



- 1 Modeling the biogeochemical effects of rotation pattern
- ² and field management practices in a multi-crop (cotton,
- ³ wheat, maize) rotation system: a case study in northern
- 4 China
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12 Abstract. The cropping system with rotations between cotton and winter wheat-summer maize (W-M) 13 is widely adopted in northern China. Optimizing the rotation pattern and related field management 14 practices of this system is crucial for reducing its negative impacts on climate and environmental 15 quality. In this study, the approach applied to identify the optimal rotation pattern with the best 16 management practice (BMP) relied on biogeochemical model simulations to determine the negative impact potential (NIP) of individual management options/scenarios and a set of constraints. The 17 optimal rotation pattern and related BMP are referred to as the scenario with the lowest NIP that 18 19 satisfies the given constraints. All the variables of interest were generated by simulation of the DeNitrification-DeComposition 95 version (DNDC95) model. The DNDC95 model validations 20 21 performed previously for a land cultivated with the W-M and presently for an adjacent area cultivated 22 with cotton showed satisfactory performance in simulating the variables of interest with available 23 observations. The simulations of rotation patterns indicated that proper rotation of cotton and the W-M 24 can simultaneously benefit crop yields, soil carbon sequestration and greenhouse gas mitigation. The 25 three-crop rotation pattern in a 6-year cycle could be optimized with 3 consecutive years of cotton and 3 continuous years of W-M cultivation. The experiments with 108 management scenarios showed that 26 27 the BMP for the optimized rotation pattern involved using 15% less nitrogen fertilizer (i.e., 94 and 366 kg N ha⁻¹ yr⁻¹ for cotton and the W-M, respectively) and 20% less irrigation water (i.e., 60-180 and 28 230-410 mm yr⁻¹ for cotton and the W-M, respectively) by sprinkling than the conventional practices, 29





30 fully incorporating crop residues, and adopting the current deep tillage (20–30 cm) for cotton but the

31 reduced tillage (10 cm) for the W-M. However, field confirmation of these BMPs resulting from the

32 model-based virtual experiments is still required.

33 1 Introduction

Cotton is the world's most important fiber crop (www.fao.org/faostat). In China, it was cultivated 34 35 in 2.0-3.9% of the annual crop harvest areas and supported cotton lint production of 5.3-7.6 million 36 metric tons during 2007-2016 (China Statistical Yearbook, 2017). Wheat and maize are two important 37 cereal crops in China and the world (www.fao.org/faostat). Accounting for 39% and 26% of the cereal 38 harvest area in China, wheat and maize produced 129 and 220 million metric tons of grains in 2016, 39 respectively (China Statistical Yearbook, 2017). The high yields of these crops are largely supported by 40 intensive field management practices, such as high addition rates of fertilizer, improved irrigation and 41 application of multiple-cropping systems (e.g., Chen et al., 2014; Galloway et al., 2004; Ju et al., 2009). 42 Northern China contains not only the second most important area of cotton production following 43 that of northwestern China (China Statistical Yearbook, 2017) but also the largest area of the winter 44 wheat-summer maize double-cropping system (i.e., both crops harvested within a year, hereinafter referred to as W-M) in the country (e.g., Cui et al., 2014). Therefore, rotations of cotton and the W-M 45 have been commonly applied in this region (e.g., Liu et al., 2010, 2014). The fiber and double-cereal 46 47 systems are typically alternated every 3-5 years. This typical practice is used not only to avoid the 48 negative effects of monoculture on cotton yields but also to maximize the economic benefits in a 49 balance of labor costs and grain prices (e.g., Liu et al., 2010, 2014; Lv et al., 2014). Relying on 50 intensified field management practices by increased fertilizer inputs, advanced irrigation methods and 51 highly mechanized operations (e.g., frequent applications of pesticides and/or herbicides), cotton 52 maintains stable yields during short-term monoculture in recent decades (Han, 2010). In addition, a 53 recent study indicated that the cotton cropping system in northern China persistently functioned as an 54 intensive carbon or net greenhouse gas (GHG) source compared to the W-M because of strong carbon 55 dioxide (CO₂) emissions during the long non-growing periods (Liu et al., 2019). These previous studies 56 revealed that the change in storage of soil organic carbon (\triangle SOC), net ecosystem GHG emission





(NEGE) and other biogeochemical processes of the multiple-cropping systems in northern China likely are closely related to the rotation pattern of cotton and the W-M. Thus, one can hypothesize that identifying and adopting optimal rotation pattern of cotton and the W-M are beneficial for soil carbon sequestration and mitigation of GHG emissions in the region.

To maintain high lint and grain yields for economic profits and food/feed supplies, the three-crop 61 62 rotation system of cotton and the W-M in northern China are characterized by high additions of 63 synthetic nitrogen fertilizers and irrigation water (e.g., Ju et al., 2009). A number of studies have found 64 that cotton and the W-M receive nitrogen fertilizer(s) at 60-140 and 550-600 kg N ha⁻¹ yr⁻¹, and irrigation water at 140-200 and 90-690 mm yr⁻¹, respectively (Ju et al., 2009; Liu et al., 2014; Wang et 65 66 al., 2008). High nitrogen and water inputs can result in extraordinary changes in biogeochemical 67 processes and thus excessive nitrogen remaining in the soils and high release potentials of nitrogenous 68 pollutants. These effects can further induce a series of environmental problems (e.g., Chen et al., 2014; 69 Collins et al., 2016). In addition to fertilization and irrigation, other field management practices, e.g., 70 tillage and crop residue treatment, can also affect the biogeochemical processes related to emissions of 71 GHGs or nitrogenous pollutants, thereby contributing to negative environmental effects (e.g., Zhang et 72 al., 2017b). Therefore, the evaluation of multiple-cropping systems is shifting from a single-goal method aimed at increasing crop yields to a multi-goal approach aiming to promote sustainability (e.g., 73 74 Cui et al., 2014; Garnett et al., 2013; Zhang et al., 2018). Multi-goal evaluation is especially necessary for identifying the best management practice (BMP) while optimizing the rotation pattern of cotton and 75 the W-M. 76

77 A multi-goal strategy of multiple-cropping systems aims to not only sustain/increase crop yields 78 but also reduce environmental costs. Specifically, it aims to simultaneously sustain/increase crop 79 productivity to ensure food security, increase SOC contents to improve soil fertility, mitigate NEGE to 80 alleviate climate warming, reduce ammonia (NH₃) volatilization and nitric oxide (NO) emission to 81 secure air quality, and abate nitrate (NO3-) leaching to protect water quality. For an ecosystem with 82 annual vegetation, the $-\Delta$ SOC is regarded as the net CO₂ emission assuming negligible losses through 83 non-respiration pathways, and the residual of the annual sum of methane (CH₄) and nitrous oxide (N₂O) 84 emissions minus the \triangle SOC is regarded as the NEGE. Based on the global warming potential (GWP) of





85 each of the three GHGs, the NEGE is expressed as a CO₂ equivalent (CO₂eq) quantity for a given time

86 horizon, such as 100 years. The 100-year GWPs, which are 1 for CO₂, 34 for CH₄ and 298 for N₂O

87 (IPCC, 2013), are usually applied to calculate the CO₂ equivalents.

88 In general, there are two methodological categories for identifying the optimal rotation pattern(s) and/or BMPs (e.g., Cui et al., 2014). One is subjective. It screens and then combines the individual 89 90 single-factorial practices with the maximal benefits (e.g., Farahbakhshazad et al., 2008). The other is 91 objective. It evaluates each decision variable with price-based proxies or other measures and then 92 calculates the negative impact potential (NIP) of individual options while generating a set of constraints. 93 Then, it screens the best option that has the minimal NIP among the options with the given constraints 94 (e.g., Cui et al., 2014; Xu et al., 2017). The subjective methods are generally not applicable to complex 95 management scenarios with opposing effects among concerned decision variables. However, it is 96 difficult to ensure comparability for the price-based proxies or other measures involved in the objective 97 methods among studies with different backgrounds unless protocol-based values are provided for 98 individual decision variables. So far, however, the protocols for the price-based proxies or other 99 measures for the decision variables, including crop yields, NEGE, emissions of NH₃ and NO, and NO₃ 100 leaching, have not been available.

