



1 **Anoxic monimolimnia: Nutrients devious feeders**

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7

8 **Abstract**

9 This study focuses on the role of the meromictic anoxic basins' internal load: a) during
10 storm events and b) under the environments' typical stratification conditions. Measurements of
11 physicochemical parameters, nutrients, chlorophyll and hydrogen sulfide, four days after an
12 anoxic crisis in Aitoliko basin as well as data obtained from a biennial basin's monitoring, were
13 used. The relationships between temporal nutrient variations in the surface layer of an anoxic
14 basin with the changes on its water column physicochemical characteristics, the changes on the
15 bottom water phosphorus and nitrogen concentration and their effect on the basin's primary
16 productivity, were studied.

17 In coastal environments, storm events could result in water column total mixing. This
18 disturbance affects almost all the ecosystem's physical, chemical and biological parameters.
19 The basin becomes anoxic, massive fish kills occur and H₂S, PO₄³⁻ and NH₄⁺ release from
20 bottom waters to the interface and surface waters promoting algal blooms. Bottom layer can
21 supply the surface waters with nutrients, even during periods of high water column
22 stratification. Small scale, usually subtle, changes in physicochemical and hydrological basin's
23 characteristics promote this supply, affecting the ecosystem's primary production and shifting
24 its quality character.

25 Keywords: anoxia, eutrophication, internal load, storm events, stratification, meromictic
26 conditions



1 **1. Introduction**

2 Coastal regions are under strong human influence which is reflected into their water quality.
3 Oligotrophic estuaries and coastal systems have changed into mesotrophic and/or eutrophic, as
4 shown by an increase in toxic algal blooms, hypoxic/anoxic events, and massive mortalities of many
5 aquatic and benthic organisms. The relevance between the worldwide eutrophication increase and
6 the rapid spread of the hypoxic/anoxic environments is confirmed during the last decades. The
7 increase of organic and nutrient loading of the coastal zone, has caused the enhanced deterioration
8 of oxygen conditions depleted in environments where anoxia/hypoxia is a natural consequence of
9 their morphology and hydrodynamics (Vollenweider et al., 1992; Petricoli et al., 1996; Sorokin, et
10 al., 1996a; 1996b; Druon et al., 2004; Diaz, 2008; Rouso et al., 2009; Kemp et al., 2009; Zhang et
11 al., 2010; Rabalais et al., 2010).

12 In strongly stratified and productive water basins, bottom water dissolved oxygen is depleted
13 due to the excessive organic matter decomposition in these depths. Distribution and recycling of
14 nutrients in the overlying water column is inextricably dependent on oxygenation and redox
15 conditions. Nitrates (NO_3^-) are the predominant nitrogen form in oxygenated epilimnia, while nitrites
16 (NO_2^-) can be detectable throughout the water column with rather low concentrations. NO_2^- profiles
17 usually presents peaks near the oxic/anoxic interface. Their levels can go through a maximum in the
18 oxic waters as part of the ammonium (NH_4^+) oxidation and NO_2^- can be formed again because of the
19 NO_3^- use as an oxidant in anoxic waters. Bottom water anoxia accelerates phosphates (PO_4^{3-}),
20 ammonium (NH_4^+) and hydrogen sulphide (H_2S) generation and recycling with the accumulation
21 from organic matter decomposition. PO_4^{3-} is also released into the pore water and diffused into the
22 overlying bottom water when iron (oxy)hydroxides in the sediments are since to FeS because FeS
23 minerals do not bind PO_4^{3-} (Rozan et al., 2002; Diaz and Rosenberg, 1995; Rabalais, 2002; Luther et
24 al., 2004).

25 In permanently stratified water basins, the isolated bottom layers are not renewed under the
26 typical environments' conditions. Storm events (e.g. prolonged and severe winds) could result in
27 stratification destruction and water column total mixing. The water column turnover brings large



1 amounts of H₂S to the surface resulting in low levels of oxygen throughout the water column and
2 H₂S in the surface water, which are the conditions necessary for fish kills. Hydrogen sulfide release
3 into shallower depths is accompanied by PO₄³⁻ and NH₄⁺ release from bottom waters to the interface
4 and surface waters promoting algal blooms. Thus more organic matter is produced fueling anoxia
5 (Dassenakis et al., 1994; Fallesen et al., 2000; Kršinić et al., 2000; Astor et al., 2003; Luther et al.,
6 2004; Brandi et al., 2008; Njiru et al., 2010). The decisive influence of the internal load (accumulated
7 H₂S, PO₄³⁻ and NH₄⁺) in the ecosystem's function has been reported by numerous researchers in
8 different types of water environments, such as the salty lake Rogoznica in Croatia (Baric et al., 2003),
9 the Mariager fjord in Denmark (Fallesen et al., 1999), the internal bays of Delaware in USA (Luther
10 et al., 2004), the Aitoliko lagoon in Greece (Dassenakis et al., 1994; Leonardos and Sinis, 1997;
11 Demetriou et al., 2010) and the Tapi estuary in India (Ram et al., 2014).

