



1           **Warming effect on nitrogen fixation in Mediterranean macrophyte sediments**

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17   **Running head:** Nitrogen fixation in Mediterranean macrophytes

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19   **Key words:** Nitrogen fixation, Coastal sediments, Macrophytes, Seagrasses, Macroalgae,

20   Warming, Thermal dependence, Activation Energy, Q<sub>10</sub>, Mediterranean Sea.



21 **Abstract**

22           The Mediterranean Sea is warming faster than the global Ocean, with important  
23 consequences for organisms and biogeochemical cycles. Warming is a major stressor for key  
24 marine benthic macrophytes. However, the effect of warming on marine N<sub>2</sub> fixation remains  
25 unknown, despite the high productivity of macrophytes in oligotrophic waters is partially  
26 sustained by the input of new nitrogen (N) into the system by N<sub>2</sub> fixation. Here, we assess the  
27 impact of warming on N<sub>2</sub> fixation rates on three key marine macrophytes: *Posidonia oceanica*,  
28 *Cymodocea nodosa*, and *Caulerpa prolifera*. We experimentally measured N<sub>2</sub> fixation rates in  
29 vegetated and bare sediments at temperatures encompassing current summer mean, projected  
30 summer mean and projected summer maximum seawater surface temperatures (SST) by the end  
31 of the century under a scenario of moderate greenhouse-gas emissions. We found that N<sub>2</sub> fixation  
32 rates in vegetated sediments were 2.8-fold higher than in bare sediments at current summer mean  
33 SST, with no differences among macrophytes. Currently, the contribution of N<sub>2</sub> fixation to  
34 macrophytes productivity could account for up to 7 %, 13.8 % and 1.8 % of N requirements for  
35 *P. oceanica*, *C. nodosa*, and *C. prolifera*, respectively. We show the temperature dependence of  
36 sediment N<sub>2</sub> fixation rates. However, the thermal response differed for vegetated sediments,  
37 where rates showed an optimum at 31 °C followed by a sharp decrease at 33 °C, and bare  
38 sediments, where rates increased along the range of experimental temperatures tested here. The  
39 activation energy and Q<sub>10</sub> were lower in vegetated sediments than in bare sediments. The  
40 projected warming is expected to increase the contribution of N<sub>2</sub> fixation to Mediterranean  
41 macrophytes' productivity. Therefore, the thermal dependence of N<sub>2</sub> fixation might have  
42 important consequences for primary production in coastal ecosystems in the context of warming.



## 43 1. Introduction

44 Global mean surface temperatures have increased 0.85 °C from 1880 to 2012 and are  
45 projected to increase between 1 and 3.5 °C by the end of the 21<sup>st</sup> century, relative to pre-  
46 industrial times (IPCC, 2013). Similarly, heatwaves are more frequent since the second half of  
47 the 20<sup>th</sup> century in Europe, Asia and Australia (IPCC, 2013; Perkins et al., 2012). Oceans store  
48 most of the accumulated heat in the biosphere, warming at an average rate of 0.11 °C per decade  
49 at surface (up to 75 m depth) since 1970 (IPCC, 2013), with longer and more frequent marine  
50 heatwaves over the last century (Oliver et al., 2018). Warming is larger in small and enclosed  
51 basins such as the Mediterranean Sea (Vargas-Yáñez et al., 2008), which is warming at 0.61 °C  
52 per decade (Belkin, 2009), with rapid migration of marine isotherms (Burrows et al., 2011) and  
53 increased thermal extremes (Difffenbaugh et al., 2007). Specifically, the maximum surface  
54 seawater temperature (SST<sub>max</sub>) in the Balearic Islands, in the western Mediterranean Sea, is  
55 projected to increase by  $3.4 \pm 1.3$  °C by 2100 under a scenario of moderate greenhouse-gas  
56 (GHG) emissions (A1B scenario from the IPCC Special Report on Emissions Scenarios,  
57 equivalent to the RCP6.0 scenario of the IPCC Fifth Assessment Report) (Jordà et al., 2012),  
58 with important consequences for marine organisms and ecosystems.

59 The effect of warming on Mediterranean marine organisms is reflected in changes in  
60 abundance, survival, fertility, phenology, and migration (Marbà et al., 2015). In particular,  
61 warming increases the mortality rates of the Mediterranean endemic and key marine angiosperm,  
62 *Posidonia oceanica* (Marba and Duarte, 2010), which is predicted to be functional extinct by  
63 2049 due to warming (Chefaoui et al., 2018; Jordà et al., 2012). In coastal benthic ecosystems,  
64 warming also affects metabolic processes driving biogeochemical cycles. Warming enhances  
65 sediment sulfate reduction rates (Robador et al., 2016), leading to an increase in sulfide



66 accumulation in coastal bare sediments (Sanz-Lázaro et al., 2011) and seagrass colonized  
67 sediments (Koch et al., 2007). In the Mediterranean Sea, heatwaves and warming trigger sulfide  
68 intrusion in *P. oceanica* shoots (García et al., 2013), which has toxic effects on plant meristems  
69 (Garcias-Bonet et al., 2008) and increases shoot mortality (Calleja et al., 2007). Sanz-Lázaro et  
70 al. (2011) found that warming enhances sediment oxygen uptake and CO<sub>2</sub> emissions in coastal  
71 sediments, boosted by the addition of labile organic matter, in a mesocosm experiment.  
72 Similarly, warming together with eutrophication have been identified as main drivers of hypoxia  
73 in a Mediterranean macroalgae *Caulerpa prolifera* meadow (Vaquer-Sunyer et al., 2012).  
74 However, studies on the effect of warming on atmospheric nitrogen fixation in coastal marine  
75 ecosystems are lacking.

