Response to Reviewer 1

Revision Review for bg-2019-194:

Comparing Stability in Random Forest Models to Map Northern Great Plains Plant Communities in Pastures Occupied by Prairie Dogs Using Pleiades Imagery

Overview

Thanks for the revised manuscript and for your care and consideration of peer review comments. I find this revised manuscript to be much clearer and that the analyses now support the objectives and purpose laid out in the introduction. These changes have notably improved the manuscript and have removed confusion in the interpretation of findings. I have mostly clarification and technical comments that I feel will aid readers in understanding this work.

Specific Comments

Line 101 – If I understand the conclusions from Juel et al. 2015, then one logical extension would be that we also need to consider having spatially relevant training data (i.e. to address your issue that models may not transfer in space and time). Consider adding some additional possible solutions and implications of classification schemes (e.g. cover amounts of functional groups vs. community type)

Some additional text has been included on Line 95 to discuss this.

Line 133 – I find the connection between "signatures on imagery" and plant community response to the timing and progression factors underdeveloped. Add a sentence or two expanding what specifically will change within your communities (with relatively uniform composition) within and between years. Maybe a specific example would help here too.

This sentence has been removed. The connection between signatures on the imagery and plant community response is discussed in greater detail in the results and discussion section. See paragraph beginning on page 304.

Line 288 – Any spatial consistency on where these are? I.e. do they represent edges of the community where precip changes may lead to this finding? Would support next few sentences.

See additional text:

These are likely occurring along transition zones between prairie dog colony edge.

Line 317 – Talk about what this this means in terms of changes in or between your community types

See additional text:

Increased cumulative biomass in 2016 may cause higher NDVI values for example in On-PDG plant communities resulting in classification shifts to Off-Cool; similarly, greater NDVI values in Off-cool in 2016 may result in some of those pixels being classified as Off-Snow.

Line 329 – Need some discussion about how the selection of your community types leads to some heterogeneity within types, but this is a needed tradeoff (to lead into next paragraph)

Paragraph has been re-structured, and selection of plant communities and changes within types brought back to prairie dog influence of vegetation.

Line 393 – You have assessed the accuracy based on your 2016 data. So additional years helped you accurately predict your training sites from 2016 (relatively homogeneous areas). Be specific about what accuracy you have measured, which really is model performance here.

Point noted, this has been changed to model performance.

Line 398 – Do you also mean here that the selection of community types to map is an important consideration. I know you did not explore this specifically but seems to be an important theme in your discussion and results. Add some discussion and concluding statements about this aspect.

Additional text added:

...recognizing that plant communities rarely exist in discrete communities is important when selecting community types to map. Combining plant community ordination results with remote sensing results can aid in understanding sources of model error and limitations of classification algorithms.

Line 610 – Here and throughout the Tables and Figures please check and revise for acronym consistency. You switch between On-PDG and On-Grass, and On-PDF and On-Forb, within and between figures and tables

Corrected

Technical Comments

Line 93 – Need parentheses around 2015

Corrected

Line 99 – Parentheses around 2018(check rest of document for formatting of refs too)

Corrected

Line 111 – Do you mean prairie systems worldwide or specifically mixed grass prairies of the U.S. Northern Great Plains? I think you need to be specific here of the geographic region this paragraph addresses.

Changed to norther great plains mixed grass prairie

Line 125 – If you have a ref to send readers to about the larger study, please add.

added

Line 171 – Add the specific station used and check citation info (I found/used ref below). South Dakota Mesonet, South Dakota State University. (2019). South Dakota Mesonet Database [database].

added

Line 179 – Last sentence probably not needed. Also consider moving sentences (lines 225-227) about removing these areas and mapping prairie dog colonies up to this spot for reader clarity.

Sentence removed and additional sentence moved up in text.

Line 270 – For consideration, is "error" the best term here? For the message in your manuscript maybe use "instability?" You have the common problem of heterogeneity in your pixels/plots which makes it hard to classify to a specific type and your analysis shows that the year used can switch these mixed pixels between classes (the stability issue you are covering).

Point noted, changed to instability

Line 298 – Nice discussion in this paragraph

Line 301 – Need reference

added

Line 400 – Clarify that this is transition "zones" between communities (and not through time)

added

Line 621 – The locations of the on plot labels were confusing to me at first. Consider making these the same color as the community points and in the figure legend discuss what the +'s in the plot represent (this may also help folks identify the labels go with these centers)

Added in the legend is what +'s mean in the plot.

Line 667 – Check acronyms for consistency (see comment on 610)

Corrected

Line 683 – Check plot labels (see comment on 610)

Corrected.

Response to Reviewer 2

A number of improvements were made but I still had a difficult time reading the manuscript. The response 'Listing an author at the beginning of a sentence is a common convention in ecological literature' assumes that the ecological literature is written well. It largely is not. For this and other reasons the paragraph beginning on line 64 should be cut in its entirety and I revert to Josh Schimel's recommendation that every sentence that begins with an author needs to be rewritten if the authors are not the subject of the sentence. Doing so will make the authors realize that the structure of the text needs to change to have a simpler logical flow that makes it more apparent why prairie dog colonies make for interesting remote sensing challenges. Much of the introduction reads like a grab-bag of random papers and I don't feel that my suggestion to improve it was taken seriously. The Results and Discussion are better but could use further improvement. These are important points because it would be nice if this interesting study was more readable.

The introduction has been restructured per recommendations, this includes bringing the significance on why mapping prairie dog plant communities are an important task ecologically to the forefront of the article. Additionally text in the introduction has been added to highlight why vegetation changes associated with increased herbivory from prairie dogs might pose challenges to remote sensing. In

addition sentences in the introduction, results and discussion have been changed unless the author is the subject. Significant portions of the results and discussion has been re-structured for clarity. Additional discussion has included impacts prairie dog herbivory have on plant communities and our ability to detect these with satellite imagery.

The abstract would benefit from a brief discussion of why prairie dog colonies are important keeping in mind the international and/or non-ecological readership who would benefit from an explanation. In brief, the manuscript is technically sound but won't reach the intended audience unless the reader can see more clearly how interesting it is.

Additional discussion has been included in the abstract as well as the introduction about why prairie dogs are important to reach a broader audience.

Minor point? 152: How did cattle impact vegetation?

Cattle can impact vegetation, but this is dependent at the intensity of grazing. At 50% utilization, livestock will have a minimal impact on plant communities. Within the larger study no difference in plant communities were detected between off-town plots where cattle were excluded and off-town plots where cattle were allowed to graze. Differences largely align along an on- town and off-town gradient.

