Leaf wax *n*-alkane patterns and compound-specific δ^{13} C of plants and topsoils from semi-arid/arid Mongolia

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Abstract. Leaf wax *n*-alkane patterns and their compound-specific δ^{13} C signatures are valuable proxies for paleoenvironmental reconstructions. So far, their potential has not been investigated in semi-arid to arid Mongolia. We have therefore analysed the leaf wax *n*-alkanes and their compound-specific δ^{13} C signature of five plant species (*Poaceae, Cyperaceae, Artemisia spp., Caragana spp.* and *Larix sp.*) and topsoils (0 – 5 cm) along two transects in central and southern Mongolia.

- Grasses show a distinct dominance of the n-C₃₁ homologue, whereas the shrubs Caragana spp. and Artemisia spp. are dominated by n-C₂₉. Larix sp. is characterized by the mid-chain n-alkanes n-C₂₃ and n-C₂₅. From plant to topsoil, n-alkane patterns show the potential to differentiate between grass covered sites from those covered by Caragana spp. n-Alkane concentrations and odd-over-even predominance (OEP) of the topsoils are distinctly influenced by mean annual temperature, mean annual precipitation and aridity, likely reflecting the degree of n-alkane degradation and biomass production. In contrast, the average
- 10 chain-length (ACL) and the *n*-alkane ratio $(n-C_{31}/n-C_{29} + n-C_{31})$ are not affected by climatic parameters and thus, not biased by climate. The compound-specific δ^{13} C signatures are strongly correlated to climate, showing a significant enrichment with increasing aridity, indicating the effect of water use efficiency. Our calibration results suggest that long-chain *n*-alkanes and their compound-specific δ^{13} C signatures have great potential to reconstruct paleoenvironmental and -climatic conditions when used in sediment archives from Mongolia.

15 1 Introduction

Leaf wax biomarkers such as long-chain *n*-alkanes (n-C₂₅ - n-C₃₅) are produced in the plant cuticle as a protection layer against environmental stress and are synthesized by the polyketide biosynthetic pathway resulting in a distinct odd-over-even predominance (OEP) (Eglinton and Hamilton, 1967; Shepherd and Wynne Griffiths, 2006). Due to their water insolubility, chemical inertness and relative resistance against biochemical degradation, leaf wax *n*-alkanes stay well preserved in sediment

20 archives over geological timescales and serve as valuable biomarkers for former environmental conditions (Eglinton and Eglinton, 2008). During the last decades, leaf wax *n*-alkanes have increasingly been used for paleoenvironmental reconstructions in

lake sediments (Aichner et al., 2017; Rach et al., 2017; Schwark et al., 2002; Sun et al., 2016), marine sediments (Castañeda et al., 2009; Rommerskirchen et al., 2006; Schefuss et al., 2005), loess-paleosol sequences (Häggi et al., 2019; Schäfer et al., 2018; Zech et al., 2013) and fluvial sediment-paleosol sequences (Bliedtner et al., 2018b).

- The relative homologue distribution of leaf wax *n*-alkanes has been used as a chemotaxonomic marker to differentiate among vegetation forms and thus reconstruct paleovegetation: the *n*-alkanes n-C₂₇ and n-C₂₉ are thought to be mainly produced by deciduous trees/shrubs, whereas n-C₃₁ and n-C₃₃ are mainly produced by grasses/herbs (Bliedtner et al., 2018a; Schäfer et al., 2016; Vogts et al., 2009). The compound-specific δ^{13} C signature of leaf wax *n*-alkanes have also been used for reconstructing changes in the vegetation composition of C₃ (-20 to -35‰) and C₄ (-10 to -14‰) plants (Castañeda et al., 2009; Lane,
- 30 2017; Rao et al., 2016; Rommerskirchen et al., 2006), and give additional paleoclimatic information about drought stress and arid conditions for C₃ plants (Aichner et al., 2010a, b; Schäfer et al., 2018). Several studies have shown a strong correlation between the δ^{13} C leaf wax signal of C₃ plants and water use efficiency (WUE) that is influenced by precipitation, temperature and evapotranspiration and describes the stomata conductance of a plant to avoid water loss (Diefendorf and Freimuth, 2017; Farquhar et al., 1982; Rao et al., 2017). Along a 400 mm isohyet in China, Wang et al. (2018b) have been shown that the
- 35 compound-specific δ^{13} C signature of leaf wax *n*-alkanes are strongly correlated to temperature, with strongest correlations observed for the average temperature of June, July and August (Wang et al., 2018b). Thus, warmer/dryer conditions cause an increase of WUE resulting in an ¹³C enrichment and cooler/wetter conditions *vice versa* (Aichner et al., 2015; Castañeda et al., 2009; Diefendorf and Freimuth, 2017; Wang et al., 2018b).
- Although leaf wax *n*-alkane patterns and compound-specific δ^{13} C have been increasingly used in sediment archives for paleoenvironmental reconstructions within the last decades, they need to be calibrated regionally on recent reference material before any paleoenvironmental reconstruction can be made. The need for regional calibrations has been emphasized by the fact that Bush and McInerney (2013) questioned whether leaf wax *n*-alkane patterns can discriminate between modern vegetation forms on a global scale while several regional studies found them discriminating on a regional scale. Although the most abundant homologues differ from region to region, a good discrimination power has been reported from Europe (Schäfer et al.,
- 2016; Zech et al., 2009, 2010), the Caucasus region (Bliedtner et al., 2018a), North- and South America (Diefendorf et al., 2015; Feakins et al., 2016; Lane, 2017) and the African rain forest and savanna (Vogts et al., 2009).

