

1 **Biogeosciences**

2 **Research Paper**

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4 **Baseline for ostracod-based northwestern Pacific and Indo-Pacific shallow-**
5 **marine paleoenvironmental reconstructions: ecological modeling of species**
6 **distributions**

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28 Abstract:

29 Fossil ostracods have been widely used for Quaternary paleoenvironmental
30 reconstructions especially in marginal marine environments (e.g. for water depth,
31 temperature, salinity, oxygen levels, pollution). But our knowledge of indicator
32 species autoecology, the base of paleoenvironmental reconstructions, remains limited
33 and commonly lacks robust statistical support and comprehensive comparison with
34 environmental data. We analysed marginal marine ostracod taxa at 52 sites in Hong
35 Kong for which comprehensive environmental data are available. We applied linear
36 regression models to reveal relationships between species distribution and
37 environmental factors for 18 common taxa (mainly species, a few genera) in our
38 Hong Kong dataset, and identified indicator species of environmental parameters. For
39 example, *Sinocytheridea impressa*, widely distributed euryhaline species throughout
40 the East and South China Seas and the Indo-Pacific, indicates eutrophication and
41 bottom-water hypoxia. *Neomonoceratina delicata*, widely known species from
42 nearshore and estuarine environments in the East and South China Seas, and the Indo-
43 Pacific, indicates heavy-metal pollution and increased turbidity. The 18 taxa used for
44 this study are widely distributed geographically and divided into Widespread
45 (throughout the northwestern Pacific and Indo-Pacific regions), Temperate [South
46 China Sea to Russia (Sea of Japan coast) and Japan], Subtropical (Indo-Pacific to the
47 East China Sea), Tropical (Indo-Pacific and South China Sea), and Globally
48 Distributed Groups. With statistical support from ecological modeling and
49 comprehensive environmental data, these results provide a robust baseline for

ostracod-based Quaternary-Anthropocene paleoenvironmental reconstructions in the tropical–extratropical northwestern Pacific and Indo-Pacific widely.

Key words: Autoecology, Distribution modeling, Indicator species, Ostracoda, Paleoenvironmental reconstruction, Proxy.

Key points

1. We provide a robust baseline for ostracod-based (microscopic Arthropods) paleoenvironmental reconstructions from Quaternary and Anthropocene marginal marine sediments.

2. The studied species have wide distributions over the tropics and extratropics of the northwestern Pacific and Indo-Pacific.

3. Ecological modeling established reliable indicator ostracod species for paleoenvironmental reconstructions.

1. Introduction

Because of their small size, high abundance and excellent fossil record, fossil ostracods (microcrustaceans), have been widely used to reconstruct Quaternary environmental conditions including water-depth, salinity, temperature, oxygen, and pollution, especially in marginal marine sediments (Boomer & Eisenhauer, 2002; Cronin, 2015; Frenzel & Boomer, 2005; Horne et al., 2012; Ruiz et al., 2005). In the northwestern Pacific and Indo-Pacific there are numerous deltas (Woodroffe et al., 2006) hosting accumulated Holocene marine sediments. Many studies have

reconstructed the depositional environments of these sediments (Alberti et al., 2013; Dong et al., 2012; Irizuki et al., 2015b; Tanaka et al., 2011; Yasuhara & Seto, 2006; Yasuhara et al., 2005; Zhou et al., 2015; Wang et al., 2018). Due to high sedimentation rates (> 1 cm per year), fossil ostracods allow the high-resolution reconstruction of human-induced environmental changes (pollution, eutrophication, bottom oxygen depletion) over the past century (Irizuki et al., 2011; Irizuki et al., 2015a; Irizuki et al., 2018; Yasuhara et al., 2003; Yasuhara et al., 2007).

Many have evaluated the autoecology of ostracod indicator species as the basis for paleoenvironmental reconstructions (Hazel, 1988; Irizuki et al., 2003; Ozawa et al., 2004; Stepanova et al., 2003; Wang et al., 1988; Yasuhara & Seto, 2006; Zhao, 1984; Zhao & Wang, 1988a, 1988b). Yet these studies tend to focus on only one or a few targeted environmental factor(s) and lack rigorous statistical evaluation, particularly statistical modeling, a common approach in contemporary ecology. This is probably due to the fact that comprehensive environmental datasets are often unavailable and an ecological modeling approach (especially regression modeling and model selection) has not been common in this field of micropaleontology.

Hong Kong constitutes an ideal location for a marine ecological modeling approach in the northwestern Pacific and Indo-Pacific regions because of extensive and intensive marine environmental monitoring program and subtropical location where tropical and temperate species coexist. This program provides robust datasets for ecological modeling, and the subtropical location allows the investigation of species with different latitudinal and geographical distributions. We employed regression modeling of Hong Kong shallow-marine ostracod species to show

statistical relationships between species abundance, distribution and environmental factors. This study allows the autoecology and statistical evaluation of common tropical and extratropical species, providing a baseline for ostracod-based shallow-marine paleoenvironmental reconstructions of the northwestern Pacific and Indo-Pacific regions.

2. Study area

Hong Kong is situated at the southeastern corner of the Pearl River (Zhujiang) Delta, and has an area of 2500 km² (Fig. 1), at 22° 12.021' to 22° 33.817' N latitude, 113° 53.388' to 114° 26.920' E longitude. It is an ideal location to study natural and/or anthropogenic impacts on benthic communities due to its complex hydrology and long history of human influence. Western Hong Kong is affected by input from the Pearl River, particularly during the summer heavy rainfall (Morton & Wu, 1975), when surface water salinity is strongly influenced by rainfall. Eastern Hong Kong waters are mainly marine and derived from the South China Sea. As one of the most urbanized coastal areas in the world, human activities including rapid industrialization, sewage discharge, trawling, dredging and land reclamation have led to a deteriorated marine benthic ecosystem (Blackmore, 1998; Hodgkiss & Yim, 1995; Hong et al., 2017; Hu et al., 2008; Morton, 1996; Morton & Blackmore, 2001; Owen & Sandhu, 2000; Shin, 1977; Tanner et al., 2000).

3. Materials and Methods

3.1. Samples and laboratory procedure

In January and July of 2011 we used a Van Veen Grab to collect 100 ml of sediment from the uppermost centimeter of the sea floor from 52 sites in Hong Kong marine

waters (Fig. 1; Supplement A), 41 of which are in open waters and 11 in typhoon shelter sites. Typhoon shelters are semi-enclosed areas of water designed to protect moored vessels in extreme weather (Environmental Protection Department, 2011). All sites are included in the Hong Kong EPD marine water and sediment quality monitoring program, which has been conducted monthly since 1986 (for water) and biannually (for sediment), providing comprehensive environmental data for all stations (see below).

Sediments were wet-sieved over a 63 μm mesh sieve and air dried or oven-dried at 40 $^{\circ}\text{C}$. The residue was dry-sieved over a 150 μm mesh sieve, and ostracod specimens larger than 150 μm were picked; smaller individuals are mostly early instar juveniles that are often not preserved (because their shells are usually thin and delicate) or difficult to identify (see Yasuhara et al., 2009 and Yasuhara et al., 2017 for more details). In samples containing fewer than 200 specimens, we picked all individuals. If there were more than 200 specimens, we picked ostracods from a split. We identified each counted specimen to species level when possible. We considered both an entire carapace or a single valve as one individual for counting.

3.2 Environmental Variables

We selected parameters from the EPD monitoring program (Table 1) for our regression modeling (see below), including surface productivity (Chlorophyll-a, Chl; Fig. 2), water depth (D; Fig. 2), bottom water dissolved oxygen (DO; Fig. 2), bottom water salinity (Sal; Fig. 2), turbidity (Tur; Fig. 2), summer bottom water temperature (ST – June to September average; Fig. 2), winter bottom water temperature (WT – November to February average; Fig. 2), mud content (MD; Fig. 2), and heavy metal concentration (Cu, Zn, and Pb; Fig. 3). These parameters are known to control

ostracod faunal properties (Cronin, 2015; Cronin & Vann, 2003; Hazel, 1988; Ikeya & Shiozaki, 1993; Irizuki et al., 2005; Irizuki et al., 2015a; Irizuki et al., 2018; Ruiz et al., 2005; Yasuhara et al., 2007; Yasuhara et al., 2012b). We used averages over the entire monitoring period (1986–2011), because the ostracods in this study were mostly dead shells, thus the samples should be considered time averaged. Bottom water DO is the average of the summer season (June–September), due to the likely importance of summer bottom water oxygen depletion.

4. Regression modeling

Multiple linear regression modeling was used to determine the relationship between common ostracod species (relative abundance of each species) and environmental parameters (Supplement B). All environmental parameters were log-transformed and zero centered. Salinity outliers (salinity <25: DS2 and DS3) and samples with low abundance (<50 specimens: ES5, MS7, VS21, VS5 & VS6) were removed from the models.

The best-fitting models were selected based on Akaike's Information Criterion (AICc) for small sample size, in which the lower score indicates the better model support considering both goodness-of-fit and model complexity (Anderson & Burnham, 2002). Akaike weights were used to summarize proportional support for all candidate models (Anderson et al., 2000) (Table 2). We considered parameter estimates averaged over models, proportional to the support that each model received (Anderson et al., 2000) (Table 3). This approach accounts for the uncertainty in model selection and thus leads to appropriately broader confidence intervals than obtained by relying only on the single, best-supported model. The relative importance of various predictor variables was measured by the sum of Akaike weights of models

that included the variables in question (Brunham & Anderson, 2002).

We explored linear dependencies by computing variance inflation factors (VIF) (Legendre & Legendre, 1998) and pairwise correlations between predictor variables to assess whether multicollinearity was likely to influence regression results (Yasuhara et al., 2012b). The degree of freedom is more than one for the geographic region variable (see below), thus we computed generalized variance inflation factors (GVIF). For continuous variables, GVIF (Table 4) is the same as VIF, but for categorical variables, GVIF has degrees of freedom (Df) equal to the number of coefficients associated with it (Hendrickx et al., 2004). Thus, we used $GVIF^{1/2df}$ to make GVIF values comparable among those with different Df. $VIF > 20$ is usually indicative of high collinearity (Legendre & Legendre, 1998). Thus we calculated an equivalent threshold of 4.47 (equal to $\sqrt{20}$) for $GVIF^{1/2df}$ to assess conlinearity. Also, adjusted $R^2 > 0.8$ indicates a strong correlation of variables (Hoffman, 2015). In all datasets, summer temperature (ST) and copper (Cu) were highly correlated ($R^2=0.8217$), and the GVIFs of ST and Cu are >20 , indicating that these correlations may influence regression results. Thus, we re-ran the linear regression modeling without ST and Cu. The new GVIFs of all variables were under 4.47 (Table 4).

We considered the degree of spatial autocorrelation in model residuals by the calculation of Moran's I index for the five best models. The neighborhood size was set as 2, 5, 10, 20 and 50 km. We found significant spatial autocorrelation in model residuals for many cases, thus we forced the geographic region variable (R) (Water Control Zones defined by EPD) to be included in all models. After this treatment, spatial autocorrelation was detected only in a few models for *Propontocypris* spp., *Stigmatocythere roesmani*, and *Hemikrithe orientalis*.

The multiple linear regression model analyses were implemented in R

programming language (R Core Team, 2016). We used 'MuMIn' (Bartoń, 2013) for model averaging and 'SPDEP' (Bivand & Piras, 2015) to measure spatial autocorrelation.