101 To identify an optimal rotation pattern of a cropping system with multiple crops and its BMP using 102 the multi-goal approach, it is essential to quantify the biogeochemical effects of various pattern options and their individual scenarios of field management practices on the multiple decision variables 103 104 mentioned above. As field experiments often focus only on the decision variables of very few rotation 105 patterns and/or management practice treatments during short periods (e.g., Ding et al., 2007; Liu et al., 106 2010, 2015; Wang et al., 2013a, b), these experimental studies alone are insufficient for quantifying the 107 comprehensive biogeochemical effects concerned in a multi-goal strategy. However, process-oriented 108 biogeochemical models have the potential to overcome this limitation of field experimental studies, 109 such as the DNDC (e.g., Chen et al., 2016; Li, 1992, 2000; Zhang et al., 2017a), DAYCENT (e.g., 110 Delgrosso et al., 2005) and LandscapeDNDC (e.g., Haas et al., 2012; Molina-Herrera et al., 2016) 111 models. For instance, the DeNitrification-DeComposition (DNDC) model has bridged primary 112 ecological drivers (e.g., climate, soil properties, vegetation and anthropogenic activities such as field





- 113 management) and biogeochemical processes (e.g., decomposition, nitrification, denitrification and 114 fermentation) through soil regulating variables (e.g., temperature, moisture and pH). The regulating 115 variables resulting from the model calculations are the key regulators of soil carbon and nitrogen 116 transformation. Thus, the model can simultaneously simulate the aforementioned decision variables, 117 i.e., crop yields; Δ SOC; emissions of CH₄, N₂O, NH₃ and NO; and NO₃⁻ leaching (e.g., Cui et al., 2014; 118 Li, 2000, 2007).
- 119 The DNDC model was selected in this work to perform a modeling case study at a field site 120 located in the most southwestern region of Shanxi Province. Model simulations were conducted to 121 investigate the biogeochemical effects of various options of cotton and the W-M rotation patterns and 122 their field management practices. The model version applied in this study, DNDC95, has been validated 123 with comprehensive field measurements of the W-M at the selected field site (Cui et al., 2014), but not 124 with the cotton cropping system. This case study aimed to (i) validate the performance of DNDC95 for 125 the cotton cropping system with comprehensive field observations; (ii) simulate the abovementioned 126 decision variables for individual scenarios of rotation pattern options and field management practices; 127 (iii) investigate the biogeochemical effects of different patterns of the rotation between cotton and the 128 W-M and those of the field management practice options; and, (iv) identify the optimal rotation pattern 129 and its BMP in light of a multi-goal strategy. These efforts were undertaken to test the hypothesis that 130 the sustainable intensification of the three-crop cultivation system aiming to enhance crop yields and 131 simultaneously reduce environmentally negative impacts can be realized through the optimization of 132 the rotation pattern and its field management practices.

133 2 Materials and methods

134 **2.1 Brief introduction to the selected field site**

The field site (34 °55.50′N, 110 °42.59′E, altitude of 348 m) selected for this modeling case study is situated at the Dongcun farm, near Yongji County, Shanxi Province, northern China. It is subjected to a temperate continental monsoon climate, with an annual precipitation of 580 mm and a mean air temperature of 14.4 ℃ in 1986–2010 (Cui et al., 2014). Cotton, winter wheat and summer maize are the major crops grown in this farm and its surrounding regions. Field experiments were carried out in





140	two adjacent lands (each 100 m wide and 200 long) of the farm, in which comprehensive observations
141	were performed for some of the multiple decision variables and some other related factors in
142	2007-2010. Both lands were used for long-term cultivation of the three crops, rotating from cotton to
143	the W-M, or the opposite every 3-8 years until 2004. In the following 6 years (2005-2010), cotton was
144	consecutively cultivated in one of the lands, while the W-M was used in the other. The soil of the land
145	cultivated with cotton during the experimental period was a clay loam, with approximately 38% clay (<
146	0.002 mm), 57% silt (0.002–0.05 mm), 5% sand (0.05–2 mm), 10.0 g $\rm kg^{-1}$ SOC, 1.1 g $\rm kg^{-1}$ total
147	nitrogen and a pH (H ₂ O) of 8.0 at 0–10 cm depth and a bulk density (0–6 cm depth) of 1.20 g \mbox{cm}^{-3}
148	(Liu et al., 2010). The soil of the land cultivated with the W-M during the experimental period was also
149	a clay loam, with approximately 32% clay, 50% silt, 18% sand, 11.3 g kg^{-1} SOC, 1.12 g kg^{-1} total
150	nitrogen, a pH(H ₂ O) of 8.7 and a bulk density of 1.17 g cm ^{-3} (Liu et al., 2011, 2012). A sprinkler
151	system was applied in both lands. For more detailed information on both experimental lands,
152	experimental design, field management practices, crop yields and data from comprehensive
153	observations, please refer to Cui et al. (2014), Liu et al. (2010, 2011, 2014, 2015) and Wang et al.
154	(2013a, b).

155 2.2 Brief introduction to the DNDC95 model

156 The DNDC95 model used in this study is one of the latest DNDC versions 157 (www.dndc.sr.unh.edu/model/GuideDNDC95.pdf). Like the earlier and later versions, the DNDC95 158 model consists of two components with six modules in total. One component includes the soil climate, 159 crop growth and decomposition modules while the other contains the nitrification, denitrification and 160 fermentation modules. Driven by given primary ecological factors, the former component simulates the 161 field states of a soil-plant system, such as soil chemical and physical status, vegetation growth and 162 organic matter decomposition. Driven by the soil regulating variables yielded and provided by the former component, the latter component simulates the core biogeochemical processes of carbon and 163 164 nitrogen transformations and physical processes of liquid and gas transportations and thus the annual 165 dynamics of net ecosystem exchanges of CO2 (NEE); emissions of CH4, N2O, NH3 and NO; and NO3⁻ 166 leaching and the inter-annual dynamics of SOC and NEGE. These features enable the model





167 simulations to investigate the integrative biogeochemical effects of the changes in rotation patterns of 168 multiple crops and/or other management practices at the site or even regional scales, provided the 169 model is validated with comprehensive field observations. The inputs used to facilitate the model 170 simulation include (i) least meteorological variables of daily precipitation and maximum/minimum 171 temperature; (ii) least soil (cultivated horizon) properties of the clay fraction, bulk density, SOC content 172 and pH; (iii) least crop parameters of yield potential, thermal degree days (TDD) for maturity, and the 173 mass fractions and carbon-to-nitrogen (C/N) ratios of grain, root and leaf plus stem; (iv) management 174 practice variables of sowing and harvest (dates), fraction of incorporated/retained residue at harvest, 175 tillage (date and depth), irrigation (date, method and water amount), and fertilization (date, type, 176 method, nitrogen amount and C/N ratio of organic manure); and (v) other variables (annual means of 177 NH_3 concentration in the atmosphere and ammonium plus NO_3^- concentration in rain water). For more 178 details of the model, please see Li et al. (1992) and Li (2000, 2007, 2016).

179 2.3 Model validation for the cotton cropping system

180 Various DNDC versions have been widely applied in cropping systems across different regions of 181 the world to assess the changes in a single decision variable of SOC variation, N_2O emission, NO release, or NO₃⁻ leaching (e.g., Chen et al., 2016; Giltrap et al., 2010; Li et al., 2017). So far, however, 182 183 validation of any version with comprehensive observations of the multiple decision variables and their 184 regulating factors has been scarce. Cui et al. (2014) validated the DNDC95 model using very intensive observations for most of the multiple decision variables and their related factors in the experimental 185 186 land cultivated with the W-M (Liu et al., 2011, 2014, 2015; Wang et al., 2013a, b). Based on the work 187 of Cui and her colleagues, the authors were able to further validate the performance of this model 188 version in simulating the similar regulating factors and decision variables of the experimental land 189 cultivated with cotton.

