



1 **Shifts in organic sulfur cycling and microbiome composition in the red-tide causing**
2 **dinoflagellate *Alexandrium minutum* during a simulated marine heat wave**

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13
14
15 **Abstract**

16
17 The biogenic sulfur compounds dimethylsulfide (DMS), dimethylsulfoniopropionate (DMSP)
18 and dimethylsulfoxide (DMSO) are produced and transformed by diverse populations of
19 marine microorganisms and have substantial physiological, ecological and biogeochemical
20 importance spanning organism to global scales. Understanding the production and
21 transformation dynamics of these compounds under shifting environmental conditions is
22 important for predicting their roles in a changing ocean. Here, we report the physiological and
23 biochemical response of *Alexandrium minutum*, a dinoflagellate with the highest reported
24 intracellular DMSP content, exposed to a 6 day increase in temperature mimicking coastal
25 marine heatwave conditions (+ 4°C and + 12°C). Under mild temperature increases (+ 4°C),
26 *A. minutum* growth was enhanced, with no measurable physiological stress response. However,
27 under an acute increase in temperature (+ 12°C), *A. minutum* growth declined, photosynthetic
28 efficiency (F_v/F_m) was impaired, and enhanced oxidative stress was observed. These
29 physiological responses were accompanied by increased DMS and DMSO concentrations
30 followed by decreased DMSP concentrations. At this higher temperature, we observed a
31 cascading stress response in *A. minutum*, which was initiated 6h after the start of the experiment
32 by a spike in DMS and DMSO concentrations and a rapid decrease in F_v/F_m . This was followed
33 by an increase in reactive oxygen species (ROS) and an abrupt decline in DMS and DMSO on
34 day 2 of the experiment. A subsequent decrease in DMSP coupled with a decline in the growth
35 rate of both *A. minutum* and its associated total bacterial assemblage coincided with a shift in
36 the composition of the *A. minutum* microbiome. Specifically, an increase in the relative
37 abundance of OTUs matching the genus *Oceanicaulis* (17.0%), *Phycisphaeraceae* SM1A02
38 (8.8%) and *Balneola* (4.9%) as well as a decreased relative abundance of *Maribacter* (24.4%),
39 *Marinoscillum* (4.7%) and *Seohaecicola* (2.7%), were primarily responsible for differences in



40 microbiome structure observed between temperature treatments. These shifts in microbiome
41 structure are likely to have been driven by either the changing physiological state of *A. minutum*
42 cells, shifts in biogenic sulfur concentrations, or a combination of both. We suggest that these
43 results point to the significant effect of heatwaves on the physiology, growth and microbiome
44 composition of the red-tide causing dinoflagellate *A. minutum*, as well as potential implications
45 for biogenic sulfur cycling processes and marine DMS emissions.

46

47 **Keywords:** DMS, DMSP, DMSO, oxidative stress, temperature stress

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49



50 1. Introduction

51

52

53 Many marine phytoplankton produce the organic sulfur dimethylsulfoniopropionate (DMSP)
54 (Zhou et al., 2009; Berdalet et al., 2011; Caruana and Malin, 2014), for which it can function as
55 an antioxidant, osmolyte, grazing deterrent and currency in reciprocal chemical exchanges with
56 heterotrophic bacteria (Stefels, 2000; Sunda et al., 2002; Wolfe et al., 1997; Kiene et al., 2000).
57 Phytoplankton-derived DMSP is in fact a major source of sulfur and carbon for marine
58 heterotrophic bacteria (Kiene et al., 2000), which in turn play a major role in the cycling and
59 turnover of organosulfur compounds in the ocean (Todd et al., 2007; Curson et al., 2011). The
60 subsequent cycling of DMSP into other biogenic sulfur molecules including dimethylsulfide
61 (DMS) and dimethylsulfoxide (DMSO) by a suite of microbial transformation pathways (Kiene
62 et al., 2000; Sunda et al., 2002) and physical drivers (Brimblecombe and Shooter, 1986) have
63 important ecological and biogeochemical implications spanning from cellular to global scales
64 (Sunda et al., 2002; Charlson et al., 1987; DeBose et al., 2008; Van Alstyne et al., 2001; Knight,
65 2012; Nevitt et al., 1995).

66

67 Among DMSP-producing phytoplankton, the dinoflagellate *Alexandrium minutum*, has the
68 highest recorded DMSP cell content, with an average concentration of 14.2 pmol cell⁻¹,
69 compared with less than 1 pmol cell⁻¹ in most other dinoflagellates (Caruana and Malin, 2014).
70 Blooms of *A. minutum* occur from the Mediterranean Sea to the South Pacific coast in sea
71 surface waters within temperature ranges of 12°C to 25°C (Laabir et al., 2011). Notably, some
72 strains of *Alexandrium*, including *A. minutum*, produce saxitoxins, which lead to paralytic
73 shellfish poisoning (PSP) and are responsible for the most harmful algal blooms in terms of
74 magnitude, distribution and consequences on human health (Anderson et al., 2012).

75

76 *A. minutum* commonly inhabits shallow coastal and estuarine waters (Anderson, 1998), which
77 are globally experiencing substantial shifts in environmental conditions, including increases in
78 sea surface temperature (SST) associated with climate change (Harley et al., 2006). Although
79 generally less studied than chronic temperature rises associated with global climate change
80 (Frölicher and Laufkötter, 2018), acute temperature increases known as marine heatwaves
81 (MHWs) (Hobday et al., 2016) have recently been demonstrated to be becoming more frequent
82 and persistent as a consequence of climate change (Oliver et al., 2018). Increases in MHW
83 occurrence are anticipated to become particularly frequent within the shallow coastal and



84 estuarine waters, where *A. minutum* blooms occur (Ummenhofer and Meehl, 2017;Anderson,
85 1998).

86

87 Coastal MHW events have recently had dramatic impacts on coastal environments. MHW
88 events in Western Australian (2011) and the Northeast Pacific (2013-2015) resulted in
89 significant ecosystem shifts with increases in novel species at the expenses of others (Frölicher
90 and Laufkötter, 2018). The 2016 MHW that was associated with El Niño Southern Oscillations
91 resulted in the mass coral bleaching of more than 90% of the Great Barrier Reef (Hughes et al.,
92 2017). While it is clear that MHWs can have severe consequences on a variety of systems and
93 organisms, their effects on marine microbes and the biogeochemical processes that they
94 mediate have rarely been investigated (Joint and Smale, 2017).

95

96 While there is evidence that increases in seawater temperature can lead to increased DMSP
97 and/or DMS concentrations in phytoplankton (McLenon and DiTullio, 2012;Sunda et al.,
98 2002), it is not clear how a shift in DMSP net production by phytoplankton under acute
99 temperature stress will alter the composition and function of their associated microbiome and
100 how, in turn, this will influence biogenic sulfur cycling processes within marine habitats. There
101 is therefore a pressing need to understand the physiological and biogeochemical consequences
102 of thermal stress on phytoplankton-bacteria interactions within the context of events such as
103 MHWs. This is particularly important, given that a shift in the composition of the
104 phytoplankton microbiome could potentially dictate atmospheric DMS fluxes depending on
105 whether the bacterial community preferentially cleave or demethylate DMSP (Todd et al.,
106 2007;Kiene et al., 2000).

