Responses to review by Editor Yakov Kuzyakov of *Biogeosciences* manuscript bg-2019-319: "*Increasing soil carbon stocks in eight permanent forest plots in China*"

We are very grateful to editor Dr. Yakov Kuzyakov for the detailed and valuable comments on our manuscript. We think our revisions are sufficient and thorough, addressing all the questions and issues during our revisions.

Editor's comments and our responses are presented below.

Editor's comments are given in *italic font*, and our responses in blue regular font.

Your submission is not prepared as correct submission:

- you have many yellow labelings - surely the first author has not improved the ms according to submission of somebody.

Response: We must apologize for the yellow labeling in the original manuscript. We are extremely sorry for the misunderstanding this has caused and for not clarifying it in the previous response letters. Actually, we had carefully merged the revisions of the coauthors, and the yellow labels were a way to mark the changes we made in the previous manuscript. For this updated manuscript, we have uploaded a "clean" version (without any track changes) to the system, and have also uploaded a "marked-up" version (showing the changes we made in the revised manuscript) after the "List of all relevant changes made in the manuscript " at the bottom of this file.

- the precision of the numbers in the text and in the Tables.

Response: Thank you for this comment. According to our understanding, values of soil organic carbon (SOC) content (%), bulk density (g cm⁻³), SOC stock (Mg C ha⁻¹), change rate of bulk density (mg cm⁻³ yr⁻¹), change rate of SOC stock (kg C ha⁻¹ yr⁻¹) and the relative change rate of SOC stock (% yr⁻¹) should be given to one decimal place throughout the revised text, in the in-text tables and also in the supplementary tables. However, considering the validity of the data, values of the change rate of SOC content (% yr⁻¹) are kept in two decimal place throughout the revised manuscript.

- the Abstract is very poor.

Response: We have clarified the methods and results in the section of *Abstract* in the revised manuscript. We added information about the four study sites and the dynamics of SOC content and bulk density during the past two decades.

- for the Fig 5 - there are much more papers from the literature.

Response: Thank you for this helpful comment. We have added SOC change rates from four public sources of literature (a boreal site, Rantakari et al., 2012; two temperate sites, Dölle and Schmidt, 2009, and Tefs and Gleixner, 2012; and a subtropical site, Tang and Li, 2013). We have also added information about sampling intervals in each site into Fig. 5 in the revised manuscript.

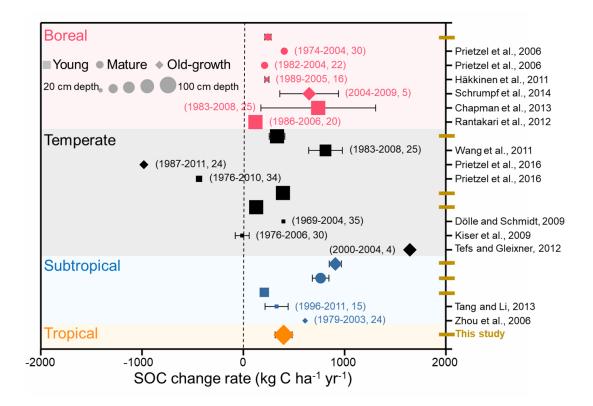


Fig. 5. Comparison of the changes in forest soil organic carbon (SOC) stocks according to repeated soil samplings and/or long-term observations. Different colors, shapes, and sizes represent different forest biomes, ages, and soil depths, respectively. The numbers in parentheses indicate the sampling times and intervals between the two soil samplings.

- some of the regressions in Fig 4 are not correct.

Response: Thank you for this helpful comment. We have deleted the function for the insignificant relationship, and retained the dashed lines to represent insignificant trends, in the new figure. We have clarified this in the figure legend in the revised manuscript.

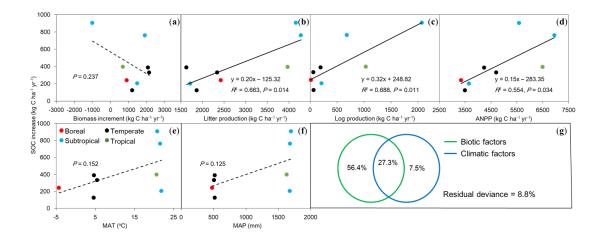


Figure 4. Relationships between rates of increase in soil organic carbon (SOC) against biotic and climatic factors in eight forests in China. (a) Biomass increment, (b) litter production, (c) log production, (d) above-ground net primary production (ANPP), (e) mean annual temperature (MAT), (f) mean annual precipitation (MAP), and (g) the relative effects of biotic (a, b and c) and climatic (e and f) factors on SOC increase rates (kg C ha⁻¹ yr⁻¹) using partial regression analyses. Solid lines indicate significant relationships (P < 0.05) and dashed lines represent insignificant trends (P > 0.05) between SOC increase rates and biotic and climatic factors.

- if your study is focused only on China, you should not publish this in an international Journal. Make your study internationally interesting.

Response: Thank you for this important comment. To increase the relevance of our manuscript, especially for international readers, we have made the following improvements in the revised manuscript. (1) We have compared our SOC change rates in boreal and temperate forests with SOC dynamics from national soil inventories of other countries, such as those from Germany, Sweden, and Denmark. (2) We have compared our regional SOC budget estimate with the results of Yang et al. (2014), who estimate SOC dynamics of China's forests by comparing measurements from literature during the 2000s with historical records derived from a national soil inventory during the 1980s. (3) We also compared the regional SOC budget estimate with a study of global forest carbon budgets (Pan et al. 2011). We found that carbon

sequestration in China's forests represented 20.8% of the total temperate regions of the northern hemisphere. The sequestration rate in China's forests is slightly higher than the mean value of the total temperate regions, relative to the forest area of China (i.e., 18.9% of the forest areas in the temperate regions). Then we added this information, and its implication for the importance of forest soil sequestration in this region, to the revised manuscript.

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List of all relevant changes made in the manuscript

We are very grateful to editor Dr. Yakov Kuzyakov for your valuable comments. We have carefully addressed all the comments and suggestions from you, which have been incorporated into the revision of our manuscript. The major changes are as follows:

1. We retained the comparison between our SOC change rates and the results from the repeated soil sampling in individual studies. These studies have primarily been conducted in the forests of Europe and the USA. In addition, we have compared the SOC change rates with SOC dynamics from national soil inventories of some European countries in the *Discussion* section. We also compared the regional SOC budget estimate with a regional assessment from China's forests (Yang et al., 2014) and a study of global forest carbon budgets (Pan et al., 2011).

2. We have rewritten the *Abstract* section, and clarified the methods and results in this section.

3. We have added some valid data into Fig. 5 and deleted the equations based on the insignificant relationship of the previous Fig. 4.

