



1   **Evaluation of bacterial glycerol dialkyl glycerol tetraether and  $^2\text{H}$ -  
2    $^{18}\text{O}$  biomarker proxies along a Central European topsoil transect**

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20 **Keywords**

21 Leaf wax *n*-alkanes, hemicellulose sugars, pH, temperature, CBT, MBT', precipitation  
22  $\delta^2\text{H}/\delta^{18}\text{O}$ , relative humidity

23 **Abstract**

24 Molecular fossils, like bacterial branched glycerol dialkyl glycerol tetraethers (brGDGTs), and  
25 the stable isotopic composition of biomarkers, such as  $\delta^2\text{H}$  of leaf wax-derived *n*-alkanes ( $\delta^2\text{H}_{n\text{-alkane}}$ )  
26 or  $\delta^{18}\text{O}$  of hemicellulose-derived sugars ( $\delta^{18}\text{O}_{\text{sugar}}$ ) are increasingly used for the  
27 reconstruction of past climate and environmental conditions. Plant-derived  $\delta^2\text{H}_{n\text{-alkane}}$  and  
28  $\delta^{18}\text{O}_{\text{sugar}}$  values record the isotopic composition of plant source water ( $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$ ),  
29 which usually reflects mean annual precipitation ( $\delta^2\text{H}/\delta^{18}\text{O}_{\text{precipitation}}$ ), modulated by  
30 evapotranspirative leaf water enrichment and biosynthetic fractionation. Accuracy and  
31 precision of respective proxies should be ideally evaluated at a regional scale. For this study,  
32 we analysed topsoils below coniferous and deciduous forests, as well as grassland soils along a  
33 Central European transect in order to investigate the variability and robustness of various  
34 proxies, and to identify effects related to vegetation. Soil pH-values derived from brGDGTs  
35 correlate reasonably well with measured soil pH-values, but systematically overestimate them  
36 ( $\Delta\text{pH} = 0.6 \pm 0.6$ ). The branched vs. isoprenoid tetraether index (BIT) can give some indication  
37 whether the pH reconstruction is reliable. Temperatures derived from brGDGTs overestimate  
38 mean annual air temperatures slightly ( $\Delta T_{\text{MA}} = 0.5^\circ\text{C} \pm 2.4$ ). Apparent isotopic fractionation ( $\varepsilon_{n\text{-alkane/precipitation}}$  and  $\varepsilon_{\text{sugar/precipitation}}$ ) is lower for grassland sites than for forest sites due to “signal  
39 damping”, i.e. grass biomarkers do not record the full evapotranspirative leaf water enrichment.  
40 Coupling  $\delta^2\text{H}_{n\text{-alkane}}$  with  $\delta^{18}\text{O}_{\text{sugar}}$  allows to reconstruct the stable isotopic composition of the  
41 source water more accurately than without the coupled approach ( $\Delta\delta^2\text{H} = \sim-21\text{‰} \pm 22$  and  
42  $\Delta\delta^{18}\text{O} = \sim-2.9\text{‰} \pm 2.8$ ). Similarly, relative humidity during daytime and vegetation period  
43 ( $\text{RH}_{\text{MDV}}$ ) can be reconstructed using the coupled isotope approach ( $\Delta\text{RH}_{\text{MDV}} = \sim-17 \pm 12$ ).  
44 Especially for coniferous sites, reconstructed  $\text{RH}_{\text{MDV}}$  values as well as source water isotope  
45 composition underestimate the measured values. This can be likely explained by understory  
46 grass vegetation at the coniferous sites contributing significantly to the *n*-alkane pool but only  
47 marginally to the sugar pool in the topsoil. The large uncertainty likely reflect the fact that  
48 biosynthetic fractionation is not constant, as well as microclimate variability. Overall, GDGTs  
49 and the coupled  $\delta^2\text{H}_{n\text{-alkane}}\text{-}\delta^{18}\text{O}_{\text{sugar}}$  approach have great potential for more quantitative  
50 paleoclimate reconstructions.  
51



## 52 1 Introduction

53 Information about the variability and consequences of past climate changes is a prerequisite for  
54 precise predictions regarding the present climate change. Molecular fossils, so called  
55 biomarkers, climate proxies have great potential to enhance our understanding about variations  
56 of past climate and environmental changes. Lipid biomarkers in particular, are increasingly  
57 used for paleoclimate and environmental reconstructions (e.g. Brincat et al., 2000; Eglinton and  
58 Eglinton, 2008; Rach et al., 2014; Romero-Viana et al., 2012; Schreuder et al., 2016). However  
59 strengths and limitations of respective proxies need known (Dang et al., 2016). For this,  
60 calibrations using modern reference samples are essential.

61 Terrestrial branched glycerol dialkyl glycerol tetraethers (brGDGTs) that are synthesized in the  
62 cell membranes of anaerobic heterotrophic soil bacteria (Oppermann et al., 2010; Weijers et al.,  
63 2010) have great potential for the reconstruction of past environmental conditions (e.g. Coffinet  
64 et al., 2017; Schreuder et al., 2016; Zech et al., 2012), although some uncertainties exist.  
65 Calibration studies suggest that the relative abundance of the individual brGDGTs varies with  
66 mean annual air temperature ( $T_{MA}$ ) and soil pH (Peterse et al., 2012; Weijers et al., 2007), at  
67 least across large, global climate gradients or along pronounced altitudinal gradients (Wang et  
68 al., 2017). However, in arid regions the production of brGDGT is limited, while isoprenoidal  
69 GDGTs (iGDGTs) produced by archaea provide the dominant part of the overall soil GDGT  
70 pool (Anderson et al., 2014; Dang et al., 2016; Dirghangi et al., 2013; Wang et al., 2013; Xie  
71 et al., 2012). The ratio of brGDGTs vs. isoprenoid GDGTs (BIT) can be used as indication  
72 whether a reconstruction of  $T_{MA}$  and pH will be reliable. Moreover, Mueller-Niggemann et al.  
73 (2016) revealed an influence of the vegetation cover on the brGDGT producing soil microbes.  
74 From field experiments, it is known, that vegetation type and mulching practice strongly effect  
75 soil temperature and moisture (Awe et al., 2015; Liu et al., 2014). Thus, multiple factors can be  
76 expected to influence soil microbial communities and GDGT production. So far, little is known  
77 about the variability of GDGT proxies on a regional scale, and a calibration study with small  
78 climate gradient but with different vegetation types might be useful.

79 Compound specific stable hydrogen isotopes of leaf wax biomarkers, such as long chain *n*-  
80 alkanes ( $\delta^2H_{n\text{-alkanes}}$ ) record the isotopic signal of precipitation and therefore past climate and  
81 environmental conditions (Sachse et al., 2004, 2006). However, various influencing factors are  
82 known all along the way from the moisture source to leaf waxes (Pedentchouk and Zhou, 2018  
83 and Sachse et al., 2012 for review). One is the evapotranspiration of leaf water (Feehans and  
84 Sessions, 2010; Kahmen et al., 2013; Zech et al., 2015), which is strongly driven by relative air  
85 humidity (RH; e.g. Cernusak et al., 2016 for review). In addition, a strong precipitation signal  
86 is known to be incorporated into long chain leaf waxes (Hou et al., 2008; Rao et al., 2009;  
87 Sachse et al., 2004). In paleoclimate studies, it is often not feasible to disentangle between the  
88 evapotranspirative enrichment from the precipitation signal. Zech et al. (2013) proposed to  
89 couple  $\delta^2H_{n\text{-alkane}}$  results with oxygen stable isotopes of hemicellulose-derived sugars ( $\delta^{18}O_{\text{sugar}}$ ).  
90 Assuming constant biosynthetic fractionation factors ( $\varepsilon_{\text{bio}}$ ) for the different compound classes  
91 (*n*-alkanes and hemicellulose sugars), the coupling enables the reconstruction of the isotopic  
92 composition of leaf water, RH and  $\delta^2H/\delta^{18}O$  of plant source water ( $\approx \delta^2H/\delta^{18}O$  of precipitation;  
93 Tuthorn et al., 2015). So far, a detailed evaluation of this approach on the European scale, as  
94 well as concerning possible effects related to vegetation changes is missing.



95 We analysed topsoil samples under coniferous, deciduous and grassland vegetation along a  
96 Central European transect in order to estimate the variability of the biomarker proxies. More  
97 specifically, we aim to test whether:

98 (i) the vegetation type has an influence on the brGDGT proxies, the  $\delta^2\text{H}_{n\text{-alkane}}$  and the  $\delta^{18}\text{O}_{\text{sugar}}$   
99 stable isotopic composition, as well as on reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  and RH.  
100 (ii) the published brGDGT proxies used for reconstructing mean annual temperature and soil  
101 pH are sensitive enough to reflect the medium changes in temperature and soil pH along our  
102 transect.  
103 (iii) the coupled  $\delta^2\text{H}_{n\text{-alkane}}-\delta^{18}\text{O}_{\text{sugar}}$  approach faithfully reflects  $\delta^2\text{H}/\delta^{18}\text{O}$  of precipitation and  
104 RH along the transect.