However, when interpreting leaf wax *n*-alkanes and their compound-specific δ^{13} C signature several potential pitfalls such as species-specific and intra-leaf variations (Diefendorf et al., 2011; Gao et al., 2015), the influence of environmental and climatic factors (Carr et al., 2014; Diefendorf and Freimuth, 2017; Farquhar et al., 1982; Hoffmann et al., 2013; Rao et al., 2017;

50 Tipple et al., 2013), the dependency on altitude (Feakins et al., 2018; Hultine and Marshall, 2000), and *n*-alkane degradation (Brittingham et al., 2017; Buggle et al., 2010; Li et al., 2018a) have to be considered and accounted for. So far, such regional calibration studies on recent leaf wax *n*-alkane patterns, compound-specific δ^{13} C as well as potential climatic biases on the leaf wax signal do not exist for Mongolia.

Thus, this study investigates leaf wax *n*-alkane patterns and compound-specific δ^{13} C of modern plants and topsoils from 55 semi-arid/arid Mongolia to evaluate their potential for regional paleovegetation and -climate reconstructions. More specifically, we tested the chemotaxonomic potential of leaf wax *n*-alkane patterns from five dominant plant species and whether their

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homologue distribution can be used to discriminate between woody shrubs and grasses/herbs on a regional scale. Moreover, we investigate differences in the compound-specific δ^{13} C signature of leaf waxes from plants and topsoils, and how the plant signal is incorporated into the topsoil. Since only the topsoils represent an averaged decadal leaf wax signal, we check for

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potential environmental/climatic influences by correlating the leaf wax *n*-alkane patterns and δ^{13} C with altitude, mean annual temperature (MAT), mean annual precipitation (MAP) and the aridity index (AI). While we test that the *n*-alkane patterns are not biased by climatic influences, we test the potential of leaf wax δ^{13} C to reflect on MAT, MAP and aridity. Additionally, we test for a potential bias from altitude on leaf wax *n*-alkane patterns and the δ^{13} C signal. Therefore, our calibration results will be a base for future robust paleovegetational and -climate reconstructions in semi-arid/arid Mongolia using leaf wax *n*-alkanes

65 from regional sediment archives. Such lacustrine, loess-paleosol and colluvial sediment archives have been increasingly investigated in semi-arid/arid Mongolia and could potentially be used for future leaf wax based paleoenvironmental reconstructions (Klinge et al., 2017; Peck et al., 2002; Prokopenko et al., 2007; Rudaya and Li, 2013; Wang et al., 2011).

2 Material and methods

2.1 Geographical setting and sampling

- 70 Semi-arid/arid Mongolia is a highly continental region characterized by harsh/long winters and hot/short summers (Dashkhuu et al., 2015). Mongolia is located at the interface of three major atmospheric circulation systems controlling the regional climate (Fig. 1). The summer climate is dominated by the East Asian Summer Monsoon (EASM) and the Westerlies that provide most of the yearly moisture and precipitation during the summer month, i.e. 75% of the annual precipitation falls in June, July and August (Rao et al., 2015; Wang and Feng, 2013). The dry and cold winter climate is dominated by the Siberian high that mostly
- 75 blocks the moisture supply from the Westerlies during winter (Peck et al., 2002). The Mongolian climate has a north-south gradient in MAT and MAP, with increasing MAT from north to south and MAP vice versa (Harris et al., 2014, Fig. 2A, B). This north-south gradient in temperature and precipitation is further reflected by the AI (Fig. 2C) and in the distribution of regional vegetation biomes with taiga and mountain/forest steppe in northern and central Mongolia and steppe/desert steppe in southern Mongolia (Hilbig, 1995; Klinge and Sauer, 2019).
- 80 Topsoils (0-5 cm) were sampled along a north-south transect (transect I, see Fig. 1 for location) in June 2016, and plants and topsoils (0 - 5 cm) along an east-west transect (transect II, see Fig. 1 for location) in July/August 2017. Along transect II, the topsoils were sampled together with the dominant plant species ($5m^2$ around the topsoil sampling site), which comprise the woody shrub Caragana spp., the grasses/herbs Poaceae, Cyperaceae and Artemisia spp., and Larix sp. as coniferous tree. Artemisia spp. summarizes different herbaceous species and perennial 'shrubby' species with a woody base (e.g. Artemisia
- *frigida*). Transect I covers an altitudinal range between 1224 and 1611 m a.s.l., MAT and MAP range from -1.7 to 5.5° C and 85 99 to 276 mm/a. Transect II covers an altitudinal range between 1333 and 2792 m a.s.l., and a range from -7.3 to $-0.5 \,^{\circ}\text{C}$ and 210.8 to 276.2 mm/a in MAT and MAP, respectively and is separated in the Ugii Nuur catchment (TS II-A), an altitudinal transect (TS II-B) and the Telmen Nuur catchment (TS II-C) (Fig. 2; Fick and Hijmans, 2017).

2.2 Leaf wax analysis

90 2.2.1 Leaf wax extraction and quantification

Total lipids of the topsoils ($\sim 35 \text{ g}$) from transect I were extracted at the University of Bern, Switzerland, using accelerated solvent extraction (Dionex ASE 200: 6.9 MPa, 100 °C) with 60 ml dichloromethane (DCM):methanol (MeOH) (9:1, v/v) over three extraction cycles as described by Bliedtner et al. (2018b). Total lipids of topsoils ($\sim 10 \text{ g}$) and plant leaves ($\sim 1 \text{ g}$) from transect II were extracted at the Friedrich-Schiller-University of Jena, Germany, using ultrasonic extraction with 20 ml DCM:MeOH (9:1, v/v) over three cycles as described by Bliedtner et al. (2018a).

