

30 source during hypoxic conditions in summer especially in the southern part of the hypoxic
31 region, which is influenced by open-ocean intrusions.

32

33 **1. Introduction**

34 In coastal seas, hypoxic conditions (oxygen concentrations lower than 2 mg L⁻¹ or 62.5
35 mmol m⁻³) are increasingly caused by rising anthropogenic nutrient loads from land (Diaz
36 & Rosenberg, 2008; Rabalais et al., 2010; Fennel and Testa, 2019). Hypoxic conditions are
37 detrimental to coastal ecosystems leading to a decrease in species diversity and rendering
38 these systems less resilient (Baird et al., 2004; Bishop et al., 2006; Wu, 2002). Hypoxia is
39 especially prevalent in coastal systems influenced by major rivers such as the northern Gulf
40 of Mexico (Bianchi et al., 2010), Chesapeake Bay (Li et al., 2016), and the Changjiang
41 Estuary (CE) in the East China Sea (Li et al., 2002).

42 The Changjiang River is the largest river in China and fifth largest in the world in terms
43 of volume transport, with an annual discharge of 9×10^{11} m³ year⁻¹ via its estuary (Liu et
44 al., 2003). The mouth of the CE is at the confluence of the southeastward Yellow Sea
45 Coastal Current and the northward Taiwan Warm Current (Figure 1). Hydrographic
46 properties in the outflow region of the CE are influenced by several different water masses
47 including fresh Changjiang Diluted Water, relatively low-salinity coastal water, more
48 saline water from the Taiwan Warm Current, and high-nutrient, low-oxygen water from
49 the subsurface of the Kuroshio (Wei et al., 2015; Yuan et al., 2008). The interactions of
50 these water masses together with wind forcing and tidal effects lead to a complicated and
51 dynamic environment.

52 Freshwater discharge reaches the minimum in winter when the strong northerly monsoon
53 (dry season) prevails and peaks in summer during the weak southerly monsoon (wet
54 season) resulting in a large freshwater (FW) plume adjacent to the estuary. Along with the
55 FW, the Changjiang River delivers large quantities of nutrients to the East China Sea (ECS)
56 resulting in eutrophication in the plume region (Li et al., 2014; Wang et al., 2016). Since
57 the 1970s, nutrient load has increased more than twofold with a subsequent increase in
58 primary production in the outflow region of the estuary (Liu et al., 2015). Hypoxia off the
59 CE was first detected in 1959 and, with a spatial extent of up to 15,000 km², is among the
60 largest coastal hypoxic zones in the world (Fennel & Testa, 2019). Although no conclusive

61 trend in oxygen minima has been observed (Wang, 2009; Zhu et al., 2011), hypoxic
62 conditions are suspected to have expanded and intensified in recent decades (Li et al., 2011;
63 Ning et al., 2011) due to the increasing nutrient loads from the Changjiang River (Liu et
64 al., 2015).

65 It is generally accepted that water-column stratification and the decomposition of
66 organic matter are the two essential factors for hypoxia generation, and this is also the case
67 for the shelf region off the CE (Chen et al., 2007; Li et al., 2002; Wei et al., 2007). High
68 solar radiation and freshwater input in summer contribute to strong vertical stratification
69 which is further enhanced by near-bottom advection of waters with high salinities (> 34)
70 and low temperatures (< 19 °C) by the Taiwan Warm Current. The resulting strong
71 stratification inhibits vertical oxygen supply (Li et al., 2002; Wang, 2009; Wei et al., 2007).
72 At the same time, high organic matter supply fuels microbial oxygen consumption in the
73 subsurface (Li et al., 2002; Wang, 2009; Wei et al., 2007; Zhu et al., 2011). It has also been
74 suggested that the TWC brings additional nutrients contributing to organic matter
75 production (Ning et al., 2011) and that the low oxygen concentrations (~ 5 mg L⁻¹) of the
76 TWC precondition the region to hypoxia (Ning et al., 2011; Wang, 2009).

77 While observational analyses suggest that hypoxia off the CE results from the interaction
78 of various physical and biogeochemical processes, quantifying the relative importance of
79 these processes and revealing the dynamic mechanisms underlying hypoxia development
80 and variability require numerical modeling (Peña et al., 2010). Numerical modeling studies
81 have proven useful for many other coastal hypoxic regions such as the Black Sea
82 northwestern shelf (Capet et al., 2013), Chesapeake Bay (Li et al., 2016; Scully, 2013), and
83 the northern Gulf of Mexico (Fennel et al., 2013; Laurent & Fennel, 2014).

84 Models have also been used to study the hypoxic region of the CE. Chen et al. (2015a)
85 used a 3D circulation model with a highly simplified oxygen consumption parameterization
86 (a constant consumption rate) to investigate the effects of physical processes, i.e.
87 freshwater discharge, and wind speed and direction, on hypoxia formation. Chen et al.
88 (2015b) examined the tidal modulation of hypoxia. The model domain in these two
89 previous studies was relatively limited encompassing only the CE, Hangzhou Bay and the
90 adjacent coastal ocean but did not cover the whole area affected by hypoxia (Wang, 2009;
91 Zhu et al., 2011). Zheng et al. (2016) employed a nitrogen cycle model coupled with a 3D

92 hydrodynamic model to examine the role of river discharge, wind speed and direction on
93 hypoxia, and also emphasized the physical controls. These previous modeling studies
94 focused on the response of hypoxia to physical factors only and did not address seasonal
95 evolution and interannual variations of hypoxia or the influence of variability in biological
96 rates.

97 More recently, Zhou et al. (2017) analyzed the seasonal evolution of hypoxia and the
98 importance of the Taiwan Warm Current and Kuroshio intrusions as a nutrient source using
99 an advanced coupled hydrodynamic-biological model. However, the baseline of their
100 model does not include sediment oxygen consumption (SOC), which is thought to be a
101 major oxygen sink in the hypoxic region off the CE (Zhang et al., 2017) and other river-
102 dominated hypoxic regions including the northern Gulf of Mexico (Fennel et al. 2013, Yu
103 et al. 2015a,b). Zhou et al. (2017) acknowledged the importance of SOC based on results
104 from a sensitivity experiment but did not quantify its role in hypoxia generation.

105 Here we introduce a new 3D physical-biological model implementation for the ECS that
106 explicitly includes nitrogen and phosphorus cycling and SOC. The model is a new regional
107 implementation for the ECS of an existing physical-biogeochemical model framework that
108 has been extensively used and validated for the northern Gulf of Mexico (Fennel et al.,
109 2011, 2013; Laurent et al., 2012; Laurent and Fennel, 2014; Yu et al., 2015b; Fennel and
110 Laurent, 2018). The hypoxic zones in northern Gulf of Mexico and off the CE have similar
111 features including the dominant influence of a major river (Changjiang and Mississippi), a
112 seasonal recurrence every summer, a typical maximum size of about 15,000 km²,
113 documented P-limitation following the major annual discharge in spring and a significant
114 contribution of SOC to oxygen sinks in the hypoxic zone (Fennel and Testa 2019). Here
115 the model is used to explore the evolution of hypoxia on interannual and intra-seasonal
116 scales and to identify the main factors contributing to the different modes of variability.
117 For this study, we performed and validated a 6-year simulation in the ECS, discuss the
118 main drivers of interannual and intra-seasonal variability, and present an oxygen budget to
119 quantify the relative importance of SOC and the influence of lateral advection of oxygen.
120 A companion study by Grosse et al. (2020) uses the same model to quantify the importance
121 of intrusions of nutrient-rich oceanic water from the Kuroshio for hypoxia development off
122 the CE.

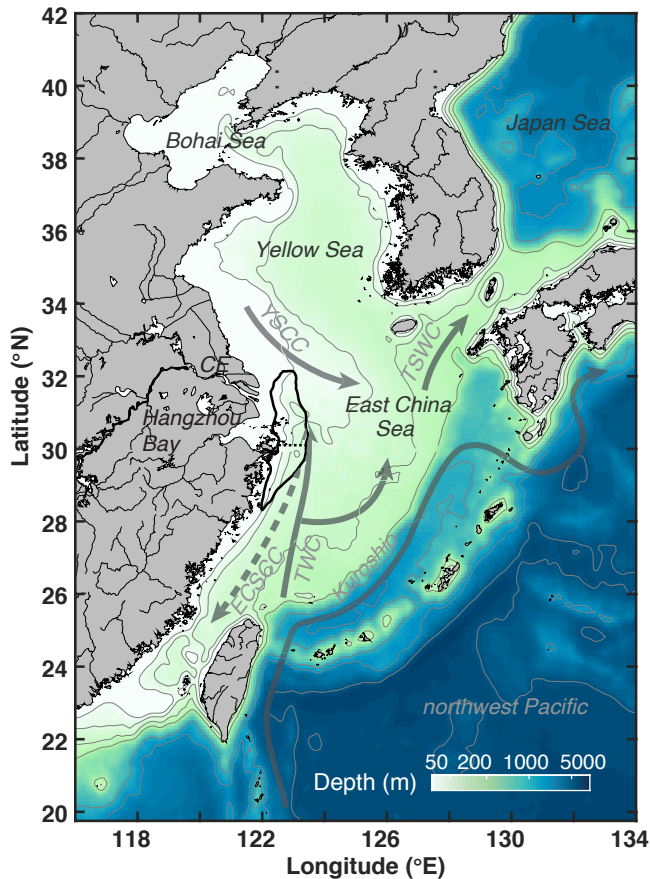
123 **2. Model description**

124 **2.1. Physical model**

125 The physical model used in this study is based on the Regional Ocean Modeling System
126 (ROMS; Haidvogel et al., 2008) and was implemented for the ECS by Bian et al. (2013a).
127 The model domain extends from 116°E to 134°E and from 20°N to 42°N (Figure 1),
128 covering the Bohai Sea, the Yellow Sea, the ECS, part of the Japan Sea and the adjacent
129 northwest Pacific, with a horizontal resolution of 1/12° (about 10 km) and 30 vertical layers
130 with enhanced resolution near the surface and bottom. The model uses the recursive
131 Multidimensional Positive Definite Advection Transport Algorithm (MPDATA) for the
132 advection of tracers (Smolarkiewicz and Margolin, 1998), a third-order upstream advection
133 of momentum, and the Generic Length Scale (GLS) turbulence closure scheme (Umlauf &
134 Burchard, 2003) for vertical mixing.

135 The model is initialized with climatological temperature and salinity from the World
136 Ocean Atlas 2013 V2 (WOA13 V2) (Locarnini et al., 2013; Zweng et al., 2013), and is
137 forced by 6-hourly wind stress, and heat and freshwater fluxes from the ECMWF ERA-
138 Interim dataset (Dee et al., 2011). Open boundary conditions for temperature and salinity
139 are prescribed from the monthly climatology (WOA13 V2), and horizontal velocities and
140 sea surface elevation at the boundaries are specified from the monthly SODA data set
141 (Carton & Giese, 2008). In addition, eight tidal constituents (M2, S2, N2, K2, K1, O1, P1
142 and Q1) are imposed based on tidal elevations and currents are extracted from the global
143 inverse tide model data set of TPXO7.2 of Oregon State University (OSU, Egbert &
144 Erofeeva, 2002). At the open boundaries, Chapman and Flather conditions are used for the
145 free surface and the barotropic velocity, respectively, and the radiation condition for the
146 baroclinic velocity. Eleven rivers are included in the model. Freshwater discharge from the
147 Changjiang River uses daily observations from the Datong Hydrological Station (DHS;
148 www.cjh.com.cn). Since daily observations are not available for the other rivers, we
149 prescribed monthly or annual climatologies (Liu et al., 2009; Tong et al., 2015; Zhang,
150 1996).

151



152

153 **Figure 1.** Bathymetry of the model domain with 30, 50, 100, 200, 1000, 2000 and 5000 m isobaths.

154 The black outline near the Changjiang Estuary (CE) and Hangzhou Bay indicates the zone typically
 155 affected by low-oxygen conditions (dotted line shows separation between northern and southern
 156 zones). Solid grey arrows denote currents present throughout the year (Kuroshio; TWC: Taiwan
 157 Warm Current; YSCC: Yellow Sea Coastal Current). The dashed grey arrow indicates the direction
 158 of the wintertime East China Sea Coastal Current (ECSCC) which flows in the opposite direction
 159 to summertime flow.