To validate the model for the cotton cultivation system, the observational data of the input items and those of some of the multiple decision variables and their related factors were collected from Liu et al. (2010, 2014) and Wang et al. (2013a, b). The daily meteorological data of 2004–2010 were directly obtained from Cui et al. (2014). Measured data (described in Sect. 2.1) were directly used for the least





194	required soil properties. The field capacity and wilting point in water-filled pore space (WFPS) used as
195	model inputs, being 0.65 and 0.2, respectively, were cited from Cui et al. (2014), who calibrated both
196	values using the observations in the adjacent land cultivated with the W-M. The crop parameters for
197	cotton were directly determined by the field measurements. The potential grain (seeds plus lint) yield
198	was 1900 kg C ha ^{-1} , which was given as 1.2 times the mean of the measured values. The mass fractions
199	and C/N ratios of grain, root and leaf plus stem were 0.41 and 25, 0.16 and 40, and 0.43 and 40,
200	respectively. The TDD from seeding to maturity was 3600 $$ C, which was given as the sum of those
201	daily averages of air temperature greater than 0 $$ °C during the period from sowing to final grain harvest.
202	The detailed management practices, including dates of operations; tillage depth; fertilizer type, amount
203	and depth; and irrigation water amount and method (Table S1), were obtained from Li et al. (2009) and
204	Liu et al. (2014). Compared with the conventional fertilizer application rate of 110–140 kg N $ha^{-1}~yr^{-1}$
205	for the cotton cropping system, the fertilizer doses of 2007 and 2008 were reduced to $66-75 \text{ kg N} \text{ ha}^{-1}$
206	yr^{-1} by local farmers to avoid the overgrowth of leaves instead of seeds or lint. The measured data used
207	to validate the model outputs were available for soil (5 cm depth) temperature and topsoil (0–6 cm)
208	moisture as major regulating factors and for those used as decision variables including N_2O and NO
209	emissions in 2007–2009 (Liu et al., 2010, 2014), CH_4 uptake fluxes during the period from March to
210	November 2010 (unpublished data of the authors), grain yields, and NEE during the period from
211	November 2008 to November 2009 (Wang et al., 2013a). For the W-M, the crop parameters and other
212	inputs used by Cui et al. (2014) were directly adopted in this study. As the validation of the model
213	outputs for the W-M was reported by Cui et al. (2014), it was not repeated in this paper.

214 2.4 Scenario settings of rotation patterns and management practices

To investigate the biogeochemical effects of rotation pattern and management practices with the goal of identifying the optimal pattern and its BMP, two levels of scenarios were explored.

The level-I scenarios considered rotation patterns. They were set for a 6-year cycle based on surveys of local farmers (Liu et al., 2010, 2011, 2014). These pattern scenarios included all the rotation combinations between cotton and the W-M as well as the monoculture of the latter cropping system. Thus, there were six level-I scenarios in total, hereinafter referred to as R_0 , R_1 ,..., R_5 , from which the





221 optimal pattern was selected. Among these scenarios, R₀ denotes the 6-year monoculture of the W-M; 222 R1 represents the 1-year cotton rotated with the 5-year W-M; and so on. Some years ago, the local 223 farmers typically did not cultivate cotton monocultures for longer than five years in the region. 224 However, heavy applications of herbicides, pesticides and germicides are often required to stabilize the 225 cotton yields of long-term monocultures. To reduce the negative environmental effects of these 226 chemicals, the longest cotton monoculture was set as 5 years in the level-I scenarios. For the transition 227 from cotton to the W-M in each pattern scenario, winter wheat sowing occurs in late October of the 228 same year, immediately following the last harvest of cotton grain. Prior to this winter wheat sowing, the 229 cotton residues are cut with machinery operations and fully incorporated into the soil by plow tillage 230 (30 cm depth). In the non-transition year, the cutting and incorporation of cotton residues occur in early 231 November (Table S2). For the opposite transition, the cotton grown after the last maize is sown in the 232 middle of April of the following year, with a bare-soil fallow period of nearly 5 months between the 233 growing seasons of the two crops. In screening for the optimal rotation pattern, the management practices of the baseline level-II scenario (Tables S2-3) were adopted to drive the simulations for all 234 235 the level-I scenarios.

236 The level-II scenarios considered the field management practices of fertilization, irrigation, crop residue treatment and tillage. From these scenarios, the BMP was identified for the optimal rotation 237 238 pattern. In this regard, the level-II scenarios were set only for the optimal rotation pattern chosen from 239 the level-I scenarios. In setting these level-II scenarios, five factors were considered, including (i) fertilizer dose, (ii) water amount and (iii) method of irrigation, (iv) incorporated/retained fraction of 240 241 crop residues, and (v) depth of tillage. This resulted in 108 level-II scenarios in total, of which one was 242 used as the baseline for all five factors, eight were used for the single-factorial settings (Table 1), and 243 the remaining ninety-nine were the multi-factorial scenarios that were generated by fully combining the 244 settings of the baseline and single-factor scenarios (details of combinations not shown). In each 245 single-factor scenario, one of the five factors was exclusively modified, i.e., improved from its baseline 246 level.

The values of the five factors were set for the baseline level-II scenario (Tables S2-3) by referring
to the observations for the conventional management practices in both experimental lands (Liu et al.,





249 2010, 2011, 2014; Wang et al., 2013a, b). These baseline management practices were widely adopted in 250 the region (Cui et al., 2014). Specifically, the dates of sowing and final harvest were set as 14 April and 251 23 October, respectively, for cotton; as 24 October and 3 June, respectively, for wheat; and as 6 June and 14 October, respectively, for maize. Doses of nitrogen fertilizers were set as 110 and 430 kg N ha⁻¹ 252 yr⁻¹ for cotton and the W-M, respectively. The fertilizer types, application methods and timings of the 253 254 conventional management practices (Liu et al., 2010, 2011, Wang et al., 2013a) were adopted for the 255 baseline scenario (Tables S1-2). Over the last few decades, the fields in the region have been mostly 256 flood-irrigated (Liu et al., 2010). Therefore, flood irrigation was chosen for the baseline. The baseline 257 timings and water amounts were set by referring to the 10- to 30-d cumulative precipitation prior to 258 individual irrigations and the recorded timings and water amounts of conventional management 259 practices in both lands during the experimental periods. Thus, the irrigation frequencies and annual 260 cumulative water amounts during the three consecutive 6-year rotation cycles for the baseline scenario (Table S3) vary from 1 to 3 times and 75 to 230 mm yr^{-1} for cotton and 4 to 6 times and 290 to 510 mm 261 yr^{-1} for the W-M, respectively. As the incorporation of crop residues has been widely adopted in 262 263 northern China, full incorporation following the final harvest of each crop was set as the baseline. Two 264 tillage depths were set in the baseline level-II scenario based on the conventional practices of the region. One is approximately 20 cm, which occurs immediately prior to the sowing of each crop. The other is 265 30 cm. This depth is used following the cutting of cotton stems immediately or shortly after the last 266 267 grain harvest of this crop. Table S2 lists in detail the baseline management practices excluding those of 268 irrigation.

269 The five non-baseline level-II scenarios include: (i) N94/366 and N77/301, which use 15% and 30% 270 lower fertilizer nitrogen doses, respectively, than the single-factor baseline scenario (i.e., N110/430); (ii) 271 I08 and I06, which reduce the water amounts of each irrigation event by 20% and 40%, respectively, 272 relative to the single-factor baseline scenario (i.e., 110); (iii) IS, which adopts sprinkling instead of 273 flood irrigation as the baseline method (i.e., IF); (iv) RIO, with no incorporation/retention of 274 aboveground residues instead of the full incorporation/retention as in the baseline (i.e., RI1); and, (v) 275 T10 and T0, which apply reduced tillage (to 10 cm depth) and no-tillage practices, respectively, for the 276 W-M, instead of the conventional tillage depth of 20 cm used as the baseline (i.e., T20).





277 Table 1 summarizes the different settings of the five management factors among the basic level-II

278 scenarios.

279 2.5 Operation of model simulation

280 For all the scenarios of both levels, the average outputs of individual decision variables for 18-year simulations, including three consecutive 6-year rotation cycles of each level-I scenario, were 281 used to assess the biogeochemical effects of rotation patterns and management practices. The 282 283 simulations were driven by the meteorological data observed at the Yuncheng station (approximately 60 km east to the experimental site) in 1996-2013, which were provided by the China Meteorological 284 285 Data Service Center (http://data.cma.cn/data/cdcindex/cid/6d1b5efbdcbf9a58.html). To stabilize the 286 carbon and nitrogen dynamics and reduce the residual effects of initial conditions (Zhang et al., 2015), especially to establish the carbon and nitrogen equilibriums among different pools, a spin-up of at least 287 288 10 years is usually required (Palosuo et al., 2012). For this reason, the model applied a spin-up of 12 289 years (i.e., a period of two 6-year rotation cycles) before performing the 18-year simulations for each 290 scenario. The spin-up was driven by the same rotation pattern and field management practices as each 291 scenario. The average yields were the averages of cotton, wheat and maize yields of all three 6-year 292 rotation cycles, while the average values of other decision variables explicitly or implicitly involved in 293 Eq. (1) were the averages of 18-year simulations for each scenario.