5. Results and discussions

The comprehensive ostracod dataset for the 52 sites and the environmental variables enabled us to elucidate distribution patterns of common ostracod taxa and their related environmental factor(s). We identified 151 species belonging to 76 genera (Supplement A). Among them, 18 common taxa (mainly species, a few genera) of *Pistocythereis bradyi*, *Bicornucythere bisanensis* s.l., *Nipponocythere bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica*, *Loxoconcha epeterseni*, *Sinocytheridea impressa*, *Neomonoceratina delicata*, *Keijella kloempritsensis*, *Neosinocythere elongata*, *Stigmatocythere roesmani*, *Hemicytheridea reticulata*, *Loxoconcha malayensis*, *Alocopocythere goujoni*, *Hemikrithe orientalis*, *Propontocypris* spp., *Neonesidea* spp. and *Xestoleberis* spp. (Supplement B) were used for regression modelling, and their relative abundances (to the total ostracod abundance in a sample) show a significant relation with environmental parameters. The best three regression models are presented in Table 2 and the model-averaged parameter estimates in Table 3. A small percentage of specimens of phytal genera (e.g., *Xestoleberis* spp., *Neonesidea* spp.) were contained in each sample, which are basically allochthonous specimens in bottom sediments transported from surrounding phytal environments. The value of allochthonous species to environmental interpretation is limited, however most ostracod specimens in each sample are composed of benthic, muddy sediment dwellers which are considered autochthonous.

Ostracods were divided into four groups based on their geographic distributions, including (a) Widespread Group; (b) Temperate Group; (c) Subtropical Group; (d) Tropical Group; and (e) Globally distributed Group (Fig. 4).

a. Widespread Group

Only one species *Pistocythereis bradyi* constitutes Widespread Group. *Pistocythereis bradyi* is widely distributed throughout the marginal marine environments of Japan, the East and South China Seas, and the Indo-Pacific (Fig. 4). Relative abundance of *P. bradyi* was highly correlated with salinity (positive), water depth (negative) and Pb (negative) (Tables 2 and 3). This indicates that the species prefers relatively shallow environments with high salinity (Figs. 2 and 5). In the Pearl River Delta and shallow South China Sea (Fig. 4), *P. bradyi* is dominant along the inner continental shelf at water depths <100 m (mostly common between 10–50 m), and at salinities from 30–40 (Li, 1985; Zhao & Wang, 1990; Zhao et al., 1986). *Pistocythereis bradyi* is a typical middle muddy bay species in Japan (Irizuki et al., 2006; Yasuhara & Irizuki, 2001; Yasuhara & Seto, 2006), and known from open bays such as Gamagyang Bay in Korea (Abe, 1988) and Malacca Strait (Whatley & Zhao, 1988b). In these studies, *P. bradyi* prefers relatively high salinity and deeper water in the inner continental shelf. Our data agree as to the preference for high salinity, but inconsistent with the literature regarding shallower water depths (Tables 2 and 3). Salinity may be more important than depth, but the restricted depth range of our sites (≤ 35 m) may also be a reason for this inconsistency. Our results indicate that *P. bradyi* is sensitive to metal pollution (Tables 2 and 3).

b. Temperate Group

251 Five species including *Bicornucythere bisanensis* s.l., *Nipponocythere*
252 *bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica* and *Loxoconcha*
253 *epeterseni* are distributed from Japan-Russian Coast to South China Sea.

254 *Bicornucythere bisanensis* s.l. is very common in most samples, the relative
255 abundance of which was significantly correlated with water depth (negative) (Tables
256 2 and 3) and this species prefers shallower environments (Figs. 2 and 5). In Chinese
257 and Japanese coastal areas, *B. bisanensis* s.l. is abundant in brackish water (salinity:
258 20–30) at depths less than 10 m (Ikeya & Shiozaki, 1993; Irizuki et al., 2006; Zhao et
259 al., 1986). Our results confirm this preference for shallow depths. In Japan,
260 *Bicornucythere bisanensis* is tolerant of anthropogenic impacts, especially
261 eutrophication and the resulting bottom water hypoxia (Irizuki et al., 2003; Irizuki et
262 al., 2011; Irizuki et al., 2015a; Irizuki et al., 2018; Yasuhara et al., 2003; Yasuhara et
263 al., 2007; Yasuhara et al., 2012a). We did not see a significant relation between
264 relative abundance and metal concentration, productivity, or dissolved oxygen.
265 Possibly, the more dominant presence of *S. impressa* and *N. delicata*, that are neither
266 dominant or distributed throughout most of Japan, could explain this difference.
267 These species may have a higher tolerance than *B. bisanensis* s.l.. Another
268 explanation may be that different morphotypes have different ecological preferences
269 (Abe, 1988), and only Form A is known to be tolerant to eutrophication and bottom-
270 water oxygen depletion (Irizuki et al., 2011; Irizuki et al., 2015a; Irizuki et al., 2018;
271 Yasuhara & Yamazaki, 2005; Yasuhara et al., 2007). Form A is less abundant in
272 Hong Kong, and due to the difficulty of juvenile identification, we did not divide *B.*
273 *bisanensis* into morphotypes. *Bicornucythere bisanensis* s.l. is widely distributed
274 throughout marginal marine environments around Japan, Russia (Sea of Japan coast),
275 and the East and South China Seas (Fig. 4).

276 Relative abundance of *N. bicarinata* correlated with productivity (negative)
277 (Tables 2 and 3). This is a typical middle bay species in Japan (Irizuki et al., 2006),
278 abundant on muddy substrates at water depths >10 m (Yasuhara & Seto, 2006;
279 Yasuhara et al., 2005). We found *N. bicarinata* to be sensitive to eutrophication,
280 preferring lower productivity (Figs 2 and 5). This species is known from marginal
281 marine environments around Japan and the East and South China Seas (Fig. 4).

282 Relative abundance of *S. quadriaculeata* correlated to productivity (positive),
283 and turbidity (negative) (Tables 2 and 3). This is a typical inner muddy bay species in
284 Japan (Irizuki et al., 2006), which prefers silty substrates in brackish waters, at
285 salinities from 20–30, and water depths of 2–7 m (Ikeya and Shiozaki, 1993). This
286 study shows a preference for waters with higher productivity but relatively low
287 turbidity (Tables 2 and 3), so that the species is abundant in Tolo Harbour (higher
288 productivity, lower turbidity) but not in Deep Bay (higher turbidity) (Figs. 2 and 5).
289 *Spinileberis quadriaculeata* is not tolerant to seasonal anoxia or oxygen depletion (0–
290 1 mg/L) in Uranouchi Bay, Japan (Irizuki et al., 2008), but we do not find a
291 significant correlation with dissolved oxygen content, probably due to the relatively
292 high bottom-water oxygen content (2.96–6.84 mg/L) in Hong Kong (Fig. 2;
293 Supplement B). *Spinileberis quadriaculeata* is widely distributed in marginal marine
294 environments around Japan, Russia (Sea of Japan coast), and the East and South
295 China Seas (Fig. 4).

296 Relative abundance of *P. japonica* correlated with water depth (positive)
297 (Tables 2 and 3). This species is known from relatively deeper waters (>40 m) in the
298 East China Sea (Ishizaki, 1981; Wang et al., 1988). At our sites, it has its greatest
299 abundance at the deeper southern sites (Fig. 5). *Phlyctocythere japonica* is distributed
300 around Japan (Yasuhara et al., 2002) and the East and South China Seas (Fig. 4).

Similarly to *P. japonica*, relative abundance of *L. epeterseni* correlated with water depth (positive), and turbidity (negative) (Tables 2 and 3). It occurs in the southern and eastern, deeper and less turbid regions of Hong Kong waters, but the trend is not very clear (Figs 2 and 5). This species is also known from the deeper parts of Osaka Bay (Yasuhara & Irizuki, 2001) and marginal marine environments around Japan (Ishizaki, 1968), the East China Sea (Hou et al., 1982), and the South China Sea (Cao, 1998) (Fig. 4). This species is reported as *Loxoconcha modesta* in Hou & Gou (2007), and also has been misidentified as *Loxoconcha viva* and *Loxoconcha sinensis* (Hou & Gou, 2007). Ishizaki (1968) described *Loxoconcha laeta* and *Loxoconcha modesta*, but these are the females and males of the same species (Ikeya et al., 2003). Ishizaki (1981) gave the new species names *Loxoconcha epeterseni* and *Loxoconcha tosamodesta* for *Loxoconcha laeta* and *Loxoconcha modesta*, respectively, because these names were junior homonyms. Since *Loxoconcha laeta* (= *epeterseni*) appears earlier than *Loxoconcha modesta* (= *tosamodesta*) in the original description (Ishizaki, 1968), we use the name *Loxoconcha epeterseni* for this species (e.g., see Ikeya et al., 2003).

c. Subtropical Group

Six species including *Sinocytheridea impressa*, *Neomonoceratina delicata*, *Keijella kloempritsensis*, *Neosinocythere elongata*, *Stigmatocythere roesmani* and *Hemicytheridea reticulata* are reported from the East China Sea to Indo-Pacific area.

Sinocytheridea impressa [= *Sinocytheridea latiovata*; see Whatley and Zhao (1988a)] is the most dominant species in this study, the relative abundance of which significantly correlated with salinity (negative), dissolved oxygen (negative), mud content (positive) and productivity (positive) (Tables 2 and 3). This species is

noticeably dominant in areas characterized by a muddy bottom including northern
Mirs Bay, Port Shelter and coastal Southern Waters (Fig. 5). It is also abundant in
Tolo Harbour, an area known for its summer hypoxia and eutrophication (Hu et al.,
2001; Sin & Chau, 1992). These results are consistent with previous studies indicating
that *S. impressa* is dominant in low salinity, nutrient-rich and turbid estuaries (Irizuki
et al., 2005; Tanaka et al., 2011), but we did not see a significant relation with
turbidity (Tables 2 and 3). *Sinocytheridea impressa* is known as a euryhaline species
widely distributed throughout the East and South China Seas [abundant in water
depths of <20 m; Whatley and Zhao (1988a)], and the Indo-Pacific (Fig. 4).

Neomonoceratina delicata (= *Neomonoceratina crispata*; see Hou & Gou,
2007) is very common in most of the samples, and the relative abundance
significantly correlates with Zn (positive), salinity (positive) and turbidity (positive)
(Tables 2 and 3). *Neomonoceratina delicata* is a nearshore species, abundant at depths
less than 30 m, at relatively high salinities (>30; Zhao and Wang, 1988). It prefers the
higher salinity waters in Hong Kong (Figs. 2 and 5; Tables 2 and 3) and in view of its
positive correlation with Zn and turbidity, is likely tolerant to human-induced
environmental stress such as pollution and eutrophication. This species is widely
known from nearshore and estuarine environments in the East and South China Seas,
and the Indo-Pacific (Fig. 4).

The relative abundance of *K. kloempritis* correlated only with water depth
(positive) (Fig. 6; Tables 2 and 3). Widely known from the tropical Indo-Pacific
region, *K. kloempritis* is abundant along the inner continental shelf of the South
and East China Seas (Fig. 4), at water depths ranging from 20–50 m and salinity close
to normal marine (Zhao & Wang, 1990). Our modelling results are consistent with
this, showing a preference for the relatively deeper waters in our study (Tables 2 and

3). Thus, this species is probably useful for reconstructing past sea-level changes in the broad tropical and subtropical Indo-Pacific and northwestern Pacific regions as a deeper water indicator.

