean biogeochemistry component of UK-ESM1, which is becoming widely used in the community. This study emphasises that small perturbations in MEDUSA process equations can produce very different model results, and the ensemble of perturbations generally encompasses the in situ observations. Therefore, our study shows promise that the ensemble of a single biogeochemical model resulting from perturbing the model structure, may produce meaningful prediction ranges of its state variables. However, our study is based on 1-D simulation, and further study with a 3-D biogeochemical model would help extend results to the global ocean. It may also be possible to further minimise the computational costs by systematically reducing the number of ensemble members whilst retaining a realistic ensemble range. Further studies could include varying the weighting of ensemble members, or reducing the number of model combinations to improve the ensemble range and to assess properly different plankton functional types and dissolved inorganic carbon. Such a perturbed biology ensemble may also be useful for data assimilation e.g. with satellite-derived chlorophyll.

Appendix A: Determining phytoplankton phenology

Before determining the initiation time, bloom timing must be identified. This is done by taking the ten years of surface chlorophyll output and breaking it down into individual years. These are then rearranged into two datasets: January-December and June-May, and the date of maximum chlorophyll concentration in each year is determined. If the peak timing occurs mostly towards the end or the beginning of the year, June to May datasets are used instead of the former. The timing is then adjusted if the calendar year has changed.

The initiation is determined by the day that chlorophyll concentration exceeds a given threshold. However, since in situ chlorophyll has some data gaps and modelled chlorophyll is not smooth, some studies have fitted a function or model to the datasets to make the chlorophyll data smoother (Platt et al., 2009; Sapiano et al., 2012; Brody et al., 2013). Here we use a 5th order polynomial curve to get a smooth fit of the bloom peaks in the data (Fig. A1), from which phenology metrics are calculated. After being fitted, a threshold of half the bloom peak concentration is chosen. To find the **peak time**, the date at which maximum chlorophyll concentration is achieved in the fitted curve is determined, and this date is used as a reference to calculate other metrics. **Amplitude** is then calculated as half of the highest peak minus the minimum concentration. **Initiation** is

the day when chlorophyll concentration goes just above the threshold towards the maximum (Brody et al., 2013). **Termination** of the bloom is defined when concentration falls below the threshold (Racault et al., 2012). If two peaks are detected the termination of the spring bloom is determined when the first bloom reduces to its minimum, just before the second bloom starts (in the first valley). **Duration** of the bloom is simply the total number of days on which chlorophyll concentration is above the threshold or termination minus initiation.

This phenology is useful to see how the bloom develops and terminates, whether the concentration increases rapidly and decreases slowly or vice versa. The phenology is summarised in Fig. A1. The curve fitting method is only applied if the data shows potential outliers especially in higher concentrations. If there is only one prominent bloom each year, as at stations ALOHA and BATS, and the data is smooth, the regular threshold method (when the concentration is above 50% of the maximum bloom, and the associated initiation and termination times), without fitting the data with a curve is applied. To avoid results being affected by how bloom phenology is determined, the same method is used for determining the metrics from both in situ and model output.

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Tables

Table 1. Parameter values for resource uptake (U), zooplankton grazing (G), and plankton mortalities (ρ and ξ for phytoplankton and zooplankton respectively), described using similar functional forms (shown in Fig. 1). In grazing equation, g_m represents maximum grazing rate, P_a is the prey, and p_a denotes the grazing preference. Starred equations are the default functional responses in MEDUSA.

Process/ Plankton type	Symbol	Meaning	Parameter value (mmol m^{-3})			
Nutrient Uptake (U)			Monod* (U_1) $\frac{n}{n+k}$	Sigmoidal (U_2) $\frac{n^2}{n^2+k^2}$	Exponential (U_3) $1-\exp(-\frac{n}{k})$	Trigonometric (U_4) $\frac{2}{\pi}\arctan\left(\frac{n}{k}\right)$
Non-diatom	kN_{nd}	shape defining constant for nitrogen	0.5	0.74	1.12	0.60
	kFe_{nd}	shape defining constant for iron	0.33 $\times 10^{-3}$	0.49 $\times 10^{-3}$	0.74 $\times 10^{-3}$	0.40 $\times 10^{-3}$
Diatom	kN_d	shape defining constant for nitrogen	0.75	1.12	1.68	0.91
	kSi_d	shape defining constant for silicon	0.75	1.12	1.68	0.91
	kFe_d	shape defining constant for iron	0.67 $\times 10^{-3}$	0.99 $\times 10^{-3}$	1.50 $\times 10^{-3}$	0.81 $\times 10^{-3}$