105

## 106 **2 Material and methods**

### 107 **2.1 Geographical setting and sampling**

108 In November 2012, we collected topsoil samples (0-5 cm depth) at 16 locations along a transect  
109 from Southern Germany to Southern Sweden (Fig. 1A) and distinguished between sites with  
110 coniferous forest (con, n = 9), deciduous forest (dec, n = 14) and grassland (grass, n = 6)  
111 vegetation cover (for more details see Schäfer et al. (2016) and Tab. S1).

112

### 113 **2.2 Database of instrumental climate variables and isotope composition of precipitation**

114 Climate data was derived from close-by weather observation stations operating by the regional  
115 institutions (Deutscher Wetterdienst (DWD) for Germany, Danmarks Meteorologiske Institut  
116 (DMI) for Denmark and the Sveriges Meteorologiska och Hydrologiska Institute (SMHI) for  
117 Sweden). The DWD provides hourly data for each station (DWD Climate Data Center, 2018b),  
118 enabling not only the calculation of  $T_{\text{MA}}$ , but also of the mean annual relative air humidity  
119 ( $\text{RH}_{\text{MA}}$ ), mean temperature and relative air humidity during the vegetation period ( $T/\text{RH}_{\text{MV}}$ ),  
120 and of daytime temperature and relative humidity averages over the vegetation period  
121 ( $T/\text{RH}_{\text{MDV}}$ ). In addition, annual precipitation observations were used to derive the mean annual  
122 precipitation amount ( $P_{\text{MA}}$ ; DWD Climate Data Center, 2018b). From the DMI, the respective  
123 climate variables were derived from published technical reports (Cappelen, 2002; Frich et al.,  
124 1997; Laursen et al., 1999). The SMHI provides open data from which we derived the climate  
125 variables for the Swedish sites (Swedish Meteorological and Hydrological Institute, 2018). For  
126 more details about the climate database used for calculations and comparisons, the reader is  
127 referred to Tab. S2.

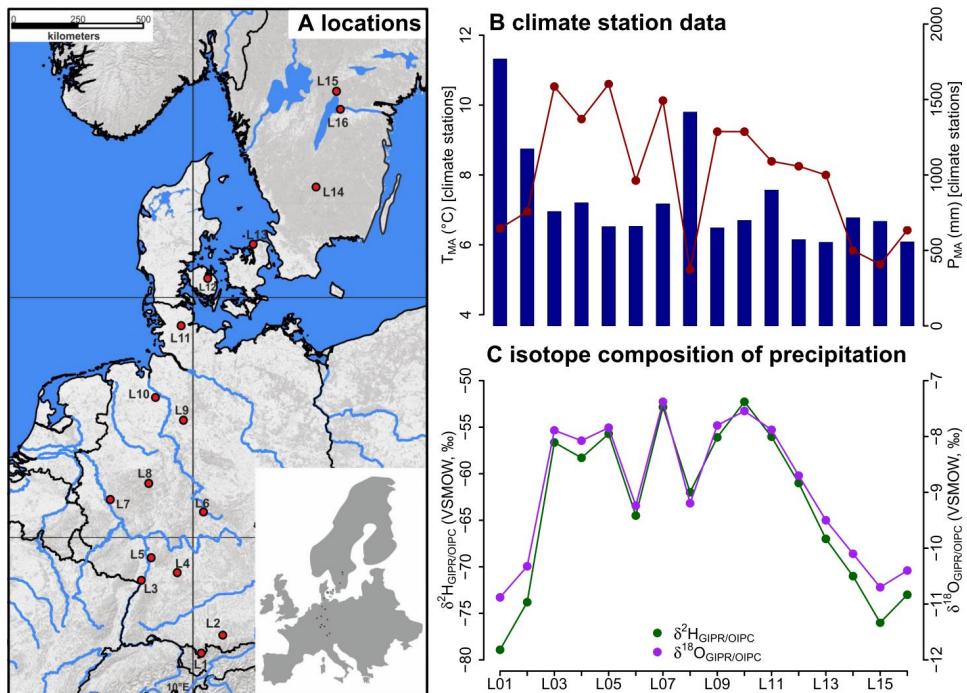
128 For comprising German precipitation  $\delta^2\text{H}/\delta^{18}\text{O}$  along the transect, we realized a regionalisation  
129 (called  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{GIPR}}$ ) using online available data from 34 German GNIP stations, 4 Austrian  
130 ANIP stations and the Groningen GNIP station (van Geldern et al., 2014; IAEA/WMO, 2018;  
131 Stumpf et al., 2014; Umweltbundesamt GmbH, 2018), following the approach of Schlotter  
132 (2007). However, instead of the multivariate regression procedure applied by Schlotter (2007),  
133 we used a random forest approach (Hothorn et al., 2006; Strobl et al., 2007, 2008) to describe  
134 the relationship of squared latitude, latitude, longitude and altitude vs. long term weighted  
135 means of precipitation  $\delta^2\text{H}/\delta^{18}\text{O}$ , and realized the prediction for the study sites. For the Danish



136 and Swedish sites, such a procedure was not possible. Hence, the annual precipitation  $\delta^2\text{H}/\delta^{18}\text{O}$   
137 values were derived from the Online Isotopes in Precipitation Calculator (OIPC, version 3.1),  
138 therefore called  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{OIPC}}$  (Bowen, 2018; Bowen and Revenaugh, 2003; IAEA/WMO,  
139 2015). The finally used  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  data are given in Tab. S1.

140 The  $T_{\text{MA}}$  along the transect ranges from 5.3 to 10.6°C, and  $P_{\text{MA}}$  ranges from 554 to 1769 mm  
141 (Fig. 1B). Precipitation  $\delta^2\text{H}/\delta^{18}\text{O}$  shows moderate changes along the transect,  $\delta^2\text{H}_{\text{GIPR/OIPC}}$   
142 varies between -52 and -79‰, and  $\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  ranges from -7.4 to -10.9‰ (Fig. 1C).

143 Correlations between  $\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  and  $P_{\text{MA}}$ , altitude of the locations,  $T_{\text{MA}}$  are given in the  
144 supplementary material (Fig. S1 to S3), along with a  $\delta^2\text{H}_{\text{GIPR/OIPC}}$  vs.  $\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  scatter plot  
145 (Fig. S4).



146  
147 **Fig. 1.** (A) Sample locations (red dots, map source: US National Park Service), (B) variations  
148 of mean annual air temperature ( $T_{\text{MA}}$ ) and mean annual precipitation ( $P_{\text{MA}}$ ) derived from close-  
149 by climate station data, and (C) hydrogen and oxygen stable isotope composition of  
150 precipitation ( $\delta^2\text{H}_{\text{GIPR/OIPC}}$  and  $\delta^{18}\text{O}_{\text{GIPR/OIPC}}$ , respectively) as derived for the sampled transect  
151 locations (see section 2.2 GIPR  $\delta^2\text{H}/\delta^{18}\text{O}$  generation procedure). The reader is referred to  
152 section 2.2 (and Tab. S1 and S2) for database and reference information of data plotted in (B)  
153 and (C).

154

### 155 2.3 Soil extractions and analysis

#### 156 2.3.1 GDGTs and pH

157 A detailed description of sample preparation for lipid analysis can be found in Schäfer et al.  
158 (2016). Briefly, 1–6 g freeze-dried and grounded soil sample was microwave extracted with 15



159 ml dichloromethane (DCM)/methanol (MeOH) 9:1 (*v:v*) at 100°C for 1 h. Extracts were  
160 separated over aminopropyl silica gel (Supelco, 45 µm) pipette columns. The nonpolar fraction  
161 (including *n*-alkanes) was eluted with hexane and further purified over AgNO<sub>3</sub> coated silica  
162 pipette columns (Supelco, 60-200 mesh) and zeolite (Geokleen Ltd.). The GDGT-containing  
163 fraction was eluted with DCM:MeOH 1:1 (*v:v*), re-dissolved in hexane/isopropanol (IPA) 99:1  
164 (*v:v*) and transferred over 0.45 µm PTFE filters into 300 µl inserts. For quantification, a known  
165 amount of a C<sub>46</sub> diol standard was added after transfer. The samples were analysed at ETH  
166 Zurich using an Agilent 1260 Infinity series HPLC-atmospheric chemical pressure ionization  
167 mass spectrometer (HPLC-APCI-MS) equipped with a Grace Prevail Cyano column (150 mm  
168 × 2.1 mm; 3 µm). The GDGTs were eluted isocratically with 90% A and 10% B for 5 min and  
169 then with a linear gradient to 18% B for 34 min at 0.2 ml min<sup>-1</sup>, where A=hexane and  
170 B=hexane/isopropanol (9:1, *v:v*). Injection volume was 10 µl and single ion monitoring of  
171 [M+H]<sup>+</sup> was used to detect GDGTs.

172 The pH of the samples was measured in the laboratory of the Soil Biogeochemistry group,  
173 Institute of Agronomy and Nutritional Sciences, Martin-Luther-University Halle-Wittenberg,  
174 in a 1:3 soil:water (*w/v*) mixture.