107

108 The aims of this study were to investigate how acute increases in temperature, such as those
109 associated with MHW events, alter the physiological state and biogenic sulfur cycling
110 dynamics of *A. minutum* and determine how these changes might influence the composition of
111 the *Alexandrium* microbiome. We hypothesized that an abrupt increase in temperature would
112 lead to physiological impairment (Falk et al., 1996;Robison and Warner, 2006;Iglesias-Prieto
113 et al., 1992;Rajadurai et al., 2005) and oxidative stress (Lesser, 2006) in *A. minutum*, leading
114 to an up-regulation of DMSP, DMS and DMSO production (McLenon and DiTullio,
115 2012;Sunda et al., 2002) in this high DMSP producer, which could ultimately lead to a shift in
116 the composition of the *A. minutum* microbiome.



117 2. Methods

118

119 2.1. Culturing and experimental design

120 Cultures of *Alexandrium minutum* (CS-324), isolated from Southern Australian coastal waters
121 (Port River, Adelaide, 11/11/1988, CSIRO, ANACC's collection) were grown in GSe medium
122 at 18°C and 50 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ under a 12:12 light:dark cycle. One month before the start
123 of each experiment, *A. minutum* cultures were acclimated over four generations to 20°C and
124 200 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ (14:10 hour light: dark cycle). Cultures were grown to a cell
125 concentration of $\sim 60,000 \text{ mL}^{-1}$ before cells were inoculated into fresh GSe medium. Six days
126 prior to the start of experiments, 20 L of GSe medium was inoculated with a cell concentration
127 of $1,140 \text{ mL}^{-1}$ (experiment 1, April 2016) and 680 mL^{-1} (experiment 2, June 2016) and aliquots
128 of 500 mL were transferred into 40 individual 750 mL sterile tissue culture flasks. Culture
129 flasks were incubated in four independent water baths (10 flasks in each) and maintained under
130 control conditions of 20°C and 200 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$. Temperature and light control was
131 achieved using circulating water heaters (Julabo, country) and programmable LED lights
132 (Hydra FiftyTwo, Company, Country). All cultures were mixed twice daily to keep cells in
133 suspension by gentle swirling.

134

135 On Day 1 (T_0), five culture flasks from each 20°C water bath were transferred to four new
136 water baths for exposure to experimental treatment temperatures (either 24°C experiment 1; or
137 32°C, experiment 2), so that each control and experimental water bath contained five flasks.
138 One culture flask from each tank was immediately sampled for baseline measurements of:
139 DMS (2 mL), DMSP and DMSO (1 mL) concentrations, photochemical efficiency (3 mL),
140 algal and bacterial cell counts (1 mL), ROS quantification (1 mL) and DNA extraction (~ 470
141 mL). The dissolved DMSP fraction was not determined because preliminary investigations
142 showed that gravity filtration was too time consuming, potentially due to clogging of filters by
143 the large *A. minutum* cells (30 μm diameter), leading to filtration artefacts for DMSP analysis,
144 as have previously been mentioned by Berdalet et al. (2011). At 18:00 on Day 1 (T_6), 12:00 on
145 Day 2 (T_{24}), 12:00 on Day 5 (T_{96}) and 12:00 on Day 6 (T_{120}), one flask from each of the eight
146 water baths was removed from the incubation conditions and sampled as described above.

147

148 2.2. Photosynthetic efficiency measurements

149 Subsamples for measurement of photosynthetic efficiency were dark adapted for 10 min under
150 aluminium foil and transferred to a quartz cuvette for Pulse Amplitude Modulated (PAM)



151 fluorometric analysis using a Water PAM (Walz GmbH, Effeltrich, Germany). Once the base
152 fluorescence (F_0) signal had stabilized (measuring light intensity 3, frequency 2s), a saturating
153 pulse (intensity 12, Width 0.8s) was used to measure the maximum quantum yield (F_v/F_M) of
154 photosystem II (PSII). As base fluorescence is dependent on cell density, the photomultiplier
155 gain was adjusted and recorded to maintain F_0 at a level of 0.2 a.u. before saturating the
156 photosystem. Samples were kept in suspension during measurements via continuous stirring at
157 minimal speed inside the quartz cuvette to avoid cells settling.

158

159 *2.3. Microalgal and bacterial cell counts*

160 Subsamples for bacterial cell counts were stained with SYBR Green at a final concentration of
161 1:10,000 and incubated in the dark for 15 min (Marie et al. 1997). Subsamples for microalgal
162 cell counts and stained subsamples for bacterial cell counts were diluted 1:10 and 1:100
163 respectively into sterile GSe medium prior to analysis with a BD Accuri C6 Flow Cytometer
164 (Becton Dickinson). Phytoplankton cells were discriminated using red auto-fluorescence and
165 side scatter (SSC), whereas bacterial populations were discriminated and quantified using
166 SYBR green fluorescence and SSC.

167

168 *2.4. Reactive oxygen species measurements*

169 The presence of reactive oxygen species (ROS) was detected within cultures using the
170 fluorescent probe 2,7-dichlorodihydrofluorescein-diacetate (CM-H2DCFDA; Molecular
171 Probes), which binds to ROS and other peroxides (Rastogi et al., 2010). The reagent was
172 thawed at room temperature for 10 min and activated using 86.5 μ L of DMSO, with 5 μ L of
173 activated reagent added to each sample (final concentration 5 μ M). Samples were vortexed for
174 5 sec and incubated at room temperature for 30 min. Samples were then centrifuged at 2,000 g
175 for 2 min, the supernatant with reagent dye was discarded, and stained cells were resuspended
176 in 1 mL of PBS, prior to quantification of fluorescence by flow cytometry. Mean green
177 fluorescence was quantified from cytograms of forward light scatter (FSC) against green
178 fluorescence. A positive (+ 10 μ L of H_2O_2) and negative (no ROS added) control of PBS were
179 run to ensure that detected cell fluorescence was completely attributable to the ROS probe.

180

181 *2.5. Sulfur analysis by gas chromatography*

182 The preparation of all blanks and samples used in the dilution steps described below were
183 prepared with sterile (0.2 μ M filtered and autoclaved) phosphate-buffered saline (PBS, salinity
184 35ppt) to avoid cell damage from altered osmolarity and to maintain similar physical properties



185 as seawater during headspace analysis by gas chromatography. Aliquots for DMS analysis were
186 transferred into 14 mL headspace vials that were immediately capped and crimped using butyl
187 rubber septa (Sigma Aldrich Pty 27232) and aluminum caps (Sigma Aldrich Pty 27227-U),
188 respectively. DMSP aliquots were 1:1 diluted with sterile PBS and DMSP was cleaved to DMS
189 by adding 1 pellet of NaOH to each vial, which was immediately capped and crimped. Samples
190 were incubated for a minimum of 30 min at room temperature to allow for the alkaline reaction
191 and equilibration to occur prior to analysis by gas chromatography (Kiene and Slezak, 2006).
192 At the end of the experiment, alkaline samples used for DMSP analysis were uncapped and left
193 to vent overnight under a fume hood. On the next day, samples were purged for 10 min with
194 high purity N₂ at an approximate flow rate of 60 mL min⁻¹ to remove any remaining DMS
195 produced from the alkaline treatment. Samples were then neutralized by adding 80 µL of 32 %
196 HCl and DMSO was converted to DMS by adding 350 µL of 12 % TiCl₃ solution to each vial,
197 which was then immediately capped and crimped (Kiene and Gerard, 1994; Deschaseaux et al.,
198 2014b). Vials were then heated in a water bath at 50°C for 1h and cooled down to room
199 temperature prior to analysis by gas chromatography.