95 DCM:MeOH (9:1, v/v) over three cycles as described by Bliedtner et al. (2018a).

The total lipid extract from both transects was separated over aminopropyl pipette columns (Supelco, 45 μm) into (i) an apolar fraction including the *n*-alkanes, (ii) a more polar fraction and (iii) an acid fraction. *n*-Alkanes were eluted with ~ 4ml hexane and additionally cleaned over coupled silver-nitrate (AgNO₃) – zeolite pipette columns. *n*-Alkanes were subsequently dissolved in hydrofluoric acid and liquid–liquid recovered with hexane. Identification and quantification of the *n*-alkanes was
performed on an Agilent 7890B gas chromatograph equipped with an Agilent HP5MS column (30 m × 320 μm × 0.25 μm film thickness) and a flame ionization detector (GC-FID). For identification and quantification, external *n*-alkane standards with a known concentration (*n*-alkane mix *n*-C₂₁ – *n*-C₄₀, Supelco) were run with each sequence.

2.2.2 Compound-specific δ^{13} C analysis

Compound-specific carbon isotopes were measured for the most abundant *n*-alkanes n-C₂₉ and n-C₃₁. Isotope measurements were performed on an Isoprime Vision isotope ratio mass spectrometer coupled to a gas chromatograph (Agilent 7890B GC) equipped with an Agilent HP5GC column ($30 \text{ m} \times 320 \text{ \mu} \text{m} \times 0.25 \text{ \mu} \text{m}$ film thickness) via a GC5 pyrolysis/combustion interface. The GC5 was operating in combustion mode with a CuO reactor at 850 °C. Samples were injected in splitless mode and measured in triplicates. *n*-Alkane standards (n-C₂₇, n-C₂₉ and n-C₃₃) with known isotopic composition (Schimmelmann standard, Indiana) were measured as duplicates after every third triplicate. The standard deviation for the triplicate measurements were

110 < 0.7%, and the standard deviation for the standards was < 0.2% (n = 102). Carbon isotopic composition is given in the delta notation (δ^{13} C) versus the Vienna Pee Dee Belemnite standard (VPDB).

2.2.3 Data analysis

n-Alkane concentrations ($\sum n$ -Alkane) are given in $\mu g g^{-1}$ dry weight and were calculated as the sum of *n*-C₂₅ to *n*-C₃₅. Additionally, we tested the *n*-alkane concentrations as the sum of *n*-C₂₃ to *n*-C₃₅ because of the high proportions of *n*-C₂₃ in

115 *Larix sp.* However, differences between both *n*-alkane concentrations are minor and thus, we present the *n*-alkane concentrations including n-C₂₃ in the supplementary material. The OEP was calculated according to Hoefs et al. (2002) and serves as a

proxy for degradation with values below five indicating enhanced *n*-alkane degradation (Buggle et al., 2010; Zech et al., 2009, 2010).

$$OEP = \frac{n - C_{27} + n - C_{29} + n - C_{31} + n - C_{33}}{n - C_{26} + n - C_{28} + n - C_{30} + n - C_{32}}$$
(1)

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The average chain length (ACL) was determined after Poynter et al. (1989) and is used to distinguish between leaf waxes predominantly produced by deciduous trees and shrubs (n-C₂₇ and n-C₂₉) and grasses and herbs (n-C₃₁ and n-C₃₃) (Bliedtner et al., 2018a).

$$ACL = \frac{27 \cdot n \cdot C_{27} + 29 \cdot n \cdot C_{29} + 31 \cdot n \cdot C_{31} + 33 \cdot n \cdot C_{33}}{n \cdot C_{27} + n \cdot C_{29} + n \cdot C_{31} + n \cdot C_{33}}$$
(2)

A normalized *n*-alkane ratio was calculated for the most abundant *n*-alkanes n-C₂₉ and n-C₃₁.

125 *n*-Alkane ratio =
$$\frac{n \cdot C_{31}}{(n \cdot C_{29} + n \cdot C_{31})}$$
 (3)

2.3 Statistical analysis

Differences in *n*-alkane concentration, OEP, ACL, *n*-alkane ratio and compound-specific δ^{13} C among plant species and between topsoils and plants were analysed using analysis of variance (ANOVA) followed by pairwise comparisons based on Tukey's honestly significant difference. Since the relative homologue distribution of the *n*-alkanes is important for the discrimination

- 130 between plant species, the *n*-alkane patterns were analysed as compositional data according to Aitchison (2003). Correlations of *n*-alkane concentration, OEP, ACL, *n*-alkane ratio and compound-specific δ^{13} C with climatic parameters and altitude were tested using weighted linear or polynomial regression. The environmental parameters MAT and MAP were derived from the WorldClim 2.0 dataset (1970-2000, 30s resolution, Fick and Hijmans, 2017) and the AI from the Global Aridity Index and Potential Evapo-Transpiration (ET0) Climate Database v2 (1970-2000, 30s resolution, Trabucco and Zomer, 2019). The
- 135 altitude was extracted from the Shuttle Radar Topography Mission (SRTM) data (Jarvis et al., 2008). Regressions were tested only for the topsoils because they represent an averaged leaf wax signal over some decades (Angst et al., 2016) and can thus be correlated with the WorldClim and the Global Aridity Index data that represents averaged climate data over 30 years. In contrast, plants only reflect an annual signal of the sampling year 2017 and cannot be correlated with the available climatic parameters from the WorldClim and the Global Aridity Index dataset, and also annual climate data for the sampling year 2017
- are not available. Model selection was based on hierarchical comparison of models with increasing polynomial order using F ratios. Goodness of fit of the final models was assessed using weighted R^2 values. All statistical analyses were done using the statistical software system R (R Core Team, 2019), and the package compositions (Boogaart, 2013) for compositional data analysis.