294 **2.6 Method for identifying the best management practices**

295 An objective method jointly relying on three constraints and NIPs was adopted in this study to 296 identify the optimal rotation pattern and BMP. These constraints included (i) stable or increased crop 297 yields, (ii) stable or increased SOC (identical with the SOC definition of Cui et al. 2014), and (iii) 298 reduced NEGE. In the present study, the NEGE was quantified as a CO₂eq based on the 100-year GWPs provided by the IPCC (2013). The NIP was used to evaluate the potential for a climatically and 299 environmentally integrative impact exerted by a rotation pattern with a given level-II scenario. The NIP 300 was measured in this study by a price-based proxy quantity in USD ha⁻¹ yr⁻¹, even though other 301 302 alternative units were used in other similar studies. The NIP was determined by the quantities of





303	idividual decision variables and their coefficients as mass-scaled price-based proxies (Eq. (1)).	

 $NIP = k_1 NEGE + k_2 NH_3 + k_3 NO + k_4 N_2 O_{ODM} + k_5 NL$ $\tag{1}$

In Eq. (1), NEGE, NH₃, NO, N₂O_{ODM}, and NL represent the multi-goal decision variables of net ecosystem GHG emission (Mg CO₂eq ha⁻¹ yr⁻¹), NH₃ volatilization, NO emission, release of N₂O as ozone layer depletion matter and hydrological nitrogen loss (mainly by NO₃⁻ leaching), respectively (kg N ha⁻¹ yr⁻¹ for all the variables of nitrogen compounds). The coefficients k_1 , k_2 , k_3 , k_4 and k_5 are mass-scaled price-based proxies of NEGE, NH₃, NO, N₂O_{ODM}, and NL, respectively. Their values presented in Cui et al. (2014) were directly used in this study, which were 7.00 USD Mg⁻¹ CO₂eq, and 5.02, 25.78, 1.33 and 1.92 USD kg⁻¹ N for k_1 , k_2 , k_3 , k_4 and k_5 , respectively.

In Eq. (1), a lower NIP indicates a better set of management practices that can exert smaller negative impacts on the climate and environment. Accordingly, the optimal rotation pattern with the BMP was identified as the scenario with the lowest NIP among those scenarios satisfying all three constraints.

315 2.7 Statistics and analysis

316 The statistical criteria of the (i) index of agreement (IA) (Eq. (2)), (ii) Nash-Sutcliffe efficiency 317 (NSI) (Eq. (3)) evaluating modeling efficiency (e.g., Moriasi et al., 2007; Nash and Sutcliffe, 1970), (iii) 318 determination coefficient (R^2) , slope and significance level of a zero-intercept univariate linear 319 regression (ZIR) of observations (o) against simulations (s) and (iv) model relative bias (MRB) (e.g., 320 Congreves et al., 2016; Willmott and Matsuura, 2005) were simultaneously used to evaluate model 321 validity. In Eqs. 2–4, k and n (k = 1, 2, ..., n) denote the kth pair and the total pair number of the values, 322 respectively, and \bar{o} represents the mean of observations. The IA index falls between 0 and 1, with a 323 value closer to 1 indicating better simulation, and vice versa. An NSI value between 0 and 1 shows acceptable model performance and otherwise, worse. The R^2 value of a ZIR, which has the same 324 325 meaning as that of a univariate linear regression including an intercept, is also calculated by Eq. (4) 326 (Jiang, 2010), wherein \hat{o} denotes the s-based prediction using the ZIR. The F-test was used to 327 determine the significance level of a ZIR. Better model performance is indicated by a significant ZIR at 328 P < 0.05 and a slope and an R^2 value being closer to 1, and vice versa. A slope less than 1 indicates an





329 overestimation by simulations and the opposite underestimation. The MRB, which is expressed as a 330 percentage, is defined by Eq. (5), wherein \bar{s} denotes the average of simulations. In this study, an 331 absolute MRB greater than two times the coefficient of variation (CV) (or a simulation falling outside 332 of the range of mean \pm two standard deviations) of spatially replicated observations is used to represent 333 a significant discrepancy based on the 95% confidence interval (CI); the opposite indicates a good 334 agreement or an insignificant discrepancy based on the CI. A positive MRB value indicates 335 overestimation by modeling whereas a negative value indicates underestimation.

IA = 1 -
$$\frac{\sum_{k=1}^{n} (s_k - o_k)^2}{\sum_{k=1}^{n} (|s_k - \bar{o}| + |o_k - \bar{o}|)^2}$$
(2)

$$NSI = 1 - \frac{\sum_{k=1}^{n} (o_k - s_k)^2}{(3)}$$

$$R^{2} = 1 - \frac{\sum_{k=1}^{n} (o_{k} - \dot{o}_{k})^{2}}{\sum_{k=1}^{n} (o_{k} - \dot{o})^{2}}$$
(4)

$$MRB = 100 \frac{\overline{s} \cdot \overline{o}}{\overline{o}}$$
(5)

In this study, ZIR analysis, variance analysis, and graphical comparison were performed with
 SPSS Statistics Client 19.0 (SPSS Inc., Chicago, USA) and Origin 8.0 (OriginLab, Northampton, MA,
 USA) software.

339 3 Results

340 **3.1 Model validation of the cotton cropping system**

 $\sum_{k=1}^{n} (o_k - o)^2$

The seasonal dynamics and magnitudes of soil (5 cm) temperature and topsoil (0–6 cm) moisture were predicted well by the model simulations (Figs. 1a–b). The sound model performance was indicated by the IA, NSI, and ZIR slope and R^2 of 0.99 and 0.81, 0.95 and 0.24, 1.05 and 0.93, and 0.95 (n = 677, P < 0.001) and 0.29 (n = 432, P < 0.001) for the temperature and moisture, respectively.

345 For the comparison of cotton yields (seeds plus lint) for two consecutive years, the simulations





- 346 and observations also agreed well, with an absolute MRB of 7%, which was less than two times the
- spatial CVs (37–44%) for the measurements (adapted from Liu et al., 2014).
- The simulated seasonal patterns and peak emissions of N₂O and NO generally matched the observations (Figs. 1c–d). The simulations showed a relatively low modeling efficiency for daily fluxes, as indicated by the IA, NSI, and ZIR slope and R^2 of 0.77, < 0, 0.48 and 0.40 (n = 592, P < 0.001), respectively, for N₂O and of 0.62, < 0, 0.37 and 0.09 (n = 614, P < 0.001), respectively, for NO. Nevertheless, the simulated annual N₂O and NO emissions in the two consecutive experimental years were comparable with the observations, with absolute MRBs of less than 3% and 6%, respectively,
- both of which were less than two times the spatial CVs (24–50%) for the measurements (adapted from
 Liu et al., 2014).
- The simulated NEE fluxes suggested that the model captured the seasonal fluctuations, which were negative in the cotton growing season, but positive or neutral during the remaining periods (Fig. 1e). The IA, NSI, and ZIR slope and R^2 were 0.72, 0.29, 0.79 and 0.31 (n = 365, P < 0.001), respectively, for the daily NEE simulation. For the annual cumulative NEE, the model simulation showed an absolute MRB of 2%, which was less than the reported uncertainty (25%) of observations (Wang et al., 2013a)
- For CH₄ uptake, the observations and simulations showed similar seasonal variations (Fig. 1f), with the IA, NSI and ZIR slope and R^2 of 0.71, < 0, 0.76 and 0.16 (n = 69, P < 0.01), respectively. The model simulation yielded an absolute MRB of 8% for the cumulative CH₄ uptake, which was smaller than two times the spatial CVs (12%) of observations (adapted from the unpublished data).
- These results suggest that the DNDC95 model could be applicable for investigating the biogeochemical effects of different rotation patterns between cotton and the W-M and those of different management practices under a given rotation pattern.