160

161 **2.2. Biological model**

162 The biological component is based on the pelagic nitrogen cycle model of Fennel et al.
 163 (2006, 2011, 2013) and was extended to include phosphate (Laurent et al., 2012; Laurent
 164 & Fennel, 2014) and riverine dissolved organic matter (Yu et al., 2015b). The model
 165 includes two forms of dissolved inorganic nitrogen (DIN), nitrate (NO₃) and ammonium
 166 (NH₄), phosphate (PO₄), phytoplankton (Phy), chlorophyll (Chl), zooplankton (Zoo), two
 167 pools of detritus, suspended and slow-sinking small detritus (SDet) and fast-sinking large

168 detritus (LDet), and riverine dissolved organic matter (RDOM). Here, riverine dissolved
169 and particulate organic nitrogen enter the pools of RDOM and SDet, respectively. The
170 remineralization rate of RDOM is an order of magnitude lower than that of SDet to account
171 for the more refractory nature of the riverine dissolved organic matter (Yu et al., 2015b).

172 At the sediment-water interface, SOC is parameterized assuming “instantaneous
173 remineralization,” i.e. all organic matter reaching the sediment is remineralized
174 instantaneously and oxygen is consumed due to nitrification and aerobic remineralization
175 at the same time. In the “instantaneous remineralization”, all phosphorus is returned to the
176 water column as PO₄ while a constant fraction of fixed nitrogen is lost due to denitrification.
177 All biogeochemical model parameters are given in Table S1 in the Supplement. A more
178 detailed model descriptions can be found in the Supplement to Laurent et al. (2017).

179 Light is vertically attenuated by chlorophyll, detritus and seawater itself. In addition, to
180 account for the effects of colored dissolved organic matter (CDOM) and suspended
181 sediments, which show relatively high values near the coast and in the river plume (Bian
182 et al., 2013b; Chen et al., 2014), a light-attenuation term dependent on water depth and
183 salinity is introduced which yields higher attenuation in shallow areas and in the FW plume.

184 Initial and boundary conditions for NO₃, PO₄ and oxygen are prescribed using the
185 World Ocean Atlas 2013 (WOA13) climatology (Garcia et al., 2013a,b). A small positive
186 value is used for the other variables. NO₃ is nudged towards climatology in the northwest
187 Pacific at depth > 200 m. Monthly nutrient loads of NO₃ and PO₄ from the Changjiang are
188 from the Global-NEWs Model (Wang et al., 2015) but were adjusted by multiplicative
189 factors of 1.20 and 1.66, respectively, to ensure a match between simulated and observed
190 nutrient concentrations in the CE (see July and Aug 2012 in Figure 2). Nutrient loads in
191 other rivers are based other published climatologies (Liu et al., 2009; Tong et al., 2015;
192 Zhang, 1996).

193 We performed an 8-year simulation from 1 January 2006 to 31 December 2013, with
194 2006-2007 as model spin up and 2008-2013 used for analysis. Model output was saved
195 daily.

196

197 **3. Results**

198 **3.1. Model validation**

199 The model is validated by comparing simulated surface and bottom temperature, salinity,
200 current patterns and strengths, surface chlorophyll, surface nitrate and bottom oxygen to
201 observations. The model reproduces remotely sensed spatial and temporal SST patterns
202 (NOAA AVHRR) very well with an annual correlation coefficient of 0.98 (Figure S1).
203 Simulated surface and bottom salinity also show similar spatial and seasonal patterns as
204 available *in situ* observations (Figures S2 and S3) with correlation coefficients of 0.77 and
205 0.84, respectively. Simulated surface and bottom temperature, when compared with
206 available *in situ* data (Figures S4 and S5), are also consistent with the observations with
207 correlation coefficients of 0.96 and 0.93.

208 The simulated current systems in the ECS and YS show typical seasonal variations as
209 follows (see also Figure S6). In winter, currents mainly flow southward on the Yellow Sea
210 and ECS shelves driven by the northerly wind. In contrast, the East China Sea Coastal
211 Current and the Korean Coastal Current flow northward in summer. The Kuroshio is

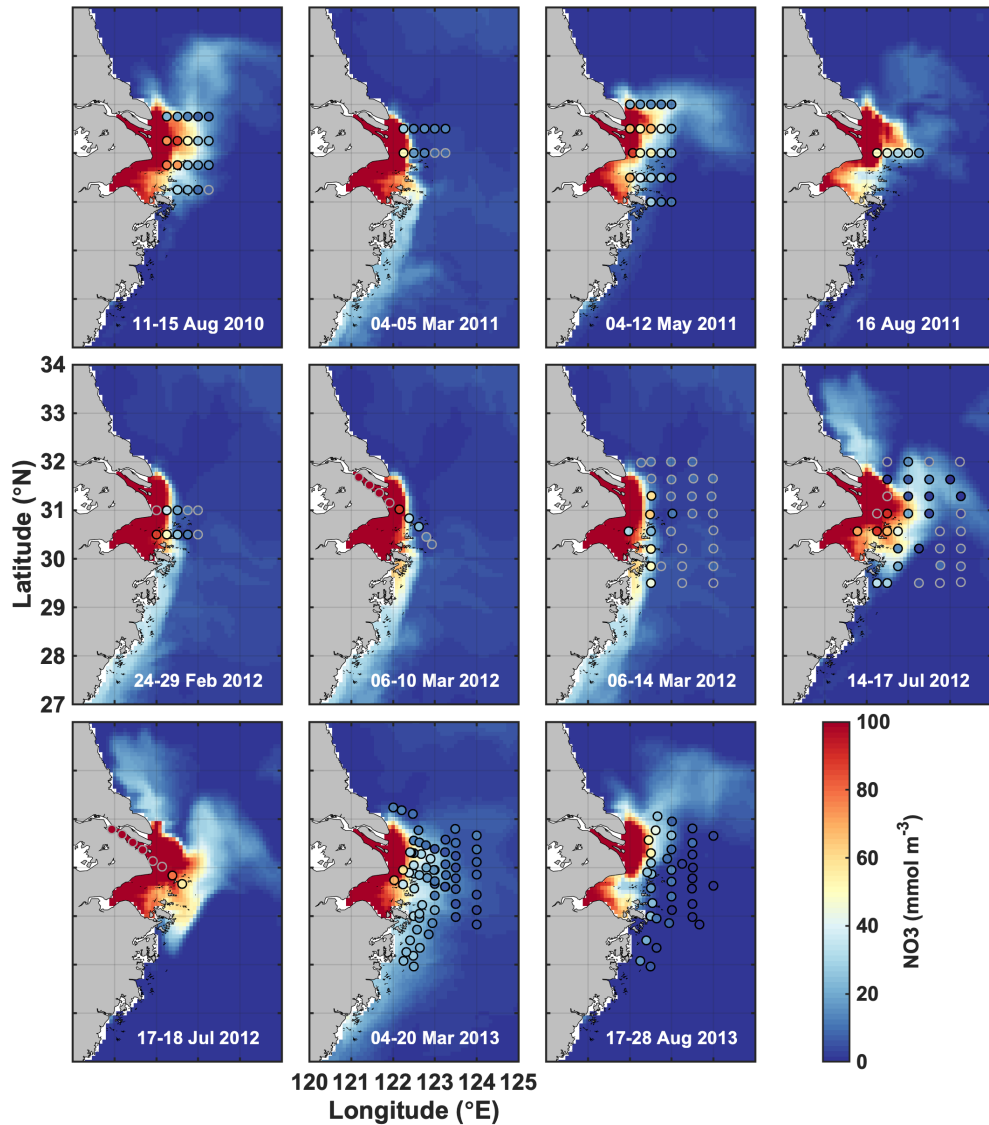


Figure 2: Simulated surface nitrate (colored map) shown for the day that marks the mid-point of the cruise dates (given in each panel) the compared to observations (dots) during 11 cruises from 2011 to 2013.

212 stronger in summer than in winter. The model captures the seasonal pattern of the current
 213 system and resolves currents in the ECS and Yellow Sea (also see Grosse et al. 2020).

214 Simulated monthly averaged (2008-2013) surface chlorophyll concentrations in May,
 215 August and November are compared with satellite-derived fields (MODIS-Terra) and
 216 agree well with correlation coefficients of 0.77, 0.94 and 0.64, respectively (Figure S7).

217 Simulated surface nitrate concentrations are shown in comparison to in situ observations
 218 in Figure 2 and agree well with a correlation coefficient of 0.84. Observations in March

219 and July of 2012 show strongly elevated concentrations in the CE and a sharp gradient in
220 the vicinity of the estuary's mouth that are well represented by the model. Likewise,
221 simulated and observed bottom oxygen distributions are compared in Figure 3 and agree
222 reasonably well overall with a correlation coefficient of 0.71 although the model
223 underestimates observed low-oxygen conditions in July of 2011 and 2013 and August
224 2013.

225 Together, these comparisons show that the model is able to reproduce important aspects
226 of the physical-biogeochemical dynamics in the study region.

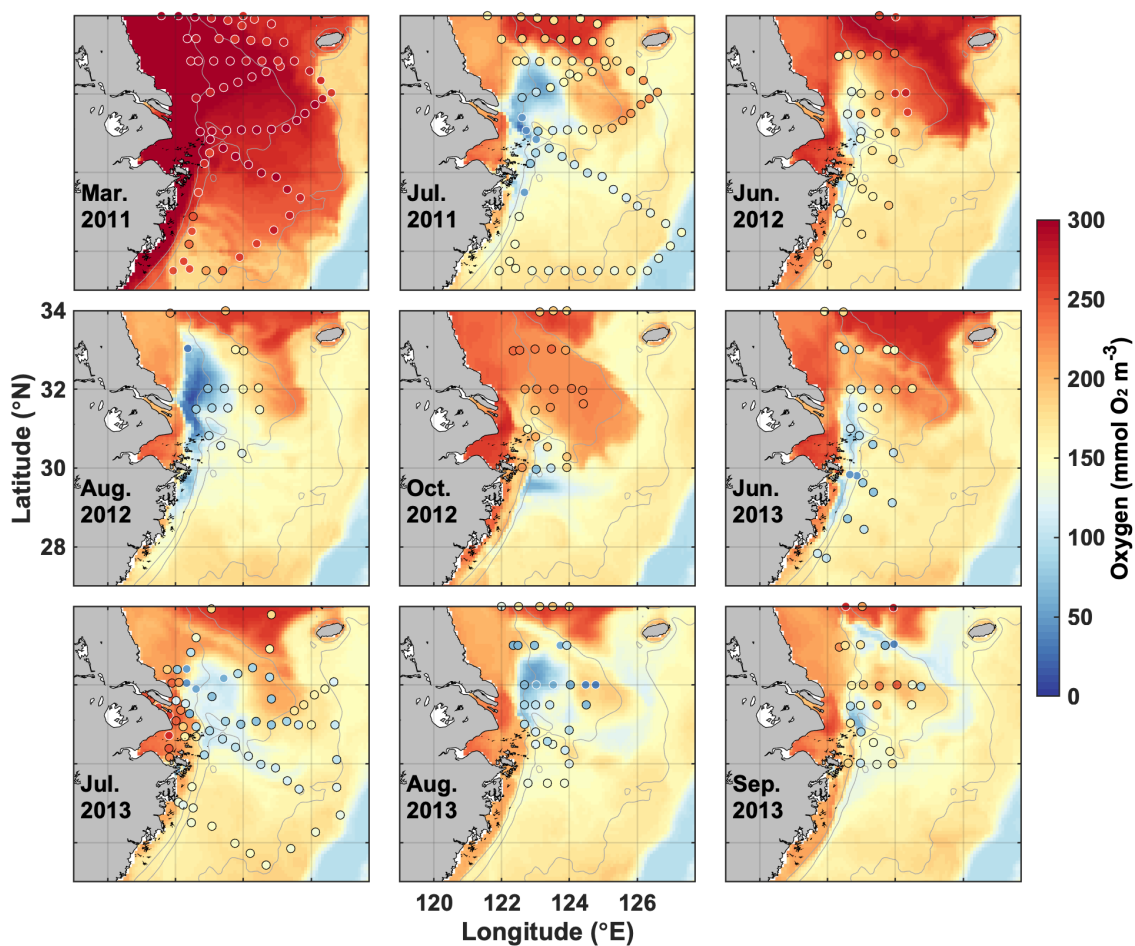


Figure 3. Simulated bottom oxygen (colored map) shown for the day that marks the mid-point of the cruise dates compared with observations (dots) during nine cruises from 2011 to 2013.