4. We have unified the precision of numbers in text and tables (Table 3, S2, S3 and S4).

5. We have carefully revised the language of the full text.

1 Increasing soil carbon stocks in eight permanent forest plots in China

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25 Abstract. Forest soils represent a major stock of organic carbon (C) in the terrestrial biosphere, but the dynamics of soil organic carbon-C (SOC) stock are poorly quantified, especially based 26 27 onlargely due to lack of direct field measurements. In this study, we investigated the 20-year changes in the SOC stocks inat eight permanent forest plots, which represent boreal (1998-28 29 2014), temperate (1992–2012), subtropical (1987–2008), and tropical forest biomes (1992– 2012) across China from southern to northern China. SOC contents increased significantly 30 from the 1990s to the 2010s, mostly in the upper 0–20 cm soil depth, and soil bulk densities 31 32 do not change significantly during the same period. As a result, **F**the averaged SOC stocks increased significantly from 125.2±85.2 Mg C ha⁻¹ in the 1990s to 133.6±83.1 Mg C ha⁻¹ in 33 the 2010s across the forest plots, with a mean increase of 127.2–907.5 kg C ha⁻¹ yr⁻¹. This 34 SOC accumulation was resulted primarily from both increasing leaf litter and fallen logs, 35 which accounts and equivalent to 3.6-16.3% of above-ground net primary production. Our 36 37 findings provided strong direct evidence that China's forest soils have been acting as 38 significant carbon C sinks, although their strength varies in with forests in with different 39 climates. 40 Keywords: soil organic carbon, carbon cycle, forest ecosystems, global change, permanent

41 plot

42 **1 Introduction**

43 Terrestrial ecosystems have absorbed approximately 30% of the carbon dioxide (CO₂) emitted 44 from human activities since the beginning of the industrial era (IPCC, 2013). Forests have 45 contributed more than half of these carbon (C) fluxes of terrestrial ecosystems (Pan et al., 2011). Since soils contain a huge C stock in forest ecosystems, even a slight change in this 46 47 stock will induce a considerable feedback to the atmospheric CO₂ concentrations (Lal, 2004; Luo et al., 2011). Thus, accurate assessment of the changes in soil organic carbon (SOC) is 48 49 critical to understanding how forest soils will respond to global climate change. However, it is difficult to capture the SOC change with short-term measurements (Smith, 2004) because the 50 51 soil C pool typically has a longer turnover time and higher spatial variability compared to the 52 vegetation C pool (Schrumpf et al., 2011; Canadell and Schulze, 2014). 53 Previous efforts have estimated the changes in regional SOC stocks with indirect 54 approaches, such as regional assessments (Yang et al., 2014) and model simulations 55 (Todd-Brown et al., 2013). These estimates often involve large uncertainties due to the 56 inherently high spatial variability of soils and lack of direct measurements representing large 57 areas (Sitch et al., 2013). One reliable approach to reducing the uncertainties is to conduct long-term monitoring of forest SOC stocks at sites that represent broader landscapes (Prietzel 58 59 et al., 2016). Unfortunately, such repeated, accurate field-based measurements of SOC stocks 60 from which to generate change estimates are generally lacking and inadequate worldwide (Zhao et al., 2019). 61 62 There are aA few soil resampling studies that have explored the SOC changes in different 63 forests, but the results are often <u>contradictorycontrary</u>. For instance, Schrumpf et al. (20112014) found that SOC in deciduous broadleaved forests in central Germany increased, 64 with a change rate of 650.0 kg C ha⁻¹ yr⁻¹ from 2004 to 2009. In contrast, Prietzel et al. (2016) 65 indicated that SOC stocks in the German Alps forests decreasedhad a significantly, decrease 66

with a average change rates of 988.2732 kg C ha⁻¹ yr⁻¹ in forests in the Alps between 1987-67 1986 and 2011, and 441.1 kg C ha⁻¹ vr⁻¹ in the Berchtesgaden region between 1976 and 2011. 68 69 Kiser et al. (2009) found that the hardwood forest soils in central Tennessee, USA, exhibited a slight C source (-11 kg C ha⁻¹ yr⁻¹) between 1976 and 2006, and that the relative change rate 70 ranged from -0.4% yr⁻¹ to 0.3% yr⁻¹ between 1976 and 2006. Chen et al. (2015) synthesized 71 global SOC changes, and found that the relative rates of change-rates inof forest SOC stocks 72 were contradictory among long-term experiments (0.249% yr⁻¹), regional comparisons (0.34%73 yr⁻¹), and repeated soil samplings (-0.14% yr⁻¹). Such discrepancies can be partly attributed to 74 the insufficient observations and inconsistent methodologies. It may also involve The different 75 76 effects of changing environmental factors and nitrogen inputs on soil C dynamics may also be 77 involved (Norby and Zak, 2011). In addition, to date these studies have were primarily been conducted in the forests of Europe and the USAUnited States, but few have been carried out 78 79 in China's forests.

80 Forests in China, with cover an area of 156 Mha (Guo et al., 2013), and range span from 81 boreal coniferous forests and deciduous broadleaved forests in the northeast to tropical rain 82 forests and evergreen broadleaved forests in the south and southwest. They include, -covering almost all major forest biomes of the Northern Hemisphere (Fang et al., 2012). Such variations 83 in climate and forest types have provided ideal opportunities venues to examine the spatial 84 85 patterns of SOC in relation to meteorological and biological factors. At the national scale, the mean annual air temperature of China has increased by more than 1 °C between 1982 and 2011, 86 87 which is considerably higher than the global average (Fang et al., 2018). Since the 1980s, the 88 government Government of China has implemented several large-scale Nationalnational-Forest forest- Protection projects. These climatic changes and conservation practices 89 90 in China have significantly stimulated carbon-C uptake into forest ecosystems (Fang et al., 2014, 2018; Feng et al., 2019). Several studies have assessed the temporal dynamics of SOC 91

92 stock across China's forests, using model simulations (Piao et al., 2009) or regional 93 assessments (Pan et al., 2011; Tang et al., 2018). However, these estimates revealed 94 contrasting trends of in SOC dynamics and also lacked direct measurements of SOC change. 95 Therefore, in this study we measured SOC density (carbon C amount per unit area) of eight permanent forest sites plots from tropical, subtropical, temperate, and boreal forests in 96 97 China duringat two periods inof the 1990s and 2010s to quantify their SOC changes. We then 98 analyzed the potential biotic and climatic drivers in the SOC dynamics across these forests. 99 We fFinally, we assessed the changes inof SOC stocks in China's forests using the site data 100 obtained from this study.