369 3.2 Biogeochemical effects of different cotton and wheat-maize rotation patterns

Figure 2 illustrates the dynamics of crop yields and each decision variable resulting from the consecutive simulations over 18 years for all rotation pattern options subject to the field management practices of the baseline scenario. Figure 3 shows the relationship between the annual average of each





- 373 decision variable and the number of consecutive years of cotton monoculture within the rotation pattern
- 374 options.
- 375 The average grain yields of cotton, wheat and maize were not significantly different among various rotation pattern options, with averages of 3.0, 5.0 and 7.0 kg dry matter ha⁻¹ for cotton, wheat 376 and maize, respectively (Figs. 2a-c). 377
- 378 For the dynamic changes in annual SOC stocks, the values were positive for the W-M, but 379 negative for the cotton, except for the first year after the transition to this fiber crop. As Fig. 2d indicates, the simulated SOC contents over the 18-year period increased for R0, R1, R2 and R3 but 380 381 decreased for R₄ and R₅. The annual average $-\Delta$ SOC significantly increased (P < 0.001) with an 382 increase in the consecutive years of cotton monoculture from 0 to 5 within the 6-year rotation cycle 383 (Fig. 3a). Rotation pattern options with baseline management showed small variations in CH_4 uptake 384 (Fig. 2e), with the annual uptake ranging from 1.80 to 1.89 kg C ha⁻¹. However, the annual averages of 385 CH_4 uptake significantly increased (P < 0.001) with an increase in the consecutive years of cotton monoculture from 0 to 5 within the 6-year rotation cycle (Fig. 3b). For N₂O, the annual emission of the 386 387 rotation options with the management baselines showed large inter-annual variations (Fig. 2f), with a 388 CV of 21-45%. Meanwhile, the annual average emissions of this gas significantly decreased from 4.3 389 to 2.6 kg N ha⁻¹ (Fig. 3c) with an increase in the consecutive years of cotton monoculture (P < 0.001). 390 As a result, increasing the consecutive years of cotton monoculture from 0 to 5 within a 6-year rotation significantly promoted (P < 0.01) the NEGE (Figs. 2g and 3d). 391
- As for the gaseous air pollutants of NH3 and NO, the simulated annual emissions ranged from 17 392 393 to 104 and 0.5 to 3.5 kg N ha⁻¹, respectively (Figs. 2h-i). Figures 3e and f show that the averages of 394 annual emissions of both gases were significantly reduced by increasing the consecutive years of cotton 395 monoculture from 0 to 5 within the 6-year rotation cycle (P < 0.001).

396 The annual NO_3^- leaching of different rotation pattern options displayed significant inter-annual 397 variations (Fig. 2j), with CVs of 40-68%. Thus, the annual averages of NO3⁻ leaching insignificantly 398 changed in response to the consecutive years of cotton monoculture within the 6-year rotation cycle 399 (Fig. 3g).

400

The NIP significantly varied among the various rotation pattern options (P < 0.001), declining





401 from 609 to 320 USD ha^{-1} yr⁻¹ with an increase in the consecutive years of cotton monoculture within 402 the 6-year rotation cycle (Fig. 3h). For the three constraints, the one for crop yields was not used in 403 screening the optimal rotation pattern as the simulations showed no obvious difference among the 404 options. Both R_0 and R_5 represent the typical rotation patterns in the region. The simulations for the 405 former indicate the greatest increase in SOC and lowest NEGE but the highest NIP, and those for the 406 latter show the greatest SOC loss and largest NEGE but the lowest NIP (Figs. 3a and d). These patterns 407 indicate that neither typical rotation pattern is optimal. Considering the trade-offs between NIP and 408 both \triangle SOC and NEGE, R₃ was chosen as the optimal rotation pattern that can balance economic 409 benefit (crop yields), soil fertility and environmental safety well.

410 **3.3 Biogeochemical effects of single-factor management alternatives**

411 Under the screened optimal rotation pattern of R₃, the biogeochemical effects of each single field 412 management practice were evaluated. 413 N94/366 had no obvious effects on annual average yields compared to N110/430 based on the 414 single-factor baseline of nitrogen doses. However, N77/301, with a further 15% reduction in doses, 415 reduced the annual averages of crop yields, especially for the W-M, by more than 10% (P < 0.05; Fig. 4a). The annual averages of △SOC and N₂O emission tended to decline by 2% and 3%, respectively, 416 417 for N94/366 and by 10% and 10%, respectively, for N77/301. As a result, the NEGE for nitrogen 418 reduction scenarios was reduced by 13-15% (Fig. 3b). For NH₃, NO emissions and NO₃⁻ leaching, the 419 annual averages were reduced by 19%, 10% and 29%, respectively, for N94/366 and 37%, 20% and 420 49%, respectively, for N77/301 (Fig. 4c). Compared with N110/430, the lower emissions of gaseous or 421 dissolved nitrogen pollutants from these two scenarios with reduced nitrogen doses resulted in a 15% 422 and 33% (P < 0.05) lower NIP, respectively (Fig. 4a).

In comparison with I10 as the baseline, I08 slightly decreased the annual average yields by less than 4% (Fig. 4a), but the annual average yield of maize was reduced by more than 10% for I06. The effects of the reduced irrigation water amount on annual \triangle SOC, CH₄ uptake, and N₂O, NH₃ and NO emissions were not consistent among simulations during the 18-year period. Thus, the annual averages of these decision variables for both I08 and I06 were not significantly different from those of I10. The





428 annual averages of NO3⁻ leaching also showed insignificant response to the reduced irrigation water 429 amount. Therefore, saving water did not lead to an obviously modified NIP compared to the 430 conventional amount of irrigation water used as the single-factor baseline (Figs. 4b and c). 431 The IS tended to increase crop yields very slightly (by 0.4–2%) while showing negligible effects on the decision variables and thus the NIP compared to IF as the baseline irrigation method. 432 433 In comparison with RI1 as the baseline, the annual averages of crop yields of the RI0 tended to 434 decrease slightly (< 3%). However, the SOC contents of RIO were generally lower than those of the 435 baseline and showed a \triangle SOC of -0.18 Mg C ha⁻¹ yr⁻¹ on average (Figs. 4a-b). Meanwhile, the annual N₂O and NO emissions of RI0 were consistently lower than those of RI1, with annual averages of 2.2 436 versus 3.6 kg N ha⁻¹ yr⁻¹ for N₂O (P < 0.05) and 1.4 versus 1.9 kg N ha⁻¹ yr⁻¹ for NO (Figs. 4b-c). The 437 large decreases in both △SOC and N₂O emissions resulted in a 56% higher annual average NEGE than 438 439 that of the baseline. In addition, RI0 decreased the annual average of NO_3^- leaching by approximately 440 31% compared to RI1. The trade-off between NEGE and NO_3^- leaching resulted in approximately comparable NIP values between RI0 and RI1 (Figs. 4a-c). 441 442 For both T10 and T0, the annual averages of crop yields were not significant changed from those 443 of T20 as the baseline with a conventional tillage depth of 20-30 cm (Fig. 4a). However, both T10 and 444 T0 enlarged the annual averages of \triangle SOC and N₂O emission to different extents compared to the baseline, with 0.21–0.18 versus 0.12 Mg C ha⁻¹ yr⁻¹ for the former and 3.9–4.2 versus 3.6 kg N ha⁻¹ 445 yr⁻¹ for the latter (Fig. 4b). Further, T10 and T0 led to a lower annual CH₄ uptake than T20, with the 446 effect of the no-tillage scenario being very significantly in particular (P < 0.01). As the increases in 447 448 \triangle SOC could offset the decrease in CH₄ uptake and the increases in N₂O emission, the annual average of NEGE for T10 decreased by 19%. Moreover, T10 and T0 showed no obvious effects on the 449 450 emissions of NH₃ and NO and leaching of NO₃⁻. Nevertheless, no statistically significant difference in

451 NIP was detected between T10 or T0 and the baseline.

According to the results above on the single-factor scenarios under the optimal rotation pattern, the following insights were obtained: reducing nitrogen fertilizer doses by 15% does not significantly affect the crop yields but obviously reduces the decision variables, thus leading to reduced NIPs; full incorporation/retention of crop residues decreases NEGE but increases NO₃⁻ leaching, thus showing no





456 obvious effect on the NIP due to the trade-off between these two opposing effects; shifting 457 conventional tillage to the reduced tillage or no-tillage practices benefits \triangle SOC while increasing N₂O 458 emission.