76 Nitrogen (N<sub>2</sub>) fixation plays a fundamental role in balancing nutrient budgets at the basin  
77 scale in the Mediterranean Sea, with most of this N<sub>2</sub> fixation associated with *P. oceanica*  
78 seagrass meadows (Béthoux and Copin-Montégut, 1986). Indeed, endophytic nitrogen-fixing  
79 bacteria have been detected in roots of *P. oceanica* (Garcias-Bonet et al., 2012; Garcias-Bonet et  
80 al., 2016) and N<sub>2</sub> fixation has been reported in leaves and roots of *P. oceanica* (Agawin et al.,  
81 2016; Lehnen et al., 2016) and *in situ* incubations (Agawin et al., 2017). Therefore, the high  
82 productivity of *P. oceanica* in the oligotrophic Mediterranean waters is partially supported by N<sub>2</sub>  
83 fixation. However, the magnitude of N<sub>2</sub> fixation rates in the rhizosphere of *P. oceanica* is still  
84 unknown, as well as N<sub>2</sub> fixation associated to other key macrophytes commonly found in the  
85 Mediterranean Sea, such as *Cymodocea nodosa* and *Caulerpa prolifera*. In addition, whether  
86 warming will affect N<sub>2</sub> fixation rates is still unknown.

87 Here, we experimentally assess the response N<sub>2</sub> fixation rates in coastal Mediterranean  
88 ecosystems to warming. We focus specifically on three key Mediterranean macrophytes: two



89 seagrass species (*P. oceanica*, *C. nodosa*) and one green macroalgae species (*C. prolifera*). We  
90 experimentally measured N<sub>2</sub> fixation rates in vegetated and bare sediments at temperatures  
91 encompassing current summer mean SST (25 and 27 °C), projected summer mean SST (29 and  
92 31 °C), and projected summer SST<sub>max</sub> (33 °C) by the end of the century under a scenario of  
93 moderate GHG emissions to assess: i) differences between vegetated and bare sediments, ii)  
94 differences among macrophyte species and iii) the thermal dependence of sediment N<sub>2</sub> fixation  
95 rates.

96

## 97 **2. Materials and methods**

98

### 99 **2.1. Study site**

100 The study was conducted with communities sampled in Pollença Bay (Mallorca, Spain), a  
101 bay located in the western Mediterranean Sea (39° 53.792' N; 3° 5.523' E). The study site was  
102 selected based on the coexistence of the three most commonly found macrophyte species in the  
103 region, including two seagrasses (*Posidonia oceanica* and *Cymodocea nodosa*) and one green  
104 macroalgae (*Caulerpa prolifera*). The three macrophytes grow close to each other in  
105 monospecific patches at 5-m depth. The study was conducted in mid-June 2017 when *in situ*  
106 daily mean (±SE) SST was 26.4 ± 0.08 °C.

107 We sampled sediment colonized by these three macrophytes and the adjacent bare  
108 sediment using sediment cores (50 cm length and 4.5 cm in diameter). We collected the sediment  
109 samples by pushing the cores down into the sediment with the help of a rubber hammer and  
110 carefully extracting at least 15 cm of undisturbed top sediment. The cores were transported  
111 immediately to the laboratory. We measured seawater salinity using a calibrated conductivity



112 meter (ProfiLine Cond 3310, WTW®, USA) and summer SST was monitored and recorded *in*  
113 *situ* every 2 hours from 2012 to 2017 using a Hobo logger (Onset Computer Corporation®, MA,  
114 USA). Full summer SST records are available for 2012, 2013, 2106, and 2017. No data is  
115 available for 2015, and only partial temporal coverage is available for 2014 (Fig. 1). The summer  
116 mean ( $\pm$ SE) SST varied from  $26.29 \pm 0.05$  °C in 2013 to  $27.03 \pm 0.04$  °C in 2017, with an  
117 average summer mean SST of  $26.54 \pm 0.17$  °C for the last years. Average summer minimum and  
118 maximum SST were 22.92 and 29.08 °C, respectively. The highest maximum SST was 29.67 °C  
119 and was registered in August 2017 (Table 1).

120

## 121 2.2. Nitrogen fixation rates

122 We measured sediment N<sub>2</sub> fixation rates by the Acetylene Reduction Assay (Capone and  
123 Taylor, 1980) in *P. oceanica*, *C. nodosa* and *C. prolifera* vegetated sediments and the adjacent  
124 bare sediment at five incubation temperatures: 25, 27, 29, 31, and 33 °C. The 25 and 27 °C  
125 temperature treatments represent the current summer mean SST, covering the *in situ* recorded  
126 average summer mean SST of  $26.54 \pm 0.2$  °C (25 % percentile = 25.81 °C and 75 % percentile =  
127 27.61 °C) from 2012 to 2017. The 29 and 31 °C temperature treatments represent the range of the  
128 projected summer mean SST by the end of the century under a scenario of moderate GHG  
129 emissions equivalent to RCP6.0, applying the projected mean SST increase of  $2.8 \pm 1.1$  °C in the  
130 region (Jordà et al., 2012) over the current summer mean SST. The 33 °C temperature treatment  
131 represent the projected summer SST<sub>max</sub> by the end of the century under a scenario of moderate  
132 GHG emissions equivalent to RCP6.0, by applying the projected SST<sub>max</sub> increase of  $3.4 \pm 1.3$  °C  
133 (Jordà et al., 2012) over the summer SST<sub>max</sub> of 29.67 °C already recorded in 2017.



