

Response to Associate Editor Decision: Reconsider after major revisions (03 Oct 2018) by David Butman

Comments to the Author:

1. After reviewing all of the interactive comments on bg-2018-227, it is recommended that the authors return to the calculated annual emissions and possibly remove this from the discussion. As outlined from all comments, the appropriate data sets may not exist to provide a meaningful annual emission estimate given the restricted sampling period. The discussion could be limited to monsoonal periods only. If the authors would like to keep this in, it is suggested that they develop a stronger presentation on what changes across seasons related to gas transfer and concentrations.

Response: Thank you for your helpful comment. According to your suggestion, annual emission estimate was deleted and CO₂ evasion upscaling was conducted during the monsoonal period in main text as follows.

“We used our measured CO₂ emission rates by FCs for upscaling flux estimates during monsoonal period given the sampling in this period and it was found to be 0.70 TgCO₂ for all rivers sampled in our study (Table 3a). The estimated emission in the monsoonal period was close to that of the revised model (0.71 ± 0.66 (95% confidence interval: 0.46-0.94) Tg CO₂), and using the determined k average, i.e., 0.69 ± 0.65 (95% confidence interval: 0.45-0.93) Tg CO₂, but slightly higher than the estimation using water-depth based model (0.54 ± 0.51 Tg CO₂) and Alin’s model (0.53 ± 0.50 Tg CO₂) (Table 3b). The higher emission, i.e., 1.66 ± 1.55 (1.08-2.23) Tg CO₂, in the monsoonal period only using flow velocity based model may be over-estimated when compared to other models, flux from determined k (Table 3b) and previous annual estimates, i.e., our earlier annual evasion of 0.64-2.33 Tg CO₂/y using TBL on the TGR river networks (Li *et al.*, 2018). Moreover, our estimated CO₂ emission in the monsoonal period also suggests that CO₂ annual emissions from rivers and streams in this area were previously underestimated, i.e., 0.03 Tg CO₂/y (Li *et al.*, 2017) and 0.37-0.44 Tg CO₂/y (Yang *et al.*, 2013) as the former used TBL model with a lower k level, and the latter employed floating chambers, but they both sampled very limited tributaries (i.e., 2-3 rivers). Therefore, measurements of k must be made mandatory along with pCO₂ measurement in the river and stream studies.” (Last paragraph in section 4.3)

Further, we strongly discussed the monsoonal sampling effects on gas transfer and pCO₂ concentrations, as well as annual evasion in the SOM. This could help do the comparison with other studies.

Seasonal changes related to gas transfer and concentrations

We used our measured CO₂ emission rates by FCs for upscaling flux estimates and it was found to be 1.39 TgCO₂/y for all rivers sampled (Table S3a). The estimated

emission was close to that of the revised model (1.40 ± 1.31 (95% confidence interval: 0.91-1.87) Tg CO₂/y), and using the determined k average, i.e., 1.37 ± 1.28 (95% confidence interval: 0.89-1.84) Tg CO₂/y, but slightly higher than the estimation using water-depth based model (1.08 ± 1.01 Tg CO₂/y) and Alin's model (1.06 ± 1.00 Tg CO₂/y) (Table S3b). The estimate was within the range of our earlier work using TBL on the TGR river networks (0.64-2.33 Tg CO₂/y) (Li *et al.*, 2018). The higher emission, i.e., 3.29 ± 3.08 (2.15-4.43) Tg CO₂/y, using flow velocity based model may be over-estimated (Table 3b). Therefore, this study suggests that CO₂ emissions from rivers and streams in this area may be underestimated, i.e., 0.03 Tg CO₂/y (Li *et al.*, 2017) and 0.37-0.44 Tg CO₂/y (Yang *et al.*, 2013) as the former used TBL model with a lower k level, and the latter employed floating chambers, but they both sampled very limited tributaries (i.e., 2-3 rivers). Therefore, measurements of k must be made mandatory along with pCO₂ measurement in the river and stream studies.

As our sampling was limited to monsoonal periods only, which could not provide a meaningful annual emission estimate given the restricted sampling period. Thus, we developed a stronger discussion on what changes across seasons related to gas transfer and pCO₂ concentrations. As outlined in the main text, riverine pCO₂ in the monsoonal season in this region was 81% the level in the dry season, and current velocity was 1.7-fold higher in monsoonal season (Li *et al.*, 2018), thus k₆₀₀ was 1.6-fold higher in the monsoonal period based our model. This could be defensible due to that prior study on the Yellow River reported that k₆₀₀ level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*, 2015), another study on the Wuding River demonstrated that k level in the wet season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Moreover, a factor of 1.4 for water area was designated based on other monsoonal rivers. Then annual emission could be estimated at 1.59 Tg CO₂/y, slightly higher than the estimation using the data in the monsoonal period only.

Table S3. CO₂ emission from total rivers sampled in the study.

(a) Upscaling using CO₂ areal flux by FC.

	Catchment Area km ²	Water surface km ²	CO ₂ areal flux mmol/m ² /d	CO ₂ emission Tg CO ₂ /y
Daning	4200	21.42	122.0 ± 239.4	0.042
Qijiang	4400	30.8	50.3 ± 177.2	0.025
TGR river	50000	377.78	217.7 ± 334.7	1.321
Total				1.39

(b) Upscaling using determined k₆₀₀ average and models (whole dataset are used here).

From determined k ₆₀₀ mean	Flow velocity-based model (Fig. 4b) (numbers in bracket is from the revised model; Fig. 4c)	Water depth-based model (Fig. 4a)	Alin's model

Mean		1.37	3.29 (1.40)	1.08	1.06
S.D.		1.28	3.08 (1.31)	1.01	1.00
95% CI	Lower	0.89	2.15 (0.91)	0.71	0.69
for	Bound				
Mean					
	Upper	1.84	4.43 (1.87)	1.46	1.43
	Bound				

A total water area of approx. 430 km² for all tributaries (water area is from Landsat ETM+ in 2015).

Table S4. Monsoonal sampling effects on annual emission

	Monsoonal season	Dry season	Monsoonal/Dry
pCO ₂ (µatm)	846	1043	0.81
ΔpCO ₂ (µatm)	446	643	0.69
k ₆₀₀ (cm/h)	48.4	31.3	1.55
CO ₂ areal flux (mmol/m ² /d)	196.9	183.4	
Water area (km ²)	602	430	
Emission (Tg CO ₂)	0.96	0.63	

2. Furthermore, it is recommended that the authors provide a stronger discussion of the bias that may occur using static chambers in a river environment. There are estimates of this bias in the literature and those values should be referenced. This method is concerning, unless in very small streams with significant turbulence induced from both surface and subsurface features.

Response: We presented strong discussion (please refer to section 4.4), and the following text was added.

Efforts have been devoted to measurement techniques (comparison of FC, eddy covariance-EC and boundary layer model-BLM) for improving CO₂ quantification from rivers because of a notable contribution of inland waters to the global C budget (), which could have a large effect on the magnitude of the terrestrial C sink. Prior studies reported inconsistent trends of CO₂ area flux by these methods. For instance, CO₂ areal flux from FC was much lower than EC (Podgrajsek et al., 2014), while areal flux from FC was higher than both EC and BLM elsewhere (Erkkilä et al., 2018), however, Schilder et al (2013) demonstrated that areal flux from BLM was 33-320% of in-situ FC measurements. Albeit unsatisfied errors of varied techniques and additional perturbations from FC exist, FC method is currently a simple and preferred measurement for CO₂ flux because that choosing a right k value remained a major challenge and others require high workloads (Martinsen et al., 2018).

Recent study further reported fundamental differences in CO₂ emission rates between ACs and freely DFs (Lorke *et al.*, 2015), i.e., ACs biased the gas areal flux higher by a factor of 2.0-5.5. However, some studies observed that ACs showed

reasonable agreement with other flux measurement techniques (Galfalk *et al.*, 2013), and this straightforward, inexpensive and relatively simple method AC was widely used (Ran *et al.*, 2017). Water-air interface CO₂ flux measurements were primarily made using ACs in our studied streams and small rivers because of relatively high current velocity; otherwise, floating chambers will travel far during the measurement period. In addition, inflatable rings were used for sealing the chamber headspace and submergence of ACs was minimal, therefore, our measurements were potentially overestimated but reasonable. We could not test the overestimation of ACs in this study, the modified FCs, i.e., DCs and integration of ACs and DCs, and multi-method comparison study including FCs, ECs and BLM should be conducted for a reliable chamber method.

3. Furthermore, the inclusion of data that is not relevant to the calculation and interpretation of k and CO₂ emissions was identified as distracting. This is in reference to the DOC/TN/TP. As outlined in two of the reviews, either identify why these are included (DOC is useful to know that you are not overestimating Alkalinity from organic acids) or remove from the manuscript.

Response: The parts of TN and TP were removed.

4. This effort will provide new knowledge and data from understudied rivers in SE Asia. That alone is a strong contribution. After addressing in detail these and the reviewer comments, this would be suitable for publication.

Response: Thanks for your very positive comment.

END of Review

Response to Anonymous Referee #1

1. General comments: The paper “Gas transfer velocities of CO₂ in subtropical monsoonal climate streams and small rivers” appears to be something of a companion piece to “Riverine CO₂ supersaturation and outgassing in a subtropical monsoonal mountainous area (Three Gorges Reservoir Region) of China” <https://doi.org/10.1016/j.jhydrol.2018.01.057> published in the Journal of Hydrology. In this current submission, the authors present k calculated from floating chamber flux measurements and using models, and discuss the implications of the differing approaches to k for making regional scale flux estimates. Using chambers to determine CO₂ fluxes, the authors then use $p\text{CO}_2$ to derive the gas transfer velocity. These flux-derived k values are compared to modeled k values. It is good to see the spatial aspects of the gas transfer velocity addressed. However, I do not feel that there is an adequate consideration of the uncertainty in the estimates/calculations provided. For the flux-derived k values, there is little provided in terms of uncertainty assessments.

Response: We thanked the referee for the comment. In our previous article, we studied the $p\text{CO}_2$ and emission rate as well as their controls from fluvial networks in the TGR area, which is based on two field works in the TGR region, and the diffusive models from other studies were used. In this study, we attempted to derive k levels and develop the gas transfer model in this area (mountainous streams and small rivers) for more accurate quantification of CO₂ areal flux, and also to serve for the fluvial networks in the Yangtze River or others with similar hydrology and geomorphology. In addition, we did more detailed field study in the two contrasting rivers Daning and Qijiang for developing models (see the sampling locations map). This study clearly showed original contribution to the current literature and this study is different than the article published in the Journal of Hydrology. We clearly state the new contributions and significances in the last paragraph of the “Introduction” as follows.

“Our recent study preliminarily investigated $p\text{CO}_2$ and air – water CO₂ areal flux as well as their controls from fluvial networks in the Three Gorges Reservoir (TGR) area (Li *et al.*, 2018). The past study was based on two field works, and the diffusive models from other continents were used. In this study, we attempted to derive k levels and develop the gas transfer model in this area (mountainous streams and small rivers) for more accurate quantification of CO₂ areal flux, and also to serve for the fluvial networks in the Yangtze River or others with similar hydrology and geomorphology. Moreover, we did detailed field campaigns in the two contrasting rivers Daning and Qijiang for models (Fig. 1). The study thus clearly stated distinct differences than the previous study (Li *et al.*, 2018) by the new contributions of specific objectives and data supplements, as well as wider significance.”

We added a section (4.4.) for “**Uncertainty assessment of $p\text{CO}_2$ and flux-derived k values**” in the part of “Discussion”.

The uncertainty of flux-derived k values mainly stem from air–water gradient of CO_2 ($\Delta p\text{CO}_2$ in ppm) and flux measurements (Golub et al., 2017; Lorke et al., 2015; Bodmer et al., 2016). Thus we provided uncertainty assessments caused by dominant sources of uncertainty from measurements of aquatic $p\text{CO}_2$ and CO_2 areal flux since uncertainty of atmospheric CO_2 measurement could be neglected.

In our study, aquatic $p\text{CO}_2$ was computed based on pH, alkalinity and water temperature rather than directly measured. Recent studies highlighted $p\text{CO}_2$ uncertainty caused by systematic errors over empiric random errors (Golub et al., 2017). Systematic errors are mainly attributed to instrument limitations, i.e., sondes of pH and water temperature. The relative accuracy of temperature meters was ± 0.1 °C according to manufacturers' specifications, thus the uncertainty of water T propagated on uncertainty in $p\text{CO}_2$ was minor (Golub et al., 2017). Systematic errors therefore stem from pH, which has been proved to be a key parameter for biased $p\text{CO}_2$ estimation calculated from aquatic C system (Li et al., 2013; Abril et al., 2015). We used a high accuracy of pH electrode and the pH meters were carefully calibrated using CRMs, and in situ measurements showed an uncertainty of ± 0.01 . We then run an uncertainty of ± 0.01 pH to quantify the $p\text{CO}_2$ uncertainty, and an uncertainty of $\pm 3\%$ was observed. Systematic errors thus seemed to show little effects on $p\text{CO}_2$ errors in our study.

Random errors are from repeatability of carbonate measurements. Two replicates for each sample showed the uncertainty of within $\pm 5\%$, indicating that uncertainty in $p\text{CO}_2$ calculation from alkalinity measurements could be minor.

The measured pH ranges also exhibited great effects on $p\text{CO}_2$ uncertainty (Hunt et al., 2011; Abril et al., 2014). At low pH, $p\text{CO}_2$ can be overestimated when calculated from pH and alkalinity (Abril et al., 2014). Samples for CO_2 fluxes estimated from pH and alkalinity showed pH average of 8.39 ± 0.29 (median 8.46 with quartiles of 8.24-8.56) ($n=115$). Thus, overestimation of calculated CO_2 areal flux from pH and alkalinity is likely to be minor. Further, contribution of organic matter to non-carbonate alkalinity is likely to be neglected because of low DOC (mean 6.67 mg/L; median 2.51 mg/L) (Hunt et al., 2011; Li et al., 2013).

Efforts have been devoted to measurement techniques (comparison of FC, eddy covariance-EC and boundary layer model-BLM) for improving CO_2 quantification from rivers because of a notable contribution of inland waters to the global C budget (), which could have a large effect on the magnitude of the terrestrial C sink. Prior studies reported inconsistent trends of CO_2 area flux by these methods. For instance, CO_2 areal flux from FC was much lower than EC (Podgrajsek *et al.*, 2014), while areal flux from FC was higher than both EC and BLM elsewhere (Erkkila *et al.*, 2018), however, Schilder et al (Schilder *et al.*, 2013) demonstrated that areal flux from BLM was 33-320% of in-situ FC measurements. Albeit unsatisfied errors of varied techniques and additional perturbations from FC exist, FC method is currently a simple and preferred measurement for CO_2 flux because that choosing a right k value remains a major challenge and others require high workloads (Martinsen *et al.*, 2018).

Recent study further reported fundamental differences in CO₂ emission rates between ACs and freely DFs (Lorke *et al.*, 2015), i.e., ACs biased the gas areal flux higher by a factor of 2.0-5.5. However, some studies observed that ACs showed reasonable agreement with other flux measurement techniques (Galfalk *et al.*, 2013), and this straightforward, inexpensive and relatively simple method AC was widely used (Ran *et al.*, 2017). Water-air interface CO₂ flux measurements were primarily made using ACs in our studied streams and small rivers because of relatively high current velocity; otherwise, floating chambers will travel far during the measurement period. In addition, inflatable rings were used for sealing the chamber headspace and submergence of ACs was minimal, therefore, our measurements were potentially overestimated but reasonable. We could not test the overestimation of ACs in this study, the modified FCs, i.e., DCs and integration of ACs and DCs, and multi-method comparison study including FCs, ECs and BLM should be conducted for a reliable chamber method.

Sampling seasonality considerably regulated riverine pCO₂ and gas transfer velocity and thus water-air interface CO₂ evasion rate (Li *et al.*, 2012; Ran *et al.*, 2015). We sampled waters in wet season (monsoonal period) due to that it showed wider range of flow velocity and thus it covered the k₆₀₀ levels in the whole hydrological season. Wet season generally had higher current velocity and thus higher gas transfer velocity (Ran *et al.*, 2015), while aquatic pCO₂ was variable with seasonality. We recently reported that riverine pCO₂ in the wet season was 81% the level in the dry season (Li *et al.*, 2018), and prior study on the Yellow River reported that k level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*, 2015), while another study on the Wuding River demonstrated that k level in the wet season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Thus, we acknowledged a certain amount of errors on the annual flux estimation from sampling campaigns during the wet season in the TGR area, while this uncertainty could not be significant because that the diluted pCO₂ could alleviate the overestimated emission by increased k level in the wet season (stronger discussion please refer to SOM).

2. pCO₂ was not measured, but rather was computed based on pH, alkalinity and temperature. This would have large uncertainties that then propagate into k estimates. Golub et al. (2017, doi: 10.1002/2017JG003794) note that “freshwater researchers must make significant efforts to standardize and reduce errors in pCO₂ predictions”. I encourage the authors to undertake a more systematic uncertainty analysis for their pCO₂ values and propagate this error into uncertainty estimates for k.

Response: Thanks for the comment. We reported the quality control such as systematic errors and random errors of pH and alkalinity, water temperature, as well as non-carbonate alkalinity effects. Please refer to the response above.

3. Further, the authors here excluded deriving k values for samples that did not have a very large gradient in CO₂ across the air-water interface. The authors chose 110 uatm as the threshold for excluding data, but this was presented without any indication of choice of threshold, making it appear rather arbitrary. Given the pH of the rivers sampled and the pCO₂ that was at times undersaturated, this appears rather problematic

in that it introduces bias that carries through to the regional estimates provided.

Response: We addressed this issue as follows.

“Prior to statistical analysis, we excluded k_{600} data for samples with the air-water $p\text{CO}_2$ gradient $<110 \mu\text{atm}$, since the error in the k_{600} calculations drastically enhances when $\Delta p\text{CO}_2$ approaches zero (Borges et al., 2004), and datasets with $\Delta p\text{CO}_2 >110 \mu\text{atm}$ provide an error of $<10\%$ on k_{600} computation (see the Fig. as follows)”

The additional section 4.4 **Uncertainty assessment of $p\text{CO}_2$ and flux-derived k values** included uncertainty of pH and scaling-up estimation, for example, effects of chemical enhancement for quite high pH values.

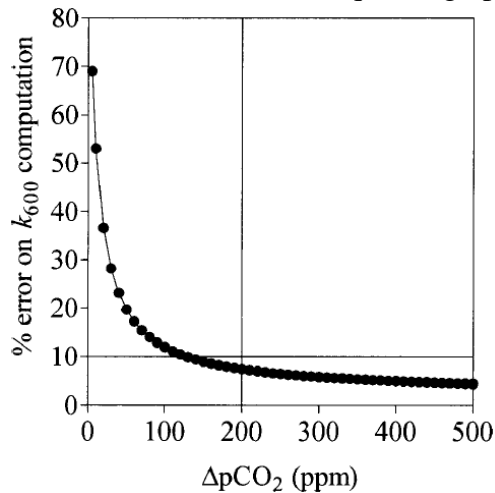


Figure Theoretical error ($\pm\%$) on the computation of the gas transfer velocity of CO_2 (k_{600}) as a function of the air–water gradient of CO_2 ($\Delta p\text{CO}_2$ in ppm), assuming a constant uncertainty on $\Delta p\text{CO}_2$ of $\pm 3\%$ (Borges et al., 2004).

Borges, A.V., Delille, B., Schiettecatte, L.S., Gazeau, F., Abril, G., Frankignoulle, M., 2004a. Gas transfer velocities of CO_2 in three European estuaries (Randers Fjord, Scheldt, and Thames). *Limnology & Oceanography* 49, 1630-1641.

4. The authors in this paper refer to their k values as “observed”, but these are in fact derived, and so need to have uncertainty better characterized. Upscaling from X floating chamber measurements to a river network draining 58000 km². How many flux measurements were made with floating chambers is not clearly stated, but it appears to be about 100 all made during summer 2016. Going from summer measurements for 100 points to annual estimates for 58000 km² also requires some consideration of error propagation and bias. Fluxes were only retained when the floating chambers yielded linearly increasing CO_2 against time, which again biases against low flux locations.

Response: We have changed “observed, or measured” to “flux-derived or derived”, and discussed the uncertainty of k values as mentioned above. We agreed that more samples could improve the CO_2 estimates, while our sampling locations were much more or at least comparable to the other publications.

A total of 115 discrete grab samples were collected (each sample consisted of three replicates). Floating chambers with replicates were deployed in 101 sites (32 sampling sites in Daning, 37 sites in TGR river networks and 32 sites in Qijiang). The sampling period covers spring and summer season, our sampling points are reasonable considering a water area of 433 km². For example, 16 sites were collected for Yangtze system to examine hydrological and geomorphological controls on $p\text{CO}_2$ (i.e., Liu et al., 2017), and 17 sites for dynamic biogeochemical controls on riverine $p\text{CO}_2$ in the Yangtze basin (Liu et al., 2016, Global Biogeochemical Cycles).

In our sampling points, all measured fluxes were retained since the floating chambers yielded linearly increasing CO_2 against time following manufacturer's specification.

Liu, S., Lu, X.X., Xia, X., Yang, X., Ran, L., 2017. Hydrological and geomorphological control on CO_2 outgassing from low-gradient large rivers: An example of the Yangtze River system. *Journal of Hydrology* 550, 26-41.

5. Of the attempted flux measurements, what fraction was discarded?

Response: We revised as follows.

“Prior to statistical analysis, we excluded k_{600} data for samples with the air-water $p\text{CO}_2$ gradient $<110 \mu\text{atm}$, since the error in the k_{600} calculations drastically enhances when $\Delta p\text{CO}_2$ approaches zero (Borges *et al.*, 2004; Alin *et al.*, 2011), and datasets with $\Delta p\text{CO}_2 >110 \mu\text{atm}$ provide an error of $<10\%$ on k_{600} computation. Thus, we discarded the samples (36.7% of sampling points with flux measurements) with $\Delta p\text{CO}_2 <110 \mu\text{atm}$ for k_{600} model development, while for the flux estimations from diffusive model and floating chambers, all samples were included.”

6. Finally, a minor point is that the authors state several times that theirs is the first determination of k for subtropical streams and small rivers. I would point the authors towards global syntheses on CO_2 evasion as well as individual studies that include k estimates.

Response: We agree the comment, and addressed this issue. Several sentences for implication of k determination and comparison with other k studies were added in the part of Discussion. We also re-organised and added sub-headings of “Discussion”.

“4.1. Determined k values relative to world rivers; 4.2. Hydraulic controls of k_{600} ; 4.3. Implications for large scale estimation; 4.4. Uncertainty assessment of $p\text{CO}_2$ and flux-derived k values”

Minor comments

7. Finally, a minor point is that the authors state several times that theirs is the first determination of k for subtropical streams and small rivers. I would point the authors towards global syntheses on CO_2 evasion as well as individual studies that include k

estimates.

Response: We have presented k levels related to global rivers (please refer to 4.1) and implications of k for large scale estimation (see section 4.3).

8. The figure S1 does not show the sample locations within the Daning or Qijiang basins. These may be the same locations as Figure 1 in Li et al. (2018) Journal of Hydrology doi: 10.1016/j.jhydrol.2018.01.057?

Response: The sampling sites and study aims are different than previous study (Please refer to the section of 2.1 and the last paragraph in the “Introduction”). In the revised Ms, we supplied the map of sampling locations in the main text as Fig. 1.

We added several sentences in the section of “INTRODUCTION” to highlight the differences between our study and previous study, as well as what is advanced by this study (please refer to the first Comment).

9. There are a number of grammatical issues throughout the paper that the authors should address.

Response: We carefully edited English, and also get helps from a native English scientist.

The additional Table was added to the Table S2 in SOM.