459 3.4 Identification of best management practices

To achieve the BMP under the optimal rotation pattern (i.e., R₃), all multifactorial scenarios were
simulated. The simulated results of all 108 scenarios were compared to evaluate the integrative
biogeochemical effects based on the three constraints and NIPs.

463 To screen the BMP, the annual averages for the 18-year simulations of the examined variables, 464 including crop yield, NEGE, NH₃ volatilization, N₂O and NO emissions, and NO₃⁻ leaching, for each 465 single- and multifactorial scenario of various management practices were analyzed. In consideration of 466 the two constraints that crop yields and SOC stocks are no less than those of the baseline, only five out 467 of the 107 scenarios, excluding that of the baseline, met these requirements. Further taking into account the third constraint that the NEGE should be reduced compared with that of the baseline, only four of 468 469 the 107 scenarios were screened. According to the final objective quantitative criteria, one of these four 470 scenarios, i.e., N94/366_I08_IS_RI1_T10, which showed the minimal NIP (23% lower than that of the 471 baseline), was identified as the BMP under R₃.

472 The identified BMP for the investigated crop rotation system showed the following management 473 features: (i) each of the cotton and W-M is cultivated for three consecutive years within a 6-year rotation cycle; (ii) the present crops and the current schedules of planting, harvesting, fertilization (date, 474 475 depth, and splits) and irrigation (date and times) are adopted; (iii) urea is applied at a 15% lower rate, namely, 94 and 366 kg N ha⁻¹ yr⁻¹ for cotton and the W-M, respectively; (iv) 20% less water than the 476 477 conventional level is irrigated by sprinkling; (v) crop residues are fully incorporated at harvest; and (vi) 478 conventional tillage (20-30 cm depth) for cotton but reduced tillage (10 cm depth) for the W-M are 479 applied.

480 In comparison to the baseline, i.e., the currently applied field management practices, the identified 481 BMP could produce stable crop yields (increase by 1% on average) and enlarge \triangle SOC (by 75% on 482 average) while decreasing NEGE (by -26% on average), NH₃ volatilization (by -19% on average), NO





483 emission (by -2% on average) and NO₃⁻ leaching (by -43% on average) (Table 2).

484 4 Discussion

485 4.1 Model performance

486 The DNDC model has been widely applied in agricultural systems across the world. The model 487 version used in this study has been comprehensively validated with measurements of soil 488 environmental variables, crop yields, and fluxes of NEE, NH₃, CH₄, N₂O and NO under various field 489 management treatments for the double-cropping system of W-M (Cui et al., 2014). Cui and her 490 colleagues reported satisfying model performance in simulating the variables of interest of the W-M. 491 The validation in the present study also showed satisfying performance of the DNDC95 model in 492 simulating the crop yields, NEE, CH₄ uptake, emissions of N₂O and NO and related soil factors of the 493 land cultivated with cotton. Both validations, namely those by Cui et al. (2014) and those by this study, 494 suggest that the DNDC95 model can be applied to quantify the decision variables and related factors of 495 the cotton and W-M rotation system under different management practices.

496 The model effectively captured the dynamics of topsoil temperature and moisture as responses to 497 seasonal variations in precipitation or irrigation events. Furthermore, the well-simulated crop yields 498 combined with the soil environmental factors provided a solid basis for further simulating the decision 499 variables to quantify the NIP under any condition. The crop yields and soil environmental factors are the indicators of essential processes of plant nitrogen uptake and the key factors regulating the 500 501 biogeochemical processes of a cropping system (Chirinda et al., 2010; Kröbel et al., 2010). For the 502 simulations of N₂O and NO emissions, discrepancies in daily emissions generally occur in DNDC 503 models and other current biogeochemical models due to the complex interactions among soil 504 environmental factors and carbon and nitrogen processes (Zhang et al., 2015). For the case of cotton in 505 this study, underestimations of daily NO fluxes mostly occurred during the spring period of both 506 experimental years. This under-prediction may have been caused by the overestimated topsoil moisture, 507 which facilitated the reduction of NO to N₂O via denitrification (del Prado et al., 2006). These negative 508 biases in simulating daily NO fluxes resulted in worse statistical evaluations. The temporal variations in 509 both observed and simulated NEE indicated that the cotton field frequently assimilated atmospheric





510 CO_2 during the vigorous growth stages but emitted CO_2 to the atmosphere during the other periods. 511 However, the abrupt NEE increases during the growing season on cloudy or rainy days were still not 512 well simulated due to the calculation of plant growth relying on cumulative daily temperature over 0 $\,$ $\,$ $\,$ $\,$ $\,$ 513 and the parameter TDD (e.g., Cui et al., 2014). For the simulations of other nitrogen losses from the 514 cotton field, the NH₃ volatilization and NO₃⁻ leaching accounted for 18-24% and 6-13%, respectively, 515 of the applied fertilizer nitrogen during the 2-year period used for model validation. These loss rates 516 were comparable with the field measurements of 10-23% for NH₃ volatilization (Li et al., 2016) and 517 16-17% for NO₃⁻ leaching (Liu et al., 2014). 518 The model validation suggests that the scientific processes determining the NO fluxes during the 519 spring period before nitrogen fertilization in cotton fields and those determining the photosynthesis and

520 ecosystem respiration in CO_2 fluxes (both as the NEE components) on cloudy and rainy days still 521 require improvement in the future study. As there was no observation of NH_3 volatilization and NO_3^- 522 leaching from the experimental cotton field, this validation study did not include both decision 523 variables. Thus, future study still requires further validation of model performance using 524 comprehensive observations covering both decision variables as well as others.

525 4.2 Biogeochemical effects of rotation pattern and management practices

526 The monoculture effect is neglected in the DNDC model. However, this neglect did not affect the 527 simulated crop yields and thus the decision variables as results of biogeochemical interactions. For the W-M, the two crops are cultivated in rotation instead of the monoculture of either of them. Therefore, 528 529 the W-M has been cultivated over the long term in many lands in northern China. The monoculture 530 effect may manifest as changes in cotton yields if this crop is continuously cultivated for a long time. In 531 practice, a period of 5 consecutive years is usually applied for the longest cotton monoculture to 532 stabilize its yields while avoiding heavy uses of herbicides, pesticides and germicides, which benefit the environment. Meanwhile, balanced elemental nutrients have been applied in cotton cultivation, in 533 534 which the crop residues remain in the field after the final seed harvest to avoid nutrient deficit. In this way, the negative effect of monoculture on cotton yields can be offset (Han, 2010). This fact is 535 536 reflected by the DNDC model that assumes balanced nutrient supplies for any crops (e.g., Li, 2017).





537	The definition of modeled SOC of Cui et al. (2014) was applied in this study, which represents the
538	total organic pool including the carbon in microbes, active humus (namely, humads) and resistant (or
539	inactive) humus and excludes visible biomass residues in any form. In fact, this definition is identical to
540	that of the SOC measured by the classic method in soil science. The simulated positive annual changes
541	in SOC for the W-M were mainly attributed to the incorporation of full aboveground residues (at rates
542	of 5.1–7.0 Mg C ha^{-1} yr ⁻¹), which were favorable for carbon sequestration (Han et al., 2016). However,
543	the negative annual changes in SOC for the cotton cropping system resulted from the shorter growing
544	season relative to that of the W-M. The longer fallow season stimulated a notable amount of CO_2
545	emission from the soil, thereby reducing soil carbon sequestration (Liu et al., 2019). As a remarkable
546	carbon sink, the W-M with incorporation of full crop residues could even totally compensate for the
547	SOC lost during the first cotton-planting year following cultivation of the W-M for some years. Thus,
548	the annual change in SOC was generally positive in the first cotton-cultivation year of the investigated
549	rotation patterns. Therefore, an appropriate rotation pattern of cotton and the W-M could balance the
550	ecosystem organic carbon budget of a 6-year cycle as well as gain an economic benefit from growing
551	the non-cereal cash crop. The rotation patterns of R_0 and R_1 acted as net GHG sinks since the increased
552	SOC exceed the increased $\mathrm{N}_2\mathrm{O}$ emission related to the W-M cultivation. In comparison, the other
553	rotation pattern options all functioned as net GHG sources. The higher application rate of fertilizer for
554	the W-M than for cotton resulted in more reactive nitrogen remained in the soil (Chen et al., 2014; Ju et
555	al., 2009), thereby stimulating more emissions of the nitrogenous air pollutants and $N_2 O$ under the
556	scenarios with fewer cotton planting years. Therefore, the $R_{\rm 3}$ should be proposed as the appropriate
557	rotation pattern of cotton and the W-M, which, among the six examined options, can allow the
558	sustainable intensification with maximal yield benefits and minimal negative impacts on the
559	environment.
560	Northam China, as the most important egricultural region, experienced an increase in error yields