175

### 176 2.3.2 $\delta^{2}\text{H}_{n\text{-alkane}}$

177 The hydrogen isotopic composition of the highest concentrated *n*-alkanes (*n*-C<sub>25</sub>, *n*-C<sub>27</sub>, *n*-C<sub>29</sub>,  
178 *n*-C<sub>31</sub>, and *n*-C<sub>33</sub>) was determined using a TRACE GC Ultra Gas Chromatography connected to  
179 a Delta V Plus Isotope Ratio Mass Spectrometer via a <sup>2</sup>H pyrolysis reactor (GC-<sup>2</sup>H-Py-IRMS;  
180 Thermo Scientific, Bremen, Germany) at the ETH Zurich. The compound-specific <sup>2</sup>H/<sup>1</sup>H ratios  
181 were calibrated against an external standard with C<sub>15</sub> – C<sub>35</sub> homologues. External standard  
182 mixtures (A4 mix from A. Schimmelmann, University of Indiana) were run between the  
183 samples for multipoint linear normalization. The H<sub>3</sub><sup>+</sup> factor was determined on each  
184 measurement day and was constant throughout the periods of the sample batches. Samples were  
185 analysed in duplicates, and results typically agreed within 4% (average difference = 1.4%). All  
186  $\delta^{2}\text{H}$  values are expressed relative to the Vienna Standard Mean Ocean Water (V-SMOW).

187

### 188 2.3.3 $\delta^{18}\text{O}_{\text{sugar}}$

189 Hemicellulose sugars were extracted and purified using a slightly modified standard procedure  
190 (Amelung et al., 1996; Guggenberger et al., 1994; Zech and Glaser, 2009). Briefly, myoinositol  
191 was added to the samples prior to extraction as first internal standard. The sugars were released  
192 hydrolytically using 4M trifluoroacetic acid for 4 h at 105°C, cleaned over glass fiber filters and  
193 further purified using XAD and Dowex columns. Before derivatization with methylboronic acid  
194 (Knapp, 1979), the samples were frozen, freeze-dried, and 3-O-methylglucose in dry pyridine  
195 was added as second internal standard. Compound-specific hemicellulose sugar <sup>18</sup>O  
196 measurements were performed in the laboratory of the Soil Biogeochemistry group, Institute of  
197 Agronomy and Nutritional Sciences, Martin-Luther-University Halle-Wittenberg, using GC-  
198 <sup>18</sup>O-Py-IRMS (all devices from Thermo Fisher Scientific, Bremen, Germany). Standard  
199 deviations of the triplicate measurements were 1.4‰ (over 29 investigated samples) for  
200 arabinose and xylose, respectively. We focus on these two hemicellulose-derived neutral sugars



201 arabinose and xylose as they strongly predominate over fucose in terrestrial plants, soils and  
202 sediments (Hepp et al., 2016 and references therein). Rhamnose concentrations were too low to  
203 obtain reliable  $\delta^{18}\text{O}$  results. All  $\delta^{18}\text{O}$  values are expressed relative to the Vienna Standard Mean  
204 Ocean Water (V-SMOW).

205

## 206 **2.4 Theory and Calculations**

207 2.4.1 Calculations used for the GDGT-based reconstructions

208 The branched and isoprenoid tetraether (BIT) index is calculated according to Hopmans et al.  
209 (2004), for structures see Fig. S5:

$$210 \text{BIT} = \frac{\text{Ia+IIa+IIIa}}{\text{Ia+IIa+IIIa+crenarchaeol}}. \quad (1)$$

211 The cyclopentane moiety number of brGDGTs correlates negatively with soil pH (Weijers et  
212 al., 2007), which led to the development of the cyclization of branched tetraethers (CBT) ratio.  
213 CBT and the CBT based pH ( $\text{pH}_{\text{CBT}}$ ) were calculated according to Peterse et al. (2012):

$$214 \text{CBT} = -\log \frac{\text{Ib+IIb}}{\text{Ia+IIa}}, \quad (2)$$

$$215 \text{pH}_{\text{CBT}} = 7.9 - 1.97 \times \text{CBT}. \quad (3)$$

216 The number of methyl groups in brGDGTs correlates negatively with  $T_{\text{MA}}$  and soil pH (Peterse  
217 et al., 2012; Weijers et al., 2007). Thus, the ratio of the methylation of branched tetraethers  
218 (MBT) ratio and the CBT ratio can be used to reconstruct  $T_{\text{MA}}$ . We use the equation given by  
219 Peterse et al. (2012):

$$220 \text{MBT}' = \frac{\text{Ia+Ib+Ic}}{\text{Ia+Ib+Ic+IIa+IIb+IIc+IIIa}}, \quad (4)$$

$$221 \text{T}_{\text{MA}} = 0.81 - 5.67 \times \text{CBT} + 31.0 \times \text{MBT}'. \quad (5)$$

222

223 2.4.2 Calculations and concepts used for the coupled  $\delta^2\text{H}$ - $\delta^{18}\text{O}$  approach

224 The apparent fractionation is calculated according to Cernusak et al. (2016):

$$225 \varepsilon_{\text{n-alkane/precipitation}} = \left( \frac{\delta^2\text{H}_{\text{n-alkane}} - \delta^2\text{H}_{\text{GIPR/OIPC}}}{1 + \delta^2\text{H}_{\text{GIPR/OIPC}}/1000} \right), \quad (6)$$

$$226 \varepsilon_{\text{sugar/precipitation}} = \left( \frac{\delta^{18}\text{O}_{\text{sugar}} - \delta^{18}\text{O}_{\text{GIPR/OIPC}}}{1 + \delta^{18}\text{O}_{\text{GIPR/OIPC}}/1000} \right). \quad (7)$$

227 The isotopic composition of leaf water ( $\delta^2\text{H}/\delta^{18}\text{O}_{\text{leaf water}}$ ) can be calculated using  $\varepsilon_{\text{bio}}$  for  $\delta^2\text{H}_{\text{n-alkane}}$  (-160‰, Sachse et al., 2012; Sessions et al., 1999) and  $\delta^{18}\text{O}_{\text{sugar}}$  (+27‰, Cernusak et al., 2003; Schmidt et al., 2001):

$$230 \delta^2\text{H}_{\text{leaf water}} = \left( \frac{1000 + \delta^2\text{H}_{\text{n-alkane}}}{1000 + \varepsilon_{\text{bio}} (\text{n-alkane})} \right) \times 10^3 - 1000, \quad (8)$$

$$231 \delta^{18}\text{O}_{\text{leaf water}} = \left( \frac{1000 + \delta^{18}\text{O}_{\text{sugar}}}{1000 + \varepsilon_{\text{bio}} (\text{sugar})} \right) \times 10^3 - 1000. \quad (9)$$

232 Zech et al. (2013) introduced the conceptual model for the coupled  $\delta^2\text{H}_{\text{n-alkane}}$ - $\delta^{18}\text{O}_{\text{sugar}}$  approach  
233 in detail. Briefly, the coupled approach is based on the following assumptions (illustrated in  
234 Fig. 8): (i) The isotopic composition of precipitation, which is set to be equal to the plant source  
235 water, typically plots along the global meteoric water line (GMWL;  $\delta^2\text{H} = 8 \times \delta^{18}\text{O} + 10$ ) in a



236  $\delta^{18}\text{O}$  vs.  $\delta^2\text{H}$  space (Craig, 1961); (ii) Source water uptake by plants does not lead to any  
237 fractionation (e.g. Dawson et al., 2002), and significant evaporation of soil water can be  
238 excluded; (iii) Evapotranspiration leads to enrichment of the remaining leaf water along the  
239 local evaporation line (LEL; Allison et al., 1985; Bariac et al., 1994; Walker and Brunel, 1990),  
240 compared to the source water taken up by the plant; (iv) The biosynthetic fractionation is  
241 assumed to be constant. In addition, isotopic equilibrium between plant source water (~  
242 weighted mean annual precipitation) and the local atmospheric water vapour is assumed.  
243 Further assumption concerns the isotope steady-state in the evaporating leaf water reservoir.  
244 The coupled approach allows for reconstructing the isotopic composition of plant source water  
245 ( $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$ ) from the reconstructed leaf water, by calculating the intercepts of the LELs  
246 with the GMWL (Zech et al., 2013). The slope of the LEL ( $S_{\text{LEL}}$ ) can be assessed by the  
247 following equation (Gat, 1971):

$$248 \quad S_{\text{LEL}} = \frac{\varepsilon_2^* + C_k^2}{\varepsilon_{18}^* + C_k^{18}}, \quad (10)$$

249 where  $\varepsilon^*$  are equilibrium isotope fractionation factors and  $C_k$  are kinetic fractionation factors.  
250 The latter equals to 25.1‰ and 28.5‰, for  $C_k^2$  and  $C_k^{18}$ , respectively (Merlivat, 1978). The  
251 equilibrium fractionation factors can be derived from empirical equations (Horita and  
252 Wesolowski, 1994) by using  $T_{\text{MDV}}$  values. For two Danish sites  $T_{\text{MDV}}$  are not available, instead  
253  $T_{\text{MV}}$  is used here (section 2.2 and Tab. S2).