Relative abundance of *N. elongata* correlated only with turbidity (positive) (Tables 2 and 3). This species occurs along the entire coast of China (Fig. 4) in marginal marine, especially estuarine environments shallower than 20 m (Dong et al., 2012; Hou & Gou, 2007; Liu et al., 2013; Liu et al., 2017; Zhao & Whatley, 1993) (Fig. 4). Known from the Indo-Pacific region, our modeling results and previous studies indicate consistently that *N. elongata* prefers shallow, turbid waters like Deep Bay and the Pearl River Estuary (Figs 2 and 6).

The relative abundance of both *S. roesmani* and *H. reticulata* correlated with Pb (negative) (Tables 2 and 3), thus they are sensitive to metal pollution (but note the significant autocorrelation with the modeling result of *S. roesmani*) and absent in areas with high metal concentrations, e.g., Tolo and Victoria Harbours (Fig. 3, 6). Relative abundance of *H. reticulata* also correlated with water depth (negative) (Tables 2 and 3). This species is abundant in Tolo Harbour and the inner part of Mirs Bay (Fig. 6), at shallow depths, and is also consistently found in very shallow waters from the Indo-Pacific (Zhao & Whatley, 1989). Their metal-pollution sensitivity is contradictory because they occur in Tolo and Victoria Harbours, both polluted regions of Hong Kong, and further research is needed to better understand these results (Figs 2 and 6). They occur in the East and South China Seas and the Indo-Pacific region (Fig. 4).

d. Tropical Group

Three species including *Loxoconcha malayensis*, *Alocopocythere goujoni* and *Hemikrithe orientalis* are distributed from the South China Sea to the Indo-Pacific.

Relative abundance of *L. malayensis* correlated with dissolved oxygen (negative) and mud content (negative) (Tables 2 and 3). It is a typical tropical species known from the Indo-Pacific and the South China Sea (Fig. 4). We did not find a correlation with temperature, likely due to the small range of variation of bottom water temperatures in Hong Kong (winter temperature: 19.10–21.49°C). This species prefers coarse sediments and is resistant to low oxygen content (Table 2 and 3), as seen by its abundance in Victoria Harbour (Figs. 2 and 6).

Relative abundance of both *A. goujoni* and *H. orientalis* correlated with natural factors only. *Alocopocythere goujoni* correlated with salinity (positive) (Tables 2 and 3) and occurs not only in Mirs Bay where the salinity is higher, but also in Deep Bay and North Western Waters where the salinity is lower than in other areas (Figs 2 and 6). The Deep Bay and North Western Waters are shallow and have relatively low oxygen content. The modeling result of this species shows a marginally insignificant but negative relationship to oxygen content and water depth with moderately high relative importance (Table 3). We explain this inconsistency by considering their preference for higher salinity and shallow water depths, and also their resistance to low oxygen conditions, but further research is needed to know their autoecology with better confidence. Relative abundance of *H. orientalis* correlated with water depth (positive) (Tables 2 and 3; but note the significant autocorrelation with the modeling result of this genus), and it is more abundant in deeper waters including southern Mirs Bay (Fig. 6). It is known from depths of 20–50 m in the South China Sea (Zhao & Wang, 1988a), and reported from tropical Indo-Pacific marginal marine environments (Fig. 4). Our regression modeling consistently shows a

positive relationship between relative abundance and winter temperatures, with moderately high relative importance, although the correlation is marginally insignificant (Table 2 and 3).

e. Globally Distributed Group

Propontocypris is known as a cosmopolitan genus. The relative abundance of *Propontocypris* spp. significantly correlated with productivity (negative) (Tables 2 and 3). This negative correlation with productivity (but note a significant autocorrelation with the modeling result of this genus) indicates that the genus prefers less eutrophic waters (Fig. 6). *Propontocypris* is a good swimmer (Maddocks, 1969), and thus may have an advantage in obtaining food in relatively food-limited environments.

Phytoplankton genera including *Neonesidea* spp. and *Xestoleberis* spp. have global distribution and are correlated with various environmental factors. The relative abundance of *Neonesidea* spp. correlated with dissolved oxygen (positive), as expected for a phytoplankton species (Smith & Kamiya, 2002; Yamada, 2007) (Table 2 and 3; Fig. 7). Similarly, phytoplankton (Irizuki et al., 2008; Sato & Kamiya, 2007; Yasuhara et al., 2002) *Xestoleberis* spp. correlated with dissolved oxygen (positive), turbidity (negative) and mud content (negative) (Table 2 and 3; Fig. 7). This taxon's habitat preference including clear water, coarse sediment, and high oxygen content are reflected in our modeling. As mentioned above, the value of allochthonous phytoplankton species to environmental interpretation is limited, but they broadly reflect adjacent phytoplankton environments.

Summary

Benthic ostracods from Hong Kong marginal marine waters studied here include widespread (i.e., one species distributed throughout the northwestern Pacific–Indo-Pacific region), temperate (i.e., five species distributed from the South China Sea to Japan and Russia), subtropical (i.e., six species distributed from the Indo-Pacific to the East China Sea), and tropical (i.e., three species distributed in the Indo-Pacific and South China Sea) species and three globally distributed genera (Fig. 4). We provide a robust baseline of autoecology for these common ostracod taxa based on rigorous statistical modeling using comprehensive environmental data. We established reliable indicator taxa for water depth, mud content, salinity, turbidity, dissolved oxygen, heavy metal pollution (Pb and Zn) and eutrophication (chlorophyll-a) (Table 5). Thus our results are applicable for future ostracod-based paleoenvironmental studies in a wide range of localities from the tropics to the extratropics, and from the Indian Ocean to the northwestern Pacific. We established pollution and eutrophication indicator species in tropical environments for the first time. Anthropocene paleoenvironmental and paleoecological studies in the tropics are urgently needed because (1) the tropics are seriously under-studied (Wilkinson et al., 2014; Yasuhara et al., 2012a), (2) tropical environments and ecosystems are vulnerable and sensitive to human influences (Jackson et al., 2001; Pandolfi et al., 2003), and (3) Indo-Pacific tropical environments have been seriously degraded by the human activity of rapidly developing countries (Bellwood et al., 2004; Jackson, 2008; Knowlton & Jackson, 2008). Our results provide useful and reliable tools for tropical Anthropocene research in the broad Indo-Pacific region.

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Captions

Fig. 1 Locality map showing the 52 sampling sites across Hong Kong, including 41 open water sites (blue dots) and 11 typhoon shelter sites (red open dots). From west to east, DS: Deep Bay; NS: North Western waters; SS: Southern waters; VS: Victoria Harbour; ES: Eastern Buffer; JS: Junk Bay; TS: Tolo Harbour; PS: Port Shelter; MS: Mirs Bay.

Fig. 2 Spatial distribution of environmental parameters in Hong Kong. Mean surface-water chlorophyll-a concentration; water depth; summer (June to September) bottom-water dissolved oxygen content; mean bottom-water salinity; mean turbidity; mean summer (June to September) bottom-water temperature; mean winter (November to February) bottom-water temperature; and mean mud content. All are averaged values of the data obtained during 1986–2013 (Table 1).

Fig. 3 Spatial distribution of environmental parameters in Hong Kong. Mean copper (Cu) concentration, mean lead (Pb) concentration, and mean zinc (Zn) concentration in surface sediments. All are averaged values of the data obtained during 1986–2013 (Table 1).

Fig. 4 Geographical distributions of the 18 taxa in the northwestern Pacific and Indo-Pacific regions, including *Pistocythereis bradyi*, *Bicornucythere bisanensis* s.l., *Nipponocythere bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica*, *Loxoconcha epeterseni*, *Sinocytheridea impressa*, *Neomonoceratina delicata*, *Keijella kloempitensis*, *Neosinocythere elongata*, *Stigmatocythere roesmani*, *Hemicytheridea reticulata*, *Loxoconcha malayensis*, *Alocopocythere goujoni*, *Hemikrithe orientalis*, *Propontocypris* spp., *Neonesidea* spp. and *Xestoleberis* spp.. The following references were used mainly to determine the geographical distributions of the species: Al Jumaily and Al-Sheikhly (1999); Dewi (1997); Dong et al. (2012); Fauzielly et al. (2013); Gu et al. (2017); Hong et al. (2017); Hou and Gou (2007); Hussain et al. (2004); Hussain et al. (2010); Hussain and Mohan (2000, 2001); Irizuki et al. (2006); Irizuki et al. (2009); Iwatani et al. (2014); Jie et al. (2013); Li (1985); Mostafawi (1992); Nishath et al. (2017); Noraswana et al. (2014); Pugliese et al. (2006); Schornikov et al. (2014); Tanaka et al. (2009); Tanaka et al. (2011); Wang et al. (1988); Wang and Zhang (1987); Wang and Zhao (1985); Zhao (1984); Zhao and Wang (1988a, 1988b, 1990); Zhao and Whatley (1993); Zhou et al. (2015). Note that *Sinocytheridea impressa* is known to be distributed in Japan, but their Japanese distribution is very limited in certain areas of the southern part of Japan (Iwasaki, 1992; Tanaka et al. 2019). Thus, we did not indicate their Japanese-Russian coast distribution in this figure.

Fig. 5 Spatial distribution of the relative abundance for *Pistocythereis bradyi*, *Bicornucythere bisanensis* s.l., *Nipponocythere bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica*, *Loxoconcha epeterseni*, *Sinocytheridea*

843 *impressa*, and *Neomonoceratina delicata* in Hong Kong. See Figure 1 for sampling
844 stations.

845

846 Fig. 6 Spatial distribution of the relative abundance for *Keijella kloempritis*,
847 *Neosinocythere elongata*, *Stigmatocythere roesmani*, *Hemicytheridea reticulata*
848 *Loxoconcha malayensis*, *Alocopocythere goujoni*, *Hemikrithe orientalis*, and
849 *Propontocypris* spp. in Hong Kong. See Figure 1 for sampling stations.

850

851 Fig. 7. Spatial distribution of the relative abundance for *Neonesidea* spp. and
852 *Xestoleberis* spp. in Hong Kong. See Figure 1 for sampling stations.

853

854

855 Table 1. Summary of marine water/sediment parameters. Note: 1. Summer: June,
856 July, August and September. 2. Winter: November, December, January and February.

857

858 Table 2. Best three regression models of the relative abundance of common species,
859 including *Pistocythereis bradyi*, *Bicornucythere bisanensis* s.l., *Nipponocythere*
860 *bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica*, *Loxoconcha*
861 *epeterseni*, *Sinocytheridea impressa*, *Neomonoceratina delicata*, *Keijella*
862 *kloempritis*, *Neosinocythere elongata*, *Stigmatocythere roesmani*, *Hemicytheridea*
863 *reticulata* *Loxoconcha malayensis*, *Alocopocythere goujoni*, *Hemikrithe orientalis*,
864 *Propontocypris* spp., *Neonesidea* spp. and *Xestoleberis* spp.. The table shows the
865 coefficient of each term, adjusted R^2 , the Akaike information criterion corrected for
866 small sample size (AICc), and the Akaike weight (AW). Bold denotes significance at

867 P < 0.05. Overall P is < 0.05 in all models. R: region. Other abbreviations are found in
868 Table 1.

869

870 Table 3. Model-averaged parameter estimates and CIs of the relative abundance for
871 common species including *Pistocythereis bradyi*, *Bicornucythere bisanensis* s.l.,
872 *Nipponocythere bicarinata*, *Spinileberis quadriaculeata*, *Phlyctocythere japonica*,
873 *Loxoconcha epeterseni*, *Sinocytheridea impressa*, *Neomonoceratina delicata*, *Keijella*
874 *kloempitensis*, *Neosinocythere elongata*, *Stigmatocythere roesmani*, *Hemicytheridea*
875 *reticulata* *Loxoconcha malayensis*, *Alocopocythere goujoni*, *Hemikrithe orientalis*,
876 *Propontocypris* spp., *Neonesidea* spp. and *Xestoleberis* spp.. CIs, confidence
877 intervals; RI, relative importance (the sum of the Akaike weights of models that
878 include the variable in question; see Materials and Methods); R, region. Other
879 abbreviations are found in Table 1. Bold denotes CIs that exclude zero. For R,
880 coefficient, lower CI, and upper CI values shown are averages of those for geographic
881 regions.