560 Northern China, as the most important agricultural region, experienced an increase in crop yields 561 by a factor of 2.8 in 1980–2008, during which the application of mineral fertilizers increased by a 562 factor of 5.1. The rapid increase in fertilizer inputs has resulted in excessive nitrogen remaining in soil, 563 posing potential risks for the environment (Chen et al., 2011; Zhang et al., 2017b). To solve this 564 problem, the reduction of fertilizer application was proposed in several previous studies (e.g., Chen et





al., 2011, 2014; Liu et al., 2012). The results of scenario analysis in this study indicated that, for the optimal rotation pattern, further reducing the farmer-optimized nitrogen doses by 15% (to 94 and 366 kg N ha⁻¹ yr⁻¹ for cotton and the W-M, respectively) could sustain the crop yields while greatly decreasing the emissions of nitrogenous gases.

569 In addition to fertilization, over-irrigation in northern China has also been ubiquitous for a long 570 time, which threatens the water security of this region due to the sharply declining groundwater table 571 and water pollution (Gao et al., 2015; Ju et al., 2009). For this reason, only management options with a 572 reduced irrigation water amount should be considered under the severe shortage of water resources, 573 even though the yields for these scenarios slightly decreased. In addition, adopting sprinkling instead of 574 flood irrigation for an equal water amount showed positive effects on crop yields, indicating irrigation 575 efficiency improvement (Zhang et al., 2017b). This result means that rather than blindly using larger 576 nitrogen doses, increasing water-use efficiency through the application of alternative irrigation 577 techniques could be a pathway to sustain crop yields with reduced nitrogen addition.

578 Through a ban on burning crop residues in fields, returning the residues to the field is strongly 579 suggested by the government. The simulated results for incorporation of full residues indicated that this 580 practice can simultaneously increase crop yields, soil carbon storage and emissions of N_2O and NO581 while decreasing the NEGE. These simulations were consistent with the results of field observations 582 (Liu et al., 2011). The consistent results indicate the beneficial effect of residue incorporation on GHG 583 balance.

584 Reduced tillage and no-tillage practices have been promoted in China in recent decades. To 585 facilitate decomposition of woody cotton residues and avoid outbreaks of diseases and pests due to the 586 continuous implementation of the no-tillage practice, adjustment of tillage practices was only applied in 587 the W-M, while deep tillage for cotton was maintained when setting the tillage scenarios. As the 588 no-tillage practice reduces gas permeability, it significantly decreases CH₄ uptake (Zhao et al., 2016). 589 However, as shown by the model simulations, the contribution of this effect on CH4 uptake to the 590 change in NEGE was much smaller compared to that of the no-tillage effect on SOC stock. Similar to 591 previous experimental studies (e.g., Zhao et al., 2016), the scenario analysis relying on model 592 simulations in this study showed that the reduced tillage or no-tillage practice could sustain crop yields





593 and increase carbon sequestration in the soil while reducing NH₃ volatilization and NO₃⁻ leaching. 594 As shown above, the influences of various field management practices on crop yields and the 595 decision variables were different from each other in most cases. This implies that an appropriate 596 combination of the single-factor management practices must exist for the rotation cropping system of 597 cotton and the W-M. The biogeochemical effects of such an appropriate combination should satisfy the 598 three constraints while resulting in the lowest NIP. Direct observations in field experiments, which 599 often have very few treatments, measure very few variables and/or parameters for each treatment and 600 represent very few conditions, are usually far less sufficient for screening such an appropriate 601 combination. However, identifying such an appropriate combination is one of the purposes of 602 developing a biogeochemical model, such as the DNDC model. The biogeochemical model validated 603 with limited observations from field experiments is in principle capable of fulfilling this task.

604 **4.3 Evaluation of the best management practice**

605 The scenario analysis succeeded in screening for the BMP. The BMP could sustain the crop yields 606 of cotton, winter wheat and summer maize of the three-crop rotation system, increase SOC stock, 607 mitigate N_2O emission and NEGE, reduce emissions of NH_3 and NO, and reduce NO_3^- leaching due to the enhanced resource use efficiency in response to the reduced nitrogen-fertilizer dose and irrigation 608 609 water amount. Hence, the BMP could result in a more drastically reduced NIP than the other examined 610 management options. In the face of the scarcity of water resources and the pollution of underground 611 water in northern China, reducing the irrigation water amount is an important pathway for adapting and 612 alleviating these challenges (Gao et al., 2015; Ju et al., 2009; Zhang et al., 2017b). Therefore, in 613 addition to the BMP, another scenario, in which the irrigation water amount is further reduced to 60% 614 of the conventional level of farmer practices, could also be chosen if the loss of maize yield by 9% on 615 average is tolerable. The model simulations for this management practice combination with 25% less 616 irrigation water than the identified BMP showed similar effects on the cotton and wheat yields and all 617 the decision variables compared to the BMP.

618 The uncertainties of the screened BMP mainly stem from three aspects.

619 The first aspect is the possible limitation of the applied model, which cannot simulate the effects





of monoculture on cotton yields. However, some studies have demonstrated that the negative effects of
cotton monoculture can be compensated for by improved field management (Han, 2010; Liu et al.,
2014). This solution has been considered as much as possible in developing the scenario settings to
avoid the unexpected monoculture influences.

624 The second aspect is the insufficient data for model validation. Due to the lack of observations, the 625 model simulations on NH_3 fluxes for cotton cultivation and those on NO_3^- leaching in both cotton and 626 the W-M cropping systems were not validated (Cui et al., 2014; this study). The DNDC model has been 627 established by following the mass conservation law. In other words, it can reflect the mass balance of 628 the nitrogen budget for the simulated soil layer (0-50 cm depth) well. This model principle implies that 629 only one budget item could be allowed to not be validated. The item is usually soil nitrogen loss 630 through the production of dinitrogen gas (N_2) mainly by denitrification, which is very difficult to 631 measure in situ (e.g., Wang et al., 2013). For both cropping systems, the nitrogen lost through this 632 pathway could be almost fully inhibited in the topsoil, wherein the soil moisture contents were often lower than 60% WFPS (Linn and Doran, 1984; Liu et al., 2011, 2014). However, most likely, N2 lost 633 634 could not be neglected in lower soil due to higher groundwater tables of approximately 1 m in depth. 635 The model principle also implies that the uncertainty for the simulation of NO_3^- leaching during cotton cultivation might have been larger than that of the W-M cropping system since the former has one more 636 637 decision variable (i.e., NH₃ volatilization) not directly validated by observations. These situations suggest that the current uncertainties in identifying the BMP may not be reduced unless the simulations 638 of both NH₃ volatilization and NO₃⁻ leaching can be validated in addition to the other validated 639 640 variables.

The last aspect is the method applied for the identification of the BMP. The current method only considers the biogeochemical effects on decision variables and crop yields as one of the three constraints. It excluded other factors, such as those related to the costs of management practices, thereby likely resulting in uncertainty. Although the method applied in this study still has some deficiencies, this case study shows its potential application in more comprehensive situations, which can be easily and automatically implemented as long as the simulations for all the decision variables and crop yields can be validated with comprehensive observations.