3 Results

145 **3.1** *n*-Alkane patterns in plants and topsoils

Leaf wax *n*-alkanes are present in all analysed plants and topsoils and show a distinct OEP (Fig. 3). For the analysed plants from transect II, the most abundant n-alkane homologues vary among the plant species: Poaceae and Cyperaceae tend to be dominated by $n-C_{31}$, and Artemisia spp. and Caragana spp. by $n-C_{29}$. Larix sp. has its dominance on the mid-chain nalkane $n-C_{25}$. The topsoils from both transects are mostly dominated by $n-C_{31}$ (Fig. 3). Figure 4 illustrates differences in 150 *n*-alkane concentrations, OEP, ACL and the *n*-alkane ratio of the analysed plant species and topsoils, and table 1 shows the corresponding level of significance. *n*-Alkane concentrations are significantly higher in plants than topsoils ($p = 5.1e^{-07}$). Plants range from 9 to $2508 \,\mu g g^{-1}$, with *Caragana spp.* having significantly the highest concentrations, and *Larix sp.* having the lowest concentrations (Fig. 4A, Tab. 1). *n*-Alkane concentrations in topsoils range from 0.2 to $59 \,\mu g \, g^{-1}$ with transect I having much lower concentrations than transect II (Fig. 4A). Plants show a wide OEP with values ranging between 4 and 155 39. Highest OEP values are observed for *Caragana spp.*, followed by *Poaceae*, *Cyperaceae*, *Artemisia spp.* and *Larix sp.* Topsoils have generally lower OEP values than plants ranging from 1.5 up to 5.5 for transect I and from 4.8 to 19 for transect II, respectively (Fig. 4B). ACL values of plants range from 28.3 to 30.8 with Larix sp. showing significantly lower ACLs compared to the other plants (Tab. 1). ACL values for topsoils are higher but in the same range as most plant species, ranging from 29.6 to 30.4 along transect I and between 29.2 and 31.8 along transect II (Fig. 4C). The *n*-alkane ratio show values 160 between 0.16 and 0.8. *Poaceae* and *Cyperaceae* tend to have the highest values, *Caragana spp.* and *Larix sp.* the lowest. The

topsoils range from 0.42 to 0.75 for both transects (Fig. 4D).

3.2 Compound-specific δ^{13} C

Compound-specific δ¹³C were measured for the most abundant *n*-alkanes *n*-C₂₉ and *n*-C₃₁ (Fig. 5). Plants show δ¹³C values between -36% to -29.5% for *n*-C₂₉, and between -35.8% and -30.3% for *n*-C₃₁, with *Larix sp.* being most enriched in 165 leaf wax ¹³C among all plant species. In comparison to plants, topsoils tend to be more enriched in ¹³C with a larger scatter and range from -33.8% to -25.6% for *n*-C₂₉, and from -34.3% to -25.2% for *n*-C₃₁ with a distinct enrichment along transect I (Fig. 5). Compound-specific δ¹³C of *n*-C₂₉ and *n*-C₃₁ differ significantly between topsoils from transect I and transect II, but not significantly between plant species and between plant species and topsoils from transect II (Tab. 1).

4 Discussion

170 4.1 *n*-Alkane patterns in plants

The plant species from Mongolia show distinct differences in their relative *n*-alkane homologue pattern (Fig. 3). The grasses *Poaceae* and *Cyperaceae* are dominated by n-C₃₁, whereas the woody shrub *Caragana spp.* is dominated by n-C₂₉. Those findings are in line with previous regional studies from the Tibetan Plateau (Cheung et al., 2015; Wang et al., 2018c), the

Chinese Loess Plateau (Liu et al., 2018) and the Caucasus region (Bliedtner et al., 2018a), which report *n*-C₃₁ being mainly
produced by grasses/herbs and *n*-C₂₉ by deciduous trees/shrubs. Although *Artemisia spp*. could be expected to be herbaceous with a dominance in *n*-C₃₁, our results show a distribution maximum at *n*-C₂₉, i.e. it is more similar to the woody shrub *Caragana spp*. than to the grass species. This does not necessarily be contradictory and corroborates the findings of Wang et al. (2018b) along a 400 mm isohyet transect from north-western to central China who reports that *Artemisia* can grow both as herbaceous and as a woody shrub (e.g. *Artemisia frigida*). In contrast to the other plant species, the coniferous tree *Larix*

- sp. is dominated by the mid-chain *n*-alkanes n-C₂₃ and n-C₂₅, resulting in significant lower ACL (Fig. 4C, Tab. 1). However, statistically significant differences between the ACL of the other plant species are not evident, although the relative homologue pattern reveal differences among them (Tab. 1, Fig. 3 and 4C). The ACL of the grasses/herbs *Poaceae* and *Cyperaceae* and *Artemisia spp.* and the woody shrub *Caragana spp.* have only slight differences and a small range between 29.6 and 29.9. This is due to a strong ACL scattering of the grasses/herbs that overlap the ACL of the woody shrubs (Fig. 4C). Thus, a clear
- 185 chemotaxonomic discrimination between grasses/herbs and woody shrubs is not given by the ACL for the investigated modern plants from Mongolia. A better chemotaxonomic discrimination is provided by the *n*-alkane ratio $n-C_{31}/(n-C_{29} + n-C_{31})$ that is based on the most abundant homologues $n-C_{29}$ and $n-C_{31}$. The *n*-alkane ratio significantly separates the grasses *Poaceae* and *Cyperaceae* from the woody shrub *Caragana spp*. and the coniferous tree *Larix sp*. (Fig. 4D). The *n*-alkane ratio of *Artemisia spp*. lies in-between the woody shrubs and grasses, most likely because of their ability to grow as both herbaceous and woody
- 190 shrubs (Fig. 4D). This is further expressed statistically as the *n*-alkane ratio of *Artemisia spp.* is equal to those calculated from both *Caragana spp.* and the grass species (Tab. 1). Beside *Artemisia spp.*, the *n*-alkane ratio has the chemotaxonomic potential to discriminate significantly between *Larix sp.* and grasses as well as between *Caragana spp.* and grasses. However, the *n*-alkane patterns from *Larix sp.* with their mid-chain dominance has to be interpreted with caution when comparing species-specific differences to the long-chain dominated plant species, that is mostly due to the fact that *Larix sp.* only produce
- small amounts of *n*-alkanes. When incorporated into the soil, the coniferous *n*-alkane signal from *Larix sp.* should become overproportional overprinted by the undergrowth of the grasses/herbs (Diefendorf et al., 2011; Schäfer et al., 2016).