134           Once in the laboratory, the sediment from the cores was extruded carefully using a  
135 plunger, and the first 10 cm of sediment below surface were collected, the rest of the sediment in  
136 the core was discarded. For each replicate, 80 ml of sediment was placed in a 500 ml glass bottle.  
137 Then, we added 200 ml of autoclaved seawater and the bottles were closed with a lid fitted with  
138 a gas-tight valve. Finally, we added 20 ml of acetylene-saturated seawater through the gas-tight  
139 valve of each bottle in order to achieve a final acetylene concentration of 4 mM. The acetylene-  
140 saturated seawater was prepared according to Wilson et al. (Wilson et al., 2012). We run the  
141 sediment incubations in triplicate for each type of sediment and each temperature treatment. The  
142 sediment incubations were run under dark conditions in five water baths (i.e. one per temperature  
143 treatment) equipped with thermometers and heaters located in a stable temperature room. The  
144 target temperature for each water bath was maintained using an IKS-AQUASTAR system which  
145 controlled and recorded the temperature every 10 minutes. During the 24 h incubations, the  
146 temperature oscillation around the target temperatures ranged from 0.3 to 0.7 °C, and the  
147 temperature accuracy was  $\pm 0.05$  °C on average (Table 2). The incubations lasted 24h, starting  
148 after the addition of acetylene-saturated seawater. We sampled the headspace five times: at the  
149 start of the experiment and at 12, 17, 20, and 24 h since the onset of the experiment. Specifically,  
150 we withdrew 3 ml of air from the headspace with a gas-tight syringe. The headspace air sample  
151 was immediately injected into a 3 ml vacuum vial for further analysis of ethylene concentration  
152 on a gas chromatographer equipped with a flame ionization detector (FID-GC, Agilent 5890)  
153 using a PoraPLOT U GC column (25 m  $\times$  0.53 mm  $\times$  20  $\mu$ m, Agilent Technologies, USA). We  
154 built a calibration curve using three ethylene standards of known concentration (1.02, 10.13 and  
155 99.7 ppm) and Helium as a balance gas, supplied by Carbueros Metálicos S.A. (Palma de  
156 Mallorca, Spain). We estimated the concentration of dissolved ethylene, from the ethylene



157 concentration in the equilibrated air as described previously (Wilson et al., 2012) and applying  
158 the solubility coefficient of ethylene according to Breitbarth et al. (Breitbarth et al., 2004) as a  
159 function of temperature and salinity.

160 We run negative controls consisting in sediment without addition of acetylene-saturated  
161 seawater in order to confirm that ethylene was not naturally produced by our samples, and  
162 autoclaved seawater used in the preparation of the incubations with addition of acetylene-  
163 saturated seawater in order to confirm that ethylene was not produced in the seawater. The  
164 ethylene production rates were converted into N<sub>2</sub> fixation rates by applying the common ratio of  
165 3 mol of acetylene:1 mol of N<sub>2</sub> (Welsh, 2000).

166 At the end of the incubation, we dried the sediment samples at 60 °C and recorded the dry  
167 weight for further calculations. Moreover, we calculated the sediment organic matter (OM)  
168 content of each replicate sediment sample by loss on ignition (Dean Jr, 1974). The sediment N<sub>2</sub>  
169 fixation rates were standardized to surface area integrated over 10 cm sediment depth by taking  
170 into account the sediment bulk density.

171

### 172 **2.3. Statistical analysis**

173 Differences in sediment OM content and bulk density among *P. oceanica*, *C. nodosa*, and  
174 *C. prolifera* vegetated sediments and bare sediment were tested with the non-parametric Kruskal  
175 Wallis test. Differences in sediment N<sub>2</sub> fixation rates among the four types of sediment were  
176 tested by Friedman test matching by temperature treatment. Differences in sediment N<sub>2</sub> fixation  
177 rates between vegetated and bare sediments were tested by non-parametric Mann-Whitney *U*  
178 test. Moreover, we tested the thermal dependence of sediment N<sub>2</sub> fixation rates in vegetated and  
179 bare sediments by fitting Arrhenius function to estimate the activation energy, derived from the





180 relationship between the natural logarithm of N<sub>2</sub> fixation rates and the inverse of the temperature  
181 multiplied by the Boltzmann's constant (Dell et al., 2011), and Q<sub>10</sub> (Raven and Geider, 1988).  
182 The activation energy and Q<sub>10</sub> of N<sub>2</sub> fixation in vegetated sediments were calculated using the  
183 increasing rates measured at four temperature treatments (25, 27, 29, and 31 °C), while the  
184 declining rates measured at 33 °C were not included. The full range of temperature treatments  
185 was used for bare sediments since no decline was detected. All statistical analyses were  
186 performed using JMP (SAS Institute Inc., USA) and PRISM (GraphPad Software Inc., USA)  
187 statistical software.