Process/ Plankton type	Symbol	Meaning	Parameter value (mmol m^{-3})	
Grazing (G)			Holling type III* (G_1) $g_m\frac{p_a P_a^2}{k_g^2+p_a P_a^2+p_b P_b^2}$	Holling type II (G_2) $g_m\frac{p_a P_a^2}{k_g(p_a P_a+p_b P_b)+p_a P_a^2+p_b P_b^2}$
Microzooplankton	k_{mi}	half saturation constant	0.80	0.46
	pmi_{nd}	grazing preference for non-diatom	0.75	0.75
	pmi_{det}	grazing preference for detritus	0.25	0.25
Mesozooplankton	k_{me}	half saturation constant	0.30	0.17
	pme_{nd}	grazing preference for non-diatom	0.15	0.15
	pme_{det}	grazing preference for detritus	0.15	0.15
	pme_d	grazing preference for diatoms	0.35	0.35
	pme_{mi}	grazing preference for microzooplankton	0.35	0.35

Process/ Plankton type	Symbol	Meaning	Parameter value (mmol m^{-3})			
Mortality (ρ,ξ)			Hyperbolic* (ρ_1,ξ_1) $\mu\frac{P}{P+k_M}P$	Linear (ρ_2,ξ_2) μP	Quadratic (ρ_3,ξ_3) μP^2	Sigmoidal (ρ_4,ξ_4) $\mu\frac{P^2}{P^2+k_M^2}P$
Non-diatom	μ_{nd}	maximum rate (day^{-1})	0.10	0.10	0.05	0.10
	k_{Mnd}	half saturation constant	0.50	-	-	0.74
Diatom	μ_d	maximum rate (day^{-1})	0.10	0.10	0.05	0.1
	k_{Md}	half saturation constant	0.50	-	-	0.74
Microzooplankton	μ_{mi}	maximum rate (day^{-1})	0.10	0.10	0.05	0.10
	k_{Mmi}	half saturation constant	0.50	-	-	0.74
Microzooplankton	μ_{mi}	maximum rate (day^{-1})	0.20	0.20	0.07	0.20
	k_{Mmi}	half saturation constant	0.75	-	-	1.12

Table 2. Location, data source, and available depth range for the five oceanographic stations

Station	Location	Source	depth range
ALOHA	22°45'N, 158°00'W	http://hahana.soest.hawaii.edu/hot/hot-dogs/interface.html	5-5000 m
BATS	32°50'N, 64°10'W	http://bats.bios.edu/	4-4000 m
Cariaco	10°30'N, 64°40'W	http://imars.marine.usf.edu/cariaco	1-1310 m
L4	50°15'N, 4°12.3'W	http://www.westernchannelobservatory.org.uk/data.php (available upon request)	surface
PAP	49°N, 16.5°W	http://projects.noc.ac.uk/pap/data	7-400m

Table 3. Error statistics, 10-year mean, and NRR of chlorophyll (mg m^{-3}) and nitrogen (mmol m^{-3}) concentration at five stations for the default run, ensemble mean, ensemble median, and the ensemble range (ensemble maximum - ensemble minimum). These are calculated from surface to 200 m depth, starting from January 1998 to December 2007. Bias is (model output) – (in situ observation). Bold text indicate the smallest RMSE. At Station L4 error statistics and mean are taken from the surface and starts from January 1999 for chlorophyll and June 2000 for nitrogen. For station PAP, error statistics are taken from 2002-2004 since in situ data is only available during that time.