254 In a  $\delta^{18}\text{O}$ - $\delta^2\text{H}$  diagram, the distance of the leaf water from the GMWL define the deuterium-  
255 excess of leaf water ( $d_{\text{leaf-water}} = \delta^2\text{H}_{\text{leaf-water}} - 8 \times \delta^{18}\text{O}_{\text{leaf-water}}$ , according Dansgaard, (1964); Fig.  
256 8). To convert  $d_{\text{leaf-water}}$  into mean RH during daytime and vegetation period ( $RH_{\text{MDV}}$ ), a  
257 simplified Craig-Gordon model can be applied (Zech et al., 2013):

$$258 \quad RH = 1 - \frac{\Delta d}{\varepsilon_2^* - 8 \times \varepsilon_{18}^* + C_k^2 - 8 \times C_k^{18}}, \quad (11)$$

259 where  $\Delta d$  is the difference in  $d_{\text{leaf-water}}$  and the deuterium-excess of source water ( $d_{\text{source-water}}$ ).

260

## 261 **2.5 Statistics**

262 In the statistical analysis we checked sample distributions for normality (Shapiro and Wilk,  
263 1965) and for equal variance (Levene, 1960). If normality and equal variances are given, we  
264 perform an Analysis of Variance (ANOVA). If that is not the case, we conduct the non-  
265 parametric Kruskal-Wallis Test. ANOVA or Kruskal-Wallis are used to find significant  
266 differences ( $\alpha=0.05$ ) between the vegetation types (deciduous, conifer and grass).

267 In order to describe the relation along a 1:1 line, the coefficient of correlation ( $R^2$ ) was  
268 calculated as  $R^2 = 1 - \sum (\text{modeled} - \text{measured})^2 / \sum (\text{measured} - \text{measured mean})^2$ . The small  
269  $r^2$  is taken as coefficient of correlation of a linear regression between a dependent (y) and  
270 explanatory variable(s). The root mean square error (RMSE) of the relationships was calculated

271 as  $RMSE = \sqrt{\left( \frac{1}{n} \sum (\text{modeled} - \text{measured})^2 \right)}$ . All data plotting and statistical analysis was  
272 realized in R (version 3.2.2; R Core Team, 2015).

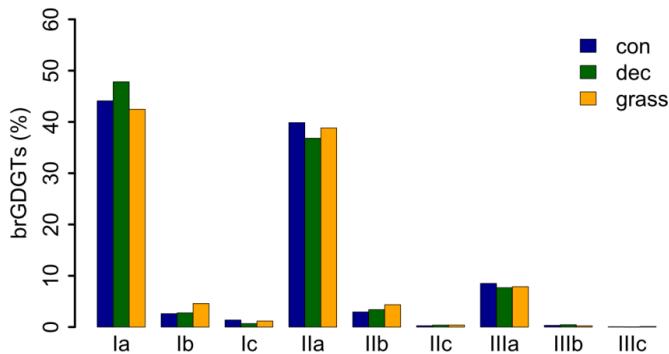
273



## 274 3 Results and Discussion

### 275 3.1 GDGT concentrations

276 GDGT Ia has the highest concentration under all vegetation types, followed by GDGT IIa and  
277 GDGT IIIa (Fig. 2). GDGT Ib, IIb and Ic occur in minor, GDGT IIc and IIIb only in trace  
278 amounts. GDGT IIIc was below the detection limit in most of the samples (Tab. S3). Although  
279 other studies document an influence of the vegetation cover on soil temperature and soil water  
280 content, which control the microbial community composition in soils (Awe et al., 2015; Liu et  
281 al., 2014; Mueller-Niggemann et al., 2016), we find no statistically different pattern of the  
282 individual brGDGTs.



283  
284 **Fig. 2.** Mean concentrations of individual brGDGTs as percentage of all brGDGTs for the three  
285 investigated types. Abbreviations: con = coniferous forest sites (n=9); dec = deciduous forest  
286 sites (n=14); grass = grassland sites (n=6).

287 Total concentrations of brGDGTs range from 0.32 to 9.17 µg/g dry weight and tend to be  
288 highest for the coniferous samples and lowest for the grasses (Fig. 3A, Tab. S3). Bulk brGDGT  
289 concentrations lie within ranges of other studies examining soils of mid latitude regions (Huguet  
290 et al., 2010b, 2010a; Weijers et al., 2011). Similar concentrations in coniferous and deciduous  
291 samples imply that brGDGT production does not strongly vary in soils below different forest  
292 types. The grass samples show lower brGDGT concentrations compared to the forest samples,  
293 but this is probably mainly due to ploughing of the grass sites and hence admixing of mineral  
294 subsoil material. Anyhow, the differences in brGDGT concentrations are not significant (p-  
295 value = 0.06).

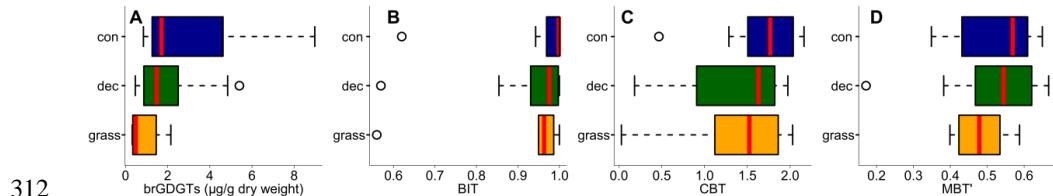
296

### 297 3.2 BIT index

298 Most of the samples have a BIT index higher than 0.9 (Fig 3B and Tab. S3). The BIT-values  
299 are typical for soils in humid and temperate climate regions (Weijers et al., 2006). However,  
300 outliers exist. The most likely source of iGDGTs in soils are Thaumarchaeota, i.e. aerobic  
301 ammonia oxidizing archaea producing Crenarchaeol and its regiosomer (Schouten et al., 2013  
302 and references therein), precipitation amounts drop below 700-800 mm (Dang et al., 2016;  
303 Dirghangi et al., 2013). The P<sub>MA</sub> data of our sampling sites mostly show precipitation > 550  
304 mm (Fig. 1B), but one has to be aware that this data is based on the climate station nearest to  
305 the respective sampling locations and microclimate effects, such as sunlight exposure, canopy



306 cover or exposition might have a pronounced influence on the brGDGT vs. iGDGT distribution.  
307 Mueller-Niggemann et al. (2016) found higher BIT indices in upland soils compared to paddy  
308 soils and stated that the management type also influences BIT values in soils. Along our  
309 transect, grass sites tend to have slightly lower BIT-values than forest sites, probably due to the  
310 absence of a litter layer and hence, no isolation mechanism preventing evaporation of soil water.  
311 Anyhow, differences between vegetation types are not significant (p-value = 0.32).



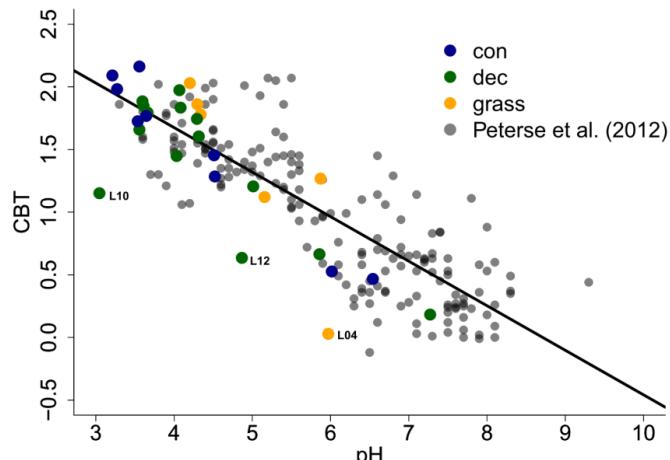
312 **Fig. 3.** (A) Total concentrations of brGDGTs in  $\mu\text{g g}^{-1}$  dry weight, as well as (B) BIT, (C) CBT  
313 and (D) MBT'. Abbreviations: con = coniferous forest sites (n=9); dec = deciduous forest sites  
314 (n=14); grass = grassland sites (n=6). Box plots show median (red line), interquartile range  
315 (IQR) with upper (75%) and lower (25%) quartiles, lowest whisker still within 1.5IQR of lower  
316 quartile, and highest whisker still within 1.5IQR of upper quartile, dots mark outliers.