	Current velocity m/s	Water depth m/s	Wind speed m/s	k_{600} cm/h	Reference
Mekong tributary	0.39±0.28	0.9±0.6	0.7±0.6	23.3±17.3	Alin et al., 2011
Yellow	1.8		1.8 (1.2-2.3)	42±17	Ran et al., 2015
Yangtze	1.2±1.5		1.2±1.1	38±40	Liu et al., 2017
Mekong stem	0.92±0.42		1.8±1.2	15±9	Alin et al., 2011

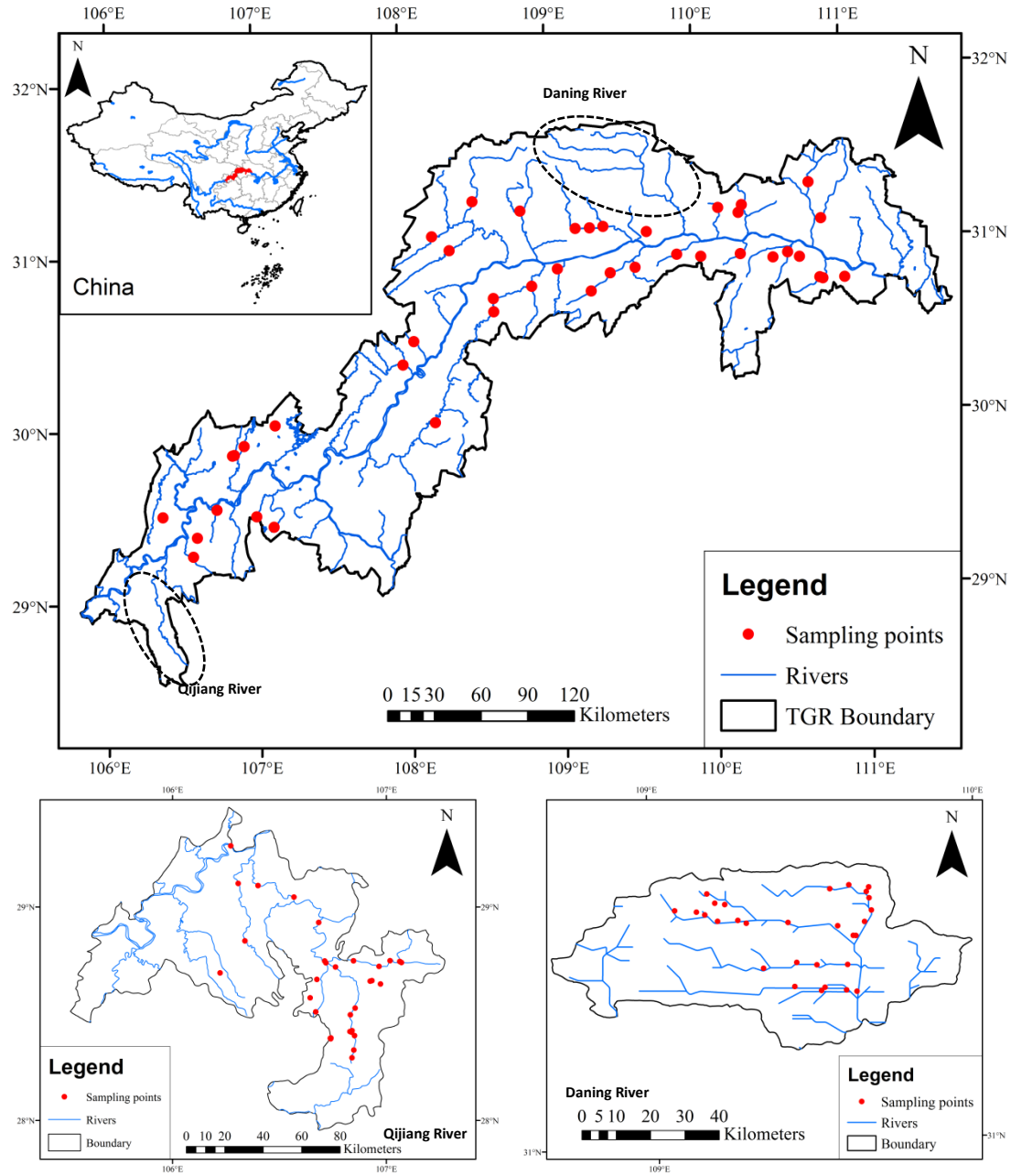


Fig. 1. Map of sampling locations of major rivers and streams in the Three Gorges Reservoir region, China (main text).

Response to Anonymous Referee #2

The manuscript reports on transfer velocities of CO₂ (K) in streams and small rivers for assessing the gas fluxes. CO₂ released from lakes and rivers has been recently recognized as an important component in the global carbon cycle. The accurate estimation of CO₂ flux is still challenging primarily due to the difficulty in obtaining an appropriate K value. The topic would be of great interest for the community of scientists working on carbon cycles and can be considered for publication. However, the current version need to revised (see below).

Response: We thanked the referee for the positive comment. We revised the Ms based on the comments and suggestions as follows.

General comments

1. As emphasized by the authors, the study focuses on the subtropical monsoonal streams and small rivers which are characterized by large seasonal variations in climate and discharge. Hence, the K value in these rivers should also have obvious seasonality. Unfortunately, the samples presented in this study were collected in the rainy season. The K value were calculated based on the one-time sampling campaign, which might result in a certain amount of errors on the annual flux estimation. Regarding this, the uncertainty of the sampling data and the calculations, as well as the reliability of the argument should be sufficiently discussed.

Response: We agreed the comment. Sampling seasonality largely impacted riverine $p\text{CO}_2$ and gas transfer velocity and thus water-air interface CO_2 evasion rate. In our Ms, we sampled waters in the rainy season due to that it showed wider range of flow velocity and thus rainy season covered the k levels in the whole hydrological season. Rainy season generally had higher current velocity and thus higher gas transfer velocity, while aquatic $p\text{CO}_2$ was variable with seasonality. Thus, we estimated CO_2 emission in monsoonal period instead of annual emission (please refer to section 4.3).

We added a section of “**4.4. Uncertainty assessment of $p\text{CO}_2$ and flux-derived k values**” in the part of “Discussion”. In this section, effects and uncertainty of sampling seasonality on errors of annual CO_2 flux estimation were included.

Sampling seasonality considerably regulated riverine $p\text{CO}_2$ and gas transfer velocity and thus water-air interface CO_2 evasion rate (Li *et al.*, 2012; Ran *et al.*, 2015). We sampled waters in wet season (monsoonal period) due to that it showed wider range of flow velocity and thus it covered the k_{600} levels in the whole hydrological season. Wet season generally had higher current velocity and thus higher gas transfer velocity (Ran *et al.*, 2015), while aquatic $p\text{CO}_2$ was variable with seasonality. We recently reported that riverine $p\text{CO}_2$ in the wet season was 81% the level in the dry season (Li *et al.*, 2018), and prior study on the Yellow River reported that k level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*, 2015), while another study on the Wuding River demonstrated that k level in the wet season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Thus, we acknowledged a certain amount of errors on the annual flux

estimation from sampling campaigns during the wet season in the TGR area, while this uncertainty could not be significant because that the diluted $p\text{CO}_2$ could alleviate the overestimated emission by increased k level in the wet season (stronger discussion please refer to SOM).

2. *In my point of view, the variation in K value of the rivers studied are obvious and need to be discussed. In addition, the spatial difference of K values is only sorted out for the three river systems (Daning, Qijiang and TGR). I would suggest the authors examine the variations of K following the physical characteristics of rivers (such as the current velocity, slope and the water depth) or/and the river orders.*

Response: Spatial differences of k values were discussed for the three rivers systems (please refer to section 4.2). We discussed the controls of physical characteristics of current velocity, slope and the water depth, while river orders were not extracted.

“This could substantiate the higher k_{600} levels and spatial changes in k_{600} values of our three river systems. For instance, similar to other turbulent rivers in China (Ran *et al.*, 2015; Ran *et al.*, 2017), high k_{600} values in the TGR, Daning and Qijiang rivers were due to mountainous terrain catchment, high current velocity (10 – 150 cm/s) (Fig. 4b), bottom roughness, and shallow water depth (10 - 150 cm) (Fig. 4a). It has been suggested that shallow water enhances bottom shear, and the resultant turbulence increases k values (Alin *et al.*, 2011; Raymond *et al.*, 2012). These physical controls are highly variable across environmental types (Figs. 4a and 4b), hence, k values are expected to vary widely (Fig. 3). The k_{600} values in the TGR rivers showed wider range (1-177 cm/h; Fig. 3; Table S1), spanning more than 2 orders of magnitude across the region, and it is consistent with the considerable variability in the physical processes on water turbulence across environmental settings. Similar broad range of k_{600} levels was also observed in the China’s Yellow basin (ca. 0-123 cm/h) (Ran *et al.*, 2015; Ran *et al.*, 2017).

Contrary to our expectations, no significant relationship was observed between k_{600} and water depth, and current velocity using the entire data in the three (TGR, Daning and Qijiang) river systems (Fig. S4). There were not statistically significant relationships between k_{600} and wind speed using separated data or combined data, and it is consistent with earlier studies (Alin *et al.*, 2011; Raymond *et al.*, 2012). Flow velocity showed linear relation with k_{600} , and the extremely high value of k_{600} was observed during the periods of higher flow velocity (Fig. S4a) using combined data. Similar trend was also observed between water depth and k_{600} values (Fig. S4b). The lack of strong correlation between k_{600} and physical factors are probably due to combined effect of both flow velocity and water depth, as well as large diversity of channel morphology, both across and within river networks in the entire catchment (60, 000 km²). This is further collaborated by weak correlations between k_{600} and flow velocity in the TGR rivers (Fig. 4), where one or two samples were taken for a large scale examination. k_{600} as a function of water depth was obtained in the TGR rivers, but it explained only 30% of the variance in k_{600} . However, model using data from Qijiang could explain 68% of the variance in k_{600} (Fig. 4b), and it was in line with general theory. Nonetheless, k_{600} from our flow velocity based model (Fig. 4b) was potentially largely overestimated with consideration

of other measurements (Alin *et al.*, 2015; Ran *et al.*, 2015; Ran *et al.*, 2017). When several extremely values were removed, k_{600} (cm/h) was parameterized as follows ($k_{600} = 62.879FV + 6.8357$, $R^2 = 0.52$, $p = 0.019$, FV-flow velocity with a unit of m/s), and this revised model was in good agreement with the model in the river networks of the Yellow River (Ran *et al.*, 2017), but much lower than the model developed in the Yangtze system (Liu *et al.*, 2017) (Fig. 4c). This was reasonable because of k_{600} values in the Yangtze system were from large rivers with higher turbulence than Yellow and our studied rivers. Furthermore, the determined k_{600} using FCs was, on average, consistent with the revised model (Table 2). These differences in relationship between spatial changes in k_{600} values and physical characteristics further corroborated heterogeneity of channel geomorphology and hydraulic conditions across the investigated rivers.”

3. The pCO_2 calculated in this paper is between 50-4830ppm, which indicates that the river pCO_2 value is sometimes much lower than that of the atmosphere, that is, the studied rivers can sometimes absorb CO_2 from the atmosphere. It seems that the annual CO_2 flux for the whole basin was calculated in this paper based on the averaged K value from the observed results using floating chamber method. The question is that is it reliable to estimate both directions of the CO_2 flux at the air-water interface (including river CO_2 outgassed to the atmosphere and the atmospheric CO_2 input to rivers) by using the same K value? Or what uncertainty will it cause?

Response: We thanked for your critical comment. Worldwide studies reported the dependence of k on flow dynamics; while k average from statistical analysis with $\Delta pCO_2 > 110 \mu atm$ was normally used in riverine CO_2 flux estimation in rivers where a broad of range of pCO_2 occurred (Alin *et al.*, 2011). Considering that k was largely dependent on hydraulic characteristics, this uncertainty was not discussed. The method employed here for scaling up CO_2 estimation was widely used (e.g., Alin *et al.*, 2011). We added one section 4.4 (see Discussion part) on systematic errors and empiric random errors of pCO_2 , CO_2 areal flux and k from pCO_2 determinations, sampling seasonality, flux measurements etc.

4. This study measured DOC, DTN, and DTP, but the authors did not mention these measurements in the discussion section. What is the relationship between these variables and the K values?

Response: k values are reported to be dominated by physical characteristics of river systems (Borges *et al.*, 2004; Alin *et al.*, 2011; Raymond *et al.*, 2012). We re-wrote this part and pattern of nutrients in the “Result” section was removed.

Specific comments

5. L 94-97: I would suggest rephrase these sentences, since they cannot convey clearly the real contribution or scientific merit of this study.

Response: We revised as follows.

“Our new contributions to the literature include (1) determination and controls of k levels for small rivers and streams in subtropical areas of China, and (2) new k models

developed in the subtropical mountainous river networks”

6. L 111-112, 117: *The classification method of the river order used here should be clarified. The number of a river order defined by different classification system may represent different size or hierarchy of a river.*

Response: We addressed this issue, and provided the map of sampling location (Fig. 1). We also supplied methodology of water areal extraction in section 2.5.

7. L 214-216: *This statement is problematic. Clearly, the studied rivers are not always supersaturated reference to atmospheric CO₂ as the pCO₂ in rivers is between 50- 4830 uatm.*

Response: We changed to “pCO₂ varied between 50 and 4830 μ atm with mean of 846 ± 819 μ atm (Table 1). There were 28.7% of samples that had pCO₂ levels lower than 410 μ atm, while the studied rivers were overall supersaturated with reference to atmospheric CO₂ and act as a source for the atmospheric CO₂.”

8. L 274-285 *These arguments need more solid evidence to support. As mentioned in the general comments, I would suggest that the authors focus on discussions on relationship between spatial change in K values and physical characteristics of rivers or/ and the river orders.*

Response: Spatial differences of k values were discussed for the three rivers systems (please refer to section 4.2). We discussed the controls of physical characteristics of current velocity, slope and the water depth.

9. L 497-498 *The water area is a very critical parameter for the calculation of CO₂ flux in a basin, so the acquisition of water area is essential and should be described more in detail. For example, what is the resolution of the satellite image? In addition, the variation of surface area of water between wet and dry seasons should be considered.*

Response: We provided the information for the acquisition of water area and citation was included (please refer to Methodology section).

2.5. Estimation of river water area

Water surface is an important parameter for CO₂ efflux estimation, while it depends on its climate, channel geometry and topography. River water area therefore largely fluctuates with much higher areal extent of water surface particularly in monsoonal season. However, most studies do not consider this change, and a fraction of the drainage area is used in river water area calculation (Zhang *et al.*, 2017). In our study, a 90 m resolution SRTM DEM (Shuttle Radar Topography Mission digital elevation model) data and Landsat images in dry season were used to delineate river network, and thus water area (Zhang *et al.*, 2018), whilst, stream orders were not extracted. Water area of river systems is generally much higher in monsoonal season in comparison to dry season, for

instance, Yellow River showed 1.4-fold higher water area in the wet season than in the dry season (Ran *et al.*, 2015). Available dry-season image was likely to underestimate CO₂ estimation.

Ran, L.S., Lu, X.X., Yang, H., Li, L.Y., Yu, R.H., Sun, H.G., Han, J.T., 2015. CO₂ outgassing from the Yellow River network and its implications for riverine carbon cycle. *Journal of Geophysical Research-Biogeosciences* 120, 1334-1347.

Zhang, T., Li, J., Pu, J., Martin, J.B., Khadka, M.B., Wu, F., Li, L., Jiang, F., Huang, S., Yuan, D., 2017. River sequesters atmospheric carbon and limits the CO₂ degassing in karst area, southwest China. *Science of The Total Environment* 609, 92-101.

Zhang, J., Li, S.Y., Dong, R.Z., Jiang, C.S., 2018. Physical evolution of the Three Gorges Reservoir in Holocene using advanced SVM on Landsat images and SRTM DEM data. *Environ Sci Pollut Res* 25, 14911-14918.

10. Finally, I would suggest the authors polish the English grammar and writing, as well as the figs presenting.

Response: We carefully edited the Ms and re-organized the Tables and Figures presentation.

Response to Anonymous Referee #3

General comments

The manuscript of Li et al. presents measured CO₂ fluxes, transport coefficients based on CO₂, and calculated pCO₂ data of running waters in a subtropical monsoonal climate zone. These data are complemented by among others water chemistry parameters such as DOC, DTN, DTP, as well as hydrogeomorphology data (e.g. water depth, flow velocity). They provide data and insights about transport coefficients for a so far understudied region and highlight the spatial variability and subsequent uncertainty for regional upscale estimates.

By investigating the key parameter for CO₂ flux estimates - the transport coefficient - in an understudied region, Li et al. address a very relevant topic. Narrowing down the uncertainties of regional upscaling estimates of riverine CO₂ fluxes is of wide interest, hence this study would make a good contribution to the literature and the subject matter is thus of interest to Biogeosciences readers.

Response: We thank you for your overall positive comments, and accordingly revised the Ms.

However, in my opinion, the manuscript has some problems:

(1) The terminology used in this manuscript is quite confusing to me. It seems to me that "streams", "rivers", "river networks" are used interchangeably (without definition and consistency), which makes it hard to follow the red line of the story. The terminology needs to be clarified and unified.

Response: Based on the delineation of river systems by Alin et al., 2011 (JGR), small rivers and streams encompass rivers with channels < 100 m. We have clarified the term in the method.

(2) The sampling design is not very clear to me. All investigated running waters seem to be in the Three Gorges Reservoir (TGR) region, but in addition two larger streams (Daning and Qijiang) were sampled. In the results and discussion, these investigated running waters are combined, sometimes split, which makes it hard to follow (in the main text and tables). In my opinion, these three "regions" need to be presented in a unified way (always separated or combined, possibly both in each table and figure), and presented more clearly in the text.

Response: We provided both separated and combined data in Tables and Figures, please refer to Figs 2 and 3, as well as Figs. S2 and S3. In addition, we also clearly stated this in the "Method" part.

"Spatial differences (Daning, Qijiang and entire tributaries of TGR region) were tested using the nonparametric Mann Whitney U-test. Multivariate statistics, such as correlation and stepwise multiple linear regression, were performed for the models of k_{600} using potential physical parameters of wind speed, water depth, and current velocity from separated data and combined data (Alin et al., 2011)."

(3) One of the main messages is the presentation of transport coefficients in a subtropical monsoonal climate zone, which is interesting, but I can imagine that there is a large difference in the wet and dry season. However, all the measurements were done in the wet season. I think this issue should be clearly acknowledged and discussed.

Response: We agreed your opinion and addressed this issue in an additional section 4.4. We

sampled in the monsoonal period as it covered the flow velocity in the whole hydrological year. Albeit k_{600} is higher in wet season than dry season, our main objectives are to develop models of k rather than the annual evasion. In this section, effects and uncertainty of sampling seasonality on errors of annual CO_2 flux estimation were also included.

“Sampling seasonality considerably regulated riverine $p\text{CO}_2$ and gas transfer velocity and thus water-air interface CO_2 evasion rate (Li *et al.*, 2012; Ran *et al.*, 2015). We sampled waters in wet season (monsoonal period) due to that it showed wider range of flow velocity and thus it covered the k_{600} levels in the whole hydrological season. Wet season generally had higher current velocity and thus higher gas transfer velocity (Ran *et al.*, 2015), while aquatic $p\text{CO}_2$ was variable with seasonality. We recently reported that riverine $p\text{CO}_2$ in the wet season was 81% the level in the dry season (Li *et al.*, 2018), and prior study on the Yellow River reported that k level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*, 2015), while another study on the Wuding River demonstrated that k level in the wet season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Thus, we acknowledged a certain amount of errors on the annual flux estimation from sampling campaigns during the wet season in the TGR area, while this uncertainty could not be significant because that the diluted $p\text{CO}_2$ could alleviate the overestimated emission by increased k level in the wet season (stronger discussion please refer to SOM).”

(4) There are two technical issues: (i) The measurements with the floating chambers are poorly described. The only information Li et al. provide is that the floating chambers were “deployed”. If the flux measurements are done in an anchored or free floating manner is critical (see e.g. Lorke, A., Bodmer, P., Noss, C., Alshboul, Z., Koschorreck, M., Somlai-Haase, C., Bastviken, D., Flury, S., McGinnis, D. F., Maeck, A., Müller, D., and Premke, K.: Technical note: drifting versus anchored flux chambers for measuring greenhouse gas emissions from running waters, Biogeosciences, 12, 7013-7024, <https://doi.org/10.5194/bg-12-7013-2015>, 2015.). Hence, this issue needs to be addressed clearly. (ii) It seems that all the flux and $p\text{CO}_2$ measurements were done distributed during the day. The fact that there is a diurnal cycle of CO_2 was not considered (see e.g. Pascal Bodmer, Marlen Heinz, Martin Pusch, Gabriel Singer, Katrin Premke, Carbon dynamics and their link to dissolved organic matter quality across contrasting stream ecosystems, Science of The Total Environment, Volume 553, 2016, Pages 574-586, ISSN 0048-9697, <https://doi.org/10.1016/j.scitotenv.2016.02.095>.), and values directly compared. This issue should at least be discussed.

Response: A new section of “4.4. Uncertainty assessment of $p\text{CO}_2$ and flux-derived k values” was added. We assessed systematic errors and random errors of measurements, drifting chamber and anchored chambers, sampling time. We carefully read the two article and the two citations were included. In addition, the following text was added in the Method section.

“Similar to other studies (Alin *et al.*, 2011), sampling and flux measurements in the day time would tend to underestimate CO_2 evasion rate (Bodmer *et al.*, 2016).”

(5) Developing models to estimate transport coefficients is meaningful, but the process of the model development is poorly described. Additionally, which data were used for the models, and which not is confusing to me (goes along with my comment (2) above).

Response: The issue was addressed as follows.

Water samples from a total of 115 sites were collected. Floating chambers with replicates were deployed in 101 sites (32 sampling sites in Daning, 37 sites in TGR river networks and 32 sites in Qijiang). The sampling period covered spring and summer season, our sampling points are reasonable considering a water area of 433 km². For example, 16 sites were collected for Yangtze system to examine hydrological and geomorphological controls on $p\text{CO}_2$ (Liu *et al.*, 2017), and 17 sites for dynamic biogeochemical controls on riverine $p\text{CO}_2$ in the Yangtze basin (Liu *et al.*, 2016). Similar to other studies, sampling and flux measurements in the day would tend to underestimate CO_2 evasion rate (Bodmer *et al.*, 2016).

Prior to statistical analysis, we excluded k_{600} data for samples with the air-water $p\text{CO}_2$ gradient $<110 \mu\text{atm}$, since the error in the k_{600} calculations drastically enhances when $\Delta p\text{CO}_2$ approaches zero (Borges *et al.*, 2004; Alin *et al.*, 2011), and datasets with $\Delta p\text{CO}_2 >110 \mu\text{atm}$ provide an error of $<10\%$ on k_{600} computation (please refer to the Fig. 1 in the bottom). Thus, we discarded the samples (36.7% of sampling points with flux measurements) with $\Delta p\text{CO}_2 <110 \mu\text{atm}$ for k_{600} model development, while for the flux estimations from diffusive model and floating chambers, all samples were included.

Multivariate statistics, such as correlation and stepwise multiple linear regression, were performed for the models of k_{600} using potential physical parameters of wind speed, water depth, and current velocity as the independent variables from both separated data and combined data (Alin *et al.*, 2011). k models were obtained by water depth using data from the TGR rivers, while by flow velocity in the Qijiang.

We also highlighted separated or combined data used in the Tables and Figures.

(6) From what I see in these data, there are several running waters undersaturated with respect to CO_2 (Fig. S2 and Fig. 1), and hence a sink of CO_2 . This aspect is totally neglected and the investigated running waters are generalized as CO_2 sources to the atmosphere. In my opinion, this aspect of influxes of CO_2 is very valuable and should be properly discussed.

Response: We addressed this issue by revision in the parts of “Result” and “Discussion”.

Firstly, we revised in the “Result” section as follows. “ $p\text{CO}_2$ varied between 50 and 4830 μatm with mean of $846 \pm 819 \mu\text{atm}$ (Table 1). There were 28.7% of samples that had $p\text{CO}_2$ levels lower than 410 μatm , while the studied rivers were overall supersaturated with reference to atmospheric CO_2 and act as a source for the atmospheric CO_2 .”

Secondly, the under-saturated $p\text{CO}_2$ levels were further examined in the “Discussion” part.

“The calculated $p\text{CO}_2$ levels were within the published range, but towards the lower-end of published concentrations compiled elsewhere (Cole and Caraco, 2001; Li *et al.*, 2013). The total mean $p\text{CO}_2$ ($846 \pm 819 \mu\text{atm}$) in the TGR, Danning and Qijiang sampled was lower than one third of global river’s average ($3220 \mu\text{atm}$) (Cole and Caraco, 2001). The lower $p\text{CO}_2$ than most of the world’s river systems, particularly the under-saturated values, demonstrated that heterotrophic respiration of terrestrially derived DOC was not significant. Compared with high alkalinity, the limited delivery DOC particularly in the Daning and Qijiang river systems (Figs. S2 and S3) also indicated that in-stream respiration was limited. These two river systems are characterized by karst terrain and underlain by

carbonate rock, where photosynthetic uptake of dissolved CO₂ and carbonate minerals dissolution considerably regulated aquatic pCO₂ (Zhang *et al.*, 2017).”

(7) *As far as I see, there is some arbitrariness regarding data handling/processing. The cut-off at 110 μatm (line 198) for the air-water CO₂ gradient for k₆₀₀ calculations, as well as “When several extremely values are removed: : :” (line 303), needs to be described/demonstrated/justified much more clear.*

Response: We provided more details on this concern.