Stations		Nitrogen profile				Chlorophyll profile				Surface chlorophyll				Integrated chlorophyll			
		r	RMSE	Bias	Mean	r	RMSE	Bias	Mean	r	RMSE	Bias	Mean	r	RMSE	Bias	Mean
PAP	Ens mean	0.23	3.26	0.61	6.59	0.42	0.32	0.06	0.48	0.45	0.51	0.22	0.66				
		(±0.07)	(±2.57)	(±5.13)	(±5.24)	(±0.37)	(±0.73)	(±0.68)	(±0.75)	(±0.38)	(±0.73)	(±0.68)	(±0.76)				
	Ens median	0.23	**3.16**	0.54	6.38	0.49	**0.29**	0.003	0.42	0.54	**0.46**	0.15	0.60				
	Default run	0.21	3.32	-0.20	5.64	0.28	0.40	0.18	0.59	0.36	0.57	0.30	0.74				
	In situ				5.83				0.42				0.44				
	NRR		1.25				1.20				1.29						
ALOHA	Ens mean	0.77	**1.06**	0.67	1.20	0.22	**0.10**	-0.06	0.06	0.22	**0.05**	-0.01	0.10	0.69	**2.73**	-0.72	3.80
		(±0.03)	(±0.19)	(±0.39)	(±0.39)	(±0.49)	(±0.04)	(±0.11)	(±0.11)	(±0.47)	(±0.09)	(±0.13)	(±0.14)	(±0.60)	(±5.49)	(±7.09)	(±10)
	Ens median	0.77	**1.06**	0.68	1.18	0.14	0.11	-0.07	0.05	0.13	**0.05**	-0.01	0.07	0.56	3.3	-1.17	3.34
	Default run	0.77	1.09	0.61	1.10	0.28	**0.10**	-0.03	0.09	0.27	0.07	0.03	0.11	0.70	4.71	1.25	5.77
	In situ				0.50				0.12				0.08				4.52
	NRR		1.39				1.29				1.07				1.01		
BATS	Ens mean	0.56	1.39	1.16	1.77	0.19	0.33	-0.12	0.05	0.22	0.33	-0.12	0.05	0.39	52.13	-19.39	6.18
		(±0.38)	(±0.84)	(±1.00)	(±1.01)	(±0.37)	(±0.05)	(±0.16)	(±0.16)	(±0.58)	(±0.15)	(±0.05)	(±0.15)	(±0.54)	(±9.40)	(±21)	(±14)
	Ens median	0.55	1.39	1.16	1.77	0.11	0.33	-0.12	0.05	0.06	0.34	-0.12	0.05	0.27	**23.30**	-17.71	4.51
	Default run	0.58	**0.73**	0.62	1.35	0.23	**0.31**	-0.07	0.10	0.28	**0.31**	-0.07	0.09	0.43	48.58	-10.77	13.14
	In situ				0.98				0.17				0.15				23.90
	NRR		1.38				1.39				1.40				1.40		
Cariaco	Ens mean	0.78	**2.97**	0.61	5.39	0.29	**0.83**	-0.02	0.49	0.13	**1.23**	0.02	0.77	0.41	**17.73**	-1.05	11.47
		(±0.08)	(±0.49)	(±2.54)	(±2.54)	(±0.34)	(±0.42)	(±0.93)	(±0.93)	(±0.22)	(±0.33)	(±1.90)	(±0.57)	(±0.40)	(±7.90)	(±17)	(±17)
	Ens median	0.76	3.24	0.51	5.29	0.20	0.88	-0.18	0.32	0.072	1.29	-0.29	0.46	0.29	19.46	-5.51	7.00
	Default run	0.76	3.29	0.59	5.37	0.22	0.87	-0.09	0.42	0.11	1.27	-0.18	0.57	0.34	18.71	-3.86	8.65
	In situ				4.78				0.51				0.76				12.52
	NRR		1.25				1.19				1.21				1.17		
L4	Ens mean	0.70	2.94	1.56	4.52					0.25	1.05	0.42	1.76				
		(±0.14)	(±2.13)	(±4.06)	(±4.06)					(±0.33)	(±1.67)	(±2.61)	(±2.61)				
	Ens median	0.68	3.10	1.73	4.69					0.21	**1.02**	0.27	1.61				
	Default run	0.52	**2.67**	1.12	4.08					0.31	1.13	0.83	2.17				
	In situ				2.96								1.34				
	NRR		1.31								1.21						

Table 4. Surface annual mean and phytoplankton phenology from in situ, ensemble mean, median, and default run. The range and NRR in the bracket are the values for changing the functional form one process at a time (shown on Fig. 8).

Stations		Annual Mean (mg m^{-3})	Initiation Time	Bloom (mg m^{-3})	Peak Height (mg m^{-3})	Amplitude (mg m^{-3})	Duration	Termination
PAP	Ens mean	0.61	01 Apr	07 May	2.07	0.96	95	26 Jul
	Range	±0.70(0.58)	±51	±45	±2.98	±1.63	±99	±124
	NRR	1.26 (1.37)	1.14	1.31	1.08	1.09	1.42	1.60
	Ens med	0.55	12 Apr	15 May	2.03	0.95	87	22 Jul
	Default run	0.71	03 Apr	05 May	2.1	0.96	99	21 Aug
	In situ	0.44	20 Apr	03 Jun	1.52	0.44	95	24 Jul
ALOHA	Ens mean	0.07	21 Mar	21 Apr	0.14	0.047	62	15 Aug
	Range	±0.13(0.11)	±89	±119	±0.28	±0.11	±95	±119
	NRR	0.84 (1.17)	1.35	1.29	0.97	1.19	1.56	1.28
	Ens med	0.063	26 Mar	02 May	0.14	0.05	85	24 Aug
	Default run	0.10	14 Mar	18 Apr	0.25	0.096	66	10 Aug
	In situ	0.084	08 May	26 May	0.14	0.048	47	23 Jun
BATS	Ens mean	0.047	02 Mar	12 Apr	0.1	0.043	89	06 Jul
	Range	±0.14(0.11)	±187	±174	±0.42	±0.19	±116	±198
	NRR	1.40 (1.39)	1.18	1.17	1.42	1.42	1.08	1.20
	Ens med	0.038	28 Feb	06 Apr	0.08	0.033	95	02 Aug
	Default run	0.091	06 Mar	25 Apr	0.29	0.13	65	19 Jun
	In situ	0.17	25 Feb	29 Mar	0.58	0.27	93	28 May
Cariaco	Ens mean	0.61	20 May	22 Jul	1.09	0.38	133	30 Sep
	Range	±1.53(1.29)	±101	±66	±2.61	±0.86	±63	±61
	NRR	0.78 (0.90)	1.48	1.40	1.39	1.42	1.88	1.55
	Ens med	0.37	22 May	14 Jul	0.83	0.34	110	06 Sep
	Default run	0.46	21 May	22 Jul	0.98	0.39	122	19 Sep
	In situ	0.61	16 Mar	21 Apr	2.39	1.15	76	01 Jun
L4	Ens mean	1.65	13 May	06 Jun	3.25	1.13	64	17 Aug
	Range	±2.48(2.14)	±100	±82	±3.12	±1.50	±78	±167
	NRR	1.00 (1.36)	1.49	1.42	1.32	1.48	1.22	1.19
	Ens med	1.49	18 May	07 Jun	3.09	1.13	70	18 Sep
	Default run	2.03	19 Apr	08 Jun	3.73	1.3	94	11 Aug
	In situ	1.20	09 Mar	11 Apr	3.58	1.64	80	28 May