101

102 **2** Materials and methods

103 **2.1** Study sites

104 From north to south, We investigated eight permanent forest plots infrom four forest sites 105 (from north to south: Great Xing'anling, Mt. Dongling, Mt. Dinghu, and Jianfengling) were-106 investigated (Fig. 1). The four sites spanned a wide range from 18.7 °N to 52.6 °N in latitude, 107 and belonged to boreal, temperate, subtropical, and tropical climate zones, respectively, with a 108 climatic difference of approximately 26 °C in mean annual temperature and 1,200 mm in 109 mean annual precipitation. The eight plots comprised included a boreal larch forest (Larix 110 gmelinii), two temperate deciduous broadleaved forests (Betula platyphylla and Quercus 111 wutaishanica), a temperate pine plantation (Pinus tabuliformis), a subtropical evergreen 112 broadleaved forest, a subtropical pine plantation (P.inus massoniana), a subtropical pine and 113 broadleaved mixed forest, and a tropical mountain rainforest (for details, see Table 1). 114 Stand characteristics of all eight plots are summarized in Table 1. The boreal larch forest 115 was a 100-year-old mature stand at the time of the first sampling (Wang et al., 2001). Three temperate forest plots (birch, oak, and pine forests) were located along an elevation gradient 116

117 onat Mt. Dongling, Beijing. Both birch and oak forest plots were 55-year-old secondary 118 forests at the time of the first sampling, dominated by *B. platyphylla* and *Q. wutaishanica*, 119 respectively. The temperate pine plantation was 30-years-old at the time of the first sampling, 120 and was dominated by P. tabuliformis (Fang et al., 2007). Three subtropical forest plots were 121 located inat Dinghu Biosphere Reserve in Guangdong Province, South China (Zhou et al., 122 2006). The subtropical evergreen broadleaved forest was an old-growth stand more than 400 123 years old, co-dominated by Castanopsis chinensis, Canarium pimela, Schima superba, and 124 Engelhardtia roxburghiana. The subtropical pine (P. massoniana) plantation was 125 approximately 40 years old at the time of the first sampling. The mature mixed pine and 126 broadleaved forests was approximately 110 years old at the time of the first sampling, and 127 which represented the mid-successional stages of monsoon evergreen broadleaved forest in 128 this region. The tropical mountain rainforest plot was located at the Jianfengling National 129 Natural Reserve, southwestern Hainan (Zhou et al., 2013).; Itit hads not been disturbed for 130 more than 300 years, and iwas dominated by species in the families Lauraceae and Fagaceae, 131 such ase.g., Mallotus hookerianus, Gironniera subaequalis, Cryptocarya chinensis, 132 Cyclobalanopsis patelliformis and Nephel-ium topengii. For detailed descriptions on these eight plots, see Supplementary Materials and Methods. 133

134

135 2.2 Soil sampling and calculation of SOC content

The first sampling was conducted between 1987 and 1998 <u>inat</u> each of the eight forests (Table
1). We re-measured the same sample plots <u>inat</u> each forest between 2008 and 2014 using
identical sampling protocols.

<u>InAt</u> each forest plot, 2–5 soil-pits were dug to collect <u>soil</u>the samples for analyzing the
 physical and chemical properties during the two sampling periods (most in the 1990s during
 the first sampling period and in the 2010s during the second sampling period). The samples

142 were taken at a-depth intervals of 10 cm down to the maximum soil depth. In brief, for the 143 boreal forest, three soil pits were established down to the 40-cm soil depth were established in 144 random locations in the growing season of in 1998. In August 2014, three soil pits were again 145 randomly excavated again to the same soil depth to allow sampling for SOC content and bulk 146 density. For the three temperate forests, two soil profiles (100 cm depth) were dug in each plot 147 to collect soil samples at 10 cm intervals during the summer of 1992. In the summer of 2012, 148 three soil profiles were dug, and soils were sampled from the same-respective horizons in 149 each soil profile (Zhu et al., 2015). For the three subtropical forests, tThe first sampling in the 150 three subtropical forests was conducted in September of 1988 infor the evergreen and the pine 151 plots, and in 1987 for the mixed plot, both at the end of the rainy season and at the beginning 152 of the dry season. Five soil pits (60 cm depth) were randomly excavated to collect samples for 153 the calculation of SOC content and bulk density. In September 2008, the soil sampling was 154 repeated. For the tropical forest, five soil profiles (100 cm depth) were established at 10 cm 155 intervals during summer in 1992 and again in summer 2012.

156 We used consistent sampling and analysis approaches to determine the bulk density and 157 SOC content between the two sampling times. Three bulk density samples were obtained for each layer using with a standard container with 100 cm³ in volume. The soil moisture was 158 159 determined by weighing to the nearest 0.1 g after 48 h oven-drying at 105 °C. The bulk 160 density was calculated as the ratio of the oven-dried mass to the container volume. Another 161 three paired samples for C analysis were air-dried, the fine roots removed-off the fine roots by hand, and sieved (2 mm mesh). The SOC content was measured using the wet oxidation 162 163 method (Nelson and Sommers, 1982). The SOC content and was calculated according to Eq.uation (1): 164

165

$$SOC = \sum_{i=1}^{n} CC_i \times Bd_i \times V_i \times HF_i$$
(1)

166 where CC_i , Bd_i , and V_i are SOC content (%), bulk density (kg m⁻³), and volume (m³) at the

167 *i*-th soil horizon, respectively. HF_i is calculated as $1 - \frac{\text{stone volume+root volume}}{V_i}$ and is a 168 dimensionless factor that represents the fine soil fraction within a certain soil volume. 169

2.3 Calculation of above-ground biomass <u>(AGB)</u> and net primary production

171 Diameter at breast height (DBH, 1.3 m) and height of all living trees with DBH > 5 cm were

172 measured <u>inat</u> each plot in <u>the 1990s</u> and 2010s. The above-ground biomass (AGB) of

173 different components (stem, bark, branches, and foliage) was estimated for all tree species

using the allometric equations (Table S1). A standard factor of 0.5 was used to convert

175 biomass to C (Leith and Whittaker, 1975). The net increment of AGB (Δ Store) was calculated

176 for each plot as the difference between the biomass in the 1990s and the 2010s. The

above-ground net primary production (ANPP, kg C ha⁻¹ yr⁻¹) was calculated from Eq.uation

178 (2):

179 $ANPP = Litterfall + \Delta Store + Mortality$ (2)

180 where Litterfall and Δ Store are litter production and above-ground net biomass increment per 181 year, respectively. Mortality (defined as above-ground dead wood production) was estimated 182 as the summed production of fallen logs and standing snags per year.