227
228
229
230

231 **3.2. Simulated oxygen dynamics**

232 The model simulates annually recurring hypoxic conditions with a typical seasonal cycle
 233 where bottom waters are well-oxygenated until April/May, hypoxic conditions establish in
 234 June or July, become more pronounced in August, and disperse in October or November
 235 (Figure 4a, c). However, the model also simulates significant interannual variability in
 236 timing and extent of hypoxia over the 6-year simulation period (Figure 4b, c). The years
 237 with largest maximum hypoxic extent are 2010 (20,520 km²), 2009 (16,660 km²), 2012
 238 (13,930 km²) and 2008 (12,720 km²) while the simulated hypoxic extent is much smaller
 239 (<5,000 km²) in 2011 and 2013. The ranking is similar when considering the time-
 240 integrated hypoxic extent (Figure 4b). The year with the largest maximum and integrated
 241 hypoxic extent (2010) also has the highest peak discharge (Figure 4a) and highest annual

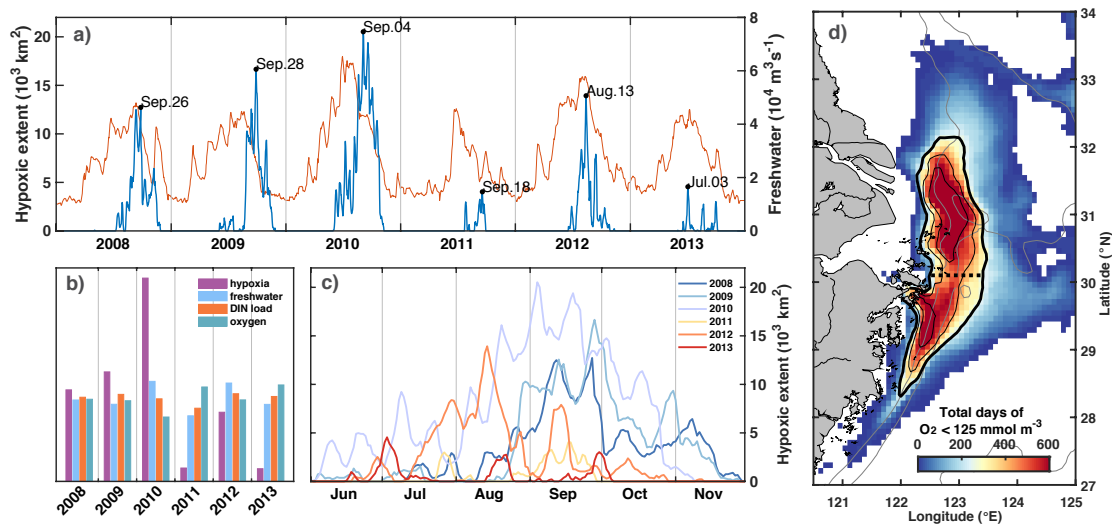


Figure 4. a) Time series of freshwater discharge and simulated hypoxic extent with peaks specified by date. b) Annual comparison of normalized time-integrated hypoxic extent, freshwater discharge, and DIN load, and summer-mean bottom oxygen concentration. c) Evolution of simulated hypoxic extent by year. d) Frequency map of number of days when bottom oxygen concentrations were below 125 mmol m⁻³ (4 mg/l). The black isolines indicate 240, 360 and 480 days (or 40, 60 and 80 days per year). The thick solid line indicates the region we refer to as the typical low-oxygen zone and the dashed line shows the demarcation between its northern and southern regions.

242 freshwater discharge ($65,400 \text{ m}^3 \text{ s}^{-1}$), although the annual discharge is similar to 2008 and
243 2012.

244 The region where low-oxygen conditions are most commonly simulated is indicated by the
245 frequency map in Figure 4d, which shows the total number of days in the 6-year simulation
246 when bottom oxygen concentrations were below 125 mmol m^{-3} (or 4 mg/l), i.e. twice the
247 hypoxic threshold. It is known from observations that there are two centers of recurring
248 hypoxic conditions: the northern core is located just to the east of the CE and Hangzhou
249 Bay and the southern core to the southeast of Hangzhou Bay. The model is consistent with
250 these observations and simulates two distinct core regions of low-oxygen conditions
251 centered at 31°N and 29.3°N . The northern core region is larger than the southern core
252 region ($9,050 \text{ km}^2$ for a threshold of 80 days per year of $< 4 \text{ mg/l}$ compared to $5,230 \text{ km}^2$).
253 We will refer to the region defined by a threshold of 40 days of $< 4 \text{ mg/l}$ of per year (solid
254 black line in Figure 1 and 4d) as the “typical low-oxygen zone” for the remainder of the
255 manuscript and demarcate the northern and southern sections by 30.1°N latitude (dashed
256 line in Figures 1 and 4d).

257 There are marked differences in the phenology of simulated hypoxic extent (Figure 4c).
258 Among the four years with largest hypoxic areas, hypoxia establishes relatively late (mid-
259 August) and lasts long (into November) in 2008 and 2009. In contrast in 2012, hypoxic
260 conditions establish earlier (June), are most pronounced in August and are eroded by mid-
261 October. In 2010, the year with the largest peak extent, hypoxia establishes already at the
262 beginning of June and is maintained until the end of October, leading to the by far largest
263 time-integrated hypoxia among the 6 years (Figure 4b). In all years there are times when
264 hypoxic extent decreases rapidly. In the following sections we explore the drivers of
265 interannual and intra-seasonal variations in low-oxygen conditions and the role of
266 biological processes and physical forcing.

267

268 *3.2.1 Interannual variations in hypoxia*

269 As mentioned above, there is significant interannual variation in hypoxic extent in the 6-
270 year simulation (Figure 4a, b, c). The years with the largest time-integrated hypoxic events
271 are 2010, 2009 and 2008 followed by 2012 with the fourth largest hypoxic extent. In 2011
272 and 2013, hypoxic conditions were much less severe than in the other 4 years. Freshwater

273 (FW) input and nutrient load are less variable with the largest FW inputs in 2010 and 2012
 274 and the lowest in 2011. In an attempt to explain the interannual variations in hypoxia, we
 275 consider first the role of riverine FW inputs and nutrient loads. More specifically, we
 276 investigate correlations of time-integrated hypoxic area, average primary production (PP),
 277 total oxygen consumption (OC) by respiration, sediment oxygen consumption (SOC) and
 278 bottom oxygen in the typical low-oxygen zone, and the spatial extent of the FW plume with
 279 annually integrated FW input and DIN load (Figure 5).

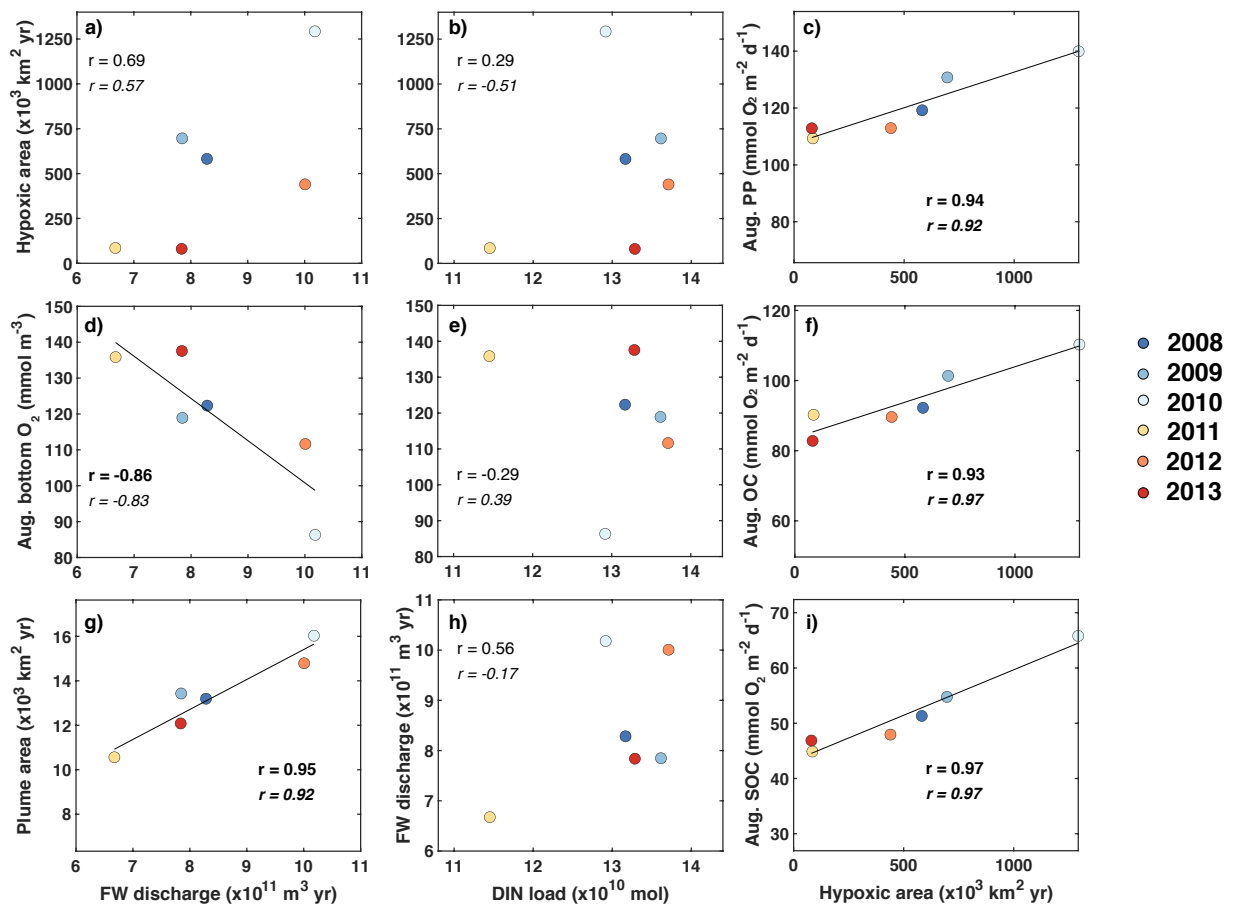


Figure 5. Correlations of time-integrated hypoxic area, average primary production, respiration and bottom oxygen in the typical low-oxygen zone in August, and the spatial extent of the FW plume (defined here as the area with surface salinity smaller than 25) with annually integrated FW input and DIN load. Correlation coefficients are given for all 6 years and, in italic font, after excluding year 2011. Significant correlations are shown in bold font and linear regressions indicated by the black line whenever the correlation is significant at $p < 0.05$.

281 There is a significant negative correlation between annual FW input and mean bottom-
282 water oxygen concentration in the low-oxygen zone of -86% and a weaker, statistically
283 insignificant positive correlation of 69% between annual FW input and integrated hypoxic
284 area (Figure 5a, d). This indicates that variations in FW input at least partly explain
285 variability in hypoxic conditions. Perhaps surprisingly, there is no convincing correlation
286 between annual FW input and annual DIN load (Figure 4h). Although the correlation
287 coefficient is 56% when all 6 years are considered, the correlation drops to -17% when the
288 low-flow year 2011 is excluded and neither of these correlations is statistically significant.
289 As expected, there is a strong positive correlation of 95% between annual FW input and
290 time-integrated plume area (Figure 4g). Plume area can thus be interpreted as a proxy of
291 FW input.