12 Water column balance is usually restored in a relatively short period of time, after a
13 holomictic event, under the influence of local meteorological and hydrological processes.
14 Geochemical processes equilibrium is initiated by dissolved oxygen concentrations being reset to the
15 normal levels and this take longer, but not more than a few months. Environmental recolonisation,
16 starts soon after the physicochemical conditions return to normal, but requires several years before
17 the ecosystem be restored to the former conditions. An ultimate biological recovery is usually
18 unattainable (Leonardos and Sinis, 1997; Fallesen et al., 2000; Baric et al., 2003).

19 The question now becomes, is whether the balance of an anoxic water ecosystem is under
20 the threat of its hypolimnetic nutrient and sulfide load only in the case of storm events and water
21 column total mixing.

22 In polymictic water basins it is clear that the accumulated nutrients in the bottom layer will
23 supply surface waters after the pycnocline overturns. Besides this mechanism of basin water quality
24 degradation, it is now recognized as a one of the largest obstacles in eutrophic environmental
25 management and restoration efforts (Conley et al., 2009). The role of internal load, in permanently
26 stratified water basins, is not so clear. In the Baltic Sea, one of the largest anoxic environments of
27 the world, internal nutrient load, is implicated for the surface layers replenishment through vertical



1 mixing processes, exacerbating the basin's eutrophic conditions caused by the high external nutrient
2 and organic inflows in the Baltic Sea. Quantitative data cannot fully support this scenario and
3 therefore further research is required (Reissmann et al., 2009).

4 In the present study, the impact of storm events on water column stability and bottom water
5 hypoxia/anoxia of enclosed coastal basins is investigated. The importance of internal loading (H_2S ,
6 PO_4^{3-} and NH_4^+) has emerged. This leads to the disturbance on the main nutrients, dissolved oxygen,
7 hydrogen sulfide and chlorophyll distribution, following total water column mixing. Additionally,
8 the relationship between temporal nutrients variations in surface layers, of permanent anoxic coastal
9 basins with; a) changes in the physicochemical characteristics of the water column, b) changes in the
10 bottom water phosphorus and nitrogen concentrations, and c) their effect on the basin's primary
11 productivity, are studied.

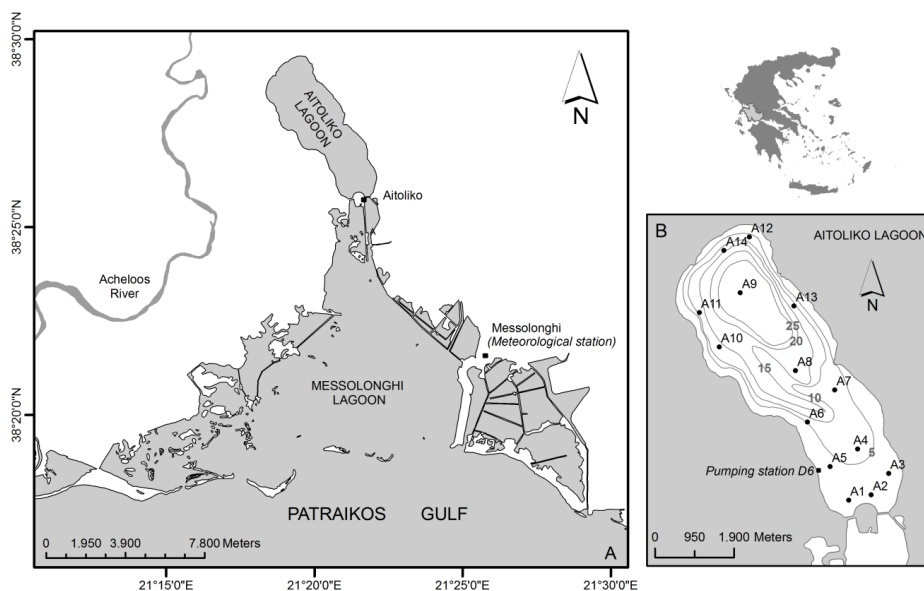
12 In order to achieve the objectives of this study, two different sets of Aitoliko basin's data
13 were used. The first one includes measurements of physicochemical parameters, nutrients,
14 chlorophyll and sulfides, four days after a storm event and the consequent anoxic crisis in Aitoliko
15 basin on 4th of December 2008. The second one contains a similar data set obtained from a biennial
16 (May 2006-May 2008) Aitoliko basin monitoring.

17

18 **2. Materials and methods**

19 *2.1. Study area*

20 Aitoliko is a semi-enclosed basin in western Greece. It covers an area of about 16 km² and
21 its maximum depth is 27.5m. It is characterized as non-typical lagoon by its depth and the fact that
22 its longitudinal axis is perpendicular to the shoreline. Aitoliko basin communicates southerly with a
23 typical shallow lagoon (Messolonghi lagoon) with mean depth of about 0.5 m. The lagoonal system
24 communicates southerly with Patraikos Gulf (maximum depth 100 m) (Fig. 1A). The two lagoons
25 are connected through shallow and narrow openings under the bridges that connect the town of
26 Aitoliko with the mainland.



1
2 Fig. 1. (A) Map of the extended study area. (B) Sampling stations in Aitoliko basin.

3

4 Aitoliko is a permanently highly-stratified coastal basin with its isolated bottom water known
5 to be anoxic and sulfidic. The large fresh water inflows are mainly arise through a pumping station
6 (Fig. 1B) which is located near the basin's sill. The inflows are implicated both for the basin's
7 permanent stratification and for the deterioration of water fluxes with its source basin (Messolonghi
8 lagoon).