188

### 189 3. Results

190 Sediment OM content was significantly different in the sediments colonized by different  
191 macrophyte species ( $\chi^2_{3,56} = 50.33$ ,  $p < 0.0001$ ). *Posidonia oceanica* sediments had the highest  
192 OM content ( $13.34 \pm 0.56$  %), whereas bare sediments had the lowest OM content ( $0.44 \pm 0.50$   
193 %, Table 2). Sediment bulk density differed among sediment types ( $\chi^2_{3,56} = 46.02$ ,  $p < 0.0001$ ,  
194 Table 3).

195 Average N<sub>2</sub> fixation rates in bare sediments was  $0.06 \pm 0.01$  (range from 0.01 to 0.09)  
196 nmol N gDW<sup>-1</sup> h<sup>-1</sup>; while, average N<sub>2</sub> fixation rates in vegetated sediments was three-fold greater  
197 at  $0.19 \pm 0.03$  (range from 0.05 to 0.9) nmol N gDW<sup>-1</sup> h<sup>-1</sup>. Within the vegetated sediments, the  
198 maximum mean N<sub>2</sub> fixation rate was detected in *C. prolifera* ( $0.22 \pm 0.05$  nmol N gDW<sup>-1</sup> h<sup>-1</sup>),  
199 whereas the minimum mean N<sub>2</sub> fixation rate was measured in *C. nodosa* ( $0.15 \pm 0.04$  nmol N  
200 gDW<sup>-1</sup> h<sup>-1</sup>). Nitrogen fixation rates differed among the four different sediment types (i.e. bare, *P.*  
201 *oceanica*, *C. nodosa* and *C. prolifera* sediments) ( $\chi^2_{3,56} = 10.68$ ,  $p = 0.005$ ) when expressed by  
202 sediment dry weight. However, once the rates were converted into aerial basis, these differences



203 were no longer significant ( $\chi^2_{3,56} = 6.12$ ,  $p > 0.05$ ), due to high variability in sediment bulk  
204 densities. Sediment  $N_2$  fixation rates were independent of OM content (linear regression,  
205  $p > 0.05$ ).

206 Nitrogen fixation rates in aerial basis were significantly higher in vegetated sediments  
207 compared to bare ones ( $U = 154$ ,  $p < 0.002$ ) when pooling all temperature treatments together,  
208 with sediments colonized by macrophytes supporting, on average, twice the nitrogen fixation rate  
209 than bare sediments (mean  $\pm$  SE =  $3.86 \pm 0.53$  and  $1.77 \pm 0.20$  mg N m<sup>-2</sup> d<sup>-1</sup>, respectively). The  
210 differences between vegetated and bare sediment remained significant at current summer SST  
211 (25-27 °C) and projected summer mean SST (29-31 °C) ( $U = 13$ ,  $p < 0.01$  and  $U = 23$ ,  $p < 0.05$ ,  
212 respectively), with  $N_2$  fixation rates of  $3.15 \pm 0.48$  and  $1.14 \pm 0.3$  mg N m<sup>-2</sup> d<sup>-1</sup> in vegetated and  
213 bare sediments, respectively, at current summer SST (25-27 °C) and  $N_2$  fixation rates of  $5.25 \pm$   
214  $1.17$  and  $2.18 \pm 0.2$  mg N m<sup>-2</sup> d<sup>-1</sup> in vegetated and bare sediments, respectively at projected  
215 summer mean SST (29-31 °C). However, vegetated and bare sediments did not differ in  $N_2$   
216 fixation rates between vegetated and bare sediments at projected summer SST<sub>max</sub> (33 °C), with  
217  $N_2$  fixation rates of  $2.49 \pm 0.248$  and  $2.21 \pm 0.15$  mg N m<sup>-2</sup> d<sup>-1</sup>, respectively (Fig. 2).

218 In vegetated sediments,  $N_2$  fixation rates increased linearly with temperature up to 31°C  
219 ( $N_2$  fixation (nmol N g DW<sup>-1</sup> h<sup>-1</sup>) =  $-0.63 + 0.03 \cdot \text{Temperature}$ ,  $R^2 = 0.11$ ,  $p < 0.05$ ), with a  
220 marked decrease from  $0.32 \pm 0.09$  nmol N g DW<sup>-1</sup> h<sup>-1</sup> at 31 °C to  $0.11 \pm 0.01$  nmol N g DW<sup>-1</sup> h<sup>-1</sup>  
221 at 33 °C (Fig. 3). Nitrogen fixation rates in bare sediments increased linearly with temperature up  
222 to 33 °C ( $N_2$  fixation (nmol N g DW<sup>-1</sup> h<sup>-1</sup>) =  $-0.11 + 0.01 \cdot \text{Temperature}$ ,  $R^2 = 0.51$ ,  $p < 0.005$ , Fig.  
223 3). The associated activation energies were  $0.91 \pm 0.39$  and  $1.25 \pm 0.39$  eV for vegetated and  
224 bare sediments, respectively (Fig. 4). Associated  $Q_{10}$  values were  $3.84 \pm 2.22$  and  $6.41 \pm 2.97$  for  
225 vegetated and bare sediments, respectively.