183

184 **2.4 Litter and fallen log production**

Annual litterfall was collected from June 2010 to June 2013 in the tropical sites $\frac{1}{25}$ from June 1990 to June 2008 in the subtropical sites $\frac{1}{25}$ from April to November of 2011–2014 in the temperate sites $\frac{1}{25}$ and from May to October of 2010–2014 in the boreal sites. Litter (leaves, flowers, fruits, and woody materials <_2 cm diameter) was collected monthly from 10–15 litter traps (1 × 1 m², 1 m above ground) in each plot to calculate annual litter production. After collection, the samples were taken to the laboratory, oven-dried at 65 °C to a constant mass and weighed. The 10–15 replicates from feach plot were averaged as the monthly mean value. Annual litter production (kg C ha⁻¹ yr⁻¹) was estimated as the sum of the monthly
production in the year of collection.

Log production represents the mortality (that is, death of entire trees) per year. Annual
log production was determined from 2010 to 2013 in tropical sites¹/₂₅ from 1989 to 1996 in
subtropical sites¹/₂₅ from 2011 to 2014 in temperate sites¹/₂₅ and from 2010 to 2014 in boreal sites.
Stocks of fallen logs were harvested and weighed during each investigated year.

198

199 2.5 Forest area and fossil fuel emission data

200 In order to figure out To calculate the amount of C sequestration size in China's forest soils, 201 we estimated the changes in the national forest SOC stocks., We usedusing the mean SOC 202 accumulation rates obtained from this study and the data of forest area for each forest type 203 documented in the national forest inventory during the period of in 1989–1993, which is close-204 to-approximates the first sampling period in the present study (Guo et al., 2013). The changes 205 in the national forest SOC stock were calculated as the product of SOC density, SOC density 206 change rate, and forest area for major forest types during the period of 1989–1993. In addition, 207 to evaluate the relative importance of forest soil C sequestration in the national C budget, we obtained the data of fossil fuel emissions during 1991-2010 from the Carbon Dioxide 208 209 Information Analysis Center (Zheng et al., 2016).

210

211 3 Results

212 **3.1** Changes in SOC

Soil organic carbonOC stocks were investigated <u>inat</u> eight permanent forest plots <u>in from</u>-four
forest sites from northern to southern China, <u>inat</u> two periods: (around <u>the 1990s</u> and 2010s).
The eight plots spanned a wide range from 18.7 °N to 52.6 °N in latitude, and included aboreal larch forest (*Larix gmelinii*) in Great Xing'anling, two temperate deciduous-

217 broadleaved forests (Betula platyphylla and Quercus wutaishanica) and a temperate pine-

218 plantation (*Pinus tabuliformis*) in Mt. Dongling, a subtropical evergreen broadleaved forest, a

219 subtropical pine plantation (*Pinus massoniana*) and a subtropical pine and broadleaved mixed-

220 forest in Mt. Dinghu, and a tropical mountain rainforest in Mt. Jianfengling (Fig. 1 and Table-

221 1).

222 The changes in SOC contents, bulk density, and SOC stocks in the top 20 cm soil layer 223 between the 1990s and the 2010s arewere shown in Fig. 2, Fig. S1 and Fig. S2. The paired 224 *t*-test analysis indicated that SOC contents in the 0-20 cm0-to 20 cm0 depth was significantly higher in the 2010s than those in the 1990s ($3.22\pm0.265\%$ vs. $2.285\pm0.63\%$; t = -5.65, P < -5.65225 0.001) (Table 2). The average rate of increase rate inof SOC content was 0.018±0.0042% yr⁻¹ 226 in the top 20 cm depth, ranging from 0.01% yr⁻¹ $\stackrel{1}{3}$ to 0.0439% yr⁻¹ across the study sites. These 227 rates of increase rates of increase in SOC content in the 0–10 cm horizon ($0.03\pm0.020\%$ yr⁻¹) 228 were three times larger than those in the 10–20 cm horizon $(0.010\pm0.0019\% \text{ yr}^{-1})$ (Table S2). 229 230 At the same time, the bulk density of the top 20 cm soil layer decreased in most-of the sites (6 ofout 8 sites), with an average rate of decrease-rate of 2.74 ± 3.768 mg cm⁻³ yr⁻¹ (Table S3). 231 232 As a result, the SOC stock in the top 20 cm soil layer was found to have increased 233 significantly in the past two decades (t = -5.85, P < 0.001, Table 2), with an average accumulation rate of 332.4 \pm 200.2 kg C ha⁻¹ yr⁻¹ (0.7 \pm ±0.40% yr⁻¹-; Fig. 2; also see Table S3). 234 235 The temperate pine plantation experienced the largest increase of in SOC stock in the top 20 cm depth (630.8 ± 111.2 kg C ha⁻¹ yr⁻¹). In contrast, the smallest rate of increase rate-was 236 observed in the subtropical mixed forest (117.3 ± 25.2 kg C ha⁻¹ yr⁻¹). It should be noted that 237 238 SOC stock in the top 20 cm depth in the subtropical evergreen old-old-growth forest increased from 35.6±6.0 Mg C ha⁻¹ in 1988 to 45.6±6.9 Mg C ha⁻¹ in 2008 (increased by 498.3±78.8 kg 239 C ha⁻¹ yr⁻¹), which lead to the highest relative accumulation rate $(1.40\pm0.22\% \text{ yr}^{-1})$ among the 240 study sites. 241

We further compared the SOC stocks of the whole soil profile between 1990s and $2010s_{\overline{1}}$ 242 with at a depth of 0-40 cm in the boreal site, 0-60 cm in the subtropical site, and 0-100 cm in 243 244 the temperate and tropical sites, between 1990s and 2010s (Fig. 3). The SOC stocks of all sampling sites in the 2010s were higher than those in the 1990s. The paired *t*-test analysis 245 revealed a significant increase in SOC stocks for the whole soil profile during the sampling 246 period (t = -4.15, P < 0.01; Table 2). The mean SOC stocks of the whole soil profile in the 247 248 eight forests increased from 125.2±85.2 Mg C ha⁻¹ in the 1990s to 133.6±83.1 Mg C ha⁻¹ in the 2010s, with an accumulation rate of 421.2±274.4 kg C ha⁻¹ yr⁻¹ and a relative increase rate 249 250 of 0.56±0.54% (Fig. 2). The SOC accumulation rates of SOC displayed large variability among different climate zones and forest types. For different climate zones, the SOC 251 252 accumulation rates in the subtropical and tropical sites were relatively higher than those in the boreal and temperate sites (Fig. 3). The greatest increase of thein SOC stock occurred in the 253 subtropical evergreen old-old-growth forest (907.5±60.1 kg C ha⁻¹ yr⁻¹), and the least-one-254 occurred in the temperate deciduous oak forest (127.2±25.3 kg C ha⁻¹ yr⁻¹; Table S3). The 255 256 relative rates of increase rates in the subtropical evergreen old-growth forest $(1.33\pm0.109\%)$ yr⁻¹) and the subtropical mixed forest $(1.495\pm0.162\% \text{ yr}^{-1})$ were higher than those in the 257 temperate forests ($0.105\pm0.01\%$ yr⁻¹ in the oak forest, $0.14\pm0.03\%$ yr⁻¹ in the pine forest, and 258 $0.492\pm0.02\%$ yr⁻¹ in the birch forest; Table S3). 259 260 In addition, the rates of SOC increase rate (127.2–907.5 kg C ha⁻¹ yr⁻¹) was equivalent to

261 3.6–16.3% of ANPP (3340.1–6944.7 kg C ha⁻¹ yr⁻¹), with the highest rate in the subtropical 262 evergreen forest (16.3±4.2%) and the lowest in the temperate oak forest (3.6±3.4%) (Tables 3_ 263 <u>and ; Table S4</u>).