9

10 2.2. Data collection

11 Four days after the storm event of the 4th of December 2008, a sampling cruise was carried
12 out in Aitoliko basin. During this cruise, continuous profiles of physicochemical parameters such as
13 temperature, conductivity, dissolved oxygen, redox potential and pH were measured in situ using a
14 Troll 9500 water quality multi-parameter instrument, at a network of fourteen (14) sampling sites
15 (Fig. 1B). Water samples were collected from the deepest part (sampling site A₉) of the Aitoliko
16 basin (Fig.1B), with 5 meters vertical intervals; using a 2.5 l Hydro Bios free flow sampler, during



1 that cruise. Water samples for sulfides, nutrient and chlorophyll determination were brought to the
2 laboratory in a portable refrigerator at 4°C.

3 Identical data were collected, on a monthly base, during a biennial (May 2006-May 2008)
4 monitoring of the Aitoliko basin.

5 There were no direct meteorological observations within the Aitoliko lagoon. Therefore, in
6 order to assess the effect of the wind on the lagoon's hydrography, wind measurements from a station
7 on the nearby Messolonghi town were used. This station is approximately 10 kilometres southeast of
8 Aitoliko town (Fig. 1A). Wind speed and wind direction time series, with 10 minutes temporal
9 resolution, were available. Daily means were calculated and studied for the analysis of the wind time
10 series.

11

12 2.3. Data processing

13 Measured temperature and conductivity data were corrected by a low pass filter to minimize
14 sharp spikes in salinity for the short-term mismatch of the sensor responses between temperature and
15 conductivity, using Matlab. Any spikes remaining in the salinity data were removed by calculating a
16 1m running average. The despiked temperature and salinity data were used to construct density
17 (σ_t) profiles.

18 In practice, density was not measured; it was calculated from *in situ* measurements of
19 pressure, temperature, and conductivity using the equation of state for sea water. For simplification,
20 physical oceanographers often quote only the last 2 digits of the density, a quantity they call *density*
21 *anomaly* or *Sigma* (s, t, p):

$$22 \sigma(s, t, p) = \rho(s, t, p) - 1 \text{ g/cm}^3 \quad (1)$$

23 When ocean surface layers are studying, compressibility can be ignored, and the sigma-t
24 quantity (written σ_t) can be used:

$$25 \sigma_t = \sigma(s, t, 0) \quad (2)$$

26 This is the density anomaly of a water sample when the total pressure on it has been reduced
27 to atmospheric pressure (*i.e.* zero water pressure), but the temperature and salinity are *in situ* values.



1 In this study the *Brünt–Väisälä frequency* or the *static stability frequency* was used,
2 evaluating the strength of density stratification in Aitoliko lagoon water column. The frequency
3 quantifies the importance of stability, and it is a fundamental variable in the dynamics of stratified
4 flow. In simplest terms, the frequency can be interpreted as the vertical frequency excited by a
5 vertical displacement of a fluid parcel.

6 In the ocean where salinity is important, the *Brünt–Väisälä frequency* (N^2) is expressed by
7 the equation:

$$8 \quad N^2 = -\frac{g}{\rho} \frac{d\rho}{dz} \quad (3)$$

9 where ρ , the potential density, depends on both temperature and salinity, and g the acceleration of
10 gravity.

11

12 2.4. Water samples storage and analysis

13 Water samples for sulfides (S^{2-}) measurements were always the first taken from the
14 sampling bottle immediately after the sampler was back on board, taking special care to obtain
15 samples with minimum aeration. To preserve samples for total sulfide determination, zinc acetate
16 and sodium hydroxide solutions were put into the bottles before filling them with sample. Four
17 drops of 2N zinc acetate solution per 100ml sample were used, while the final pH was always at
18 least 9. The bottles were always filled completely and stoppered immediately. Water samples for
19 sulfide analyses were measured unfiltered and freshly within 24h using the iodometric method
20 (APHA, 1998).

21 The different nitrogen forms, ammonia, nitrite and nitrate, were determined using the
22 indophenol blue, the colorimetric and the cadmium reduction method respectively. The ascorbic acid
23 method was used in order to determine orthophosphates in the water samples (APHA, 1998).

24 Chlorophyll, acetone extraction was following as soon as the samples were transferred to the
25 laboratory and about twenty hours later chlorophyll-a, b and c was determined using the trichromatic
26 colorimetric method (APHA, 1998).