28		
29	TITLE	
30 31	Comparing Stability in Random Forest Models to Map Northern Great Plains Plant Communities in Pastures Occupied by Prairie Dogs Using Pleiades Imagery	
32	Jameson Brennan ^{a,} , Patricia Johnson ^a , and Niall Hanan ^b	
33	^a South Dakota State University West River Agricultural Center 1905 N Plaza Dr. Rapid City,	
34	SD 57702	
35	^b Jornada Basin LTER, New Mexico State University Plant and Environmental Sciences Las	
36	Cruces, NM 88003	
37	Corresponding author: Jameson Brennan	
38	Email: Jameson.brennan@sdstate.edu	Format
39	Second Author email: Patricia.johnson@sdstate.edu	Format
40	Third Author email: nhanan@ad nmsu edu	Format
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ABSTRACT

56	The use of high resolution imagery in remote sensing has Black tailed prairie dogs
57	(Cynomys ludovicianus) have been described as a keystone species and important for grassland
58	conservation, yet many concerns exist over the potential to improve understanding of patch level
59	variability inimpact of prairie dogs on plant structure and community composition that may be
60	lost at coarser scales. Random forest (RF) is a machine learning technique that has gained
61	considerable traction in remote sensing applications due to its biomass production and
62	consequently livestock production. The ability to produce accurate classifications map plant
63	communities in pastures colonized by prairie dogs can provide land managers with highly
64	dimensional data and relatively efficient computing times. an opportunity to optimize rangeland
65	production while balancing conservation goals. The aim of this study was to test the ability of
66	random forest (RF) to classify five plant communities located both on and off prairie dog towns
67	in mixed grass prairie landscapes of north central South Dakota, assess the stability of RF models
68	among different years, and determine the utility of utilizing remote sensing techniques to identity
69	prairie dog colony extent. During 2015 and 2016, Pleiades satellites were tasked to image the
70	study site for a total of five monthly collections each summer (June-October). Training polygons
71	were mapped in 2016 for the five plant communities and used to train RF models. Both the 2015
72	and 2016 RF models had low (1%) out of bag error rates. However, comparisons between the
73	predicted plant community maps using the 2015 imagery and one created with the 2016 imagery
74	indicate over 32.9% of pixels changed plant community class between 2015 and 2016. The
75	results show that while RF models may predict with a high degree of accuracy, overlap of plant
76	communities and inter-annual differences in rainfall may cause instability in fitted models. A
77	final RF model combining both 2015 and 2016 data yielded the lowest error rates, and was also
78	highly accurate in determining prairie dog colony boundaries.

79	Keywords
80	Remote sensing, random forest, rangelands, plant ecology, high resolution imagery
81	
82	INTRODUCTION
83	Within the Northern Great Plains mixed grass prairie ecosystem, black tailed prairie dog
84	colonization is an issue of concern for livestock producers (Miller et al. 2007). Competition
85	between prairie dogs and livestock is a major concern for land managers looking to optimize
86	beef production while still conserving wildlife species (Augustine and Springer 2013). Prairie
87	dogs have been identified as a keystone species, and are often seen as ecosystem engineers
88	providing habitat to a number of other plant and wildlife species (Davidson et al. 2010; Kotliar et
89	al. 1999). Prairie dogs can also reduce availability of forage for livestock by directly reducing
90	the quantity of forage available (through direct consumption, clipping plants to increase predator
91	detection, and building soil mounds), and by changing species composition (Derner et al. 2006).
92	Within the mixed grass prairie, C3 mid-grasses tend to decrease and C4 short-grasses increase
93	along an increasing gradient of grazing intensity (Irisarri et al. 2016). Due to repeated
94	defoliation, older core areas of prairie dog towns often become characterized by extensive areas
95	of bare ground and low vegetation production, which is generally limited to annual forb and
96	dwarf shrub species. Pastures containing extensive areas of bare ground due to prairie dog
97	colonization may potentially depress livestock forage intake rates and ultimately beef production.
98	The ability to map the extent and monitor the impact of prairie dogs on the landscape can help
99	land managers looking to optimize livestock production on prairie dog occupied rangelands.

Remote sensing of rangelands greatly improves our ability to study and understand 100 complex ecological interactions across the landscape. As technology advances, monitoring of 101 rangelands via remote sensing platforms will facilitate research products freely available to land 102 103 managers (Browning et al. 2015). One of the main advantages of remote sensing data is its capacity to cover wide areas, allowing assessment of plant communities at landscape level scales 104 105 as compared to traditional point-based assessments (Ramoelo et al. 2015; Yu et al. 2018). Numerous studies have demonstrated the utility of remote sensing applications in monitoring 106 rangeland condition, including mapping of vegetation communities, plant species composition, 107 108 biomass estimation, and impact of grazing intensity on the landscape (Goodin and Henebry 1997; Blanco et al. 2008; Franke et al. 2012). Additionally, successive images throughout a 109 110 growing season may potentially capture phenological changes associated with differences in C-3 and C-4 plant species composition (Goodin and Henebry 1997). 111

Advances in remote sensing technology have facilitated the mapping and assessment of a 112 broad range of habitats at different scales (Corbane et al. 2015). For example, Schmidtlein et al. 113 114 (2007) used hyperspectral imagery at 2m resolution in combination with ordination techniques to map functional plant group gradients in a Bavarian pasture. Within the Delaware Gap National 115 Recreation Area, multiple Landsat 7 scenes were used (30m resolution) with classification tree 116 algorithms to map forest and plant communities for the National Park Service Vegetation 117 Mapping Program (de Colstoun et al. 2003). In Majella National Park, Italy, 4m resolution 118 119 imagery was used with normalized difference vegetation index (NDVI) to map and predict grass and herbaceous biomass variability over a 200 km² area (Cho et al. 2007). 120

Many methods for accurately classifying plant communities using remote sensing
 techniques have been used in numerous ecological and natural resource studies. One method,