882

883 Table 4. GVIF value for Environmental Variables. Df, degree of freedom; R, region.
884 Other abbreviations are found in Table 1.

885

886 Table 5. Summary of autoecology for common ostracod taxa. Chl: Chlorophyll-a; D:
887 Water Depth; DO: Dissolved Oxygen; MD: Mud Content; Sal: Salinity; Tur:
888 Turbidity; WT: Winter Temperature; Pb: Lead; Zn: Zinc; R: Region. + and - marks
889 indicate significant positive and negative correlations, respectively.

890

891

892 Supplement A. Ostracod faunal list.

893

894 Supplement B. Dataset used for the regression modeling.

895

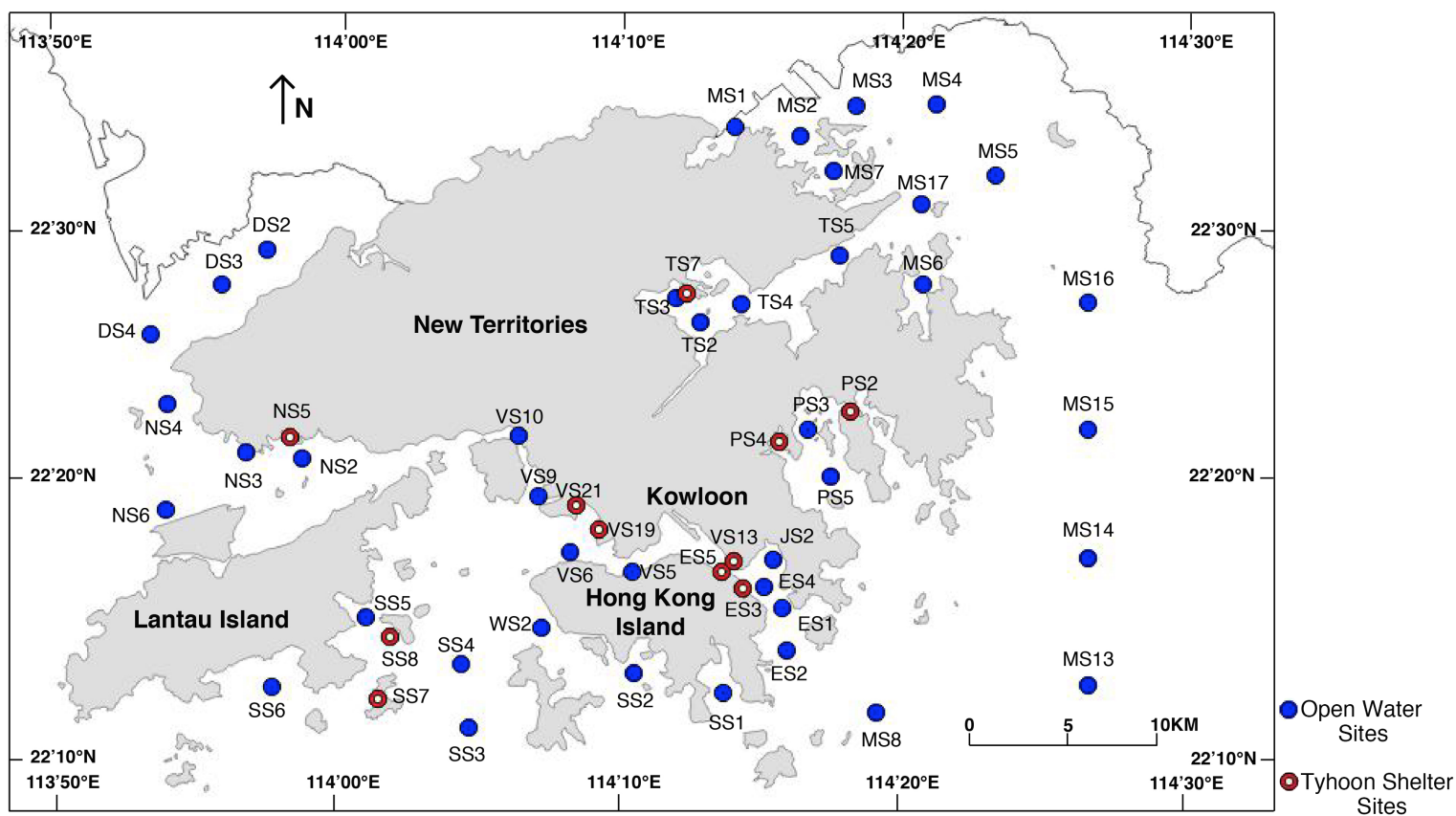


Fig. 1

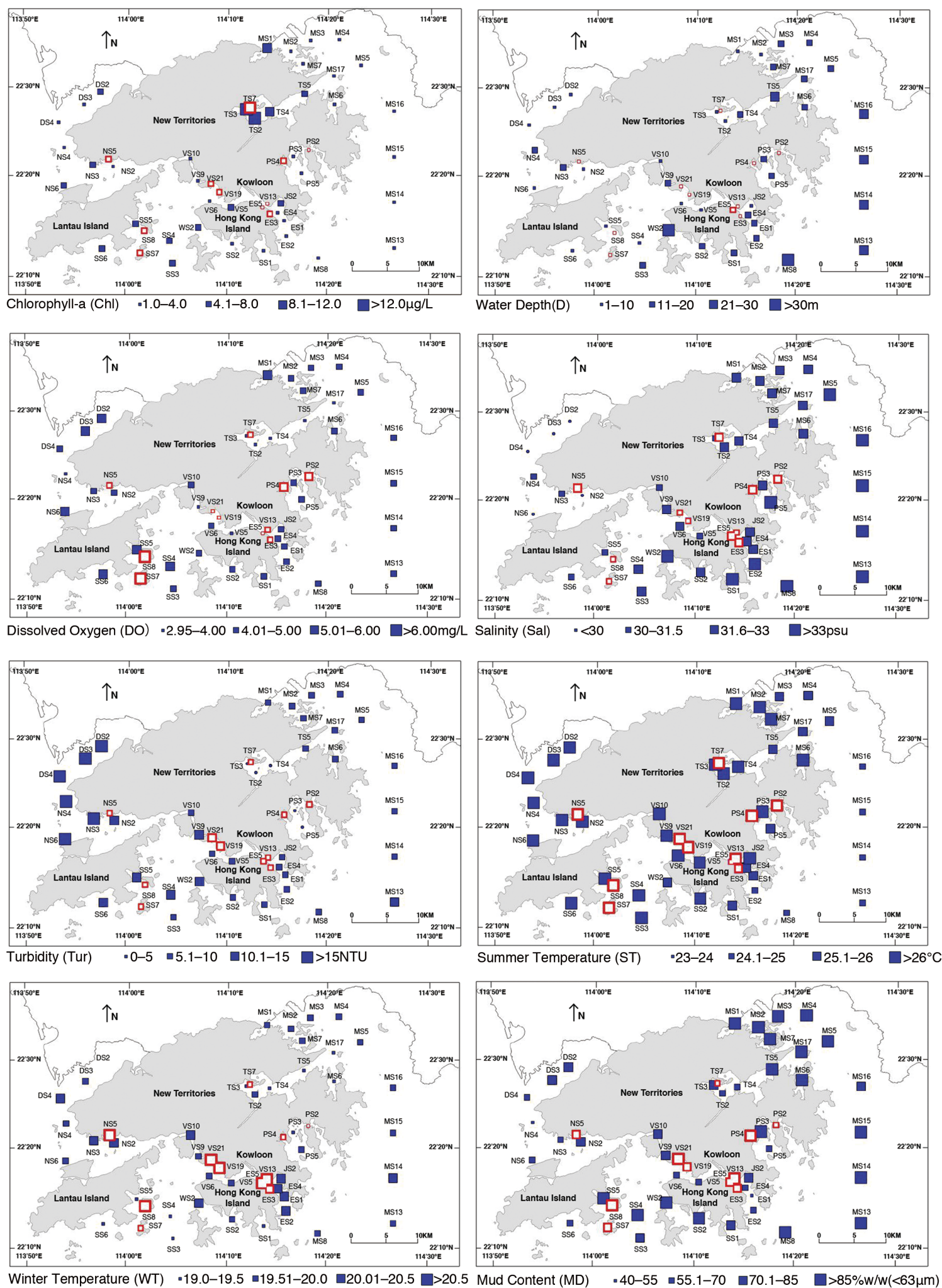


Fig. 2

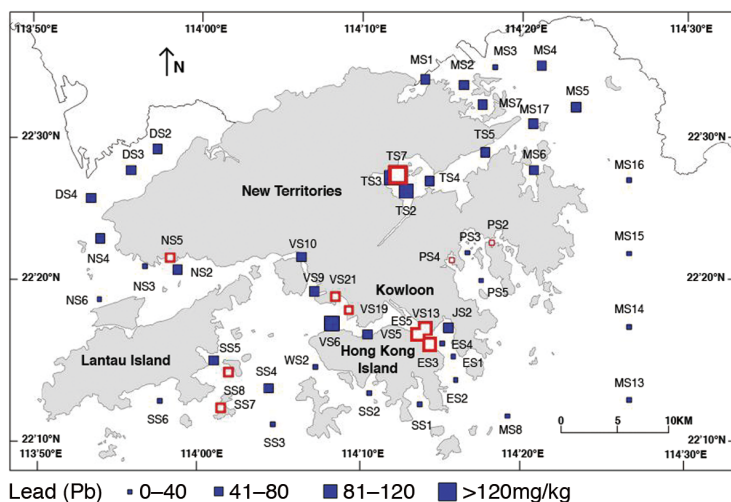
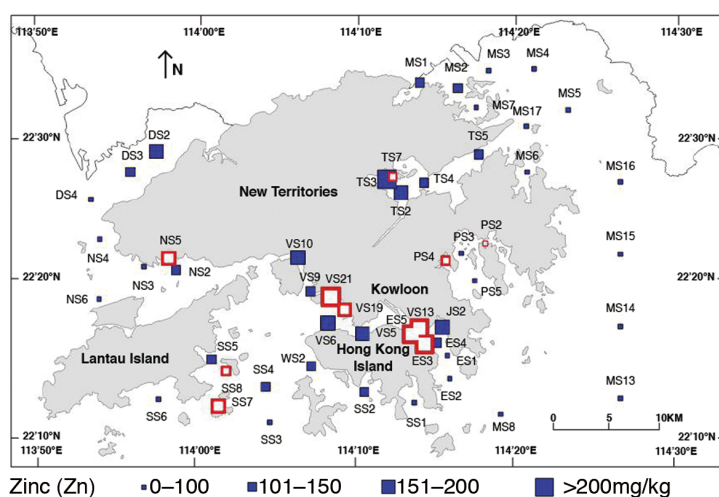
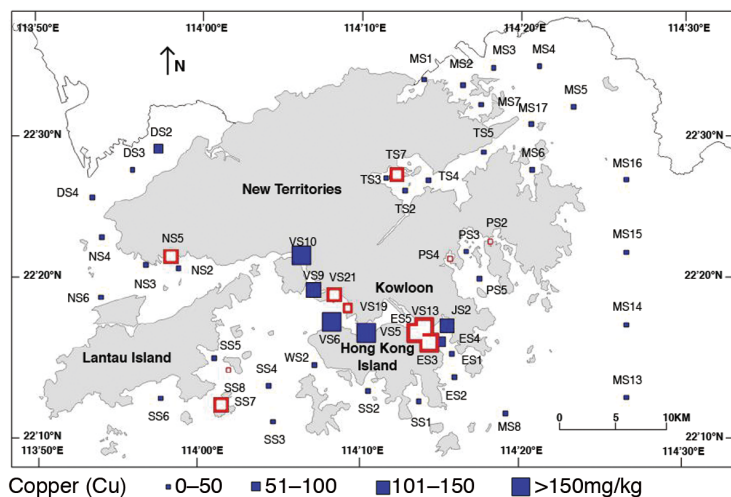
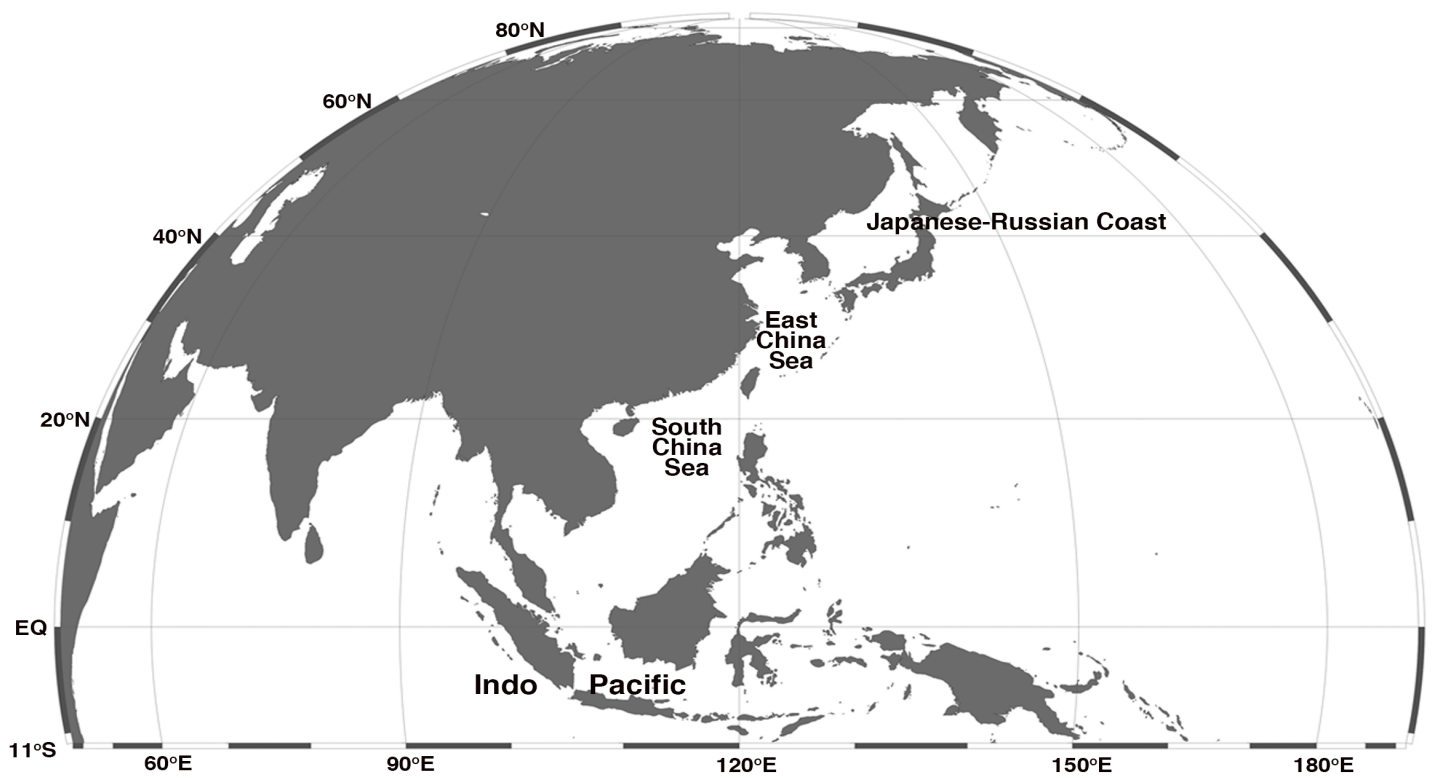
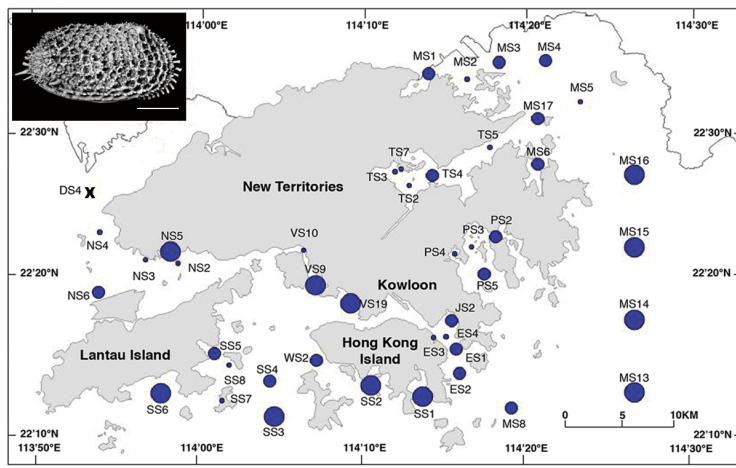


Fig. 3

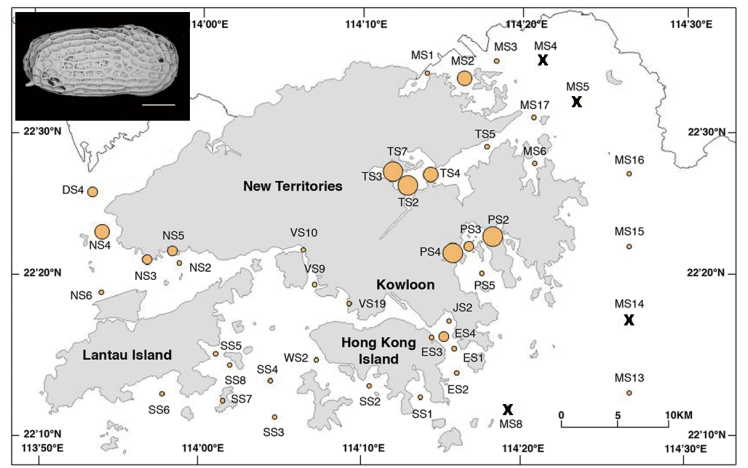