200

201 DMS, DMSP and DMSO samples were analyzed by 500 µL direct headspace injections using
202 a Shimadzu Gas Chromatograph (GC-2010 Plus) coupled with a flame photometric detector
203 (FPD) set at 180°C with instrument grade air and hydrogen flow rates set at 60 mL min⁻¹ and
204 40 mL min⁻¹, respectively. DMS was eluted on a capillary column (30 m x 0.32 mm x 5 µm)
205 set at 120°C using high purity Helium (He) as the carrier gas at a constant flow rate of 5 mL
206 min⁻¹ and a split ratio of five. A six-point calibration curve and PBS blanks were run by 500
207 µL direct headspace injections prior to subsampling culture flasks using small volumes of
208 concentrated DMSP.HCl standard solutions (certified reference material WR002, purity 90.3
209 ± 1.8% mass fraction, National Measurement Institute, Sydney, Australia) that were diluted in
210 sterile PBS to a final volume of 2 mL. A 5-point calibration curve was run prior to DMSO
211 analysis using DMSO standard solutions (Sigma Aldrich Pty, D2650) diluted in PBS to a final
212 volume of 2 mL and converted to DMS with TiCl₃ in the same manner as the experimental
213 samples. PBS blanks treated with NaOH and TiCl₃ were also run along with the calibration
214 curves.

215

216 *2.6. DNA extraction*

217 Following sub-sampling for the physiological and biogenic sulfur measurements described
218 above, the remaining 400 mL within each culture flask was filtered onto a 47 mm diameter,



219 0.22 μm polycarbonate filter (Millipore) with a peristaltic pump at a rate of 80 rpm to retain
220 cells for DNA analysis. The filters were subsequently stored in cryovials, snap frozen with
221 liquid nitrogen and stored at -80°C until extraction. DNA extraction was performed using a
222 bead-beating and chemical lysis based DNA extraction kit (PowerWater DNA Isolation Kit,
223 MoBio Laboratories) following the manufacturer's instructions. DNA quantity and purity were
224 checked for each sample using a Nanodrop 2000 (Thermo Fisher Scientific, Wilmington, DE,
225 USA). Three replicate samples with the highest DNA quantity and purity from the control and
226 treatment tanks, collected at the beginning (T_0) and end (T_{120}) of the experiment, were
227 subsequently sequenced.

228

229 *2.7.16S rRNA amplicon sequencing and bioinformatics*

230 To characterize the bacterial assemblage structure (microbiome) of *A. minutum* cultures, we
231 employed 16S rRNA amplicon sequencing. We amplified the V1-V3 variable regions of the
232 16S rRNA gene using the 27F (AGAGTTTGATCMTGGCTCAG, Lane, 1991) and 519R
233 (GWATTACCGCGGCKGCTG, Turner et al., 1999) primer pairing, with amplicons
234 subsequently sequenced using the Illumina MiSeq platform (Ramaciotti Centre for Genomics;
235 Sydney, NSW, Australia) following the manufacturer's guidelines. Raw data files in FASTQ
236 format were deposited in the National Center for Biotechnology Information (NCBI) Sequence
237 Read Archive (SRA) under the study accession number PRJNA486692.

238

239 Bacterial 16S rRNA gene sequencing reads were analysed using the QIIME pipeline (Caporaso
240 et al., 2010; Kuczynski et al., 2012). Briefly, paired-end DNA sequences were joined, de novo
241 Operational Taxonomic Units (OTUs) were defined at 97% sequence identity using UCLUST
242 (Edgar, 2010) and taxonomy was assigned against the SILVA v128 database (Quast et al.,
243 2012; Yilmaz et al., 2013). Chimeric sequences were detected using usearch61 (Edgar, 2010)
244 and together with chloroplast OTUs were filtered from the dataset. Sequences were then
245 aligned, filtered and rarefied to the same depth to remove the effect of sampling effort upon
246 analysis.

247

248 *2.8. Statistical analysis*

249 Repeated measures analysis of variance (rmANOVA) models were fitted to the data to quantify
250 the effects of temperature and time (fixed factors) on all response variables measured in this
251 experiment (cell density, F_v/F_m , ROS, DMS, DMSP and DMSO concentrations) using IBM
252 SPSS Statistics 20. Assumptions of sphericity were tested using Mauchly's test. In cases where



253 this assumption was violated, the degrees of freedom were adjusted using the Greenhouse-
254 Geisser correction factor. Bonferroni adjustments were used for pairwise comparisons. Each
255 variable was tested for the assumption of normality and log, ln or sqrt transformations were
256 applied when necessary.

257

258 For sequencing data, alpha diversity parameters of the rarefied sequences and Jackknife
259 Comparison of the weighted sequence data (beta diversity) were calculated in
260 QIIME (Caporaso et al., 2010). A two-way PERMANOVA with Bray-Curtis similarity
261 measurements was performed on abundance data of taxonomic groups that contained more
262 than 1% of total generated OTUs (represent 90.23% of the data) using PAST (Hammer et al.,
263 2008). In addition, PAST was used to perform non-metric multidimensional scaling (nMDS)
264 analysis and isolate the environmental parameters (normalised as follows: $(x - \text{mean}) / \text{stdev}$) that
265 contributed the most to the differences between groups using the Bray-Curtis similarity
266 measure. SIMPER analysis performed with the White *t*-test was used to identify the taxonomic
267 groups that significantly contributed the most to the shift in bacterial composition in *A.*
268 *minutum* cultures over time and between temperature treatments.

269

270 3. Results

271

272 3.1. Algal growth and physiological response

273 *A. minutum* cell abundance exponentially increased over time in both the control (20°C) and
274 24°C temperature treatment, but a significantly faster growth rate ($p = 0.001$, *t*-test) occurred
275 at 24°C ($2.66 \pm 0.01 \text{ d}^{-1}$; average \pm SE) compared to the 20°C control ($2.57 \pm 0.01 \text{ d}^{-1}$), resulting
276 in significantly greater cell abundance at 96h ($p = 0.007$) and 120h ($p < 0.001$) (rmANOVA,
277 **Table 1, Fig. 1a**). On the other hand, the 32°C treatment resulted in decreased growth rates
278 ($2.40 \pm 0.02 \text{ d}^{-1}$ versus $2.58 \pm 0.02 \text{ d}^{-1}$; *t*-test) and significantly lower cell abundance, relative
279 to the control, at all time points from 6h after the start of the experiment ($p \leq 0.03$; rmANOVA,
280 **Table 1, Fig. 1b**). *A. minutum* abundance demonstrated a marked decline on day 5 in the 32°C
281 treatment.

282

283 No significant difference in the maximum quantum yield (F_v/F_m) of *A. minutum* cultures
284 occurred between 20°C and 24°C until 120h after the start of the experiment, where a
285 significantly lower F_v/F_m occurred in the 24°C treatment ($p = 0.01$; rmANOVA, **Table 1, Fig.**
286 **2a**). In contrast, F_v/F_m was significantly lower in *A. minutum* cultures maintained at 32°C



287 compared to the 20°C at all time points from 6h after the start of the experiment ($p \leq 0.01$)
288 (rmANOVA, **Table 1, Fig. 2b**). However, on days 5 and 6, the F_V/F_M of cultures kept at 32°C
289 recovered to values (0.72 ± 0.008) close to those of the control (0.75 ± 0.004) (**Fig. 2B**).