Prior to statistical analysis, we excluded k₆₀₀ data for samples with the air-water pCO₂ gradient <110 μatm, since the error in the k₆₀₀ calculations drastically enhances when ΔpCO₂ approaches zero (Borges *et al.*, 2004; Alin *et al.*, 2011), and datasets with ΔpCO₂ <110 μatm provide an error of <10% on k₆₀₀ computation” (see the Fig. 1 in the bottom)

Regarding ““When several extremely values are removed: : :” (L 303) for revised model, we supplied data including extremely data and excluding extremely data in Fig. 4 (the original Fig. 3).

(8) *CO₂ fluxes were measured, while pCO₂ and transport coefficients were calculated. This should be clearly stated throughout the manuscript to be transparent.*

Response: We highlighted this in the section of “Method” (see section 2.4), and changed “measured k or observed k to flux-derived k or derived k or calculated k”.

(9) *I am not a native English speaker, but I think that the manuscript should be revised for the English language. (see exemplarily in the specific comments and technical corrections below)
I think this is a valuable study, but the combination of the points mentioned above make the manuscript hard to follow and the conclusions and main messages drawn in the current state of the manuscript too general. In a revised version, the study would get more shaped, more detailed and informative, and the conclusions and main messages can be specified and more related to the investigated region.*

Response: We edited the English and revised the Ms based on comments and suggestions.

Specific comments

Abstract:

Line 20: Indicate how many river networks (see general comment (2))

Response: 60 rivers

Line 24: As far as I understood not when all data were included. Please be more specific here.

Response: Corrected and provided details in “Method” section.

Lines 30 – 33: This is not really new. Maybe you can specify this statement for the investigated region?

Response: Addressed. The sentence was revised as follows.

“We concluded that simple parameterization of k as a function of morphological characteristics was site specific and hence highly variable in river systems of the upper Yangtze. k models should be developed for stream studies to evaluate the contribution of these regions to atmospheric CO_2 .”

Introduction:

Line 41: Bastviken et al., 2011 totally focuses on CH_4 . I suggest replacing this reference with a more suitable one.

Response: Bastviken et al., 2011 was replaced by “Raymond et al., 2013; Butman and Raymond, 2011”

Line 42: But you did not present "new accurate measurement techniques" in your study, what are your reasons to mention this in the introduction?

Response: “new” was removed.

Lines 50 – 58: This equation is pretty standard knowledge and can just be described in words here. The equation can be moved to the methods.

Response: Yes, it is very standard knowledge, while the text would be more shape, and easily to follow with equations here.

Line 63: The standardized transport coefficient (k_{600}) should be explained here.

Response: k_{600} (the standardized transfer coefficient at a temperature of 20°C)

Line 80: You set the scene of seasonal precipitation, but in the study, you only measure in the wet season. This is contradictory. This issue should be discussed.

Response: We changed “concentrated seasonal precipitation” to “hydrological seasonality”

Lines 84-89: Kind of repetition and partially contradictory to the text in lines 43-49.

Response: The topic is different. The former focuses on flux determination, while the latter talked about k measurement method. We clarified the main text.

Lines 92-99: The relevance of the study (first time in this region, etc.), the objectives and how these objectives were approached should be written more clearly. At this point, the input parameters for the model development is totally unclear.

Response: We re-wrote this part as follows, we also provide details for k models.

To contribute to this debate, extensive investigation was firstly accomplished for determination of k in rivers and streams of the upper Yangtze using FC method. Models of k were further developed using hydraulic properties from flux measurements and TBL model.

Our recent study preliminarily investigated $p\text{CO}_2$ and air – water CO_2 areal flux as well as their controls from fluvial networks in the Three Gorges Reservoir (TGR) area (Li *et al.*, 2018). The past study was based on two field works, and the diffusive models from other continents were used. In this study, we attempted to derive k levels and develop the gas transfer model in this area (mountainous streams and small rivers) for more accurate quantification of CO_2 areal flux, and also to serve for the fluvial networks in the Yangtze River or others with similar hydrology and geomorphology. Moreover, we did detailed field campaigns in the two contrasting rivers Daning and Qijiang for models (Fig. 1). The study thus clearly stated distinct differences than the previous study (Li *et al.*, 2018) by the new contributions of specific objectives and data supplements, as well as wider significance.

Our new contributions to the literature include (1) determination and controls of k levels for small rivers and streams in subtropical areas of China, and (2) new k models developed using hydraulic parameters in the subtropical mountainous river networks.

Materials and methods:

Line 105: In my opinion, Figure S1 should go to the main text. There are no sampling points for Daning and Qijiang, which is confusing to me.

Response: We supplied the map in the main text as Fig. 1 (see Fig. 2 in the bottom).

Lines 105-109: Please add a reference for this statement.

Response: “Li et al., 2018” was cited here.

Lines 110-118: Please see my general comment (2): Please restructure this, make clear how many running waters were sampled where, the size of the sampled running waters (Strahler stream order is fine), and why in these three regions. Otherwise, it is hard to follow your storyline.

Response: We provided details in the section 2.2.

Water samples from a total of 115 sites were collected. Floating chambers with replicates were deployed in 101 sites (32 sampling sites in Daning, 37 sites in TGR river networks and 32 sites in Qijiang). The sampling period covered spring and summer season, our sampling points are reasonable considering a water area of 433 km². For example, 16 sites were collected for Yangtze system to examine hydrological and geomorphological controls on $p\text{CO}_2$ (Liu *et al.*, 2017), and 17 sites for dynamic biogeochemical controls on riverine $p\text{CO}_2$ in the Yangtze basin (Liu *et al.*, 2016). Similar to other studies, sampling and flux measurements in the day would tend to underestimate CO_2 evasion rate (Bodmer *et al.*, 2016). In our sampling points, all measured fluxes were retained since the floating chambers yielded linearly increasing CO_2 against time following manufacturer’ specification.

Prior to statistical analysis, we excluded k_{600} data for samples with the air-water $p\text{CO}_2$ gradient <110 μatm , since the error in the k_{600} calculations drastically enhances when $\Delta p\text{CO}_2$ approaches zero (Borges *et al.*, 2004; Alin *et al.*, 2011), and datasets with $\Delta p\text{CO}_2 > 110 \mu\text{atm}$ provide an error of <10% on k_{600} computation. Thus, we discarded the samples (36.7% of sampling points with flux measurements) with $\Delta p\text{CO}_2 < 110 \mu\text{atm}$ for k_{600} model development, while for the flux estimations from diffusive model and floating chambers, all samples were included.

Line 141-142: What is "PP"?

Response: EGM-4 (Environmental Gas Monitor; PP SYSTEMS Corporation, USA)

Line 148: I don't really understand what you mean by this sentence, please revise.

Response: Changed to "All the solvents and reagents used in experiments were of analytical - reagent grade"

Lines 155-156: This sentence is confusing to me, please revise.

Response: Changed "The relationship was yielded when $z=1$ ($U_{10}=1.208 \times U_1$). " to " $U_{10}=1.208 \times U_1$ as we measured the wind speed at a height of 1 m (U_1). "

Line 158: Do you mean CO2SYS? If yes, please add the corresponding reference.

Response: "Lewis et al., 1998" was cited.

Lewis, E., Wallace, D., Allison, L.J., 1998. Program developed for CO₂ system calculations.; Brookhaven National Lab., Dept. of Applied Science, Upton, NY (United States); Oak Ridge National Lab., Carbon Dioxide Information Analysis Center, TN (United States), p. Medium: ED; Size: 40 p.

Line 167: What was the brand of the tubing?

Response: Changed to "CO₂ impermeable rubber-polymer tubing"

Line 170: What is DC?

Response: "DC" was removed.

Line 173: Please see my general comment (4) (i)

Response: Uncertainty of chambers was discussed in the additional section 4.4. In our study, both drifting chamber and anchored chambers were used, which is dependent on *in situ* current velocity. The following text was added.

"In sampling sites with low and favorable flow conditions (Fig. S1), freely drifting chambers (DC) were executed, while sampling sites in rivers and streams with higher flow velocity were conducted with anchored chambers (AC) (Ran *et al.*, 2017). AC would create overestimation of CO₂ emissions in our studied region (Lorke *et al.*, 2015)."

Line 177: The units are confusing to me. Why is there two times pressure and temperature? Please double check if the units match up in the end, to me they do not.

Response: We have carefully checked and revised.

Line 187: Please be more specific: k was calculated by reorganizing Eq (1)

Response: Revised.

Line 192: Sc to the power of 0.5? This seems weird. What do you mean here?

Response: Corrected.

Line 198: Please justify the cut-off at 110 _atm. Maybe add a figure to the supplementary material.

Response: Revised. Please refer to general comment (7).

Line 203: I read about water depth and current velocity the first time here. These measurements need to be described before.

Response: Measurements of water depth and current velocity were added in the part of “Methods”.

Line 213: The pH is quite high. This in combination with influxes of CO₂ requires at least a short discussion about chemical enhancement.

Response: We revised the text in the section of “Result” (see the first paragraph below), and added text (see the second paragraph below) in the Discussion section (4.1).

“pH varied from 7.47 to 8.76 with exceptions of two quite high values of 9.38 and 8.87 (mean: 8.39 ± 0.29 from total dataset). Much lower pH was observed in TGR rivers (8.21 ± 0.33) (Table 1; $p < 0.05$; Fig. S2).”

“Higher pH levels were observed in Daning and Qijiang ($p < 0.05$ by ANOVA), where more carbonate rock exists that are characterized by karst terrain. Our pH range was comparable to the recent study on the karst river in China (Zhang *et al.*, 2017). Quite high values (i.e., 9.38 and 8.87) were recorded in the investigated sites, where chemical enhancement would increase the influx of atmospheric CO₂ to alkaline waters (Wanninkhof and Knox, 1996), while 1.7% of sampling sites that were strongly affected by chemical enhancement were not significant on a regional scale. This chemical enhancement of CO₂ influx was also reported to be limited in high-pH rivers (Zhang *et al.*, 2017).”

Wanninkhof, R., Knox, M., 1996. Chemical enhancement of CO₂ exchange in natural waters. *Limnology and Oceanography* 41, 689-697.

Line 214: Please see my general comment (6)

Response: Addressed. Please refer to general comment (6)

Lines 218-222: This paragraph belongs to the discussion section.

Response: We moved this to Discussion (see 4.1)

Lines 223-227: This paragraph should be revised because it is not very clearly written. Please add the p-values to the text in case of significances.

Response: We re-wrote this part, and nutrients were removed.

Lines 235-237: What is the meaning of this ratio? Please add a few words what the reader can get from this information.

Response: Removed.

Line 242: These models and how you developed them should be explained better (in the method section).

Response: We provided details in the section of “Method”.

Lines 246-248: I do not understand this sentence. What do you mean by "binned"?

Response: Changed to “combined”

Discussion:

Lines 270-274: How does this paragraph support the discussion of your study?

Response: These texts could support the discussion on relationship between spatial change in k values and physical characteristics (i.e., current velocity, slope and the water depth) of three river systems.

Spatial differences of k values and their controls of physical characteristics of current velocity, slope and the water depth were discussed for the three rivers systems (please refer to section 4.2).

“This could substantiate the higher k_{600} levels and spatial changes in k_{600} values of our three river systems. For instance, similar to other turbulent rivers in China (Ran *et al.*, 2015; Ran *et al.*, 2017), high k_{600} values in the TGR, Daning and Qjiang rivers were due to mountainous terrain catchment, high current velocity (10 – 150 cm/s) (Fig. 4b), bottom roughness, and shallow water depth (10 - 150 cm) (Fig. 4a). It has been suggested that shallow water enhances bottom shear, and the resultant turbulence increases k values (Alin *et al.*, 2011; Raymond *et al.*, 2012). These physical controls are highly variable across environmental types (Figs. 4a and 4b), hence, k values are expected to vary widely (Fig. 3). The k_{600} values in the TGR rivers showed wider range (1-177 cm/h; Fig. 3; Table S1), spanning more than 2 orders of magnitude across the region, and it is consistent with the considerable variability in the physical processes on water turbulence across environmental settings. Similar broad range of k_{600} levels was also observed in the China’s Yellow basin (ca. 0-123 cm/h) (Ran *et al.*, 2015; Ran *et al.*, 2017).

Contrary to our expectations, no significant relationship was observed between k_{600} and water depth, and current velocity using the entire data in the three (TGR, Danning and Qjiang) river systems

(Fig. S4). There were not statistically significant relationships between k_{600} and wind speed using separated data or combined data, and it is consistent with earlier studies (Alin *et al.*, 2011; Raymond *et al.*, 2012). Flow velocity showed linear relation with k_{600} , and the extremely high value of k_{600} was observed during the periods of higher flow velocity (Fig. S4a) using combined data. Similar trend was also observed between water depth and k_{600} values (Fig. S4b). The lack of strong correlation between k_{600} and physical factors are probably due to combined effect of both flow velocity and water depth, as well as large diversity of channel morphology, both across and within river networks in the entire catchment (60, 000 km²). This is further collaborated by weak correlations between k_{600} and flow velocity in the TGR rivers (Fig. 4), where one or two samples were taken for a large scale examination. k_{600} as a function of water depth was obtained in the TGR rivers, but it explained only 30% of the variance in k_{600} . However, model using data from Qijiang could explain 68% of the variance in k_{600} (Fig. 4b), and it was in line with general theory. Nonetheless, k_{600} from our flow velocity based model (Fig. 4b) was potentially largely overestimated with consideration of other measurements (Alin *et al.*, 2015; Ran *et al.*, 2015; Ran *et al.*, 2017). When several extremely values were removed, k_{600} (cm/h) was parameterized as follows ($k_{600} = 62.879FV + 6.8357$, $R^2 = 0.52$, $p = 0.019$, FV-flow velocity with a unit of m/s), and this revised model was in good agreement with the model in the river networks of the Yellow River (Ran *et al.*, 2017), but much lower than the model developed in the Yangtze system (Liu *et al.*, 2017) (Fig. 4c). This was reasonable because of k_{600} values in the Yangtze system were from large rivers with higher turbulence than Yellow and our studied rivers. Furthermore, the determined k_{600} using FCs was, on average, consistent with the revised model (Table 2). These differences in relationship between spatial changes in k_{600} values and physical characteristics further corroborated heterogeneity of channel geomorphology and hydraulic conditions across the investigated rivers.”

Lines 286-288: No significances: : : But still, you developed the models considering all data? This is not at all clear to me. Did you split/separate the data set for the models? This is not clear in Table 2. Please see my general comment (2). I think these data/models are valuable, but at the moment they seem arbitrary and should be better explained. This would help to give them more weight.

Response: We now provided details of model developing in the “Method” (see the response to general comment (2), (5) and (7). We also clearly stated the separated or combined data for models in the captions of Tables and Figs.

Lines 286-309: I see a lot of results here, which are presented in the discussion for the first time. The part presenting pure results should be moved to the results section.

Response: Revised. The following text was moved to the “RESULT”.

“Contrary to our expectations, no significant relationship was observed between k_{600} and water depth, and current velocity using the entire data in the three river systems (TGR, Danning and Qijiang) (Fig. S4). There were not statistically significant relationships between k_{600} and wind speed using separated data or combined data. Flow velocity showed linear relation with k_{600} , and the extremely high value of k_{600} was observed during the periods of higher flow velocity (Fig. S4a) using combined data. Similar trend was also observed between water depth and k_{600} values (Fig. S4b). k_{600} as a function of water depth was obtained in the TGR rivers, but it explained only 30% of the variance in k_{600} . However, model using data from Qijiang could explain 68% of the variance in k_{600} (Fig. 4b), and it was in line with general theory.”

Line 303: Please justify the removal of "extremely values". Maybe add a figure to the supplementary material. If there is no objective criteria and justification, I do not see why data should be removed.

Response: We provided details in the Method and also supplied figure with or without extremely values in the main text (Fig. 4).

Lines 327-333: Why discussing k values and not k600 values? I think this needs to be unified/consistent throughout the manuscript.

Response: We unified to k_{600} though k_{600} and k were widely discussed in previous studies.

Conclusion:

Lines 358-360: Very general, but actually, the regions had to be separated/ split, no?

Response: The words were removed.

Lines 368-369: I think you should focus the conclusion on the investigated region.

Response: "in the river systems of the upper Yangtze River" was added.

Tables:

Table 3: b) Why not presenting k600 values here which can be directly compared with other studies?

Response: We unified to "k₆₀₀"

Figures:

Fig. S1: Was always everything sampled at each point? "Samples" should be replaced by "sampling point"

Response: We moved it to the main text as Fig. 1, and "Samples" was changed to "Sampling point".

Fig. S4: There is no reference to this figure in the main text.

Response: Fig. S4 was cited in the main text.

Technical corrections

Line 22: Delete "were"

Line 44: add "by" or "via" before "floating chambers"

Line 59: Replace "precisely" with "well"

Line 127: "consisted" instead of "consists"

Line 139: "pH sonde" "was" instead of "is"

Line 225: "Daning" instead of "Danning"

Response: All the typos were corrected.

carbon dioxide transfer velocity and flux in low-gradient river systems and implications for regional carbon budgets. *Journal of Geophysical Research-Biogeosciences* 116.

Bodmer, P., Heinz, M., Pusch, M., Singer, G., Premke, K., 2016. Carbon dynamics and their link to dissolved organic matter quality across contrasting stream ecosystems. *Science of the Total Environment* 553, 574-586.

Borges, A.V., Delille, B., Schiettecatte, L.S., Gazeau, F., Abril, G., Frankignoulle, M., 2004. Gas transfer velocities of CO₂ in three European estuaries (Randers Fjord, Scheldt, and Thames). *Limnology and Oceanography* 49, 1630-1641.

Cole, J.J., Caraco, N.F., 2001. Carbon in catchments: connecting terrestrial carbon losses with aquatic metabolism. *Marine and Freshwater Research* 52, 101-110.

Li, S., Ni, M., Mao, R., Bush, R.T., 2018. Riverine CO₂ supersaturation and outgassing in a subtropical monsoonal mountainous area (Three Gorges Reservoir Region) of China. *Journal of Hydrology* 558, 460-469.

Li, S.Y., Lu, X.X., Bush, R.T., 2013. CO₂ partial pressure and CO₂ emission in the Lower Mekong River. *Journal of Hydrology* 504, 40-56.

Li, S.Y., Lu, X.X., He, M., Zhou, Y., Li, L., Ziegler, A.D., 2012. Daily CO₂ partial pressure and CO₂ outgassing in the upper Yangtze River basin: A case study of the Longchuan River, China. *Journal of Hydrology* 466, 141-150.

Liu, S., Lu, X.X., Xia, X., Yang, X., Ran, L., 2017. Hydrological and geomorphological control on CO₂ outgassing from low-gradient large rivers: An example of the Yangtze River system. *Journal of Hydrology* 550, 26-41.

Liu, S., Lu, X.X., Xia, X., Zhang, S., Ran, L., Yang, X., Liu, T., 2016. Dynamic biogeochemical controls on river pCO₂ and recent changes under aggravating river impoundment: An example of the subtropical Yangtze River. *Global Biogeochemical Cycles* 30, 880-897.

Ran, L., Li, L., Tian, M., Yang, X., Yu, R., Zhao, J., Wang, L., Lu, X.X., 2017. Riverine CO₂ emissions in the Wuding River catchment on the Loess Plateau: Environmental controls and dam impoundment impact. *Journal of Geophysical Research-Biogeosciences* 122, 1439-1455.

Ran, L.S., Lu, X.X., Yang, H., Li, L.Y., Yu, R.H., Sun, H.G., Han, J.T., 2015. CO₂ outgassing from the Yellow River network and its implications for riverine carbon cycle. *Journal of Geophysical Research-Biogeosciences* 120, 1334-1347.

Wanninkhof, R., Knox, M., 1996. Chemical enhancement of CO₂ exchange in natural waters. *Limnology and Oceanography* 41, 689-697.

Zhang, T., Li, J., Pu, J., Martin, J.B., Khadka, M.B., Wu, F., Li, L., Jiang, F., Huang, S., Yuan, D., 2017. River sequesters atmospheric carbon and limits the CO₂ degassing in karst area, southwest China. *Science of The Total Environment* 609, 92-101.

1 **Gas transfer velocities of CO₂ in subtropical monsoonal climate streams and**
2 **small rivers**

3

4 **Siyue Li^{a*}, Rong Mao^a, Yongmei Ma^a, Vedula V. S. S. Sarma^b**

5 a. Research Center for Eco-hydrology, Chongqing Institute of Green and Intelligent
6 Technology, Chinese Academy of Sciences, Chongqing 400714, China

7 b. CSIR-National Institute of Oceanography, Regional Centre, Visakhapatnam, India

8

9 **Correspondence**

10 **Siyue Li**

11 *Chongqing Institute of Green and Intelligent Technology (CIGIT),*

12 *Chinese Academy of Sciences (CAS).*

13 *266, Fangzheng Avenue, Shuitu High-tech Park, Beibei, Chongqing 400714, China.*

14 *Tel: +86 23 65935058; Fax: +86 23 65935000*

15 *Email: syli2006@163.com*

16 **Abstract**

17 CO₂ outgassing from rivers is a critical component for evaluating riverine carbon
18 cycle, but it is poorly quantified largely due to limited measurements and modeling of
19 gas transfer velocity (k) in subtropical streams and rivers. We measured CO₂ flux
20 rates, and calculated k and partial pressure (pCO₂) in 60 river networks of the Three
21 Gorges Reservoir (TGR) region, a typical area in the upper Yangtze River with
22 monsoonal climate and mountainous terrain. The ~~observed-determined~~ k₆₀₀ values
23 (~~k₆₀₀~~=48.4±53.2 cm/h) ~~were~~ showed large variability due to spatial variations in
24 physical controls on surface water turbulence. Our flux-derived k values
25 ~~measurements~~ using chambers were comparable with model derived from flow
26 velocities. Unlike in open waters, k is more pertinent to flow velocity and water depth
27 in the studied ~~small~~ river systems. Our results show that TGR river networks emitted
28 approx. 0.74-4 Tg CO₂/y during monsoonal period using varying approaches such as
29 chambers, ~~derived~~ measured k values and developed k model. This study suggests that
30 incorporating scale-appropriate k measurements into extensive pCO₂ investigation is
31 required to refine basin-wide carbon budgets in the subtropical streams and small
32 rivers. We concluded that simple parameterization of k as a function of morphological
33 characteristics was site specific and hence highly variable in rivers of the upper
34 Yangtze. k models should be developed for stream studies to evaluate the contribution
35 of these regions to atmospheric CO₂.

36

37 **Key words:** CO₂ outgassing, riverine C flux, flow velocity, physical controls, Three

38 Gorge Reservoir, Yangtze River

带格式的：下标

39 1. Introduction

40 Rivers serve as a significant contributor of CO₂ to the atmosphere (Cole *et al.*,
41 2007; Tranvik *et al.*, 2009; Li *et al.*, 2012; Raymond *et al.*, 2013). As a consequence,
42 accurate quantification of riverine CO₂ emissions is a key component to estimate net
43 continental carbon (C) flux (Raymond *et al.*, 2013). More detailed observational data
44 and ~~new~~ accurate measurement techniques are critical to refine the riverine C budgets
45 (Raymond and Cole, 2001; Li and Bush, 2015). Generally two methods are used to
46 estimate CO₂ areal fluxes from the river system, such as direct measurements floating
47 chambers (FCs), and indirect calculation of thin boundary layer (TBL) model that
48 depends on gas concentration at air-water gradient and gas transfer velocity, *k* (Guerin
49 *et al.*, 2007; Xiao *et al.*, 2014). Direct measurements are normally laborious, while the
50 latter method shows ease and simplicity and thus is preferred (Butman and Raymond,
51 2011; Li *et al.*, 2012; Li *et al.*, 2013; Lauerwald *et al.*, 2015; Ran *et al.*, 2015).

52 The areal flux of CO₂ (*F*, unit in mmol/m²/d) *via* the water–air interface by TBL
53 is described as follows:

$$54 F = k \times K_h \times \Delta pCO_2 \quad (1)$$

$$55 K_h = 10^{-(1.11 + 0.016 * T - 0.00007 * T^2)} \quad (2)$$

56 where *k* (unit in m/d) is the gas transfer velocity of CO₂ (also referred to as piston
57 velocity) at the *in situ* temperature (Li *et al.*, 2016). ΔpCO_2 (unit in μatm) is the
58 air-water gradient of *p*CO₂ (Borges *et al.*, 2004). K_h (mmol/m³/ μatm) is the
59 aqueous-phase solubility coefficient of CO₂ corrected using temperature (*T* in °C) (Li
60 *et al.*, 2016).