1 **3. Results**

2 *3.1. Nutrients, chlorophyll and water quality characteristics of the Aitoliko basin*

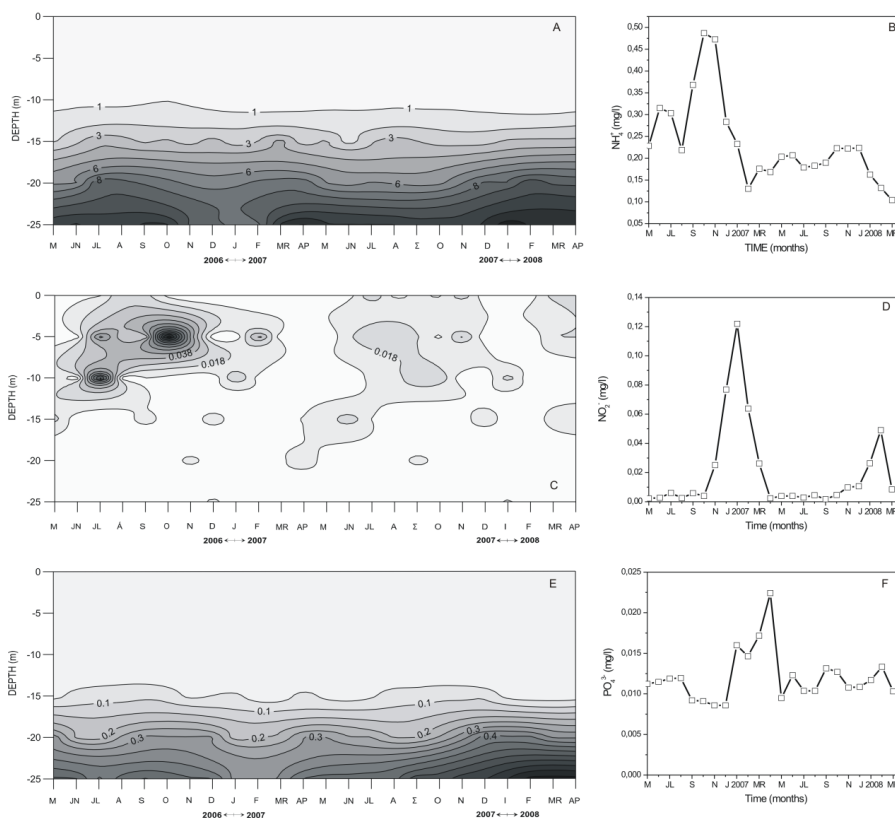
3 In this section, the data obtained during the biennial Aitoliko basin's monitoring are
4 presented, focusing on the analyses of nutrients, chlorophylls and sulfides spatiotemporal variations.

5 Relatively low concentrations of ammonium were determined at the surface layer of the
6 Aitoliko basin throughout the sampling period. More specifically, the average NH_4^+ concentrations
7 ranged between 0.1mg/l and 0.5mg/l at the upper 10m of the basin's water column. From this depth
8 to the basin's bottom a sharp increase of the ammonium concentrations was recorded, while the
9 maximum value of about 12mg/l was determined at the deepest sampling depth of 25m (Fig. 2A).

10 Focusing on the basin's epilimnion, remarkable temporal changes in ammonium
11 concentration were recorded. The relative high values, of about 0.25-0.5mg/l, identified in the surface
12 layer of the Aitoliko basin during the summer and autumn months of 2006 was sharply reduced
13 during the winter period 2006-2007 taking values in the range of 0.14-0.28mg/l. Comparable NH_4^+
14 concentrations (mean value of about 0.2mg/l), were measured in all the surface samples from the
15 spring to the autumn of 2007. A further reduction of the determined ammonium values was observed
16 during winter 2007-2008. In the March 2008 surface samples, 0.11mg/l NH_4^+ (mean value) were
17 measured (Fig. 2B). During the sampling period, characteristic temporal changes in the NH_4^+
18 concentrations were recorded in the Aitoliko basin bottom layer as well. In particular, NH_4^+
19 concentration of up to 10mg/l were determined at the maximum sampling depth (25m) from May to
20 October, while in the following winter months ammonium concentration decreased to below 8mg/l.
21 From the first spring months of 2007, NH_4^+ concentrations in the basin's hypolimnion returned to
22 the summer 2006 level, while a slight increase in its values (maximum concentration of 12mg/l) was
23 observed during spring of 2008 (Fig. 2A).

24

25



1

2 Fig. 2. (A) Depth – time diagram of ammonium concentration (NH_4^+) in mg/l, in the deepest sampling station
3 A₉ during the period May 2006-May 2008. (B) Monthly variation of mean NH_4^+ epilimnetic values during the
4 period May 2006-May 2008. (C) Depth – time diagram of nitrate concentration (NO_3^-) in mg/l, in the deepest
5 sampling station A₉ during the period May 2006-May 2008. (D) Monthly variation of mean NO_2^- epilimnetic
6 values during the period May 2006-May 2008. (E) Depth – time diagram of phosphate concentration (PO_4^{3-})
7 in mg/l, in the deepest sampling station A₉ during the period May 2006-May 2008. (F) Monthly variation of
8 mean PO_4^{3-} epilimnetic values during the period May 2006-May 2008.

9

10 Nitrates were detectable only in the basin's surface layer. The measured concentrations
11 varied from 0mg/l to 0.06mg/l, with the maximum values to characterize the depths of 5m and 10m
12 in July and October of 2006. The measured NO_3^- concentrations characterized by seasonal variation,
13 with the highest measured values to be recorded during the period July-November for both years of



1 basin's monitoring. The second year maximum was significantly lower than the first year
2 corresponding one (Fig. 2C).