264

3.2 Relationships between SOC change rates and biotic and climatic variables

266 To understand the possible mechanisms for the <u>rates of SOC</u> increase rates as described above,

we analyzed the driving forces for this significantly increased SOC stock using measurements 267 268 of AGB growth rate, above-ground litter and fallen log production, and ANPP (Table 3). The 269 linear regression analysis showed that there was no significant correlation between SOC change rates and AGB growth rate (P => 0.23705; Fig. 4a). The SOC accumulation rates 270 were positively and significantly associated with annual litter ($R^2 = 0.66$, P = 0.01; Fig. 4b) 271 and fallen log production ($R^2 = 0.69$, P = 0.01; Fig. 4c). The SOC accumulation rates across 272 these forests were closely associated with the observed ANPP ($R^2 = 0.55$, P = 0.034,; Fig. 4d), 273 274 and also showed an increasing trend with increasing mean annual temperature and 275 precipitation, despite insignificant (both P > 0.1; Figs. 4e and Fig. 4f). The multiple regression 276 analysis indicated the relative effects of biotic factors (AGB growth rate, litter and fallen log 277 productions) and climatic factors (mean annual temperature and precipitationMAT and MAP) 278 on the rates of SOC increase-rates (Fig. 4g). When the effects of climatic factors were under 279 control, the biotic factors independently explained 56.4% of the variations. By comparison, 280 when the effects of biotic factors were under control, only 7.5% of the variations were 281 explained by the climatic factors.

282

283 4 Discussion

284 4.1 SOC accumulation

Previous evidence of the forest SOC changes comes mainly from individual experiments
(Prietzel et al., 2006; Kiser et al., 2009; Häkkinen et al., 2011) or regional comparisons
(Lettens et al., 2005; Pan et al., 2011; Ortiz et al., 2013) in European and American forests. In
this study, we performed a broad-scale forest soil resampling to evaluate changes in SOC
stock across eight permanent forest plots in China. Our measurements suggest that SOC
stocks exhibited a significant accumulation in these forests from the 1990s to the 2010s, at the
accumulation rate of 127.2–907.5 kg C ha⁻¹ yr⁻¹. These accumulation rates are comparable to

292	those of the other studies that were primarily conducted in boreal and temperate forests in the
293	other regions (-11.0–812.0 kg C ha ⁻¹ yr ⁻¹ , Fig. 5). In detail, the <u>rate of</u> SOC accumulation-rate
294	of the boreal forest in the present study was estimated as 243.4 kg C ha ⁻¹ yr ⁻¹ , which was
295	within the range of boreal forests in European and American forests (210115.6-652-740.0 kg
296	C ha ⁻¹ yr ⁻¹) (Prietzel et al., 2006; Häkkinen et al., 2011; <u>Rantakari et al., 2012;</u> Chapman et al.,
297	2013; Schrumpf et al., 2014). The rates of SOC accumulation rates in of the three temperate
298	forests ranged from 127.2 to 390.8 kg C ha ⁻¹ yr ⁻¹ , comparable to the regional comparisons
299	data of 200 <u>.0</u> kg C ha ⁻¹ yr ⁻¹ in the temperate forests of China (Yang et al., 2014). Evidence
300	from soil inventory-based studies of SOC dynamics also demonstrated that soil of boreal and
301	temperate forests in European countries is likely to accumulate C (Berg et al., 2009; Nielsen
302	et al., 2012; Grüneberg et al., 2014). The mean rate of SOC accumulation in the humus layers
303	of boreal forests in Sweden was estimated to be 251.0 kg C ha ⁻¹ yr ⁻¹ during the period 1961-
304	2002 (Berg et al., 2009). Nielsen et al. (2012) assessed the rates of SOC change in Denmark's
305	broadleaved deciduous and coniferous forests by two soil inventories conducted during 1990
306	and 2005. The estimated rates of SOC change in the broadleaved and coniferous forests were
307	90.0 and 310.0 kg C ha ⁻¹ yr ⁻¹ , respectively. Two soil inventories provided data for analysis of
308	the mineral soils of forests in Germany, which were found to have sequestrated 410.0 kg C
309	ha ⁻¹ yr ⁻¹ during the period of 1987–2008 (Grüneberg et al., 2014). Therefore, evidence from
310	long-term observations, and from the repeated soil sampling in individual studies and in
311	national soil inventory reports, suggests that soils of boreal and temperate forests in the
312	northern hemisphere have functioned as C sinks during past decades.
313	In other subtropical and tropical forest ecosystems, the direct evidence of regarding SOC
314	dynamics is relatively scarce. However, based on the estimates from regional comparisons,
315	Pan et al. (2011) showed that <u>global</u> tropical forests of the world-wereas a C-source of 1.38 4
316	Pg C ha ⁻¹ yr ⁻¹ from 1990 to 2007. At the global scale, tropical land-use changes have caused a

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317 sharp drop in forest area, which also led to a large release of C from release in tropical forest 318 soils. Without land-use change and deforestation, soils inof the subtropical and tropical forests 319 have been functioning as a considerable C sink during the past two decades in this study (627.6±370.1 and 397.9±84.2 kg C ha⁻¹ yr⁻¹, respectively, Table 3). Not only catastrophic-320 321 land-use changes, but also slightLimited forest management (e.g., litter and dead wood 322 harvest), as well as catastrophic land-use changes, can result in the loss of C from forest soil-323 earbon. Prietzel et al. (2016) reported a large loss of SOC in forests in the German Alps-324 forests, where half of the woody biomass and dead wood hadhave been harvested over the 325 recent decades. On the one hand, the harvesting of the forest floor canould decrease litter and 326 dead wood inputs into soils and subsequently lead to the loss of soil Cearbon (Davidson and 327 Janssens, 2006). On the other hand, a decreased in the amount of the forest floor maycould lead to an increase inof soil erosion, especially in the mountain forests (Evans et al., 2013). 328 329 Additionally, the high-elevation ecosystems are expected to be more sensitive to warming 330 than other regions, with associated changes in soil freezing and thawing events and in snow 331 cover, which maymight be another reason for the SOC losses in forests in theof German Alps-332 forests.