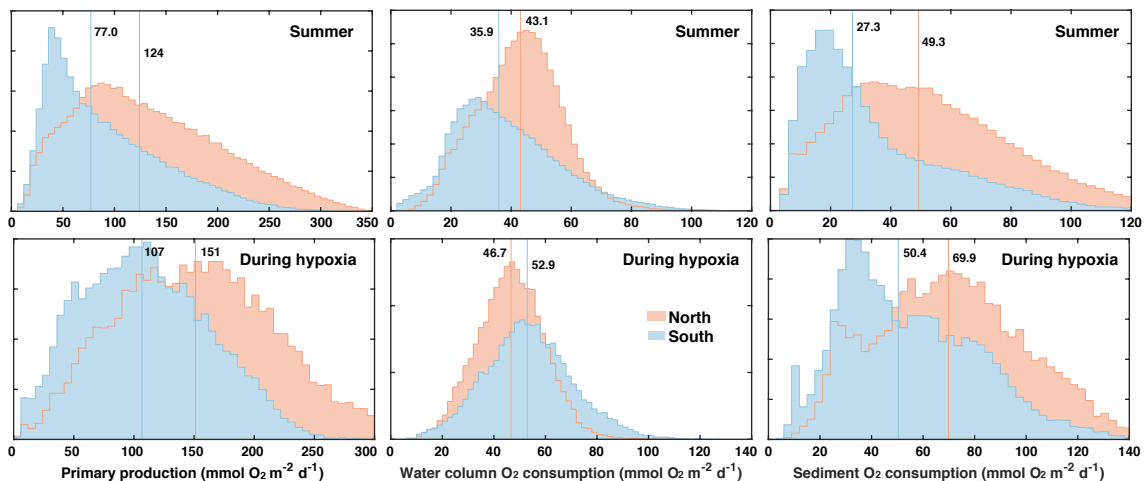
292 In contrast to the positive correlations between FW input and hypoxia as well as
293 bottom oxygen, correlations between the annual DIN load and integrated hypoxic area as
294 well as mean bottom-water oxygen are much weaker and insignificant (Figure 4b, e). This
295 implies that interannual variations in DIN load do not explain year-to-year variations in
296 hypoxia. However, the correlations between integrated hypoxic area and mean rates of PP
297 and OC (especially SOC) in August are significant and strong at 94% and 93% (97%),
298 respectively (Figure 5c, f, i). The high correlation between hypoxic area and OC is
299 primarily driven by SOC. Clearly, biological processes are important drivers of hypoxia
300 and contribute to its interannual variability, but they do not appear to result from variations
301 in DIN load. More relevant are variations in FW load, which explain interannual variations
302 in hypoxia at least partly.

303 Clearly, other factors than riverine inputs must be at play in driving interannual
304 variations. For example, comparing the years 2010 and 2012, both had very similar FW
305 input and DIN load, but differed in severity of hypoxia (Figure 5a, b). Likewise, the years
306 2009 and 2013 were very similar in terms of FW input and DIN load, but very different in
307 hypoxic extent. Next, we investigate the role of biological and physical drivers of intra-
308 seasonal and interannual variability in hypoxia.

309
310
311
312

313 3.2.2 Biological drivers of intra-seasonal variability in hypoxia

314 In order to explore whether biological rates are related to the presence of FW, and whether
315 the correlations that emerged when relating mean annual quantities also hold on shorter
316 time scales, it seems prudent to distinguish between the northern and southern regions. The
317 bathymetry in the northern zone is slightly deeper than in the southern zone (median depth
318 of 28.5 m versus 24.6 m) and several biological rates with direct relevance to oxygen
319 dynamics are different between the two zones (Figure 6). During the summer months (June
320 to September), primary production (PP), oxygen consumption in the water column
321 (WOC=OC-SOC), and SOC are larger in the northern zone with medians of 124 compared
322 to 77.0 mmol O₂ m⁻² d⁻¹ for PP, of 43.1 versus 35.9 mmol O₂ m⁻² d⁻¹ for WOC, and 49.3
323 versus 27.3 mmol O₂ m⁻² d⁻¹ for SOC. During hypoxic conditions, PP and SOC are also
324 notably larger in the northern zone with medians of 151 versus 107 mmol O₂ m⁻² d⁻¹ for PP
325 and 69.9 versus 50.4 mmol O₂ m⁻² d⁻¹ for SOC. In the water column, the difference is
326 reversed and WOC larger in the southern than the northern zone (52.9 versus 46.7 mmol
327 O₂ m⁻² d⁻¹). Because of these different characteristics, we consider the northern and
328 southern zones of the typical low-oxygen region separately.

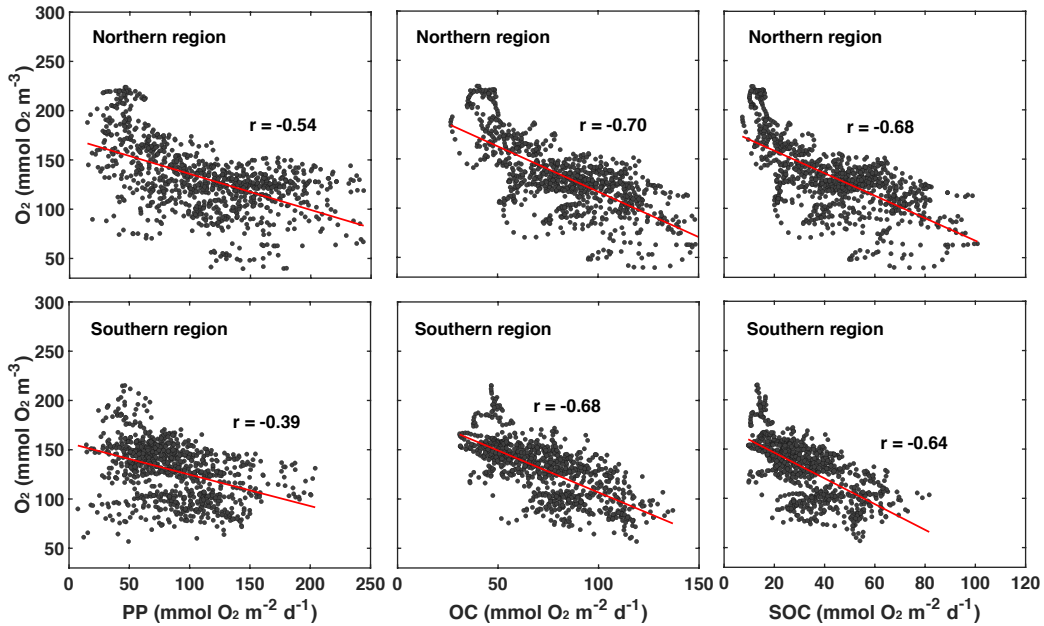


329 **Figure 6:** Histograms primary production and water-column and sediment respiration during the
330 summer months (June to September) and during hypoxic conditions in the northern and southern
331 parts of the typically hypoxic zone. Medians are indicated by vertical lines.
332

333

334 The annual correlations presented in the previous section indicate that biological rates are
335 important drivers for interannual variability but not due to variations in nutrient load.

336 Variability in annual FW input is a better predictor. In order to better understand how
 337 variability in FW is related to biological rates and thus hypoxia, we first explore whether
 338 significant relationships exist between daily biological rates, bottom-water oxygen, and the
 339 presence of FW in the two zones. Since annual FW input is highly correlated with the
 340 extent of the FW plume (see Figure 5g), daily plume extent can be used as a measure of
 341 FW presence and compared to daily rates of PP, OC, SOC, and bottom oxygen.

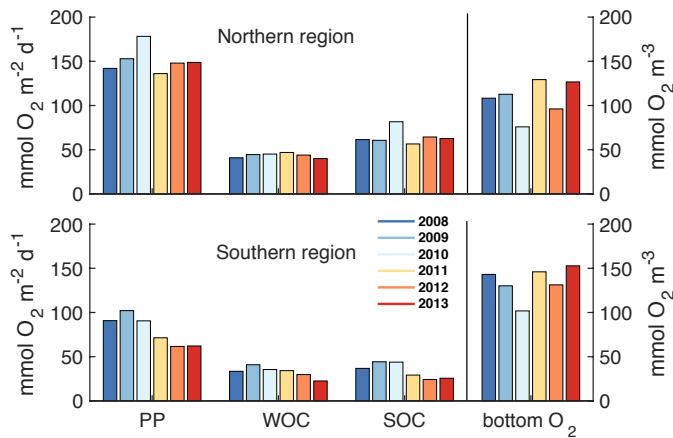


342
 343 **Figure 7.** Correlations of daily averaged rates of PP, OC and SOC plotted with daily mean
 344 bottom oxygen concentration in the northern and southern regions of the low-oxygen zone in
 345 summer. The correlations are all significant. Correlation coefficients and slope and intercept of
 346 linear regressions (indicated by red lines) are given in Table 1.
 347

Relationships between bottom oxygen (mmol m^{-3}) in northern region and											
PP ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)			OC ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)			SOC ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)					
r	a	b	r	a	b	r	a	b			
-0.54	-0.36	172	-0.70	-0.92	209	-0.68	-1.14	181			
Same for the southern region											
-0.39	-0.32	157	-0.68	-0.85	192	-0.64	-1.30	172			
Relationships between plume area (10^3 km^2 ; defined by surface salinity < 29) in northern region and											
PP ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)			OC ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)			SOC ($\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$)			Bottom oxygen (mmol m^{-3})		
0.62	6.04	47.6	0.49	2.48	57.7	0.51	2.05	22.0	-0.56	-3.74	171
Same for the southern region											
0.43	3.78	64.6	0.56	3.18	57.8	0.43	1.50	24.7	-0.49	-3.52	149

348 **Table 1.** Correlation coefficients and parameters of a linear model fit (of the form $y=ax+b$) between
 349

350 Daily PP, OC, and SOC are all significantly and negatively correlated with bottom-water
 351 oxygen (Figure 7, Table 1). This confirms that local production of organic matter and the
 352 resulting biological oxygen consumption are important for hypoxia development.
 353 However, it is also obvious that variability around the best fit is large (Figure 7).
 354 Furthermore, bottom oxygen and biological rates are significantly correlated with the
 355 extent of the FW plume (Table 1). This suggests that variability in the presence of FW
 356 contributes to variability in hypoxia not only by increasing vertical stratification and thus
 357 inhibiting vertical supply of oxygen to the subsurface but also because PP and respiration
 358 is larger in the river plume. Likely, large FW plumes stimulate more widespread biological
 359 production and thus oxygen consumption.



360
 361 **Figure 8.** Mean August rates of PP, WOC, and SOC and mean bottom oxygen concentration in the
 362 northern and southern regions for 6-year simulation.

363
 364 However, variations in biological rates alone do not explain variability in bottom
 365 oxygen concentrations and hypoxia. In Figure 8, we show August mean rates of PP, WOC,
 366 and SOC as well as mean bottom oxygen in the northern and southern zones for all years.
 367 In the northern zone, WOC is remarkably similar in all years. PP and SOC are also similar
 368 except in 2010 when SOC is higher. The low bottom oxygen concentration in 2010 could
 369 be explained by the relatively higher SOC; however, 2012 also had relatively low bottom
 370 oxygen while biological rates were similar to the other years with higher bottom oxygen.
 371 Likewise in the southern zone, differences in PP, WOC, and SOC among the years do not
 372 explain differences in bottom oxygen, as the years with the lowest bottom oxygen (2010

373 and 2012) are not the years with the highest PP and oxygen consumption rates. Next, we
374 analyse the role of wind forcing, direction and stratification.

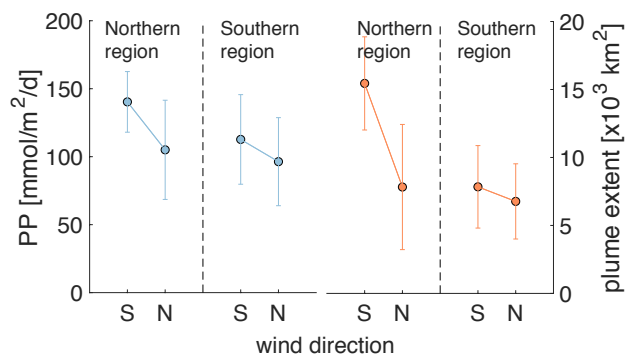
375

376 3.2.3 Physical drivers of intra-seasonal variability in hypoxia

377 We focus our analysis of physical drivers on wind direction and wind strength and
378 their relation to FW plume location and extent because the latter has already been identified
379 as an explanatory variable for interannual variations.

380 Wind direction is relevant because for most of June, July, and August winds blow
381 predominantly from the south, but switch to predominantly northerly winds between the
382 2nd half of August and the end of September. As a result of the northward, upwelling
383 favorable winds in the early summer, the FW plume is spread offshore and overlaps
384 primarily with the northern zone. After the switch to mostly southward, downwelling-
385 favorable directions, the FW plume moves southward, becomes more contained near the
386 coast, and grows in its southward extent as it is transported by a coastal current. Wind
387 direction has a demonstrable impact on PP and the extent of the FW plume as shown in
388 Figure 9 for the month of September. Especially in the region, PP and plume extent are
389 notably larger during southerly winds when the FW plume is more spread out, than during
390 northerly winds when the plume is restricted to the coastal current.