3 Two seasonal maxima were observed in the vertical distribution of nitrites throughout the
4 Aitoliko basin water column. The first one characterized the basin's surface layer during the period
5 November 2006-March 2007. The maximum concentration of about 0.18mg/l was determined in the
6 depth of 0m, in January 2007. The second one characterized the period December 2007-March 2008.
7 During that time, nitrite concentrations in the Aitoliko basin's surface layer (maximum concentration
8 ~0.05mg/l) were significantly lower than those determined during the corresponding period of the
9 first monitoring year. In all the other samples, nitrite concentrations were lower than 0.01mg/l (Fig.
10 2D).

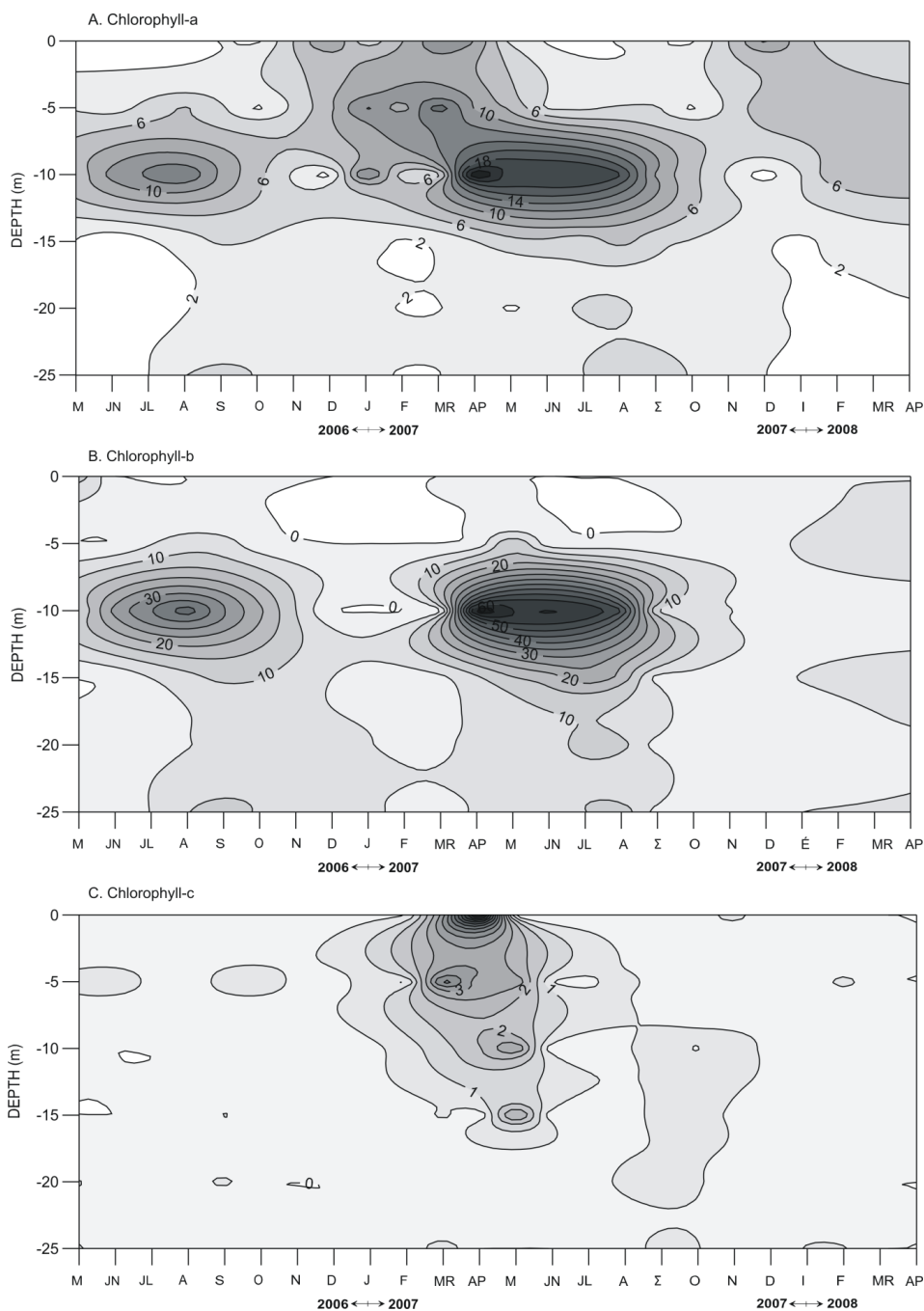
11 Relatively low values of orthophosphates characterized the surface layer of the Aitoliko
12 basin throughout the sampling period. More specifically, the average PO_4^{3-} identified in the surface
13 10m of the Aitoliko basin water column varied from 0.005mg/l to 0.023mg/l. A remarkable increase
14 on PO_4^{3-} concentration, in the 2006-2007 winter period, was recorded at the Aitoliko basin
15 epilimnion. The maximum values (~0.023mg/l PO_4^{3-}) were identified in the first spring months of
16 2007; while in May onwards the mean epilimnetic PO_4^{3-} values were 0.012mg/l (Fig. 2E). A sharp
17 increase of the phosphate concentrations with depth was recorded, and a maximum value of about
18 0.75mg/l was determined at the deepest sampling point of 25m. Just as in the case of the ammonium
19 measured values, orthophosphate concentrations in Aitoliko basin bottom layer, varied temporally.
20 In particular, PO_4^{3-} concentrations up to 0.5mg/l were determined at the maximum sampling depth
21 (25m) during the summer and autumn months of 2006. In the winter period of the first sampling year,
22 a slight decrease in the PO_4^{3-} concentration was recorded (maximum concentration of 0.45mg/l).
23 From the first spring months of 2007, PO_4^{3-} concentrations in the basin's hypolimnion returned to the
24 summer 2006 level, while a slight increase in its values (maximum concentration of 0.75mg/l) was
25 observed during spring of 2008 (Fig. 2F).