#### 226 4. Discussion

227 The overall average N<sub>2</sub> fixation rate found in Mediterranean vegetated sediments at  
228 current summer mean SST ( $3.15 \pm 0.48$  mg N m<sup>-2</sup> d<sup>-1</sup>) is within the range of those rates reported  
229 for sediments colonized by temperate seagrass species (from 1.2 to 6.5 mg N m<sup>-2</sup> d<sup>-1</sup> in *Zostera*  
230 *marina* sediments in the North Sea (McGlathery et al., 1998) and the Northwest Atlantic coast  
231 (Capone, 1982), and from 0.1 to 7.3 mg N m<sup>-2</sup> d<sup>-1</sup> in *Zostera noltii* in the Northeast Atlantic coast  
232 (Welsh et al., 1996)), but lower than the rates reported for tropical and subtropical seagrass  
233 species (see references in (Welsh, 2000)). The overall N<sub>2</sub> fixation rates in vegetated sediments  
234 are higher than in bare sediments, consistent with the long-recognized role of marine plants in  
235 enhancing N<sub>2</sub> fixation rates (Capone, 1988). Specifically, the vegetated sediments supported 3 to  
236 4-fold higher N<sub>2</sub> fixation rates than bare sediments did.

237 The N<sub>2</sub> fixation rates we measured in *P. oceanica* sediments at current summer mean  
238 SST ( $2.86 \pm 1.26$  mg N m<sup>-2</sup> d<sup>-1</sup>) are higher than the rates reported by Agawin et al. (ranging from  
239 0.06 to 1.51 mg N m<sup>-2</sup> d<sup>-1</sup>) in summer, at similar seawater temperature, using benthic bell-jar  
240 chambers containing *P. oceanica* shoots and the underlying sediment (Agawin et al., 2017).  
241 However, the different methodological approaches make comparisons difficult: while sediment  
242 slurries might slightly overestimate rates due to sediment structure disturbance and increase in  
243 organic matter availability, incubation chambers might underestimate rates due to poor diffusion  
244 of acetylene into the sediment (Welsh, 2000). Nevertheless, the N<sub>2</sub> fixation rates in bare  
245 sediments at current summer mean SST ( $1.14 \pm 0.31$  mg N m<sup>-2</sup> d<sup>-1</sup>) are very similar to those  
246 measured by benthic bell-jar in bare sediment adjacent to a *P. oceanica* meadow (from 0.01 to  
247 1.99 mg N m<sup>-2</sup> d<sup>-1</sup>) (Agawin et al., 2017), suggesting that these differences in N<sub>2</sub> fixation rates in  
248 *P. oceanica* sediment might also be due to variability among sites. The N<sub>2</sub> fixation rates in *C.*



249 *nodosa* and *C. prolifera* sediments reported here are the first reports, to the best of our  
250 knowledge, for these two important Mediterranean macrophyte species. Indeed, the analysis of  
251 sediment N stocks in a *C. nodosa* meadow in the Mediterranean Sea suggested that N<sub>2</sub> fixation  
252 might be contributing to enhance the N stocks compared to bare sediments (Pedersen et al.,  
253 1997). The similar stable N isotope composition of *C. nodosa* tissues and those of *P. oceanica* in  
254 the Mediterranean (Fourqurean et al., 2007) also suggests that they use similar sources of  
255 nitrogen. The N<sub>2</sub> fixation rates at current summer mean SST in *C. prolifera* sediments found here  
256 ( $0.17 \pm 0.04$  nmol N gDW<sup>-1</sup> h<sup>-1</sup>) are similar to the sediment N<sub>2</sub> fixation rates associated to the  
257 invasive *C. taxifolia* in Monaco ( $0.12 \pm 0.09$  nmol N gDW<sup>-1</sup> h<sup>-1</sup>) but 20-fold lower than the N<sub>2</sub>  
258 fixation rates reported for *C. taxifolia* in France ( $3.96 \pm 1.99$  nmol N gDW<sup>-1</sup> h<sup>-1</sup>) (Chisholm and  
259 Moulin, 2003).

260         Although the sediments colonized by these three macrophyte species hold similar rates,  
261 the contribution of sediment N<sub>2</sub> fixation to the productivity of each plant is different. Taking into  
262 account the average net production ( $2.63$  and  $1.47$  g DW m<sup>-2</sup> d<sup>-1</sup> for *P. oceanica* and *C. nodosa*,  
263 respectively (Duarte and Chiscano, 1999) and  $5.16$  g DW m<sup>-2</sup> d<sup>-1</sup> for *C. prolifera* (Marbà  
264 unpublished)) and the tissue nitrogen content (from 1.55 to 1.63 % for *P. oceanica*, from 1.91 to  
265 2.28 % for *C. nodosa* (Duarte, 1990; Fourqurean et al., 2007) and from 3 to 4.9 % for *C. prolifera*  
266 (Morris et al., 2009)), the mean measured sediment N<sub>2</sub> fixation rates detected at current summer  
267 mean SST (25 and 27 °C) could account for 6.7 to 7 %, 11.5 to 13.8 % and 1.1 to 1.8 % of the  
268 nitrogen requirements for *P. oceanica*, *C. nodosa*, and *C. prolifera*, respectively. The calculated  
269 contribution of N<sub>2</sub> fixation to seagrass growth requirements fall within the range of the N<sub>2</sub>  
270 fixation contributions reported for temperate seagrasses, ranging from 5 to 12 % for *Z. marina*  
271 and *Z. noltii*, respectively (Welsh, 2000). The calculated contribution of N<sub>2</sub> fixation to fulfill the



272 macrophytes' growth requirements points N<sub>2</sub> fixation as partially supporting the high  
273 productivity of these primary producers in the Mediterranean oligotrophic waters.