Figure 1. Nearly identical curves which describes resource uptake (a), zooplankton grazing (b), and phytoplankton mortality (c). Figure (a) shows four uptake functions, which have been optimised to the default uptake function, monod (U_1). Figure (b) shows two grazing functional forms, the holling type III (G_1) and type II (G_2) functions. Four phytoplankton mortality functions are shown on figure (c), whereby hyperbolic is the default function. The optimisation method is describe in section 2.1, 2.2, and 2.3. Table 1 describes the function's equations and parameters.

Figure 2. SeaWIFs-derived mean 1998 chlorophyll-a (mg m^{-3}) overlain with the 5 oceanographic stations time series site (Red dots). These stations are located in different oceanic regions: oligotrophic (ALOHA and BATS), coastal (L4 and Cariaco), and abyssal plain (PAP).

Figure 3. Chlorophyll and nitrogen profiles from ensemble mean ((a) and (d) respectively), in situ observations ((c) and (f) for chlorophyll and nitrogen respectively), and 75th and 25th quartile range of concentrations at each depth ((b) for chlorophyll and (e) for nitrogen) at station PAP. The range are obtained by averaging the concentrations from all ensemble members for 10 years at each depths. Black dots in the second column show the mean concentration of the ensemble mean over the time series (from January 1998-December 2007). White solid line in (a) shows mixed layer depth.

Figure 4. Inter-annual mean of surface chlorophyll from all the study sites ((a)-(e)) and the 10-year annual mean (g), all measured in mg m^{-3}. The boxplots show the ensemble annual means. Blue cross is the in situ observation, red open circle, black dot, and blue stars are the ensemble mean, median, and the default run respectively. The blue box is the 75th (top) and 25th (bottom) quartiles. Red line is the median. The whiskers are the ensemble minimum and maximum mean of surface chlorophyll. Annual mean values and NRR are described in Table 4.

Figure 5. 10-year monthly mean surface chlorophyll from all the study sites ((a)-(e)), showing the seasonal dynamics of surface chlorophyll (mg m^{-3}). The boxplots show the ensemble seasonal means. Blue cross is the in situ observation, red open circle, black dot, and blue stars are the ensemble mean, median, and the default run respectively. The blue box is the 75th (top) and 25th (bottom) quartiles. The red line is the median. The whiskers are the ensemble minimum and maximum mean of surface chlorophyll. In station PAP, in situ data for December is not available due to low light and high cloud cover.

Figure 6. Inter-annual variability of averaged 200 m integrated nitrogen, from all the study sites ((a)-(e)), and the annual mean (f). Since the in situ data for PAP does not always cover the first 200m, the overall mean nitrogen concentration from all depth

is used instead. For station L4, in situ nitrogen is only collected on the surface. Blue cross is the in situ observation, red open circle, black dot, and blue stars are the ensemble mean, median, and default run respectively. The blue box is the 75th(top) and 25th(bottom) quartiles. Red line is the median, and the whiskers are the ensmeble minimum and maximum of the integrated nitrogen. In station L4 and PAP data for nitrogen is only available from 2000-2007 and 2002-2004 respectively.

Figure 7. 10-year monthly mean of averaged 200 m integrated nitrogen from all the study sites ((a)-(e)), showing the seasonal dynamics of nitrogen (mmol m^{-3}). For station PAP, the nitrogen shown is the overall profile, and in L4, the in situ nitrogen concentration is only available at the surface. The boxplot shows the ensemble monthly means. Blue cross is the in situ observation, red open circle, black dot, and blue stars are the ensemble mean, median, and the default run respectively. The blue box is the 75th (top) and 25th (bottom) quartiles. The red line is the median. The whiskers are the ensemble minimum and maximum mean of integrated nitrogen. In station PAP, the in situ data is only collected from 2002-2004 and L4 from 2000-2007.

Figure 8. Annual mean of surface chlorophyll when changing only one process at a time (blue box), overlain with annual mean of all ensemble members (green box) at five oceanographic stations. Ensemble mean and median plotted in the figure (shown in red open circle and black closed circle), are the from the 128 ensemble members.