123	random forest classification (RF), has gained considerable traction in the remote sensing
124	community for its ability to produce accurate classifications, handle highly dimensional data, and
125	provide efficient computing times (Belgiu and Drăguț 2016)RF is seen as an improvement over
126	simple classification tree analysis by reducing noise and misclassification of outliers (Laliberte et
127	al. 2007; Nitze et al. 2015). RF is an ensemble decision tree classifier which combines bootstrap
128	sampling to construct several individual decision trees from which a class probability is assigned
129	(Mellor et al. 2013). RF builds each tree using a deterministic algorithm selecting a random set
130	of variables and a random sample from the calibration data set (Ramoelo et al. 2015).
131	The utility of random forest algorithms has been demonstrated in remote sensing
132	applications at multiple scales. Lowe and Kulkarni (2015) showed that RF was effective at
133	producing highly accurate classification maps using two Landsat scenes (30m resolution).
134	Ramoelo et al. (2015) successfully used RF modeling to predict leaf nitrogen content using
135	World-View 2 satellite images (2m resolution) in grassland and forest communities. Similarly,
136	Mutanga et al. (2012) concluded that RF regression modelling provided an effective
137	methodology for variable selection and predicting biomass in wetland environments using high
138	resolution satellite imagery (2m).
139	Considerable research has focused on the application of RF classification across different
140	plant communities at various scales, however, concerns existacross many plant communities at
141	multiple scales (Mutanga et al. 2012; Lowe and Kulkarni 2015; Ramoelo et al. 2015). Concerns
142	exist, however, over the transferability of these models to different sites, across seasons, or years.
143	For example, Juel et al. 2015 showed that RF models have shown to have a high degree of
144	classification accuracy for elassifyingmapping fine scale coastal vegetation using digital
 145	elevation maps and high resolution orthophoto imagery, but model accuracy decreased

146	significantly when applied to spatially separated sites, showing a lack of stability in the model.
147	Corcoran et al. (2013) showed (Juel et al. 2015). Selecting spatially releveant training data or
148	including species level cover data may help improve or explain differences observed when
149	transferring models between sites. Incorporating additional seasons of data may also improve
150	RF model accuracy; previous research has shown an improvement of RF model accuracy in
151	classifying wetlands in northern Minnesota with the inclusion Landsat 5 images across two years
152	using full season data versus summer only, and fall only models. Jones (Corcoran et al.
153	20182013). Longer term studies have also demonstrated the utility of using RF modeling with
154	30m Landsat data to monitor rangeland cover across the western United States over a 33 year
155	period. (Jones et al. 2018). Results of these studies suggest the scale and seasonality of the
156	imagery may play an important role in the stability and accuracy of RF models.
157	The stability in RF models to accurately map plant communities within prairie dog
158	occupied pastures may be particularly important for managers looking to monitor prairie dog
159	colony expansion or contraction over time. While classification rates are often reported in
160	studies, the potential overlap in plant community composition is rarely explored as a potential
161	source of error within the models. MostMany research studies focus solely on spectral
162	differences in plant communities and fail to analyze community differences on the ground at the
163	species level (de Colstoun et al. 2003; Geerken et al. 2005). Lastly, while the focus of many of
164	these remote sensing studies is on mapping plant communities at landscape scales to study land
165	use changes and address conservation related issues, very little research has examined the
166	impacts of animal species on plant community composition, and how this might affect
167	elassification accuracy. This may be especially important within prairie dog occupied
168	rangelands, where shifts in plant community composition may be driven more by the presence or

absence of an herbivore species versus elevation, soils, or other landscape features. These
 herbivory induced changes in plant community may facilitate or hamper classification schemes.
 The ability to accurately map plant communities within prairie dog occupied pastures can help
 improve management of rangelands colonized by prairie dogs, yet little research has explored the
 possibility of utilizing remote sensing as a tool to do so.

174 Within the mixed grass prairie ecosystem, black tailed prairie dog colonization is an issue of concern for livestock producers (Miller et al. 2007). Competition between prairie dogs and 175 livestock is a major concern for land managers looking to optimize beef production while still 176 conserving wildlife species (Augustine and Springer 2013). Prairie dogs can reduce availability 177 of forage for livestock by directly reducing the quantity of forage available (through direct 178 179 consumption, clipping plants to increase predator detection, and building soil mounds), and by changing species composition (Derner et al. 2006). Older-core areas of prairie dog towns often 180 become characterized by extensive areas of bare ground and low vegetation production, which is 181 generally limited to annual forb and dwarf shrub species. Pastures containing extensive areas of 182 bare ground due to prairie dog colonization may potentially depress livestock forage intake rates 183 and ultimately beef production. The ability to accurately map prairie dog colonies using remote 184 sensing will help improve our understanding of the impact of prairie dogs on plant communities, 185 and help inform land management decisions within rangelands occupied by prairie dogs. 186

A large collaborative study from 2012-2016 was conducted to evaluate livestock production on mixed-grass prairie pastures with varying levels of prairie dog occupation. _A major goal of the larger study was to determine which plant communities on the pastures cattle preferred to graze, and how those preferences shifted within and between years-<u>(Olson et al.</u> 2016). Plant communities on the site were categorized based on location (on- or off-town) and

192 visually apparent dominant plant functional groups. Thus, plant community as defined for this study was a collection of species within an area of a relatively uniform composition different 193 from neighboring patches. Differences in neighboring patches were evident by differences in 194 dominant functional group (forb vs grass) or differences in photosynthetic pathways (C3 vs C4 195 196 grasses). We expected the plant communities to remain relatively stable during the study, 197 however their signatures on satellite imagery could change within and between years as a result of the timing and magnitude of rainfall, timing of green up, phenological progression, and other 198 199 factors. The overall goal of this paper, then, was to develop maps that accurately classify plant 200 communities based on satellite imagery collected between years. Specific objectives of this study were to 1) determine differences in the five identified plant communities based on species 201 composition, 2) assess the utility of using a RF model with high resolution satellite imagery to 202 classify plant communities of interest within a mixed grass prairie ecosystem containing prairie 203 dogs, 3) determine the stability of the RF model when using subsequent years of satellite 204 imagery with identical training data, and 4) determine the ability of high resolution satellite 205 206 imagery to accurately *classifymap* prairie dog towns. Our ability to map and understand these plant communities' at large scales will give researchers insight into applying RF models across 207 208 years using high resolution imagery. Research from this study will allow us to better assess how 209 prairie dogs drive changes in plant communities, and provide a new tool to map the extent and impact of prairie dog colonization on the landscape to better inform land management decisions. 210

211

METHODS

212 Study site

The study area (45.74N, 100.65W) was located near McLaughlin, South Dakota on a
northern mixed-grass prairie ecosystem. Native prairie pastures (810 ha total area) were leased