61 $\Delta p\text{CO}_2$ can be measured precisely-well in various aquatic systems, however, the
62 accuracy of the estimation of flux is depended on the k value. Broad ranges of k for
63 CO_2 (Raymond and Cole, 2001; Borges *et al.*, 2004; Raymond *et al.*, 2012) were
64 reported due to variations in techniques, tracers used and governing processes. k is
65 controlled by turbulence at the surface aqueous boundary layer, hence, k_{600} (the
66 standardized gas transfer velocity at a temperature of 20 °C) is parameterized as a
67 function of wind speed in open water systems of reservoirs, lakes, and oceans (Borges
68 *et al.*, 2004; Guerin *et al.*, 2007; Wanninkhof *et al.*, 2009). While in streams and small
69 rivers, turbulence at the water-air interface is generated by shear stresses at streambed,
70 thus k is modeled using channel slope, water depth, and water velocity in particular
71 (Alin *et al.*, 2011; Raymond *et al.*, 2012). Variable formulations of k have been
72 established by numerous theoretical, laboratory and field studies, nonetheless, better
73 constraint on k levels is still required as its levels are very significant and specific due
74 to large heterogeneity in hydrodynamics and physical characteristics of river networks.
75 This highlights the importance of k measurements in a wide range of environments
76 for the accurate upscaling of CO_2 evasion, and for parameterizing the physical
77 controls on k_{600} . However, only few studies provide information of k for riverine CO_2
78 flux in Asia (Alin *et al.*, 2011; Ran *et al.*, 2015), and those studies do not address the
79 variability of k in China's small rivers and streams.

80 Limited studies demonstrated that higher levels of k in the Chinese large rivers
81 (Alin *et al.*, 2011; Ran *et al.*, 2015; Liu *et al.*, 2017; Ran *et al.*, 2017), which
82 contributed to much higher CO_2 areal flux particularly in China's monsoonal rivers

83 that are impacted by ~~concentrated seasonal precipitation~~hydrological seasonality. The
84 monsoonal flow pattern and thus flow velocity is expected to be different than other
85 rivers in the world, as a consequence, k levels should be different than others, and
86 potentially is higher in subtropical monsoonal rivers.

87 Considerable efforts, such as purposeful (Jean-Baptiste and Poisson, 2000;
88 Crusius and Wanninkhof, 2003) and natural tracers (Wanninkhof, 1992) and FCs
89 (Borges *et al.*, 2004; Guerin *et al.*, 2007; Alin *et al.*, 2011; Prytherch *et al.*, 2017),
90 have been carried out to estimate accurate k values. The direct determination of k by
91 FCs is more popular due to simplicity of the technique for short-term CO₂ flux
92 measurements (Raymond and Cole, 2001; Xiao *et al.*, 2014; Prytherch *et al.*, 2017).
93 Prior reports, however, have demonstrated that k values and the parameterization of k
94 as a function of wind and/or flow velocity (probably water depth) vary widely across
95 rivers and streams (Raymond and Cole, 2001; Raymond *et al.*, 2012). To contribute to
96 this debate, extensive investigation was firstly accomplished for determination of k in
97 rivers and streams of the upper Yangtze using FC method. Models of k were further
98 developed using hydraulic properties by flux measurements and TBL model. Our new
99 contributions to the literature include (1) ~~providing first~~ determination and controls
100 of k levels for small rivers and streams in subtropical areas of China, and (2)
101 ~~comparisons of two methods for CO₂ areal fluxes by FCs and models developed~~new
102 models developed in the subtropical mountainous river networks. The outcome of this
103 study is expected to help in accurate estimation of CO₂ evasion from subtropical
104 rivers and streams, and thus refine riverine C budget over a regional/basin scale.

105 Our recent study preliminarily investigated $p\text{CO}_2$ and air – water CO_2 areal flux
106 as well as their controls from fluvial networks in the Three Gorges Reservoir (TGR)
107 area (Li *et al.*, 2018). The past study was based on two field works, and the diffusive
108 models from other continents were used. In this study, we attempted to derive k levels
109 and develop the gas transfer model in this area (mountainous streams and small rivers)
110 for more accurate quantification of CO_2 areal flux, and also to serve for the fluvial
111 networks in the Yangtze River or others with similar hydrology and geomorphology.
112 Moreover, we did detailed field campaigns in the two contrasting rivers Daning and
113 Qijiang for models (Fig. 1). The study thus clearly stated distinct differences than the
114 previous study (Li *et al.*, 2018) by the new contributions of specific objectives and
115 data supplements, as well as wider significance.

116

117 **2. Materials and methods**

118 **2.1. Study areas**

119 All field measurements were carried out in the rivers and streams of the Three
120 Gorges Reservoir (TGR) region ($28^{\circ}44'–31^{\circ}40'N$, $106^{\circ}10'–111^{\circ}10'E$) that is locating
121 in the upper Yangtze River, China (Fig. [1S4](#)). This region is subject to humid
122 subtropical monsoon climate with an average annual temperature ranging between 15
123 and 19°C . Average annual precipitation is approx. 1250 mm with large intra- and
124 inter-annual variability. About 75% of the annual total rainfall is concentrated in April
125 through September (Li *et al.*, 2018).

126 The river sub-catchments include large scale river networks covering the

127 majority of the tributaries of the Yangtze in the TGR region, i.e., ~~32 first-order~~
128 ~~tributaries and 16 second-order~~48 tributaries were collected. These tributaries have
129 drainage areas that vary widely from 100 to 4400 km² with width ranging from 1 m to
130 less than 100 m. The annual discharges from these tributaries have a broad spectrum
131 of 1.8 – 112 m³/s. Detailed samplings were conducted in the two largest rivers of
132 Daning (35 sampling sites) and Qijiang (32 sites) in the TGR region. These two river
133 basins drain catchment areas of 4200 and 4400 km² ~~with maximal third-order~~
134 ~~tributaries~~The studied river systems had width < 100 m, we thus defined them as
135 small rivers and streams. The Daning and Qijiang river systems are underlain by
136 widely carbonate rock, and locating in a typical karst area. The location of sampling
137 sites is deciphered in Fig. ~~1S1~~. The detailed information on sampling sites and
138 primary data are presented in the Supplement Materials (Appendix Table A1). The
139 sampling sites are outside the Reservoirs and are not affected by dam operation.

140

141 **2.2. Water sampling and analyses**

142 Three fieldwork campaigns from the main river networks in the TGR region
143 were undertaken during May through August in 2016 (i.e., 18-22 May for Daning, 21
144 June-2 July for the entire tributaries of TGR, and 15-18 August for Qijiang). A total
145 of 115 discrete grab samples were collected (each sample consisteds of three
146 replicates). Running waters were taken using pre acid-washed 5-L high density
147 polyethylene (HDPE) plastic containers from depths of 10 cm below surface. The
148 samples were filtered through pre-baked Whatman GF/F (0.7- μ m pore size) filters on

149 the sampling day and immediately stored in acid-washed HDPE bottles. The bottles
150 were transported in ice box to the laboratory and stored at 4 °C for analysis.

151 Concentrations of dissolved organic carbon (DOC) ~~and nutrients, dissolved total~~
152 ~~nitrogen (DTN), and dissolved total phosphorus (DTP)~~ were determined within 7 days
153 of water collection (Mao *et al.*, 2017).

154 Water temperature (T), pH, DO saturation (DO%) and electrical conductivity
155 (EC) were measured *in situ* by the calibrated multi-parameter sondes (HQ40d HACH,
156 USA, and YSI 6600, YSI incorporated, USA). pH, the key parameter for $p\text{CO}_2$
157 calculation, was measured to a precision of ± 0.002 , and pH sonde ~~was~~ calibrated by
158 the certified reference materials (CRMs) before measurements with an accuracy is
159 better than 0.2%. Atmospheric CO_2 concentrations were determined *in situ* using ~~PP~~
160 ~~Systems~~ EGM-4 (Environmental Gas Monitor; PP SYSTEMS Corporation, USA).

161 Total alkalinity was measured using a fixed endpoint titration method with 0.0200
162 mol/L hydrochloric acid (HCl) on the sampling day. DOC concentration was
163 measured using a total organic carbon analyzer (TOC-5000, Shimadzu, Japan) with a
164 precision better than 3% (Mao *et al.*, 2017). ~~DTN and DTP concentrations were~~

165 ~~determined using a continuous-flow autoanalyzer (AA3, Seal Analytical, Germany)~~
166 ~~and/or spectrophotometer following peroxodisulfate oxidation (Ebina *et al.*, 1983)~~

167 ~~All the solvents and reagents used in experiments were of analytical reagent grade.~~
168 ~~chemical were used for all experiments.~~

170 Concomitant stream width, depth and flow velocity were determined along the

带格式的：突出显示

带格式的：突出显示

带格式的：突出显示

171 | cross section. Wind speed at 1 m over the water surface (U_1) and air temperature (T_a)
172 | were measured with a Testo 410-1 handheld anemometer (Germany). Wind speed at
173 | 10 m height (U_{10} , unit in m/s) was calculated using the following formula (Crusius
174 | and Wanninkhof, 2003):

$$175 \quad U_{10} = U_Z \left[1 + \frac{(C_{d10})^{1/2}}{K} \times \ln\left(\frac{10}{z}\right) \right] \quad (3)$$

176 | where C_{d10} is the drag coefficient at 10 m height (0.0013 m/s), and K is the von
177 | Karman constant (0.41), and z is the height (m) of wind speed measurement. ~~The~~
178 | ~~relationship was yielded when $z=1$ ($U_{10}=1.208 \times U_1$)~~ as we measured the wind speed at
179 | a height of 1m (U_1).

带格式的：下标

180 | Aqueous $p\text{CO}_2$ was computed from the measurements of pH, total alkalinity, and
181 | water temperature using CO_2 system (k_1 and k_2 are from Millero, 1979) (Lewis *et al.*,
182 | 1998), which have been identified as high quality data (Borges *et al.*, 2004; Li *et al.*,
183 | 2012; Li *et al.*, 2013).

184

185 | 2.3. Water-to-air CO_2 fluxes using FC method

186 | FCs (30 cm in diameter, 30 cm in height) were deployed to measure air-water
187 | CO_2 fluxes and transfer velocities. They were made of cylindrical polyvinyl chloride
188 | (PVC) pipe with a volume of 21.20 L and a surface area of 0.071 m^2 . These
189 | non-transparent, thermally insulated vertical tubes, covered by aluminum foil, were
190 | connected *via* CO_2 impermeable rubber-polymer tubing (with outer and inner
191 | diameters of 0.5 cm and 0.35 cm, respectively) to a portable non-dispersive infrared

192 CO₂ analyzer EGM-4 (PPSystems). Air was circulated through the EGM-4 instrument
193 via an air filter using an integral ~~DC~~ pump at a flow rate of 350 ml/min. The chamber
194 method was widely used and more details of advantages and limits on chambers were
195 reviewed elsewhere (Borges *et al.*, 2004; Alin *et al.*, 2011; Xiao *et al.*, 2014).

196 Chamber measurements were conducted by deploying two replicate chambers or
197 one chamber for two times at each site. In sampling sites with low and favorable flow
198 conditions (Fig. S1), freely drifting chambers (DCs) were executed, while sampling
199 sites in rivers and streams with higher flow velocity were conducted with anchored
200 chambers (ACs) (Ran *et al.*, 2017). ACs would create overestimation of CO₂
201 emissions in our studied region (Lorke *et al.*, 2015). Data were logged automatically
202 and continuously at 1-min interval over a given span of time (normally 5-10 minutes)
203 after enclosure. The CO₂ area flux (mg/m²/h) was calculated using the following
204 formula.

$$205 \quad F = 60 \times \frac{dp_{CO_2} \times M \times P \times T_0}{dt \times V_0 \times P_0 \times T} H \quad (4)$$

206 Where dp_{CO_2}/dt is the rate of concentration change in FCs ($\mu\text{l}/\text{min}$); M is the
207 molar mass of CO₂ (g/mol); P is the atmosphere pressure of the sampling site (Pa); T
208 is the chamber absolute temperature of the sampling time (K); V₀, P₀, T₀ is the molar
209 volume (22.4 l/mol), P₀ is atmosphere pressure (101325 Pa), and T₀ is absolute
210 temperature (273.15 K) –under the standard condition; H is the chamber height above
211 the water surface (m) (Alin *et al.*, 2011). We accepted the flux data that had a good
212 linear regression of flux against time ($R^2 \geq 0.95$, $p < 0.01$) following manufacturer?
213 specification. In our sampling points, all measured fluxes were retained since the

带格式的: Default, 缩进: 首行缩进: 2 字符, 定义网格后自动调整右缩进, 调整中文与西文文字的间距, 调整中文与数字的间距

214 floating chambers yielded linearly increasing CO₂ against time.

215 Water samples from a total of 115 sites were collected. Floating chambers with
216 replicates were deployed in 101 sites (32 sampling sites in Daning, 37 sites in TGR
217 river networks and 32 sites in Qijiang). The sampling period covered spring and
218 summer season, our sampling points are reasonable considering a water area of 433
219 km². For example, 16 sites were collected for Yangtze system to examine
220 hydrological and geomorphological controls on pCO₂ (Liu *et al.*, 2017), and 17 sites
221 for dynamic biogeochemical controls on riverine pCO₂ in the Yangtze basin (Liu *et*
222 *al.*, 2016). Similar to other studies, sampling and flux measurements in the day would
223 tend to underestimate CO₂ evasion rate (Bodmer *et al.*, 2016).

带格式的：下标

225 **2.4. Calculations of the gas transfer velocity**

226 The k was ~~computed with~~ calculated by reorganizing Eeq (1). To make comparisons, k
227 is normalized to a Schmidt (Sc) number of 600 (k₆₀₀) at a temperature of 20 °C.

$$228 \quad k_{600} = k_T \left(\frac{600}{S_{CT}} \right)^{-0.5} \quad (5)$$

$$229 \quad S_{CT} = 1911.1 - 118.11T + 3.4527T^2 - 0.04132T^3 \quad (5)$$

230 Where k_T is the measured values at the *in situ* temperature (T, unit in °C), S_{CT} is the
231 Schmidt number of temperature T. Dependency of ~~k proportional to Sc^{-0.5}~~ -0.5 was
232 employed here as measurement were made in ~~highly~~ turbulent rivers and streams in
233 this study (Wanninkhof, 1992; Borges *et al.*, 2004; Alin *et al.*, 2011).

235 **2.5. Estimation of river water area**

236 Water surface is an important parameter for CO₂ efflux estimation, while it
237 depends on its climate, channel geometry and topography. River water area therefore
238 largely fluctuates with much higher areal extent of water surface particularly in
239 monsoonal season. However, most studies do not consider this change, and a fraction
240 of the drainage area is used in river water area calculation (Zhang *et al.*, 2017). In our
241 study, a 90 m resolution SRTM DEM (Shuttle Radar Topography Mission digital
242 elevation model) data and Landsat images in dry season were used to delineate river
243 network, and thus water area (Zhang *et al.*, 2018), whilst, stream orders were not
244 extracted. Water area of river systems is generally much higher in monsoonal season
245 in comparison to dry season, for instance, Yellow River showed 1.4-fold higher water
246 area in the wet season than in the dry season (Ran *et al.*, 2015). Available dry-season
247 image was likely to underestimate CO₂ estimation.

带格式的：下标

249 **2.65. Data processing**

250 Prior to statistical analysis, we excluded k_{600} data for samples with the air-water
251 $p\text{CO}_2$ gradient $<110 \mu\text{atm}$, since the error in the k_{600} calculations drastically enhances
252 when $\Delta p\text{CO}_2$ approaches zero (Borges *et al.*, 2004; Alin *et al.*, 2011), and datasets
253 with $\Delta p\text{CO}_2 >110 \mu\text{atm}$ provide an error of $<10\%$ on k_{600} computation. Thus, we
254 discarded the samples (36.7% of sampling points with flux measurements) with
255 $\Delta p\text{CO}_2 <110 \mu\text{atm}$ for k_{600} model development, while for the flux estimations from
256 diffusive model and floating chambers, all samples were included.

257 Spatial differences (Daning, Qijiang and entire tributaries of TGR region) were

258 tested using the nonparametric Mann Whitney U-test. Multivariate statistics, such as
259 correlation and stepwise multiple linear regression, were performed for the models of
260 k_{600} using potential physical parameters, ~~such as of~~ wind speed, water depth, and
261 current velocity ~~as the independent variables~~ (Alin *et al.*, 2011). Data analyses were
262 conducted from both separated data and combined data of river systems. k models
263 were obtained by water depth using data from the TGR rivers, while by flow velocity
264 in the Qijiang, whilst, models were not developed for combined data. All statistical
265 relationships were significant at $p < 0.05$. The statistical processes were conducted
266 using SigmaPlot 11.0 and SPSS 16.0 for Windows (Li *et al.*, 2009; Li *et al.*, 2016).

267

268 3. Results

269 3.1. CO₂ partial pressure and key water quality variables

270 The significant spatial variations in water temperature, pH, $p\text{CO}_2$, and DOC ~~and~~
271 nutrients (DTN and DTP) were observed among Daning, TGR and Qijiang rivers
272 whereas alkalinity did not display such variability (Fig. S2). pH varied from 7.47 to
273 8.76-9.38 (mean of 8.39 ± 0.29) with exceptions of two quite high values of 9.38 and
274 8.87 (total mean: 8.39 ± 0.29), and Much lower pH was observed in TGR rivers (8.21
275 ± 0.33) (Table 1; $p < 0.05$; Fig. S2). $p\text{CO}_2$ varied between 50 and 4830 μatm with mean
276 of $846 \pm 819 \mu\text{atm}$ (Table 1). There were 28.7% of samples that had $p\text{CO}_2$ levels
277 lower than 410 μatm , suggesting that all three while the studied rivers were are
278 overall supersaturated with reference to atmospheric CO₂ and act as a source for the
279 atmospheric CO₂. The $p\text{CO}_2$ levels were 2.1 to 2.6-fold higher in TGR rivers than

带格式的：突出显示

280 Daning ($483 \pm 294 \mu\text{atm}$) and Qijiang Rivers ($614 \pm 316 \mu\text{atm}$) (Fig. S2). ~~The~~
281 ~~calculated $p\text{CO}_2$ levels were within the published range, but towards the lower end of~~
282 ~~published concentrations compiled elsewhere (Cole and Caraco, 2001; Li et al., 2013).~~
283 ~~The total mean $p\text{CO}_2$ ($846 \mu\text{atm}$) in the TGR, Daning and Qijiang sampled is lower~~
284 ~~than one third of global river's average ($3220 \mu\text{atm}$) (Cole and Caraco, 2001).~~

域代码已更改

285
286 The much higher concentrations of dissolved organic carbon (DOC) ~~and~~
287 ~~dissolved nutrients (DTN and DTP) (Fig. S3) were~~ observed in the TGR rivers
288 ~~($p < 0.01$ by Mann-Whitney Rank Sum Test; Fig. S3) than Qijiang and Daning. In~~
289 ~~comparison to Daning, Qijiang showed significantly Relatively higher~~
290 concentrations of DOC ~~and DTN were observed in Qijiang than Daning River but~~
291 ~~mean TDP was much higher in latter than former river ($p < 0.05$; Fig. S3; Table 4).~~

带格式的：突出显示

带格式的：字体：(默认) Times
New Roman, 小四, 非加粗

带格式的：突出显示

293 3.2. CO₂ flux using floating chambers

294 The calculated CO₂ areal fluxes were higher in TGR rivers (217.7 ± 334.7
295 $\text{mmol/m}^2/\text{d}$, $n = 35$), followed by Daning ($122.0 \pm 239.4 \text{ mmol/m}^2/\text{d}$, $n = 28$) and
296 Qijiang rivers ($50.3 \pm 177.2 \text{ mmol/m}^2/\text{d}$, $n = 32$) (Fig. 24). The higher CO₂ evasion
297 from the TGR rivers is consistent with high riverine $p\text{CO}_2$ levels. The mean CO₂
298 emission rate was $133.1 \pm 269.1 \text{ mmol/m}^2/\text{d}$ ($n = 95$) in all three rivers sampled. The
299 mean CO₂ flux differed significantly between TGR rivers and Qijiang (Fig. 24). ~~The~~
300 ~~ratio of mean to median of CO₂ areal flux ranged between 1.4 (Qijiang) and 2.6 (TGR~~
301 ~~rivers).~~

302

303 3.3. k levels

304 ~~Samples with $\Delta p\text{CO}_2$ less than 110 μatm were excluded for k_{600} calculations, thus~~
305 ~~a~~ total of 64 data were used (10 for Daning River, 33 for TGR rivers and 21 for
306 Qijiang River) to develop k model after removal of samples with $\Delta p\text{CO}_2$ less than 110
307 μatm (Table 2). No significant variability in k_{600} values were observed among the
308 three rivers sampled (Fig. 32). The mean k_{600} (unit in cm/h) ~~was~~ relatively higher in
309 Qijiang (60.2 ± 78.9), followed by Daning (50.2 ± 20.1) and TGR rivers (40.4 ± 37.6),
310 while the median k_{600} (unit in cm/h) was higher in Daning (50.5), followed by TGR
311 rivers (30.0) and Qijiang (25.8) (Fig. 32; Table S1). ~~Combined Binned~~ k_{600} data were
312 averaged to 48.4 ± 53.2 cm/h (95% CI: 35.1-61.7), and it is 1.5-fold higher than the
313 median value (32.2 cm/h) (Fig. 32).

带格式的：下标

314 Contrary to our expectations, no significant relationship was observed between
315 k_{600} and water depth, and current velocity using the entire data in the three river
316 systems (TGR, Danning and Qijiang) (Fig. S4). There were not statistically significant
317 relationships between k_{600} and wind speed using separated data or combined data.
318 Flow velocity showed linear relation with k_{600} , and the extremely high value of k_{600}
319 was observed during the periods of higher flow velocity (Fig. S4a) using combined
320 data. Similar trend was also observed between water depth and k_{600} values (Fig. S4b).
321 k_{600} as a function of water depth was obtained in the TGR rivers, but it explained only
322 30% of the variance in k_{600} . However, model using data from Qijiang could explain
323 68% of the variance in k_{600} (Fig. 4b), and it was in line with general theory.

带格式的：缩进：首行缩进：2 字符

带格式的: 缩进: 首行缩进: 0 字符

带格式的

带格式的: 下标

带格式的: 下标

带格式的: 下标

域代码已更改

带格式的: 字体: (默认) Times New Roman, 非突出显示

324

325

326 4. Discussion

327 4.1. Determined k values relative to world rivers

328 We derived first-time the k values in the subtropical streams and small rivers.

329 Our ~~determined~~measured k_{600} levels with a 95% CI of 35.1 to 61.7 (mean: 48.4) cm/h

330 is compared well with a compilation of data for streams and small rivers (e.g., 3-70

331 cm/h) (Raymond *et al.*, 2012). Our determined k_{600} values are greater than the global

332 rivers' average (8 - 33 cm/h) (Butman and Raymond, 2011; Raymond *et al.*, 2013),

333 and much higher than mean for tropical and temperate large rivers (5-31 cm/h) (Alin

334 *et al.*, 2011). These studies evidences that k_{600} values are highly variable in streams

335 and small rivers (Alin *et al.*, 2011; Ran *et al.*, 2015). Though the mean k_{600} in the

336 TGR, Daning and Qijiang is higher than global mean, however, it is consistent with

337 k_{600} values in the main stream and river networks of the turbulent Yellow River ($42 \pm$

338 17 cm/h) (Ran *et al.*, 2015), and Yangtze (38 ± 40 cm/h) (Liu *et al.*, 2017) (Table S2).

339 The calculated pCO_2 levels were within the published range, but towards the

340 lower-end of published concentrations compiled elsewhere (Cole and Caraco, 2001;

341 Li *et al.*, 2013). The total mean pCO_2 (846 ± 819 μatm) in the TGR, Danning and

342 Qijiang sampled ~~was~~is lower than one third of global river's average (3220 μatm)

343 (Cole and Caraco, 2001). The lower pCO_2 than most of the world's river systems,

344 particularly the under-saturated values, demonstrated that heterotrophic respiration of

345 terrestrially derived DOC was not significant. Compared with high alkalinity, the

346 limited delivery of DOC particularly in the Daning and Qijiang river systems (Figs.
347 S2 and S3) also indicated that in-stream respiration was limited. These two river
348 systems are characterized by karst terrain and underlain by carbonate rock, where
349 photosynthetic uptake of dissolved CO₂ and carbonate minerals dissolution
350 considerably regulated aquatic pCO₂ (Zhang *et al.*, 2017).

带格式的：下标

351 Higher pH levels were observed in Daning and Qijiang river systems ($p < 0.05$ by
352 Mann-Whitney Rank Sum Test), where more carbonate rock exists that are
353 characterized by karst terrain. Our pH range was comparable to the recent study on
354 the karst river in China (Zhang *et al.*, 2017). Quite high values (i.e., 9.38 and 8.87)
355 were recorded in some investigated sites, where chemical enhancement would
356 increase the influx of atmospheric CO₂ to alkaline waters (Wanninkhof and Knox,
357 1996), while 1.7% of sampling sites that were strongly affected by chemical
358 enhancement were not significant on a regional scale. This chemical enhancement of
359 CO₂ influx was also reported to be limited in high-pH rivers (Zhang *et al.*, 2017).