290

291 3.2. Reactive oxygen species (ROS)

292 Significantly lower concentrations of ROS were measured at 24°C than at 20°C at 96h ($p =$
293 0.003) and 120h ($p = 0.03$) (rmANOVA, **Table 1, Fig. 2c**). In contrast, significantly greater
294 concentrations of ROS were measured at 32°C than 20°C at 24h ($p < 0.001$), 96h ($p = 0.001$)
295 and 120h ($p = 0.01$) after the start of the experiment (rmANOVA, **Table 1, Fig. 2d**). In-line
296 with the recovery in measured F_V/F_M , ROS concentrations in cultures kept at 32°C started to
297 decline to values closer to those of the control on days 5 and 6 of the experiment (**Fig. 2d**). A
298 significant negative correlation between F_V/F_M levels and ROS concentrations was observed
299 under the 32°C temperature treatment ($R^2 = 0.623$; $p = 0.02$, $n = 18$; **Fig. 3**).

300

301 3.3. Biogenic sulfur dynamics

302 Cellular concentrations of DMSP, DMS and DMSO ranged from 444 ± 33 to 1681 ± 175 fmol
303 cell^{-1} , from 13 ± 1.02 to 87 ± 5 fmol cell^{-1} and from 9 ± 1.41 to 94 ± 24 fmol cell^{-1} , respectively,
304 over both experiments (**Fig. 4**). Concentrations of all three sulfur compounds slowly decreased
305 over time in all *A. minutum* cultures regardless of the temperature treatment. No significant
306 difference in DMSP concentration was recorded between 20°C and 24°C throughout
307 experiment 1 ($p > 0.05$; rmANOVA, **Table 1, Fig. 4a**), whereas significantly less DMSP was
308 measured in cells at 32°C than in the 20°C control at 96h ($p = 0.02$; rmANOVA, **Table 1, Fig.**
309 **4b**).

310

311 Significantly lower DMS concentrations were measured at 24°C compared to 20°C at 24h (p
312 < 0.001) and 120h ($p = 0.002$) (rmANOVA, **Table 1, Fig. 4c**). In contrast, DMS was
313 significantly higher at 32°C than 20°C 6h after the start of the experiment ($p = 0.008$;
314 rmANOVA, **Table 1, Fig. 4d**). A similar pattern was observed for DMSO, where relative to
315 the controls, it was significantly lower at 24°C 24h after the start of the experiment ($p = 0.001$;
316 rmANOVA, **Table 1, Fig. 4e**) and significantly greater at 32°C after 6h and 24h ($p < 0.05$, **Fig.**
317 **4f**).

318

319

320



321 3.4. Bacterial abundance and composition

322 Bacterial cell abundance exponentially increased over time at both 20°C and 24°C (**Fig. 5a**).
323 Bacterial abundance was significantly greater at 24°C than at 20°C 120 h after the start of the
324 experiment ($p = 0.05$; rmANOVA, **Table 1, Fig. 5a**). However, no significant difference ($p >$
325 0.05 , t -test) in bacterial growth rate was observed between 20°C ($4.15 \pm 0.05 \text{ d}^{-1}$) and 24°C
326 ($4.18 \pm 0.01 \text{ d}^{-1}$). In contrast, bacterial growth rate was significantly lower at 32°C than at the
327 20°C control ($4.05 \pm 0.01 \text{ d}^{-1}$ versus $4.23 \pm 0.02 \text{ d}^{-1}$; $p < 0.001$, t -test) (**Fig. 5b**), resulting in
328 significantly lower bacterial cell densities at 24h ($p = 0.002$), 96h ($p = 0.002$) and 120h ($p <$
329 0.001) relative to the control (rmANOVA, **Table 1, Fig. 5b**).

330

331 The composition of the initial (T_0) *A. minutum* microbiome was consistent across all samples,
332 but then diverged significantly with time and between temperature treatments (**Fig. 6a-b**; Bray-
333 Curtis similarity measurement, Shepard plot stress = 0.0587). A significant temporal shift in
334 bacterial composition occurred at both 20°C and 32°C, with dissimilarities in community
335 composition between T_0 and T_{120} of 27% and 42% occurring respectively (SIMPER analysis).
336 Notably, bacterial communities at 32°C differed significantly (two-way PERMANOVA; $p <$
337 0.05) to 20°C at T_{120} , with 32% dissimilarity in community composition. These differences
338 were primarily driven by increased relative abundance of bacterial Operational Taxonomic
339 units (OTUs) within the *Oceanicaulis* (17%), *Phycisphaeraceae SM1A02* (8.8%) and *Balneola*
340 (4.9%) genus along with a decline in the relative abundance of OTUs matching *Maribacter*
341 (24%), *Marinoscillum* (4.7%) and *Seohaecicola* (2.7%) (*Rhodobacter* family) in the 32°C
342 treatment (White test, **Fig. 6c**), with all taxa cumulatively contributing to 63% of the OTU
343 differences between temperature treatments at T_{120} (SIMPER analysis). In the 32°C treatment,
344 differences in microbiome composition between T_0 and T_{120} were driven by the elevated levels
345 of ROS, while in the control (20°C) the community shift was principally driven by differences
346 in bacterial and algal cell abundance (**Fig. 6a**; MDS analysis). Similarly, the elevated
347 concentration of ROS as well as the lower FV/FM, lower algal and bacterial cell abundance
348 and lower DMSP, DMS and DMSO concentrations in the 32°C drove the differences in
349 microbiome composition between the temperature treatments (**Fig. 6b**; MDS analysis)

350

351 4. Discussion

352

353 Climate change induced shifts within marine ecosystems are predicted to fundamentally alter
354 the physiology of planktonic organisms and the biogeochemical transformations that they
355 mediate (Finkel et al., 2009; Tortell et al., 2008; Hallegraeff, 2010). Rising seawater



356 temperatures are one of the major impacts of climate change on marine ecosystems (Harley et
357 al., 2006), and can be manifested both as long-term gradual increases (IPCC, 2007, 2013) or
358 intense episodic marine heatwaves (Frölicher and Laufkötter, 2018; Hobday et al., 2016).
359 Although less examined to date than chronic temperature increases, MHWs are predicted to
360 become more frequent and severe (Oliver et al., 2018) and have been proposed as a mechanism
361 for triggering toxic algal blooms (Ummenhofer and Meehl, 2017). Against this back-drop of
362 changing environmental conditions, microbial production and cycling of dimethylated sulfur
363 compounds could be particularly relevant because they simultaneously play a role in the stress
364 response of marine phytoplankton (Berdalet et al., 2011; Deschaseaux et al., 2014a; Sunda et
365 al., 2002; Wolfe et al., 2002; Stefels and van Leeuwe, 1998) and have been predicted to have
366 biogeochemical feed-back effects that are relevant for local climatic processes (Charlson et al.,
367 1987).

