

Reply to reviewer #1

General comments:

The present paper examined the seasonal and annual variations of dissolved N₂O in a time-series station located in the southwestern Baltic Sea. The results show the coupled variations between the N₂O anomalies, the oxygen concentrations, and nutrients.

The paper presents a valuable new dataset of N₂O and related biogeochemical parameters in a marine region subject to extensive human activities and so nutrients inputs, responsible of the deoxygenation in the Baltic Sea. After the revision, I consider that the manuscript is highly interesting and provide relevant information about processes occurring to the N₂O in the Boknis Eck. The paper is well written and structured, with an appropriate description of the state of the art, objectives are clearly outlined and discussion precisely referenced. The main strength of the paper is the monthly sampling undertaken during twelve years.

However, there are several weaknesses in the paper. First, the authors make the discussion of the results based on data that are included in this study. There are references to data that exist but are not shown. But, in order to discuss about upwelling and hydrographic changes, about algal blooms and ammonium changes, the salinity, temperature, chlorophyll and ammonium data should be included and shown in this paper.

Both reviewers requested to show additional temperature, salinity, chlorophyll *a* and ammonium data. These data were not explicitly shown in our manuscript because they have been published already in Lennartz et al.: Long-term trends at the Boknis Eck time series station (Baltic Sea), 1957–2013: does climate change counteract the decline in eutrophication? *Biogeosciences*, 11, 6323–6339, <https://doi.org/10.5194/bg-11-6323-2014>. However, in order to provide this obvious lacking information, we decided to include the seasonal and annual variations of temperature, salinity, chlorophyll *a* and ammonium in a new supplement to our manuscript.

Secondly, the paper lack of a proper description of the water masses presents at BE and their temporal variability.

Considering that all the measurements were conducted at one fixed location, it is difficult to investigate the water masses based on our data. However, the hydrological condition at BE are not complicated. As is written in Lines 86-87, there is no pronounced river input, and saline water from the North Sea plays a dominant role. By showing the seasonal and annual variations of temperature, salinity, dissolved oxygen and other parameters, readers gain a comprehensive idea about the hydrographic conditions at BE during 2005-2017.

We rewrite the lines 86-87 which read now:

There is no significant river runoff to Eckernförde Bay. Hence, the hydrographical conditions are mainly dominated by saline water input from the North Sea and less saline water from the Baltic Proper, which is typical for that region.

Specific comments:

-Lines 129-130: How did you shifted the data to the 15th? Include procedure and assumptions in the text.

Since the time of sampling varied every month (usually 20-40 days interval), it would be easier for comparison and data analysis if all the samples were collected with a regular spacing. In this case, we ignored the slight time difference and assumed that all of the samples were collected on the same day every month.

We have replaced lines 129-130 with the following sentence: Sampling time varied for every month (usually 20-40 day interval), but for the statistical analysis, data was assumed to be regularly spaced as differences on weekly scales were minor.

Line 170: Could you explicitly explain in the text how did you computed Sc , instead just give the reference? What is the equation for computing Sc ?

Sc was computed as:

$$Sc = \nu / D_{N_2O}$$

$$D_{N_2O} = 3.16 \times 10^{-6} e^{-18370/RT}$$

where ν is the kinematic viscosity of seawater, which is calculated from the empirical equations given in Siedler and Peters (1986), and D_{N_2O} is the diffusion coefficient of N_2O in seawater. R is the universal gas constant and T is the water temperature in K. We used the D_{N_2O} from Rhee (2000).

We have incorporated this part into section 2.4.

Lines 176-184: The comparison of the range of concentrations found between Boknis and other time-series would be better move later in the text, since the reader at this point does not have enough information about the causes that differentiate it from other time-series. The authors should better discuss not only the different magnitudes of the N_2O concentrations, but also the site-specific processes responsible of such differences.

Thank you for the suggestion. The purpose of the comparison is just to give a general idea about the values of the few time-series N_2O measurement published so far, because it might be different from the normal cruises which only last for days or months. A comparison of N_2O concentrations between different time-series analysis is not the major topic of the manuscript, and a discussion about the site-specific processes requires detailed information on the environmental variability of the time-series stations, which does not fit in the scope of the manuscript. In this case, we keep this part in section 3.1.

Lines 207...: In case there is additional information in the BTS, such as chlorophyll, during the study period, show the data in figures instead to refer to previous studies.

Chlorophyll *a* data were added and are now shown in a new supplement.

Lines 235: are there NH_4 data available at the study site during the study period? In that case, it would be better to show them for the discussion instead to appeal to a reference NH_4^+ data were added and are now shown in a new supplement.

Lines 238-239: “Denitrification is inhibited by the presence of O_2 and thus nitrification is presumably responsible for the high N_2O concentrations in winter/early spring.” This statement is not correct at all. The production of N_2O by denitrification can occur at suboxic and hypoxic environments. Please, modify this sentence.

Thank you for pointing out the problem. We modified the text to “Denitrification is inhibited by the presence of high concentrations of dissolved O₂ (> 20 μmol L⁻¹) and...”

Line 239-240: The authors should normalized the N₂O and pH to a constant temperature. Otherwise, temperature changes can be the responsible of this relationship because of thermodynamics changes and not necessarily due to nitrification. In fact, it is not as clear the positive correlation between the N₂O and pH in figure 5, since for pH higher than 7.6, there is no apparent trend between N₂O and pH. The relationship between pH and N₂O obtained during incubations experiments described by Rees et al. (2016) cannot be directly compare to this study, since the experimental conditions and approaches are completely different. The authors should rewrite the entire paragraph.

We need to delete this part from the manuscript because after double-checking the data, we realize that some of the pH values were not calibrated properly. After re-calibration, the relationship between N₂O and pH no long exists. We are very sorry about the mistake.

Lines 263-269: Again, the temperature salinity and Chla information at Boknis are mentioned in the text, but data are missed. If data for these parameters exist, the authors should include in the manuscript. It would reinforce some of the statement that now could look only speculative.

Temperature, salinity and Chlorophyll *a* data were added and are now shown in a new supplement.

Lines 287-288: “Although the observed temperatures and salinities during October 2016–April 2017 were comparable to other years,..”. Please, show temperature and salinity.

Data is now shown in the supplement (see replies above).

Lines 295-296: “Considering the classical view of N₂O consumption via denitrification under hypoxic and anoxic conditions”. This is contrary to the statement done at lines 238-239. Consider to rewrite the first one.

Thank you pointing this out. We have revised the first one as suggested.

Lines 304-306; 308-309: The authors should make use of temperature, salinity or density to show changes in water masses.

Temperature and salinity data can be found in the supplement. Mixed layer variations can be seen in Fig. 4. As is mentioned above, it is difficult to show the changes in water masses. We suggested that the low-N₂O water was a result of advection because vertical exchange can be excluded. However, we do not have any evidence since we did not measure dissolved N₂O from adjacent waters.

Lines 313: Instead of “presence” it would be more correct “concentration/level”

We have revised it as suggested.

Lines 320: “We did not observe an exceptional spring algae bloom in 2017”. Please, consider to include Chla or POM to support this statement.

Chlorophyll a data is now shown in the supplement (see replies above). Unfortunately, POM data are not available. Secchi depth, a proxy of water transparency, is slightly lower in March 2017 than the average value. This could be used to support the statement.

We modified the text to “Secchi depth, a proxy of water transparency, was 3.8 m in March 2017, which is only slightly lower compared to the monthly average value for March (4.5 ± 1.8 m). There was no exceptional spring algae bloom and thus we infer that assimilative uptake of nutrients by phytoplankton was not responsible for the low nutrients concentrations.”

Lines 319: Why can not be shown the Chla data?

Chlorophyll a data is now shown in the supplement.

Lines 331-335: The author should also discuss the potential dependence of rates on temperature and its impact on the seasonal variations of N₂O production/consumption trough the text.

Unfortunately, we did not measure N₂O production/consumption rates at BE. A discussion about the potential temperature dependence of rates is, thus, too speculative. Besides, there is no significant temperature anomaly during the low-N₂O-event. In this case, we suggest that the impact of temperature on the low-N₂O event could be excluded.

Lines 356-357: Please, consider to support this statement with the salinity data 371-373. Please, show the density (or temperature and salinity) record to track the upwelling event in autumn 2017. Lines 377-378: Please, show the chlorophyll data. Lines 385_386: Please, show the ammonium data.

The data are now shown in the supplement.

Lines 394-399: This is a very speculative paragraph as it is written. Could you give any evidence for these potential explanations of the homogeneous distribution of N₂O?

Although the oxic/anoxic interface, where enhanced N₂O production occurs, lasts for several months, the high N₂O concentrations were usually observed only in late autumn (Fig. 4). We agree this paragraph is speculative. There are just some “potential explanations” for why there is no enhanced N₂O in early autumn. Unfortunately, we do not have any further evidence to support the conjecture.

-Section 3.5: The author should evaluate in the results the impact of the dissolved gas analysis uncertainty in the air-sea flux computation and the uncertainty introduced in the net seasonal and annual air-sea NO fluxes.

The uncertainty of flux density, which is mainly derived from K_{N_2O} , with a minor contribution from the error of trace gas analysis, was estimated to be 20% (Wanninkhof, 2014). The average flux density at BE was $3.5 \pm 12.4 \mu\text{mol m}^{-2} \text{d}^{-1}$. With a large uncertainty in the flux density, it is difficult to compute meaningful seasonal/annual fluxes. In this case, we only discussed the variation of flux density in section 3.5.

We have added the uncertainty of flux density in section 3.5.

Lines 416-424: The authors show that N₂O concentration change seasonally, but the saturation stay almost constant. So, how can the author affirm that emissions are controlled by temperature?

There is a seasonality in surface N₂O concentrations but not for the N₂O saturation. During the transformation of concentrations into saturations, the effect of temperature on the saturation is more essential than the effect of salinity. In summer when surface N₂O concentrations are low, N₂O saturations are increased by the relative high temperature (because the equilibrium concentration is decreased). In winter the N₂O concentrations are high, but N₂O saturations are decreased because of the high N₂O solubility at low temperature condition. Temperature is “buffering” the variation of saturation and thus affecting N₂O emissions, and this is our point of “a modulating role”.