Species	Indo Pacific	South China Sea	East China Sea	Japanese-Russian Coast
<i>Pistocythereis bradyi</i>	(a)			
<i>Bicornucythere bisanensis</i> s.l.				
<i>Nipponocythere bicarinata</i>				
<i>Spinileberis quadriaculeata</i>		(b)		
<i>Phlytocythere japonica</i>				
<i>Loxoconcha epeterseni</i>				
<i>Sinocytheridea impressa</i>				
<i>Neomonoceratina delicata</i>				
<i>Keijella kloempritsensis</i>				
<i>Neosinocythere elongata</i>	(c)			
<i>Stigmatocythere roesmani</i>				
<i>Hemicytheridea reticulata</i>				
<i>Loxoconcha malayensis</i>				
<i>Alocopocythere goujoni</i>	(d)			
<i>Hemikrithe orientalis</i>				
<i>Propontocypris</i> spp.				
<i>Neonesidea</i> spp.	(e)			
<i>Xestoleberis</i> spp.				

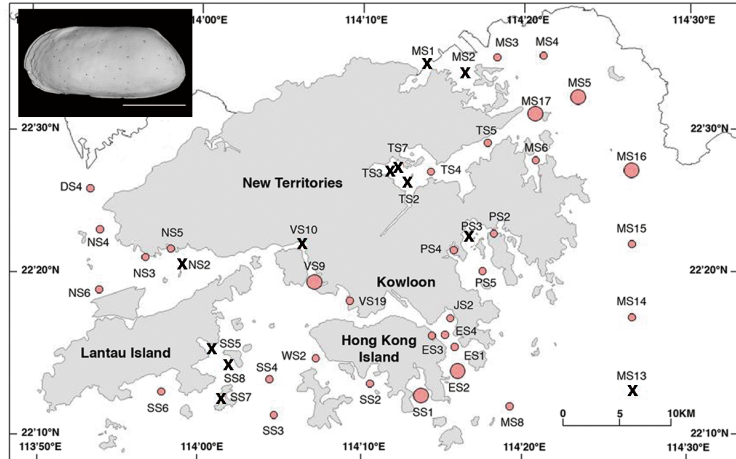
Fig. 4



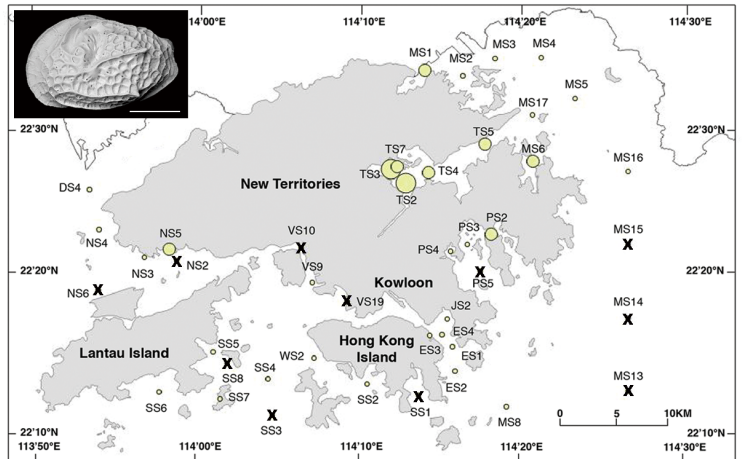
Pistocythereis bradyi 0 < • ≤ 5% 5 < • ≤ 10% 10 < • ≤ 15% x absent



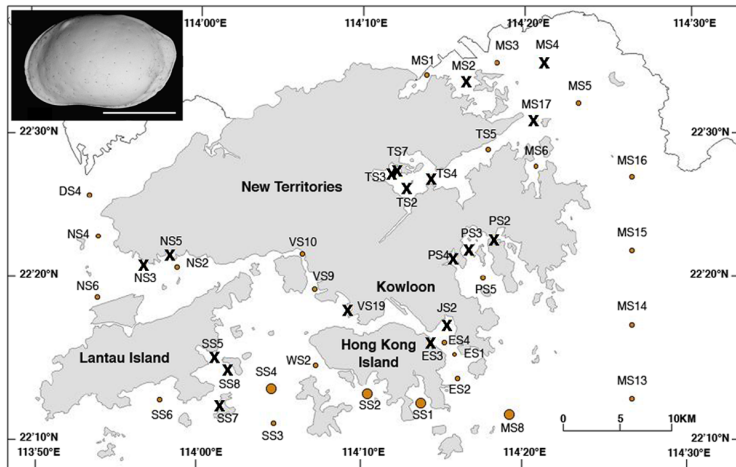
Bicornucythere bisanensis s.l. 0 < • ≤ 5% 5 < • ≤ 10% 10 < • ≤ 15% 15 < • ≤ 20% x absent