368

369 This study investigated the biogenic sulfur cycling dynamics of *A. minutum*, and its
370 microbiome, in response to an intense, short-term thermal stress event, akin to the marine heat-
371 wave events occurring with increasing frequency within coastal habitats (Oliver et al., 2018).
372 Indeed, MHWs have been defined as an abrupt increase in temperature of at least 3 to 5°C
373 above climatological average that lasts for at least 3 to 5 days (Hobday et al., 2016). Large
374 increases in temperature of about 8°C above the yearly average led to red-tides of exceptional
375 density in San Francisco Bay (Cloern et al., 2005). While a 12°C increase in temperature
376 constitutes an extreme scenario of MHWs, even for coastal habitats, this experimental
377 temperature was selected with the intention to induce thermal stress in *A. minutum*.

378

379 *4.1. Effects of thermal stress on A. minutum growth, physiology and ROS production*

380

381 A 4°C increase in temperature resulted in faster algal growth and lower oxidative stress,
382 indicating that 24°C was close to a temperature optimum for this strain of *Alexandrium*. This
383 is perhaps not surprising considering that *Alexandrium* species are capable of growing under a
384 wide range of temperatures from 12°C to 25°C (Laabir et al., 2011). In contrast, a 12°C increase
385 in temperature resulted in a rapid and clear cascade of physiological responses, indicative of
386 an acute thermal stress response in *A. minutum*. Overall, *A. minutum* cells exposed to 32°C
387 immediately exhibited slower growth relative to the 20°C control, suggesting that a 12°C
388 increase in temperature rapidly led to either an increase in cell death rate or a decrease in cell
389 division (Rajadurai et al., 2005; Veldhuis et al., 2001). The slower growth rate at 32°C was



390 coupled with a drop in photosynthetic efficiency and an increase in ROS concentrations, which
391 are both common stress responses to thermal stress in marine algae (Lesser, 2006;Falk et al.,
392 1996;Robison and Warner, 2006;Iglesias-Prieto et al., 1992). In fact, these two physiological
393 responses are often interconnected as increased ROS production generally occurs in both the
394 chloroplast and mitochondria of marine algae exposed to thermal stress, causing lipid
395 peroxidation and ultimately leading to a loss in thylakoid membrane integrity (Falk et al., 1996)
396 and a decrease in the quantum yield of PSII (Lesser, 2006). This was reflected in the positive
397 correlation observed between the maximum quantum yield of PSII and ROS concentrations.

398

399 Although photosynthetic efficiency remained impaired and ROS concentrations remained high
400 under 32°C until the end the experiment, both biomarkers of stress started to return to values
401 closer to those of the 20°C control by day 5 and 6 of the experiment. This was most likely at
402 the expense of a decline in algal abundance since slow growth often coincides with concurrent
403 cellular repair and photosystem activity recovery (Robison and Warner, 2006). The differential
404 physiological response between 24°C and 32°C indicates that although cultures of this strain
405 of *A. minutum* appear to be highly resistant to temperature changes, an abrupt increase in
406 temperature of 12°C simulating an extreme marine heatwave led to a prolonged (4 day) stress
407 response. It could also suggest an acclimation period necessary for such an abrupt shift in
408 temperature, especially since recovery (in F_v/F_m and ROS levels) was observed towards the
409 end of the experiment.

410

411 *4.2. Biogenic sulfur cycling as a response to thermal stress in A. minutum*

412

413 Biogenic organic compounds are key compounds in the stress response of phytoplankton, with
414 evidence they can be used in responses to changes in temperature (Van Rijssel and Gieskes,
415 2002;Stefels, 2000). An up-regulation of the biogenic sulfur yield was expected as a stress
416 response to increased temperature in *A. minutum*, through either an increase in cellular DMSP
417 concentrations, or an increase in DMS via the cleavage of DMSP (McLenon and DiTullio,
418 2012;Berdalet et al., 2011;Wolfe et al., 2002;Sunda et al., 2002). No significant change in
419 DMSP concentrations was observed between the control and 24°C treatment, where, as
420 described above, physiological responses converged to indicate that 24°C was in fact a more
421 optimal growth temperature for this organism. This temperature optimum was associated with
422 lower cellular DMS and DMSO concentrations than in the 20°C control. Since algal stress
423 responses often result in increased cellular sulfur concentrations in dinoflagellates (McLenon



424 and DiTullio, 2012; Berdalet et al., 2011), it is perhaps not surprising that DMS and DMSO
425 concentrations decreased under what appear to have been more optimal growth temperature
426 conditions.

427

428 In contrast to the decreases in DMS and DMSO observed at 24°C, exposure to 32°C resulted
429 in spikes in DMS and DMSO 6h after the start of the experiment, which accompanied decreased
430 algal growth and impaired photosystem II. The increases in DMS and DMSO observed in the
431 32°C treatment may have resulted from enhanced DMSP exudation from phytoplankton cells
432 during cell lysis (Simó, 2001), resulting in an increasing pool of dissolved DMSP made readily
433 available to bacteria and phytoplankton DMSP-lyases (Riedel et al., 2015; Alcolombri et al.,
434 2015; Todd et al., 2009; Todd et al., 2007). Indeed, although DMSP-lyases can be present both
435 extracellularly and intracellularly in marine bacteria (Yoch et al., 1997), algal DMSP-lyases
436 seem to be exclusively located extra-cellularly (Stefels and Dijkhuizen, 1996), indicating that
437 DMSP cleavage to DMS is mainly possible when DMSP exudes from phytoplankton cells
438 during lysis (Simó, 2001). However, it is notable that lower DMSP concentrations in the 32°C
439 treatment than in the control only occurred on day 4, whereas the spike in DMS and DMSO
440 were evident at the outset of the experiment (6h). Since this decrease in DMSP at 96h was not
441 coupled with an increase in DMS, this could alternatively be indicative of assimilation of
442 DMSP-sulfur by bacterioplankton for *de novo* protein synthesis (Kiene et al., 2000), with this
443 demethylation pathway often accounting for more than 80% of DMSP turnover in marine
444 surface waters. The spike in DMSO measured 6h after the increase in temperature to 32°C most
445 likely indicated rapid DMS oxidation by ROS under thermal stress (Sunda et al., 2002; Niki et
446 al., 2000). At that time however, we found no evidence for ROS build up in *A. minutum*
447 cultures, possibly because ROS concentrations were kept in check by sufficient DMS synthesis
448 and active DMS-mediated ROS scavenging (Lesser, 2006; Sunda et al., 2002). In contrast, 24h
449 after the start of the experiment, increased ROS coincided with an abrupt decline in DMS and
450 DMSO, perhaps suggestive of serial oxidation via active ROS scavenging of both DMS to
451 DMSO and DMSO to methane sulfinic acid (MSNA) (Sunda et al., 2002).

452

453 The only previous study that has examined sulfur responses to stress exposure in *A. minutum*
454 examined the effect of physical turbulence by shaking *A. minutum* cultures for up to four days
455 (Berdalet et al., 2011). While the authors of that study also observed slower cell growth as a
456 response to stress exposure, in contrast to our study, cellular DMSP concentrations increased
457 by 20%. Here, a drop in DMSP concentration was observed at 96h between the control and



458 temperature treatment. Therefore, even though DMSP concentrations were quantified with a
459 similar approach as in Berdalet et al. (2011) (no filtration of the samples with assuming that
460 particulate DMSP concentrations overrule dissolved DMSP and DMS concentrations), it seems
461 that heat stress and turbulence triggered a dissimilar sulfur response to stress in *A. minutum*.