274 We experimentally demonstrate that N<sub>2</sub> fixation rates in coastal sediments are thermal  
275 dependent, both in vegetated and bare sediments, in agreement with higher rates typically  
276 reported in tropical and subtropical meadows compared to temperate and cold seagrass systems  
277 (Herbert, 1999;McGlathery, 2008;Welsh, 2000); although an experimental demonstration was  
278 lacking. The thermal dependence, as reflected by the activation energy and Q<sub>10</sub>, for N<sub>2</sub> fixation  
279 rates was, however, higher in bare sediments than in vegetated sediments, possibly due to  
280 different bacterial communities. Westrich and Berner (1988) also found that sulfate reduction  
281 exhibited a more pronounced thermal dependence in sediments supporting lower rates. The  
282 activation energies for N<sub>2</sub> fixation in vegetated sediments ( $0.91 \pm 0.4$  eV or  $87.8 \pm 37.6$  KJ mol<sup>-1</sup>)  
283 and in bare sediments ( $1.25 \pm 0.4$  eV or  $120.6 \pm 37.6$  KJ mol<sup>-1</sup>) are within the range of the  
284 activation energy reported for sediment sulfate reduction (range from 36 to 132 KJ mol<sup>-1</sup>  
285 (Robador et al., 2016;Westrich and Berner, 1988)) and for sediment organic matter degradation  
286 (range from 54 to 125 KJ mol<sup>-1</sup> (Middelburg et al., 1996)). The Q<sub>10</sub> values associated to sediment  
287 N<sub>2</sub> fixation ( $3.84 \pm 2.22$  and  $6.41 \pm 2.97$  for vegetated and bare sediments, respectively) are  
288 higher than those reported for sediment sulfate reduction (from 1.6 to 3.4 (Robador et al., 2016)),  
289 but still similar to those values associated to organic matter degradation (from 2.2 to 6.3  
290 (Middelburg et al., 1996)). Moreover, the thermal response differed for vegetated sediments,  
291 where N<sub>2</sub> fixation rates showed an optimum at 31 °C followed by a sharp decrease at 33 °C, and  
292 bare sediments, where N<sub>2</sub> fixation rates increased along the range of experimental temperatures  
293 tested here. The thermal response of N<sub>2</sub> fixation in vegetated sediments found here is similar to  
294 the thermal response reported for N<sub>2</sub> fixation in soil crusts (Zhou et al., 2016) and seagrass



295 rhizosphere (Garcias-Bonet et al., 2018), with an increase in rates up to 30 and 29 °C,  
296 respectively, and a marked decrease in rates at temperatures above the optimum. The forecasted  
297 warming by the end of the century could potentially increase N<sub>2</sub> fixation rates by 36.7% in  
298 vegetated sediments and 46.8% in bare sediments. However, the decrease in N<sub>2</sub> fixation rates in  
299 vegetated sediments at 33 °C would imply a reduction of a third in the contribution of N<sub>2</sub> fixation  
300 to the macrophytes' productivity during heatwaves. Therefore, the thermal dependence of N<sub>2</sub>  
301 fixation in vegetated sediments might have important consequences for primary production in  
302 coastal ecosystems in the context of warming. However, this may not be the case for *P.*  
303 *oceanica*, as this species is projected to be critically compromised, to the extent that functional  
304 extinction is possible, with projected Mediterranean warming rates by 2050 (Chefaoui et al.,  
305 2018; Jordà et al., 2012).

306

### 307 **Conclusion**

308 Mediterranean macrophyte meadows are sites of intense N<sub>2</sub> fixation rates, twice as high  
309 as those in adjacent bare sediments. As these rates increased greatly with warming, realized  
310 warming of the Mediterranean Sea is expected to have led to enhanced N<sub>2</sub> fixation rates, both for  
311 bare and vegetated sediments, with further warming leading to further increase in N<sub>2</sub> fixation  
312 rates.

313

### 314 **Author contribution**

315 NG-B, RVS, CMD and NM designed the study. NG-B and RVS performed the field  
316 work, and sample and data analysis. NG-B, RVS, CMD and NM interpreted the results. NG-B



317 wrote the first draft of the manuscript. All authors contributed substantially to the final  
318 manuscript.