391



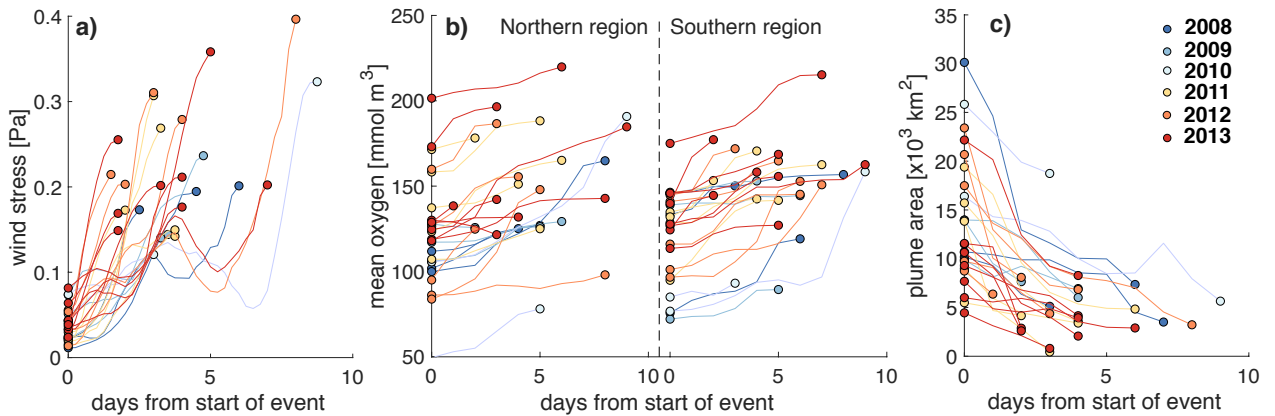
392

393 **Figure 9.** Mean PP and FW plume extent in the northern and southern regions averaged over days
394 with north and south wind (i.e. when direction is +/- 45° of true north or south) and wind strength
395 >0.03 Pa for in September.

396

397 Wind strength is relevant because storm events can erode vertical stratification and
398 thus lead to resupply of oxygen to bottom waters due to vertical mixing. We investigated
399 the effect of wind strength on bottom oxygen, hypoxia, and the extent of the FW plume by
400 first inspecting time series of these variables (Figure S8). We isolated all event during the
401 months June to September and, in Figure 10, show the corresponding changes in wind
402 stress, mean bottom oxygen in the northern and southern zones, and the extent of the FW
403 plume. We diagnosed these events as follows. First, we identified all days when the wind
404 stress exceeded 0.12 Pa. Then we detected the minima in wind stress adjacent to the high-
405 wind days by searching for minima in wind stress within 3 days prior and 3 days after the
406 high-wind days. The periods within these minima are used as analysis period for each wind
407 event. In four instances the wind stress exceeded the threshold within 5 days of a previous
408 wind event. Those subsequent high-wind events were combined into one. We identified
409 the minimum in bottom oxygen (maximum in FW plume area) at the beginning of the event
410 and the maximum in oxygen (minimum in FW area) after the maximum in wind stress was
411 reached.

412 Figure 10a illustrates rapid increases in wind stress typically within 2 to 4 days. The
413 only exceptions are the 4 events where two storms occurred in rapid succession and the
414 combined event lasted longer (up to 8 days) until maximum wind stress was reached. The
415 year with the most wind events is 2013 (with 8 in total including one of the combined long-
416 lasting event). The year with the least events is 2010 (2 events) followed by 2009 (3 events).
417 Most of these events resulted in notable increases in mean bottom oxygen, typically by 10
418 to 30 mmol m⁻³, but up to 100 mmol m⁻³ in 2010 in the southern zone (Figure 10b). In the
419 rare cases where bottom oxygen did not increase or slightly decreased, bottom oxygen was
420 already elevated before the wind event. The wind events strongly affected the extent of the
421 FW plume (Figure 10c) by mixing the freshwater layer with underlying ocean water. The
422 effects are largest when the FW plume was most expansive. This analysis shows the
423 significant role of storm events in disrupting the generation of low-oxygen conditions and
424 ventilating bottom waters.

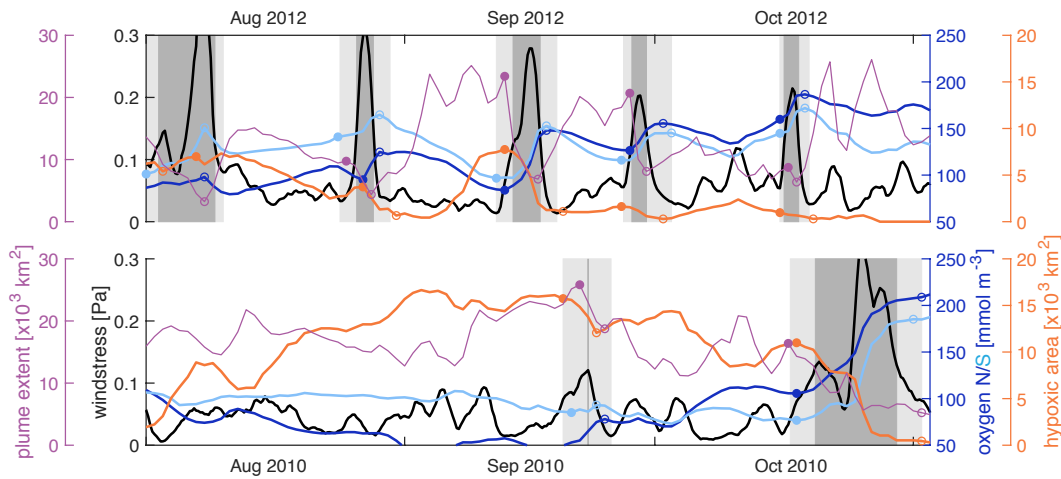


425

426 **Figure 10.** Evolution of a) wind stress, b) bottom mean oxygen in the northern and southern
 427 regions, and c) extent of the FW plume during high-wind events. These events are defined by wind
 428 stress exceeding 0.12 Pa.

429

430 In section 3.2.1 above, we noted that while the years 2010 and 2012 had very similar
 431 FW input and DIN load, 2010 had a much larger hypoxic area. Likewise, the years 2009
 432 and 2013 were very similar in terms of FW input and DIN load, but 2009 had a much larger
 433 hypoxic area. Considering the frequency and severity of high-wind events explains the
 434 differences in both cases.



435

436 **Figure 11.** Wind stress (black), mean bottom oxygen in the northern and southern zones (dark and
 437 light blue), total hypoxic extent (orange), and FW plume extent (purple) throughout August,
 438 September and October of 2010 and 2012. The filled and open circles indicate a variables' value at

439 the beginning and after high-wind events. High-wind days/events are indicated by the dark/light
440 gray shading.

441 Figure 11 shows the wind stress, mean bottom oxygen in the northern and southern
442 zones, and total hypoxic extent and FW plume extent in 2012 and 2010. In 2012, there
443 were 5 high-wind events during the months of August, September, and October that all
444 coincided with increases in bottom oxygen, decreases in hypoxic extent when a hypoxic
445 zone was established at the beginning of the event, and decreases in FW plume extent.
446 Inspection of the evolution of bottom oxygen is especially instructive. While bottom
447 oxygen concentrations declined during periods with average or low wind, they were
448 essentially reset at a much higher level during each wind event. Whenever the FW plume
449 was extensive at the beginning of a high-wind event, it was drastically reduced during the
450 event. In 2010 bottom oxygen was at similar levels to 2012 at the beginning of August but
451 dropped to low levels throughout August, especially in the northern zone, and remained
452 low with widespread hypoxia until a major wind event in the second half of October
453 ventilated bottom waters. Except for a very short event in the second half of September,
454 there were no high-wind events from August until mid-October in 2010.

455 The differences in hypoxia in 2009 and 2013 can also be explained by the frequency
456 and intensity of high-wind events. In 2013, there were 8 high-wind events from July to
457 October that led to an almost continuous ventilation of bottom waters while in 2009 there
458 were only 3 such events during the same period (Figure S8). Low to average winds from
459 mid-August to early October of 2009 coincided with a decline in bottom oxygen and
460 establishment of an expansive hypoxic zone throughout most of September.

461 These analyses show that wind direction and strength play an important role in
462 determining the location of the hypoxic zone (i.e. northern versus southern region) and the
463 extent and severity of hypoxic conditions.

464

465 ***3.3 Oxygen budgets for the northern and southern regions***

466 In order to further investigate the roles of physical and biological processes in regulating
467 hypoxia, oxygen budgets are calculated from daily model output for the period from March
468 to August for the northern and southern hypoxic regions. Considering that hypoxic
469 conditions occur near the bottom, we evaluate an oxygen budget not only for the whole

470 water column but also for its lower portion which typically becomes hypoxic. To account
471 for variations in the thickness of the hypoxic layer, which tends to be thicker in deeper
472 waters (similar to observations by Ning et al., 2011), we include the bottommost 12 layers
473 of our model grid. Because of the model's terrain-following vertical coordinates, the
474 thickness of these 12 model layers varies with total depth as shown. The terms considered
475 in the budget are air-sea flux, lateral physical advection and diffusion, vertical turbulent
476 diffusion (for the subsurface budget only), PP, WOC (including respiration and
477 nitrification), and SOC. Each term is integrated vertically over the whole water column and
478 also over the bottom-most 12 layers and then averaged for the northern and southern
479 regions for each month (Figure 12, Table S2).

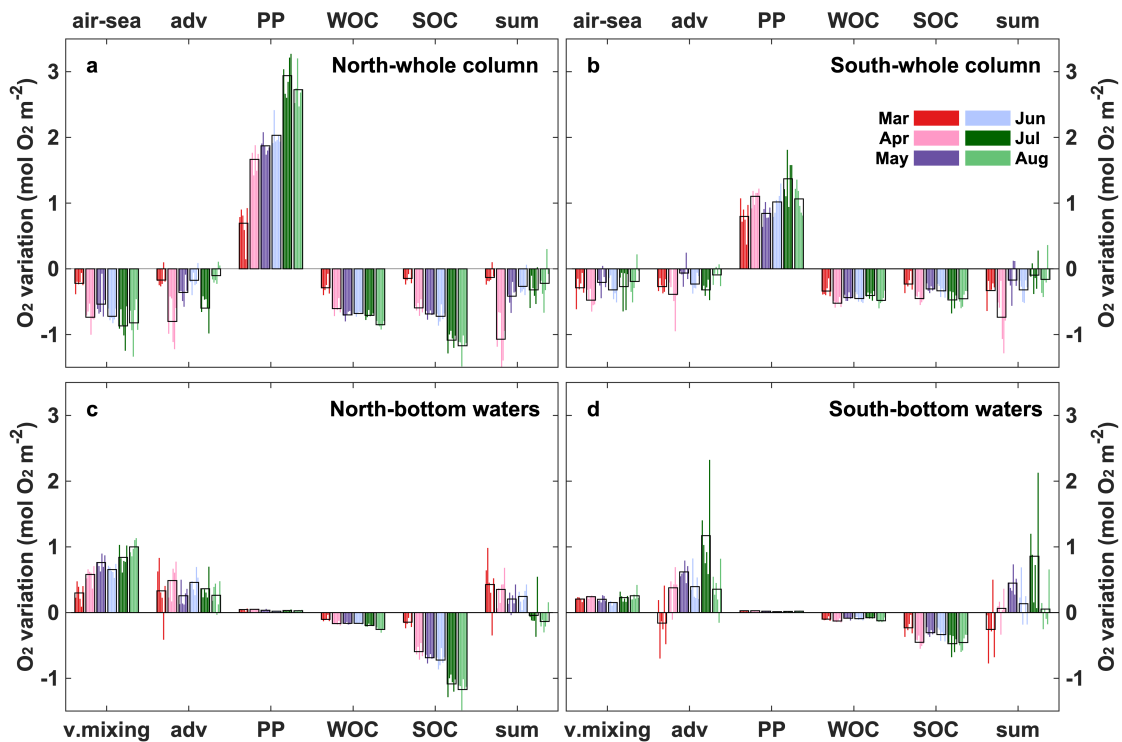


Figure 12. Monthly averaged (2008-2013) oxygen budgets for the whole water column and subsurface water from March to August in the northern and southern hypoxic regions. Adv represents lateral advection and lateral diffusion which is comparatively small, while v.mixing represents vertical turbulent diffusion, which is only relevant for the subsurface budget. Thin color bars represent individual years whereas the black bars are the 6-year average.