462

463 Overall, a 12°C increase in temperature led to lower photosynthetic efficiency, increased
464 oxidative stress and slower cell growth in the red-tide mediating dinoflagellate *A. minutum*.
465 This physiological stress response was coupled with a differential biogenic sulfur cycling as
466 shown by spikes in DMS and DMSO as well as lower DMSP concentrations, most likely
467 translating ROS scavenging and DMSP uptake by bacterioplankton, respectively.

468

469 4.3. A shift in *A. minutum* associated-bacteria composition triggered by thermal stress

470

471 In light of DMSP and related biogenic sulfur compounds constituting an important source of
472 carbon and sulfur to phytoplankton-associated bacteria (Kiene et al., 2000), it follows that any
473 shift in biogenic sulfur concentrations could influence the microbiome composition of *A.*
474 *minutum*. Indeed, the most pronounced temporal shift in the composition of the bacterial
475 community associated with *A. minutum* occurred in the 32°C treatment. This shift was
476 primarily characterized by a statistically significant increase in the relative abundance of OTUs
477 classified as members of the *Oceanicaulis*, *Phycisphaeraceae* and *Balneola* and a significant
478 decrease in OTUs classified as members of the *Maribacter*, *Marinoscillum* and *Seohaecicola*.
479 To predict any potential role of these key OTUs in biogenic sulfur cycling processes, we
480 screened the genomes of members of these groups using BLAST for four genes commonly
481 involved in DMSP metabolism: *dmdA*, CP000031.2 (Howard et al., 2006); *dddP*, KP639186
482 (Todd et al., 2009); *tmm*, JN797862 (Chen et al., 2011); and *dsyB*, KT989543 (Kageyama et
483 al., 2018). A BLAST query of the sequences in the NCBI nucleotide collection (nr/nt) database
484 revealed that previously sequenced members of the genera *Maribacter* (taxid:252356, 357
485 sequences), *Oceanicaulis* (taxid:153232, 36 sequences), *Marinoscillum* (taxid:643701, 23
486 sequences), *Seohaecicola* (taxid:481178, 18 sequences) and *Balneola* (taxid:455358, 44
487 sequences) did not possess any homologs of these sulfur cycling genes. While no homologs
488 were found in the genus *SMIA02*, perhaps because very little genomic information is available
489 for this genus. However a close phylogenetic relative to *SMIA02* (99% query cover, 80%
490 identical, E-value = 0.0), and also a member of the *Phycisphaeraceae* family (*P. mikurensis*
491 10266; genbank accession numbers AP012338.1), possessed significant homologues to all four



492 query genes involved in DMSP metabolism: *dmdA* (92% identical, E-value < 0.001), *dddP*
493 (87% identical, E-value = 0.003), *tmm* (82% identical, E-value = 0.002) and *dsyB* (92%
494 identical, E-value < 0.001). It is thus possible that the spike in DMS and DMSO concentrations
495 in the early stage of the 32°C heat treatment was a consequence of (or contributed to) the
496 preferential recruitment of *Phycisphaeraceae* SM1A02.

497

498 Some members of the *Rhodobacter* family such as several members of the *Roseobacter* genus
499 and *Rhodobacter sphaeroides* are known to possess homologues of either or both *dmdA* and
500 *ddd* genes, which are responsible for DMSP demethylation and DMSP-to-DMS cleavage,
501 respectively (Howard et al., 2006;Curson et al., 2008). However, none of the available
502 reference genomes from *Seohaecicola*, a member of the *Rhodobacteraceae*, possessed any
503 homologs of targeted biogenic sulfur cycling. Similarly, members of the *Maribacter*, which
504 was the main contributor to the difference in microbiome structure between the control and
505 thermal stress treatment, are known not to possess DMSP/DMS transformation pathways
506 (Kessler et al., 2018). Hence, the decline of these taxa in the heat stress treatments, where an
507 upshift in biogenic sulfur availability occurred, is perhaps not surprising.

508

509 Ultimately, the quick conversion of DMSP to DMS (Wolfe et al., 2002) and oxidation of DMS
510 to DMSO (Sunda et al., 2002) was potentially caused by (or led to) a shift in microbiome
511 composition towards the preferential growth of sulfur-consuming bacteria (e.g.
512 *Phycisphaeraceae* SM1A02) at the expense of other types of bacteria (e.g. *Seohaecicola*).
513 Alternatively, the observed shifts in microbiome structure may have occurred independently to
514 the biogenic sulfur cycling processes and was instead related to other metabolic shifts in the
515 heat-stressed *A. minutum*. Notably, the temporal shift in bacterial composition under thermal
516 stress was associated with increased cellular ROS at the end of the experiment, indicating a
517 potential link to oxidative stress. However, in light of the phylogenetic patterns discussed
518 above, this correlation could also reflect a secondary correlation driven by a sulfur-related
519 cascade response, whereby an increase in ROS could have led to an up-regulation of DMSP
520 synthesis (McLenon and DiTullio, 2012;Sunda et al., 2002) and DMSP exudation from *A.*
521 *minutum* cells (Simó, 2001).

522

523

524

525



526 **5. Conclusion**

527

528 Abrupt and intense increases in seawater temperatures associated with MHWs are predicted to
529 become more frequent and intense (Oliver et al., 2018) and have the potential to influence the
530 structure of coastal microbial assemblages and the nature of the important biogeochemical
531 processes that they mediate. Here, we hypothesized that an acute increase in temperature,
532 mimicking a coastal MHW, would trigger a physiological and biochemical stress response in
533 the DMSP-producing dinoflagellate *A. minutum*. This response was indeed observed, with
534 evidence for impaired photosynthetic efficiency, oxidative stress, spikes in DMS and DMSO
535 concentrations, a drop in DMSP concentration and a shift in the composition of the *A. minutum*
536 microbiome. These patterns are indicative of a profound shift in the physiological state and
537 biochemical function of an ecologically relevant dinoflagellate under MHW conditions and
538 suggest that MHWs have the potential to not only influence the composition and interactions
539 of coastal microbial food-webs, but re-shape sulfur budgets in coastal waters.



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Author contribution:

ED, KP and JS devised the experimental design. ED and JOB conducted the thermal stress experiments, including sampling and sample analysis. NS and JOB processed sequencing data while ED processed the physiological and sulfur data. ED wrote the manuscript with significant contributions from all co-authors.

Competing interests:

The authors declare that they have no conflict of interest.



Figure captions

Figure 1 – Algal cell abundance in *A. minutum* cultures in experiment 1 (20°C and 24°C) (A) and experiment 2 (20°C and 32°C) (B); average \pm SE, $n = 4$.

Figure 2 – Photosynthetic efficiency (A, B) and reactive oxygen species (ROS) (C, D) in *A. minutum* cultures in experiment 1 (20°C and 24°C) (A, C) and experiment 2 (20°C and 32°C) (B, D); average \pm SE, $n = 4$.

Figure 3 – Correlation between the photosynthetic efficiency and reactive oxygen species (ROS) in *Alexandrium minutum* under the 32°C thermal stress treatment; $n = 18$.

Figure 4 – DMSP (A, B), DMS (C, D) and DMSO (E, F) concentrations in *A. minutum* cultures in experiment 1 (20°C and 24°C) (A, C, E) and experiment 2 (20°C and 32°C) (B, D, F); average \pm SE, $n = 4$.