26 As expected, higher chlorophyll-a concentrations characterized the surface layer of the
27 Aitoliko basin throughout the sampling period. More specifically, during the summer months of



1 2006, chlorophyll-a concentrations higher than $12\mu\text{g/l}$ were determined at the depth of 10m, while
2 the 5 surface meters were characterized values lower than $4\mu\text{g/l}$. At the deep layer (below the depth
3 of 15m) chlorophyll-a values varied from $1\mu\text{g/l}$ to $3\mu\text{g/l}$. In October, November and December of
4 2006 at the entire basin's water column notably low chlorophyll-a concentrations were determined.
5 Since January 2007, a gradual increase was recorded in the measured chl-a values. In the winter
6 period, chl-a concentration increases were noticeable at the 5 surface meters, while during the spring
7 months the higher values were determined at the depth of 10m. Concentrations higher than $22\mu\text{g/l}$
8 were measured at the depth of 10m in April 2007, when the chlorophyll-a values of the basin's
9 shallower layer were about $10\mu\text{g/l}$. During the summer period high chlorophyll-a concentrations up
10 to $18\mu\text{g/l}$ were determined at the depth of 10m. In agreement with the first sampling year,
11 chlorophyll-a distribution, low chl-a concentrations characterized the entire basin's water column in
12 the late fall of 2007, while an increasing trend of the parameter's values at the surface of Aitoliko
13 was evident during the first months of 2008. In the early spring of 2008 a chlorophyll-a concentration
14 of $7.4\mu\text{g/l}$ was measured at the depth of 10m (Fig. 3A).

15 The seasonality of chlorophyll-a distribution in Aitoliko water column characterized
16 chlorophyll-b variations as well. The surface 5m, layer of the Aitoliko basin was characterized by
17 chlorophyll-b concentration lower than $5\mu\text{g/l}$ throughout the sampling period. From June to October
18 2006 high concentrations of chlorophyll-b were identified in the depth of 10m, with the maximum
19 value of $43\mu\text{g/l}$ to be measured at this depth in August. Nearly zero chl-b concentrations were
20 identified throughout the basin's water column during the winter period 2006-2007. In April 2007,
21 $61\mu\text{g/l}$ chl-b was identified at the depth of 10m, whereas high concentrations were determined at this
22 depth throughout the spring and summer period of 2007. From October 2007 to April 2008 notably
23 low chlorophyll-b values characterized the entire Aitoliko basin water column (Fig. 3B).



1
2 Fig. 3. Depth – time diagram of chl-a (A), chl-b (B) and chl-c (C) in $\mu\text{g/l}$, in the deepest sampling station A₉
3 during the period May 2006-May 2008.



1 Chlorophyll-c distribution was differentiated from those of chl-a and b. In general, low
 2 concentrations (0-0.5µg/l) of chlorophyll-c were determined in Aitoliko basin. Only exception was
 3 the spring of 2007 during which relative high values of chl-c determined up to the depth of 15m. The
 4 maximum concentration characterized April 2007, where about 9µg/l was determined on the basin's
 5 surface (0m) (Fig. 3C).

6 During the biennial monitoring period, accumulated sulfides reached concentrations up to
 7 56.8mg/l near the bottom. In table 1 the determined sulfide concentrations at the maximum sampling
 8 depth (25m) throughout the sampling period (May 2006-May 2007), are presented. The higher
 9 sulfide values characterized the summer periods of both sampling years, while reduced
 10 concentrations were determined on February 2007 (31mg/l) and February 2008 (30mg/l).

11
 12 Table 1. Sulfide concentrations at the maximum sampling depth (25m) throughout the sampling period (May
 13 2006-May 2007). Mean monthly D₆ pumping station, discharge, during the period May 2006-May2008.

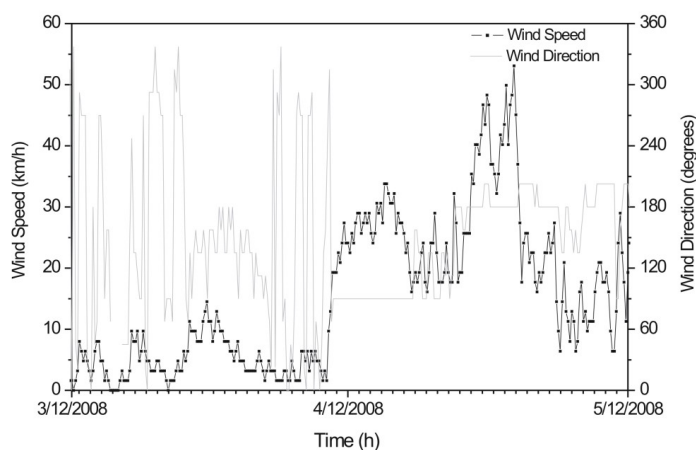
Year	Month	Sulfide concentration A9 (25m) (mg/l)	D ₆ Discharge (x10 ⁶ m ³ /month)
2006	M	41,1	1,02
	JN	41	1,16
	JL	47,6	1,33
	AU	42,6	1,65
	S	47,8	1,23
	O	48,8	0,77
	N	50,8	0,71
2007	D	50,8	0,69
	J	49	0,63
	F	31	0,56
	MR	44,24	0,70
	AP	52,2	0,78
	M	52,92	0,72
	JN	46,12	0,69
	JL	55,16	0,64
	AU	51,8	0,64
	S	39,4	0,66
	O	46,1	0,78
2008	N	50,67	0,85
	D	55,6	0,80
	J	55,36	0,74
	F	30	0,71
	MR	44,6	0,87
	AP	55,48	0,98
	M	56,8	1,28