4.2 Compound-specific δ^{13} C of plants

The compound-specific δ^{13} C values of the *n*-alkanes from our investigated plants from transect II show consistent δ^{13} C values among the plant species, except *Larix sp.*, and are in a typical range of C₃ plants (Fig. 5; Tipple and Pagani, 2007). Although

- 200 Pyankov et al. (2000) have reported C_4 plants in Mongolia among 16 plant families including *Poaceae*, those are not evident along our sampled plant transect. While some C_4 plants have been found in the Khangai Mountains, their distribution is mainly limited to the semiarid-steppe and semi-desert areas in southern Mongolia and the Gobi desert, i.e. beyond our plant sampling sites (Pyankov et al., 2000; Su et al., 2011). Statistically significant differences did not exist between the most abundant homologues n- C_{29} and n- C_{31} (p = 1), indicating that no different fractionation has occured during biosynthesis (Wang et al.,
- 205 2018b). Consistent δ^{13} C values between the most abundant homologues are in good agreement with compound-specific δ^{13} C analyses of three *Artemisia* species (*Artemisia argyi, capilares* and *scoparia*) along a 400 mm isohyet in China (Wang et al., 2018b). While no differences are found between the δ^{13} C values of the grasses/herbs and woody shrubs, only *Larix sp.* is

enriched up to 2%, but still in the range C₃ plants (Fig. 5, Tab. 1). Such an enrichment of coniferous trees compared to other plants might be explained by differences in species-specific fractionation (Diefendorf et al., 2015).

210 4.3 Comparing *n*-alkane patterns and compound-specific δ^{13} C of plants versus topsoils

4.3.1 The leaf wax signal from plants to topsoils along transect II

Along transect II, modern plants have higher *n*-alkane concentration than the topsoils, with *Artemisia spp.* and *Caragana spp.* having significantly higher *n*-alkane concentrations than the respective topsoils (Fig. 4A, Tab. 1). Thus, lower *n*-alkane concentration in the topsoils indicate that *n*-alkanes become diluted during the incorporation from plant biomass into the topsoil (Fig. 4A). Likewise, the OEP decreases from plants to topsoil and indicate enhanced organic matter degradation (Buggle et al.,

- 215 (Fig. 4A). Likewise, the OEP decreases from plants to topsoil and indicate enhanced organic matter degradation (Buggle et al., 2010; Schäfer et al., 2016) and microbial alteration (Schulz et al., 2012) (Fig. 4B). Despite possible degradation effects during soil developement, the topsoils show distinct OEP values between 4.8 and 19, still indicating a good preservation (Zech et al., 2009). Along transect II, one exception in terms of higher *n*-alkane concentration and OEP is TSC10 Ah1, showing 59µg g⁻¹ and 19, respectively (Fig. 1, sampling site 25). Site TSC10 is characterized by stagnating soil conditions with a distinct organic
- 220 rich topsoil, limiting organic matter degradation and microbial alteration of *n*-alkanes (Hoefs et al., 2002). Thus, TSC10 Ah1 remains exceptional and not comparable to the other topsoils from transects II. Overall, decreasing concentrations and OEP values from plants to topsoils are in good agreement with other regional studies (Bliedtner et al., 2018a; Howard et al., 2018; Li et al., 2018b; Schäfer et al., 2016; Zech et al., 2009).

For the topsoils, *n*-C₃₁ is on average the most abundant *n*-alkane homologue, indicating a typical *n*-alkane pattern produced

- by grasses (Bliedtner et al., 2018a, Fig. 3). The only exceptions are sites covered with *Caragana spp.* (n = 8) where higher amounts of *n*-C₂₉ are evident within the respective topsoils, and the two *Caragana* covered topsoils TLC4 Ah1 and TLC6 Ah1 (Fig. 1, sampling sites 40 and 42) even show a dominance of *n*-C₂₉ (Suppl. Mat.). Thus, the dominant *n*-C₂₉ signal produced by the woody shrubs is also reflected in the respective topsoils. This is further expressed by lower ACLs and *n*-alkane ratios for those topsoils, which explains the scattering towards *n*-C₂₉ in ACL and 0.4 for the *n*-alkane ratio, respectively (Fig. 4C, D).
- At sites covered with *Larix sp.*, the mid-chain length dominance of *Larix sp.* is not reflected in the respective topsoils, which are mainly dominated by *n*-C₃₁ *n*-alkanes. Thus, *n*-alkanes from *Larix sp.* must become strongly diluted from plant to topsoil and the topsoils reflect mostly the *n*-alkanes from the grassy undergrowth like previously shown by Schäfer et al. (2016) for several coniferous sites including *Larix, Picea, Abies and Pinus*. Compared to the plants, the isotopic signature of the topsoils is slightly more enriched but reveal no statistical significance (Fig. 5, Tab. 1), which is in line with previous studies and might reflect an enrichment by diagenesis from litter to topsoil or a change in vegetation composition (Wu et al., 2019, and references).
- therein).