Figure 5 – Bacterial cell abundance in *A. minutum* cultures in experiment 1 (20°C and 24°C) (A) and experiment 2 (20°C and 32°C) (B); average \pm SE, $n = 4$.

Figure 6 – Multi-dimensional scaling (MDS) of the three phylogenetic groups defined by 16s sequencing of the bacteria population associated with *A. minutum* cultures grown under control conditions (20°C) and acute thermal stress (32°C) at T₀ and T₁₂₀ (**A**) and MDS excluding the T₀ control (**B**). Vectors represent the factors that most likely drove the shift in bacterial composition between groups. The taxonomic groups that significantly contributed to the difference in bacterial composition between T₀ and T₁₂₀ at 32°C ⁽¹⁾, between T₀ and T₁₂₀ at 20°C ⁽²⁾ and between 32°C and 20°C at T₁₂₀ ⁽³⁾ appear in bold next to the heatmap (**C**), with scaling being based on relative abundance.



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Table 1. Output of repeated measures analysis of variance (rmANOVA) for algal (CELLS_A) and bacterial (CELLS_B) cell abundance, photosynthetic efficiency (F_v/F_M), oxidative stress (ROS), dimethylsulfoniopropionate (DMSP), dimethylsulfide (DMS) and dimethylsulfoxide (DMSO) concentrations as a function of temperature (24°C or 32°C) and time. Numbers in bold indicate significant data based on the level of significance $p < 0.05$. df_1 = numerator df; df_2 = denominator df.

Parameters	24°C – mild thermal stress			32°C – mild thermal stress			
	temperature	time	temperature × time	temperature	time	temperature × time	
CELLS _A	<i>F</i>	4.04	335	4.16	27.47	237.62	8.28
	<i>df</i> ₁	1	4	4	1	2.04	2.04
	<i>df</i> ₂	6	24	24	6	12.26	12.26
	<i>p</i>	0.91	< 0.001	0.01	< 0.001	< 0.001	0.005
CELLS _B	<i>F</i>	2.13	52.2	1.35	32.56	199.8	22.26
	<i>df</i> ₁	1	1.29	1.29	1	4	4
	<i>df</i> ₂	6	7.74	7.74	6	24	24
	<i>p</i>	0.2	< 0.001	0.3	0.001	< 0.001	< 0.001
F_v/F_M	<i>F</i>	0.42	33.43	6.90	48.79	12.58	13.11
	<i>df</i> ₁	1	4	4	1	1.19	1.19
	<i>df</i> ₂	6	24	24	5	5.93	5.93
	<i>p</i>	0.54	< 0.001	0.001	0.001	0.01	0.01
ROS	<i>F</i>	37.26	6.30	5.88	33.23	8.85	8.41
	<i>df</i> ₁	1	4	4	1	2.32	2.32
	<i>df</i> ₂	6	24	24	6	13.9	13.9
	<i>p</i>	0.001	0.001	0.002	0.001	0.003	0.003
DMSP	<i>F</i>	0.79	31.16	0.95	3.03	15.18	3.17
	<i>df</i> ₁	1	1.56	1.56	1	4	4
	<i>df</i> ₂	6	9.35	9.35	6	24	24
	<i>p</i>	0.41	< 0.001	0.4	0.13	< 0.001	0.03
DMS	<i>F</i>	51.5	38.73	2.01	5.08	30.77	5.23
	<i>df</i> ₁	1	2.14	2.14	1	4	4
	<i>df</i> ₂	6	12.87	12.87	6	24	24
	<i>p</i>	< 0.001	< 0.001	0.17	0.07	< 0.001	0.004
DMSO	<i>F</i>	36.56	26.64	7.21	4.68	14.74	7.14
	<i>df</i> ₁	1	4	4	1	4	4
	<i>df</i> ₂	6	24	24	6	24	24
	<i>p</i>	0.001	< 0.001	0.001	0.07	< 0.001	0.001