Figure 9. Time series (from January 1998-December 2007) of ensemble mean and in situ, and range of chlorophyll and nitrogen concentrations at oligotrophic stations. Station ALOHA is shown on (a)-(f) and BATS is shown on (g)-(l). White solid line in (b) and (g) represents mixed layer depth. (b), (d), (h), and (j) are the 75th and 25th percentile range of chlorophyll ((b) for ALOHA and (h) for BATS) and nitrogen ((d) for ALOHA and (j) BATS) over the depth. The range is obtained by averaging the chlorophyll and nitrogen concentrations of each ensemble members over the time series at each depth. Black dots in (b), (d), (h), and (j) are the mean of the ensemble. Ensemble mean chlorophyll profiles (shown on (a) and (g)) and nitrogen ((e) and (k)) are obtained from all of the ensemble members. *In situ* chlorophyll are shown in (c) and (i), and nitrogen are shown in (g) and (l), for ALOHA and BATS respectively.

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Figure 12. 10-year mean and RMSE of surface chlorophyll (mg m^{-3}) at five stations from all ensemble members. The first panel ((a)-(e)) shows surface chlorophyll mean and RMSEs are shown on the second panel ((f)-(j)). Concentrations and RMSEs are arranged by the lowest chlorophyll (top left) mean to the highest (bottom right), depending on the oceanographic regions. For station PAP, the sequence is sorted based on coastal station. The y-axis shows combination of uptake ($U_1, U_2, U_3,$ and U_4)

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Figure 13. 10-year mean and RMSE of nitrogen (mmol m^{-3}), at five stations from all ensemble members. The first panel ((a)-(e)) shows nitrogen mean and RMSEs are shown on the second panel ((f)-(j)). Concentrations and RMSEs are arranged by the lowest chlorophyll (top left) mean to the highest (bottom right), depending on the oceanographic regions. For station PAP, the sequence is sorted based on coastal station. The y-axis shows combination of uptake ($U_1, U_2, U_3,$ and U_4) and grazing (G_1

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Figure 14. Time series of chlorophyll and nitrogen profile of ensemble mean, their range, and in situ concentrations at the coastal stations Cariaco (a-f) and L4 (g-h) from January 1998-December 2007. (a) and (d) show chlorophyll and nitrogen ensemble mean at Cariaco respectively. White solid line in (a) is the mixed layer depth. (b) and (e) shows the 75th and 25th percentile of chlorophyll and nitrogen concentrations at each depth. The black dots are the mean of the ensemble. These range

15 are obtained form the 10-year mean concentrations at each depth. Since in situ chlorophyll and nitrogen were taken at the surface in station L4, only surface time series were shown in (g-h). The grey shades on chlorophyll, shown in (g), and nitrogen, shown in (h) time series show 75th and 25th percentile of the range. Blue and red dots are in situ concentrations for chlorophyll and nitrogen respectively.

Figure 15. Phytoplankton phenology metrics at the five stations. Blue cross is the in situ, red, black, and blue dots are the

20 ensemble mean, median, and the default run respectively. The timings and concentrations are averaged annually from January 1998 to December 2007.