Environmental information of the plants compound-specific δ^{13} C signal only reflects one vegetation period, whereas the topsoils compound-specific δ^{13} C signal reflects environmental variability on decadal timescales, which might explain the topsoil δ^{13} C enrichment. However, one topsoil (TLC4 Ah1) shows a strong ¹³C enrichment up to ~ -25%. Such an enrichment

might be explained by *n*-alkane contributions from succulent plants, which tend to be more enriched in 13 C within the range of

C₃ plants (Boom et al., 2014). Succulents were growing on stone rich, thin topsoils in the catchment of Lake Telmen (Fig. 1, TS II-C). For comparison, we sampled the succulent *Orostachys malacophylla* from the Telmen catchment and analysed their compound-specific δ^{13} C isotopes that yield -24.7% for *n*-C₂₉ (*n* = 1) and -25.03% for *n*-C₃₁ (*n* = 1). Thus, increased inputs of succulent δ^{13} C might best explain the enriched isotopes in the Telmen catchment and the extreme value of $\sim -25\%$ from site TLC4.

4.3.2 The leaf wax signal of the topsoils along both transects

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The topsoils of both Mongolian transects show distinct differences in *n*-alkane concentration and OEP, which are higher along transect II and decrease along transect I. This is mostly due to the fact that *n*-alkane production and degradation is influenced by the climatic gradient along transect I (see section 4.4 and Fig. 2 for more detailed discussion). Beside some n-C₂₉ dominated sites with *Caragana spp.*, the ACL and the *n*-alkane ratio show the dominance of n-C₃₁ which indicate the *n*-alkane origin from the grasses *Poaceae* and *Cyperacea* (Fig. 4, Bliedtner et al., 2018a; Schäfer et al., 2016; Vogts et al., 2009; Zech et al., 2010). This is further expressed by the results of ANOVA, because the ACL and *n*-alkane ratios from topsoils and grasses are not statistically different (Tab. 1).

4.4 Climatic influences on topsoil *n*-alkane patterns and compound-specific δ^{13} C

- To test potential climatic influences on our *n*-alkane proxies, we correlate them with MAT, MAP and AI (Fig. 6). The *n*-alkane patterns show that *n*-alkane concentrations in topsoils and their preservation (OEP) are correlated to climatic parameters, but also with altitude, indicating higher *n*-alkane concentrations and a better *n*-alkane preservation above 2000 m a.s.l. (Fig. 7, transect II-B). Since altitude generally controls MAT (R² = 0.624) and MAP (R² = 0.395), we suggest that variations along the investigated transects are primarily climate induced. Thus, we detected MAT (R² = 0.517, p =< 1e 04) as main climatic control parameter on *n*-alkane concentrations for the topsoils from both transects. Correlations of *n*-alkane concentration with MAP (R² = 0.448, p =< 1e 04) and AI (R² = 0.258, p = 0.0031) are likewise significant indicating a clear correlation (Fig. 6). We conclude that lower *n*-alkane concentrations probably indicate reduced biomass production and enhanced *n*-alkane degradation in the topsoils. The former even intensifies when combined with livestock grazing (Kölbl et al., 2011, and references therein), especially along transect I (sampling sites 1 17) where biomass production is reduced and overgrazing occur
 extensively (Fig. 6). In contrast, sites along transect II-B (above 2000 m a.s.l.) describe a favourable area for plant growth and
- thus an increase in *n*-alkane concentrations and OEP (Fig. 7). However, especially transect II-B and II-C are characterized by a distinct scatter, which is mainly the result of variations in plant physiology and site-specific/micro-climatic characteristics which are not covered by the reanalysis data.

Previous studies have shown correlations between the production of the most abundant homologues with climatic parameters and altitude, i.e. common vegetation proxies such as the ACL and *n*-alkane ratio could reflect changes in MAT since plants tend to produce longer *n*-alkanes as a protection against water loss (Bush and McInerney, 2013; Feakins et al., 2016; Sachse et al., 2006; Tipple et al., 2013; Wang et al., 2018b, a). However, this could not be observed by the *n*-alkanes from Mongolian topsoils, since the ACL and the *n*-alkane ratio as common vegetation proxies show no correlations with MAT, MAP, AI or altitude (Fig. 6, 7).

- 275 In contrast, compound-specific δ^{13} C of the topsoils correlate significantly with climatic parameters. Our results show an enrichment in ¹³C with increasing temperature, aridity and decreasing precipitation (Fig. 6). This climate induced enrichment in ¹³C follows mainly the north-south gradient in decreasing MAP and increasing MAT along transect I from central Mongolia into the Gobi Desert. The only exceptions are the extreme values near Lake Telmen, which are mostly due to the input of the ¹³C enriched succulent *Orostachys malacophylla*. As already proposed by Diefendorf et al. (2010), MAP is an intense predictor
- on δ¹³C (n-C₂₉: R² = 0.683, p =< 1e 04, n-C₃₁: R² = 0.343, p =< 1e 04), which is further expressed in distinct linear correlations with the AI (Fig. 6). However, δ¹³C correlates also with altitude, i.e. leaf wax ¹³C shows a significant depletion with increasing altitude. Previous studies have shown a strong positive correlations between δ¹³C and altitude, indicating enhanced ¹³C enrichment due to an altitude effect (Feakins et al., 2018) and/or environmental/climatic plant physiological adaptions like a decrease in stomatal conductance with increasing altitude (Hultine and Marshall, 2000, and references therein). However, 285 our data shows the opposite behaviour, indicating that the observed ¹³C enrichment is based on stomata conductance induced
- by climate (water stress) and not by altitude (Fig. 6, 7). Thus, the 13 C enrichment basically indicates the photorespiration of C₃ plants, affected by water availability and evapo-transpiration and thus, the WUE (Tipple and Pagani, 2007; Diefendorf and Freimuth, 2017).