To clarify this point we rewrite the paragraph:

We found a weak seasonal cycle for surface N₂O concentrations, with high N₂O concentrations occurring in winter/early spring and low concentrations occurring in summer/autumn, but no such cycle for N₂O saturation. The seasonality in concentration but not in saturation could be largely attributed to the effect of temperature on N₂O solubility: In summer when surface N₂O concentrations are low, N₂O saturations are increased by the relative high temperature; and vice versa in winter. Although salinity also affects N₂O solubility, its contribution is negligible compared to temperature. Temperature alleviated the fluctuation of surface N₂O saturation and thus affected the sea-to-air N₂O fluxes. We conclude that temperature plays a modulating role for N₂O emissions.

Lines 476-484: Unless the author do not include salinity and temperature, they should not used them to conclude the hydrographic conditions at Boknis Eck. Further studies about the hydrography at the BE would complete the picture together with the biogeochemical data at the BE time-series station.

We agree that changing hydrographic conditions will affect N₂O cycling at Boknis Eck as well. This, however, will require modeling studies which need to take into account the ongoing environmental changes of temperature, deoxygenation, changing frequency of North Sea water inflow etc. We think that a detailed discussion of potential future projections of the environmental variability of Boknis Eck/Eckernförde Bay is beyond the scope of the manuscript.

Reply to reviewer #2

Quantifying the concentrations and dynamics of dissolved N₂O in seawater is important for understanding the climate change, but conducting measurements of sufficient duration to determine trends over seasonal, interannual, and decadal time frames for any marine ecosystem remains a challenging task. A long-term Time-Series Station like Boknis Eck in the Eckernförde Bay can provide invaluable information for documenting the role of oceans in relation to N₂O, hence this type of study is significant for our scientific understanding. The paper is well-written and clear. I only have few comments/suggestions below:

Lines 95-96: more information is needed on seawater sample collection.

The sampling procedure is described in detail in lines 95-100 on page 3. Thus we do not see a need to revise the text.

Lines 108-109: The N₂O concentration of standard gases should be provided.

We have added this information in the method section.

Line 154-155: : : dry mole fractions of atmospheric N₂O at the time of the sampling. This description is not the fact since atmospheric N₂O at the time of the sampling has not been measured and the monthly average of N₂O data measured at Mace Head was used to N₂Oeq.

Thank you for pointing this out. We modified the sentence. It reads now: “Since the atmospheric N₂O mole fractions were not measured at the BE Time-Series Station, atmospheric dry mole fractions of N₂O were derived from the monthly average of N₂O data at Mace Head, Ireland (AGAGE, <http://agage.mit.edu/>), instead.”

Lines 185-193: NH₄⁺ concentrations should be provided in the text as well as Figure 2.

Both reviewers requested to show additional temperature, salinity, chlorophyll *a* and ammonium data. These data were not explicitly shown in our manuscript because they have been published already in Lennartz et al.: Long-term trends at the Boknis Eck time series station (Baltic Sea), 1957–2013: does climate change counteract the decline in eutrophication? Biogeosciences, 11, 6323–6339, <https://doi.org/10.5194/bg-11-6323-2014>. However, in order to provide this obvious lacking information, we decided to include the seasonal and annual variations of temperature, salinity, chlorophyll *a* and ammonium in a new supplement to our manuscript.

Lines 221-222: The expression caused misunderstanding.

We would like to change it to “The seasonal variations of NO₂⁻ and NO₃⁻ were significantly correlated with each other ([NO₃⁻]=11.59[NO₂⁻]-0.51, R²=0.80, n=72, p<0.0001) and high concentrations were observed for both in winter.”

Lines 239-243: Does N₂O correlate with NH₄⁺?

There is no straightforward relationship between N₂O and NH₄⁺. By the way, we realized that there were some problems with the calibration of pH data and, therefore, this part will be deleted.

Line 462: The year of 2015 should be 2005?

Thank you for pointing out the mistake. It should be 2005, which is the beginning of the N₂O measurement.

Line 476: There is no temperature data provided at all in this study but with a conclusion 'Temperature plays a modulating role for the N₂O emission at the BE Time-Series Station'. I suggest to provide t data in Figure 2 and more t data provided in related discussion in the text

We would like to show temperature data in the supplement. Also will rewrite the relevant text in section 3.5 to explain how temperature modulates N₂O emissions in. See our reply to reviewer #1 as well.

Figure 2: It would be better for the authors to provide the vertical profiles of t, s, density, NH₄⁺ and Chl a.

The data is now shown in the supplement.

Figure 6: The vertical profiles of hydrological parameters, such as t, s and density are needed to help understand the possible influence of physical processes on N₂O distribution as discussed between lines 301 and 309.

Vertical profiles of temperature, salinity, NH₄⁺ and Chl *a* are shown in supplement. Mixed layer variations can be seen in Fig. 4.

Figure 8: title is needed for x- and y-axis at figure b and d

Figure 9: title is needed for x- and y-axis

We have added new titles in the figures.

According to the comments from the reviewers, the following changes are made in the manuscript:

1. Show seasonal and annual variations of temperature, salinity, chlorophyll *a* and ammonium in a new supplement to our manuscript.
2. Rewrite the lines 86-88.
3. Add more details about standard gases in line 109.
4. Rewrite the lines 129-131.
5. Rewrite the lines 156-158.
6. Add more details in *Sc* computation in Section 2.4 (lines 165-171).
7. Rewrite the lines 221-223.
8. Rewrite the lines 238-239.
9. Delete the discussion about the relationship between N₂O and pH in section 3.2, and the corresponding figure was removed as well.
10. Modify the sentence in line 309.
11. Rewrite the lines 316-320.
12. Rewrite the lines 414-422.
13. Add the uncertainty of flux density computation in lines 428-429.
14. Add new titles in Fig. 7 and 8.

Besides the changes mentioned above, we also revised:

1. The computation of $\Delta\text{N}_2\text{O}$ and AOU were removed from section 2.4 because the corresponding discussion was no longer in Results and Discussion.
2. Few typos in the manuscript were corrected.

A multi-year observation of nitrous oxide at the Boknis Eck Time-Series Station in the Eckernförde Bay (southwestern Baltic Sea)

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Abstract. Nitrous oxide (N₂O) is a potent greenhouse gas and it is involved in stratospheric ozone depletion. Its oceanic production is mainly influenced by dissolved nutrient and oxygen (O₂) concentrations in the water column. Here we examined the seasonal and annual variations of dissolved N₂O at the Boknis Eck (BE) Time-Series Station located in Eckernförde Bay (southwestern Baltic Sea). Monthly measurements of N₂O started in July 2005. We found a pronounced seasonal pattern for N₂O with high concentrations (supersaturations) in winter/early spring and low concentrations (undersaturations) in autumn when hypoxic/anoxic conditions prevail. Unusually low N₂O concentrations were observed during October 2016–April 2017, which was presumably a result of prolonged anoxia and the subsequent nutrient deficiency. Unusually high N₂O concentrations were found in November 2017 and this event was linked to the occurrence of upwelling which interrupted N₂O consumption via denitrification and potentially promoted ammonium oxidation (nitrification) at the oxic/anoxic interface. Nutrient concentrations (such as nitrate, nitrite and phosphate) at BE are decreasing since 1980s, but oxygen concentrations in the water column are still decreasing. Our results indicate a close coupling of N₂O anomalies to O₂ concentration, nutrients and stratification. Given the long-term trends of declining nutrient and oxygen concentrations at BE, a decrease in N₂O concentration, and thus emissions, seems likely due to an increasing number of events with low N₂O concentrations.

1. Introduction

Long-term observation with regular measurement intervals can be an effective way to monitor seasonal and interannual variabilities as well as to decipher short- and long-term trends of an ecosystem, which are required to make projections of the future ecosystem development (see e.g. Ducklow et al., 2009). Recently, multi-year time-series measurements of nitrous oxide (N₂O), a potent greenhouse gas and a major threat to ozone depletion (IPCC, 2013; Ravishankara et al., 2009), have been reported from the coastal upwelling areas off central Chile (Farías et al., 2015) and off Goa (Naqvi et al., 2010), in the North Pacific Subtropical Gyre (Wilson et al., 2017), and in Saanich Inlet (Capelle et al., 2018).

N₂O production in the ocean is generally dominated by microbial nitrification ($\text{NH}_4^+ \rightarrow \text{NO}_2^- \rightarrow \text{NO}_3^-$) and denitrification ($\text{NO}_3^- \rightarrow \text{NO}_2^- \rightarrow \text{N}_2\text{O} \rightarrow \text{N}_2$). During bacterial/archaeal nitrification, N₂O is produced as a by-product with enhanced N₂O production under low oxygen (O₂) conditions (e.g. Goreau et al., 1980; Löscher et al., 2012). N₂O is produced as an intermediate during bacterial denitrification (Codispoti et al., 2005). N₂O could be further consumed via denitrification to dinitrogen, however, this process is inhibited with the presence of O₂ because of the low O₂ tolerance of the enzyme involved (Bonin et al. 1989). This incomplete pathway is called partial denitrification and can lead to N₂O accumulation (e.g. Naqvi et al., 2000; Farías et al., 2009).

The oceans including coastal areas contribute ~25% of the natural and anthropogenic N₂O emissions (IPCC, 2013), with disproportionately high emissions from coastal and estuarine areas (Bange, 2006). N₂O emissions from coastal areas strongly depend on nitrogen inputs (Seitzinger and Kroeze, 1998; Zhang et al., 2010). The increasing input of nitrogen (i.e. eutrophication) has become a worldwide problem in coastal waters leading to enhanced productivity and severe O₂ depletion caused by enhanced degradation of organic matter (Breitburg et al., 2018; Rabalais et al., 2014). The decline in O₂ concentration (i.e. deoxygenation), either in coastal waters or the open ocean, might result in favorable conditions for N₂O production (Codispoti et al., 2001; Nevison et al., 2003). The results of a model study by Kroeze and Seitzinger (1998) indicated a significant increase of N₂O in European coastal waters for 2050. Moreover, it has been suggested that N₂O production and emissions are very likely to increase in the near future, especially in the shallow suboxic/anoxic coastal systems (Naqvi et al., 2000; Bange, 2006). However, model projections show a net decrease in future global oceanic N₂O emission during the 21st century (Martinez-Rey et al., 2015; Landolfi et al., 2017; Battaglia and Joos, 2018).