215 from 2012-2016; pastures were continuously stocked with yearling steers from June-October of each year to achieve 50% utilization. Of the 810 ha, approximately 186 ha were occupied by 216 black-tailed prairie dogs (*Cynomys ludovicianus*). Predominant soils at the site were clays and 217 loams. Ecological sites, and the plant communities they support vary widely; Loamy and Clayey 218 219 were the predominant Ecological Sites at the site with inclusions of Dense Clay, Shallow Clay, 220 and Thin Claypan (Barth et al. 2014). Plant species dominating the site were largely native, including western wheatgrass (Pascopyrum smithii Rydb.), green needlegrass (Nassella viridula 221 Trin.), and needle-and-thread (Hesperostipa comata Trin. & Rupr), intermixed with blue grama 222 223 (Bouteloua gracilis Willd. Ex Kunth), buffalograss (Bouteloua dactyloides Nutt.), and sedges (*Carex* spp.). The most common non-native species on the site was Kentucky bluegrass (*Poa* 224 225 pratensis Boivin & Love). Woody draws occupied moist drainage areas; vegetation consists 226 primarily of bur oak (Quercus macrocarpa Nutt.), American plum (Prunus americana Marshall), and chokecherry (Prunus virginiana L.). These draws were frequently flanked by snowberry-227 dominated patches (Symphoricarpos occidentalis Hook.). Plant communities on areas occupied 228 by prairie dog towns on the site were largely dominated by western wheatgrass and shortgrasses 229 (buffalograss, blue grama, and sedges) intermixed with patches of bare ground and annual forb 230 231 dominated areas. Common annual forbs on prairie dog towns included prostrate knotweed (Polygonum aviculare L.), fetid marigold (Dyssodia papposa Vent.), dwarf horseweed (Convza 232 ramosissima Cronquist), and scarlet globemallow (Sphaeralcea coccinea Nutt.). A weather 233 234 station has been maintained on site from May 2013 operated by South Dakota Mesonet. Mean annual rainfall at the site is 446 mm and average growing season (May through September) 235 236 temperature is 15.3°C (South Dakota Climate and Weather 2017).

237 Five plant communities of interest for our study site were identified: 1) Forb-dominated sites on prairie dog towns (On-Forb), 2) Grass-dominated sites on prairie dog towns (On-Grass), 238 3) Snowberry-dominated sites off-town (Off-Snow), 4) Cool season grass-dominated sites off-239 town (Off-Cool), and 5) Warm season-dominated sites off-town (Off-Warm). An additional 240 plant community labeled 'Draws' was delineated visually within ArcGIS software due to 241 difficulty in mapping these areas in the field. Areas delineated as Draws were removed from the 242 analysis area. As mentioned prior, these areas are dominated by bur oak, chokecherry, and 243 American plum, and occupied lower lying drainage areas on the site. 244

245 Training sites

To facilitate classification, training site polygons were mapped for On-Forb, On-Grass, 246 Off-Cool, Off-Warm, and Off-Snow plant communities using ArcPad for Trimble GPS units in 247 the summer of 2016. Twenty training sites were mapped for each of the plant communities 248 except Off-Warm, for which only 8 sites were mapped due to the difficulty of finding 249 homogenous stands of warm season grasses. Plant species in the Northern Great Plains are 250 dominated by cool season species; warm season species, where they occur, are typically 251 intermixed into stands of cool season species. Training sites for each plant community were 252 253 selected from across the entire study area to capture potential site differences across research 254 pastures. Sites were mapped in the field by walking the perimeter of the plant community patch with a Trimble GPS unit. Training polygon perimeter boundaries were always at least 3 meters 255 256 interior of patch edge to minimize error introduced to the training data as a result of GPS signal noise. Identified patches were then converted into a polygon shapefile within ArcGIS to be used 257 258 as training polygons for the RF classification algorithm. Within each training site polygon, three 259 0.25 m² plots were randomly located by tossing plot frames into the area of interest to determine

sampling area. Within each plot, percent cover by species was recorded in the summer of 2016at the time of polygon mapping.

262 Plant Community Analysis

Plant community analysis was performed on vegetation data collected from the three 263 $0.25m^2$ plots measured in each training polygon. Differences between plant community 264 265 compositions were determined using a Multi-Response Permutation Procedure (MRPP) with the Sorensen Bray-Curtis distance method. MRPP is a nonparametric procedure used for testing 266 hypotheses between two or more groups (Mitchell et al. 2015). Differences in community 267 268 compositions were analyzed for all plant communities, and pairwise comparisons generated. To analyze trends in species composition between plant community plots, Non-metric 269 Multidimensional Scaling (NMS) ordination was used (Kruskal 1964). Only species that 270 271 occurred in 3 or more plots were included in the ordination analysis. NMS analysis was conducted using the Sorensen Bray-Curtis distance method with 250 iterations and a stability 272 criterion of 0.00001. Analysis was repeated five times to confirm ordination pattern in the data. 273 Similarity index matrices were generated to compare plot differences between plant communities 274 and averaged by plant community. All ordination analyses (MRPP and NMS) were performed 275 276 using PC-ORD 6 software (McCune and Mefford 2002).

277 Imagery

During the summers of 2015 and 2016, Pleiades satellites were tasked to image the study site. Pleiades satellites, which are members of the SPOT family of satellites, are operated by AIRBUS Defense and Space. This platform was chosen due to its high spatial resolution (0.5 m pan chromatic, 2 m multispectral) and four band spectral resolution: pan chromatic (480-830 nm), red (600-720nm), green (490-610 nm), blue (430-550 nm), and near infrared (750-950 nm).

283 Pleiades satellites were designed for commercial tasking and monitoring, allowing multiple revisits to a project site. A total of ten image collections were acquired in the summer of 2015 284 and 2016 (five each year) from June through October during the 1st-15th of each month (Table 1). 285 Image collection times were chosen to correspond to the time periods when cattle were actively 286 grazing on the site. Multispectral images were pan-sharpened and orthorectified by the image 287 288 provider (Apollo Imaging Corp). Each monthly image collection was converted into an NDVI image. Areas delineated as Draws were removed from the analysis area. In addition, boundaries 289 of the prairie dog town were mapped using a handheld Trimble GPS unit to compare predicted 290 291 colony location with ground truth location.