带格式的：定义网格后自动调整右缩进，调整中文与西文文字的间距，调整中文与数字的间距

带格式的：字体：(默认) Times New Roman

361 **4.2. Hydraulic controls of k_{600}**

带格式的：缩进：首行缩进：0 厘米

带格式的：下标

带格式的：下标

362 It has been well established that k_{600} is governed by a multitude of physical
363 factors particularly current velocity, wind speed, stream slope and water depth, of
364 which, wind speed is the dominant factor of k in open waters such as large rivers and
365 estuaries (Raymond and Cole, 2001; Crusius and Wanninkhof, 2003; Borges *et al.*,
366 2004; Alin *et al.*, 2011). In contrast k_{600} in small rivers and streams is closely linked to
367 flow velocity, water depth and channel slope (Alin *et al.*, 2011; Raymond *et al.*, 2012).

带格式的：下标

368 Several studies reported that the combined contribution of flow velocity and wind
369 speed to k is significant in the large rivers (Beaulieu *et al.*, 2012; Ran *et al.*, 2015).
370 Thus, k_{600} values are higher in the Yellow River (ca. 0-120 cm/h) as compared to the
371 low-gradient River Mekong (0-60 cm/h) (Alin *et al.*, 2011; Ran *et al.*, 2015), due to
372 higher ~~wind speed and~~ flow velocity in the Yellow River (1.8 m/s) than Mekong river
373 (0.9 ± 0.4 m/s), resulting in greater surface turbulence and higher k_{600} level in the
374 Yellow (42 ± 17 cm/h) than Mekong river (15 ± 9 cm/h). This could substantiate the
375 higher k_{600} levels and spatial changes in k_{600} values of our three river systems. For
376 instance, similar to other turbulent rivers in China (Ran *et al.*, 2015; Ran *et al.*, 2017).
377 ~~The higher~~ k_{600} values in the TGR, Daning and Qjiang rivers ~~were~~ due to
378 mountainous terrain catchment, high current velocity (10 – 150 cm/s) (Fig. 43b),
379 bottom roughness, and shallow water depth (10 - 150 cm) (Fig. 43a). It has been
380 suggested that shallow water enhances bottom shear, and the resultant turbulence
381 increases k values (Alin *et al.*, 2011; Raymond *et al.*, 2012). These physical controls
382 are highly variable across environmental types (Figs. 43a and 43b), hence, k values
383 are expected to vary widely (Fig. 32). The k_{600} values in the TGR rivers showed wider
384 range (1-177 cm/h; Fig. 32; Table S1), spanning more than 2 orders of magnitude
385 across the region, and it is consistent with the considerable variability in the physical
386 processes on water turbulence across environmental settings. Similar broad range of
387 k_{600} levels was also observed in the China's Yellow basin (ca. 0-123 cm/h) (Ran *et al.*,
388 2015; Ran *et al.*, 2017).

389 ~~Contrary to our expectations, no significant relationship was observed between~~

带格式的：下标

带格式的：下标

带格式的：下标

带格式的：下标

带格式的：下标

带格式的：下标

390 ~~k₆₀₀ and water depth, and current velocity using the entire data in the TGR, Danning~~
391 ~~and Qjiang rivers (Fig. S5). There were not statistically significant relationships~~
392 ~~between k and wind speed and it is~~ Absent relationships between riverine k₆₀₀ and
393 wind speed were consistent with earlier studies (Alin *et al.*, 2011; Raymond *et al.*,
394 2012). ~~Flow velocity showed linear relation with k, and the extremely high value of k~~
395 ~~was observed during the periods of higher flow velocity (Fig. S5a). Similar trend was~~
396 ~~also observed between water depth and k values (Fig. S5b).~~ The lack of strong
397 correlation between k₆₀₀ and physical factors using the combined data were
398 probably due to combined effect of both flow velocity and water depth, as well as
399 large diversity of channel morphology, both across and within river networks in the
400 entire catchment (60, 000 km²). This is further collaborated by weak correlations
401 between k₆₀₀ and flow velocity in the TGR rivers (Fig. 43), where one or two samples
402 were taken for a large scale examination. We provided new insights into k₆₀₀.
403 ~~parameterized using current velocity. k₆₀₀ as a function of water depth was obtained in~~
404 ~~TGR rivers, but it explained only 30% of the variance in k₆₀₀. However, model using~~
405 ~~data from Qijiang could explain 68% of the variance in k₆₀₀ (Fig. 3b), and it was in~~
406 ~~line with general theory.~~ Nonetheless, k₆₀₀ from our flow velocity based model (Fig.
407 43b) ~~is was potentially~~ largely overestimated with consideration of other
408 measurements (Alin *et al.*, 2015; Ran *et al.*, 2015; Ran *et al.*, 2017). When several
409 extremely values ~~are were~~ removed, k₆₀₀ (cm/h) ~~is was~~ parameterized as follows (k₆₀₀
410 = 62.879FV + 6.8357, R² = 0.52, p = 0.019, FV = flow velocity with a unit of m/s), and
411 this revised model ~~is was~~ in good agreement with the model in the river networks of

带格式的：非突出显示

域代码已更改

带格式的：下标

带格式的：下标

带格式的：非突出显示

带格式的：下标

域代码已更改

域代码已更改

域代码已更改

带格式的：下标

带格式的：下标

带格式的：下标

带格式的：下标

412 the Yellow River (Ran *et al.*, 2017), but much lower than the model developed in the
413 Yangtze system (Liu *et al.*, 2017) (Fig. 43c). This ~~is~~^{was} reasonable because of k_{600}
414 values in the Yangtze system ~~are~~^{were} from large rivers with higher turbulence than
415 Yellow and our studied rivers. Furthermore, the ~~measured-determined~~ k_{600} using FCs
416 was, on average, consistent with the revised model (Table 2). These differences in
417 relationship between spatial changes in k_{600} values and physical characteristics further
418 corroborated heterogeneity of channel geomorphology and hydraulic conditions
419 across the investigated rivers.

420 The subtropical streams and small rivers are biologically more active and are
421 recognized to exert higher CO₂ areal flux to the atmosphere, however, their
422 contribution to riverine carbon cycling is still poorly quantified because of data
423 paucity and the absence of k in particular. Larger uncertainty of riverine CO₂ emission
424 in China was anticipated by use of k_{600} from other continents or climate zones. For
425 instance, k_{600} for CO₂ emission from tributaries in the Yellow River and karst rivers
426 was originated from the model in the Mekong (Zhang *et al.*, 2017), and Pearl (Yao *et*
427 *al.*, 2007), Longchuan (Li *et al.*, 2012), and Metropolitan rivers (Wang *et al.*, 2017),
428 which are mostly from temperate regions. Our k_{600} values will therefore largely
429 improve the estimation of CO₂ evasion from subtropical streams and small rivers, and
430 improve to refine riverine carbon budget. More studies, however, are clearly needed
431 to build the model, based on flow velocity and slope/water depth given the difficulty
432 in k quantification on a large scale.

433

4.3. Implications for large scale estimation

We compared CO₂ areal flux by FCs and models developed here (Fig. 43) and other studies (Alin *et al.*, 2011) (Tables 2 and 3). CO₂ evasion was estimated for rivers in China with k values ranged between 8 and 15 cm/h (Yao *et al.*, 2007; Wang *et al.*, 2011; Li *et al.*, 2012) (Table S2). These estimates of CO₂ evasion rate were considerably lower than using present k_{600} values (48.4 ± 53.2 cm/h). For instance, CO₂ emission rates in the Longchuan River (e.g., k=8 cm/h) and Pearl River tributaries (e.g., k=8-15 cm/h) were 3 to 6 times higher using present k values compared to earlier estimates. We found that the determined k_{600} average was marginally beyond the levels from water depth based model and the model developed by Alin *et al.* (Alin *et al.*, 2011), while equivalent to the flow velocity based revised model, resulting in similar patterns of CO₂ emission rates (Table 2). Hence selection of k values would significantly hamper the accuracy of the flux estimation. Therefore k must be estimated along with $p\text{CO}_2$ measurements to accurate flux estimations.

We used our measured CO₂ emission rates by FCs for upscaling flux estimates during monsoonal period given the sampling in this period and it was found to be $0.701\text{--}0.39$ TgCO₂/y for all rivers sampled in our study (Table 3a). The estimated emission in the monsoonal period was close to that of the revised model ($0.711\text{--}0.40 \pm 0.661\text{--}0.34$ (95% confidence interval: $0.4691\text{--}0.941\text{--}0.87$) Tg CO₂/y), and using the determined k average, i.e., $0.691\text{--}0.37 \pm 0.651\text{--}0.28$ (95% confidence interval: $0.4589\text{--}0.931\text{--}0.84$) Tg CO₂/y, but slightly higher than the estimation using water-depth based model ($0.541\text{--}0.08 \pm 0.511\text{--}0.04$ Tg CO₂/y) and Alin's model ($0.531\text{--}0.06 \pm 0.501\text{--}0.00$

带格式的：下标

456 Tg CO₂/y) (Table 3b). ~~The estimate was within the range of our earlier work using~~
457 ~~TBL on the TGR river networks (0.64-2.33 Tg CO₂/y) (Li *et al.*, 2018).~~ The higher
458 emission, i.e., ~~1.663.29 ± 1.553.08 (1.082.15-2.234.43)~~ Tg CO₂/y, using flow velocity
459 based model may be over-estimated when compared to other models, flux from
460 determined k (Table 3b) and previous annual estimates, i.e., our earlier annual evasion
461 of 0.64-2.33 Tg CO₂/y using TBL on the TGR river networks (Li *et al.*, 2018).
462 ~~Moreover, our estimated CO₂ emission in the monsoonal period Therefore, this study~~
463 also suggests that CO₂ annual emissions from rivers and streams in this area ~~may~~
464 ~~be~~ were previously underestimated, i.e., 0.03 Tg CO₂/y (Li *et al.*, 2017) and 0.37-0.44
465 Tg CO₂/y (Yang *et al.*, 2013) as the former used TBL model with a lower k level, and
466 the latter employed floating chambers, but they both sampled very limited tributaries
467 (i.e., 2-3 rivers). Therefore, measurements of k must be made mandatory along with
468 pCO₂ measurement in the river and stream studies.

带格式的：下标

带格式的：字体：倾斜

带格式的：下标

469

470 4.4. Uncertainty assessment of pCO₂ and flux-derived k₆₀₀ values

带格式的：下标

471 The uncertainty of flux-derived k values mainly stem from ΔpCO₂ (unit in ppm)
472 and flux measurements (Lorke *et al.*, 2015; Bodmer *et al.*, 2016; Golub *et al.*, 2017).
473 Thus we provided uncertainty assessments caused by dominant sources of uncertainty
474 from measurements of aquatic pCO₂ and CO₂ areal flux since uncertainty of
475 atmospheric CO₂ measurement could be neglected.

476 In our study, aquatic pCO₂ was computed based on pH, alkalinity and water
477 temperature rather than directly measured. Recent studies highlighted pCO₂

478 uncertainty caused by systematic errors over empiric random errors (Golub *et al.*,
479 2017). Systematic errors are mainly attributed to instrument limitations, i.e., sondes of
480 pH and water temperature. The relative accuracy of temperature meters was ± 0.1 °C
481 according to manufacturers' specifications, thus the uncertainty of water T propagated
482 on uncertainty in $p\text{CO}_2$ was minor (Golub *et al.*, 2017). Systematic errors therefore
483 stem from pH, which has been proved to be a key parameter for biased $p\text{CO}_2$
484 estimation calculated from aquatic carbon system (Li *et al.*, 2013; Abril *et al.*, 2015).
485 We used a high accuracy of pH electrode and the pH meters were carefully calibrated
486 using CRMs, and *in situ* measurements showed an uncertainty of ± 0.01 . We then run
487 an uncertainty of ± 0.01 pH to quantify the $p\text{CO}_2$ uncertainty, and an uncertainty of $\pm 3\%$
488 was observed. Systematic errors thus seemed to show little effects on $p\text{CO}_2$ errors in
489 our study.

带格式的：字体：小四，倾斜，字体颜色：黑色

490 Random errors are from repeatability of carbonate measurements. Two replicates
491 for each sample showed the uncertainty of within $\pm 5\%$, indicating that uncertainty in
492 $p\text{CO}_2$ calculation from alkalinity measurements could be minor.

493 The measured pH ranges also exhibited great effects on $p\text{CO}_2$ uncertainty (Hunt
494 *et al.*, 2011; Abril *et al.*, 2015). At low pH, $p\text{CO}_2$ can be overestimated when
495 calculated from pH and alkalinity (Abril *et al.*, 2015). Samples for CO_2 fluxes
496 estimated from pH and alkalinity showed pH average of 8.39 ± 0.29 (median 8.46 with
497 quartiles of 8.24-8.56) ($n=115$). Thus, overestimation of calculated CO_2 areal flux
498 from pH and alkalinity is likely to be minor. Further, contribution of organic matter to

带格式的：字体：倾斜

499 non-carbonate alkalinity is likely to be neglected because of low DOC (mean 6.67
500 mg/L; median 2.51 mg/L) (Hunt *et al.*, 2011; Li *et al.*, 2013).

501 Efforts have been devoted to measurement techniques (comparison of FC, eddy
502 covariance-EC and boundary layer model-BLM) for improving CO₂ quantification
503 from rivers because of a notable contribution of inland waters to the global C budget,
504 which could have a large effect on the magnitude of the terrestrial C sink. Whilst,
505 prior studies reported inconsistent trends of CO₂ area flux by these methods. For
506 instance, CO₂ areal flux from FC was much lower than EC (Podgrajsek *et al.*, 2014),
507 while areal flux from FC was higher than both EC and BLM elsewhere (Erkkila *et al.*,
508 2018), however, Schilder *et al.* (Schilder *et al.*, 2013) demonstrated that areal flux
509 from BLM was 33-320% of in-situ FC measurements. Albeit unsatisfied errors of
510 varied techniques and additional perturbations from FC exist, FC method is currently
511 a simple and preferred measurement for CO₂ flux because that choosing a right k
512 value remains a major challenge and others require high workloads (Martinsen *et al.*,
513 2018).

514 Recent study further reported fundamental differences in CO₂ emission rates
515 between ACs and freely DFs (Lorke *et al.*, 2015), i.e., ACs biased the gas areal flux
516 higher by a factor of 2.0-5.5. However, some studies observed that ACs showed
517 reasonable agreement with other flux measurement techniques (Galfalk *et al.*, 2013),
518 and this straightforward, inexpensive and relatively simple method AC was widely
519 used (Ran *et al.*, 2017). Water-air interface CO₂ flux measurements were primarily
520 made using ACs in our studied streams and small rivers because of relatively high

带格式的：下标

带格式的：非突出显示

带格式的：非突出显示

带格式的：缩进：首行缩进：2 字符，定义网格后不调整右缩进，无孤行控制，不调整西文与中文之间的空格，不调整中文和数字之间的空格，制表位：不在 4.36 字符 + 8.72 字符 + 13.09 字符 + 17.45 字符 + 21.81 字符 + 26.17 字符 + 30.53 字符 + 34.9 字符 + 39.26 字符 + 43.62 字符 + 47.98 字符 + 52.34 字符 + 56.7 字符 + 61.07 字符 + 65.43 字符 + 69.79 字符

521 current velocity; otherwise, floating chambers will travel far during the measurement
522 period. In addition, inflatable rings were used for sealing the chamber headspace and
523 submergence of ACs was minimal, therefore, our measurements were potentially
524 overestimated but reasonable. We could not test the overestimation of ACs in this
525 study, the modified FCs, i.e., DCs and integration of ACs and DCs, and multi-method
526 comparison study including FCs, ECs and BLM should be conducted for a reliable
527 chamber method.

528 Sampling seasonality considerably regulated riverine $p\text{CO}_2$ and gas transfer
529 velocity and thus water-air interface CO_2 evasion rate (Li *et al.*, 2012; Ran *et al.*,
530 2015). We sampled waters in wet season (monsoonal period) due to that it showed
531 wider range of flow velocity and thus it covered the k_{600} levels in the whole
532 hydrological season. Wet season generally had higher current velocity and thus higher
533 gas transfer velocity (Ran *et al.*, 2015), while aquatic $p\text{CO}_2$ was variable with
534 seasonality. We recently reported that riverine $p\text{CO}_2$ in the wet season was 81% the
535 level in the dry season (Li *et al.*, 2018), and prior study on the Yellow River reported
536 that k level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*,
537 2015), while another study on the Wuding River demonstrated that k level in the wet
538 season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Thus, we
539 acknowledged a certain amount of errors on the annual flux estimation from sampling
540 campaigns during the wet season in the TGR area, while this uncertainty could not be
541 significant because that the diluted $p\text{CO}_2$ could alleviate the overestimated emission
542 by increased k level in the wet season (stronger discussion please refer to SOM).

带格式的：字体：(中文)+中文正文(宋体)，10磅，字体颜色：自动设置

带格式的：首行缩进：2字符，定义网格后不调整右缩进，不调整西文与中文之间的空格，不调整中文和数字之间的空格

带格式的：下标

543

544 **5. Conclusion**

545 We provided first determination of gas transfer velocity (k) in the subtropical
546 streams and small rivers in the upper Yangtze. High variability in k values (mean 48.4
547 ± 53.2 cm/h) was observed, reflecting the variability of morphological characteristics
548 on water turbulence both within and across river networks. ~~The determined k using~~
549 ~~floating chambers (FCs) was comparable to our newly water depth based model,~~
550 ~~while substantially lower than flow velocity based model.~~ We highlighted that k
551 estimate from empirical model should be pursued with caution and the significance of
552 incorporating k measurements along with extensive $p\text{CO}_2$ investigation is highly
553 essential for upscaling to watershed/regional scale carbon (C) budget.

554 Riverine $p\text{CO}_2$ and CO_2 areal flux showed pronounced spatial variability with
555 much higher levels in the TGR rivers. The CO_2 areal flux was averaged at $133.1 \pm$
556 269.1 mmol/m²/d using FCs, the resulting emission in the monsoonal period was
557 around ~~0.74-39~~ Tg CO_2/y , similar to the scaling up emission with the determined k ,
558 and the revised flow velocity based model, while marginally above the water depth
559 based model. More work is clearly needed to refine the k modeling in the river
560 systems of the upper Yangtze River for evaluating regional C budgets.

561

562 **Acknowledgements**

563 This study was funded by “the Hundred-Talent Program” of the Chinese Academy of
564 Sciences (R53A362Z10; granted to Dr. Li), and the National Natural Science

565 Foundation of China (Grant No. 31670473). We are grateful to Mrs. Maofei Ni and
566 Tianyang Li, and Miss Jing Zhang for their assistance in the field works. Users can
567 access the original data from an Appendix. Special thanks are given to editor and
568 anonymous reviewers for improving the Ms.

569 **References**

- 570 Abril, G., Bouillon, S., Darchambeau, F., Teodoru, C.R., Marwick, T.R., Tamooh, F., Omengo, F.O.,
571 Geeraert, N., Deirmendjian, L., Polsenaere, P., Borges, A.V., 2015. Technical Note: Large overestimation
572 of pCO₂ calculated from pH and alkalinity in acidic, organic-rich freshwaters. *Biogeosciences* 12,
573 67-78.
- 574 Alin, S.R., Maria, D.F.F.L.R., Salimon, C.I., Richey, J.E., Holtgrieve, G.W., Krusche, A.V., Snidvongs, A.,
575 2015. Physical controls on carbon dioxide transfer velocity and flux in low - gradient river systems and
576 implications for regional carbon budgets. *Journal of Geophysical Research Biogeosciences* 116,
577 248-255.
- 578 Alin, S.R., Raseira, M., Salimon, C.I., Richey, J.E., Holtgrieve, G.W., Krusche, A.V., Snidvongs, A., 2011.
579 Physical controls on carbon dioxide transfer velocity and flux in low-gradient river systems and
580 implications for regional carbon budgets. *Journal of Geophysical Research-Biogeosciences* 116.
- 581 Beaulieu, J.J., Shuster, W.D., Rebholz, J.A., 2012. Controls on gas transfer velocities in a large river.
582 *Journal of Geophysical Research-Biogeosciences* 117.
- 583 Bodmer, P., Heinz, M., Pusch, M., Singer, G., Premke, K., 2016. Carbon dynamics and their link to
584 dissolved organic matter quality across contrasting stream ecosystems. *Science of the Total*
585 *Environment* 553, 574-586.
- 586 Borges, A.V., Delille, B., Schiettecatte, L.S., Gazeau, F., Abril, G., Frankignoulle, M., 2004. Gas transfer
587 velocities of CO₂ in three European estuaries (Randers Fjord, Scheldt, and Thames). *Limnology and*
588 *Oceanography* 49, 1630-1641.
- 589 Butman, D., Raymond, P.A., 2011. Significant efflux of carbon dioxide from streams and rivers in the
590 United States. *Nature Geoscience* 4, 839-842.
- 591 Cole, J.J., Caraco, N.F., 2001. Carbon in catchments: connecting terrestrial carbon losses with aquatic
592 metabolism. *Marine and Freshwater Research* 52, 101-110.
- 593 Cole, J.J., Prairie, Y.T., Caraco, N.F., McDowell, W.H., Tranvik, L.J., Striegl, R.G., Duarte, C.M., Kortelainen,
594 P., Downing, J.A., Middelburg, J.J., Melack, J., 2007. Plumbing the Global Carbon Cycle: Integrating
595 Inland Waters into the Terrestrial Carbon Budget. *Ecosystems* 10, 172-185.
- 596 Crusius, J., Wanninkhof, R., 2003. Gas transfer velocities measured at low wind speed over a lake.
597 *Limnology and Oceanography* 48, 1010-1017.
- 598 Ebina, J., Tsutsui, T., Shirai, T., 1983. SIMULTANEOUS DETERMINATION OF TOTAL NITROGEN AND TOTAL
599 PHOSPHORUS IN WATER USING PEROXODISULFATE OXIDATION. *Water Research* 17, 1721-1726.
- 600 Erkkila, K.-M., Ojala, A., Bastviken, D., Biermann, T., Heiskanen, J.J., Lindroth, A., Peltola, O., Rantakari,
601 M., Vesala, T., Mammarella, I., 2018. Methane and carbon dioxide fluxes over a lake: comparison
602 between eddy covariance, floating chambers and boundary layer method. *Biogeosciences* 15,
603 429-445.
- 604 Galfalk, M., Bastviken, D., Fredriksson, S.T., Arneborg, L., 2013. Determination of the piston velocity for
605 water-air interfaces using flux chambers, acoustic Doppler velocimetry, and IR imaging of the water
606 surface. *Journal of Geophysical Research-Biogeosciences* 118, 770-782.
- 607 Golub, M., Desai, A.R., McKinley, G.A., Remucal, C.K., Stanley, E.H., 2017. Large Uncertainty in
608 Estimating p
609 CO₂
610 From Carbonate Equilibria in Lakes. *Journal of Geophysical Research: Biogeosciences* 122, 2909-2924.
- 611 Guerin, F., Abril, G., Serca, D., Delon, C., Richard, S., Delmas, R., Tremblay, A., Varfalvy, L., 2007. Gas

612 transfer velocities of CO₂ and CH₄ in a tropical reservoir and its river downstream. *Journal of Marine*
613 *Systems* 66, 161-172.

614 Hunt, C.W., Salisbury, J.E., Vandemark, D., 2011. Contribution of non-carbonate anions to total
615 alkalinity and overestimation of $p\text{CO}_2$ in New England and New Brunswick rivers.
616 *Biogeosciences* 8, 3069-3076.

617 Jean-Baptiste, P., Poisson, A., 2000. Gas transfer experiment on a lake (Kerguelen Islands) using He-3
618 and SF₆. *Journal of Geophysical Research-Oceans* 105, 1177-1186.

619 Lauerwald, R., Laruelle, G.G., Hartmann, J., Ciais, P., Regnier, P.A.G., 2015. Spatial patterns in CO₂
620 evasion from the global river network. *Global Biogeochemical Cycles* 29, 534-554.

621 Lewis, E., Wallace, D., Allison, L.J., 1998. Program developed for CO₂ system calculations. ;
622 Brookhaven National Lab., Dept. of Applied Science, Upton, NY (United States); Oak Ridge National
623 Lab., Carbon Dioxide Information Analysis Center, TN (United States), p. Medium: ED; Size: 40 p.

624 Li, S., Bush, R.T., 2015. Revision of methane and carbon dioxide emissions from inland waters in India.
625 *Global Change Biology* 21, 6-8.