480 For the whole water column (Figure 12a, b), biological processes (PP, WOC, and SOC)
 481 greatly exceed physical processes (air-sea exchange and advective transport) in affecting
 482 oxygen. PP is always greater than the sum of WOC and SOC in the whole column
 483 indicating autotrophy in spring and summer. Advection is negative, acting as an oxygen
 484 sink and offsetting 21% of PP on average in the northern and southern regions. Of the two
 485 biological oxygen consumption terms (WOC and SOC), WOC accounts for half of total
 486 respiration. Negative air-sea flux indicates oxygen outgassing into the atmosphere and is
 487 due to photosynthetic oxygen production and decreasing oxygen solubility. However, since
 488 hypoxia only occurs in the subsurface, the subsurface budget below is more instructive.

489 When considering only subsurface waters (Figure 12c, d), the influence of PP decreases
 490 markedly, accounting for less than 2% of that in the whole water column. Vertical turbulent

491 diffusion acts as the largest oxygen source in the subsurface layer. SOC is the dominant
492 oxygen sink accounting for 80% of the total biological oxygen consumption. As
493 photosynthetic oxygen production increases gradually from spring to summer (Figure 12a,
494 b) WOC and SOC also increase as they are closely associated with photosynthetically
495 produced organic matter. Vertical oxygen diffusion tends to covary with PP, implying an
496 oxygen gradient driven by photosynthetic oxygen production in the upper layer. Lateral
497 advection of oxygen is negative in March only (early in the hypoxic season) mainly in the
498 southern region, but becomes positive later. This suggests that early in the hypoxic season,
499 import of low-oxygen water contributes to hypoxia generation but advection switches to
500 an oxygen source later. Overall, oxygen sources and sink terms are similar in the northern
501 and southern regions.

502

503 **4. Discussion**

504 We implemented and validated a state-of-the-art physical-biological model for the ECS.
505 The implementation is based on a model that was previously developed and extensively
506 used for the northern Gulf of Mexico (Fennel et al. 2011, Laurent et al. 2012, Yu et
507 al.2015b), a region that is similar to the ECS in that it receives large inputs of FW and
508 nutrients from a major river and develops extensive, annually recurring hypoxia (see Table
509 1 in Fennel and Testa (2019). Our model is more comprehensive than previous models for
510 the ECS.

511 A 6-year simulation was performed and validated. The model faithfully represents
512 patterns and variability in surface and bottom temperature and salinity, surface chlorophyll
513 and nitrate distributions, bottom oxygen, and correctly simulates the major current patterns
514 in the region (see Section 3.1 and Supplement). We thus deem the model's skill as
515 sufficient for the analysis of biological and physical drivers of hypoxia generation
516 presented here.

517 The model simulates annually recurring hypoxic conditions but with significant
518 interannual and intra-seasonal variability and marked differences in phenology of hypoxic
519 conditions from year to year (Figure 4a, b, c). Interannual variability in hypoxic conditions
520 is much larger than variations in FW input, nutrient load, and bottom oxygen
521 concentrations (Figure 4b) because small variations in oxygen can lead to large changes in

522 hypoxic area when bottom oxygen is near the hypoxic threshold. Interannual variability in
523 hypoxic area is partly explained by variations in annual FW input, consistent with previous
524 studies (Zheng et al., 2016; Zhou et al., 2017). While the correlation between time-
525 integrated hypoxic area and FW input is insignificant, there is a strong and significant
526 negative correlation between mean bottom oxygen in August and annual FW input (Figure
527 5). Annual FW input is also correlated strongly and significantly with the annually
528 integrated spatial extent of the FW plume, which is a useful metric for extent of the region
529 directly influenced by riverine inputs which induce strong density stratification and high
530 productivity.

531 Surprisingly, DIN load is not correlated with FW input, hypoxic area, and mean bottom
532 oxygen in August (Figure 5). This is in contrast to the northern Gulf of Mexico where DIN
533 load is highly correlated with both FW input and nutrient load and frequently used as a
534 predictor of hypoxic extent (Scavia et al. 2017, Laurent and Fennel 2019). However, the
535 lack of correlation between hypoxia and DIN load in the ECS should not be interpreted as
536 biological processes being unimportant in hypoxia generation, just that variations in DIN
537 load do not explain year-to-year differences. In fact, hypoxic area and biological rates (i.e.
538 mean August PP, OC, and SOC) are strongly and significantly correlated (Figure 5),
539 emphasizing the dominant role of biological oxygen consumption. The fact that riverine
540 variations in DIN load do not seem to have an effect suggests that riverine nutrient inputs
541 are large enough to saturate the region with nutrients, similar to the northern Gulf of
542 Mexico where small reductions in nutrient load have a relatively small effect (Fennel and
543 Laurent 2018).

544 Variations in riverine FW input only partly explain interannual variations in hypoxia.
545 For example, the years 2010 and 2012 had similar FW inputs and DIN loads but the hypoxic
546 area was 4 times larger in 2010 than 2012 (Figure 5a). Similarly, 2009 and 2013 had the
547 same FW inputs and nutrient loads but 2009 experienced extensive hypoxia while there
548 was almost none in 2013. In order to elucidate these differences, we investigated biological
549 and physical drivers of intra-seasonal variability.

550 In the ECS, two distinct zones of low oxygen have been observed (Li et al., 2002; Wei
551 et al., 2007; Zhu et al., 2016, 2011). The model simulates these two zones, referred to as
552 the northern and southern zone, consistent with observations (Figure 4d) and with generally

553 higher PP and SOC in the northern zone (Figure 6). Because of these differences we treated
554 the two zones separately in our analysis of intra-seasonal drivers.

555 We found daily biological rates (i.e. PP, OC, SOC) to be significantly correlated with
556 bottom oxygen in both zones, but with relatively large variability around the best linear fit
557 (Figure 7). The biological rates and bottom oxygen are also significantly correlated with
558 the extent of the FW plume (Table 1). Again, these results emphasize the dominant role of
559 biological oxygen consumption, and its relation to riverine inputs, in hypoxia generation
560 but leave a significant fraction of the variability unexplained.

561 We conducted an analysis of the effects of wind direction and strength on hypoxia. Wind
562 direction has a notable effect on the geographic distribution of hypoxia. Southerly,
563 upwelling-favorable winds lead to a more widespread eastward extension of the FW plume
564 with elevated PP and vertical density stratification (Figure 9). Northerly, downwelling-
565 favorable winds create a coastally trapped southward jet that moves FW southward and
566 constrains the plume close to the coast. A similar behavior has been described for the
567 northern Gulf of Mexico (Feng et al. 2014).

568 Wind strength turned out to be an important factor in hypoxia evolution. We identified
569 high-wind events and showed that whenever bottom oxygen is low, a high-wind event will
570 lead to a partial reoxygenation of bottom waters and decrease hypoxic extent (Figure 10).
571 The impact of high-wind events is also visible in the extent of the FW plume, which is
572 drastically reduced during high winds because FW is mixed. The frequency of high-wind
573 events during summer explains the differences in hypoxic area between 2010 and 2012
574 (Figure 11) and 2009 and 2013 (Figure S8). In 2009 and 2010 there were only few high-
575 wind events during summer while 2012 and 2013 experienced a sequence of storms that
576 led to partial reoxygenation of the water column throughout the summer and thus impeded
577 the development hypoxia.

578 We calculated oxygen budgets for the northern and southern regions considering the
579 whole water-column and the near-bottom layer only. The subsurface budget is particularly
580 useful in providing insights into when and where lateral advection amplifies or mitigates
581 hypoxia and illustrates that SOC is the dominant oxygen sink in the subsurface. The relative
582 importance of WOC and SOC had not previously been quantified for this region due to
583 lack of concurrent WOC and SOC observations and lack of models that realistically

584 account for both processes. The budget for the whole water column is less useful because
585 it is dominated by the oxygen sources, sinks and transport in the surface layer, which does
586 not experience hypoxia and thus is not relevant.

587 The importance of SOC in our model is consistent with recent observational studies in
588 the ECS. SOC on the coastal shelves in the Yellow Sea and ECS has been estimated to
589 range from 1.7 to 17.6 mmol O₂ m⁻² d⁻¹ (mean rate of 7.2 mmol O₂ m⁻² d⁻¹) from April to
590 October except August by Song et al. (2016), and from 9.1 to 62.5 mmol O₂ m⁻² d⁻¹ (mean
591 of 22.6 ± 16.4 mmol O₂ m⁻² d⁻¹) from June to October in Zhang et al. (2017). Simulated
592 SOC in the typical low-oxygen zone falls within the range observed by Zhang et al. (2017)
593 with a mean rate of 20.6 ± 19.2 mmol O₂ m⁻² d⁻¹ between April and October. Based on
594 observations, Zhang et al. (2017) already suggested that SOC is a major contributor to
595 hypoxia formation in below-pycnocline waters, which is further corroborated by our model
596 results. It is also consistent with the modelling study of Zhou et al. (2017), who did not
597 include SOC in the baseline version of their model but showed in a sensitivity study that
598 inclusion of SOC simulates hypoxic extent more realistically. Our results are in line with
599 findings from the northern Gulf of Mexico hypoxic zone where WOC is much larger than
600 SOC below the pycnocline, while SOC is dominant in the bottom 5 m where hypoxia
601 occurs most frequently in summer (Quiñones-Rivera et al., 2007; Yu et al., 2015b).

602 The finding that lateral oxygen transport can act as a net source to subsurface water is
603 also new. On seasonal scales, oxygen advection in the subsurface varies from an oxygen
604 sink in spring to a source in summer, especially in the southern hypoxic region, implying
605 that the TWC becomes an oxygen source when oxygen is depleted in the hypoxic region.
606 This aspect was neglected in previous studies which only emphasized the role of advection
607 as an oxygen sink promoting hypoxia formation (Ning et al., 2011; Qian et al., 2015). The
608 TWC originates from the subsurface of the Kuroshio northeast to Taiwan Island, and thus
609 represents an intrusion onto the continental shelves from the open ocean (Guo et al., 2006).
610 In addition to oxygen advection, nutrients are transported supporting primary production
611 on the ECS shelves (Zhao & Guo, 2011, Grosse et al. 2020). The intrusion of the TWC and
612 the Kuroshio accompanied by relatively cold and saline water, and nutrient and oxygen
613 transport, is thought to influence hypoxia development (Li et al., 2002; Wang, 2009; Zhou

614 et al., 2017) but no quantification of the relative importance has occurred until now (see
615 companion paper by Grosse et al., 2020, using the same model).

616

617 **5. Conclusions**

618 In this study, a new 3D coupled physical-biological model for the ECS was presented
619 and used to explore the spatial and temporal evolution of hypoxia off the CE and to quantify
620 the major processes controlling interannual and intra-seasonal oxygen dynamics.
621 Validation shows that the model reproduces the observed spatial distribution and temporal
622 evolution of physical and biological variables well.

623 A 6-year simulation with realistic forcing produced large interannual and intra-seasonal
624 variability in hypoxic extent despite relatively modest variations in FW input and nutrient
625 loads. The interannual variations are partly explained by variations in FW input but not
626 DIN load. Nevertheless, elevated rates of biological oxygen consumption are of paramount
627 importance for hypoxia generation in this region, as shown by the high correlation between
628 hypoxic area, bottom oxygen, and biological rates (PP, OC, SOC) on both annual and
629 shorter time scales.

630 Other important explanatory variables of variability in hypoxia are wind direction and
631 strength. Wind direction affects the magnitude of PP and the spatial extent of the FW plume,
632 because southerly, upwelling favorable winds tend to spread the plume over a large area
633 while northerly, downwelling-favorable winds push the plume against the coast and induce
634 a coastal current that contains the FW and moves it downcoast. Wind strength is important
635 because high-wind events lead to a partial reoxygenation whenever bottom oxygen is low
636 and can dramatically decrease the extent of the FW plume. The frequency of high-wind
637 events explains some of the interannual differences in hypoxia, where years with similar
638 FW input, nutrient load, and mean rates of oxygen consumption have display very different
639 hypoxic extents because high-wind events lead to partial reoxygenation of bottom waters.