Figure 1

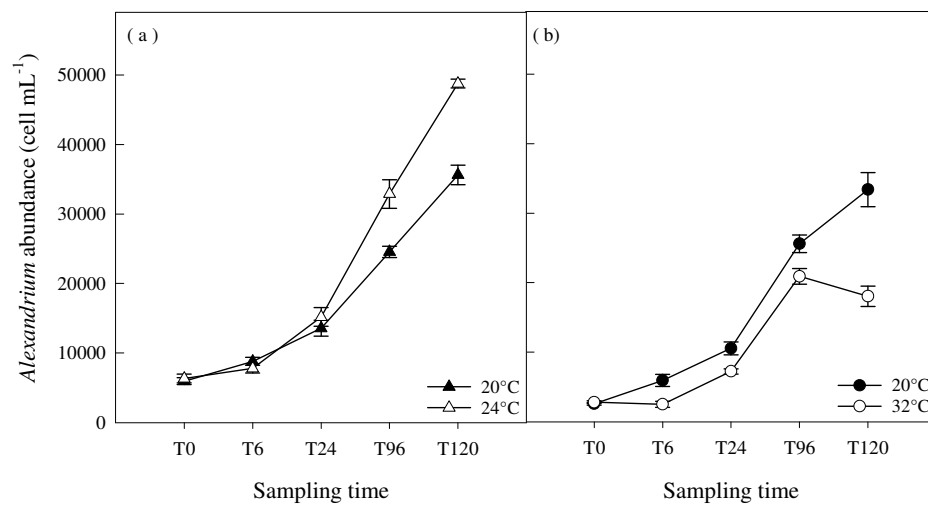




Figure 2

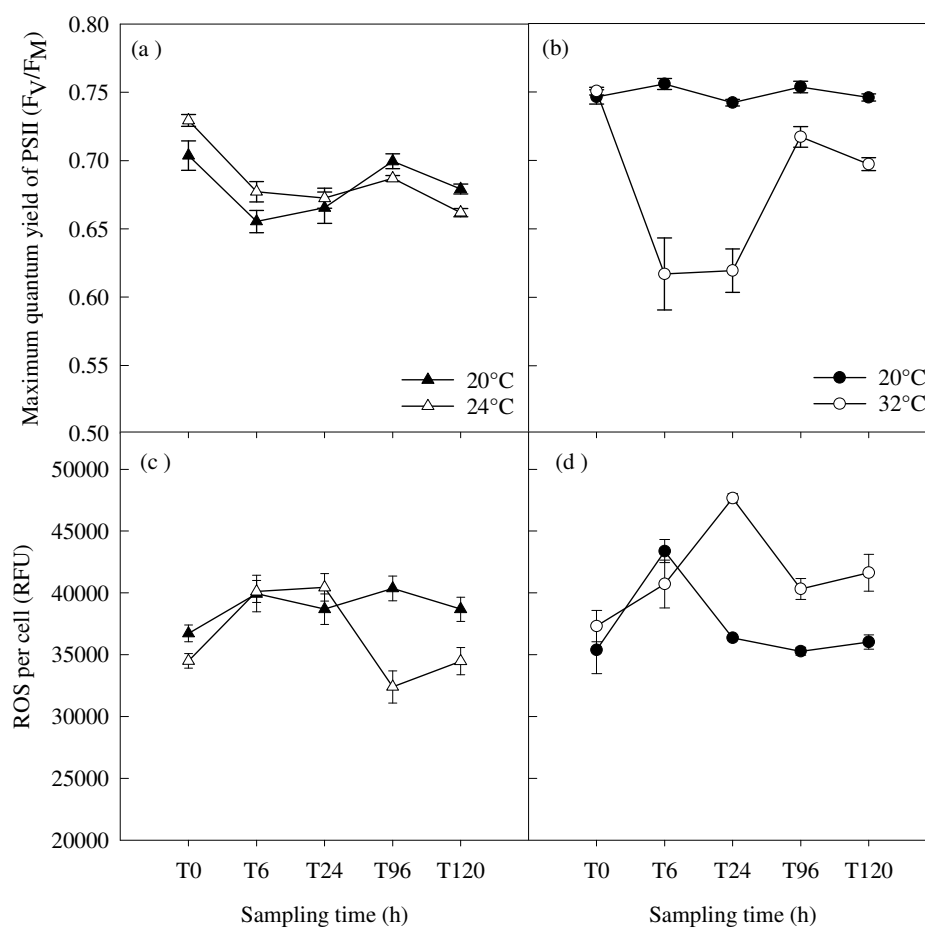




Figure 3

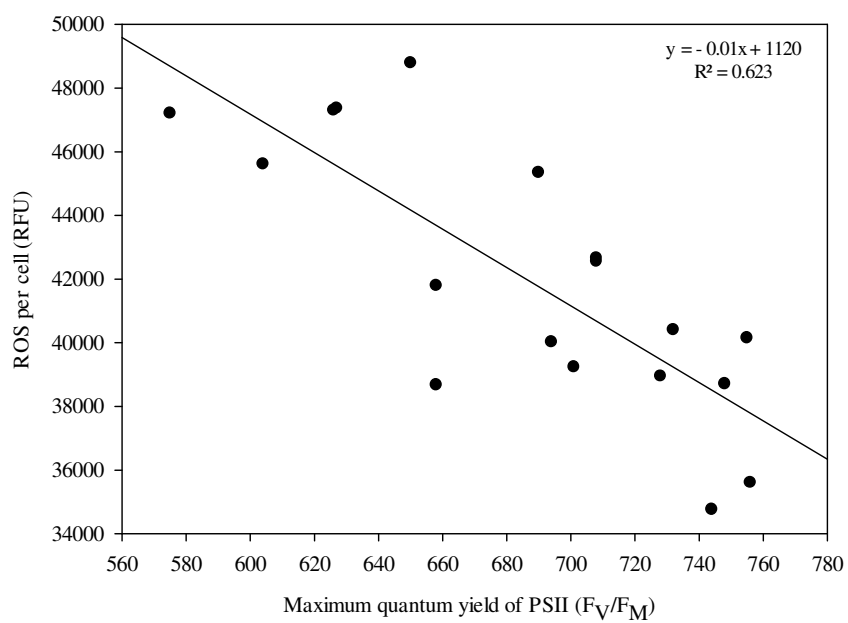




Figure 4

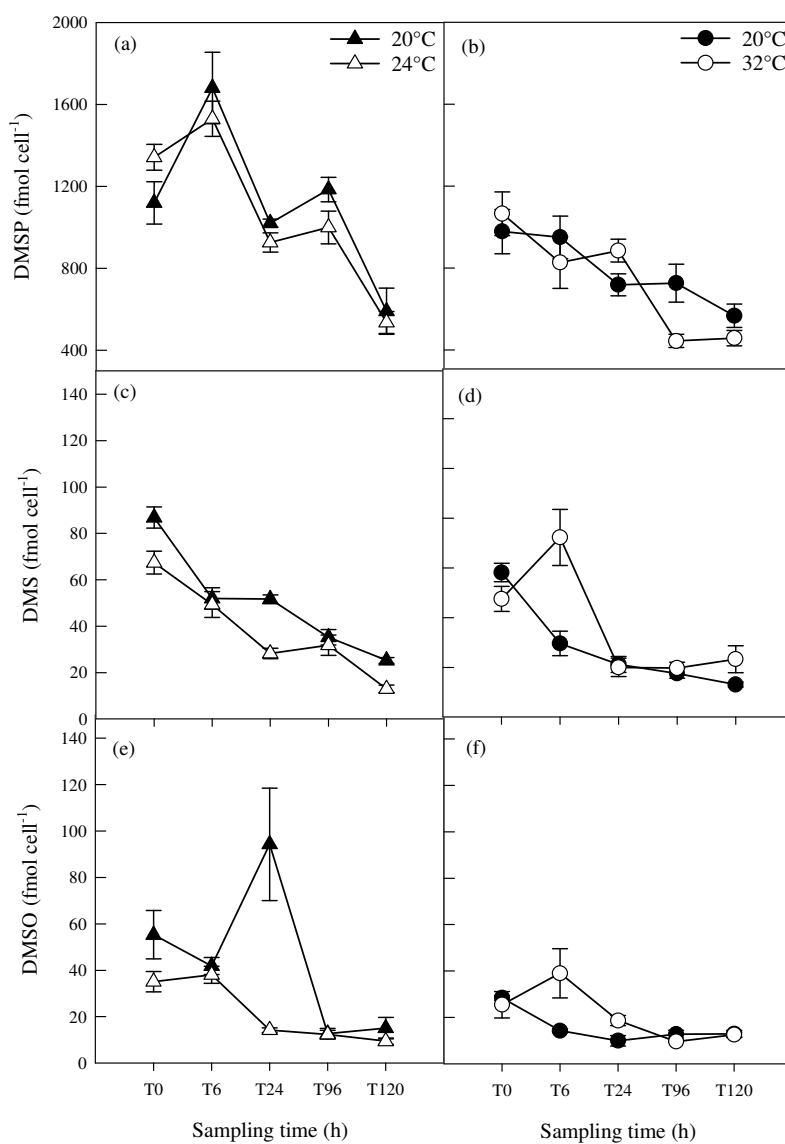




Figure 5

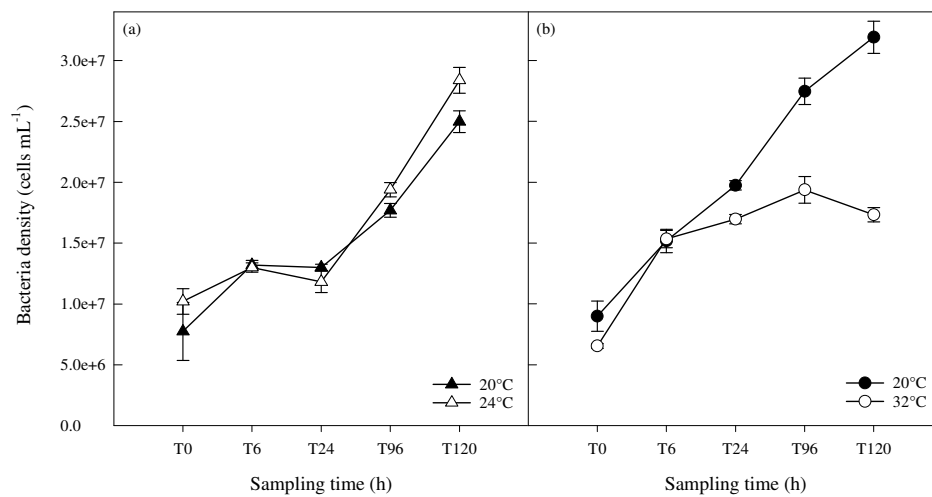




Figure 6