318

### 319 **3.3 CBT-derived pH**

320 The CBT ratio shows a pronounced variation independent of vegetation type with values  
321 between 0.03 and 2.16 (Fig 3C). The coniferous samples tend to be highest, but the differences  
322 between vegetation types are not significant (p-value = 0.48). The CBT index can be related to  
323 pH in acidic and/or humid soils (e.g. Dirghangi et al., 2013; Mueller-Niggemann et al., 2016;  
324 Peterse et al., 2012; Weijers et al., 2007) but might be an indicator of soil water content and  
325 hence, precipitation in more arid and alkaline soils (e.g. Dang et al., 2016). There is a  
326 pronounced correlation between CBT and soil pH (Fig. 4), which is in good agreement with  
327 other studies from mid latitude regions where precipitation is relatively high (Anderson et al.,  
328 2014 and references therein). Moreover, the CBT to pH relationship in terms of slope and  
329 intersect in our dataset ( $\text{CBT} = -0.47 \times \text{pH} + 3.5$ ,  $r^2 = 0.7$ , p-value < 0.0001, n = 29) is well  
330 comparable to the correlation described for the global calibration dataset of Peterse et al. (2012)  
331 ( $\text{CBT} = -0.36 \times \text{pH} + 3.1$ ,  $r^2 = 0.7$ , p-value < 0.0001, n = 176).

332 However, there are some outliers in the CBT-pH correlation, which need a further examination  
333 (see locations grass L04, dec L10 and dec L12 as marked in Figs. 4 and 5). The outliers show  
334 lower BIT indices (< 0.85, Tab. S3). Even though the data from the nearest climate station  
335 suggest no abnormal P<sub>MA</sub>. Local effects such as differences in the amount of sunlight exposure,  
336 nutrient availability for brGDGT producing organisms or, most likely soil water content might  
337 influence the brGDGT production at these locations (Anderson et al., 2014; Dang et al., 2016).  
338 A lower BIT index as well as a lower CBT occur when soil water content decreases (Dang et  
339 al., 2016; Sun et al., 2016) or when aeration is high and less anoxic microhabitats for GDGT  
340 producing microbes exist (e.g. Dirghangi et al., 2013).

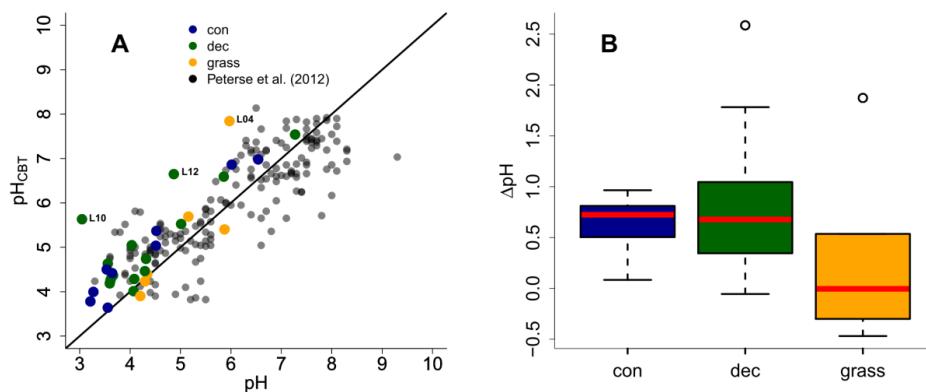


341

342 **Fig. 4.** CBT to pH relationship in our dataset in comparison to the global calibration dataset  
 343 from Peterse et al. (2012) ( $CBT = -0.36 \times pH + 3.1$ ,  $r^2 = 0.7$ ,  $p$ -value < 0.0001,  $n = 176$ , black  
 344 line). Abbreviations: con = coniferous forest sites ( $n=9$ ); dec = deciduous forest sites ( $n=14$ );  
 345 grass = grassland sites ( $n=6$ ).

346

347 As the CBT and pH are similarly correlated in our dataset and the global dataset of Peterse et  
 348 al. (2012), the CBT-derived pH correlated well with the actual pH (Fig. 5A;  $R^2 = 0.3$ ).  
 349 Expressed as  $\Delta pH$  (CBT-derived pH - measured pH), there is a tendency that the GDGTs result  
 350 in an overestimation of the real pH for the forest sites (Fig. B). Yet a Kruskal-Wallis test shows  
 351 no statistically significant difference between the vegetation types, with a  $p$ -value of 0.13. The  
 352 overall  $\Delta pH$  of  $0.6 \pm 0.6$  shows that the reconstruction of soil pH using brGDGTs works well  
 353 along this transect.



354

355 **Fig. 5.** (A) Correlation between measured pH and reconstructed soil pH ( $pH_{CBT}$ ) from our  
 356 transect data in comparison to the global calibration dataset from Peterse et al. (2012) ( $R^2 = 0.7$ ,  
 357 RMSE = 0.75,  $n = 176$ ). Black line indicates the 1:1 relationship. (B) Boxplots of  $\Delta pH$  (refers  
 358 to  $pH_{CBT}$ -pH). Box plots show median (red line), interquartile range (IQR) with upper (75%)

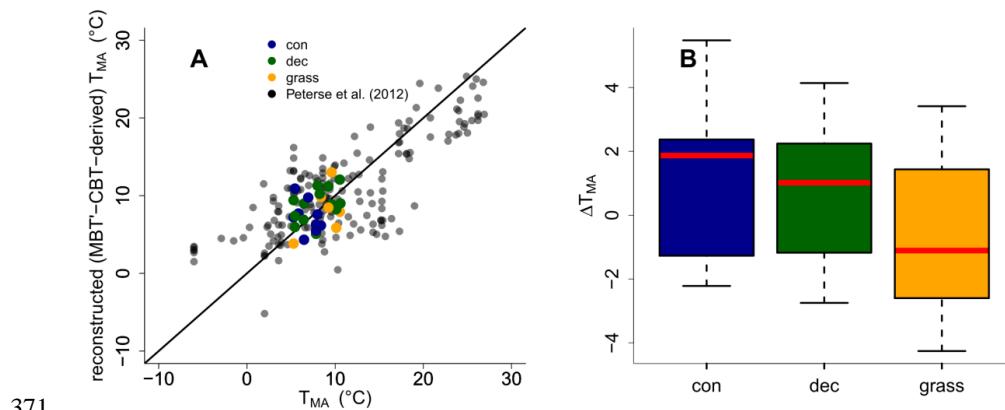


359 and lower (25%) quartiles, lowest whisker still within 1.5IQR of lower quartile, and highest  
360 whisker still within 1.5IQR of upper quartile, dots mark outliers. Abbreviations: con =  
361 coniferous forest sites (n=9); dec = deciduous forest sites (n=14); grass = grassland sites (n=6).

362

### 363 **3.4 MBT'-CBT-derived T<sub>MA</sub> reconstructions**

364 The MBT' shows high variability with values ranging from 0.17 to 0.67 no statistical  
365 differences between vegetation types (p-value = 0.54; Fig. 3D, Tab. S3). When comparing  
366 reconstructed (MBT'-CBT-derived) T<sub>MA</sub> with climate station T<sub>MA</sub>, the data plot close to the 1:1  
367 line, and fit well into the global dataset of Peterse et al. (2012) (Fig. 7A). The  $\Delta T_{MA}$  reveal an  
368 overall offset of  $0.5^{\circ}\text{C} \pm 2.4$  and there is no statistically difference between vegetation types  
369 (Fig. 7B). The standard deviation in  $\Delta T_{MA}$  of  $\pm 2.4$  is well in line with the RMSE of 5.0 for the  
370 global calibration dataset (Peterse et al., 2012).



371  
372 **Fig. 6.** (A) Correlation between climate station T<sub>MA</sub> and reconstructed (MBT'-CBT-derived)  
373 T<sub>MA</sub>. For comparison, the global calibration dataset from Peterse et al. (2012) is shown. The  
374 black line indicates the 1:1 relationship. (B) Boxplots of  $\Delta T_{MA}$  (refers to reconstructed T<sub>MA</sub>-  
375 T<sub>MA</sub> from climate stations) in the different vegetation types from our transect study. Box plots  
376 show median (red line), interquartile range (IQR) with upper (75%) and lower (25%) quartiles,  
377 lowest whisker still within 1.5IQR of lower quartile, and highest whisker still within 1.5IQR of  
378 upper quartile, dots mark outliers. Abbreviations: con = coniferous forest sites (n=9); dec =  
379 deciduous forest sites (n=14); grass = grassland sites (n=6).