333

4.2 Links between biotic and climatic factors and <u>in</u> SOC accumulations

<u>The f</u>Forest biomass of China has functioned as a significant C sink over-the recent decades
(Pan et al., 2011; Fang et al., 2014, 2018). <u>The increase in C accumulation by Increased</u>
vegetation-<u>C accumulation</u> supplied more C inputs into soils, including inputs of litter, woody
debris₁ and root exudates, and resulted in SOC accumulation (Zhu et al., 2017). However, the_
<u>rate of</u> SOC change-rate did not increase with the <u>rate of</u> biomass change-rate in this study
(Table S4). We found that soil in the subtropical old-growth forest increased at the highest
sink rate of 907.5±60.1 kg C ha⁻¹ yr⁻¹, but thate vegetation functioned as a significant C source

(-1000.3±78.2 kg C ha⁻¹ yr⁻¹). This <u>wasis</u> because the relatively higher annual litterfall and
fallen log production occur<u>red</u> in the old-growth forest, which subsequently result<u>ed</u>s in soil C
accumulation (Fig. 4). The positive (but not significant) trend between climatic factors and
SOC dynamics <u>mayeould be</u> largely <u>be</u> induced by the internal correlations between climatic
and biotic factors (Fig. 4).

347 The heterotrophic respiration of global forest soil has increased significantly over the past decades (Bond-Lamberty et al., 2018), suggesting that the increment of in the rate of soil 348 349 carbonC input-rate outweighs that of the rate of soil Ccarbon output-rate. The increasing 350 heterotrophic respiration of forest soil is mainly due to the ongoing climate changes, and 351 especially to increasing temperature. Whilst Tthe increment inof forest growth rate is due to 352 increasing temperature, together with increasing CO₂ and nitrogen fertilization (Norby et al., 353 2010; Feng et al., 2019). Thus, the sensitivity of forest net primary production NPP to ongoing climate changes should outweigh that of respiration. Additionally, Wwe also found that SOC 354 stock increased from 68.4 Mg C ha⁻¹ to 86.6 Mg C ha⁻¹, albeit the biomass C stock decreased 355 356 significantly from 1988 to 2008 in the subtropical old-growth plot. Meanwhile, Tthe 357 greatesthighest amount of litter and dead wood production and standing crop occurred in the old-growth plots, which resulted in a relatively higher soil C sequestration in the old-growth 358 359 plot compared to other plots (Fig.ure 4, Table S4). Compared with climatic factors, Bbiotic 360 factors explained the variation inof SOC dynamics better than climatic factors. Note that Iin 361 this study, we did not, however, measure the root-derived Ccarbon inputs to the SOC, although the below-ground production also makes a significant contribution to the SOC 362 363 accumulation (Nadelhoffer and Raich, 1992; Majdi, 2001; Pausch and Kuzyakov, 2018). Above-ground inputs are mineralized from litter and dead wood, and below-ground inputs 364 365 may benefit from interactions with soils-In addition to above ground inputs being mineralized from litter and dead wood, below-ground inputs may have more opportunity for interactions-366

with soils (Rasse et al., 2005). Even if the effect of the climatic factors were controlled and
 below-ground biotic factors were not included in the analysis, the above-ground biotic factors
 <u>ew</u>ould explain 56.4% of the variation <u>in theof rate of SOC accumulation-rate</u>.

370

371 4.3 <u>Regional c</u>Carbon budget of China's forests

The <u>rate of SOC accumulation-rate</u> (421.2 ± 274.4 kg C ha⁻¹ yr⁻¹, Fig. 2 and Table S3) is more than <u>one-half of the vegetation C uptake rate in China's forests (702.0 kg C ha⁻¹ yr⁻¹) (Guo et al., 2013; Fang et al., 2018). This result suggests that China's forest soils have contributed to a negative feedback to climate warming during the past two decades, rather than the positive feedback predicted by coupled <u>Cearbon-climate models</u> (Cox et al., 2000; He et al., 2016; Wang et al., 2018).–</u>

378 If we roughly use the inventory-based forest area of 138.8 Mha in China (Guo et al., 2013) 379 and extend the current SOC sink rates obtained in this study to all the forests in the country, 380 China's forest soils have sequestered approximately 1.14±0.53 Pg C overduring the past two decades (57.1±26.57 Tg C yr⁻¹). This C accumulation would be equivalent to 2.4–6.8% of the 381 382 country's fossil CO₂ emissions during the contemporary period (1991–2010) (Zheng et al., 383 2016). By comparing forest SOC data obtained from published literatures during the 2000s and a national soil inventory during the 1980s, Yang et al. (2014) estimated significant 384 Ccarbon accumulation in the forest soils of China. Although they did not estimate the national 385 386 C budget of these forest soils, we can calculate the national C sequestration rate of forest soil as 67.2 Tg C yr⁻¹, based on the C sequestration rates and forest areas of the different forest 387 388 types in their study. Our results further confirm the assessment, based on repeated 389 measurements at eight permanent forest plots, that soils in China's forests have functioned as 390 a carbon C sink for atmospheric CO_2 over during the past two decades. 391 According to previous estimates, the C sinks of three C sectors, i.e.: forest vegetation

392 biomass (Fang et al., 2014), dead wood, and litter (Zhu et al., 2017), duringover the past two decades were 740.9, 3.94, and 32.8 Tg C yr⁻¹, respectively (Table S5). If incorporating these 393 previous estimates are incorporated into the soil C accumulation rate of 57.1±26.57 Tg C yr⁻¹ 394 in the current study, then China's forests maycould have sequestered a total of ~135134.7 Tg 395 396 C per year between the 1990s to and the 2010s., Thiswhich is equivalent to 14.5% of the 397 contemporary fossil CO₂ emissions in the country (Zheng et al., 2016). According to the 398 estimate of Pan et al. (2011), the C sink rate of forests in the temperate regions of the northern hemisphere was 647.1 Tg C yr⁻¹. The C sequestration of China's forests represents 20.8% of 399 400 the total temperate regions. The sequestration rate of China's forests is slightly higher than the mean value of the total temperate regions, relative to the forest area of China (i.e., 18.9% of 401 the forest areas in the temperate regions). This result indicates that the role of forest soils in 402 403 the regional C cycle cannot be ignored, although a large uncertainty about the national C 404 budget of forest soils remains in our estimates.

405

406 **4.4 Uncertainty analysis**

We investigated the SOC stocks at-<u>in</u> eight permanent plots across four forest biomes in China.
These plots spanned a long-term timescale (approximately 20 years) and a broad spatial scale
(approximately 34 ° of latitude). We also measured several C fluxes (i.e., biomass change rate,
production of litterfall and dead wood) that were relevant to the <u>rate of SOC change rates</u>.
Even so, the following three <u>factors aspects-may introduce produce</u> uncertainties related to the
estimation of SOC dynamics.