626 Li, S., Bush, R.T., Ward, N.J., Sullivan, L.A., Dong, F., 2016. Air-water CO₂ outgassing in the Lower Lakes
627 (Alexandrina and Albert, Australia) following a millennium drought. *Science of the Total Environment*
628 542, 453-468.

629 Li, S., Gu, S., Tan, X., Zhang, Q., 2009. Water quality in the upper Han River basin, China: The impacts
630 of land use/land cover in riparian buffer zone. *Journal of Hazardous Materials* 165, 317-324.

631 Li, S., Ni, M., Mao, R., Bush, R.T., 2018. Riverine CO₂ supersaturation and outgassing in a subtropical
632 monsoonal mountainous area (Three Gorges Reservoir Region) of China. *Journal of Hydrology* 558,
633 460-469.

634 Li, S., Wang, F., Luo, W., Wang, Y., Deng, B., 2017. Carbon dioxide emissions from the Three Gorges
635 Reservoir, China. *Acta Geochimica* <https://doi.org/10.1007/s11631-017-0154-6>

636 Li, S.Y., Lu, X.X., Bush, R.T., 2013. CO₂ partial pressure and CO₂ emission in the Lower Mekong River.
637 *Journal of Hydrology* 504, 40-56.

638 Li, S.Y., Lu, X.X., He, M., Zhou, Y., Li, L., Ziegler, A.D., 2012. Daily CO₂ partial pressure and CO₂
639 outgassing in the upper Yangtze River basin: A case study of the Longchuan River, China. *Journal of*
640 *Hydrology* 466, 141-150.

641 Liu, S., Lu, X.X., Xia, X., Yang, X., Ran, L., 2017. Hydrological and geomorphological control on CO₂
642 outgassing from low-gradient large rivers: An example of the Yangtze River system. *Journal of*
643 *Hydrology* 550, 26-41.

644 Liu, S., Lu, X.X., Xia, X., Zhang, S., Ran, L., Yang, X., Liu, T., 2016. Dynamic biogeochemical controls on
645 river pCO₂ and recent changes under aggravating river impoundment: An example of the subtropical
646 Yangtze River. *Global Biogeochemical Cycles* 30, 880-897.

647 Lorke, A., Bodmer, P., Noss, C., Alshboul, Z., Koschorreck, M., Somlai-Haase, C., Bastviken, D., Flury, S.,
648 McGinnis, D.F., Maeck, A., Mueller, D., Premke, K., 2015. Technical note: drifting versus anchored flux
649 chambers for measuring greenhouse gas emissions from running waters. *Biogeosciences* 12,
650 7013-7024.

651 Mao, R., Chen, H., Li, S., 2017. Phosphorus availability as a primary control of dissolved organic carbon
652 biodegradation in the tributaries of the Yangtze River in the Three Gorges Reservoir Region. *Science of*
653 *the Total Environment* 574, 1472-1476.

654 Martinsen, K.T., Kragh, T., Sand-Jensen, K., 2018. Technical note: A simple and cost-efficient automated
655 floating chamber for continuous measurements of carbon dioxide gas flux on lakes. *Biogeosciences* 15,

656 5565-5573.

657 Podgrajsek, E., Sahlee, E., Bastviken, D., Holst, J., Lindroth, A., Tranvik, L., Rutgersson, A., 2014.

658 Comparison of floating chamber and eddy covariance measurements of lake greenhouse gas fluxes.

659 *Biogeosciences* 11, 4225-4233.

660 Prytherch, J., Brooks, I.M., Crill, P.M., Thornton, B.F., Salisbury, D.J., Tjernstrom, M., Anderson, L.G.,

661 Geibel, M.C., Humborg, C., 2017. Direct determination of the air-sea CO₂ gas transfer velocity in Arctic

662 sea ice regions. *Geophysical Research Letters* 44, 3770-3778.

663 Ran, L., Li, L., Tian, M., Yang, X., Yu, R., Zhao, J., Wang, L., Lu, X.X., 2017. Riverine CO₂ emissions in the

664 Wuding River catchment on the Loess Plateau: Environmental controls and dam impoundment impact.

665 *Journal of Geophysical Research-Biogeosciences* 122, 1439-1455.

666 Ran, L.S., Lu, X.X., Yang, H., Li, L.Y., Yu, R.H., Sun, H.G., Han, J.T., 2015. CO₂ outgassing from the Yellow

667 River network and its implications for riverine carbon cycle. *Journal of Geophysical*

668 *Research-Biogeosciences* 120, 1334-1347.

669 Raymond, P.A., Cole, J.J., 2001. Gas exchange in rivers and estuaries: Choosing a gas transfer velocity.

670 *Estuaries* 24, 312-317.

671 Raymond, P.A., Hartmann, J., Lauerwald, R., Sobek, S., McDonald, C., Hoover, M., Butman, D., Striegl,

672 R., Mayorga, E., Humborg, C., Kortelainen, P., Duerr, H., Meybeck, M., Ciais, P., Guth, P., 2013. Global

673 carbon dioxide emissions from inland waters. *Nature* 503, 355-359.

674 Raymond, P.A., Zappa, C.J., Butman, D., Bott, T.L., Potter, J., Mulholland, P., Laursen, A.E., McDowell,

675 W.H., Newbold, D., 2012. Scaling the gas transfer velocity and hydraulic geometry in streams and small

676 rivers. *Limnology & Oceanography Fluids & Environments* 2, 41-53.

677 Schilder, J., Bastviken, D., van Hardenbroek, M., Kankaala, P., Rinta, P., Stötter, T., Heiri, O., 2013.

678 Spatial heterogeneity and lake morphology affect diffusive greenhouse gas emission estimates of lakes.

679 *Geophysical Research Letters* 40, 5752-5756.

680 Tranvik, L.J., Downing, J.A., Cotner, J.B., Loiselle, S.A., Striegl, R.G., Ballatore, T.J., Dillon, P., Finlay, K.,

681 Fortino, K., Knoll, L.B., 2009. Lakes and reservoirs as regulators of carbon cycling and climate.

682 *Limnology & Oceanography* 54, 2298-2314.

683 Wang, F., Wang, B., Liu, C.Q., Wang, Y., Guan, J., Liu, X., Yu, Y., 2011. Carbon dioxide emission from

684 surface water in cascade reservoirs-river system on the Maotiao River, southwest of China.

685 *Atmospheric Environment* 45, 3827-3834.

686 Wang, X.F., He, Y.X., Yuan, X.Z., Chen, H., Peng, C.H., Zhu, Q., Yue, J.S., Ren, H.Q., Deng, W., Liu, H.,

687 2017. pCO₂ and CO₂ fluxes of the metropolitan river network in relation to the urbanization of

688 Chongqing, China. *Journal of Geophysical Research-Biogeosciences* 122, 470-486.

689 Wanninkhof, R., 1992. RELATIONSHIP BETWEEN WIND-SPEED AND GAS-EXCHANGE OVER THE OCEAN.

690 *Journal of Geophysical Research-Oceans* 97, 7373-7382.

691 Wanninkhof, R., Asher, W.E., Ho, D.T., Sweeney, C., McGillis, W.R., 2009. Advances in Quantifying

692 Air-Sea Gas Exchange and Environmental Forcing. *Annual Review of Marine Science* 1, 213-244.

693 Wanninkhof, R., Knox, M., 1996. Chemical enhancement of CO₂ exchange in natural waters. *Limnology*

694 *and Oceanography* 41, 689-697.

695 Xiao, S., Yang, H., Liu, D., Zhang, C., Lei, D., Wang, Y., Peng, F., Li, Y., Wang, C., Li, X., Wu, G., Liu, L., 2014.

696 Gas transfer velocities of methane and carbon dioxide in a subtropical shallow pond. *Tellus Series*

697 *B-Chemical and Physical Meteorology* 66.

698 Yang, L., Lu, F., Wang, X., Duan, X., Tong, L., Ouyang, Z., Li, H., 2013. Spatial and seasonal variability of

699 CO₂ flux at the air-water interface of the Three Gorges Reservoir. *Journal of Environmental Sciences*

700 25, 2229-2238.
701 Yao, G.R., Gao, Q.Z., Wang, Z.G., Huang, X.K., He, T., Zhang, Y.L., Jiao, S.L., Ding, J., 2007. Dynamics Of
702 CO₂ partial pressure and CO₂ outgassing in the lower reaches of the Xijiang River, a subtropical
703 monsoon river in China. *Science of the Total Environment* 376, 255-266.
704 Zhang, J., Li, S., Dong, R., Jiang, C., 2018. Physical evolution of the Three Gorges Reservoir using
705 advanced SVM on Landsat images and SRTM DEM data. *Environmental Science and Pollution Research*
706 25, 14911-14918.
707 Zhang, T., Li, J., Pu, J., Martin, J.B., Khadka, M.B., Wu, F., Li, L., Jiang, F., Huang, S., Yuan, D., 2017. River
708 sequesters atmospheric carbon and limits the CO₂ degassing in karst area, southwest China. *Science*
709 *of The Total Environment* 609, 92-101.
710

711 **Table 1.** Statistics of all the data from three river systems (separated statistics please
 712 refer to Figs. S2 and S3 in the Supplementary material)s.

		Water T (°C)	pH	Alkalinity (µeq/l)	pCO ₂ (µatm)	DO%	DOC (mg/L)
Number		115	115	115	115	56	114
Mean		22.5	8.39	2589.1	846.4	91.5	6.67
Median		22.8	8.46	2560	588.4	88.8	2.51
Std. Deviation		6.3	0.29	640.7	818.5	8.7	7.62
Minimum		11.7	7.47	600	50.1	79.9	0.33
Maximum		34	9.38	4488	4830.4	115.9	37.48
Percentiles	25	16.3	8.24	2240	389.8	84.0	1.33
	75	29	8.56	2920	920.4	99.1	9.96
95% CI for Mean	Lower Bound	21.4	8.33	2470.8	695.2	89.1	5.26
	Upper Bound	23.7	8.44	2707.5	997.6	93.8	8.09

713

714 CI-Confidence Interval.

715 | **Table 2.** Comparison of different model for CO₂ areal flux estimation using combined
 716 | data (unit is mmol/m²/d for CO₂ areal flux and cm/h for k₆₀₀).
 717

CO ₂ areal flux ^a	From FC	Flow velocity-based model (Fig. 43b)	Water depth-based model (Fig.3a)	Alin's model	
k ₆₀₀	48.4 ^b	116.5 ^c	38.3	37.6	
Mean	198.1	476.7	156.6	154.0	
S.D.	185.5	446.2	146.6	144.2	
95% CI for Mean	Lower Bound	129.5	311.5	102.3	100.6
	Upper Bound	266.8	641.8	210.8	207.4

718

719 CI-Confidence Interval

720 a-CO₂ areal flux is based on TBL model.

721 b-mean level that is determined using floating chambers (FC).

722 c-This figure is revised to be 49.6 cm/h if the model ($k_{600} = 62.879FV + 6.8357$, $R^2 =$
 723 | 0.52 , $p=0.019$) is used (Fig. 43c), and the corresponding CO₂ areal flux is 203 ± 190
 724 | mmol/m²/d.

725

726 | **Table 3.** CO₂ emission during monsoonal period (May through Oct.) from total rivers
 727 | sampled in the study.

728 | (a) Upscaling using CO₂ areal flux by FC during monsoonal period.

	Catchment Area km ²	Water surface km ²	CO ₂ areal flux mmol/m ² /d	CO ₂ emission Tg CO ₂ /y
Danang	4200	21.42	122.0 ± 239.4	0.042 <u>0.021</u>
Qijiang	4400	30.8	50.3 ± 177.2	0.025 <u>0.0125</u>
TGR river	50000	377.78	217.7 ± 334.7	1.32 <u>0.666</u>
Total				1.39 <u>0.70</u>

729 |
 730 | (b) Upscaling using determined k₆₀₀ average and models (whole dataset are used
 731 | here).

	From determined k ₆₀₀ mean	Flow velocity-based model (Fig. 43b) (numbers in bracket is from the revised model; Fig. 43c)	Water depth-based model (Fig. 43a)	Alin's model
Mean				1.06 <u>0.53</u>
S.D.	0.69 <u>1.37</u>	1.66 (0.71) <u>3.29 (1.40)</u>	0.54 <u>1.08</u>	1.06 <u>0.50</u>
95% CI				
for				
Lower				
Bound				1.00 <u>0.35</u>
Mean	0.45 <u>0.89</u>	1.08 (0.46) <u>2.15 (0.91)</u>	0.36 <u>0.71</u>	0.69 <u>0.35</u>
Upper				
Bound	0.93 <u>1.84</u>	2.23 (0.94) <u>4.43 (1.81)</u>	0.74 <u>1.46</u>	1.43 <u>0.72</u>

732 | A total water area of approx. 430 km² for all tributaries (water area is from Landsat

733 | ETM+ in 2015); CO₂ emission upscaling was conducted during the monsoonal period

734 | because of the sampling in this period.

带格式的：下标

带格式的：下标

带格式表格

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

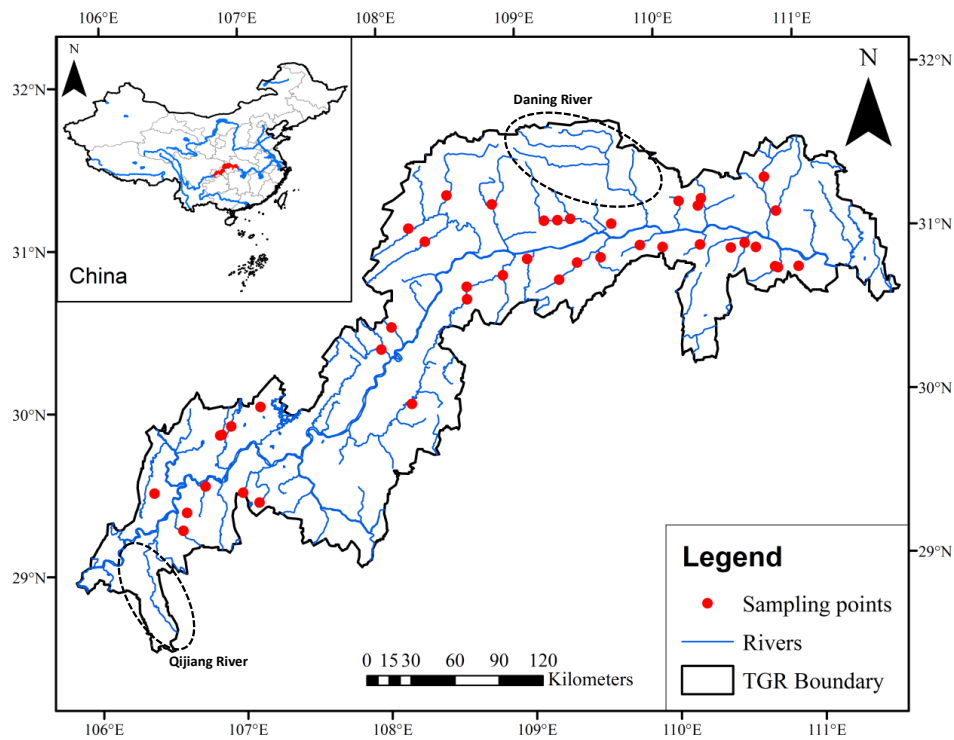
带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

带格式的：字体：(默认) Times
New Roman, 小五

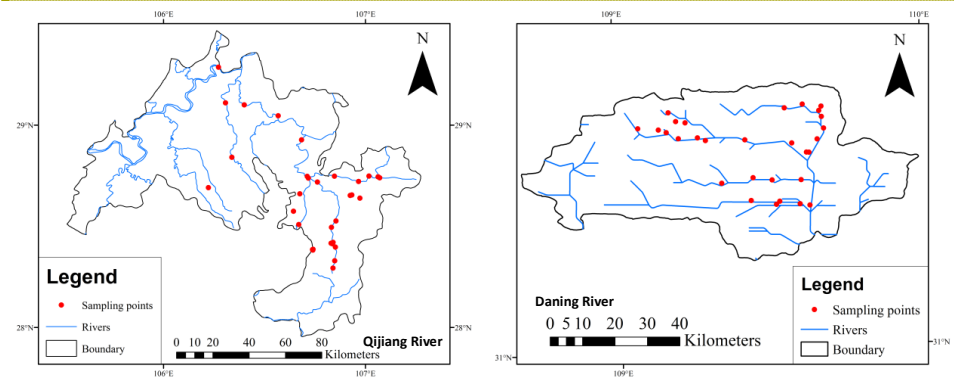
带格式的：字体：(默认) Times
New Roman, 小五

带格式的：下标



带格式的: 字体颜色: 文字 1
带格式的: 字体: 小四, 加粗, 字体颜色: 文字 1

735



带格式的: 字体: 小四, 加粗, 字体颜色: 文字 1

736

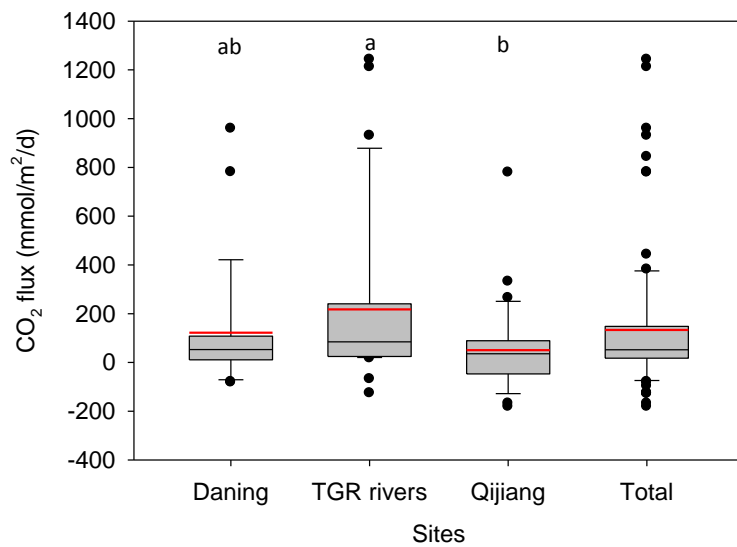
737

Fig. 1. Map of sampling locations of major rivers and streams in the Three Gorges Reservoir region, China.

带格式的: 字体: (默认) Times New Roman

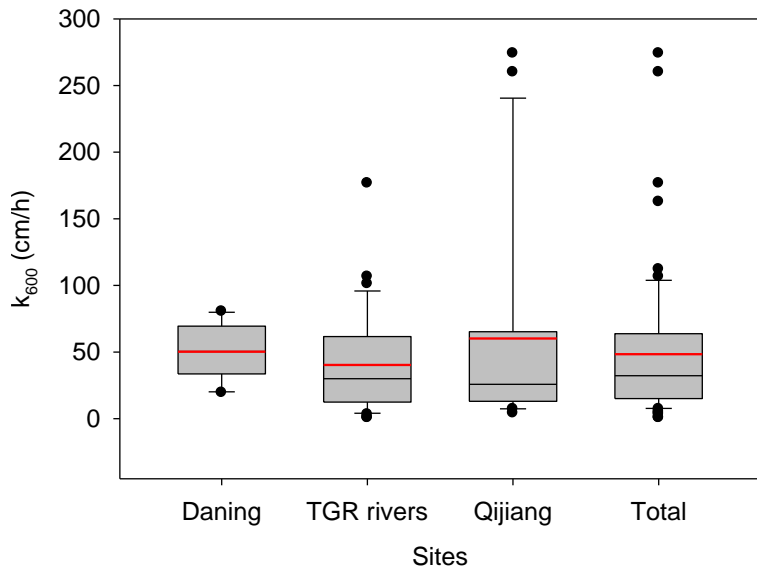
738

739



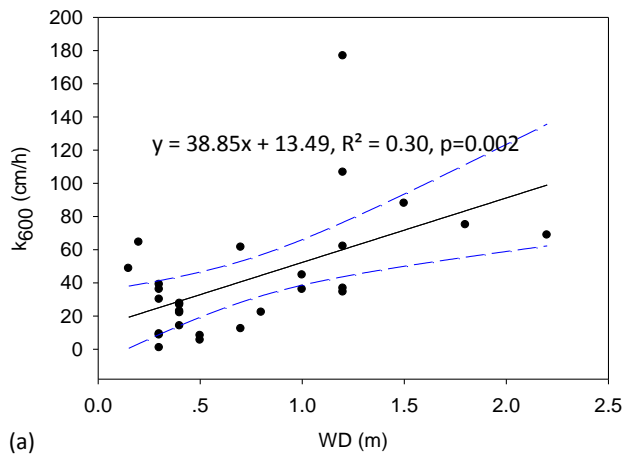
740

741 **Fig. 21.** Boxplots of CO₂ emission rates by floating chambers in the [investigated three](#)
 742 [river systems subtropical rivers](#) (different letters represent statistical differences at
 743 $p < 0.05$ by **Mann-Whitney Rank Sum Test**). (the black and red lines, lower and upper
 744 edges, bars and dots in or outside the boxes demonstrate median and mean values,
 745 25th and 75th, 5th and 95th, and <5th and >95th percentiles of all data, respectively).
 746 (For interpretation of the references to color in this figure legend, the reader is
 747 referred to the web version of this article) ([Total means combined data from three](#)
 748 [river systems](#)).

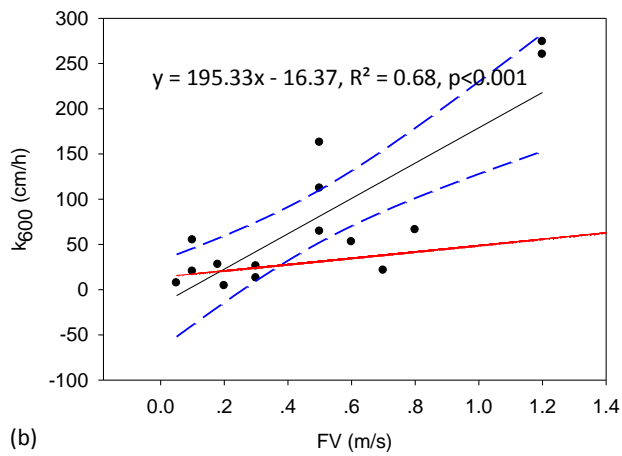


749

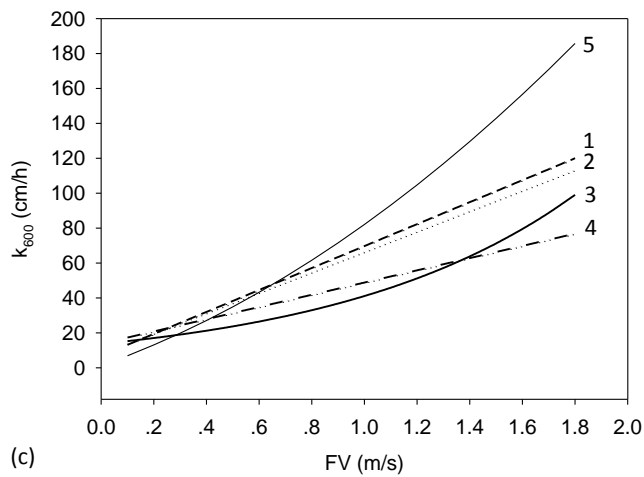
750 | **Fig. 32.** Boxplots of k_{600} levels in the investigated three river systems subtropical rivers
 751 (there is not a statistically significant difference in k among sites by Mann-Whitney
 752 Rank Sum Test). (the black and red lines, lower and upper edges, bars and dots in or
 753 outside the boxes demonstrate median and mean values, 25th and 75th, 5th and
 754 95th, and <5th and >95th percentiles of all data, respectively). (For interpretation of
 755 the references to color in this figure legend, the reader is referred to the web version
 756 | of this article) (Total means combined data from three river systems).



757 (a)



758 (b)



759 (c)

760

761 | Fig. 43. The k_{600} as a function of water depth (WD) using data from TGR rivers (a),

762 | flow velocity (FV) using data from in Qijiang (b), and comparison of the developed
763 | model with other models (c) (others without significant relationships between k and
764 | physical factors are not shown). The solid lines show regression, the dashed lines
765 | represent 95% confidence band, and the red dash-dotted line represents the model
766 | developed by Alin et al (2011) (~~in panel b~~, if several extremely values are removed in
767 | panel b, the revised model would be $k_{600} = 62.879FV + 6.8357$, $R^2 = 0.52$, $p=0.019$) (in
768 | panel c, 1-the revised model, 2-model from Ran et al., 2017, 3-model from Ran et al.,
769 | 2015, 4-model from Alin et al., 2011, 5-model from Liu et al., 2017) (1- $k_{600} =$
770 | $62.879FV + 6.8357$; 2- $k_{600} = 58.47FV+7.99$; 3- $k_{600} = 13.677\exp(1.1FV)$; 4- $k_{600} = 35 FV$
771 | $+ 13.82$; 5- $k_{600} = 6.5FV^2 + 12.9FV+0.3$) (unit of k in models 1-4 is cm/h, and unit of
772 | m/d for model 5 is transferred to cm/h).