The Baltic Sea is a nearly enclosed, marginal sea with a very limited access to the open ocean via the North Sea. The restricted water exchange with the North Sea and extensive human activities, such agriculture, industrial production and sewage discharge in the catchment area led to high inputs of nutrients to the Baltic Sea. As a result, the areas affected by anoxia have been expanding in the deep basins of the central Baltic Sea (Carstensen et al., 2014). In order to control this situation, the Helsinki Commission (HELCOM) was established in 1974 and a series of measures have been taken to prevent anthropogenic nutrient input into the Baltic Sea. Consequently, the nutrient inputs (by riverine loads, direct point-sources and, for nitrogen, atmospheric deposition) to the Baltic Sea are declining (HELCOM, 2018a). However, the number of low O₂ (i.e. hypoxic/anoxic) events in coastal waters of the Baltic Sea is increasing and deoxygenation is still going on (Conley et al., 2011; Lennartz et al., 2014). The deoxygenation in the Baltic Sea can affect the production/consumption of N₂O. Our group has been monitoring dissolved N₂O concentrations at the Boknis Eck Time-Series Station, located in Eckernförde Bay (southwestern Baltic Sea), for more than a decade. In this study, we present monthly measurements of N₂O and biogeochemical parameters such as nutrients and O₂ from July 2005 to December 2017. The major objectives of our study were: 1) to decipher the seasonal

pattern of N₂O distribution in the water column, 2) to identify short-term and long-term trends of the N₂O concentrations, 3) to explore the potential role of nutrients and O₂ for N₂O production/consumption, and 4) to quantify the sea-to-air N₂O flux density at the time-series station.

2. Material and methods

2.1 Study site

Sampling at the Boknis Eck (BE) Time-Series Station (www.bokniseck.de) started on 30 April 1957 and, therefore, it is one of the oldest continuously operated time-series stations in the world. The BE station is located at the entrance of the Eckernförde Bay (54°31' N, 10°02' E, Fig. 1) in the southwestern Baltic Sea. The water depth of the sampling site is 28 m. Various physical, chemical and biological parameters are measured on a monthly basis (Lennartz et al., 2014). There is no significant river runoff to Eckernförde Bay. Hence, the hydrographical conditions are mainly dominated by saline water input from the North Sea and less saline water from the Baltic Proper, which is typical for that region. Seasonal stratification usually starts to develop in April and lasts until October, during which hypoxia or even anoxia (characterized by the presence of hydrogen sulphide, H₂S) sporadically occurs, as a result of restricted vertical water exchange and bacterial decomposition of organic matter in the bottom water (Hansen et al., 1999; Lennartz et al., 2014). Thus, BE is a natural laboratory to study the influence of O₂ variations and anthropogenic nutrient loads on N₂O production/consumption.

2.2 Sample collection and measurement

Monthly sampling of N₂O at the BE Time-Series Station started in July 2005. Triplicate samples were collected from six depths (1, 5, 10, 15, 20 and 25 m). Seawater was drawn from 5 L Niskin bottles into 20 mL brown glass vials after overflow. The vials were sealed with rubber stoppers and aluminum caps. The bubble-free samples were poisoned with 50 µL of a saturated mercury chloride (HgCl₂) solution and then stored in a cool, dark place until measurement. The general storage time before measurements of the N₂O concentrations was less than three months.

The static headspace-equilibrium method was adopted to measure the dissolved N₂O concentrations in the vials. 10 mL helium (99.9999 %, AirLiquide, Düsseldorf, Germany) headspace was created in each vial with a gas-tight glass syringe (VICI Precision Sampling, Baton Rouge, LA). Samples were vibrated with Vortex (G-560E, Scientific Industries Inc., New York, USA) for 20 seconds and then left for at least two hours until equilibrium. 9.5 mL subsample of the headspace was subsequently injected into a GC-ECD (gas chromatograph equipped with the electron capture detector) system (Hewlett-Packard 5890 Series II, Agilent Technologies, Santa Clara, CA, USA), which was calibrated with two standard gas mixtures (N₂O in synthetic air, 320 ppb and 1000 ppb, Deuste-Steininger GmbH, Mühlhausen, Germany and Westfalen AG, Münster, Germany) prior to the measurement. The average precision of the

measurements, calculated as the median standard deviation from triplicate measurements, was 0.4 nM. Triplicates with a standard deviation of >10% were omitted. More details about the N₂O measurement can be found in Kock et al. (2016). Dissolved oxygen (O₂) concentrations were measured by Winkler titrations (Grasshoff et al., 1999). Nutrient concentrations were measured by the Segmented Continuous Flow Analysis (SCFA, Grasshoff et al., 1999). A more detailed summary of the parameters measured and methods applied can be found in Lennartz et al. (2014).

2.3 Times series analysis

A time-series can be decomposed into three main components, i.e. trend, cycle and residual component (Schlittgen and Streitberg, 2001). We used the Mann–Kendall test and wavelet analysis to detect the trend and periodical cycles in the time-series data, respectively. As for the residual component, we highlight unusual high/low N₂O concentrations during 2005-2017 and discuss the potential causes for these events.

2.3.1 Wavelet analysis

In order to decipher periodical cycles of the parameters collected at the BE Time-Series Station, a wavelet analysis method was adopted. Wavelet analysis enables the detection of the period and the temporal occurrence of repeated cycles in time-series data. One of the requirements for wavelet analysis is a regular, continuous time-series. Since there is data missing (maximum 2 months in a row) in the BE time-series, due to terrible weather or the ship's unavailability, missing data was interpolated from the previous and following months. Sampling time varied for every month (usually 20-40 day interval), but for the statistical analysis, data was assumed to be regularly spaced as differences on weekly scales were minor. Considering the band width in both frequency and time domain, a Morlet mother wavelet with a wave number of 6 was chosen (Torrence and Compo, 1998). The mother wavelet was then scaled between the frequency of a half-year cycle and the length of the time-series with a stepsize of 0.25. The wavelet analysis was conducted with the MatLab code by Torrence and Compo [2004]. More information about the method can be found on the website <http://paos.colorado.edu/research/wavelets/>.

2.3.2 Mann–Kendall test

Mann–Kendall test (MKT) is a non-parametric statistical test to assess the significance of monotonic trends for time-series measurements. It tests the null hypothesis that all variables are randomly distributed against the alternative hypothesis that a monotonic trend, either increase or decrease, exists in the time-series on a given significance level α (here $\alpha=0.05$). MKT is flexible for data with missing values and the results are not impacted by the magnitude of extreme values, which makes it a widely used test in hydrology and climatology (e.g. Xu et al., 2003; Yang et al., 2004). However, MKT is sensitive to serial correlation in the time-series. The presence of positive serial correlation would increase the probability of trend detection even though no such trend exists (Kulkarni and von Storch, 1995). In order to avoid this situation, data from 12

months were tested individually. It is assumed that there is no residual effect left from the same month last year, considering that the nitrogen species are rapidly biologically cycled. The Matlab function from Simone (2009) was used for the MKT.

2.4 Calculation of saturation and sea-to-air flux density

N₂O saturations (S_{N_2O} , %) were calculated as:

$$S_{N_2O} = 100 \times N_2O_{obs}/N_2O_{eq} \quad (1)$$

where N_2O_{obs} and N_2O_{eq} (in nM) are the observed and equilibrated N₂O concentrations in seawater, respectively. N_2O_{eq} was computed as a function of surface seawater temperature, in situ salinity (Weiss and Price, 1980) and the dry mole fractions of atmospheric N₂O at the time of the sampling. Since the atmospheric N₂O mole fractions were not measured at the BE Time-Series Station, atmospheric dry mole fractions of N₂O were derived from the monthly average of N₂O data at Mace Head, Ireland (AGAGE, <http://agage.mit.edu/>), instead.

N₂O flux density (F_{N_2O} , in $\mu\text{mol m}^{-2} \text{d}^{-1}$) was calculated as:

$$F_{N_2O} = k_{N_2O} \times (N_2O_{obs} - N_2O_{eq}) \quad (2)$$

where k_{N_2O} (in cm h^{-1}) is the gas transfer velocity calculated with the method given by Nightingale et al. (2000), as a function of the wind speed and the Schmidt number (Sc). The wind speed data were obtained from the Kiel Lighthouse (see: www.geomar.de/service/wetter/), which is approximately 20 km away from the BE Time-Series Station. The wind speed was normalized to 10 m (u_{10}) to calculate k_{N_2O} (Hsu et al., 1994). k_{N_2O} was adjusted by multiplying with $(Sc/600)^{0.5}$, and Sc was computed as:

$$Sc = v/D_{N_2O} \quad (3)$$

$$D_{N_2O} = 3.16 \times 10^{-6} e^{-18370/RT} \quad (4)$$

where v is the kinematic viscosity of seawater, which is calculated from the empirical equations given in Siedler and Peters (1986), and D_{N_2O} is the diffusion coefficient of N₂O in seawater. R is the universal gas constant and T is the water temperature in K.

3. Result and discussion

3.1 Overview

N₂O concentrations at the BE Time-Series Station showed significant temporal and depth-dependent variations from 2005 to 2017 (Fig. 2). N₂O concentrations fluctuated between 1.2 and 37.8 nM, with an overall average of 13.9 ± 4.2 nM. This value was higher than the results from the surface water of Station ALOHA (5.9–7.4 nmol kg⁻¹, average 6.5 ± 0.3 nmol kg⁻¹, Wilson et

al., 2017), which is reasonable considering the weak anthropogenic impact in the North Pacific Subtropical Gyre. The N_2O concentrations at BE were much lower than those measured at the time-series station in the coastal upwelling area off Chile ($2.9\text{--}492\text{ nM}$, average $39.4\pm 29.2\text{ nM}$ in the oxyclines and $37.6\pm 23.3\text{ nM}$ in the bottom waters, Farías et al., 2015) and a quasi-time series station off Goa (Naqvi et al., 2010), where significant N_2O accumulations were observed in subsurface waters at both locations. Our measurements were comparable to the time-series station from Saanich Inlet ($\sim 0.5\text{--}37.4\text{ nM}$, average 14.7 nM , Capelle et al., 2018), a seasonally anoxic fjord which has similar hydrographic conditions as BE.