5 Conclusions

- 290 This study investigates leaf wax *n*-alkane patterns and compound-specific δ^{13} C of modern plants and topsoils from semiarid/arid Mongolia to test their chemotaxonomic potential and dependency on climate. Our results provide the first regional calibration of leaf wax *n*-alkanes for semi-arid/arid Mongolia with the following results:
- i. *Caragana spp., Artemisia spp.* and grasses (*Poaceae* and *Cyperacea*) from semi-arid/arid Mongolia show distinct differences in their relative *n*-alkane patterns. *n*-Alkanes from the grasses are clearly dominated by *n*-C₃₁, whereas the woody shrub *Caragana spp.* is dominated by *n*-C₂₉. Since *Artemisia* species can grow both as herbaceous and woody shrubs, *Artemisia spp.* shows not a typical *n*-C₃₁ dominance but is rather more equal to *Caragana spp.* with a dominance in *n*-C₂₉. *Larix sp.* is dominated by the mid-chain *n*-alkanes *n*-C₂₃ and *n*-C₂₅. However, *Larix sp.* produces only few amounts of *n*-alkanes and their dominance of mid-chain *n*-alkanes are not distinct in the respective topsoils. Thus, *n*-alkanes are not useful to reconstruct changes in the abundance of *Larix sp.* Although the ACL reveal no potential to discriminate between plant species, the most abundant *n*-alkanes *n*-C₂₉ and *n*-C₃₁ allow to discriminate between woody shrubs and grasses, which is expressed in the *n*-alkane ratio *n*-C₃₁/*n*-C₂₉ + *n*-C₃₁.
 - **ii.** From plants to topsoils of transect II, the decrease of *n*-alkane concentrations and OEP values indicate *n*-alkane dilution with mineral soil components and ongoing *n*-alkane degradation. The *n*-alkane pattern of the topsoils are mainly characterized by a dominance of n-C₃₁, indicating dominant input from grasses. *Caragana* covered sites tend to reflect the homologue

- 305 pattern of *Caragana spp.*, with *n*-C₂₉ being the most dominant *n*-alkane. Topsoils under *Larix sp.* are dominated by the input from the grassy undergrowth. There are no significant differences in compound-specific δ^{13} C between plant species and topsoils. Topsoils tend to be 2‰ enriched compared to the plants, indicating diagenesis from litter to topsoil.
- iii. *n*-Alkane concentrations and OEP values from Mongolian topsoils are significantly correlated to climatic parameters and decrease with increasing MAT and decreasing MAP. In contrast, our data indicate that the *n*-alkane patterns from the topsoils (ACL, *n*-alkane ratio) are not influenced by climatic parameters and thus, the *n*-alkane ratio can reliably be used to detect and reconstruct differences between the vegetation forms of grasses and woody shrubs. Although increasing altitude correlates with increasing *n*-alkane concentrations and OEP, altitude has no influences on ACL and *n*-alkane ratio. For compound-specific δ¹³C of the topsoils, strong correlations exist with increasing MAT and decreasing MAP, indicating an enhanced enrichment in ¹³C with increasing aridity and drought stress. Although, δ¹³C and altitude are also significantly negative correlated, the influence of altitude is negligible since MAT decrease and MAP increase with altitude. Thus, leaf wax δ¹³C is a valuable proxy for changes in climate and water use efficiency in semi-arid/arid Mongolia.

Our results show, that the *n*-alkane homologues $n-C_{29}$ and $n-C_{31}$ have the chemotaxonomic power to differentiate between grasses and the woody shrub *Caragana spp*. Future studies on plant *n*-alkane patterns should include a detailed identification

- 320 of plants regarding different species of each plant genus to reveal the full power of the *n*-alkane ratio as a vegetation proxy. This is particularly the case for different *Artemisia* species, which can so far not be separated from grasses and woody shrubs. While the *n*-alkane patterns are not biased by climatic influences, compound-specific δ^{13} C indicate a strong climatic dependency. Thus, *n*-alkanes and their compound-specific δ^{13} C can be potentially used as valuable proxies for future leaf wax based paleoenvironmental reconstruction in sediment archives from semi-arid/arid Mongolia.
- 325 Data availability. The dataset that is used in this study is available in the supplementary material.

Competing interests. JS, MB, PS, JSch, EB and RZ declare that they have no conflict of interest.

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Figure 1. Map of Mongolia (SRTM digital elevation model). The black/white circles mark the sampling sites along transect I and II. Black arrows indicate the influence of three major atmospheric circulation systems: the Westerlies, the East Asian Summer Monsoon and the Siberian High. Submaps show (TS II-A) the Ugii Nuur catchment, (TS II-B) an altitude transect near the Tsagaan Nuur catchment and (TS II-C) the Telmen Nuur catchment in more detail.



Figure 2. Climate and environmental conditions of Mongolia. Mean annual temperature (A), mean annual precipitation (B) and the aridity index (C). MAT and MAP are based on the WorldClim Dataset of Fick and Hijmans (2017), the AI is based on the Global Aridity Index and Evapo-Transpiration Climate Database v2 of Trabucco and Zomer (2019).