773

Supporting material

Gas transfer velocities of CO₂ in subtropical monsoonal climate streams and small rivers

Siyue Li*, Rong Mao, Yongmei Ma, V.V.S.S. Sarma

Appendix Table A1. Dataset in relation to our study.

Seasonal changes related to gas transfer and concentrations

We used our measured CO₂ emission rates by FCs for upscaling flux estimates and it was found to be 1.39 TgCO₂/y for all rivers sampled (Table S3a). The estimated emission was close to that of the revised model (1.40 ± 1.31 (95% confidence interval: 0.91-1.87) Tg CO₂/y), and using the determined k average, i.e., 1.37 ± 1.28 (95% confidence interval: 0.89-1.84) Tg CO₂/y, but slightly higher than the estimation using water-depth based model (1.08 ± 1.01 Tg CO₂/y) and Alin's model (1.06 ± 1.00 Tg CO₂/y) (Table S3b). The estimate was within the range of our earlier work using TBL on the TGR river networks (0.64-2.33 Tg CO₂/y) (Li *et al.*, 2018). The higher emission, i.e., 3.29 ± 3.08 (2.15-4.43) Tg CO₂/y, using flow velocity based model may be over-estimated (Table 3b). Therefore, this study suggests that CO₂ emissions from rivers and streams in this area may be underestimated, i.e., 0.03 Tg CO₂/y (Li *et al.*, 2017) and 0.37-0.44 Tg CO₂/y (Yang *et al.*, 2013) as the former used TBL model with a lower k level, and the latter employed floating chambers, but they both sampled

带格式的: 字体: 加粗

带格式的: 缩进: 首行缩进: 0 厘米

very limited tributaries (i.e., 2-3 rivers). Therefore, measurements of k must be made mandatory along with pCO₂ measurement in the river and stream studies.

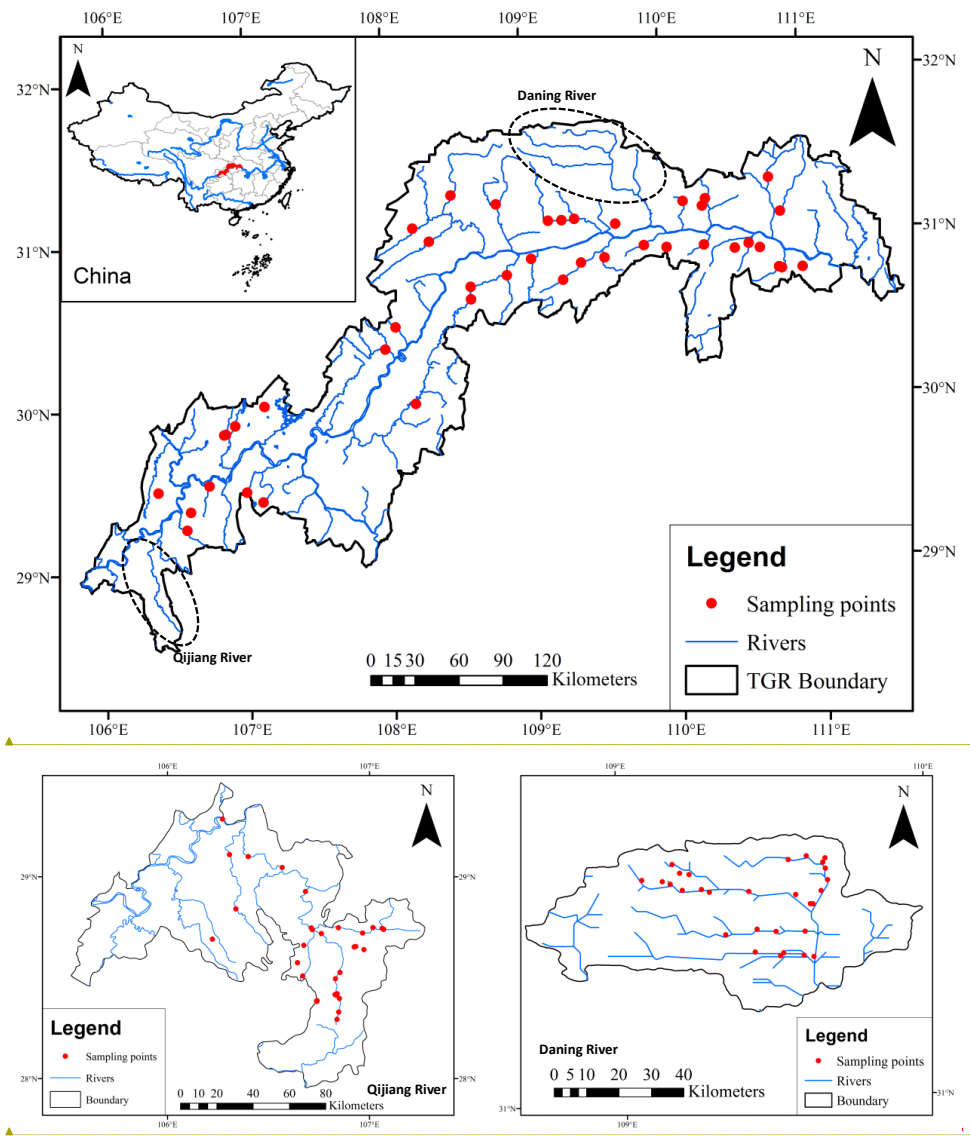
带格式的：下标

As our sampling was limited to monsoonal periods only, which could not provide a meaningful annual emission estimate given the restricted sampling period. Thus, we developed a stronger discussion on what changes across seasons related to gas transfer and pCO₂ concentrations. As outlined in the main text, riverine pCO₂ in the monsoonal season in this region was 81% the level in the dry season, and current velocity was 1.7-fold higher in monsoonal season (Li *et al.*, 2018), thus k₆₀₀ was 1.6-fold higher in the monsoonal period based our model. This could be defensible due to that prior study on the Yellow River reported that k₆₀₀ level in the wet season was 1.8-fold higher than in the dry season (Ran *et al.*, 2015), another study on the Wuding River demonstrated that k level in the wet season was 83%-130% of that in the dry season (Ran *et al.*, 2017). Moreover, a factor of 1.4 for water area was designated based on other monsoonal rivers. Then annual emission could be estimated at 1.59 Tg CO₂/y, slightly higher than the estimation using the data in the monsoonal period only.

带格式的：缩进：首行缩进： 2 字符

带格式的：下标

带格式的：下标



带格式的: 字体颜色: 文字 1

带格式的: 字体: 小四, 加粗, 字体颜色: 文字 1

带格式的: 字体: 小四, 加粗, 字体颜色: 文字 1

Fig. S1. Map of sampling locations of major rivers and streams in the Three Gorges Reservoir region, China.

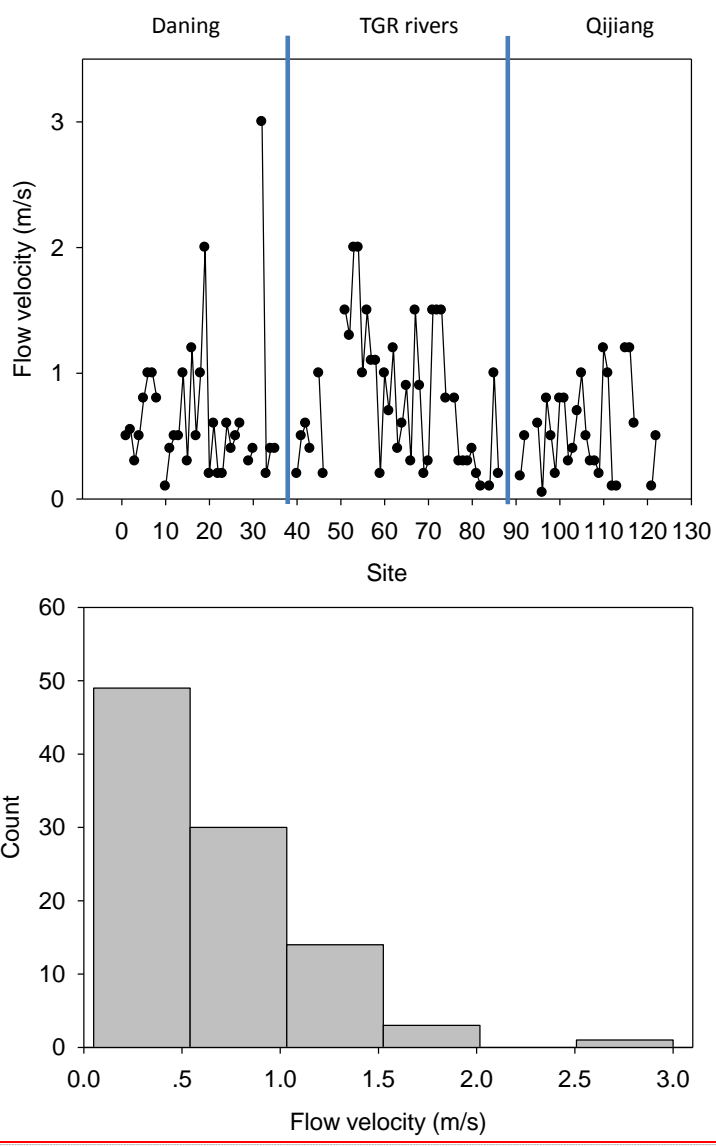


Fig. S1. Current velocity in our studied rivers.

域代码已更改

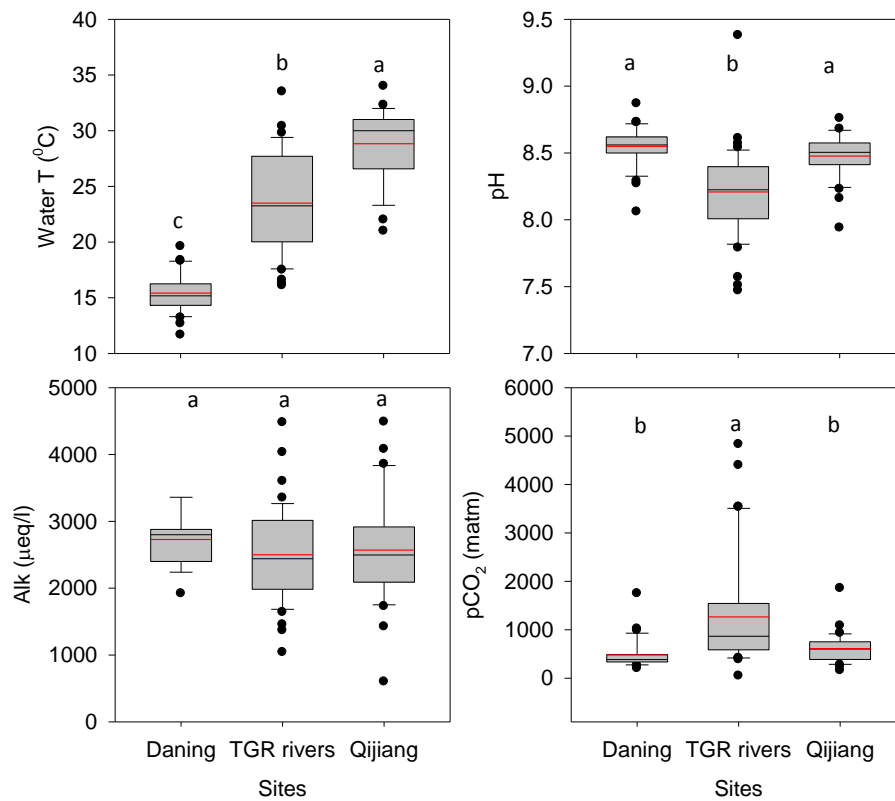


Fig. S2. Boxplots of water temperature, pH, alkalinity and pCO₂ in the subtropical rivers (different letters represent statistical differences at p<0.05). (the black and red lines, lower and upper edges, bars and dots in or outside the boxes represent median and mean values, 25th and 75th, 5th and 95th, and <5th and >95th percentiles of all data, respectively). (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article).

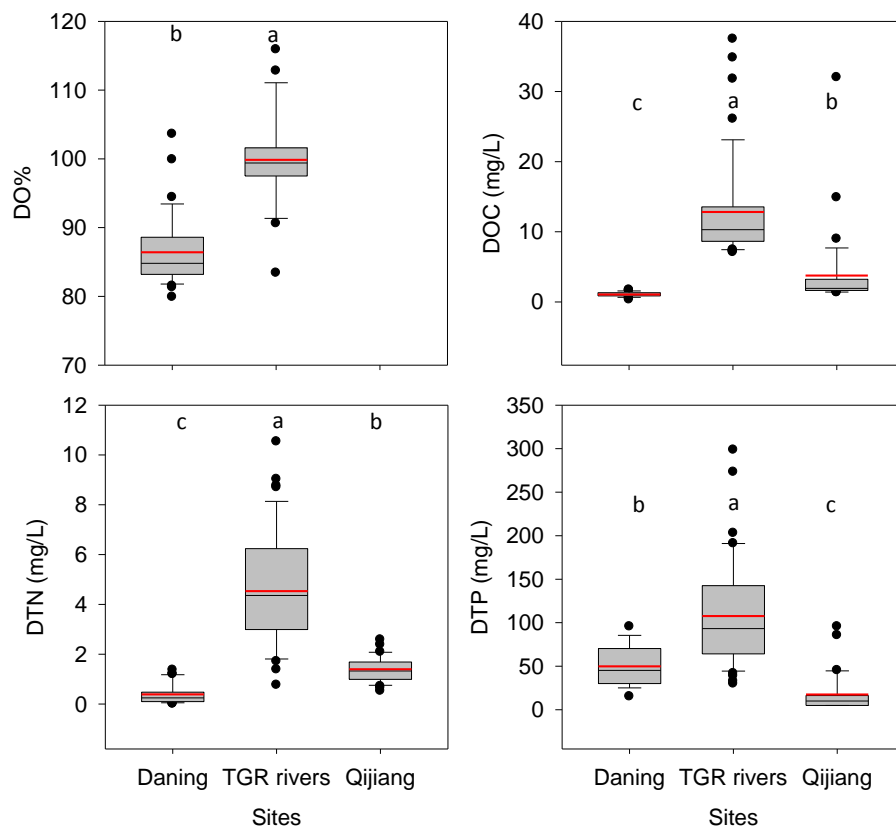


Fig. S3. Boxplots of organic carbon and nutrients in the subtropical rivers (different letters represent statistical differences at $p < 0.05$). (the black and red lines, lower and upper edges, bars and dots in or outside the boxes demonstrate median and mean values, 25th and 75th, 5th and 95th, and $<5^{\text{th}}$ and $>95^{\text{th}}$ percentiles of all data, respectively). (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article).

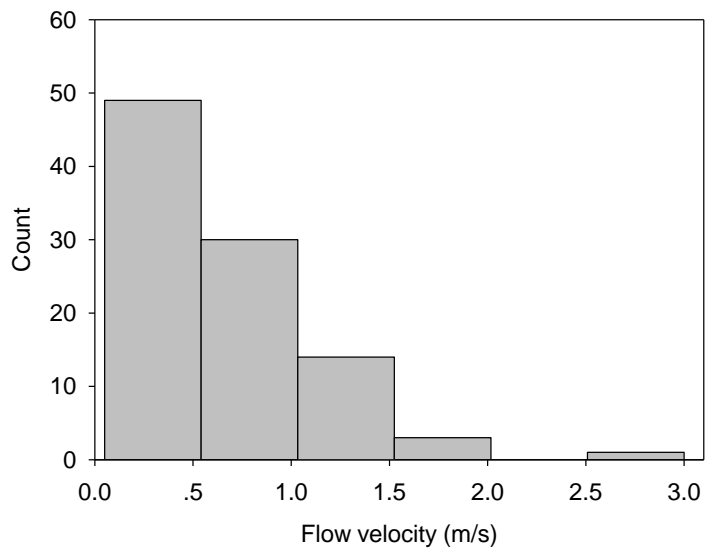
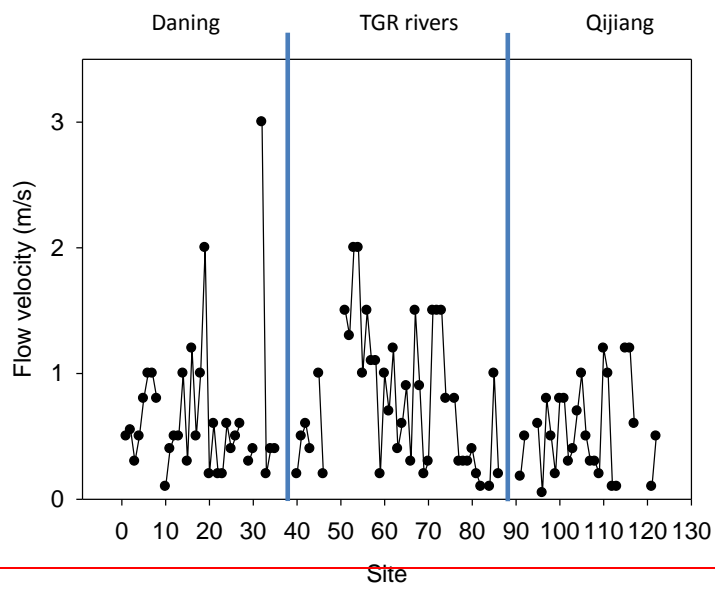


Fig. S4. Current velocity in our studied rivers.

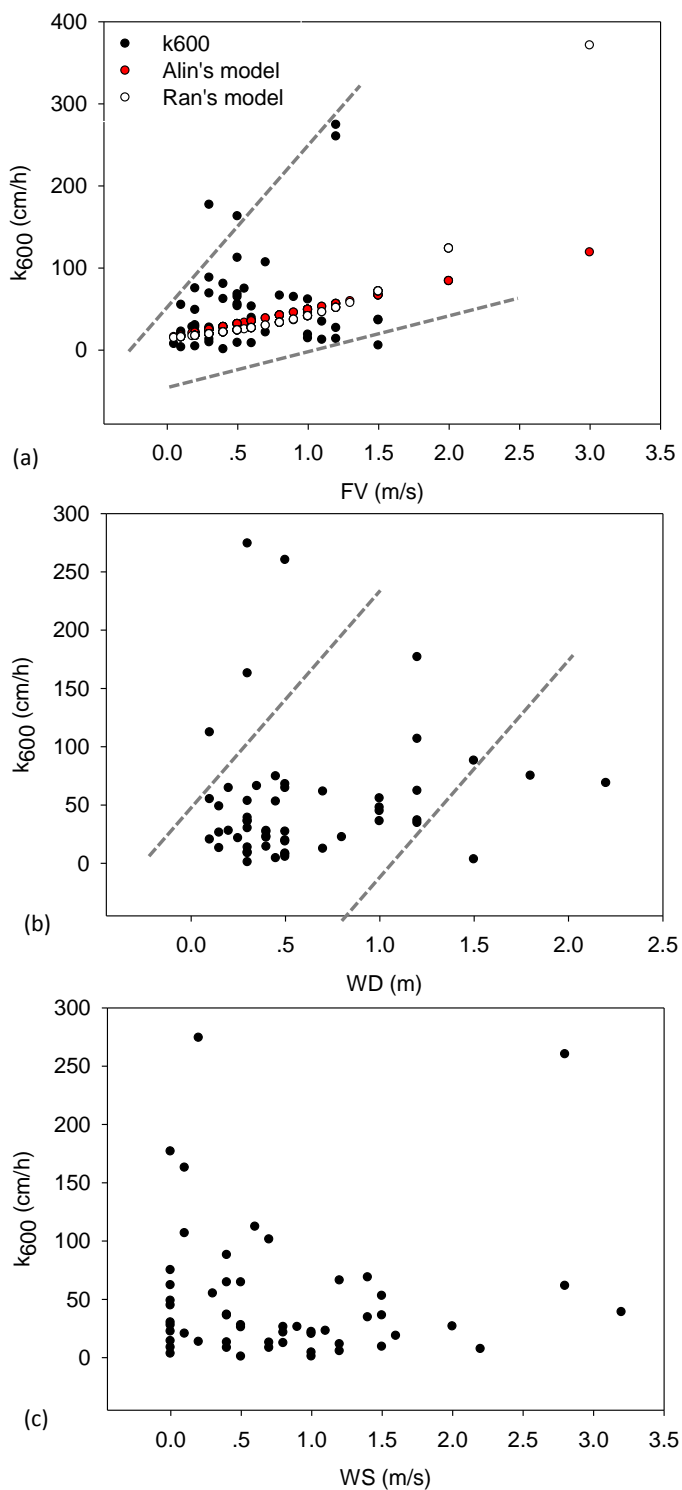


Fig. S45. Plots of k_{600} and flow velocity (FV, unit in m/s) (a), WD (water depth, unit in m) (b), and WS (wind speed, unit in m/s) (c). The circle and red-dotted lines in Fig. 3a

represent the model developed by Ran et al. (2015) and Alin et al (2011), respectively.

Table S1. Estimated k_{600} levels in our rivers.

		Danang	TGR rivers	Qijiang	Total
Mean		50.23	40.36	60.20	48.41
Median		50.46	30.04	25.81	32.22
Std. Deviation		20.10	37.60	78.85	53.20
Minimum		19.48	0.75	4.22	0.75
Maximum		80.50	176.71	274.12	274.12
Percentiles	25	35.809	12.628	13.213	16.303
	75	67.754	61.537	64.827	63.194
95% CI for Mean	Lower Bound	35.86	27.02	24.31	35.12
	Upper Bound	64.61	53.69	96.10	61.70

CI-Confidence Interval

Table S2. Comparison of k values adopted by previous studies.

River	Sites	Climate	k_{600}	Mean pCO ₂	CO ₂ degassing flux	References
			cm/h	(μ atm)	mmol/m ² /d	
Longchuan River	China	Subtropic	8	2100	156.2	Li et al., 2012
Upper stream of Maotiao River	China	Subtropic	10	3740	294.5	Wang et al., 2011
Pearl River	China	Humid subtropic	8-15	2600	189-356	Yao et al., 2007
Yangtze (Datong)	China	Subtropic	8	1297	38-148	Wang et al., 2007
Rivers in TGR area*	China	Subtropic	48.4 \pm 53.2	846 \pm 819	133.1 \pm 269.1	This study
Yangtze*	China	Subtropic	38 \pm 40	1000-2035		Liu et al., 2017
Yellow river network*	China	Temperate	42.1 \pm 16.9	2810 \pm 1985	856 \pm 409	Ran et al., 2015
Wuding*	China	Temperate	29-37	881	116-218	Ran et al., 2017

*k levels are from measurements

	Current velocity	Water depth	Wind speed	k_{600}	Reference
	m/s	m/s	m/s	cm/h	
Mekong tributary	0.39 \pm 0.28	0.9 \pm 0.6	0.7 \pm 0.6	23.3 \pm 17.3	Alin et al., 2011
Yellow	1.8		1.8 (1.2-2.3)	42 \pm 17	Ran et al., 2015
Yangtze	1.2 \pm 1.5		1.2 \pm 1.1	38 \pm 40	Liu et al., 2017
Mekong stem	0.92 \pm 0.42		1.8 \pm 1.2	15 \pm 9	Alin et al., 2011

带格式的: 下标

带格式的: 下标

Table S3. CO₂ emission from total rivers sampled in the study.

(a) Upscaling using CO₂ areal flux by FC.

	<u>Catchment Area</u> <u>km²</u>	<u>Water surface</u> <u>km²</u>	<u>CO₂ areal flux</u> <u>mmol/m²/d</u>	<u>CO₂ emission</u> <u>Tg CO₂/y</u>
<u>Daning</u>	<u>4200</u>	<u>21.42</u>	<u>122.0 ± 239.4</u>	<u>0.042</u>
<u>Qijiang</u>	<u>4400</u>	<u>30.8</u>	<u>50.3 ± 177.2</u>	<u>0.025</u>
<u>TGR river</u>	<u>50000</u>	<u>377.78</u>	<u>217.7 ± 334.7</u>	<u>1.321</u>
<u>Total</u>				<u>1.39</u>

(b) Upscaling using determined k₆₀₀ average and models (whole dataset are used here).

	<u>From</u> <u>determined</u> <u>k₆₀₀ mean</u>	<u>Flow velocity-based model (Fig. 4b)</u> <u>(numbers in bracket is from the</u> <u>revised model; Fig. 4c)</u>	<u>Water depth-based model</u> <u>(Fig. 4a)</u>	<u>Alin's</u> <u>model</u>
<u>Mean</u>	<u>1.37</u>	<u>3.29 (1.40)</u>	<u>1.08</u>	<u>1.06</u>
<u>S.D.</u>	<u>1.28</u>	<u>3.08 (1.31)</u>	<u>1.01</u>	<u>1.00</u>
<u>95% CI</u> <u>for</u> <u>Mean</u>	<u>Lower</u> <u>Bound</u> <u>0.89</u>	<u>2.15 (0.91)</u>	<u>0.71</u>	<u>0.69</u>
	<u>Upper</u> <u>Bound</u> <u>1.84</u>	<u>4.43 (1.87)</u>	<u>1.46</u>	<u>1.43</u>

A total water area of approx. 430 km² for all tributaries (water area is from Landsat ETM+ in 2015).