380

### 381 **3.5 Apparent fractionation of $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in the different vegetation types**

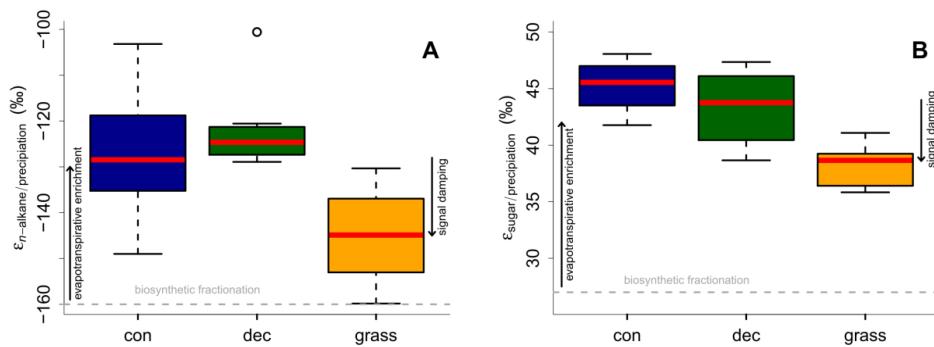
382 The  $\delta^2\text{H}$  values could be obtained for *n*-alkanes C<sub>27</sub>, C<sub>29</sub> and C<sub>31</sub> in all samples and additionally  
383 at two locations for *n*-C<sub>25</sub> and *n*-C<sub>33</sub> at six other locations. The  $\delta^2\text{H}_{n\text{-alkane}}$  values, calculated as  
384 mean of *n*-C<sub>25</sub> to *n*-C<sub>31</sub>  $\delta^2\text{H}$ , ranges from -156 to -216‰. Pooled standard deviations show an  
385 overall average of 3.6‰. The  $\delta^{18}\text{O}_{\text{sugar}}$  values, calculated as the area weighted means for  
386 arabinose and xylose, ranges from 27.7 to 39.4‰. The average weighted mean standard  
387 deviation is 1.4‰. The compound-specific isotope data is summarized along with the  
388 calculations in Tab. S4.



389 Apparent fractionation ( $\epsilon_{n\text{-alkane}/\text{precipitation}}$ ) is on the order of -120 to -150‰, i.e. a bit less than  
 390 the biosynthetic fraction of -160‰. This implies that evapotranspirative enrichment is ~ 10 to  
 391 40‰ (Fig. 7A).  $\epsilon_{n\text{-alkane}/\text{precipitation}}$  is lower for grass sites compared to the forest sites. Differences  
 392 are significant between deciduous and grass sites ( $p\text{-value} = 0.005$ ). This finding supports the  
 393 results of other studies (Kahmen et al., 2013; Liu and Yang, 2008; McInerney et al., 2011), and  
 394 can be named “signal damping”. Grasses do not only incorporate the evaporatively-enriched  
 395 leaf water only but also unenriched leaf water in the growth and differentiation zone of grasses  
 396 (Gamarra et al., 2016; Liu et al., 2017).

397 The grass-derived hemicellulose sugar biomarkers do not fully record the evapotranspirative  
 398 enrichment of the leaf water, either, as indicated by lower apparent fractionation ( $\epsilon_{\text{sugar}/\text{precipitation}}$ )  
 399 in Fig. 7B. The differences are significant between forest and grass sites ( $p\text{-value} < 0.005$ ). This  
 400 is in agreement with a study on cellulose extracted from grass blades (Helliker and Ehleringer,  
 401 2002), and again, the “signal damping” can be explained with incorporation of enriched leaf  
 402 water and non-enriched stem water.

403 Based on the comparison of evapotranspirative enrichment between forest and grass sites, the  
 404 “signal damping” can be quantified to be ~ 31% for the hemicellulose sugars, and ~ 49% for  
 405 the  $n$ -alkanes. This is in agreement with other studies that reported a loss of 22% of the leaf  
 406 water enrichment for hemicellulose sugars (Helliker and Ehleringer, 2002) and 39 to 62% loss  
 407 of the leaf water enrichment for  $n$ -alkanes (Gamarra et al., 2016).



408  
 409 **Fig. 7.** Apparent fractionation (A)  $\epsilon_{n\text{-alkane}/\text{precipitation}}$  and (B)  $\epsilon_{\text{sugar}/\text{precipitation}}$ . Biosynthetic  
 410 fractionation factors according to section 2.4.2. Box plots show median (red line), interquartile  
 411 range (IQR) with upper (75%) and lower (25%) quartiles, lowest whisker still within 1.5IQR  
 412 of lower quartile, and highest whisker still within 1.5IQR of upper quartile, dots mark outliers.  
 413 Abbreviations: con = coniferous forest sites (n=9); dec = deciduous forest sites (n=11 and 14  
 414 for  $n$ -alkanes and sugars, respectively); grass = grassland sites (n=4 and 6 for  $n$ -alkanes and  
 415 sugars, respectively). The figure conceptually illustrates the effect of biosynthetic fractionation  
 416 and evapotranspirative enrichment as well as “signal damping”.

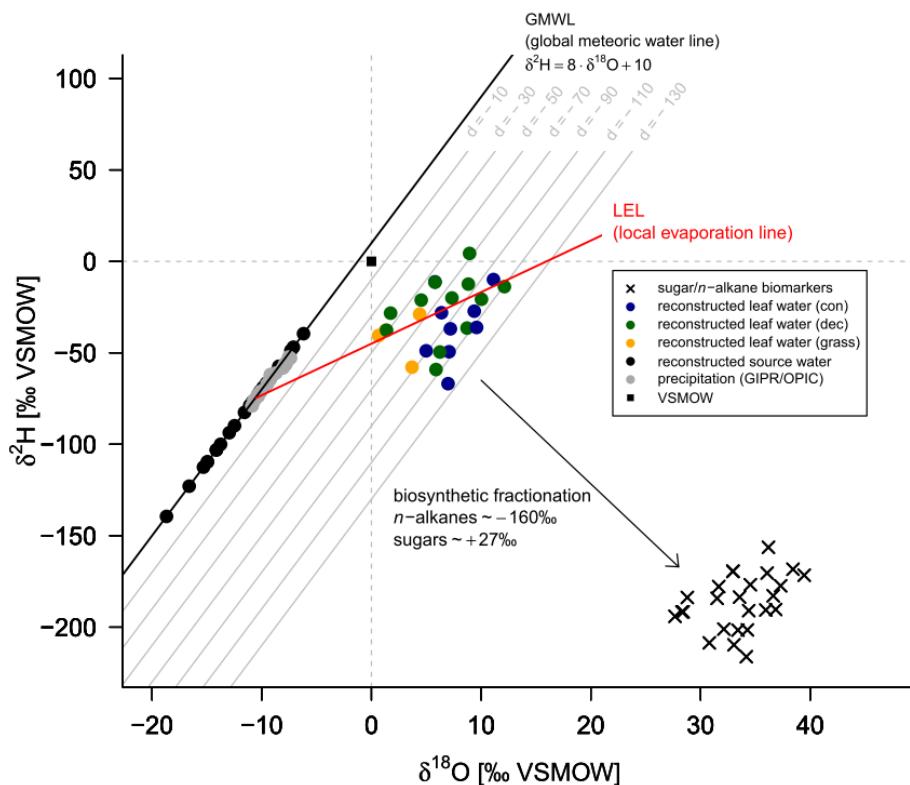
417

### 418 3.6 $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$ reconstructions

419 The  $\delta^2\text{H}$  versus  $\delta^{18}\text{O}$  diagram shown in Fig. 8 graphically illustrates the reconstruction of  
 420  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{leaf-water}}$  (colored dots) from  $\delta^2\text{H}_{n\text{-alkane}}/\delta^{18}\text{O}_{\text{sugar}}$  (crosses), as well as the reconstruction



421 of  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  (black dots). For reconstructing  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$ , LELs with an average  
 422 slope of  $2.8 \pm 0.1$  (Eq. 10) can be generated through every leaf water point and the intercepts of  
 423 these LELs with the GMWL.



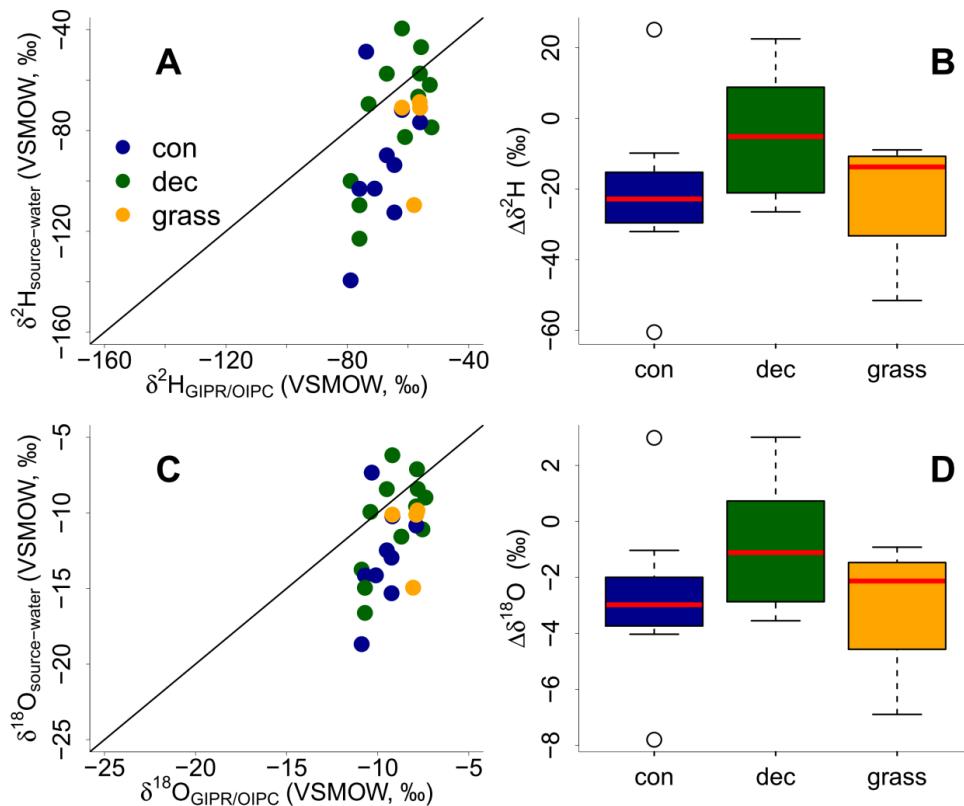
424  
 425 **Fig. 8.**  $\delta^2\text{H}$  vs.  $\delta^{18}\text{O}$  diagram illustrating  $\delta^2\text{H}_{n\text{-alkane}}$  and  $\delta^{18}\text{O}_{\text{sugar}}$ , reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{leaf-water}}$   
 426 (according Eqs. 8 and 9) and reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  in comparison to GIPR/OIPC-  
 427 based  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{precipitation}}$ . Abbreviations: con = coniferous forest sites (n=9); dec = deciduous  
 428 forest sites (n=11); grass = grassland sites (n=4).