648 5 Conclusions

649 An approach was proposed and applied to identify the optimal rotation pattern with the best other 650 field management practice (BMP) based on the negative impact potentials (NIPs) of individual 651 options/scenarios and a set of constraints. The NIP of an option or a scenario was defined as the linear function of five decision variables, including net ecosystem greenhouse gas emission (NEGE), 652 653 ammonia volatilization, nitric oxide release, emission of nitrous oxide in the form of ozone layer 654 depletion matter, and nitrate leaching. This study used three variables, i.e., crop yield, soil organic 655 carbon (SOC) content, and NEGE, to specify the applied constraints that were stable/increased crop 656 yields and SOC contents and reduced NEGE. All the decision variables and those related to constraints 657 were generated through model simulation. The optimal management option was referred to as the 658 scenario with the lowest NIP among those simultaneously satisfying the three constraints.

659 This case study focuses on a multi-crop cultivation system with rotation between cotton and winter wheat-summer maize (W-M) alternations, which has been widely adopted in northern China. As 660 661 previously indicated by the validation studies for a land cultivated with the W-M and presently 662 indicated for an adjacent area cultivated with cotton, version 95 of the DeNitrification-DeComposition (DNDC95) model showed satisfactory performance in simulating the decision and constraint variables 663 664 with available field observations. This indicates a solid basis for using simulations of the model to 665 evaluate the biogeochemical effects of rotation pattern and other field management practices of the 666 multi-cropping system. Relying on the DNDC95 simulations driven by 113 management scenarios, the 667 effect of various field management options on the variables of interest were investigated, and the 668 optimal rotation pattern and its BMP were identified. The simulation results of rotation patterns 669 indicated that proper rotation of cotton and the W-M can simultaneously be beneficial for the economic 670 benefit (crop yields), soil fertility (soil carbon sequestration) and climate (GHG mitigation). Thus, the 671 three-crop rotation pattern of a 6-year cycle could be optimized with 3 consecutive years of cotton with 3 continuous years of W-M. The BMP under this rotation pattern could use 15% less fertilizer and 20% 672 673 less irrigation water by sprinkling compared to the currently adopted conventional practices, fully incorporate crop residues and adopt conventional deep tillage (to 20-30 cm) for cotton but reduced 674 675 tillage (to 10 cm) for the W-M. This study emphasizes the necessity of comprehensive observations that





- 676 fully cover the decision and constraint variables and related soil factors to facilitate effective BMP
- 677 screening relying on virtual experiments using a biogeochemical model such as the DNDC model.

678 Author contribution

- 679 Xunhua Zheng and Chunyan Liu contributed to the further development in data management and
- 680 enhanced scientific efficacy and effectiveness. Wei Zhang performed the simulations and prepared the
- 681 manuscript with contributions from all co-authors. Chunyan Liu, Kai Wang, Rui Wang and Zhisheng
- 462 Yao designed and carried out the experiments. Feng Cui and Siqi Li applied the model in adjacent land
- 683 of winter wheat-summer maize cropping system.

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843 Table 1 Detailed management practices for each level-II scenario.

Scenarios N ^a		RI ^b	Tillage ^c IrrA ^d		IrrM ^e	
Baseline	C: 110; W-M: 430 (N110/430)	Full (RI1)	C: 20/30; W-M: 20 (T20)	C: 1-3 times, 75-230 mm yr ⁻¹ MW: 4-6 times, 290-510 mm yr ⁻¹ (I10)	IF (IF)	
N94/366	C: 94; W-M: 366	s.a.b	s.a.b	s.a.b	s.a.b	
N77/301	C: 77; W-M: 301	s.a.b	s.a.b	s.a.b	s.a.b	
108	s.a.b	s.a.b	s.a.b	20% less	s.a.b	
I06	s.a.b	s.a.b	s.a.b	40% less	s.a.b	
IS	s.a.b	s.a.b	s.a.b	s.a.b	IS	
RI0	s.a.b	Zero	s.a.b	s.a.b	s.a.b	
T10	s.a.b	s.a.b	C: 20/30; W-M: 10	s.a.b	s.a.b	
Т0	s.a.b	s.a.b	C: 20/30; W-M: 0	s.a.b	s.a.b	

^a C, cotton; W-M, wheat-maize; N, nitrogen application rate (kg N ha⁻¹ yr⁻¹);

845 ^b RI, residue incorporation;

846 ^c Tillage, tillage depth (cm);

^d IrrA, irrigation water amount;

848 e IrrM, irrigation method; IF, flood irrigation; IS, sprinkling. s.a.b, same as the baseline. The

849 management practices other than those listed in this table were the same for all the scenarios and

850 described in the text and Tables S2-3. The code in each pair of parentheses denotes the abbreviation of

851 baseline for each management factor.





- 852 Table 2 Simulated decision and constraint variables and negative impact potential (NIP) for the
- 853 management practices under the optimal rotation pattern.

Scenarios ^a	CY	WY	MY	∆SOC	CH_4	N_2O	NEGE	NH_3	NO	NL	NIP
BAS	3.54	4.98	6.91	0.12	-1.86	3.59	1.16	56	1.85	59	453
BMP	3.59	5.06	6.99	0.21	-1.82	3.64	0.86	45	1.82	34	348
BMP _a	3.55	5.01	6.31	0.20	-1.83	3.62	0.89	45	1.84	38	359

^a BMP, best management practice. CY, WY and MY, yields of cotton, wheat and maize, respectively

855 (Mg dry matter $ha^{-1} yr^{-1}$). For the other variable names and units, refer to the footnotes of Fig. 2. BAS,

BMP and BMP_a , baseline management practices, the best management practices (BMP) and the BMP

 $857 \qquad alternative, \ which \ are \ encoded \ as \ N110/430_I10_IF_RI1_T20, \ N94/366_I08_IS_RI1_T10, \ N94/IS_$

858 N94/366_I06_IS_RI1_T10, respectively. For the meanings of these codes, refer to Table 1.





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Figure 1: Observed and simulated daily mean soil (5 cm) temperature, soil (0–6 cm) moisture, nitrous oxide (N₂O) and nitric oxide (NO) fluxes, net ecosystem exchanges of carbon dioxide (NEE), and methane (CH₄) emissions. The solid- and dashed-line arrows indicate the dates of fertilization and irrigation, respectively. The measurement errors were not shown in panels a, b and e. The vertical bar for each observation in panels c, d and f indicates two times standard deviations to represent the uncertain range at the 95% confidence interval. The legend in panel a applies for all subfigures.







866

867Figure 2: Simulated crop yields, cumulative changes in soil organic carbon (\triangle SOC), methane (CH₄) and,868nitrous oxide (N₂O) release, net ecosystem greenhouse gases emission (NEGE), ammonia (NH₃) volatilization,869nitric oxide (NO) emission and nitrate leaching (NL) over a 18-year period (spanning three 6-year rotation870cycles) for each rotation pattern scenario. For the definitions of the rotation pattern options, i.e., R₀, R₁,...,871R₅, refer to the text in Sect. 2.4. The legend in panel e applies for all subfigures.







872

873 Figure 3: Simulated effects of rotation options under the baseline management practices on decision 874 variables and negative impact potential (NIP). For the definitions of the rotation pattern options, i.e., R₀, 875 R₁,..., R₅, refer to the text in Sect. 2.4. The y-axis units are Mg C ha⁻¹ yr⁻¹ for the opposite of mean annual increase in soil organic carbon stock (- Δ SOC), kg C ha⁻¹ yr⁻¹ for methane (CH₄) emission, kg N ha⁻¹ yr⁻¹ 876 for emissions of nitrous oxide (N2O), ammonia (NH3) and nitrous oxide (NO), and nitrate leaching (NL), Mg 877 $CO_2eq ha^{-1} yr^{-1}$ for net ecosystem greenhouse gas emission (NEGE), and USD $ha^{-1} yr^{-1}$ for NIP. The CO_2eq 878 was based on the 100-year global warming potentials, i.e., 34 for CH₄ and 298 for N₂O (IPCC, 2013). The 879 880 NIP was calculated using Eq. (1) presented in the text.







881

Figure 4: Averages of crop grain yields and negative impact potential (NIP), soil organic carbon change (Δ SOC), emissions of methane (CH₄) and nitrous oxide (N₂O), and net ecosystem greenhouse gases emission (NEGE), ammonia (NH₃) volatilization, nitrate leaching (NL) and nitric oxide (NO) emission over a 18-year period for different management scenarios under the optimal rotation pattern (i.e., R₃, see Fig. 2 for its details). For the definitions of each scenario, refer to Table 1. Each NIP was calculated using Eq. (1)

887 presented in the text.