NO_2^- concentrations fluctuated between below detection limit of $0.1\text{ }\mu\text{M}$ and $1.6\text{ }\mu\text{M}$, with an average of $0.2\pm 0.3\text{ }\mu\text{M}$. NO_3^- concentrations varied from below detection limit of $0.3\text{ }\mu\text{M}$ to $17.9\text{ }\mu\text{M}$, with an average of $2.0\pm 2.8\text{ }\mu\text{M}$. The temporal and spatial distributions of nitrite (NO_2^-) and nitrate (NO_3^-) were similar during 2005–2017. A clear O_2 seasonality can be seen with severe O_2 depletion in the bottom waters during summer and autumn. Anoxia with the presence of H_2S were detected in September/October 2005, September 2007, September/October 2014, and September–November 2016. All of the extremely low N_2O concentrations ($<5\text{ nM}$) were observed in the bottom waters in autumn, coinciding with hypoxia/anoxia, while the high N_2O concentrations ($>20\text{ nM}$) sporadically occurred at different depths either in spring or autumn.

3.2 Seasonal cycle

Significant cycles at different frequencies were detected via wavelet analysis at the BE Time-Series Station during 2005–2017 (Fig. 3). A half-year NO_2^- cycle sporadically occurred in 2007–2009, 2013 and 2015. There is a seasonal NO_2^- variability (at the frequency of 1 year) between 2007 and 2016 (times before 2007 and after 2016 were outside the conic line), except during 2010–2012, when high NO_2^- concentrations were not observed in winter (Fig. 2). A biennial cycle of NO_2^- could be observed as well during 2008–2015. The NO_3^- concentrations were dominated by an annual cycle and a minor half-year cycle. The biennial cycle only occurred in 2008 and 2009. A remarkable seasonal variability of dissolved O_2 prevailed all the time, which is also obvious from the times series data shown in Fig. 2. The annual N_2O cycle became gradually more and more evident until 2014, then declined and reoccurred less intensely in 2016. The periodical cycle was also present at other frequencies, indicated by the broadening of the red area before 2015 in Fig. 2d. For example, a biennial N_2O cycle occurred during 2013–2015.

The half-year cycles of NO_2^- and NO_3^- were probably associated with algae blooms which usually occur in each spring and autumn (Fig. S1 and S2). Since the time between the two blooms differed between years, the cycles were weak and thus not present in every year. Due to the fact that there was no half-year O_2 cycle at all, nutrients apart from O_2 might be the “drivers” of the sporadic half-year N_2O cycle in 2008 and 2015, because N_2O production depends on the concentration of the bioavailable nitrogen compounds (Codispoti et al., 2001).

Generally the wavelet analysis indicated a strong annual cycle for NO_2^- , NO_3^- , dissolved O_2 and N_2O at the BE Time-Series Station, which enabled us to explore the seasonal pattern with annual mean data. Although extreme values were excluded as a result of averaging, the smoothed results generally reflect the seasonality of these parameters. Here, we focus on the annual cycle.

The annual mean vertical distribution of dissolved O_2 , NO_2^- , NO_3^- and N_2O are shown in Fig. 4. Due to the development of stratification, the mixed layer was shallow in summer and deep in late autumn/winter. O_2 depletion was observed in bottom waters from late spring until late autumn. The seasonal variations of NO_2^- and NO_3^- were significantly correlated with each other ($[\text{NO}_3^-]=11.59[\text{NO}_2^-]-0.51$, $R^2=0.80$, $n=72$, $p<0.0001$) and high concentrations were observed for both in winter. Minimum N_2O concentrations were found in the bottom waters during September and October, presumably as a result of consumption during denitrification under anoxic condition (Codispoti et al., 2005). High N_2O concentrations were observed in late spring and late autumn, respectively. In late spring N_2O accumulated in the bottom waters because the stratification prevented mixing of the water column. In late autumn, however, N_2O could be ventilated to the surface and thus emitted to the atmosphere due to the breakdown of the stratification. The high N_2O concentrations could be attributed to enhanced N_2O production via nitrification and/or denitrification within the oxic/anoxic interface (Goreau et al., 1980; Codispoti et al., 1992). Since there is no clear O_2 concentration threshold, N_2O production from both nitrification and the onset of denitrification overlap at oxic/anoxic interface. To this end, direct N_2O production measurements (i.e. nitrification/denitrification rates) are required to decipher which process dominates the formation of the different N_2O maxima.

High N_2O concentrations prevailed all over the water column in winter/early spring. NH_4^+ is released from the sediment into bottom waters due to the degradation of organic matter, especially after the autumn algae bloom (Fig. S1 and S2). The stratification usually completely breaks down at this time of the year and the water column becomes oxygenated. Denitrification is inhibited by the presence of high concentrations of dissolved O_2 ($> 20 \mu\text{mol L}^{-1}$) and thus nitrification is presumably responsible for the high N_2O concentrations in winter/early spring.

3.3 Trend analysis

The MKTs were conducted for the surface (1m) and bottom (25m) N_2O concentrations and saturations of the individual 12 months, respectively. Significant decreasing trends were detected for the concentrations in the bottom waters for February and August (Table 1a), and for the saturations in the surface for September and in the bottom for August and November (Table 1b). These results indicated that some systematical changes in N_2O took place at BE. For example, the significant decrease in N_2O concentration/saturation in August might be associated with the increasing temperature, which reinforces the stratification and accelerates O_2 consumption in the bottom waters (Lennartz et al., 2014). As a result, hypoxia/anoxia starts earlier and thus enables the onset of denitrification to consume N_2O . During most of the months, trends in N_2O concentration and saturation were not significant during 2005–2017.

A significant nutrient decline has been observed at the BE Time-Series Station since the mid-1980s, however, Lennartz et al. (2014) found that bottom O₂ concentrations were still decreasing over the past 60 years. The ongoing oxygen decline was attributed to the temperature-enhanced O₂ consumption in the bottom water (Meier et al., 2018) and a prolongation of the stratification period at the BE Time-Series Station (Lennartz et al., 2014). Please note that the trends in nutrients and O₂ concentrations were detected based on the data collection which lasted for approximately 30 and 60 years, respectively, while the N₂O observations at BE Time-Series Station has lasted for only 12.5 years. Further MKT analysis for nutrients, temperature and oxygen for months with significant trends in N₂O concentrations did not show any significant results ($p>0.05$). The significant trends in N₂O concentrations thus do not seem to be directly related to one of these parameters, and we cannot state a reason for the significant trends of N₂O concentration in February and the N₂O saturation in September and November at this point. Presumably, a longer monitoring period for N₂O is required to detect corresponding trends in N₂O and oxygen or nutrients.

3.4 Extreme events

3.4.1 Low N₂O concentrations during October 2016–April 2017

Besides the low N₂O concentrations occurring in autumn, we observed a band of pronounced low N₂O concentrations which started in October 2016 and lasted until April 2017 (Fig. 5). In this period N₂O concentrations varied between 5.5–13.9 nM, with an average of 8.4 ± 2.0 nM. This is approximately 40% lower than the average N₂O concentration during the entire measurement period 2005–2017. The average N₂O saturation during 2005–2017 was $111\pm 30\%$, while from October 2016 to April 2017, the N₂O saturations were as low as 43–93% (average $62\pm 10\%$).

Undersaturated N₂O waters have been previously reported from the Baltic Sea: Rönner (1983) observed a N₂O surface saturation of 79% in the central Baltic Sea and attributed the undersaturation to upwelling of N₂O-depleted waters. Bange et al. (1998) found a minimum N₂O saturation of 91% in the southern Baltic Sea where the hydrographic conditions were significantly influenced by riverine runoff. Walter et al. (2006) reported a mean N₂O saturation of $79\pm 11\%$ for shallow stations (<30 m) in the southwestern Baltic Sea in October 2003. The low-N₂O event at BE was unusual because the concentrations were much lower than those reported values and it lasted for more than half a year.

Although the observed temperatures and salinities during October 2016–April 2017 were comparable to other years (Fig. S1), it is difficult to evaluate the role of physical mechanism in the low-N₂O event because of insufficient data for water mass exchange at the BE Time-Series Station. Here we mainly focused on the chemical or biological processes. Anoxia events with the presence of H₂S were observed in the bottom waters for three months in a row during September–November 2016. This is an unusual long period and is unprecedented at the BE

Time-Series Station. In December 2016 the stratification did not completely break down. Although the water column was generally oxygenated, bottom O₂ concentrations were the lowest observed during the past ten years. Considering the classical view of N₂O consumption via denitrification under hypoxic and anoxic conditions, we inferred that denitrification accounted for low N₂O concentrations in the bottom layer. However, the question still remains where the low-N₂O-concentration water in the upper layers came from.

In September 2016, low N₂O concentrations were only observed in the bottom waters where the anoxia occurred. However, the situation was different in the following months. During October/November 2016, N₂O concentrations were homogeneously distributed in the water column. Although the stratification gradually started to break down in late autumn, the density gradient was still strong enough to keep the bottom waters at anoxic conditions and prevented the low-N₂O-concentration to reach the surface. Thus we inferred that the unusual low N₂O concentrations in the upper layers (above 20 m) were probably resulting from advection of adjacent waters. Due to the fact that the upper layers were well-mixed and oxygenated, in situ N₂O consumption in the water column could be neglected. We suggest therefore, that the N₂O depleted waters were resulting from consumption of N₂O in bottom waters elsewhere and then they were upwelled and transported to BE. Hence, N₂O consumption via denitrification might have been, directly or indirectly, responsible for the low N₂O concentrations during October–November 2016.

In December 2016, the bottom waters were ventilated with O₂. Although N₂O consumption by denitrification should have been inhibited by the high concentrations of O₂ (Codispoti et al., 2001), the N₂O concentrations did not restore to their normal level under suboxic conditions. Since January 2017, the whole water column was well mixed and oxygenated. Usually a significant nutrient supply could be observed starting in November (Fig. 4) as a result of remineralization and vertical mixing, but the average NO₂⁻ and NO₃⁻ concentrations during November 2016–April 2017 were 0.2 and 1.4 μM, respectively, which was about 50% and 60% lower than in other years. Ammonium (NH₄⁺) and chlorophyll *a* concentrations during this period were comparable to that of other years (Fig. S1). Secchi depth, a proxy of water transparency, was 3.8 m in March 2017, which is only slightly lower compared to the monthly average value for March (4.5±1.8 m). There is no exceptional spring algae bloom and thus we infer that assimilative uptake of nutrients by phytoplankton was not responsible for the low nutrients concentrations. The nutrient deficiency might be attributed to enhanced nitrogen removal processes like denitrification or anammox (Voss et al., 2005; Hietanen et al., 2007; Hannig et al., 2007) during the prolonged period of anoxia in autumn 2016. During the low N₂O event, we found that N₂O concentrations were positively correlated with both NO₂⁻ ([N₂O]=7.02[NO₂⁻]+7.36, R²=0.29, n=24, p<0.01) and NO₃⁻ ([N₂O]=0.80[NO₃⁻]+7.36, R²=0.51, n=24, p<0.0001). These results indicate that the development and maintenance of the low-N₂O-concentration was closely associated with nutrient deficiency. Especially after the breakdown of

the stratification, when denitrification was no longer a significant N₂O sink, nutrients might have become a limiting factor for N₂O production.