292 Random Forest model

293 For the RF model, the Random Forest package of the Comprehensive R Archive Network (CRAN) implemented by Liaw and Wiener (2002) was utilized. Training data were constructed 294 295 by stacking all satellite imagery spectral bands (Red, Blue, Green, and NIR) and NDVI bands for each month of each year (25 total dimensions per year) to create a raster stack for each year's 296 imagery (2015 and 2016). To train the model, pixel values were extracted from the satellite 297 imagery raster stack for each training polygon mapped in the field. The random forest models 298 299 were built using 200 decision trees and default number of nodes at each split (sqrt(n)), with plant community data as the response category (On-Grass, On-Forb, Off-Cool, Off-Warm, and Off-300 Snow) and spectral band values as the predictor. Models were checked for error stabilization, for 301 302 all models error rates stabilized around 50 trees. Yearly models (2015 and 2016) were built for output comparison. A combined years model was also constructed using all available spectral 303 304 data from 2015 and 2016 (50 dimensions).

305	Within the random forest package, Out of Bag (OOB) error rates were calculated by
306	reserving one-third of the training data to test the accuracy of the predictions. Models were then
307	used to predict class belonging for 2015 and 2016 raster stacks and the combined 2015 and 2016
308	stack using the 'predict' function within program R. To assess the stability of the RF models
309	from year to year, the "crosstab" function in the raster package in program R was used to
310	calculate the number of pixels that changed class from 2015 to 2016. The output was used to
311	calculate percent of pixels that were unchanged from 2015 to 2016 model predictions and
312	percent of pixel change that occurred between years for plant community predictions.
313	Results and Discussion
314	Plant Community
315	MRPP pairwise comparisons results showed a significant difference between all plant
316	communities ($P < 0.001$). Differences are evident between plant communities in the 2-D plot of
317	the NMS ordination (final stress = 20.01 , instability < 0.00001 after 66 iterations), with some
318	overlap occurring between communities (Figure 1). Plant communities on-town and off-town
319	are clustered at opposite ends of the ordination plot, with the greatest distance being between On-
320	Forb and Off-Snow. Archer et al. (1987) showed in a detrended Detrended correspondence
321	analysis of plant communities ranging from uncolonized, 2 years post colonization, and 4-6 years
322	post colonization, showed that uncolonized sites were clustered at one extreme and the 4-6 year
323	sites at the other extreme- (Archer et al. 1987). Interestingly, Off-Warm and On-Grass
324	communities are clustered closer in ordination space. Plant communities shifts on-town towards
325	those dominated by shortgrass species have been documented (Agnew et al. 1986; Koford 1958),
326	and is probably attributable to the high grazing resistance of the C4 species blue grama and
327	buffalograss (Derner et al. 2006).

328 Similarity index differences between plant communities were generated from a Sorensen (Bray-Curtis) distance matrix, and can be seen in Table 2. While there is some overlap between 329 plant communities, in general similarities between plant communities are low (< 29%), with the 330 331 greatest similarity index differences generated from a Sorensen (Bray-Curtis) distance 332 matrixdistance occurring between the On-Forb communities and the off-town communities 333 (Table 2). Based on how plant communities were selected in this study, we expected plant community composition to be distinct between groups. Though plant communities are defined 334 by dominant functional group in this study, the amount of overlap occurring demonstrates that 335 336 other functional groups and species exist within these distinct patches, which may be a potential 337 source of error instability in classification models.

338 Random Forest Model Results

339 Results from the RF modelmodels show low OOB misclassification error rates for each individual plant community (Table 3) indicating a high degree of accuracy in the model. Overall 340 341 the OOB model error rates were 0.9% and 1.12% for the 2015 and 2016 model respectively. OOB accuracy is an unbiased estimate of the overall classification accuracy eliminating the need 342 for cross-validation (Breiman 2001). <u>Lawrence et al. (2006) showed</u> OOB error rates have been 343 344 shown to be reliable estimates of class accuracy for identifying invasive species. Similarly, OOB 345 error rates have been reported to be reliable in (Lawrence et al. 2006), and mapping corn and 346 soybean fields across multiple years (Zhong et al. 2014). Belgiu and Drăgut (2016) in their 347 review of RF applications in remote sensing acknowledge that the reliability of OOB error measurements needs to be further tested using a variety of datasets in different scenarios 348 349 Consistency in error rates for plant communities appears to indicate stability in the 2015 350 and 2016 RF models which used identical training sites on consecutive yearly satellite imagery.

351 However, when comparing yearly predicted plant community maps, differences between community classifications are slightly more pronounced, indicating the models may not be as 352 353 stable as predicted based solely on the OOB error rates. Overall a total of 67.04% pixels remained unchanged in their plant community 354 355 classification from 2015 to 2016 (Table 4). Of the pixels that changed classification between 356 years, 15.13 were on-town to off-town transitions, 2.26 were on-town to on-town transitions, and 15.57 were off-town to off-town plant community transitions. It is unlikely in this northern 357 mixed-grass prairie ecosystem that all the changes in plant communities indicated by 358 359 classification of pixels were real changes from one plant community type to another over one year. In the absence of a major disturbance event, such major shifts in species composition 360 typically occur much more slowly (Vermeire et al. 2018). The results from the plant community 361 analysis indicate training sites were chosen appropriately to account for differences in species 362 composition on the ground, therefore apparent changes are much more likely due to factors that 363 affect the spectral signature of the vegetation. Factors that may potentially affect spectral 364 signatures could include changes resulting from prairie dog herbivory, changes in precipitation 365 regimes, or changes occurring along plant community transition zones. 366 367 The pixels changing from On-Grass to Off-Cool represented the highest percentage of pixels that changed plant community classification at 7.28%. Johnson-Nistler et al. (2004) 368 observed up to 7 times more standing dead forage present%; these are likely occurring along 369 370 transition zones at the prairie dog colony edge. Both On-Grass and Off-Cool plant communities have western wheatgrass as a dominant species. Similarity in species dominance may explain 371 372 some of the challenges to distinguishing between some on and off colony plant communities. 373 Difficulty in classifying Off-Cool and On-PDG may also be due to subtle vegetation changes