640 A model-derived oxygen budget shows that SOC is larger than WOC in the subsurface
641 of the hypoxic region. Lateral advection of oxygen in the subsurface switches from an
642 oxygen sink in spring to a source in summer especially in the southern region and is likely
643 associated with open-ocean intrusions onto the coastal shelf supplied by the Taiwan Warm
644 Current.

645 **Acknowledgments:** HZ was supported by the National Key Research and Development
646 Program of China (2016YFC1401602 and 2017YFC1404403) and the China Scholarship
647 Council (CSC). The authors thank the crew of the Dongfanghong2 for providing much help
648 during the sampling cruises, and Compute Canada for access to supercomputer time. KF
649 acknowledges support from the NSERC Discovery Program.

650 **Code/Data Availability:** The ROMS model code is available at <http://myroms.org>.
651 NOAA AVHRR and MODIS-Terra are available at
652 <https://www.nodc.noaa.gov/SatelliteData/ghrsst/> and <http://oceancolor.gsfc.nasa.gov/>.
653 The model results are available on request to the authors.

654 **Author Contributions:** The manuscript is based on HZ's PhD thesis (in Chinese). CB
655 implemented the physical model. HZ added the biological component, performed model
656 simulations, and wrote the first version of the manuscript with input from KF and AL. For
657 the manuscript revision, AL reran the model simulation, AL and KF performed additional
658 analyses, and KF revised the text with input from all co-authors.

659 **Competing Interests:** The authors declare they have no competing interests.

660

661 **References**

- 662 Baird, D., Christian, R. R., Peterson, C. H., & Johnson, G. A.: Consequences of hypoxia on
663 estuarine ecosystem function: Energy diversion from consumers to microbes. *Ecological*
664 *Applications*, 14(3), 805–822. <https://doi.org/10.1890/02-5094>, 2004.
- 665 Bian, C., Jiang, W., & Greatbatch, R. J.: An exploratory model study of sediment transport
666 sources and deposits in the Bohai Sea, Yellow Sea, and East China Sea. *Journal of Geophysical*
667 *Research: Oceans*, 118(11), 5908–5923. <https://doi.org/10.1002/2013JC009116>, 2013a.
- 668 Bian, C., Jiang, W., Quan, Q., Wang, T., Greatbatch, R. J., & Li, W.: Distributions of suspended
669 sediment concentration in the Yellow Sea and the East China Sea based on field surveys during
670 the four seasons of 2011. *Journal of Marine Systems*, 121–122, 24–35,
671 <https://doi.org/10.1016/j.jmarsys.2013.03.013>, 2013b.
- 672 Bianchi, T. S., DiMarco, S. F., Cowan, J. H., Hetland, R. D., Chapman, P., Day, J. W., & Allison,
673 M. A.: The science of hypoxia in the northern Gulf of Mexico: A review. *Science of the Total*
674 *Environment*, 408(7), 1471–1484. <https://doi.org/10.1016/j.scitotenv.2009.11.047>, 2010.

675 Bishop, M. J., Powers, S. P., Porter, H. J., & Peterson, C. H.: Benthic biological effects of
676 seasonal hypoxia in a eutrophic estuary predate rapid coastal development. *Estuarine, Coastal
677 and Shelf Science*, 70(3), 415–422. <https://doi.org/10.1016/j.ecss.2006.06.031>, 2006.

678 Capet, A., Beckers, J. M., & Grégoire, M.: Drivers, mechanisms and long-term variability of
679 seasonal hypoxia on the Black Sea northwestern shelf - Is there any recovery after
680 eutrophication? *Biogeosciences*, 10(6), 3943–3962. <https://doi.org/10.5194/bg-10-3943-2013>,
681 2013.

682 Carton, J. A., & Giese, B. S.: A Reanalysis of Ocean Climate Using Simple Ocean Data
683 Assimilation (SODA). *Monthly Weather Review*, 136(8), 2999–3017,
684 <https://doi.org/10.1175/2007MWR1978.1>, 2008.

685 Chen, C. C., Gong, G. C., & Shiah, F. K., Hypoxia in the East China Sea: One of the largest
686 coastal low-oxygen areas in the world. *Marine Environmental Research*, 64(4), 399–408.
687 <https://doi.org/10.1016/j.marenvres.2007.01.007>, 2007.

688 Chen, J., Cui, T., Ishizaka, J., & Lin, C.: A neural network model for remote sensing of diffuse
689 attenuation coefficient in global oceanic and coastal waters: Exemplifying the applicability of
690 the model to the coastal regions in Eastern China Seas. *Remote Sensing of Environment*, 148,
691 168–177. <https://doi.org/10.1016/j.rse.2014.02.019>, 2014.

692 Chen, X., Shen, Z., Li, Y., & Yang, Y.: Physical controls of hypoxia in waters adjacent to the
693 Yangtze Estuary: A numerical modeling study. *Marine Pollution Bulletin*, 97(1–2), 349–364.
694 <https://doi.org/10.1016/j.marpolbul.2015.05.067>, 2015a.

695 Chen, X., Shen, Z., Li, Y., & Yang, Y.: Tidal modulation of the hypoxia adjacent to the Yangtze
696 Estuary in summer. *Marine Pollution Bulletin*, 100(1), 453–463,
697 <https://doi.org/10.1016/j.marpolbul.2015.08.005>, 2015b.

698 Dee, D. P., Uppala, S. M., Simmons, A. J., Berrisford, P., Poli, P., Kobayashi, S., ... Vitart, F.:
699 The ERA-Interim reanalysis: Configuration and performance of the data assimilation system.
700 *Quarterly Journal of the Royal Meteorological Society*, 137(656), 553–597.
701 <https://doi.org/10.1002/qj.828>, 2011.

702 Diaz, R. J., & Rosenberg, R.: Spreading dead zones and consequences for marine ecosystems.
703 *Science*, 321(5891), 926–929. <https://doi.org/10.1126/science.1156401>, 2008.

704 Egbert, G. D., & Erofeeva, S. Y.: Efficient inverse modeling of barotropic ocean tides. *Journal of*
705 *Atmospheric and Oceanic Technology*, 19(2), 183–204. [https://doi.org/10.1175/1520-0426\(2002\)019<0183:EIMOBO>2.0.CO;2](https://doi.org/10.1175/1520-0426(2002)019<0183:EIMOBO>2.0.CO;2), 2002.

707 Feng, Y., Fennel, K., Jackson, G.A., DiMarco, S.F. & Hetland, R.D.: A model study of the
708 response of hypoxia to upwelling favorable wind on the northern Gulf of Mexico shelf, *Journal*
709 *of Marine Systems* 131, 63-73, 2014.

710 Fennel, K., and Testa, J.M.: Biogeochemical controls on coastal hypoxia, *Annual Review of*
711 *Marine Science*, 11, 105-130, <https://doi.org/10.1146/annurev-marine-010318-095138>, 2019.

712 Fennel, K. and Laurent, A.: N and P as ultimate and proximate limiting nutrients in the northern
713 Gulf of Mexico: implications for hypoxia reduction strategies, *Biogeosciences*, 15, 3121-3131,
714 <https://doi.org/10.5194/bg-15-3121-2018>, 2018.

715 Fennel, K., Hetland, R., Feng, Y., & DiMarco, S.: A coupled physical-biological model of the
716 Northern Gulf of Mexico shelf: Model description, validation and analysis of phytoplankton
717 variability. *Biogeosciences*, 8(7), 1881–1899. <https://doi.org/10.5194/bg-8-1881-2011>, 2011.

718 Fennel, K., Hu, J., Laurent, A., Marta-Almeida, M., & Hetland, R.: Sensitivity of hypoxia
719 predictions for the northern Gulf of Mexico to sediment oxygen consumption and model
720 nesting. *Journal of Geophysical Research: Oceans*, 118(2), 990–1002.
721 <https://doi.org/10.1002/jgrc.20077>, 2013.

722 Fennel, K., Wilkin, J., Levin, J., Moisan, J., O'Reilly, J., & Haidvogel, D.: Nitrogen cycling in
723 the Middle Atlantic Bight: Results from a three-dimensional model and implications for the
724 North Atlantic nitrogen budget. *Global Biogeochemical Cycles*, 20(3), 1–14.
725 <https://doi.org/10.1029/2005GB002456>, 2006.

726 Garcia, H. E., Boyer, T. P., Locarnini, R. A., Antonov, J. I., Mishonov, A. V., Baranova, O. K., ...
727 Johnson, D. R.: *World Ocean Atlas 2013. Volume 3: dissolved oxygen, apparent oxygen*
728 *utilization, and oxygen saturation. NOAA Atlas NESDIS 75*, 2013a.

729 Garcia, H. E., Locarnini, R. A., Boyer, T. P., Antonov, J. I., Baranova, O. K., Zweng, M. M., ...
730 Johnson, D. R.: *World Ocean Atlas 2013, Volume 4 : Dissolved Inorganic Nutrients*
731 *(phosphate, nitrate, silicate). NOAA Atlas NESDIS 76 (Vol. 4)*, 2013b.

732 Grosse, F., Fennel, K., Zhang, H., Laurent, A.: Quantifying the contributions of riverine vs.
733 oceanic nitrogen to hypoxia in the East China Sea, *Biogeosciences*, [https://doi.org/10.5194/bg-](https://doi.org/10.5194/bg-2019-342)
734 2019-342, accepted for publication

735 Guo, J. S., X. M. Hu and Y. L. Yuan: A diagnostic analysis of variations in volume transport
736 through the Taiwan Strait using satellite altimeter data, *Advances in Marine Science*, 23(1):
737 20 – 26 (in Chinese with English abstract), 2005.

738 Haidvogel, D. B., Arango, H., Budgell, W. P., Cornuelle, B. D., Curchitser, E., Di Lorenzo, E., ...
739 Wilkin, J., *Ocean forecasting in terrain-following coordinates: Formulation and skill*

740 assessment of the Regional Ocean Modeling System. *Journal of Computational Physics*,
741 227(7), 3595–3624. <https://doi.org/10.1016/j.jcp.2007.06.016>, 2008.

742 Laurent, A., & Fennel, K.: Simulated reduction of hypoxia in the northern Gulf of Mexico due to
743 phosphorus limitation. *Elementa: Science of the Anthropocene*, 2(1), 000022.
744 <https://doi.org/10.12952/journal.elementa.000022>, 2014.

745 Laurent, A., Fennel, K.: Time-evolving, spatially explicit forecasts of the northern Gulf of
746 Mexico hypoxic zone. *Environmental Science & Technology*, 53, 14,449–14,458, doi:
747 10.1021/acs.est.9b05790, 2019.

748 Laurent, A., Fennel, K., Hu, J., & Hetland, R.: Simulating the effects of phosphorus limitation in
749 the Mississippi and Atchafalaya river plumes. *Biogeosciences*, 9(11), 4707–4723.
750 <https://doi.org/10.5194/bg-9-4707-2012>, 2012.

751 Laurent, A., Fennel, K., Cai, W.-J., Huang, W.-J., Barbero, L., Wanninkhof, R.: Eutrophication-
752 Induced Acidification of Coastal Waters in the Northern Gulf of Mexico: Insights into Origin
753 and Processes from a Coupled Physical-Biogeochemical Model. *Geophys. Res. Lett.*, 44 (2),
754 946–956. <https://doi.org/10.1002/2016GL071881>, 2017.

755 Li, D., Zhang, J., Huang, D., Wu, Y., & Liang, J.: Oxygen depletion off the Changjiang (Yangtze
756 River) Estuary. *Science in China Series D: Earth Science*, 45(12), 1137.
757 <https://doi.org/10.1360/02yd9110>, 2002.

758 Li, H. M., Tang, H. J., Shi, X. Y., Zhang, C. S., & Wang, X. L.: Increased nutrient loads from the
759 Changjiang (Yangtze) River have led to increased Harmful Algal Blooms. *Harmful Algae*, 39,
760 92–101. <https://doi.org/10.1016/j.hal.2014.07.002>, 2014.

761 Li, M., Lee, Y. J., Testa, J. M., Li, Y., Ni, W., Kemp, W. M., & Di Toro, D. M.: What drives
762 interannual variability of hypoxia in Chesapeake Bay: Climate forcing versus nutrient loading?
763 *Geophysical Research Letters*, 43(5), 2127–2134. <https://doi.org/10.1002/2015GL067334>,
764 2016.