In general, the low-N₂O-concentration event during October 2016–April 2017 can be divided into two parts: in the stratified waters during October–November 2016, O₂ played a dominant role and N₂O was consumed via denitrification under anoxic conditions. In the well-mixed water column during December 2016–April 2017, nutrient deficiency seemed to have constrained N₂O production via nitrification under suboxic/oxic conditions.

In recent years a novel biological N₂O consumption pathway, called N₂O fixation, which transforms N₂O into particulate organic nitrogen via its assimilation, has been reported (Farías et al., 2013). This process can take place under extreme environmental conditions even at very low N₂O concentrations. Cornejo et al. (2015) reported that N₂O fixation might play a major role in the coastal zone off central Chile where seasonally occurring surface N₂O undersaturation was observed. The relatively high N₂ fixation rates in the Baltic Sea (Sohm et al., 2011) highlight the potential role of N₂O fixation (Farías et al., 2013). However, we cannot quantify the role of biological N₂O fixation for the N₂O depletion in the Baltic Sea due to the absence of N₂O assimilation measurements.

3.4.2 High N₂O concentrations in November 2017

High N₂O concentrations were observed at the BE Time-Series Station in November 2017. The average value reached 35.4±1.5 nM, which was the highest concentration measured during the entire sampling period from 2005 to 2017. Dissolved N₂O was homogeneously distributed in the water column, but this event did not last long. In December, dissolved N₂O returned to normal levels and the average concentration in the water column was comparable to that of other years. Average N₂O saturation in November 2017 was 322±10%, which was also the highest for the past 12.5 years. This value was much higher than the maximum surface N₂O saturation reported by Rönner (1983) in the central Baltic Sea, but was comparable to the results observed in the southern Baltic Sea (312%, Bange et al., 1998). Bange et al. (1998) linked the enhanced N₂O concentrations to riverine runoff because those samples were collected in an estuarine area, however, the riverine influence around the BE Time-Series Station is negligible. As a result, the impact of fresh water input can be excluded.

Dissolved O₂ seemed to play a dominant role in the high N₂O concentrations. Enhanced N₂O production usually occurred at the oxic/anoxic interface, which was closely linked to the development of water column stratification. In general the breakdown of the stratification is faster than its establishment at the BE Time-Series Station. As a result, it took about half a year for bottom O₂ saturation to gradually decrease from ~80% to almost 0% (i.e. anoxia), but only two months to restore normal saturation level in 2010 (Fig. 6). In late autumn, surface water penetrated into the deep layers via vertical mixing and eroded the oxic/anoxic interface. The entire water column quickly became oxygenated and the enhanced N₂O production was stopped.

Hypoxia/anoxia at BE is usually observed in the bottom waters in autumn, but in September 2017, hypoxic water (O_2 saturation < 20 %, which was close to the criterion for hypoxia, see Naqvi et al., 2010) was found in the subsurface layer (10 m) as well. Surface O_2 saturation was only ~50%, which was the lowest during the sampling period 2005–2017. The density gradient of the water column in September 2017 was much lower than in other years. These results indicate the occurrence of an upwelling event at BE Time-Series Station in autumn 2017, which might be a result of the saline water inflow from the North Sea considering the change of salinity in the water column (Fig. S1). Strong vertical mixing has interrupted the hypoxia/anoxia and bottom O_2 saturation reached ~60% in October 2017. The presence of O_2 prevented N_2O consumption via denitrification, as a result, we did not observe a significant N_2O decline during that period (Fig. 5).

Considering the fact that a significant autumn algae bloom was observed in autumn 2017 (as indicated by high chlorophyll a concentrations, see Fig. S1), severe O_2 depletion in the bottom water could be expected. Although the bottom O_2 saturation was only slightly lower in November than in October, we speculate that even lower O_2 saturation (but not anoxia) might have occurred between October and November. The “W-shaped” O_2 saturation curve (see Fig. 6) suggests that the stratification did not completely break down in October and that there might have been a reestablishment of the oxic/anoxic interface providing favorable conditions for enhanced N_2O production. Due to the degradation of organic nitrogen, NH_4^+ is released from the sediment into bottom waters (Dale et al., 2011), especially in autumn when O_2 is low (Fig. S2). NH_4^+ concentrations in November 2017 were lower than in other years (Fig. S1), and NO_2^- concentrations were higher (Fig. 5), indicating that nitrification occurred in bottom waters. To this end, we suggest that the reestablishment of the oxic/anoxic interface promoted ammonium oxidation (the first step of nitrification). In this case, N_2O could have temporarily accumulated because its consumption via denitrification was blocked. Meanwhile, the relatively low density gradient (i.e. low stratification) allowed upward mixing of the excess N_2O to the surface. However, we inferred that that this phenomenon would only last for a few days due to the rapid breakdown of stratification at the BE Time-Series Station.

Due to the development of the pronounced stratification, the oxic/anoxic interface prevailed in summer/early autumn as well, but we did not observe N_2O accumulation during these months. One of the potential explanations is that enhanced N_2O production only took place within particular depths where strong O_2 gradient existed, but our vertical sampling resolution was too low to capture this event. Also enhanced N_2O production might be covered by the weak mixing which brought low- N_2O water from the bottom to the surface.

The upwelling event played different roles in autumn 2016 and 2017. First, upwelling took place somewhere else but at BE because of the strong density and O_2 gradient in the water column during autumn 2016. Second, bottom water remained anoxic in autumn 2016, while the compensated water for upwelling in 2017 penetrated through stratification and brought O_2 into bottom water (Fig. 6), which caused enhanced N_2O production. Similarly, autumn upwelling was

detected in 2011 and 2012 when we found relatively low O_2 concentrations in subsurface layers (10 m) (Fig. 2), but we did not observe an increase in bottom O_2 concentrations and N_2O concentrations remained low during that time. These upwelling events seem to be driven by saline water inflow considering the prominent increase in salinity, but the mechanism dominates O_2 input into bottom water before the stratification break down remains unclear.

3.5 Flux density

During 2005–2017, surface N_2O saturations at the BE Time-Series Station varied from 56 % to 314 % (69–194 % excluding the extreme values discussed in Sect. 3.4), with an average of 111 ± 30 % (111 ± 20 % without the extreme values). Generally the water column at BE was slightly oversaturated with N_2O . Our results are in good agreement with the estimated mean surface N_2O saturation for the European shelf (113%, Bange, 2006).

We found a weak seasonal cycle for surface N_2O concentrations, with high N_2O concentrations occurring in winter/early spring and low concentrations occurring in summer/autumn, but no such cycle for N_2O saturation (Fig. 4; Fig. 7). The seasonality in concentration but not in saturation could be largely attributed to the effect of temperature on N_2O solubility: In summer when surface N_2O concentrations are low, N_2O saturations are increased by the relative high temperature; and vice versa in winter. Although salinity also affects N_2O solubility, its contribution is negligible compared to temperature. Temperature alleviated the fluctuation of surface N_2O saturation and thus affected the sea-to-air N_2O fluxes. We conclude that temperature plays a modulating role for N_2O emissions.

The wind speed (u_{10}) at the BE Time-Series Station ranged from 1.1 to 14.0 m s^{-1} , with an average of $7.0 \pm 2.7 \text{ m s}^{-1}$. N_2O flux densities varied from -19.0 to $105.7 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$ (-14.1 – $30.3 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$ without the extreme values), with an average of $3.5 \pm 12.4 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$ ($3.3 \pm 6.5 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$ without the extreme values). However, the true emissions might have been underestimated because our monthly sampling resolution is insufficient to capture short-term N_2O accumulation events due to the fast breakdown of stratification in autumn. The uncertainty introduced in the flux density computation was estimated to be 20% (Wanninkhof, 2014). The flux densities at the BE Time-Series Station are comparable to those reported by Bange et al. (1998, 0.4 – $7.1 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$) from the coastal waters of the southern Baltic Sea, but are slightly lower than the average N_2O flux density reported by Rönner (1983, $8.9 \text{ } \mu\text{mol m}^{-2} \text{ d}^{-1}$) from the central Baltic Sea. Please note that the results of Rönner (1983) were obtained only from the summer season and therefore are probably biased because of missing seasonality.

In December 2014, a strong saline water inflow from the North Sea was observed, which was the third strongest ever recorded (Mohrholz et al., 2015). Although the salinity in December 2014 was comparable to other years, a remarkable increase in salinity was observed in the following several months. However, we did not detect a significant N_2O anomaly or enhanced emission during that time. Similarly, Walter et al. (2006) investigated the impact of the North Sea water

inflow on N₂O production in the southern and central Baltic Sea in 2003. The oxygenated water ventilated the deep Baltic Sea and shifted anoxic to oxic condition which led to enhanced N₂O production, but the accumulated N₂O was unlikely to reach the surface due to the presence of a permanent halocline (Walter et al., 2006).

Although we observed extremely high N₂O flux density in November 2017, the low-N₂O-concentration (<10 nM) events have become more and more frequent during the past ten years (Fig. 2). This phenomenon seldom occurred before 2011, but remarkable low N₂O concentrations can be seen in 2011 and 2013, and to a less extent in 2012 and 2014. Similar events lasted for several months in 2015 and for even more than half a year during 2016–2017. The most striking was that the low-N₂O-concentration water was not only detected in bottom waters, but also at surface which would significantly impact the air-sea N₂O flux densities. Although the MKT result did not give a significant trend for the N₂O flux densities, the data presented in Fig. 8 suggest a potential decline of N₂O flux densities from the coastal Baltic Sea, challenging the conventional view that N₂O emissions from coastal waters would most probably increase in the future, which was based on the hypothesis of increasing nutrient loads into coastal waters. Due to an effective reduction of nutrient inputs, the severe eutrophication condition in the Baltic Sea has been alleviated (HELCOM, 2018b), but ongoing deoxygenation points to the fact that it will take a longer time for coastal ecosystems to feedback to reduced nutrient inputs because other environmental changes such as warming may override decreasing eutrophication (Lennartz et al., 2014).