374 likely induced by the level of herbivory. Research on a South Dakota mixed grass prairie showed that prairie dogs remove over four times more biomass than cattle grazing on-town 375 376 (Gabrielson 2009). Up to 7 times more standing dead forage and 60% less standing crop 377 biomass has been reported on uncolonized sites compared to colonized areas, mainly attributed to prairie dogs clipping vegetation which greatly reduced the amount of grasses that reached 378 379 maturity- (Johnson-Nistler et al. 2004). Areas either less maintained on-town by prairie dogs or grazed by cattle repeatable off-town may show up-similar spectral signatures. Additionally, On-380 Grass and Off-Cool plant communities have western wheatgrass as a dominant species, and 381 382 similarity in species dominance between these communities may explain yearly shifts in predictions. Of the pixels that changed classification between years, 15.13 were on-town to off-383 town transitions, 2.26 were on town to on town transitions, and 15.57 were off town to off town 384 plant community transitions. 385 386 Differences It is unlikely in this northern mixed-grass prairie ecosystem that all the changes in plant communities indicated by classification of pixels were real changes from one 387 plant community type to another over one year. In the absence of a major disturbance event, 388 such major shifts in species composition typically occur much more slowly. The results from the 389 plant community analysis indicate training sites were chosen appropriately to account for 390 differences in species composition on the ground, therefore apparent changes are much more 391 likely due to factors that affect the spectral signature of the vegetation. One explanation for the 392 393 difference in year to year classification could <u>also</u> be attributed to the interannual variability of rainfall between 2015 and 2016 (Figure 2). Yearly rainfall patterns can result in large 394 differences in NDVI and biomass measurements across years (Wehlage et al. 2016). While 395 396 overall total rainfall between years was similar, differences in timing of precipitation that

397 occurred likely affected timing of green up and dormancy for many of the cool and warm season species on the site. This, then, would create different NDVI patterns between years (Figure 3). 398 399 Wehlage et al. (2016) for example, found that yearly rainfall differences resulted in large differences in NDVI and biomass measurements across two years in a dry mixed-grass prairie. 400 401 Goward and Prince (1995) suggested that the relationship between NDVI and annual rainfall in 402 any given year also depends on the previous year history of rainfall at the site, and Oesterheld et al. (2001) showed. Previous research has shown that annual above ground primary production of 403 404 shortgrass communities is related to current as well as previous two years precipitation-405 (Oesterheld et al. 2001). The above average rainfall at the study site in 2015 could have added to the increase in average NDVI in 2016 when compared to 2015 through an increase in cumulative 406 biomass or production at the site. Increased cumulative biomass in 2016 may cause higher 407 NDVI values for example in On-PDG plant communities resulting in classification shifts to Off-408 Cool; similarly, greater NDVI values in Off-cool in 2016 may result in some of those pixels 409 410 being classified as Off-Snow. Another possible cause for changes in plant community classifications between years is 411 412 overlap of plant community species where two plant communities share a boundary. The edges 413 of plant One issue with using categorically classified vegetation maps is that plant communities in the NGP space are seldom sharp; more often there is a transition zone, where species from 414 each community intermingle. This, rarely mutually exclusive, and tend to change along with 415 416 variability in phenological development of different plants (e.g. cool season vs. warm season) 417 associated a continuum with precipitation, as mentioned above, could result in pixels appearing to be associated with one plant community in one year and its neighboring plant community the 418 419 next. It should also be noted that plantenvironmental gradients (Equihua 1990). Plant

communities in the region, which are predominantly comprised of cool season grasses, often
include varying levels of warm season species; and snowberry thickets often have an understory
of grasses, especially near the perimeter. Thus one should expect some level of spectral mixing
within each community, and the possibility that climatic factors could result in changes in NDVI
values that, at least initially, might suggest apparent changes between The challenge of
accurately classifying plant communities.

As noted above, one issue with using categorically classified vegetation maps is that plant 449 communities in space are rarely mutually exclusive, and tend to change along aan ecological 450 451 continuum withmay be further exacerbated by changes induced by prairie dogs, where transition zones are less defined by environmental gradients (Equihua 1990).and more defined by the level 452 of herbivory. Thus, within bothand between on-town and off-town plant communities, transition 453 454 zones are likely to account for a portion of the classification change between plant communities 455 between years (Figure 4). -Alternative approaches to mapping plant communities can be the 456 recognition of fuzzy properties enabling a single point in space to exhibit characteristics of a number of plant communities (Duff et al. 2014; Fisher 2010). For example, Schmidtlein et al. 457 (2007) used NMS of species data in combination with imaging spectroscopy to produce 458 459 ordination maps of community structure. While fuzzy classification maps are more likely to give a better picture of plant community composition on a per pixel basis, they are also more difficult 460 461 to use to draw inferences of species dominance-and, livestock use across landscapespatterns, and extent of prairie dog colonization. 462

A final RF model combining all available bands and NDVI values for 2015 and 2016 reduced error rates for all plant communities below 1% (Table 3). While we have shown that <u>lower error rates may not result in more stable predictions, using all available data for a model</u> Formatte

466	will likely improve accuracy and result in a more accurate thematic map. <u>Other studies have</u>
467	reported increases in classification accuracy in RF models with the addition of combined
468	seasonal images, hyperspectral data, LiDAR images, radar (SAR) images, and ancillary
469	geographical data such as elevation and soil types (Corcoran et al. 2013; Pu et al. 2018; Shi et al.
470	2018; Xia et al. 2018; Yu et al. 2018; Zhou et al. 2018). RF models have the ability to handle
471	highly dimensional correlated data, and data combined from multiple different data sources
472	across different temporal scales; however, one disadvantage to using non-parametric classifiers
473	such as RF and decision trees is that they require a large number of observations to accurately
474	estimate the mapping function (James et al. 2014). Thus the incorporation of additional predictor
475	variables as well as additional training data will likely result in higher accuracy rates.
476	The variable importance graph of the combined model indicates that NDVI variables
477	contribute the most to the model over individual bands (Figure 5). Similar results were observed
478	byIn classifying vegetation morphology in a savanna grassland, Mishra and Crews 2014, where
479	found spectral classification features (mean NDVI or ratio NDVI) were the most significant-for
480	classifying vegetation morphology in a savanna grassland. The variable importance plot from
481	the combined data model also indicates that different months between years contribute highly to
482	the classification accuracy. Of the ten most important variables in the model, 6 were from 2015
483	and 4 from 2016, suggesting additional data in the model is likely to yield greater classification
484	accuracy. Zhou et al. (2018) using RF models showed that a combination of four seasons of
485	Sentinel-1 images and a GaoFen-1 satellite winter image produced the highest classification rate
486	of urban land cover scenes over individual seasonal images. Likewise, several otheryears' data
487	in the model is likely to yield greater classification accuracystudies have reported increases in
488	classification accuracy in RF models with the addition of combined seasonal images,