First, the sampling times and intervals <u>betweenof</u> SOC investigations were different across the sites. The first sampling was performed from 1987 to 1998 and the second sampling was carried out from 2008 to 2014. As a result, the sampling interval ranged from 16 years in <u>the</u> boreal forest plot to 21 years in the subtropical mixed forest plot (Table 1). Non-uniform

sampling times and intervals <u>maycould</u> lead to uncertainties <u>in relation toof</u> SOC stocks across
the forest plots.

419 Second, the depth of soil varied substantially, ranging from 40 cm in the boreal site to 100 cm in the temperate and tropical sites. In addition, different numbers (2-5) of soil profiles 420 were dug infor different plots were dug during the first sampling period. To ensure consistency 421 422 between the two sampling times, the same number of soil profiles were dug, and in similar 423 locations, to perform SOC stock investigations during the second sampling periodsoil profiles-424 with the same number and similar locations were dug to perform the SOC stocks investigation 425 during the second sampling period. We performed continuous observation for of litterfall and 426 dead wood production, but the observation times and durations varied across the plots. 427 Variabilitynces inof these items maymight reduce the comparability of SOC dynamics among 428 plots.

Finally, the <u>rates of SOC</u> change <u>rates of in</u> our study and <u>in</u> inventory-based forest areas and forest types were used to estimate the <u>Ccarbon</u> budget of forest soil <u>inof</u> China. However, only eight permanent forest plots were observed in this study, <u>and this</u> will inevitably lead to uncertainty <u>with respect to for national estimations</u>.

433

434 5 Conclusions

The SOC stocks within the top 20 cm depth-increased by 2.4—12.6 Mg C ha⁻¹ across the forests during the past two decades, with an annual accumulation rate of 332.4 ± 200.2 kg C ha⁻¹. If all horizons of soil horizon profiles were included, the soils may have been found to have sequestered 3.6—16.3% of the annual net primary production across the investigated sites, and the averaged accumulated rate (421.2 kg C ha⁻¹ yr⁻¹) may have been is more than <u>one-half of the vegetation C uptake rate (702.0 kg C ha⁻¹ yr⁻¹) in China's forests. These results</u> demonstrate that these forest soils have functioned as an important C sink over the recent

442	decades, although the phenomenon may not happenoccur uniformly everywhere in forests
443	around the world <u>wide</u> . Forest soils store large amounts of C ₂ and accumulate <u>it</u> \subseteq steadily and
444	often slowly, but will release it rapidly release C to the atmosphere once they are disturbed.
445	

446 *Data availability.* All relevant data are available from the corresponding author upon
447 request.

448

449 Author contributions. JF designed the research; JZ and JF designed the data analysis. JZ, JF,

450 ZZ, LJ, XH, HY, GL, CW and GZ performed SOC measurements. JF, YL, CJ and GL

451 designed sampling and analytical programmes and performed data quality control. JZ, JF, CW,

452 SZ, PL, JZ, ZT, CZ, RB and YP contributed to the writing of the manuscript.

453

454 *Competing interests.* The authors declare no competing interests.

455

456 *Financial support.* This work was partly funded by National Key Research and

457 Development Program of China (2017YFC0503906), National Natural Science Foundation of

458 China (31700374, 31621091), and the US Forest Service (07-JV-11242300-117).

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Table 1. Location, forest type, mean annual temperature (MAT), and mean annual precipitation (MAP) at eight forest plots in four climate zones,

Site	Forest	Origin	Latitude (°)	Longitude (°)	Elevation (m)	Area (m ²)	MAT (°C)	MAP (mm)	Study period
Great Xing'anling (Boreal)	Larch	Mature	52 °38'42.06"N	123 %6'7.80''E	466	20×30, 25×40	-4.3	477	1998–2014
	Birch	Secondary	39 °57'05.82''N	115 °25'38.93"E	1,350	30×35	4.7	519	1992-2012
Mt. Dongling (Temperate)	Oak	Secondary	39 °57'26.66''N	115 °25'29.14"E	1,150	30×40	4.6	519	1992–2012
(Temperate)	Pine	Plantation	39 °57'33.94''N	115 °25'39.40"E	1,050	20×30	5.5	506	1992–2012
	Evergreen	Old growth	23 °10'11.21"N	112 32'21.97"E	275	50×50	20.9	1698	1988–2008
Mt. Dinghu (Subtropical)	Mixed	Mature	23 9'58.51"N	112 °32'23.32"E	265	30×40	21.6	1680	1987–2008
Subtropleary	Pine	Plantation	23 °10'02.75"N	112 °32'30.59"E	250	30×40	21.9	1677	1988–2008
Jianfengling (Tropical)	Evergreen	Old growth	18 %3'47.01"N	108 °53'23.79"E	870	100×100	20.6	1628	1992–2012

652 together with forest origin and study periods.

Table 2. Results of the paired-samples *t* tests for soil organic carbon (SOC) content, bulk

655 density, and SOC stock at different soil depths in the eight forest plots between the 1990s and

656 the 2010s.

Soil horizon	SOC content			Bulk density			SOC stock		
	t	$d\!f$	Р	t	df	Р	t	df	Р
0 - 10 cm	-4.22	7	<0.01	2.19	7	0.06	-6.50	7	<0.001
10 - 20 cm	-4.09	7	<0.01	3.30	7	<0.05	-3.26	7	<0.05
Top 20 cm	-5.65	7	<0.001	1.01	7	0.35	-5.85	7	<0.001
Whole soil profile	-	-	-	-	-	-	-4.15	7	<0.01

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- **Table 3.** Measured <u>carbon C</u> stocks and fluxes of the four forest sites in China during the
- 659 1990s and the 2010s. AGB, above-ground biomass; ANPP, above-ground net primary-

660 production. For details, see Table S1 in the supplementary information.