4. Conclusions

The seasonal and inter-annual N₂O variations at the BE Time-Series Station from July 2005 to December 2017 were driven by the prevailing O₂ regime and nutrients availability. We found a pronounced seasonal cycle with low N₂O concentrations (undersaturations) occurring in hypoxic/anoxic bottom waters in autumn and enhanced concentrations (supersaturations) all over the water column in winter/early spring. Significant decreasing trends for N₂O concentrations were found for few months, while most of the year, no significant trend was detectable in the period of 2005–2017. During 2005–2017, no significant trends were present for O₂ and nutrients either, but these parameters all show significant decreasing trends on longer time scales (~60 years) at BE. Our results show the strong coupling of N₂O with O₂ and nutrient concentrations, and suggest similar changes on comparable time scales. Further monitoring of N₂O at BE time series station is thus important to detect changes. Further studies on N₂O production/consumption by nitrification and denitrification and analysis of the characteristic N₂O isotope signature might be very helpful to decipher the potential roles of O₂/nutrients for N₂O cycling.

Temperature plays a modulating role for the N₂O emission at the BE Time-Series Station. Although the hydrographic condition at BE is generally dominated by the inflow of saline North Sea water, this did not affect N₂O production and its emissions to the atmosphere. It seems that

events with extremely low N₂O concentrations and thus reduced N₂O emissions became more frequent in recent years. Our results provide a new perspective onto potential future patterns of N₂O distribution and emissions in coastal areas. Continuous measurement at the BE Time-Series Station with a focus on late autumn would be of great importance for monitoring and understanding the future changes of N₂O concentrations and emissions in the southwestern Baltic Sea.

Data availability

Data are available from the Boknis Eck Database: www.bokniseck.de

Author contribution

X.M., S.T.L. and H.W.B. designed the study and participated in the fieldwork. N₂O measurements and data processing were done by X.M. and S.T.L. X.M. wrote the manuscript with contributions from S.T.L. and H.W.B.

Competing interests

The authors declare that they have no conflict of interest.

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Table 1. The results of the Mann-Kendall test for the surface and bottom N₂O concentrations and saturations of the 12 individual months.

Table 1a. MKT results for N₂O concentrations

Month	January		February		March		April	
Depth/m	1	25	1	25	1	25	1	25
p	0.09	0.19	0.11	0.03(-)	0.19	0.63	0.09	0.30
Month	May		June		July		August	
Depth/m	1	25	1	25	1	25	1	25
p	0.63	0.24	0.15	0.95	0.16	0.16	0.20	0.03(-)
Month	September		October		November		December	
Depth/m	1	25	1	25	1	25	1	25
p	0.25	0.76	0.36	0.76	0.67	0.16	0.10	0.30

Table 1b. MKT results for N₂O saturations

Month	January		February		March		April	
Depth/m	1	25	1	25	1	25	1	25
p	0.37	0.24	0.15	0.15	0.19	0.63	0.11	0.19
Month	May		June		July		August	
Depth/m	1	25	1	25	1	25	1	25
p	0.19	1	0.37	0.54	0.10	0.43	0.20	0.02(-)
Month	September		October		November		December	
Depth/m	1	25	1	25	1	25	1	25
p	0.04(-)	0.85	0.06	0.43	0.20	0.03(-)	0.16	0.36

p indicates the p-value of the test, which is the probability, under the null hypothesis, of obtaining a value of the test statistic as extreme or more extreme than the value computed from the sample.

(-) indicates a rejection of the null hypothesis at α significance level and a decreasing trend is detected.

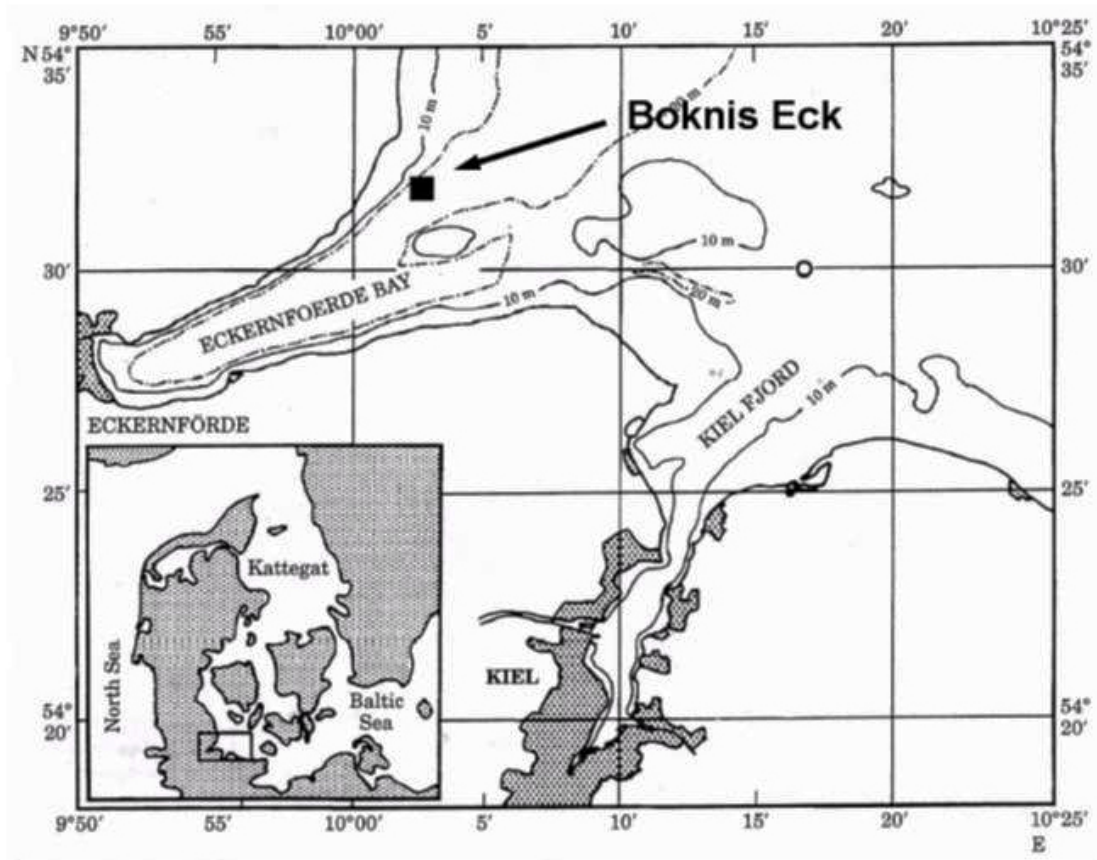


Fig. 1 Location of the Boknis Eck Time-Series Station in the Eckernförde Bay, southwestern Baltic Sea. (Map from Hansen et al., 1999)

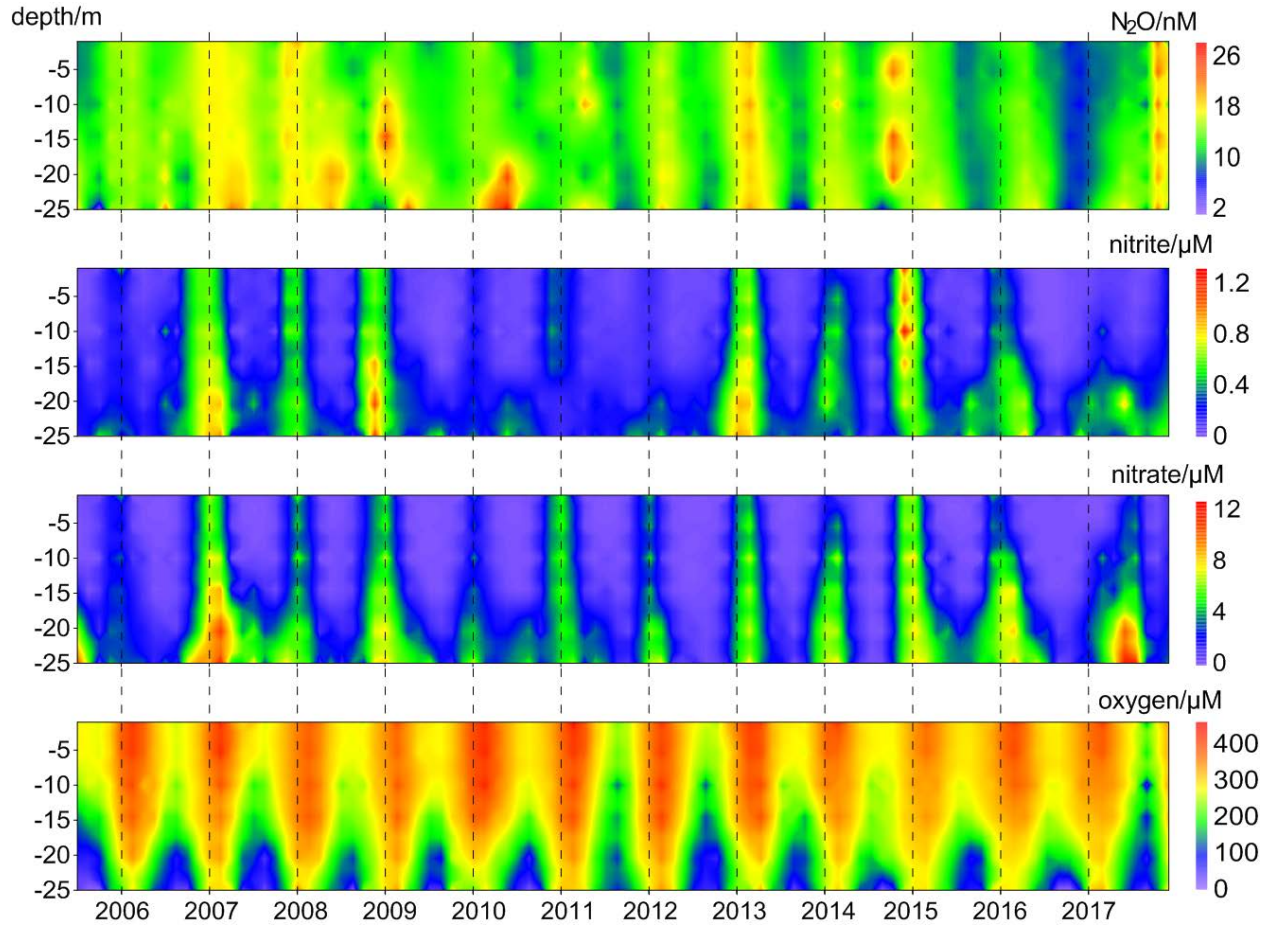


Fig. 2 Vertical distributions of dissolved O_2 , NO_2^- , NO_3^- , and N_2O from the BE Time-Series Station during 2005–2017.