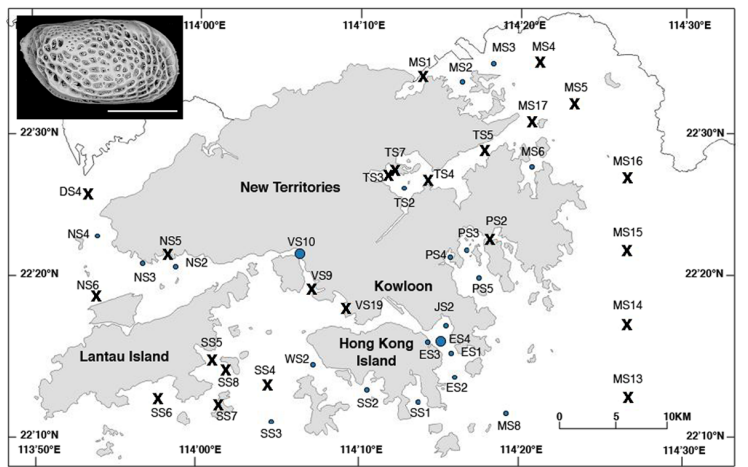
Nipponocythere bicarinata 0 < • ≤ 5% 5 < • ≤ 10% x absent



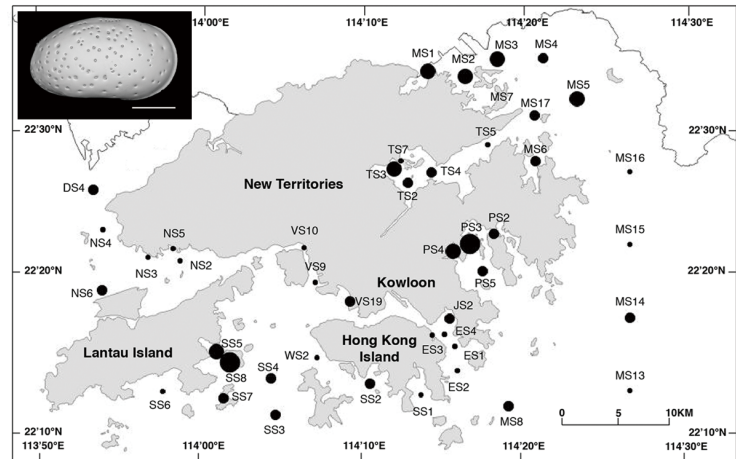
Spinileberis quadriaculeata 0 < • ≤ 5% 5 < • ≤ 10% 10% ≤ • x absent



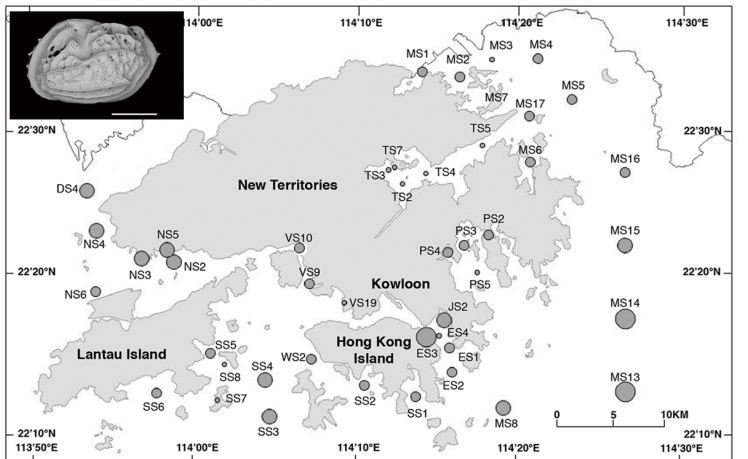
Phlyctocythere japonica 0 < • ≤ 3.5% 3.5 < • ≤ 7% x absent



Loxoconcha epeterseni 0 < • ≤ 3% 3 < • ≤ 6% x absent



Sinocytheridea impressa 0 < • ≤ 20% 20 < • ≤ 40% 40 < • ≤ 60% 60 < • ≤ 80%



Neomonoceratina delicata 0 < • ≤ 10% 10 < • ≤ 20% 20 < • ≤ 30% 30% < •

Fig. 5

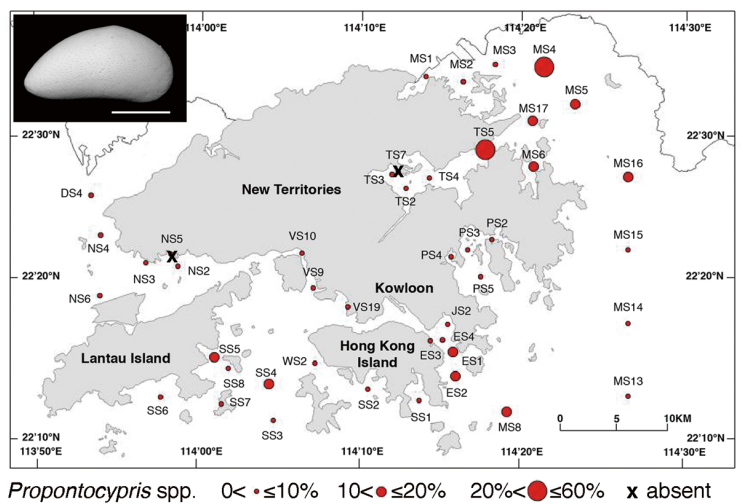
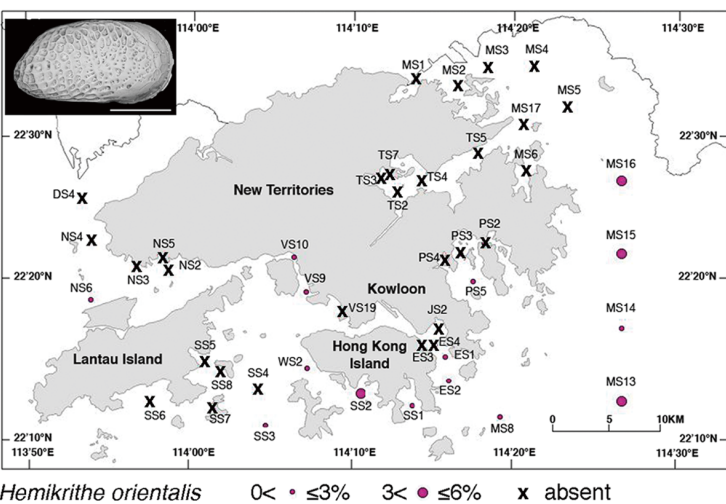
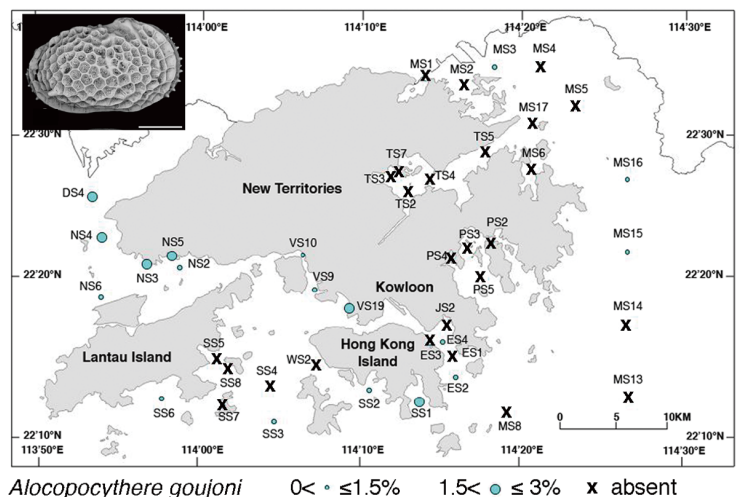
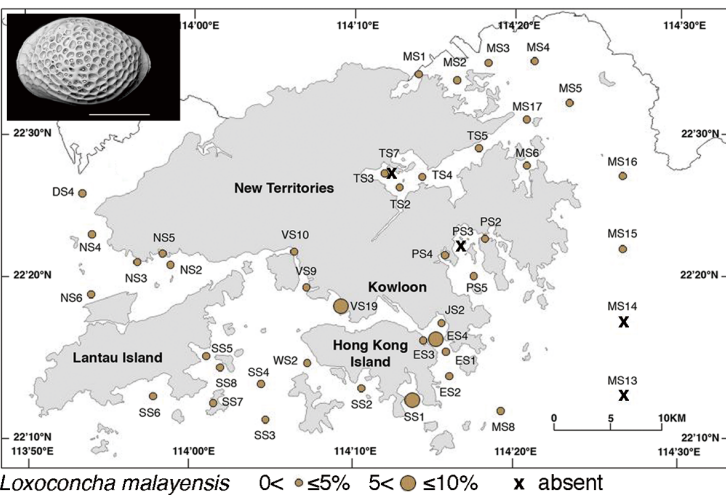
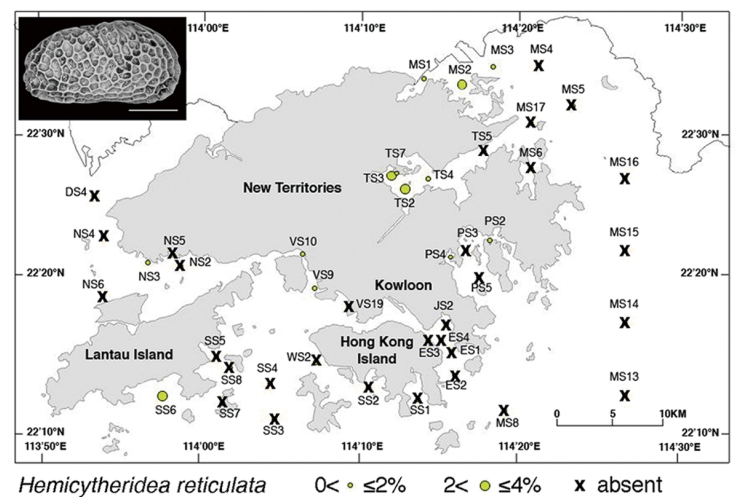
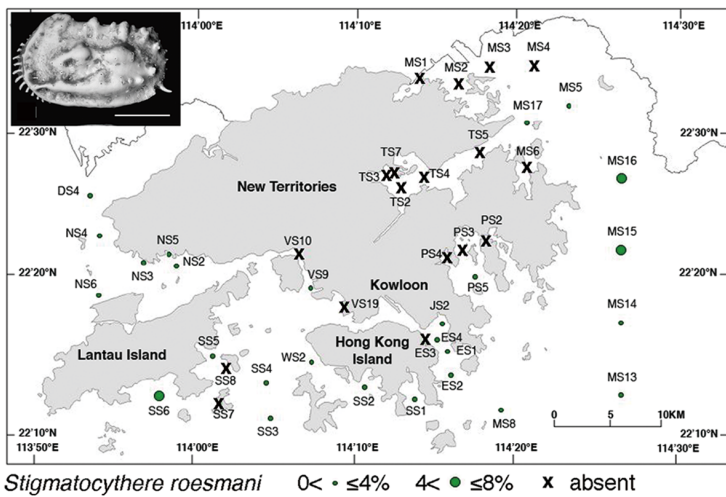
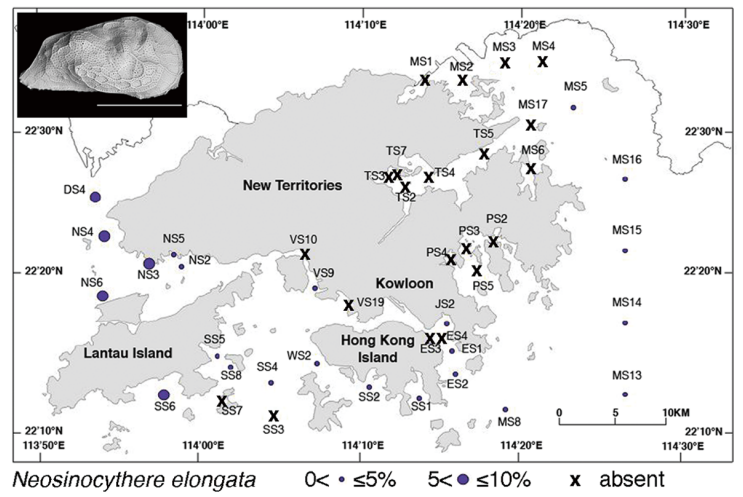
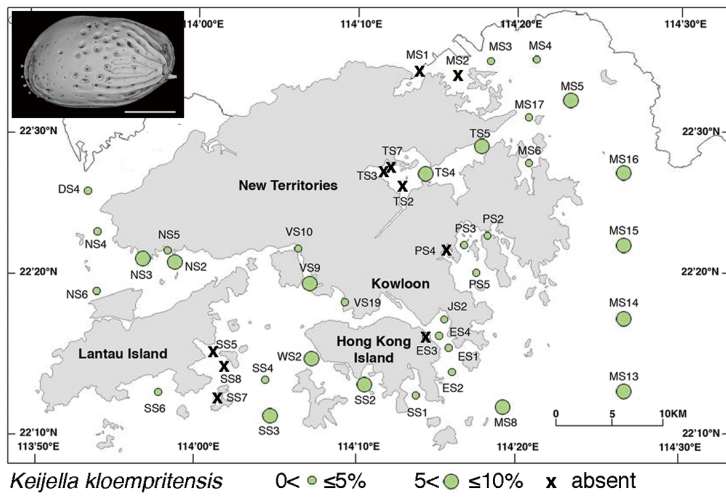