489 hyperspectral data, LiDAR images, radar (SAR) images, and aneillary geographical data such as elevation and soil types (Corcoran et al. 2013; Pu et al. 2018; Shi et al. 2018; Xia et al. 2018; Yu 490 et al. 2018). RF models have the ability to handle highly dimensional correlated data, and data 491 492 combined from multiple different data sources across different temporal scales. The internal information provided by the model, such as variable importance, can be a useful tool for 493 494 researchers to select features of greatest importance to reduce computation times in the instance of large datasets. At the size of our study area (810 ha) and a maximum of 50 variables, the 495 combined 2015-16 data model only slightly added to computation time, but not enough to 496 497 warrant feature trimming from the dataset. Land managers looking to classify prairie dog colonies on more extensive grasslands may look to including only NDVI variables into training 498 datasets to increase computational efficiency. 499

500 **Remote Sensing Prairie Dog Colonies**

Visual comparison of the predicted on-town plant communities versus off-town plant 501 communities show a clearly defined boundary between areas colonized by prairie dogs and areas 502 503 not colonized (Figure 6). Results from mapping colony boundaries with a hand held GPS device estimated the colony to be 276 ha in 2012 to 186 ha in 2015. Total colony acreage estimated 504 from summing the pixel area occupied by the On-Grass and On-Forb community pixels from the 505 combined 2015-2016 RF model was 246 ha. Previous research has demonstrated that 506 colonization by prairie dogs and subsequent increases in grazing pressure can result in significant 507 508 differences between on- and off-town plant community composition and production (Coppock et al., 1983; Winter et al. 2002; Johnson-Nistler et al. 2004; Geaumont et al. 2019). The results of 509 510 our study demonstrate that these differences are significant enough to be identified using remote 511 sensing techniques. Interestingly, a considerable portion of the area misclassified as on-town is

535	from a previously colonized area that had been poisoned in 2013, suggesting that, at least
536	spectrally, these areas still resemble plant communities similar to those actively colonized. The
537	higher area estimate from the RF model is likely the result of transition areas controlled two
538	years prior. Additionally, most other pixels misclassified as on-town are likely drainage areas
539	with high bare ground off-town, whose variability was not captured in the dataset. Results from
540	mapping colony boundaries with a hand held GPS device estimated the colony to be 276 ha in
541	2012 to 186 ha in 2015. Total colony acreage estimated from summing the pixel area occupied
542	by the On-Grass and On-Forb community pixels from the combined 2015-2016 RF model was
543	246 ha. As mentioned prior, the higher area estimate in from the RF model is likely the result of
544	transition areas controlled two years prior. One prior study had sought identify prairie dog
545	colonies using 30m Landsat imagery, however concluded that the scale was too course for
546	accurately measuring prairie dog towns (Wolbrink et al. 2002). High resolution satellite imagery
547	used in this study appears capable at capturing fine scale transitions that occur between plant
548	communities along the on town off town gradient
	communities along the on-town on-town gradient.
549	The RF model was also able to accurately predict older core areas of prairie dog towns
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549 550 551 552 553	Communities along the on-town off-town gradient. The RF model was also able to accurately predict older core areas of prairie dog towns (On-forb) often characterized by a high percentage bare ground, low vegetation production, and dominance by annual forb and dwarf shrub species (Coppock et al., 1983). Area estimates of On-Forb were 33 ha and 32 ha in 2015 and 2016 respectively. State and transition models for prairie dog towns developed within Custer State Park South Dakota, found older core areas were
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549 550 551 552 553 554 555 556 557	The RF model was also able to accurately predict older core areas of prairie dog towns (On-forb) often characterized by a high percentage bare ground, low vegetation production, and dominance by annual forb and dwarf shrub species (Coppock et al., 1983). Area estimates of On-Forb were 33 ha and 32 ha in 2015 and 2016 respectively. State and transition models for prairie dog towns developed within Custer State Park South Dakota, found older core areas were considered undesirable for management due to losses of native grasses, increased bare ground, potential for erosion, extensive presence of exotic species, and increased inputs to restore to a more desirable state (Hendrix 2018). The ability to monitor these older core areas of prairie dog towns remotely may help land managers limit sites from becoming highly degraded, and serve as

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<u>a useful tool for land managers concerned over balancing wildlife conservation with losses in</u> <u>livestock production.</u>

584 Conclusions

Stability of models is important when applying similar techniques across different sites, 585 plant communities, and in this case years. Differences in year-to-year NDVI values may alter 586 587 classification results, and the addition of two years' worth of data likely resulted in improved elassification accuracy.model performance. One of the main benefits to RF classification in 588 remote sensing is the relatively fast computing time (Belgiu and Drăgut 2016), and, given the 589 590 availability of free satellite imagery, researchers would be prudent to include multiple images across years and seasons in their model to improve accuracy. Furthermore, while the desired 591 592 outcome is often to produce thematic maps, recognizing that plant communities rarely exist in 593 discrete communities is important when trying to interpret remotely sensed classification maps.selecting community types to map. Combining plant community ordination results with 594 remote sensing results can aid in understanding sources of model error and limitations of 595 classification algorithms. This is likely to be magnified as pixel size decreases, resulting in fine 596 597 scale predictions which may be more susceptible to plant community transitions. Lastly, results 598 zones. Results from this study indicate that plant community changes induced by prairie dogs are significant enough to be detected via remote sensing techniques. Land managers looking to 599 600 optimize rangeland health on pastures occupied by prairie dogs may potentially utilize high 601 resolution imagery to monitor colony size and make recommendations of appropriate stocking rates based on extent of colonization. 602

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Tables and Figures

Table 1. Acquisition dates of Pleiades satellite imagery tasked for each month (June – October) in 2015 and 2016.

2015 Dates of Acquisition	2016 Dates of Acquisition
6/1/2015	6/5/2016
7/9/2015	7/2/2016
8/4/2015	8/2/2016
9/1/2015	9/11/2016
10/8/2015	10/1/2016

1020 Table 2. Similarity index (Sorensen (Bray-Curtis) distance method) values averaged by plot 1021 across plant <u>communities</u>.