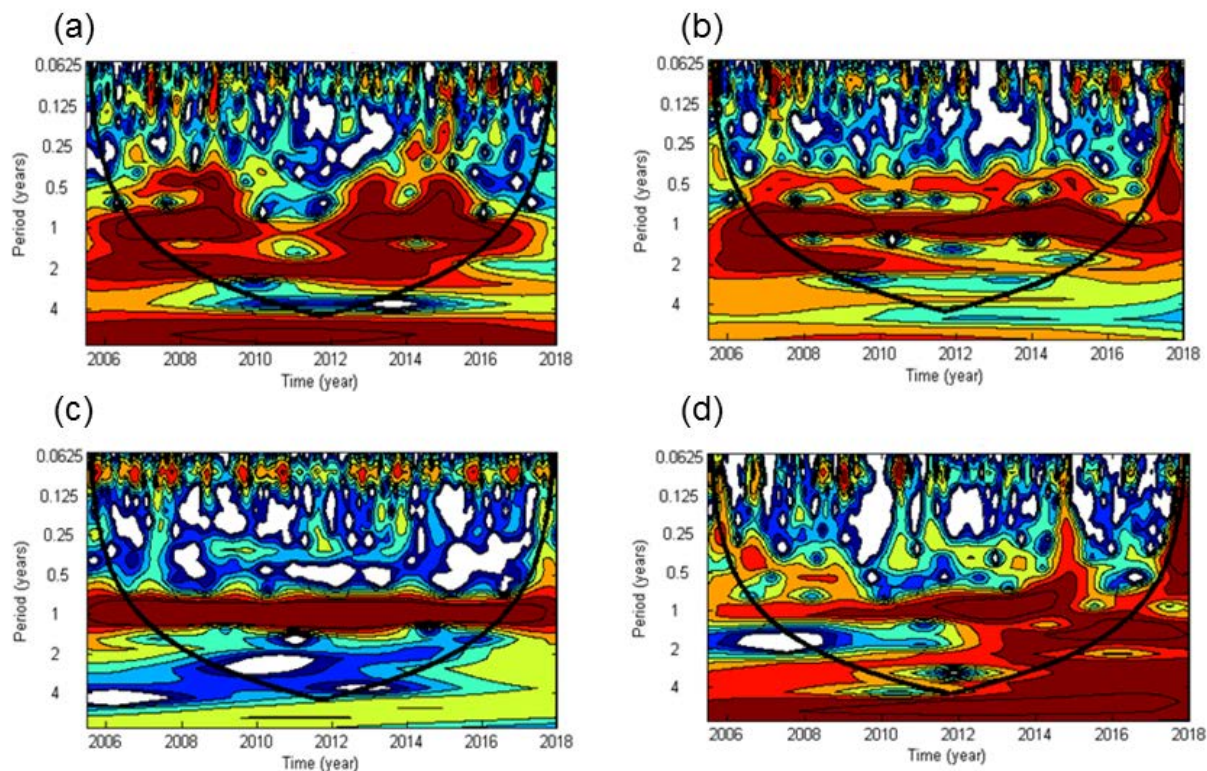


Fig. 3 Wavelet power spectra of NO_2^- (a), NO_3^- (b), dissolved O_2 (c) and N_2O (d) from the BE Time-Series Station. Red areas indicate high, blue indicate low power. The black conic line indicates the significant area where boundary effects can be excluded.

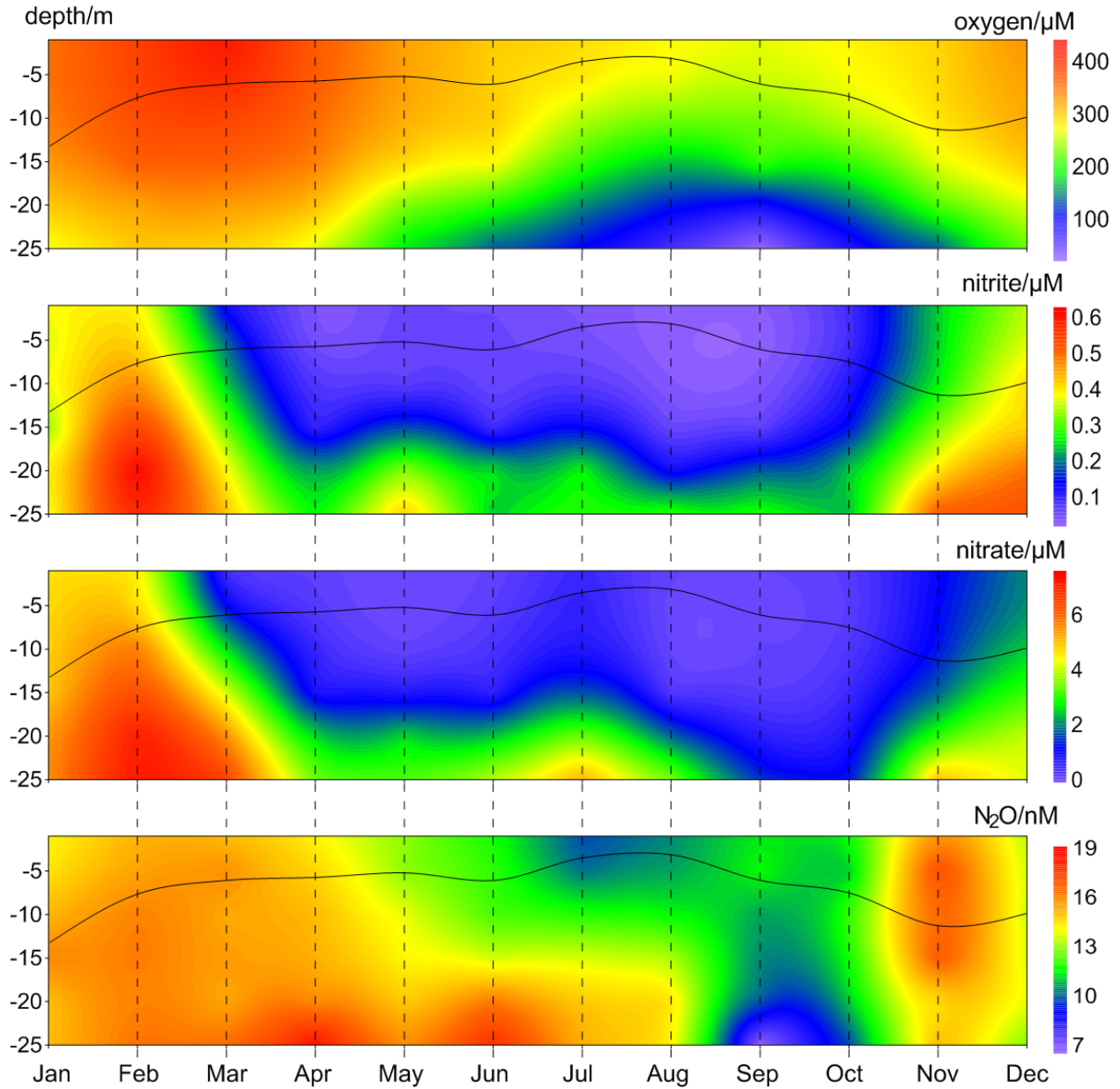


Fig. 4 Average vertical distributions of dissolved O₂, NO₂⁻, NO₃⁻, and N₂O from the BE Time-Series Station during 2005–2017. The black line indicates the mixed layer depth, which was calculated based on a potential density anomaly of 0.15 kg m⁻³ from the sea surface (1m).