Figure A1. Determining phenology using a combination of threshold method and curve fit at station L4, here the initiation is when the fitted curve is above 50% of the maximum peak, however the termination is on the first valley.
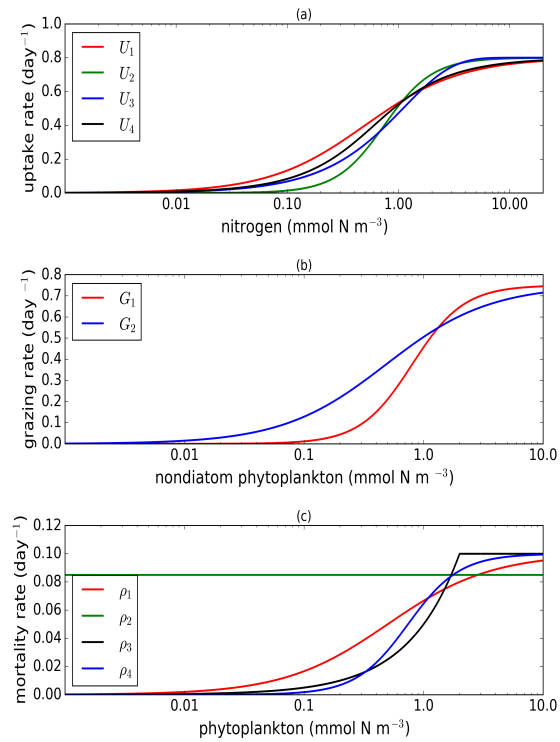