	SimiliaritySimilarity Index
Community Comparison ¹	(%)
Off-Cool vs. Off-Snow	28.2
Off-Cool vs. Off-Warm	27.8
Off-Cool vs. On-PDG	27.7
Off-Snow vs. Off-Warm	21.6
On-PDG vs. On-PDF	17.8
Off-Snow vs. On-PDG	17.3
Off-Warm vs. On-PDG	17.3
Off-Cool vs. On-PDF	7.9
Off-Snow vs. On-PDF	6.2
Off-Warm vs. On-PDF	6.2

¹Plant communities on prairie dog towns are grass-dominated (On-Grass) and forb-dominated

1024 (On-Forb); plant communities in off-town areas are cool season grass-dominated (Off-Cool),

1025 warm season grass-dominated (Off-Warm), and snowberry-dominated (Off-Snow).







Figure 1. NMS ordination plots for plant communities located on and off of prairie dog towns,
based on plant cover by species data collected in 2016 on the study site in north central South
Dakota. <u>The '+' symbol followed by the community name represent the weighted mean</u>
(centroid) of the multivariate dataset. Plant communities on prairie dog towns are grassdominated (On-Grass) and forb-dominated (On-Forb); plant communities in off-town areas are
cool season grass-dominated (Off-Cool), warm season grass-dominated (Off-Warm), and
snowberry-dominated (Off-Snow).

COI	nbined year random fo	orest models.			
	Plant Community ¹	2015 Model	2016 Model	2015-2016 Combined Model	Formatt
	Off-Cool	0.20%	0.40%	0.04%	Eormatt
	Off-Snow	2.2%	1.9%	0.69%	Formati
	Off-Warm	3.2%	5.3%	0.73%	Formatt
	On-Grass	0.40%	0.60%	0.09%	Formatt
	On-Forb	0.60%	0.70%	0.19%	Formatt
.079 .080 .081 .082	¹ Plant communities (On-Forb); plant cor warm season grass-c	on prairie dog towr nmunities in off-tov lominated (Off-War	ns are grass-dominated wn areas are cool seas rm), and snowberry-d	d (On-Grass) and forb-dominated son grass-dominated (Off-Cool), ominated (Off-Snow).	Formatt
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1120Table 4: Percent of pixels within each plant community that remain unchanged and that changed1121class belonging between 2015 and 2016 models.

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	1			Percent of Total	Formatt
Transition	2015 PC ¹	2016 PC	Total Pixels	Pixels	
A	Off-Cool	Off-Cool	9712857	31.03	Eormatt
	On-	On-			Formate
A	PDGGrass	PDGGrass	6427817	20.54	Formatt
Unchanged Pixels	Off-Snow	Off-Snow	3401264	10.87	Formatt
	On-	On-	007171	2.02	Formatt
A	PDF Forb	PDFForb	88/151	2.83	/ Formatt
A	Off-Warm	Off-Warm	555635	1.78	Formatt
	On-		2250200	7.00	Eormatt
A	PDGGrass	Off-Cool	2278390	7.28	
A	Off-Cool	Off-Snow	1468042	4.69	Formatt
		On-	10(0070	4.02	Formatt
A		PDG <u>Grass</u>	1262373	4.03	/ Formatt
A	Off-Snow	Off-Cool	11/4565	3.75	Formatt
A	Off-Warm	Off-Cool	729511	2.33	Eormatt
A	Off-Cool	Off-Warm	716503	2.29	
A	Off-Warm	Off-Snow	629212	2.01	Formatt
	On-		()())	2.00	Formatt
A	PDG <u>Grass</u>	Off-Snow	626695	2.00	Formatt
	DI-	DEForb	362/17	1 16	Formatt
A	On-	<u>- Dr<u>1'010</u></u>	302417	1.10	Formatt
Changed Pixels	PDEForb	PDGGrass	343774	1 10	Formatt
Changed I lixels		On-	515771	1.10	
	Off-Snow	PDGGrass	281061	0.90	Formatt
A	Off-Snow	Off-Warm	155213	0.50	Formatt
A	On-				Formatt
	PDGGrass	Off-Warm	82450	0.26	Formatt
A	On-				Formatt
	PDF <u>Forb</u>	Off-Cool	72758	0.23	
		On-			Formatt
A	Off-Cool	PDF <u>Forb</u>	69188	0.22	Formatt
		On-			Formatt
A	Off-Warm	PDGGrass	43132	0.14	Formatt
	On-				Formatt
A	PDF <u>Forb</u>	Off-Snow	19575	0.06	Eormatt
	0.00 111	On-	570	0.00	
	Off-Warm	PDF Forb	5/3	0.00	Formatt

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		On- PDF Forb	Off-Warm	314	0.00		Formatte
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1122	A	Off-Snow	PDF <u>Forb</u>	17	0.00		Formatte
1133 1134 1135 1136 1137	¹ Plant communities dominated (On-For (Off-Cool), warm s	(PC) on prair b); plant comr eason grass-do	ie dog towns ar nunities in off- ominated (Off-`	re grass-domina town areas are c Warm), and sno	ted (On-Grass) and fort cool season grass-domin wberry-dominated (Off)- 1ated ² -Snow).	Formatte
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Figure 2: Monthly and cumulative growing season precipitation patterns for 2015 and 2016 recorded at a weather station located on the study area in north central SD (45.737296 N, -100.657540 W)(South Dakota Mesonet 2018).







NDVI Comparisons for Plant Communities 2015 vs. 2016





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- 1152 Figure 4: Random forest classification maps from 2015 and 2016 of one pasture in the study area
- in north central South Dakota. Plant communities on prairie dog towns are grass-dominated
- 1154 (On-Grass) and forb-dominated (On-Forb); plant communities in off-town areas are cool season
- 1155 grass-dominated (Off-Cool), warm season grass-dominated (Off-Warm), and snowberry-
- 1156 dominated (Off-Snow).
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Combined 2016 & 2016 Model



1159 Figure 5: Variable importance reported as mean decrease in accuracy. Ten most important

variables are shown, with B1 and B4 corresponding to spectral bands 1 and 4 respectively from

1161 Pleiades image. Variable importance is determined by the model output as the decrease in

accuracy due to the exclusion of that variable during the out of bag error calculation process.

1163 Higher mean decrease in accuracy variables are more important in classifying the data.

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1169 Figure 6: Random forest classification map created from predictions from the combined 2015

and 2016 models. Off-town areas were created by combining the predicted off-town plant

1171 communities (Off-Cool, Off-Warm, and Off-Snow) and on-town plant communities (On-Grass

and On-Forb). The prairie dog boundary was mapped using a handheld GPS unit, the outlined

1173 2012 prairie dog boundary was former prairie dog colony poisoned in 2013.