319

### 320 **Conflict of interests**

321 Authors declare that they have no conflict of interest

322

### 323 **Acknowledgements**

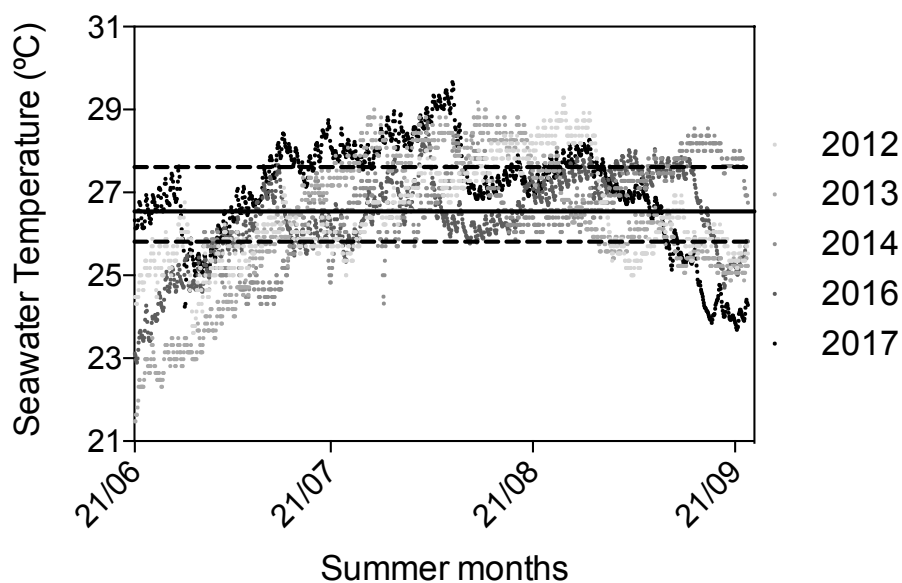
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326 University of Science and Technology to CMD. We thank Carlos Alex Morell Lujan Williams  
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329 **Figures**

330 Fig. 1. Summer *in situ* SST (from June 21<sup>st</sup> to September 22<sup>nd</sup>) from 2012 to 2017 at the study  
331 site. Solid line indicates average summer median SST and dashed lines indicate the 25 and 75%  
332 percentiles. Average summer median SST and percentiles do not include data from 2014 due to  
333 partial temporal coverage in temperature records. No data available for 2015.

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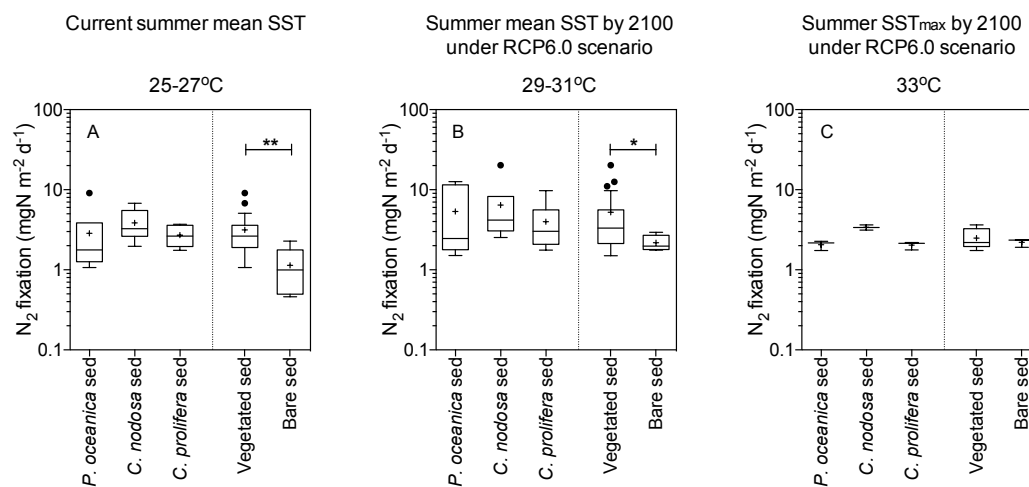
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337 Fig. 2. Box plot of the  $N_2$  fixation rates of sediments colonized by different macrophytes (as well  
 338 as grouping all vegetated sediments together) and bare sediment measured at current summer  
 339 SST (25-27 °C, (A)), projected summer mean SST by 2100 under RCP6.0 scenario (29-31 °C,  
 340 (B)), and projected summer SST<sub>max</sub> by 2100 under RCP6.0 (33 °C, (C)).  
 341

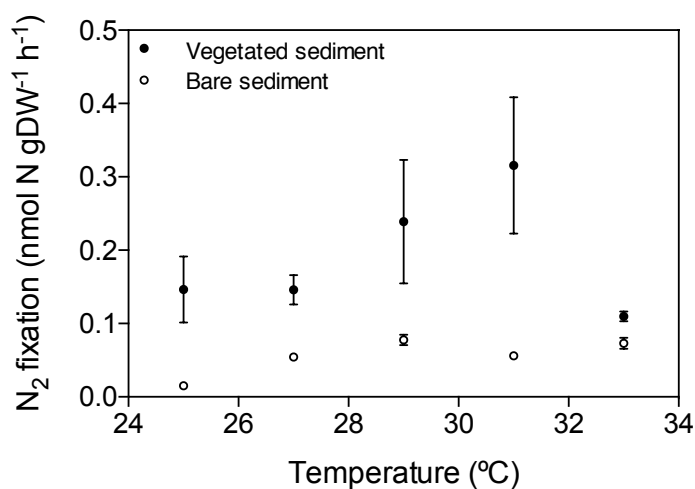


342

343



344 Fig. 3. Relationship of experimental incubation temperature and sediment N<sub>2</sub> fixation rates in  
345 vegetated (black dots) and bare sediments (white dots). Dots indicate mean values and errors  
346 bars indicate standard error of the mean.  
347