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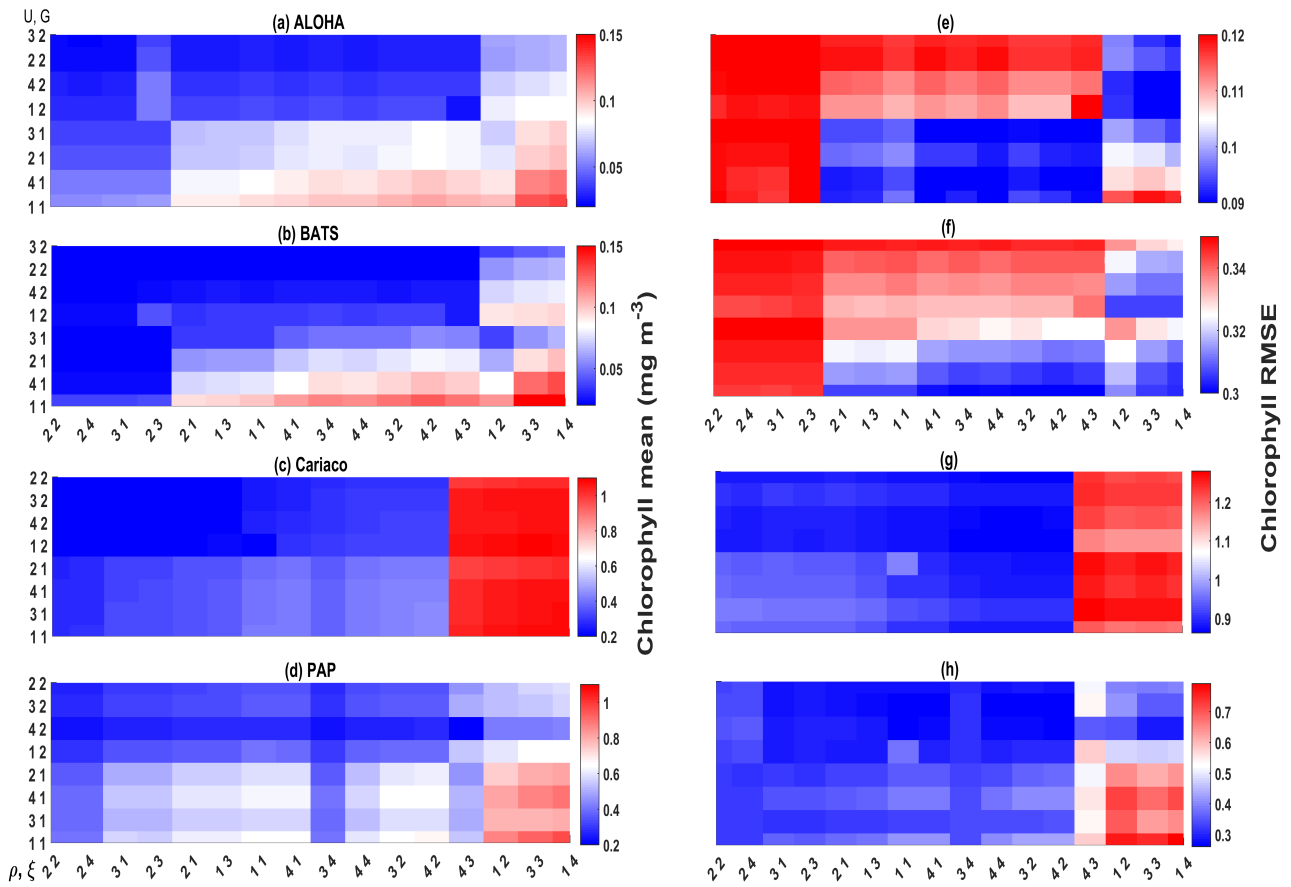
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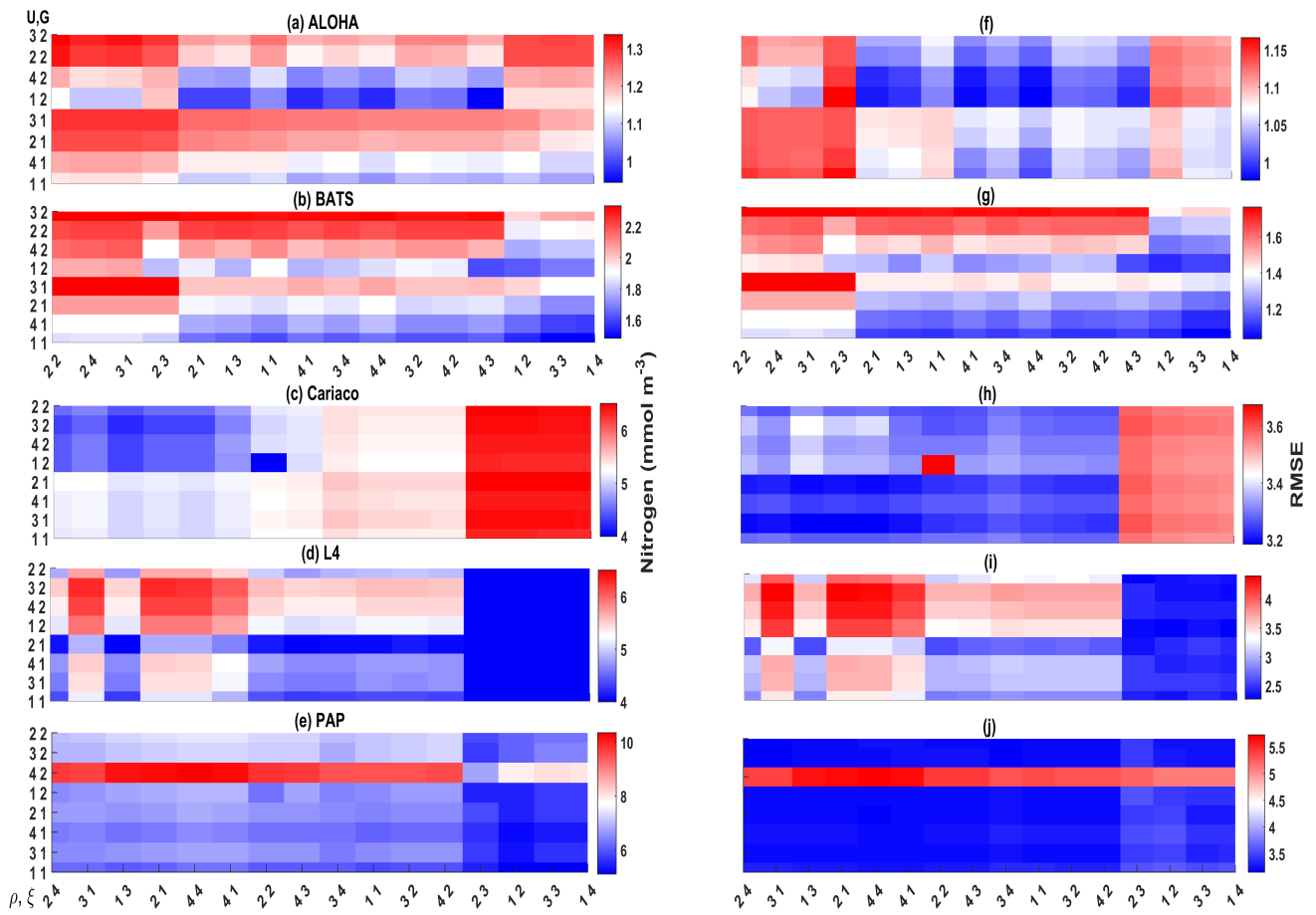
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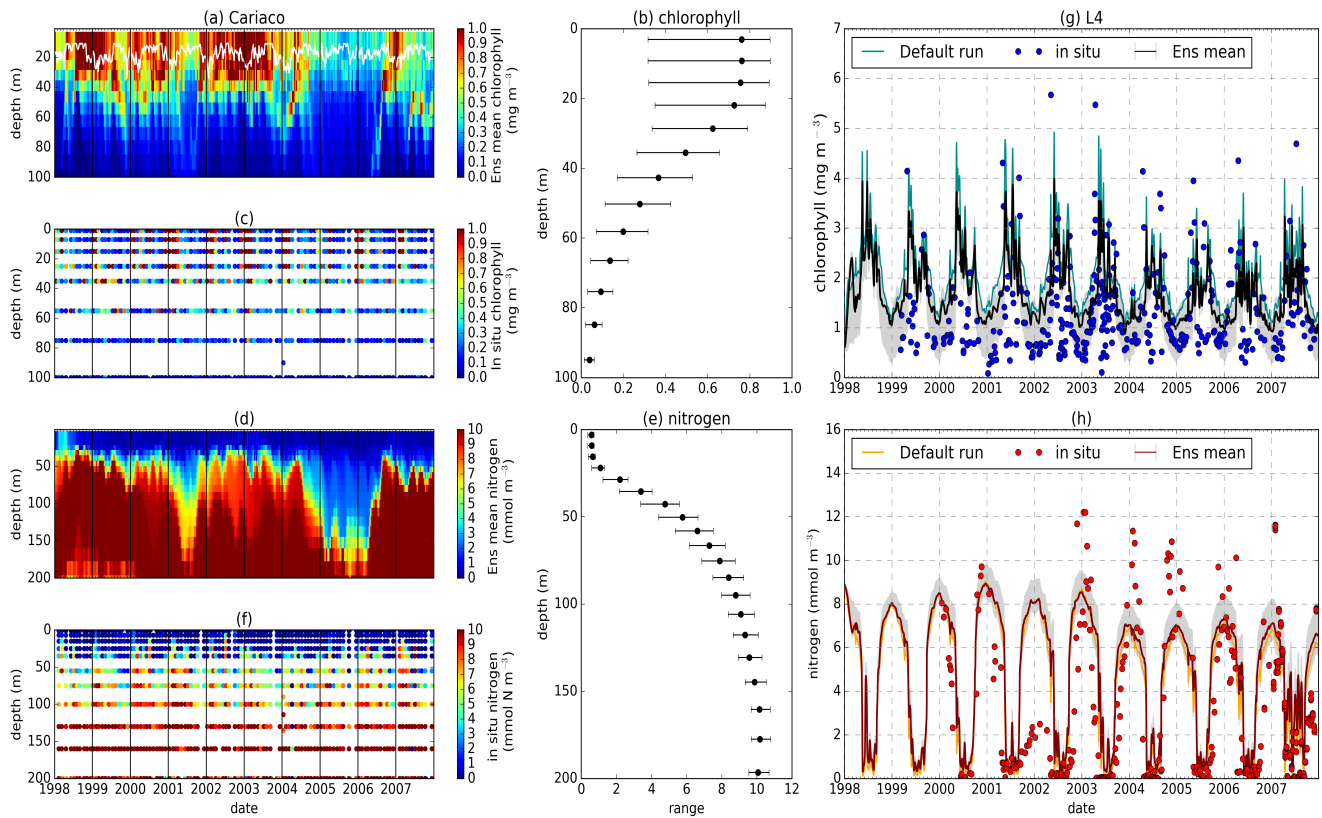