429  
 430 The reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  results can be compared with the  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  data  
 431 (Fig. 9). This comparison reveals that the coupled  $\delta^2\text{H}_{n\text{-alkane}}\text{-}\delta^{18}\text{O}_{\text{sugar}}$  approach yields more  
 432 accurate  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  results than hitherto applied  $\delta^2\text{H}_{n\text{-alkane}}$  single isotope approaches.  
 433 However, the range of the reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  values is clearly larger than in  
 434  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  values.  $\delta^2\text{H}$  is systematically underestimated by  $\sim 21\text{‰} \pm 22$  (Fig. 9B) and  
 435  $\delta^{18}\text{O}$  by  $\sim 2.9\text{‰} \pm 2.8$  (Fig. 9D). The type of vegetation seems to be not particularly relevant (p-  
 436 value = 0.18 for  $\Delta\delta^2\text{H}$  and p-value = 0.34 for  $\Delta\delta^{18}\text{O}$ ). Nevertheless, the systematic offsets tend  
 437 to be lowest for the deciduous sites ( $\Delta\delta^2\text{H}/\Delta\delta^{18}\text{O}$  is closer to zero with  $\sim 5\text{‰} \pm 15$  and  $\sim 1.1\text{‰}$   
 438  $\pm 2.1$ ), followed by grass sites ( $\sim 14\text{‰} \pm 20$  and  $\sim 2.1\text{‰} \pm 2.6$ ). In comparison, the coniferous  
 439 sites show the largest offsets ( $\sim 23\text{‰} \pm 26$  for  $\Delta\delta^2\text{H}$   $\sim 3.0\text{‰} \pm 3.3$  for  $\Delta\delta^{18}\text{O}$ ). Differences are,  
 440 however, not statistically significant. The systematic offset and the large variability might have



more specific reasons, and we suggest that this is related to the type of vegetation. Deciduous trees produce lots of leaf waxes and sugars (e.g. Prietzel et al., 2013; Zech et al., 2012a), and all biomarkers reflect and record the evapotranspirative enrichment of the leaf water (e.g. Cernusak et al., 2016; Tuthorn et al., 2014). The coupled approach and the leaf water reconstruction based on the *n*-alkane and sugar biomarkers thus works well. However, coniferous trees produce quite low amounts of *n*-alkanes (Diefendorf and Freimuth, 2016; Zech et al., 2012a), while sugar concentrations are as high as in other vascular plants (e.g. Hepp et al., 2016; Prietzel et al., 2013). For the coniferous soil samples this means that the *n*-alkanes stem most likely from the understory whereas the sugars originate from grasses and coniferous needles. When the understory is dominated by grass species then the *n*-alkane biomarkers do not record the full leaf water enrichment signal, whereas the sugars from the needles do. The reconstructed leaf water for the coniferous sites is therefore too negative concerning  $\delta^2\text{H}$ , and reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  values thus also become too negative (Fig. 8). Concerning the grass sites the following explanation can be found. Correcting for “signal damping” makes the reconstructed leaf water points more positive and shifts them in Fig. 8 up and right. As the “signal damping” is stronger for  $\delta^2\text{H}$  than for  $\delta^{18}\text{O}$  the corrected leaf water points are now above the uncorrected ones. The corrected leaf water points leads to more positive reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  values for the grass sites.

Vegetation type specific rooting depths could partly cause the overall high variability in reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$ . Deep rooting species most likely use the water from deeper soil horizons and/or shallow ground water, which is equal to the (weighted) mean annual precipitation (e.g. Herrmann et al., 1987). Shallow rooting plants take up water from upper soil horizons, which is influenced by seasonal variations in  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{precipitation}}$  and by soil water enrichment (Dubbert et al., 2013). Thus, the overall assumption that the source water of the plants reflects the local (weighted) mean precipitation might be not fully valid for all sites. Moreover, a partly contribution of root-derived rather than leaf-derived sugar biomarkers in our topsoil samples is very likely. This does, by contrast, not apply for *n*-alkanes, which are hardly produced in roots (Zech et al., 2012b and the discussion).



469  
 470 **Fig. 9.** Correlation of reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  vs. precipitation  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{IPR/OIPC}}$  (A and  
 471 C). Black lines indicate 1:1 relationship. Differences between reconstructed source water and  
 472 precipitation ( $\Delta\delta^2\text{H}/\delta^{18}\text{O} = \delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}} - \delta^2\text{H}/\delta^{18}\text{O}_{\text{IPR/OIPC}}$ ) for the three different  
 473 vegetation types (B and D). Box plots show median (red line), interquartile range (IQR) with  
 474 upper (75%) and lower (25%) quartiles, lowest whisker still within 1.5IQR of lower quartile,  
 475 and highest whisker still within 1.5IQR of upper quartile. Abbreviations: con = coniferous  
 476 forest sites (n=9); dec = deciduous forest sites (n=11); grass = grassland sites (n=4).

477 Moreover, the high variability within the vegetation types could be caused by variability in  $\varepsilon_{\text{bio}}$   
 478 of  $^2\text{H}$  in  $n$ -alkanes, as well as  $^{18}\text{O}$  in sugars. There is an ongoing discussion about the correct  
 479  $\varepsilon_{\text{bio}}$  for  $^{18}\text{O}$  in hemicellulose sugars (Sternberg, 2014 vs. Zech et al., 2014), and  $\varepsilon_{\text{bio}}$  is probably  
 480 not constant over all vegetation types. This translates into errors concerning leaf water  
 481 reconstruction and thus for reconstructing  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  values (Eq. 9 and Fig. 8).  
 482 Likewise, the  $\varepsilon_{\text{bio}}$  values reported in the literature for  $^2\text{H}$  of  $n$ -alkanes can be off from -160‰  
 483 by tens of permille (Feakins and Sessions, 2010; Tipple et al., 2015; Feakins et al., 2016;  
 484 Freimuth et al., 2017). The degree to which hydrogen originates from NADPH rather than leaf  
 485 water is important, because NADPH is more negative (Schmidt et al., 2003). The wide range  
 486 in biosynthetic  $^2\text{H}$  fractionation factors is therefore also related to the carbon and energy  
 487 metabolism state of plants (Cormier et al., 2018).

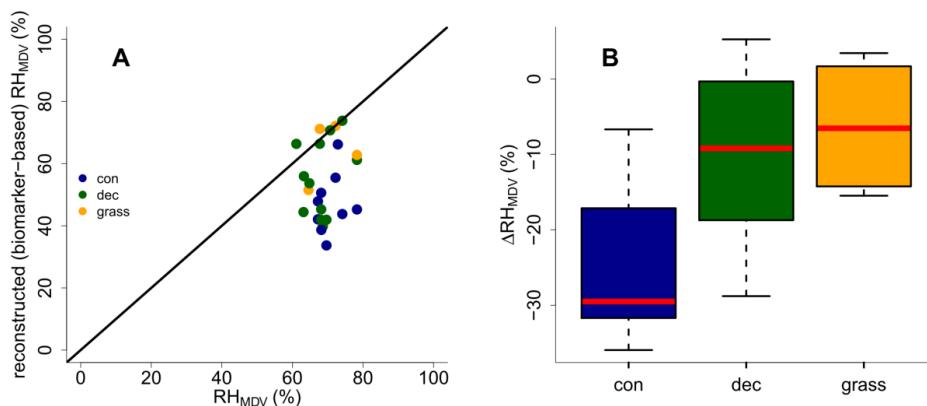


488 **3.7 RH reconstruction**

489 Reconstructed  $\text{RH}_{\text{MDV}}$  ranges from 34 to 74%, while  $\text{RH}_{\text{MDV}}$  from climate station data range  
490 from 61 to 78% (Fig. 10A). Biomarker-based values thus systematically underestimate the  
491 station data ( $\Delta\text{RH}_{\text{MDV}} = -17\% \pm 12$ ; Fig. 10B). Yet the offsets are much less for deciduous tree  
492 and grass sites ( $\Delta\text{RH}_{\text{MDV}} = -10\% \pm 12$  and  $-7\% \pm 9$ , respectively). The offsets for the coniferous  
493 sites are  $-30\% \pm 11$ , and significantly larger than for the deciduous and grass sites (p-values <  
494 0.05).