14



1 3.2. The storm event of December 2008

2 *Description of the event:* On the 4th of December, easterly to southerly winds, with speed up
3 to 50km/h were blowing in the area of Aitoliko basin (Fig. 4). These winds caused, the forcible enter
4 of water from Messolonghi lagoon into Aitoliko basin disturbing its water column balance and
5 leading to a holomictic event.



6
7 Fig. 4. Wind speed and wind direction, during 3-4/12/2008.

8
9 When the wind stopped, low dissolved oxygen concentrations were reported at the south part
10 of the basin (Dimitriou et al., 2010) and zero oxygen was recorded in the northern part. This
11 represents the spatial character in this anoxic crisis. During the following hours, birds were observed
12 on the surface of the water feeding, while fishes which had reached the shore were gulping at the
13 surface of the water in an attempt to obtain oxygen. The fish kills, that followed this storm event
14 reflected the direct impact of the anoxic crisis in the ecosystem's biology.

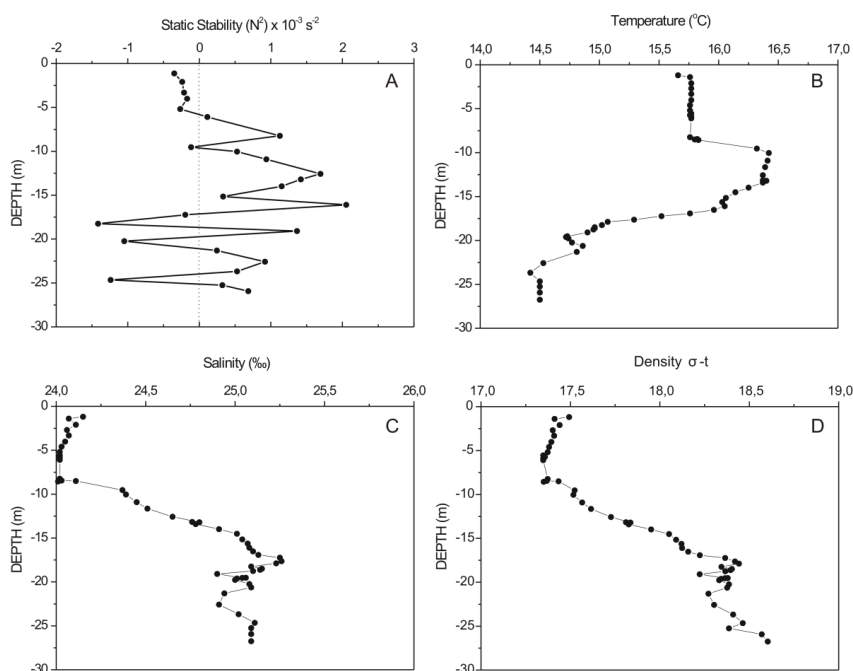
15 *Physical-chemical parameters:* Just after the holomictic event, the basin's water column was
16 nearly homogenous and quite unstable. After an intense storm capable for basin's mixing, water
17 column static stability is reduced at all depths. Alternating of low negative and positive static stability



1 values illustrated water column instabilities and indicated the mixing depth in the Aitoliko basin
2 water column (Fig.5A).

3 Slight differences between surface and bottom temperature, salinity and density (σ -t)
4 values were recorded in Aitoliko basin water column four days after the anoxic crisis of the 4th of
5 December 2008. The measured surface temperature, salinity and σ -t were about 15.5°C, 24‰
6 and 16.5 respectively, when the corresponding bottom values were just about 16.5°C, 25‰ and 17.5.
7 In the vertical distributions of these parameters, residual characteristics of the Aitoliko basin
8 stratification were retained (Fig. 5B, C, D).

9



10

11 Fig. 5. Vertical profiles of static stability (A) temperature (B), salinity (C) and density (σ -t) (D) in the
12 deepest station A₉ during December 2008.