765 Li, X., Bianchi, T. S., Yang, Z., Osterman, L. E., Allison, M. A., DiMarco, S. F., & Yang, G.:
766 Historical trends of hypoxia in Changjiang River estuary: Applications of chemical biomarkers
767 and microfossils. *Journal of Marine Systems*, 86(3–4), 57–68, 2011.
768 <https://doi.org/10.1016/j.jmarsys.2011.02.003>

769 Liu, K. K., Yan, W., Lee, H. J., Chao, S. Y., Gong, G. C., & Yeh, T. Y.: Impacts of increasing
770 dissolved inorganic nitrogen discharged from Changjiang on primary production and seafloor
771 oxygen demand in the East China Sea from 1970 to 2002. *Journal of Marine Systems*, 141,
772 200–217. <https://doi.org/10.1016/j.jmarsys.2014.07.022>, 2015.

773 Liu, S. M., Hong, G.-H., Ye, X. W., Zhang, J., & Jiang, X. L.: Nutrient budgets for large Chinese
774 estuaries and embayment. *Biogeosciences Discussions*, 6(1), 391–435.
775 <https://doi.org/10.5194/bgd-6-391-2009>, 2009.

776 Liu, S. M., Zhang, J., Chen, H. T., Wu, Y., Xiong, H., & Zhang, Z. F.: Nutrients in the
777 Changjiang and its tributaries. *Biogeochemistry*, 62(1), 1–18, 2003.

778 Locarnini, R. A., Mishonov, A. V., Antonov, J. I., Boyer, T. P., Garcia, H. E., Baranova, O.
779 K., ... Seidov, D.: World Ocean Atlas 2013. Vol. 1: Temperature. S. Levitus, Ed.; A.
780 Mishonov, Technical Ed.; NOAA Atlas NESDIS, 73, 40. [https://doi.org/10.1182/blood-2011-
781 06-357442](https://doi.org/10.1182/blood-2011-06-357442), 2013.

782 Ni, X., Huang, D., Zeng, D., Zhang, T., Li, H., & Chen, J.: The impact of wind mixing on the
783 variation of bottom dissolved oxygen off the Changjiang Estuary during summer. *Journal of
784 Marine Systems*, 154, 122–130. <https://doi.org/10.1016/j.jmarsys.2014.11.010>, 2016.

785 Ning, X., Lin, C., Su, J., Liu, C., Hao, Q., & Le, F.: Long-term changes of dissolved oxygen,
786 hypoxia, and the responses of the ecosystems in the East China Sea from 1975 to 1995. *Journal
787 of Oceanography*, 67(1), 59–75. <https://doi.org/10.1007/s10872-011-0006-7>, 2011.

788 Peña, A., Katsev, S., Oguz, T., & Gilbert, D.: Modeling dissolved oxygen dynamics and hypoxia.
789 *Biogeosciences*, 7(3), 933–957. <https://doi.org/10.5194/bg-7-933-2010>, 2010.

790 Qian, W., Dai, M., Xu, M., Kao, S. ji, Du, C., Liu, J., ... Wang, L.: Non-local drivers of the
791 summer hypoxia in the East China Sea off the Changjiang Estuary. *Estuarine, Coastal and
792 Shelf Science*, 1–7. <https://doi.org/10.1016/j.ecss.2016.08.032>, 2015.

793 Quiñones-Rivera, Z. J., Wissel, B., Justić, D., & Fry, B.: Partitioning oxygen sources and sinks in
794 a stratified, eutrophic coastal ecosystem using stable oxygen isotopes. *Marine Ecology
795 Progress Series*, 342, 69–83. <https://doi.org/10.3354/meps342069>, 2007.

796 Rabalais, N. N., Díaz, R. J., Levin, L. A., Turner, R. E., Gilbert, D., & Zhang, J.: Dynamics and
797 distribution of natural and human-caused hypoxia. *Biogeosciences*, 7, 585–619.
798 <https://doi.org/10.5194/bg-7-585-2010>, 2010.

799 Scavia, D., Bertani, I., Obenour, D. R., Turner, R. E., Forrest, D. R. & Katin, A.: Ensemble
800 modeling informs hypoxia management in the northern Gulf of Mexico, *P. Natl. Acad. Sci.
801 USA*, 114, 8823–8828, 2017.

802 Scully, M. E.: Physical controls on hypoxia in Chesapeake Bay: A numerical modeling study.
803 *Journal of Geophysical Research: Oceans*, 118(3), 1239–1256,
804 <https://doi.org/10.1002/jgrc.20138>, 2013.

805 Smolarkiewicz, P. K., & Margolin, L. G.: MPDATA: A finite-difference solver for geophysical
806 flows. *Journal of Computational Physics*, 140, 459-480, 1998.

807 Song, G., Liu, S., Zhu, Z., Zhai, W., Zhu, C., & Zhang, J.: Sediment oxygen consumption and
808 benthic organic carbon mineralization on the continental shelves of the East China Sea and the
809 Yellow Sea. *Deep-Sea Research Part II: Topical Studies in Oceanography*, *124*, 53–63.
810 <https://doi.org/10.1016/j.dsr2.2015.04.012>, 2016.

811 Tong, Y., Zhao, Y., Zhen, G., Chi, J., Liu, X., Lu, Y., ... Zhang, W.: Nutrient Loads Flowing into
812 Coastal Waters from the Main Rivers of China (2006–2012). *Scientific Reports*, *5*, 16678.
813 <https://doi.org/10.1038/srep16678>, 2015.

814 Umlauf, L., & Burchard, H.: A generic length-scale equation for geophysical. *Journal of Marine*
815 *Research*, *61*(2), 235–265. <https://doi.org/10.1357/002224003322005087>, 2003.

816 Wang, B.: Hydromorphological mechanisms leading to hypoxia off the Changjiang estuary.
817 *Marine Environmental Research*, *67*(1), 53–58,
818 <https://doi.org/10.1016/j.marenvres.2008.11.001>, 2009.

819 Wang, B., Wei, Q., Chen, J., & Xie, L.: Annual cycle of hypoxia off the Changjiang (Yangtze
820 River) Estuary. *Marine Environmental Research*, *77*, 1–5,
821 <https://doi.org/10.1016/j.marenvres.2011.12.007>, 2012.

822 Wang, B., Chen, J., Jin, H., Li, H., Huang, D., & Cai, W.-J.: Diatom bloom-derived bottom water
823 hypoxia off the Changjiang Estuary, with and without typhoon influence, *Limnology and*
824 *Oceanography*, *62*, 1552-1569, <https://doi.org/10.1002/lno.10517>, 2017.

825 Wang, H., Dai, M., Liu, J., Kao, S. J., Zhang, C., Cai, W. J., ... Sun, Z.: Eutrophication-Driven
826 Hypoxia in the East China Sea off the Changjiang Estuary. *Environmental Science and*
827 *Technology*, *50*(5), 2255–2263. <https://doi.org/10.1021/acs.est.5b06211>, 2016.

828 Wang, J., Yan, W., Chen, N., Li, X., & Liu, L.: Modeled long-term changes of DIN:DIP ratio in
829 the Changjiang River in relation to Chl- α and DO concentrations in adjacent estuary. *Estuarine,*
830 *Coastal and Shelf Science*, *166*, 153–160. <https://doi.org/10.1016/j.ecss.2014.11.028>, 2015.

831 Wei, H., He, Y., Li, Q., Liu, Z., & Wang, H.: Summer hypoxia adjacent to the Changjiang
832 Estuary. *Journal of Marine Systems*, *67*(3–4), 292–303,
833 <https://doi.org/10.1016/j.jmarsys.2006.04.014>, 2007.

834 Wei, H., Luo, X., Zhao, Y., & Zhao, L.: Intraseasonal variation in the salinity of the Yellow and
835 East China Seas in the summers of 2011, 2012, and 2013. *Hydrobiologia*, *754*(1), 13–28.
836 <https://doi.org/10.1007/s10750-014-2133-9>, 2015.

837 Wu, R. S. S.: Hypoxia: From molecular responses to ecosystem responses. *Marine Pollution*
838 *Bulletin*, *45*(1–12), 35–45. [https://doi.org/10.1016/S0025-326X\(02\)00061-9](https://doi.org/10.1016/S0025-326X(02)00061-9), 2002.

839 Yu, L., Fennel, K., & Laurent, A.: A modeling study of physical controls on hypoxia generation
840 in the northern Gulf of Mexico. *Journal of Geophysical Research C: Oceans*, *120*(7), 5019–
841 5039. <https://doi.org/10.1002/2014JC010634>, 2015a.

842 Yu, L., Fennel, K., Laurent, A., Murrell, M. C., & Lehrter, J. C.: Numerical analysis of the
843 primary processes controlling oxygen dynamics on the Louisiana shelf. *Biogeosciences*, *12*(7),
844 2063–2076. <https://doi.org/10.5194/bg-12-2063-2015>, 2015b.

845 Yuan, D., Zhu, J., Li, C., & Hu, D.: Cross-shelf circulation in the Yellow and East China Seas
846 indicated by MODIS satellite observations. *Journal of Marine Systems*, *70*(1–2), 134–149.
847 <https://doi.org/10.1016/j.jmarsys.2007.04.002>, 2008.

848 Zhang, H., Zhao, L., Sun, Y., Wang, J., & Wei, H.: Contribution of sediment oxygen demand to
849 hypoxia development off the Changjiang Estuary. *Estuarine, Coastal and Shelf Science*, *192*,
850 149–157. <https://doi.org/10.1016/j.ecss.2017.05.006>, 2017.

851 Zhang, J.: Nutrient elements in large Chinese estuaries. *Continental Shelf Research*, *16*(8), 1023–
852 1045. [https://doi.org/10.1016/0278-4343\(95\)00055-0](https://doi.org/10.1016/0278-4343(95)00055-0), 1996.

853 Zhao, L., & Guo, X.: Influence of cross-shelf water transport on nutrients and phytoplankton in
854 the East China Sea: A model study. *Ocean Science*, *7*(1), 27–43. [https://doi.org/10.5194/os-7-](https://doi.org/10.5194/os-7-27-2011)
855 [27-2011](https://doi.org/10.5194/os-7-27-2011), 2011.

856 Zheng, J., Gao, S., Liu, G., Wang, H., & Zhu, X.: Modeling the impact of river discharge and
857 wind on the hypoxia off Yangtze Estuary. *Natural Hazards and Earth System Sciences*, *16*(12),
858 2559–2576. <https://doi.org/10.5194/nhess-16-2559-2016>, 2016.

859 Zhou, F., Chai, F., Huang, D., Xue, H., Chen, J., Xiu, P., ... Wang, K.: Investigation of hypoxia
860 off the Changjiang Estuary using a coupled model of ROMS-CoSiNE. *Progress in*
861 *Oceanography*, *159*, 237–254. <https://doi.org/10.1016/j.pocean.2017.10.008>, 2017.

862 Zhou, F., Huang, D., Ni, X., Xuan, J., Zhang, J., & Zhu, K.: Hydrographic analysis on the multi-
863 time scale variability of hypoxia adjacent to the Changjiang River Estuary. *Shengtai Xuebao/*
864 *Acta Ecologica Sinica*, *30*(17), 4728–4740, 2010.

865 Zhu, J., Zhu, Z., Lin, J., Wu, H., & Zhang, J.: Distribution of hypoxia and pycnocline off the
866 Changjiang Estuary, China. *Journal of Marine Systems*, *154*, 28–40.
867 <https://doi.org/10.1016/j.jmarsys.2015.05.002>, 2016.

868 Zhu, Z.-Y., Zhang, J., Wu, Y., Zhang, Y.-Y., Lin, J., & Liu, S.-M.: Hypoxia off the Changjiang
869 (Yangtze River) Estuary: Oxygen depletion and organic matter decomposition. *Marine*
870 *Chemistry*, *125*(1–4), 108–116. <https://doi.org/10.1016/j.marchem.2011.03.005>, 2011.

871 Zweng, M. M., Reagan, J. R., Antonov, J. I., Mishonov, A. V., Boyer, T. P., Garcia, H. E., ...
872 Bidlle, M. M., World Ocean Atlas 2013, Volume 2: Salinity. NOAA Atlas NESDIS (Vol. 119).
873 <https://doi.org/10.1182/blood-2011-06-357442>, 2013.
874