Table S4. Monsoonal sampling effects on annual emission

	<u>Monsoonal season</u>	<u>Dry season</u>	<u>Monsoonal/Dry</u>
<u>pCO₂ (µatm)</u>	<u>846</u>	<u>1043</u>	<u>0.81</u>
<u>ΔpCO₂ (µatm)</u>	<u>446</u>	<u>643</u>	<u>0.69</u>
<u>k₆₀₀ (cm/h)</u>	<u>48.4</u>	<u>31.3</u>	<u>1.55</u>
<u>CO₂ areal flux</u> <u>(mmol/m²/d)</u>	<u>196.9</u>	<u>183.4</u>	
<u>Water area (km²)</u>	<u>602</u>	<u>430</u>	
<u>Emission (Tg CO₂)</u>	<u>0.96</u>	<u>0.63</u>	

References

Alin, S.R., Rasera, M., Salimon, C.I., Richey, J.E., Holtgrieve, G.W., Krusche, A.V., Snidvongs, A., 2011. Physical controls on carbon dioxide transfer velocity and flux in low-gradient river systems and implications for regional carbon budgets. *Journal of Geophysical Research-Biogeosciences* 116. G01009.

- Li, S.Y., Lu, X.X., He, M., Zhou, Y., Li, L., Ziegler, A.D., 2012. Daily CO₂ partial pressure and CO₂ outgassing in the upper Yangtze River basin: A case study of the Longchuan River, China. *Journal of Hydrology* 466, 141-150.
- Liu, S., Lu, X.X., Xia, X., Yang, X., Ran, L., 2017. Hydrological and geomorphological control on CO₂ outgassing from low-gradient large rivers: An example of the Yangtze River system. *Journal of Hydrology* 550, 26-41.
- Ran, L.S., Lu, X.X., Yang, H., Li, L.Y., Yu, R.H., Sun, H.G., Han, J.T., 2015. CO₂ outgassing from the Yellow River network and its implications for riverine carbon cycle. *Journal of Geophysical Research-Biogeosciences* 120, 1334-1347.
- Wang, F., Wang, B., Liu, C.Q., Wang, Y., Guan, J., Liu, X., Yu, Y., 2011. Carbon dioxide emission from surface water in cascade reservoirs–river system on the Maotiao River, southwest of China. *Atmospheric Environment* 45, 3827-3834.
- Wang, X.F., He, Y.X., Yuan, X.Z., Chen, H., Peng, C.H., Zhu, Q., Yue, J.S., Ren, H.Q., Deng, W., Liu, H., 2017. pCO₂ and CO₂ fluxes of the metropolitan river network in relation to the urbanization of Chongqing, China. *Journal of Geophysical Research-Biogeosciences* 122, 470-486.
- Yao, G.R., Gao, Q.Z., Wang, Z.G., Huang, X.K., He, T., Zhang, Y.L., Jiao, S.L., Ding, J., 2007. Dynamics Of CO₂ partial pressure and CO₂ outgassing in the lower reaches of the Xijiang River, a subtropical monsoon river in China. *Science of the Total Environment* 376, 255-266.

**Dataset APPENDIX TABLE A1
(Daning River)**

Time	River	Latitude	Longitude	Air pressure	a.s.l.	T(water)	pH	EC(Cond)	EC(SPC)	TDS	DO	DO	River width	River depth	Water velocity	Alkalinity	DOC	TDN	TDP
				hPa	m	°C		µs/cm	µs/cm	mg/L	%	mg/L	m	m	m/s	µeq/L	mg/L	mg/L	µg/L
2016.5.19,9:21	XJH1	31°23'54"	109°31'25"	979	290	15.58	8.27	266.8	325.3	211	81.3	8.09	5	0.5	0.5	3200	1.09	1.37	30.15
2016.5.19,10:25	XJH2	31°23'48"	109°30'48"	982.1	240	18.294	8.06	353.1	405	263	88.9	8.36	10	0.45	0.55	3360	1.57	1.21	25.13
2016.5.19,11:00	XJH3	31°24'20"	109°31'26"	988	160	18.138	8.29	347.2	399.5	260	86.1	8.12	10	0.5	0.3	3360	1.74	1.07	45.23
2016.5.19,11:42	XJH4	31°23'54"	109°33'33"	989	210	18.275	8.39	349	400.4	260	83.7	7.87	10	0.3	0.5	3360	1.32	1.20	30.15
2016.5.19,15:20	HXH1	31°27'48"	109°21'14"	967	340	14.326	8.58	151.5	190.2	124	87	8.91	15	0.5	0.8	2320	1.15	0.54	30.15
2016.5.19,16:00	HXH2	31°28'49"	109°26'0"	973	320	14.136	8.53	159.1	200.7	130	84	8.63	8	2	1	2240	0.71	0.16	75.38
2016.5.19,16:32	HXH3	31°27'54"	109°30'8"	979	260	14.473	8.53	162.9	203.8	132	84.4	8.61	10	0.5	1	2240	0.67	0.38	55.28
2016.5.19,17:07	HXH4	31°27'46"	109°36'49"	987	260	15.177	8.55	211.7	260.6	169	84.8	8.51	20	0.8	0.8	2400	1.28	0.36	40.20
2016.5.19,17:32	DNH1	31°27'35"	109°38'6"	1000	230	15.221	8.56	220.4	271.1	176	84.6	8.48				2400	1.39	0.12	85.43
2016.5.20,10:19	DXH1	31°39'48"	109°33'7"	958	540	16.248	8.53	218.4	262	170	84.1	8.25	0.5	0.1	0.1	2800	1.41	0.18	40.20
2016.5.20,10:55	DXH2	31°39'46"	109°33'6"	965	560	15.106	8.55	226.3	279	181	84.4	8.49	50	0.3	0.4	2800	1.21	0.33	50.25
2016.5.20,11:20	DXH3	31°39'3"	109°39'46"	970	430	14.55	8.62	222.6	278.1	181	83.6	8.51	2	0.3	0.5	2880	0.89	0.26	75.38
2016.5.20,11:54	DXH4	31°38'3"	109°40'11"	973	130	16.153	8.6	241	289.5	188	82.5	8.12	25	0.5	0.5	2800	1.20	0.17	55.28
2016.5.20,14:09	DXH5	31°36'16"	109°40'34"	974	360	16.389	8.71	242.2	289.9	188	87.6	8.57	20	1	1	2720	0.98	0.66	15.08
2016.5.20,14:40	DXH6	31°34'23"	109°39'11"	981	320	16.417	8.71	238.5	285.2	185	88.6	8.66	25	0.5	0.3	2800	1.34	0.33	30.15
2016.5.20,15:03	DNH2	31°31'31"	109°37'18"	985	270	14.886	8.62	189.4	234.7	153	89.7	9.07	50	1	1.2	2880	0.66	0.12	40.20
2016.5.20,15:35	DNH3	31°28'1"	109°38'9"	987	260	16.689	8.73	193.5	229.9	149	94.4	9.16	50	1	0.5	2480	0.97		40.20
2016.5.20,16:23	XXH1	31°32'16"	109°37'5"	983	260	14.772	8.67	176.2	219	142	92.8	9.4	25	1.5	1	2400	1.02	0.21	95.48
2016.5.20,16:53	DXH7	31°32'13"	109°36'39"	984	280	15.984	8.61	228.9	276.5	180	89	8.78	8	1.5	2	2560	0.99	0.08	50.25
2016.5.20,18:15	DXH8	31°39'48"	109°40'15"		500	15.84	8.56	279.1	338.2	220	82.8	8.2	2	0.2	0.2	3120	1.19	0.24	75.38

2016.5.21;10:18	XXH2	31°39'48"	109°10'8"	922	860	11.671	8.54	131.5	176.4	115	81.5	8.84	4	0.5	0.6	2640	0.95	0.04	45.23
2016.5.21;10:57	XXH3	31°38'11"	109°11'58"	888		15.278	8.73	239.1	293.5	191	82	8.22	1.5	0.1	0.2	2400	0.78	0.42	30.15
2016.5.21;11:11	XXH4	31°37'46"	109°14'17"	947	670	13.934	8.71	187.7	238	155	85.8	8.85	15	1	0.2	2880	1.62	0.46	55.28
2016.5.21;13:15	XXH5	31°37'12"	109°4'31"	912	930	13.93	8.57	187.6	237.9	155	79.9	8.24	4	0.2	0.6	2800	0.87	0.45	70.35
2016.5.21;13:47	XXH6	31°36'10"	109°7'53"	934	780	12.688	8.5	135.8	177.6	115	83.2	8.82	8	0.5	0.4	2800	0.33	0.17	85.43
2016.5.21;14:24	XXH7	31°37'7"	109°8'27"	929	780	13.358	8.56	200.5	257.9	168	82.8	8.65	3	0.5	0.5	2960	1.59	0.07	85.43
2016.5.21;14:55	XXH8	31°36'29"	109°9'59"	939	700	14.696	8.46	210.7	262.4	171	83	8.42	3	0.3	0.6	2880	0.53	0.08	25.13
2016.5.21;15:21	XXH9	31°35'21"	109°12'16"	944	660	14.319	8.57	151.7	190.6	124	83.9	8.58	25	0.3		2240	0.87	0.11	30.15
2016.5.21;15:43	XXH10	31°35'23"	109°16'4"	951	600	15.092	8.6	155.3	191.6	125	85	8.55	25	0.4	0.3	2240	0.69	0.08	25.13
2016.5.21;16:16	XXH11	31°35'23"	109°17'28"	945	660	13.219	8.46	125.9	162.5	106	84.9	8.9	4	0.2	0.4	1920	0.70	0.01	35.18
2016.5.21;17:05	XXH12	31°34'37"	109°25'54"	967	490	16.142	8.87	198.9	239.5	156	103.6	10.19				2720	0.95	0.01	65.33
2016.5.21;18:07	XXH13	31°33'54"	109°34'13"	982	340	14.455	8.46	176.3	220.8	144	92.1	9.39	8	2.5	3	2560	1.13	0.33	65.33
2016.5.22;9:26	DNH4	31°23'25"	109°37'27"	997	210	16.187	8.5	213.2	256.2	166	86.8	8.54	25		0.2	2800	1.19	0.09	30.15
2016.5.22;9:44	XJH5	31°23'26"	109°37'25"	997	180	18.37	8.35	358.7	410.7	267	86.1	8.08	8		0.4	3360	1.32	1.16	70.35
2016.5.22;10:15	DNH5	31°23'38"	109°37'49"	997	230	19.618	8.68	292	325.5	212	99.9	9.15	2	0.2	0.4	3200	1.10	0.72	45.23

(TGR rivers)

Time	Code	Latitude	Longitude	asl	T(air)	Wind speed	T(water)	pH	ORP	EC	DO	DO	River width	River depth	Water velocity	Alk	DOC	DTN	DTP
				m	°C	m/s	°C		mv	µs/cm	mg/L	%	m	m	m/s	µeq/L	mg/L	mg/L	µg/L
2016.6.21;13:41	BLXX	29°30'30"	106°59'26"	432	39.1	0.6	29.1	7.57		233						1880	26.08	3.88	76.12
2016.6.21;15:00	BLTH	29°26'47"	107°06'11"	700	33.2	1.1	24.7	7.47		225			15	0.4	0.2	1976	37.48	7.03	120.89

2016.6.22:09:04	BNB	30°01'50"	108°11'32"	616	28.2	0	20	7.96	125.8			11	0.3	0.5	1040	14.31	1.94	82.84	
2016.6.22:10:00	BLH	29°17'53"	108°55'00"	518	29.9	0.4	23.5	7.84	202.3			80	0.5	0.6	1840	22.03	3.09	143.28	
2016.6.22:12:00	BHJH	30°22'07"	107°59'10"	181	33.5	1	29	7.79	418			4	0.3	0.4	3128	13.56	3.09	138.81	
2016.6.22:13:14	BRXH	30°30'10"	108°03'45"	215	36.1	0.5	33.5	7.82	344						3112	31.80	7.19	108.95	
2016.6.22:18:00	BDH	31°18'07"	108°28'52"	193	32.5	2.8	29.3	8.25	128	976	8.49	115.9	6	0.7	1	2152	10.30	1.71	44.78
2016.6.23:10:00	BNH	31°06'21"	108°12'02"	43	31.9	0	30.4	8.06	159.3	359	6.61	90.6	6	0.15	0.2	4032	34.82	3.45	140.30
2016.6.23:11:25	BPLH	31°01'18"	108°18'58"	152	33.3	0.9	29.4	8.15	155.9	269	6.07	83.4				1688	10.47	2.96	140.30
2016.6.23:15:47	BMDX	30°44'15"	108°36'07"	217	31.6	0.8	22.2	8.42	115.2	203	8.34	101				1640	7.54	3.22	29.85
2016.6.23:16:22	BSBH	30°39'42"	108°36'07"	256	32.5	0	28.9	8.61	81	240	7.63	104.1				2536	16.56	3.58	62.69
2016.6.23:18:56	BDXH	30°48'04"	108°51'47"	132	31.2	0.7	29.4	8.21	58.9	870	8.36	112.8				3256	10.00	1.81	41.79
2016.6.24:14:34	BCTH	30°53'48"	109°02'20"	138	25.4	1.2	22.2	7.89	99.1	102.3	8.43	99.2	4	0.5	1.5	1368	9.95	2.66	95.52
2016.6.25:12:00	BMXH	30°45'40"	109°15'37"	558	22.3	2.8	17.7	8.14	85.5	211	8.98	98.9	20	0.4	1.3	2488	7.30	6.34	77.61
2016.6.25:13:30	BMXH-2	30°51'44"	109°23'34"	317	23.5	0.6	17.8	8.19	92.2	320	9	96.7	25	0.5	2	2704	8.68	4.43	71.64
2016.6.25:15:10	BJPH	30°53'19"	109°33'38"	149	23.3	1.4	17.6	8.38	64.1	265	8.89	99.4	30	1	2	2392	10.23	4.34	50.75
2016.6.26:12:23	BGDH	30°57'15"	109°50'36"	366	27	0	16.1	8.26	38.3	278	9.44	98.3	30	1.5	1	3016	9.92	1.97	59.70
2016.6.26:15:36	BBLH	30°56'09"	110°00'17"	197	25.1	0	16.6	8.36	7.9	279	8.54	94.3	30	0.5	1.5	3112	7.45	4.66	98.51
2016.6.27:11:32	BWFH	30°56'26"	110°16'15"	402	28.1	1.4	16.3	8.4	9.03	303	9.03	101.2	6	1.2	1.1	2520	8.53	5.42	57.46
2016.6.27:14:07	BLFD	30°56'16"	110°35'19"	216	30.4	0.8	21.3	8.35	48.6	290	8.56	99.5	4	0.7	1.1	2216		4.39	68.66
2016.6.27:15:00	BQGH	30°54'45"	110°29'20"	237	31.2	0	22.8	8.45	29.1	324	8.61	101.4	7	0.4	0.2	2904	9.81	6.88	80.60
2016.6.27:16:48	BJZH	30°54'30"	110°39'58"	220	29.5	0	22.1	8.45	33.5	320	8.14	101.2	7	0.4	1	2664	8.64	6.18	47.76
2016.6.27:18:00	BJWX	30°46'41"	110°49'00"	152	28.3	0.1	18.3	8.39	25.5	280	9.52	103.8	25	1.2	0.7	2616	8.33	4.76	104.48
2016.6.27:18:30	BJWX-2	30°47'04"	110°47'44"	242	29.2	2	20.1	8.33	21.8	314	8.63	98.2	3	0.4	1.2	2616	8.65	4.07	73.13
2016.6.28:10:20	BMPH	30°46'47"	110°57'44"	137	29.3	1	21.1	8.35	32.5	353	8.63	96.8	1.2	0.4	0.4	3352	11.20	3.61	161.19
2016.6.28:14:40	BGLH	31°07'27"	110°49'21"	120	33.9	3.2	21.5	8.52	-1.6	321	8.3	101.8	15	0.3	0.6	3008	22.38	8.76	44.78
2016.6.28:16:20	BXXH	31°20'07"	110°45'02"	185	33.1	1.5	19.9	8.51	31.9	330			12	0.3	0.9	3200	11.82	2.34	176.12
2016.6.29:10:44	BLTH-XXGQ	30°13'32"	110°32'47"	745	26.9	1.1	17.6	8.57	64.3	208	8.26	98.7	7	1.2	0.3	2616	7.08	1.81	191.04

2016.6.29;11:32	BSNX	31°13'22"	110°17'39"	125	33.7	0.4	19.3	9.38	8.6	232	22	1.2	1.5	2368	7.44	1.91	32.84
2016.6.29;14:32	BLXB	31°10'45"	110°16'07"	171	37.8	0.4	29.8	8.49	-2	378	6	0.2	0.9	3600	11.09	4.63	58.21
2016.6.29;16:31	BBCH	31°12'47"	110°08'03"	631	33.8	0	27.7	8.24	3.5	438	5	0.3	0.2	4480	21.81	9.03	86.57
2016.6.29;17:29	BCaoT	31°05'31"	109°38'46"	193	33.6	1.5	28.7	8.54	96.8	501	4	0.3	0.3	3176	11.23	8.70	202.98
2016.6.30;10:40	BCJH	31°07'56"	109°21'21"	129	30.7	0.4	22.8	8.5	20.3	371	22	1	1.5	2616	8.35	3.61	128.36
2016.6.30;11:10	BMeiXH	31°07'37"	109°15'54"	140	27.3	5	21.6	8.38	22.7	218	30	0.4	1.5	1968	9.53	4.60	85.82
2016.6.30;11:40	BCJB	31°07'35"	109°10'06"	288	27.8	0.5	22.9	8.34	23.8	162.8	35	1.2	1.5	1456	7.22	6.24	73.13
2016.6.30;15:21	BTXH	31°14'17"	108°48'00"	164	25	0.8	17.5	8.3	9.7	162	25	1.5	0.8	1936	8.28	0.76	38.81
2016.7.01;10:31	BGaoTH	30°17'31"	107°27'33"	331	25.9	0.7	23.3	7.93	32.3	249				2290	9.28	6.93	273.13
2016.7.01;11:28	BGJH	30°12'35"	107°29'47"	354	28.5	0	23.3	8.14	29.9	274	15	2.2	0.8	2270	12.18	3.52	101.49
2016.7.01;12:30	BQXH	30°06'21"	107°31'06"	378	29.2	1.4	23.4	8.17	69.3	258	25	2.2	0.3	2260	11.80	4.57	147.76
2016.7.01;14:41	BLXH	29°50'30"	107°08'58"	190	33.3	0	25.9	7.51	54.6	379	30	1.2	0.3	2340	12.30	8.08	191.04
2016.7.01;15:54	BTXH	30°01'59"	107°07'23"	328	29.6	0.4	24.7	7.86	40.7	377	15	1.5	0.3	2100	13.95	10.54	298.51
2016.7.02;10:50	BWTH	29°51'42"	106°50'08"	194	32.2	0	23.3	8.16	36.3	300	30	1.2	0.4	2664	10.59	2.01	89.55
2016.7.02;11:17	BYLH	29°51'54"	106°50'49"	149	34.6	0	24.4	8.17	37.6	264	45	1.8	0.2	1880	9.13	1.38	144.78
2016.7.02;11:37	BDH-YB	29°54'59"	106°54'51"	134	33.9	0	27.6	7.99	59.1	322	45	1.5	0.1	2296	10.86	4.93	94.78
2016.7.02;14:49	BCTH-YL	29°32'59"	106°43'35"	181	35.5	0.4	27.7	8.18	44	270				2008	10.72	4.57	135.82
2016.7.02;16:04	BHXH	29°23'19"	106°35'37"	208	31.2	0	23.2	8.08	73.2	303	20	0.8	0.1	2176	9.86	6.37	144.78
2016.7.02;17:25	BYPH	29°16'46"	106°34'01"	118	31	0	24.2	8.1	74.9	232	35	1	1	1912	8.43	3.55	91.79
2016.7.02;18:23	BLTH-BSY	29°30'35"	106°22'00"	271	31.3	0.4	28.1	7.87	12.6	462	6	0.6	0.2	3144	14.15	6.24	159.70

(Qijiang River)

Date	Time	River	T(air) °C	Wind speed m/s	T(water) °C	pH	EC µs/cm	TDS mg/L	River width m	River depth m	Water velocity m/s	Alk µeq/L	DOC mg/L	TDN mg/L	TDP mg/L
2016.8.15	13:30	BZD1	39.6	0.5	30	8.41	351.6	229	3	20	0.18	2920	1.34	2.58	10.05
	14:30	BZD2	40.5	1	25	8.59	317	206	20	40	0.5	2472	1.61	2.09	10.05

	15:30	BZD3	37.3	0.7	31	8.67	301.8	196				2656	2.32	1.99	5.03
	15:40	BZD4	34.4	0.6	31	8.52	269.1	175				1824	14.88	0.53	5.03
	16:10	BZD5	33.7	1.6	32	8.52	282.5	184	20	80	0.6	2336	1.48	0.87	5.03
	16:45	BZD6	34.4	2.2	30	8.3	448.2	292	2	50	0.05	2216	1.65	1.89	5.03
2016.8.16	9:10	BSY1	31	0.4	22	8.64	322.3	210	3	60	0.8	3056	2.00	1.01	5.03
	9:50	BSY2	33.1	0.5	23.3	8.49	345	224	6	50	0.5	2896	1.65	0.97	10.05
	10:25	BSY3	31.6	1	23.5	7.94	374.7	244	20	45	0.2	2520	1.39	1.12	
	13:40	BSK1	32.2	1	26.8	8.63	513.8	334	1	15	0.8	2552	4.46	1.99	5.03
	14:20	BSK2	36.1	1.2	21	8.27	338.6	220	2	35	0.8	2288	32.01	2.05	10.05
	15:00	BSK3	41.8	0.5	30.5	8.67	286.7	186	12	50	0.3	2600	3.72	0.97	10.05
	15:40	BSK4	35.4	1.3	31	8.16	416.2	271	1.5	60	0.4	1896	1.88	0.99	5.03
	16:00	BSK5	38.3	0.8	26.5	8.68	334.5	218	3	25	0.7	2048	2.53	1.57	10.05
	16:20	BSK6	40	0.8	29	8.56	381.2	248	2	50	1	1728	1.50	1.29	20.10
	17:00	BSK7	38.1	0.6	32.3	8.76	335.2	218	2	10	0.5	1984	1.65	0.86	5.03
2016.8.17	9:30	BSK8	30.2	0.8	23.3	8.48	368.9	240	11	15	0.3	3112	1.64	1.55	10.05
	10:00	BSK9	32.3	0.4	26	8.4	362.9	236	11	15	0.3	2632	1.72	1.45	45.23
	11:00	BSK10	34.4	0.2	29	8.56	367.4	239	20	35	0.2	2368	1.86	1.29	
	11:24	BSK11	35.5	2.8	29.1	8.52	410.8	207	5	50	1.2	2544	1.40	0.71	5.03
	12:45	BSK12	38.5	1.6	30	8.46	474.8	309	12	50	1	2904	1.44	1.33	40.20
	15:00	BYD1	35	0.3	29.5	8.45	663.6	431	1	10	0.1	2240	2.25	2.38	10.05
	16:30	BYD2	32.7	0.8	30	8.58	217.6	142	4	10	0.1	1808	4.74	1.03	5.03
	17:45	BYD3	33	0.5	29	8.45	1087.9	707				4488	2.00	1.62	15.08
	18:14	BYD4	32.6	0.2	32	8.55	1099.5	716	5	30	1.2	3592	2.49	1.58	15.08
	17:00	BYD5	32	0.2	30.5	8.53	1051.9	684	4	30	1.2	4080	1.73	1.26	5.03
	19:30	BQJ1	32.3	1.5	30	8.56	623	405	40	45	0.6	3784	1.83	1.43	15.08
2016.8.18	10:00	BQJ2	37	1.2	29.5	8.35	396.4	258				2416	2.99	1.26	95.48

11:20	BPH1	33.2	0.7	31.9	8.49	1350.3	878				3856	8.98	1.47	25.13
15:00	BSX1	41.4	0.1	32	8.42	150.9	98				600	3.46	0.64	5.03
16:00	BSX2	41.2	1	34	8.23	179.9	117	8	10	0.1	1424	3.31	1.33	30.15
16:30	BQG3	37.1	0.1	32	8.43	422.2	274	15	30	0.5	2376	2.46	1.71	85.43