Fig. 6

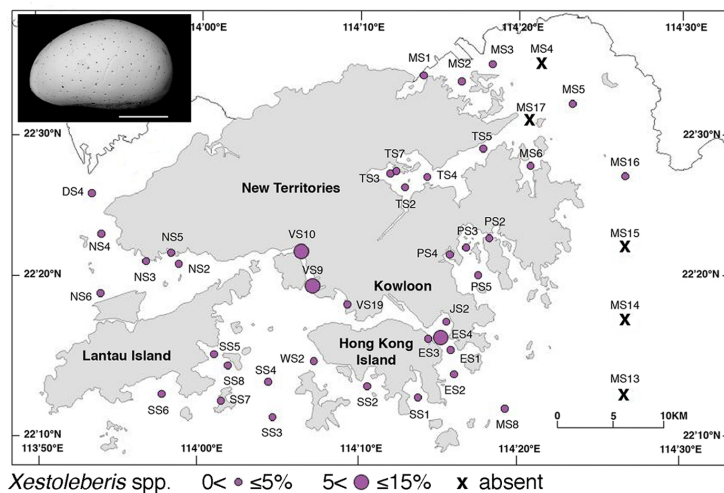
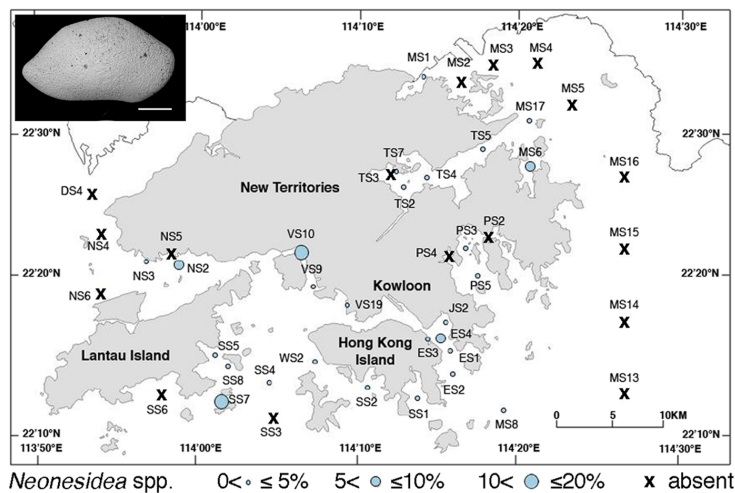


Fig. 7

Table 1.

Parameter (Abbreviation)	Unit	Sampling Depth / Material	Season
Chlorophyll- <i>a</i> (Chl)	µg/L	Surface water	All year
Water Depth (D)	m		All year
Dissolved Oxygen (DO)	mg/L	Bottom water	Summer ¹
Mud Content (MD)	%w/w (<63µm)	Bottom sediment	All year
Salinity (Sal)		Bottom water	All year
Turbidity (Tur)	NTU	Bottom water	All year
Summer Temperature (ST)	°C	Bottom water	Summer ¹
Winter Temperature (WT)	°C	Bottom water	Winter ²
Copper (Cu)	mg/kg	Bottom sediment	All year
Lead (Pb)	mg/kg	Bottom sediment	All year
Zinc (Zn)	mg/kg	Bottom sediment	All year

Table 2.

Model	Chl	D	DO	MD	Sal	Tur	WT	Pb	Zn	R	R ²	AICc	AW
<i>Pistocythereis bradyi</i>													
1		-0.12			2.67			-0.21		-0.06	0.61	-163.64	0.52
2		-0.13			2.87	0.04		-0.20		-0.05	0.61	-159.79	0.08
3		-0.12	0.07		2.83			-0.21		-0.07	0.60	-159.42	0.06
<i>Bicornucythere bisanensis</i> s.l.													
1		-0.09		-0.23						0.02	0.60	-137.80	0.12
2		-0.08								-0.01	0.56	-136.77	0.07
3				-0.29				0.18		0.00	0.58	-136.77	0.05
<i>Nipponocythere bicarinata</i>													
1	-0.07		-0.12							0.03	0.43	-200.57	0.11
2	-0.09									0.03	0.39	-200.38	0.10
3	-0.09	-0.03	-0.16							0.04	0.45	-199.30	0.06
<i>Spinileberis quadriaculeata</i>													
1	0.06					-0.09				-0.04	0.63	-183.88	0.10
2		-0.05								0.02	0.60	-182.94	0.06
3	0.08					-0.09	-0.75			-0.04	0.65	-182.66	0.05
<i>Phlyctocythere japonica</i>													
1		0.04								0.00	0.28	-238.60	0.19
2		0.03			0.33					-0.02	0.30	-237.01	0.09
3		0.03						-0.02		0.00	0.27	-235.62	0.04
<i>Loxoconcha epeterseni</i>													
1		0.03			-0.33	-0.03				0.01	0.49	-288.61	0.11
2	0.02	0.04		-0.04	-0.34	-0.04				0.00	0.56	-287.76	0.07
3	0.02	0.04			-0.34	-0.04				0.00	0.51	-287.31	0.06
<i>Sinocytheridea impressa</i>													
1	0.33		-1.47	0.75	-9.13					0.34	0.55	-37.29	0.27
2	0.41		-1.48		-8.75					0.37	0.50	-35.24	0.10
3			-1.12	0.94	-10.30					0.44	0.49	-34.78	0.08
<i>Neomonoceratina delicata</i>													
1					3.77	0.34			0.41	-0.13	0.48	-82.57	0.29
2			0.43		5.02	0.31			0.39	-0.20	0.50	-81.21	0.15
3					3.47	0.33		-0.20	0.55	-0.12	0.48	-79.32	0.06
<i>Keijella kloempitensis</i>													
1		0.06						-0.06		-0.02	0.56	-202.12	0.12
2		0.08								-0.02	0.53	-201.51	0.09
3		0.06		0.09				-0.08		-0.02	0.58	-200.86	0.06

Model	Chl	D	DO	MD	Sal	Tur	WT	Pb	Zn	R	R ²	AICc	AW
<i>Neosinocythere elongata</i>													
1	-0.02					0.09				-0.03	0.76	-244.70	0.09
2				-0.06		0.08				-0.03	0.76	-244.42	0.08
3						0.09				-0.03	0.74	-244.15	0.07
<i>Stigmatocythere roesmani</i>													
1								-0.08		0.00	0.23	-225.23	0.07
2					0.36			-0.07		-0.02	0.25	-224.06	0.04
3			0.10		0.67			-0.07		0.30	0.30	-224.03	0.04
<i>Hemicytheridea reticulata</i>													
1		-0.03						-0.04		0.01	0.23	-270.79	0.14
2	0.02	-0.02						-0.04		0.01	0.26	-269.44	0.07
3	0.03							-0.03		0.00	0.19	-268.64	0.05
<i>Loxoconcha malayensis</i>													
1		-0.03	-0.14	-0.12						0.03	0.55	-229.55	0.10
2	0.04		-0.13	-0.12						0.02	0.55	-229.42	0.09
3		-0.04	-0.14	-0.10				-0.04		0.03	0.57	-228.47	0.06
<i>Alocopocythere goujoni</i>													
1					0.21					-0.04	0.68	-331.97	0.09
2		-0.01			0.28				-0.01	-0.04	0.71	-330.52	0.05
3		-0.01			0.27			-0.02		-0.04	0.71	-330.45	0.04
<i>Hemikrithe orientalis</i>													
1		0.03					0.46			0.00	0.19	-249.44	0.11
2		0.03	0.07							0.00	0.18	-248.67	0.07
3		0.04	0.06				0.42			0.00	0.23	-248.60	0.07
<i>Propontocypris</i> spp.													
1	-0.28			0.46						0.04	0.20	-72.47	0.15
2	-0.24									0.07	0.13	-71.74	0.10
3	-0.24			0.49			-2.46			0.02	0.22	-70.76	0.06
<i>Neonesidea</i> spp.													
1	-0.12			-0.25					0.19	0.05	0.28	-149.00	0.09
2	-0.11		0.33	-0.24			-1.13		0.20	0.05	0.31	-147.86	0.05
3			0.21							0.02	0.08	-147.40	0.04
<i>Xestoleberis</i> spp.													
1		0.04	0.15	-0.11		-0.08		0.05		-0.02	0.62	-225.55	0.09
2		0.03	0.15	-0.09		-0.09				-0.02	0.59	-225.29	0.08
3				-0.09		-0.06				0.00	0.52	-224.60	0.06

Table 3.