495 Too low reconstructed  $\text{RH}_{\text{MDV}}$  values for the coniferous sites make sense in view of the  
496 previously discussed option that soils contain *n*-alkanes from the understory (which is  
497 dominated by grass species), while sugars stem from needles and grasses. As explained earlier  
498 already, the “signal damping” leads to too negative reconstructed  $\delta^2\text{H}_{\text{leaf-water}}$  (whereas  $\delta^{18}\text{O}$  is  
499 affected less by the “signal damping”), and too negative  $\delta^2\text{H}_{\text{leaf-water}}$  translates into  
500 overestimated d-excess and underestimated RH values. In Fig. 8, a correction for this require  
501 moving the coniferous leaf water data points upwards towards more positive  $\delta^2\text{H}$  values, thus  
502 the distance between the leaf water and the source water is reduced.

503 The underestimation of RH for the deciduous and grass sites could be partly associated with the  
504 use of the GMWL as baseline for the coupled  $\delta^2\text{H}_{n\text{-alkane}}-\delta^{18}\text{O}_{\text{sugar}}$  approach. The deuterium-  
505 excess of the LMWLs is generally lower than the  $+10\text{‰}$  of the GMWL, while the slopes of the  
506 LMWLs are well comparable to the GMWL (Stumpp et al., 2014). In addition, if soil water  
507 evaporation occurred before water uptake by the plants, this would lead to an underestimation  
508 of biomarker-based  $\text{RH}_{\text{MDV}}$  values. It can be furthermore assumed that plant metabolism is  
509 highest during times with direct sunshine and high irradiation, i.e. during noon at sunny days.  
510 The relevant RH could therefore be lower than the climate station-derived  $\text{RH}_{\text{MDV}}$ . Indeed,  
511 already climate station  $\text{RH}_{\text{MDV}}$  is considerable lower than  $\text{RH}_{\text{MA}}$  and  $\text{RH}_{\text{MV}}$  (Tab. S1).



513 **Fig. 10.** (A) Comparison of reconstructed (biomarker-based)  $\text{RH}_{\text{MDV}}$  values and climate station  
514  $\text{RH}_{\text{MDV}}$  data. The black line indicates the 1:1 relationship. (B) Differences between  
515 reconstructed and climate station  $\text{RH}_{\text{MDV}}$  values ( $\Delta\text{RH}_{\text{MDV}} = \text{reconstructed} - \text{climate station}$   
516  $\text{RH}_{\text{MDV}}$ ) for the three different vegetation types along the transect. Abbreviations: con =  
517 coniferous forest sites (n=9); dec = deciduous forest sites (n=11); grass = grassland sites (n=4).



518 The uncertainty of reconstructed RH<sub>MDV</sub> values are large for all three investigated vegetation  
519 types, and again these uncertainties are probably also related to  $\varepsilon_{\text{bio}}$ , which is most likely not  
520 constant as assumed for our calculations. Moreover, microclimate variability is underestimated  
521 in our approach. As mentioned in sections 2.4.2 and 3.6, in the coupled approach not only the  
522 source water of the plants is equated with (weighted) mean annual precipitation, but also an  
523 isotopic equilibrium between the source water and the (local) atmospheric water vapour is  
524 assumed. However, in areas with distinct seasonality this might be not fully valid. To account  
525 for this lack of equilibrium between precipitation and local atmospheric water vapour, apparent  
526  $\varepsilon$  values can be calculated with data from Jacob and Sonntag, (1991). As shown by Hepp et al.  
527 (2018) those values can be used to achieve alternative RH reconstructions based on the coupled  
528  $\delta^2\text{H}_{n\text{-alkane}}-\delta^{18}\text{O}_{\text{sugar}}$  approach. Such calculated RH<sub>MDV</sub> values are on average 1.5% more  
529 negative than the original values. However, this difference in RH is far below the analytical  
530 uncertainties of the compound-specific biomarker isotope analysis.

531 Finally, the integration time of the investigated topsoils has to be discussed. Unfortunately, no  
532  $^{14}\text{C}$  dates are available for the soil samples. However, most likely the organic matter has been  
533 built up over a longer timescale than the available climate data, which is used for comparison.  
534 In combination with vegetation changes/management changes throughout that period, this  
535 could surely lead to a less tight relationship of the reconstructions compared to the climate  
536 station data. Root input of arabinose and xylose seems to be of minor relevance in our topsoil  
537 samples. Otherwise, the reconstructed  $\delta^{18}\text{O}_{\text{sugar}}$  values would be too negative resulting in  
538 RH<sub>MDV</sub> overestimations, which is not observed.

539

## 540 4 Conclusions

541 We were able to show that

- 542 (i) the vegetation type does not significantly influence the brGDGT concentrations and  
543 proxies, yet the coniferous sites tend to have higher brGDGT concentrations, BIT  
544 indices and CBT/MBT' ratios, while grass sites tend to be lowest.
- 545 (ii) CBT faithfully records soil pH with a median  $\Delta\text{pH}$  of  $0.6 \pm 0.6$ , The CBT  
546 overestimates the real pH particularly at the forest sites.
- 547 (iii) CBT-MBT'-derived  $T_{\text{MA}}$  reflect the climate station-derived  $T_{\text{MA}}$  values with a  
548 median  $\Delta T_{\text{MA}}$  of  $0.5^{\circ}\text{C} \pm 2.4$ , but again slightly too high reconstruction for the forest  
549 sites were observed.
- 550 (iv) differences in the apparent fractionation between the investigated vegetation types  
551 are caused by “signal damping”, i.e. the grasses do not see and record the full  
552 evaporative enrichment of leaf water.
- 553 (v) the reconstructed  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  reflects the  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{GIPR/OIPC}}$  with a systematic  
554 offset for  $\delta^2\text{H}$  of  $\sim -21\text{‰} \pm 22$  and for  $\delta^{18}\text{O}$  of  $\sim -2.9\text{‰} \pm 2.8$  (based on overall medians  
555 of  $\Delta\delta^2\text{H}/\delta^{18}\text{O}$ ). This is caused by too negative reconstructions for coniferous and  
556 grass sites. For coniferous sites, this can be explained with *n*-alkanes originating  
557 from understory grasses, and for the grass sites the “signal damping” more effect  
558  $\delta^2\text{H}$  than  $\delta^{18}\text{O}$ . This leads to too negative reconstructed  $\delta^2\text{H}_{\text{leaf-water}}$  values and thus  
559 to too negative  $\delta^2\text{H}/\delta^{18}\text{O}_{\text{source-water}}$  reconstruction.



560 (vi) reconstructed (biomarker-based) RH<sub>MDV</sub> values tend to underestimate climate  
561 station-derived RH<sub>MDV</sub> values ( $\Delta$ RH<sub>MDV</sub> = ~ -17% ±12). For coniferous sites the  
562 underestimations are strongest, which can be explained with understory grasses  
563 being the main source of *n*-alkanes for the investigated soils under coniferous  
564 forests.

565 Overall, our study highlights the great potential of GDGTs and the coupled  $\delta^2\text{H}_{n\text{-alkane}}$ - $\delta^{18}\text{O}_{\text{sugar}}$   
566 approach for more quantitative paleoclimate reconstructions. Taking into account effects of  
567 different vegetation types improves correlations and reconstructions. This holds particularly  
568 true for the coupled  $\delta^2\text{H}_{n\text{-alkane}}$ - $\delta^{18}\text{O}_{\text{sugar}}$  approach, which is affected by “signal damping” of the  
569 grass vegetation. Assuming constant biosynthetic fractionation is likely a considerable source  
570 of uncertainty. Climate chamber experiments would be very useful to further evaluate and refine  
571 the coupled  $\delta^2\text{H}_{n\text{-alkane}}$ - $\delta^{18}\text{O}_{\text{sugar}}$  approach, because uncertainties related to microclimate  
572 variability can be reduced. Field experiments like ours suffer from the fact that biomarker pools  
573 in the sampled topsoils may have been affected by past vegetation and climate changes.

574

### 575 **Acknowledgements**

576 We thank L. Wüthrich, H. Veit, T. Sprafke, A. Groos (all University of Bern), A. Kühnel  
577 (Technical University of Munich) for constructive discussions and statistical advices, and M.  
578 Schaaerschmidt (University of Bayreuth), C. Heinrich and M. Benesch (Martin-Luther-  
579 University Halle-Wittenberg) for laboratory assistance during  $\delta^{18}\text{O}_{\text{sugar}}$  analysis and pH  
580 measurements, respectively. The Swiss National Science Foundation (PP00P2 150590) funded  
581 this research. J. Hepp greatly acknowledges the support by the German Federal Environmental  
582 Foundation (DBU) in form of his PhD-fellowship.

583

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