Figure 3. *n*-Alkane patterns of plants (transect II) and topsoils (transect I and II) from Mongolia. The bars show the mean values \pm standard deviations.



Figure 4. *n*-Alkane concentration $(n-C_{25} - n-C_{35})$ (A.), OEP (B.), ACL (C.) and *n*-alkane ratio $(n-C_{31}/(n-C_{29} + n-C_{31}))$ (D.) of plants and topsoils from Mongolia (n = *Larix sp.* = 7, *Cyperaceae* = 6, *Poaceae* = 10, *Caragana spp.* = 8, *Artemisia spp.* = 13, topsoils TS I = 17, topsoils TS II = 35). The boxplots indicate median values (red lines), mean values (black squares), interquartile ranges with lower (25%) and upper (75%) quartiles (box), outlayers (whiskers) and investigated samples (grey circles). See Tab. 1 for statistics.



Figure 5. Compound-specific δ^{13} C of plants and topsoils from Mongolia. (A.) Compound-specific δ^{13} C₂₉ (n: *Larix sp.* = 5, *Cyperaceae* = 5, *Poaceae* = 8, *Caragana spp.* = 7, *Artemisia spp.* = 13, topsoils TS I = 16, topsoils TS II = 34). (B.) δ^{13} C₃₁ (n: *Larix sp.* = 4, *Cyperaceae* = 5, *Poaceae* = 9, *Caragana spp.* = 6, *Artemisia sp.* = 13, topsoils TS I = 16, topsoils TS II = 34). The boxplots indicate median values (red lines), mean values (black squares), interquartile ranges with lower (25%) and upper (75%) quartiles (box), outlayers (whiskers) and investigated samples (grey circles). See Tab. 1 for statistics.



Figure 6. *n*-Alkane concentration (*n*-C₂₅ - *n*-C₃₅), OEP, ACL, *n*-alkane ratio (*n*-C₃₁/(*n*-C₂₉ + *n*-C₃₁)) and compound-specific δ^{13} C (*n*-C₂₉ and *n*-C₃₁) from Mongolian topsoils plotted against climatic parameters (MAP, MAT, AI; (Fick and Hijmans, 2017; Trabucco and Zomer, 2019)). Red trend lines illustrate linear regressions, black lines polynomial regressions. Bold values indicate significance ($\alpha = 0.05$).



Figure 7. *n*-Alkane concentration (*n*-C₂₅ - *n*-C₃₅), OEP, ACL, *n*-alkane ratio (*n*-C₃₁/(*n*-C₂₉ + *n*-C₃₁)) and compound-specific δ^{13} C (*n*-C₂₉ and *n*-C₃₁) from Mongolian topsoils plotted against altitude (m a.s.l.) (Jarvis et al., 2008). Red trend lines illustrate linear regressions, black lines polynomial regressions. Bold values indicate significance ($\alpha = 0.05$).

Table 1. ANOVA p-values indicating differences among plant species and between topsoils and plants for *n*-alkane pattern and compound-specific δ^{13} C (*n*-C₂₉ and *n*-C₃₁). Bold values indicate significance ($\alpha = 0.05$). TS = topsoils.

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|---------------------------------|--|-------|-------|------------------------|---------------------------------|---------------------------------|
| | $\sum n$ -Alkane (n -C ₂₅ - n -C ₃₅) | OEP | ACL | <i>n</i> -Alkane ratio | δ ¹⁵ C ₂₉ | δ ¹⁵ C ₃₁ |
| Poaceae - Cyperaceae | 1.000 | 0.979 | 1.000 | 0.999 | 1.000 | 0.999 |
| Larix sp Cyperaceae | 0.505 | 0.032 | 0.000 | 0.010 | 0.203 | 0.315 |
| Poaceae - Artemisia spp. | 0.999 | 0.392 | 0.986 | 0.069 | 0.998 | 1.000 |
| Poaceae - Caragana spp. | 0.000 | 0.000 | 0.861 | 0.000 | 0.814 | 0.995 |
| Poaceae - Larix sp. | 0.325 | 0.001 | 0.000 | 0.007 | 0.164 | 0.381 |
| Cyperaceae - Artemisia spp. | 0.999 | 0.948 | 0.959 | 0.079 | 0.995 | 0.991 |
| Cyperaceae - Caragana spp. | 0.000 | 0.000 | 0.802 | 0.000 | 0.812 | 0.965 |
| Caragana spp Artemisia spp. | 0.000 | 0.000 | 0.991 | 0.185 | 0.928 | 0.999 |
| Larix sp Artemisia spp. | 0.137 | 0.083 | 0.000 | 0.806 | 0.230 | 0.425 |
| Larix sp Caragana spp. | 0.000 | 0.000 | 0.004 | 0.946 | 1.826 | 0.734 |
| TS Transect II - Poaceae | 0.063 | 0.000 | 0.204 | 0.801 | 0.085 | 0.198 |
| TS Transect II - Cyperaceae | 0.259 | 0.075 | 0.658 | 0.99 | 0.162 | 0.219 |
| TS Transect II - Artemisia spp. | 0.006 | 0.251 | 0.012 | 0.000 | 0.132 | 0.181 |
| TS Transect II - Caragana spp. | 0.000 | 0.000 | 0.009 | 0.000 | 0.994 | 0.799 |
| TS Transect II - Larix sp. | 1.000 | 0.573 | 0.000 | 0.000 | 0.818 | 0.996 |
| TS Transect I - TS Transect II | 0.003 | 0.000 | 0.082 | 0.273 | 0.000 | 0.006 |
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