Term	RI	Coefficient	Lower CI	Upper CI	Term	RI	Coefficient	Lower CI	Upper CI
<i>Pistocythereis bradyi</i>					<i>Loxoconcha epeterseni</i>				
R	1.00	-0.06	-0.15	0.04	R	1.00	0.00	-0.02	0.03
Sal	1.00	2.71	1.53	3.88	D	0.94	0.03	0.00	0.05
D	1.00	-0.12	-0.19	-0.06	Tur	0.76	-0.03	-0.06	0.00
Pb	0.97	-0.21	-0.32	-0.10	Sal	0.53	-0.30	-0.62	0.02
Tur	0.12	0.04	-0.07	0.15	MD	0.38	-0.04	-0.08	0.01
Zn	0.11	-0.05	-0.23	0.13	Chl	0.30	0.02	-0.01	0.04
DO	0.10	0.06	-0.18	0.31	DO	0.26	0.04	-0.02	0.10
WT	0.10	0.33	-0.98	1.64	Pb	0.17	0.01	-0.02	0.05
Chl	0.09	0.02	-0.08	0.11	WT	0.13	-0.13	-0.48	0.22
MD	0.08	-0.01	-0.20	0.18	Zn	0.12	0.01	-0.03	0.04
<i>Bicornucythere bisanensis</i> s.l.					<i>Sinocytheridea impressa</i>				
R	1.00	0.01	-0.11	0.12	R	1.00	0.36	-0.09	0.81
MD	0.63	-0.26	-0.52	0.00	Sal	0.99	-9.26	-14.61	-3.91
D	0.62	-0.08	-0.16	0.00	DO	0.91	-1.42	-2.48	-0.37
Pb	0.38	0.14	-0.05	0.33	MD	0.75	0.82	0.07	1.57
DO	0.20	0.17	-0.15	0.48	Chl	0.74	0.82	0.03	0.70
Chl	0.19	0.06	-0.07	0.19	WT	0.20	0.82	-2.34	8.89
Tur	0.19	0.07	-0.08	0.22	Zn	0.13	0.82	-0.68	0.51
Zn	0.16	-0.01	-0.28	0.25	Tur	0.11	0.82	-0.61	0.30
Sal	0.12	-0.02	-1.86	1.82	D	0.11	0.82	-0.29	0.41
WT	0.10	0.06	-1.80	1.92	Pb	0.10	0.82	-0.62	0.46
<i>Nipponocythere bicarinata</i>					<i>Neomonoceratina delicata</i>				
R	1.00	0.02	-0.04	0.09	R	1.00	-0.15	-0.42	0.12
Chl	0.92	-0.08	-0.14	-0.02	Zn	0.94	0.41	0.11	0.70
DO	0.50	-0.14	-0.30	0.02	Sal	0.93	4.06	1.00	7.13
D	0.30	-0.03	-0.08	0.02	Tur	0.84	0.33	0.06	0.59
WT	0.25	-0.03	-0.08	0.02	DO	0.33	0.45	-0.17	1.06
Sal	0.24	0.51	-0.35	1.36	Pb	0.17	-0.18	-0.67	0.30
Tur	0.19	-0.04	-0.11	0.04	MD	0.14	0.25	-0.33	0.84
Zn	0.18	-0.04	-0.12	0.04	D	0.11	-0.01	-0.22	0.20
Pb	0.13	-0.03	-0.11	0.06	WT	0.11	-0.86	-4.32	2.60
MD	0.10	0.00	-0.13	0.13	Chl	0.09	0.00	-0.24	0.23
<i>Spinileberis quadriaculeata</i>					<i>Keijella kloempritsensis</i>				
R		-0.01	-0.10	0.07	R	1.00	-0.02	-0.09	0.04
Tur	0.66	-0.09	-0.17	0.00	D	0.86	0.06	0.02	0.11
Chl	0.54	0.06	0.00	0.13	Pb	0.58	-0.08	-0.17	0.01
D	0.45	-0.05	-0.10	0.01	Sal	0.35	0.68	-0.21	1.58
WT	0.32	-0.76	-1.82	0.31	MD	0.25	0.08	-0.05	0.21
DO	0.19	0.10	-0.10	0.30	Tur	0.24	0.05	-0.03	0.14
MD	0.17	-0.07	-0.23	0.09	Chl	0.22	-0.04	-0.11	0.03
Pb	0.17	-0.05	-0.16	0.07	Zn	0.17	0.02	-0.11	0.14
Zn	0.14	0.03	-0.09	0.15	WT	0.14	0.33	-0.55	1.22
Sal	0.13	0.11	-1.03	1.26	DO	0.11	0.03	-0.14	0.19
<i>Phlyctocythere japonica</i>					<i>Neosinocythere elongata</i>				
R	1.00	-0.01	-0.05	0.03	R	1.00	-0.03	-0.07	0.00
D	0.80	0.03	0.01	0.06	Tur	1.00	0.08	0.04	0.13
Sal	0.42	0.46	-0.10	1.02	MD	0.33	-0.05	-0.13	0.02
Pb	0.21	-0.03	-0.08	0.03	Chl	0.30	-0.02	-0.05	0.01
Zn	0.20	-0.02	-0.07	0.03	WT	0.26	-0.33	-0.85	0.18
WT	0.15	0.22	-0.38	0.81	Zn	0.21	-0.02	-0.06	0.02
DO	0.14	0.03	-0.08	0.15	Pb	0.19	-0.02	-0.07	0.02
Tur	0.12	0.00	-0.05	0.05	D	0.12	0.00	-0.03	0.03
Chl	0.12	0.00	-0.04	0.05	DO	0.12	-0.02	-0.11	0.07
MD	0.11	0.00	-0.09	0.08	Sal	0.10	0.04	-0.40	0.49

Term	RI	Coefficient	Lower CI	Upper CI	Term	RI	Coefficient	Lower CI	Upper CI
<i>Stigmatocythere roesmani</i>					<i>Hemikrithe orientalis</i>				
R	1.00	-0.01	-0.07	0.05	R	1.00	0.00	-0.04	0.03
Sal	0.61	0.62	-0.04	1.27	D	0.77	0.03	0.00	0.05
Pb	0.61	-0.06	-0.12	0.00	WT	0.46	0.43	-0.07	0.93
Tur	0.37	0.04	-0.01	0.10	DO	0.44	0.08	-0.02	0.17
Zn	0.31	-0.05	-0.11	0.02	Sal	0.30	0.37	-0.17	0.90
DO	0.31	0.09	-0.04	0.22	Pb	0.22	-0.03	-0.08	0.02
MD	0.23	-0.06	-0.16	0.04	Tur	0.17	0.02	-0.03	0.07
Chl	0.19	-0.02	-0.07	0.02	Chl	0.14	-0.01	-0.05	0.03
WT	0.19	0.35	-0.33	1.02	Zn	0.11	0.00	-0.05	0.05
D	0.14	-0.01	-0.05	0.04	MD	0.10	-0.01	-0.08	0.07
<i>Hemicytheridea reticulata</i>					<i>Propontocypris</i> spp.				
R	1.00	0.01	-0.02	0.03	R	1.00	0.05	-0.19	0.30
Pb	0.67	-0.04	-0.07	0.00	Chl	0.80	-0.26	-0.47	-0.04
D	0.66	-0.02	-0.04	0.00	MD	0.58	0.50	-0.03	1.03
Chl	0.49	0.03	0.00	0.05	WT	0.29	-2.60	-6.48	1.28
WT	0.24	-0.24	-0.64	0.16	Pb	0.19	-0.18	-0.55	0.20
MD	0.23	-0.03	-0.09	0.02	Sal	0.17	-1.48	-4.88	1.92
Zn	0.17	-0.01	-0.06	0.05	DO	0.15	-0.27	-0.99	0.45
DO	0.13	0.02	-0.05	0.10	Zn	0.14	-0.03	-0.48	0.41
Tur	0.13	-0.01	-0.05	0.03	D	0.13	0.04	-0.17	0.26
Sal	0.11	-0.05	-0.43	0.34	Tur	0.10	0.01	-0.29	0.32
<i>Loxoconcha malayensis</i>					<i>Neonesidea</i> spp.				
R	1.00	0.02	-0.02	0.06	R	1.00	0.02	-0.09	0.13
DO	0.87	-0.14	-0.25	-0.03	DO	0.72	0.30	0.02	0.57
MD	0.85	-0.11	-0.20	-0.02	Zn	0.58	0.15	-0.01	0.31
D	0.48	-0.03	-0.06	0.00	MD	0.46	-0.23	-0.49	0.03
Chl	0.40	0.03	-0.01	0.08	Chl	0.42	-0.09	-0.21	0.02
Pb	0.29	-0.04	-0.10	0.02	WT	0.32	-1.13	-2.69	0.42
WT	0.26	0.41	-0.21	1.03	Tur	0.24	-0.09	-0.22	0.05
Zn	0.20	-0.03	-0.10	0.04	D	0.23	0.05	-0.04	0.15
Sal	0.15	-0.26	-0.91	0.39	Pb	0.17	0.05	-0.13	0.24
Tur	0.11	-0.01	-0.07	0.05	Sal	0.15	-0.60	-2.20	1.01
<i>Alocopocythere goujoni</i>					<i>Xestoleberis</i> spp.				
R	1.00	-0.03	-0.05	-0.02	R	1.00	-0.01	-0.06	0.05
Sal	0.78	0.22	0.03	0.41	Tur	0.77	-0.07	-0.12	-0.01
D	0.46	-0.01	-0.02	0.00	MD	0.66	-0.10	-0.19	0.00
DO	0.43	-0.03	-0.07	0.01	DO	0.63	0.13	0.00	0.26
Pb	0.31	-0.01	-0.03	0.01	D	0.42	0.03	-0.01	0.07
Zn	0.29	-0.01	-0.03	0.01	Pb	0.33	0.04	-0.02	0.11
WT	0.24	0.13	-0.08	0.33	Sal	0.27	0.46	-0.24	1.15
Tur	0.18	0.01	-0.01	0.03	WT	0.19	-0.37	-1.06	0.31
MD	0.14	-0.01	-0.04	0.02	Zn	0.14	0.00	-0.09	0.08
Chl	0.11	0.00	-0.01	0.01	Chl	0.10	0.00	-0.05	0.05

Table 4.

Environmental Variables	GVIF	Df	GVIF ^{1/2Df}
Chl	7.40	1	2.72
D	6.41	1	2.53
DO	4.49	1	2.12
MD	3.72	1	1.93
Sal	9.43	1	3.07
Tur	6.67	1	2.58
Pb	9.98	1	3.16
Zn	10.05	1	3.17
WT	2.39	1	1.55
R	1266.30	9	1.49

Table 5

Taxa	Chl	D	DO	MD	Sal	Tur	WT	Pb	Zn	R
<i>Pistocythereis bradyi</i>		-			+			-		
<i>Bicornucythere bisanensis</i> s.l.		-								
<i>Nipponocythere bicarinata</i>	-									
<i>Spinileberis quadriaculeata</i>	+					-				
<i>Phlyctocythere japonica</i>		+								
<i>Loxoconcha epeterseni</i>		+				-				
<i>Sinocytheridea impressa</i>	+		-	+	-					
<i>Neomonoceratina delicata</i>					+	+			+	
<i>Keijella kloempitensis</i>		+								
<i>Neosinocythere elongata</i>						+				
<i>Stigmatocythere roesmani</i>								-		
<i>Hemicytheridea reticulata</i>		-						-		
<i>Loxoconcha malayensis</i>			-	-						
<i>Alocopocythere goujoni</i>					+					
<i>Hemikrithe orientalis</i>		+								
<i>Propontocypris</i> spp.	-									
<i>Neonesidea</i> spp.			+							
<i>Xestoleberis</i> spp.			+	-		-				