Parameter	Boreal	Temperate	Subtropical	Tropical	
Carbon pool (Mg C ha ⁻¹)*					
AGB	91.1±25.0	89.6±17.4	107.0±41.7	213.6±41.4	
Litter	4.4±0.0	3.9±1.3	2.1±0.7	1.8±0.2	
Dead wood	1.3±0.5	4.5 ± 1.2	7.3±6.7	5.7±0.8	
Soil	69.4±6.2	231.6±14.6	67.2±19.5	102.6±19.9	
Ecosystem total	166.2±31.7	329.6±34.5	183.7±68.5	323.7±62.3	
Carbon flux (kg C ha ⁻¹ yr ⁻¹)					
AGB growth	899.4±411.0	1809.5±521.2	798.7±1572.4	684.1±145.0	
litterfall	2424.2±283.1	1946.7±361.2	3385.4±1444.6	3970.0±279.8	
Fallen log	13.0±3.7	106.1±74.5	986.7±967.3	1034.2±71.6	
Standing snag	3.5±1.8	276.7±111.1	220.0±135.7	803.4±62.4	
ANPP	3340.1±698.8	4139.0±607.7	5390.8±1655.3	6491.6±559.2	
Soil accumulation	243.4±31.1	283.6±138.5	627.6±370.1	397.9±84.2	
Ratio of soil accumulation to ANPP (%)	7.3±7.8	6.7±2.8 (3.6~9.2)	11.0±5.3 (5.7~16.3)	6.1±3.3	

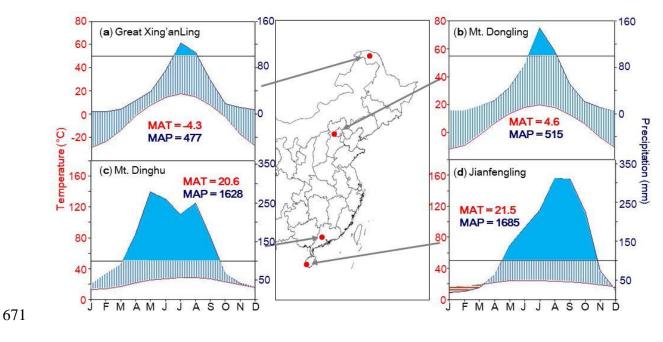
Note: Carbon pool of each ecosystem component at the time of the second sampling (2010s).

662 AGB, above-ground biomass; ANPP, above-ground net primary production. For details, see

663 <u>Table S1 in the supplementary information.</u>

664 Figures

Figure 1. Locations and climatic conditions of the sites. (a) Great Xing_an-ling, the boreal
site, (b) Mt. Dongling, the temperate site, (c) Mt. Dinghu, the subtropical site, and (d)
Jianfengling, the tropical site. The blue and red lines in climatic diagrams are the monthly
mean values of precipitation and temperature, respectively. The blue areas indicate the period
in the year when the precipitation exceeds exceeded 100 mm per month. MAT, mean annual
temperature; and MAP, mean annual precipitation.



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Figure 2. Mean soil organic carbon (SOC) content (a), bulk density (b), SOC stock (c) and
their relative change rates (d) within 0—20 cm soil depth in the 1990s and the 2010s for the
four forest sites in China. For more details, see Tables S2 in the supplementary information.

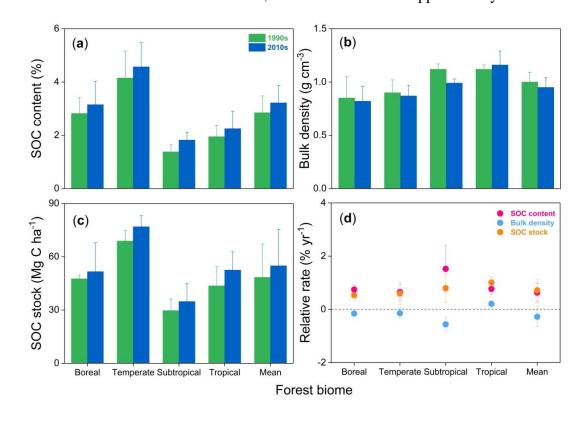
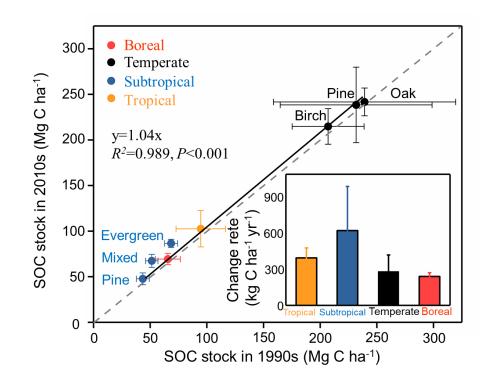
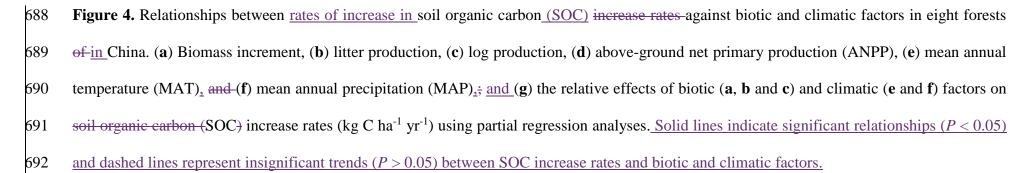


Figure 3. Comparison of soil organic carbon (SOC) stocks in eight forest plots of in China between the 1990s and the 2010s. The soil organic carbon (SOC) stocks in all forests during the two periods are above the 1:1 line, suggesting that all these forests have increased their SOC stock during the study period. The inset graph shows the SOC sink rates by forest biomes (i.e., boreal, temperate, subtropical, and tropical forests), which are categorized from the eight forest plots. SOC stocks and change rates are presented as means ± 1 SD. For details, see Fig.ure 1, Table 1, and Supplementary-Table S1.



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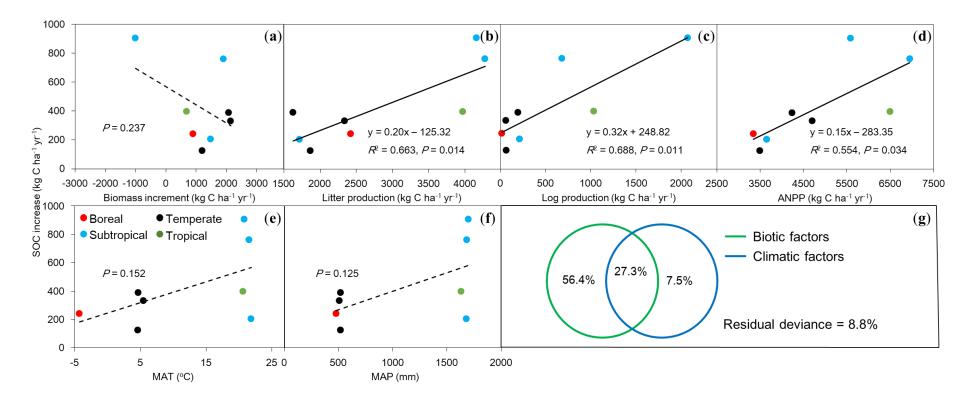


Figure 5. Comparison of the changes in forest soil organic carbon (SOC) stocks according to
repeated soil samplings and/or long-term observations. Different colors, shapes, and sizes
represent different forest biomes, ages, and soil depths, respectively. <u>The numbers in</u>
parentheses indicate the sampling times and intervals between the two soil samplings.