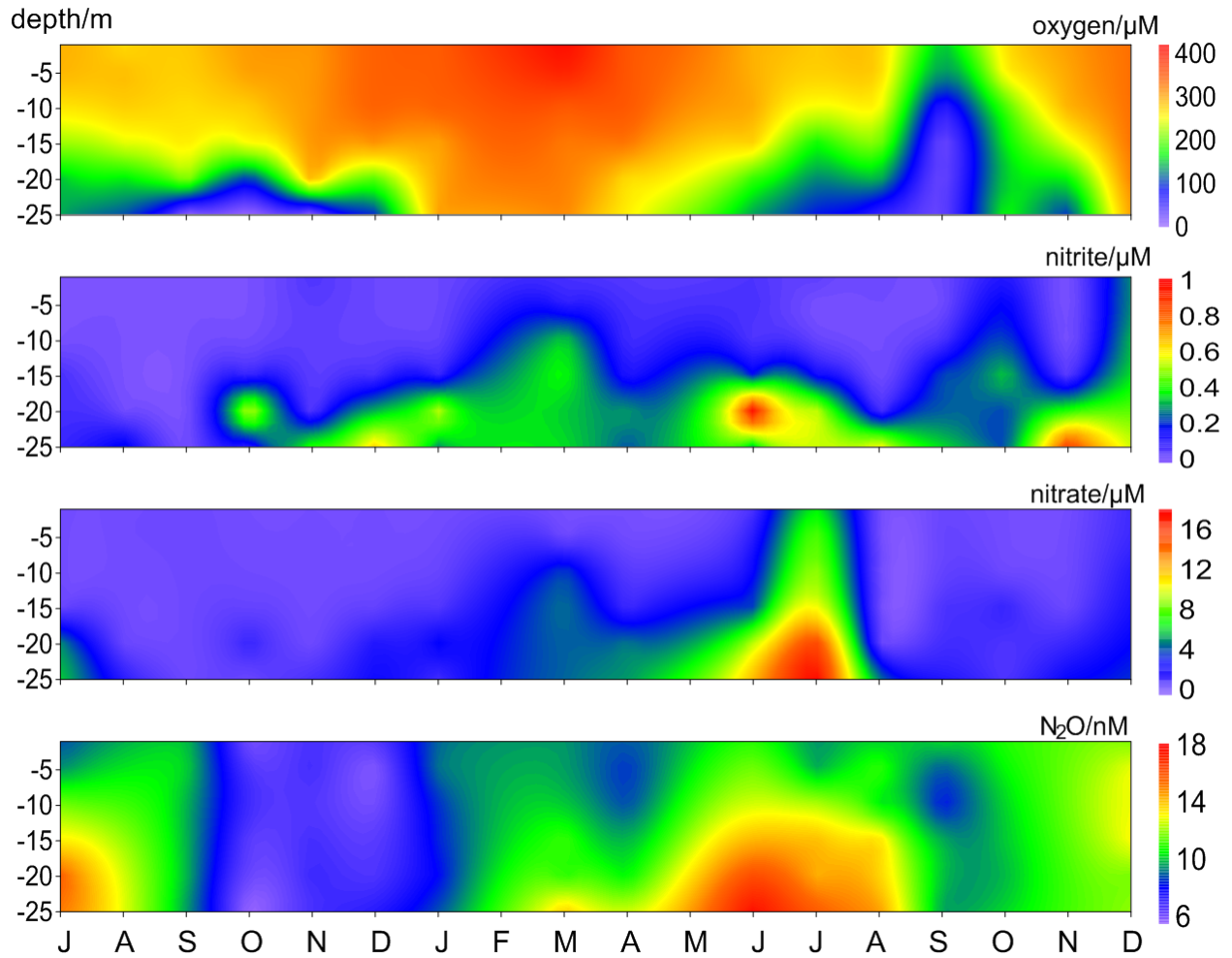
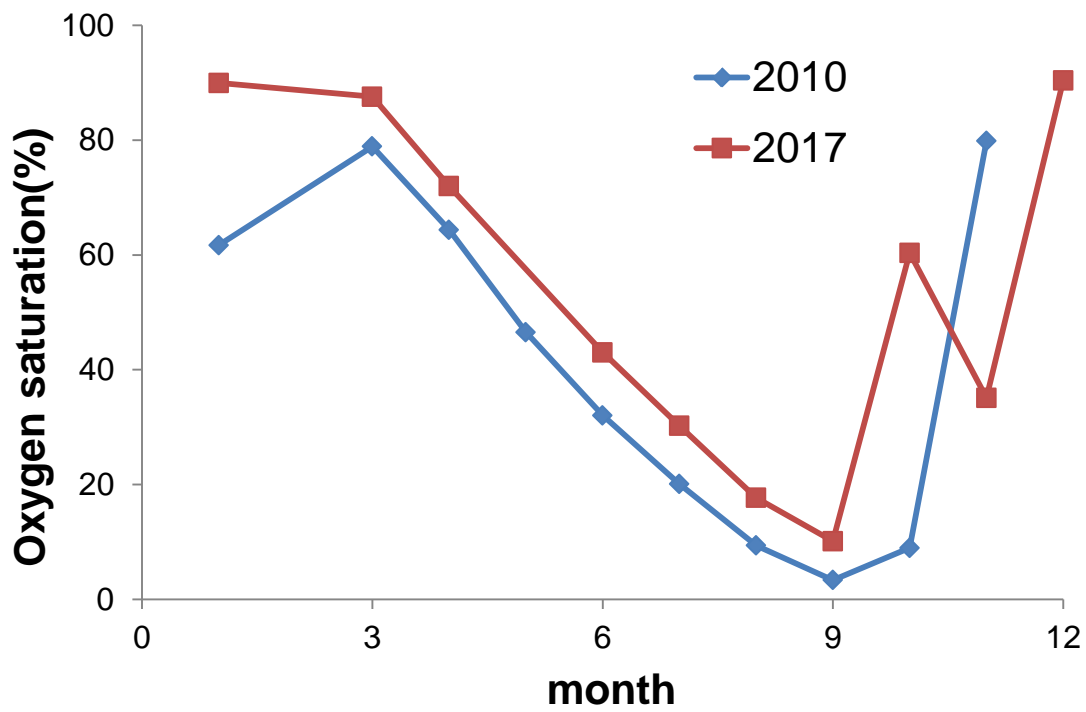
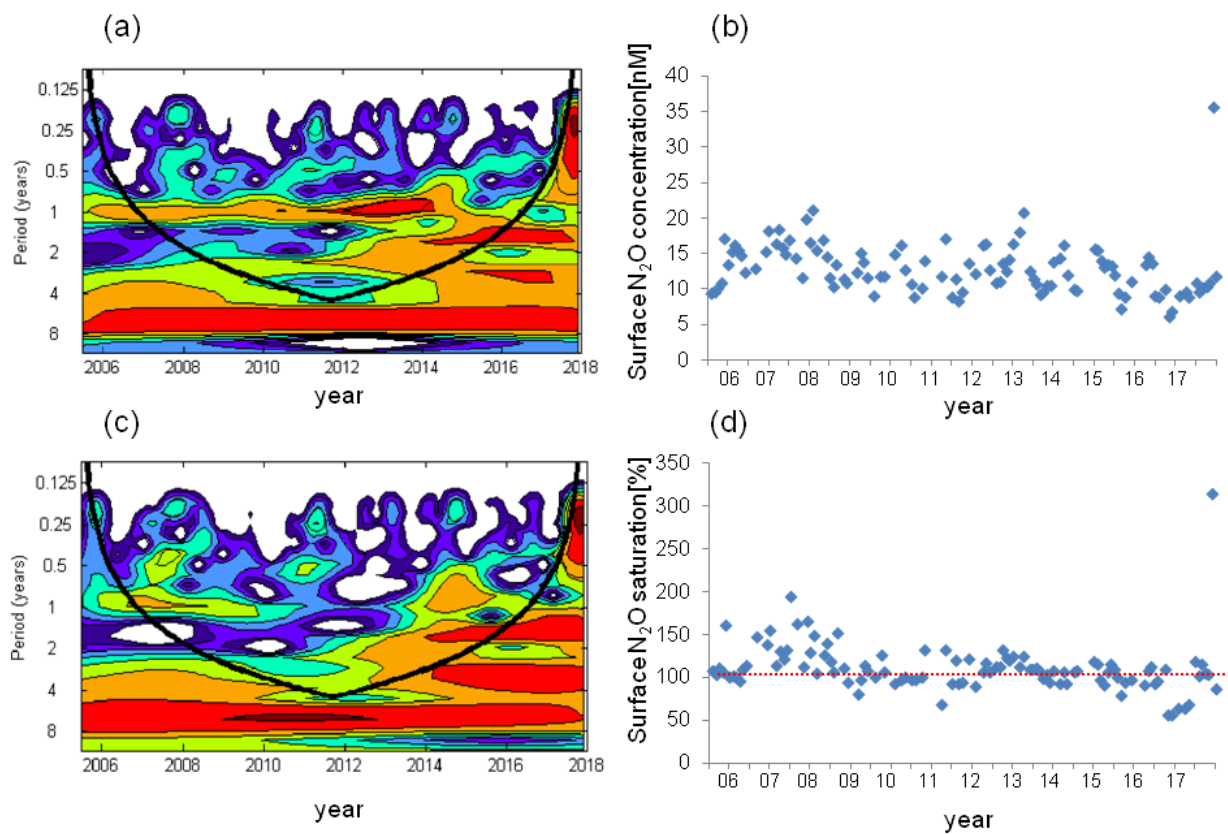


Fig. 5 Vertical distribution of dissolved O_2 , NO_2^- , NO_3^- , and N_2O from the BE Time-Series Station during July 2016–December 2017. Please note that the high N_2O concentrations in November 2017 were removed for better visualization.



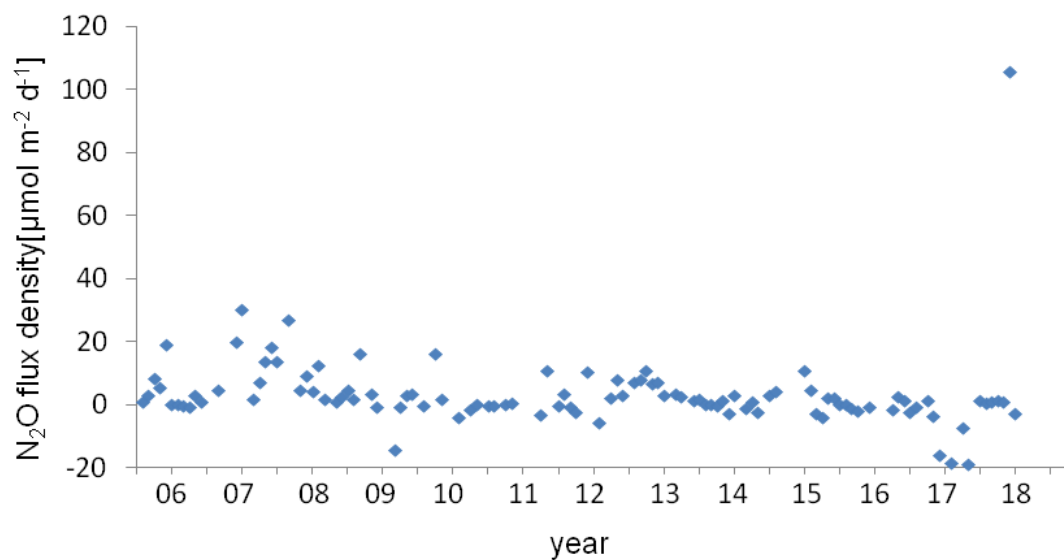
702

703 Fig. 6 Variations of bottom O₂ saturation in 2010 (blue) and 2017 (red).



704

705 Fig. 7 Wavelet analysis and the variation of surface N_2O concentrations (a, b) and surface N_2O saturations (c,
 706 d). The dashed red line in (d) indicates the saturation of 100%.



707
 708 Fig. 8 Variation of N₂O flux density at the BE Time Series-Station during 2005–2017. Negative values
 709 indicated N₂O influx from the atmosphere and positive values indicated N₂O efflux to the atmosphere.