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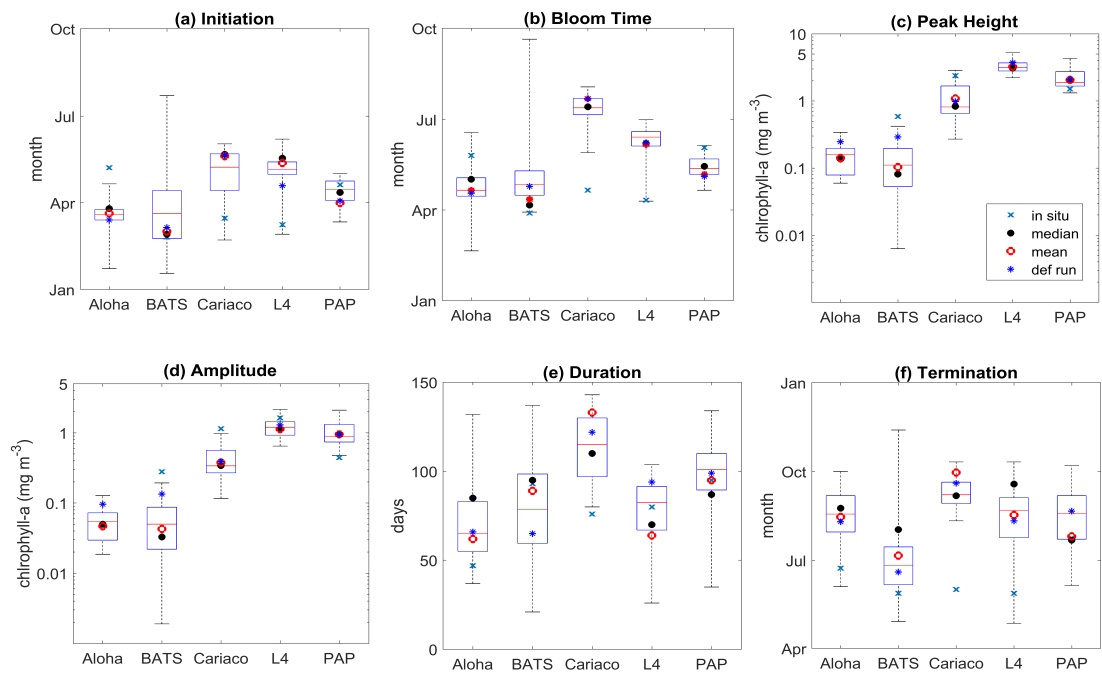
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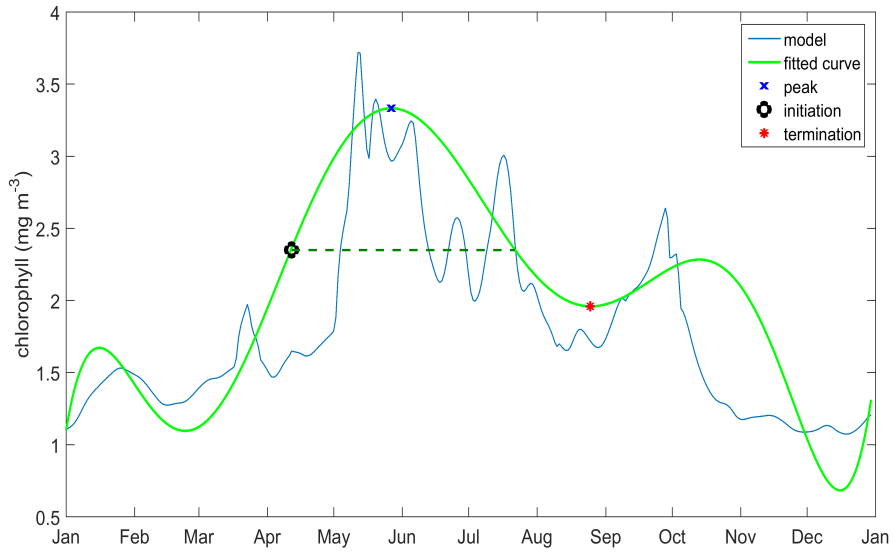
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