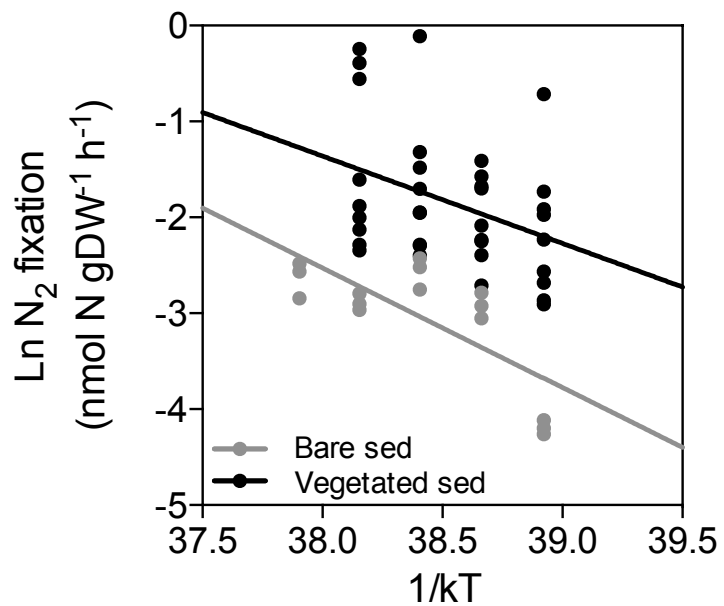
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351 Fig. 4. Arrhenius plot for N<sub>2</sub> fixation rates in vegetated (black dots and line) and bare sediments  
352 (grey dots and line).  
353



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356

357 **Tables**

358

359 Table 1. Summary of *in situ* summer SST at the study site (Pollença Bay, Mallorca) from June  
 360 21<sup>st</sup> to September 22<sup>nd</sup> from 2012 to 2017, showing the mean  $\pm$  SEM, minimum (Min) and  
 361 maximum (max), median, and 25 and 75% percentiles. nd: no data, na: not available due to  
 362 partial temporal coverage.

363

| Year                        | Summer Surface Seawater Temperature (SST) |                      |              |                |                |
|-----------------------------|---|----------------------|--------------|----------------|----------------|
|                             | Mean ( $\pm$ SEM)                         | Min - Max            | Median       | 25% percentile | 75% percentile |
| 2012                        | 26.44 $\pm$ 0.03                          | 23.61 – 29.29        | 26.23        | 25.70          | 27.28          |
| 2013                        | 26.29 $\pm$ 0.05                          | 21.48 – 29.18        | 26.28        | 25.23          | 27.98          |
| 2014                        | na  | na – 28.55           | na           | na             | na             |
| 2015                        | nd  | nd                   | nd           | nd             | nd             |
| 2016                        | 26.39 $\pm$ 0.03                          | 22.90 – 28.69        | 26.43        | 25.82          | 27.24          |
| 2017                        | 27.03 $\pm$ 0.04                          | 23.69 – 29.67        | 27.21        | 26.50          | 27.92          |
| <b>Average 2012 – 2017*</b> | <b>26.54 <math>\pm</math> 0.17</b>        | <b>22.92 – 29.08</b> | <b>26.54</b> | <b>25.81</b>   | <b>27.61</b>   |

364

365 \*including data from 2012, 2013, 2016, and 2017 for which full summer records are available.

366

367



368 Table 2. Experimental temperature set up used in the sediment incubations, indicating the five  
369 target temperatures and the actual mean ( $\pm$ SEM) temperature recorded in each incubation water  
370 bath during the 24h incubations. The ranges (minimum – maximum values) are shown in  
371 parenthesis.

372

| Target temperature (°C) | Incubation temperature (°C)       |
|-------------------------|-----------------------------------|
| 25 °C                   | 25.07 $\pm$ 0.01<br>(24.9 – 25.2) |
| 27 °C                   | 27.02 $\pm$ 0.01<br>(26.6 – 27.2) |
| 29 °C                   | 29.05 $\pm$ 0.01<br>(28.8 – 29.3) |
| 31 °C                   | 31.08 $\pm$ 0.01<br>(30.7 – 31.4) |
| 33 °C                   | 33.05 $\pm$ 0.01<br>(32.7 – 33.2) |

373

374

375



376 Table 3. Sediment organic matter content and bulk density in sediments colonized by different  
377 macrophyte species and bare sediment in Pollença bay in June 2017. Mean values ( $\pm$ SEM) and  
378 the ranges (minimum – maximum values) are shown.

379

|   | <i>P. oceanica</i><br>sediment     | <i>C. nodosa</i><br>sediment      | <i>C. prolifera</i><br>sediment  | Bare<br>sediment                 |
|---|------------------------------------|-----------------------------------|----------------------------------|----------------------------------|
| Sediment organic<br>matter content (%)                | 12.58 $\pm$ 0.95<br>(8.17 – 20.63) | 6.14 $\pm$ 0.39<br>(4.49 – 10.22) | 0.74 $\pm$ 0.12<br>(0.04 – 1.31) | 0.44 $\pm$ 0.08<br>(0.02 – 1.41) |
| Sediment bulk density<br>(g DW sed cm <sup>-3</sup> ) | 0.54 $\pm$ 0.03<br>(0.29 – 0.69)   | 1.01 $\pm$ 0.03<br>(0.74 – 1.24)  | 0.46 $\pm$ 0.02<br>(0.32 – 0.56) | 0.96 $\pm$ 0.02<br>(0.84 – 1.09) |

380

381



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