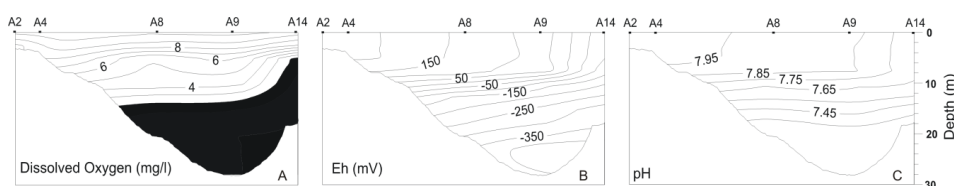
13

14 During the sampling of the 8th of December 2008, a well oxygenated surface layer down to
15 about 5m deep was recorded at the southern and the central part of the Aitoliko basin. Below that



1 depth dissolved oxygen concentrations decreased rapidly. The hypoxic and anoxic zones were spatial
2 related and reached their maximum extent in the northern part of the lagoon (Fig. 6A). It is interesting
3 to mention that the hypoxic zone extended below 14m, 8m and 4m at A₉, A₁₂ and A₁₄ stations
4 respectively. Concurrently, zero oxygen concentrations were measured below 18m, 11m and 10m at
5 A₉, A₁₂ and A₁₄ stations respectively.

6



7

8 Fig. 6. Dissolved oxygen (A), Eh (B) and pH (C) vertical distribution across A₂-A₄-A₈-A₁₄ cross section,
9 during December 2008.

10

11 Redox potential and pH values were in accordance with the vertical and spatial distribution
12 of dissolved oxygen (Fig. 6B, C). At the northern part of Aitoliko lagoon lower surface redox
13 potential (~-80) and pH (~7.6) values existed, compared to the rest of the Aitoliko basin. Moreover,
14 bottom layer redox potential decreased to about -400mV, reflecting pH (~ 7.3-7.4) and dissolved
15 oxygen decreases.

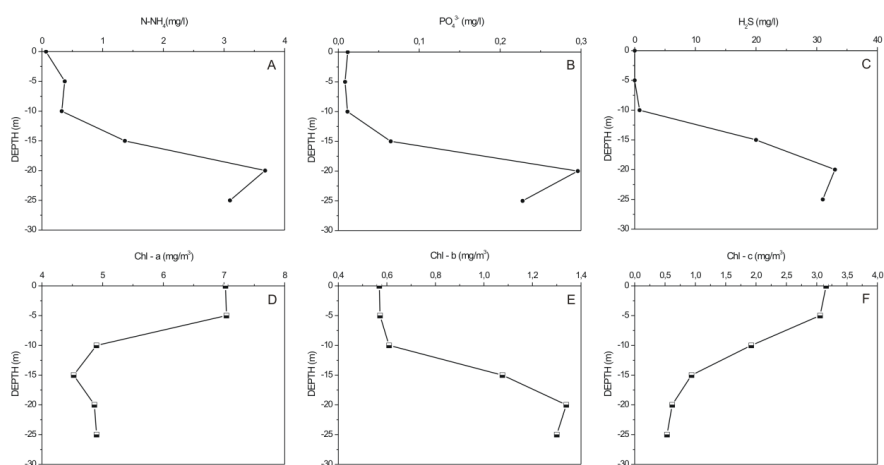
16 *Nutrients, sulfides and chlorophyll:* During the holomictic period (December 2008), nitrates
17 and nitrites were below the detection limits throughout the Aitoliko basin water column. The mean
18 ammonium concentration determined at the 10 surface meters was about 0.25mg/l, while a maximum
19 value of 3.67mg/l was characterized the 20m depth. A slight decrease of NH₄⁺ (3.09mg/l) was
20 recorded at greater depths (Fig. 7A).

21 Orthophosphates vertical distribution after the anoxic crisis is presented in Fig. 7B. Down to
22 the depth of 15m the measured PO₄³⁻ concentration was lower than 0.06mg/l. A peak of 0.3mg/l was
23 determined at the depth of 20m while a slight decrease (0.23mg/l) of orthophosphate concentrations
24 characterized greater depths.



1 About 0.8mg/l sulfides were determined in the 10 meter surface waters, while a maximum
2 value of 33mg/l was characterized the 20m depth. A slight decrease of sulfides concentration
3 (31mg/l) was recorded at greater depths (Fig. 7C).

4



5

6 Fig. 7. Vertical profiles of ammonia (A), phosphate (B), sulfides (C) and chlorophyll-a, b, c (D, E, F) in
7 Aitoliko lagoon during December 2008.

8

9 In the first days of December 2008, just after the total mixing event in Aitoliko basin, relative
10 high chlorophyll-a concentrations were determined throughout the water column. The five surface
11 meters were characterized by 7µg/l chl-a, while from this depth and deeper, 5µg/l chl-a were
12 determined (Fig. 7D). In the surface layer 0.67µg/l chl-b were measured while at greater depths
13 parameter's increase was recorded (Fig. 7E). In contrast, the maximum chl-c values (3µg/l)
14 characterized the basin's surface and they decreased with depth's increase (Fig. 7F).

15

16 4. Discussion

17 Permanent stratified coastal basins, like Aitoliko, are characterized by anoxic and sulfidic
18 hypolimnia. In enclosed or silled basins, anoxia occurs naturally, through their morphology and high
19 water residence times and its origin is rarely linked with eutrophication phenomena.



1 Anoxia in Aitoliko lagoon is known since 18th century, as historical records are referred in
2 total mixing events, massive fish kills and H₂S sulfide release. Measurements of physicochemical
3 parameters of Aitoliko go back to 1951. During the period 1951-2004, bottom layer temperatures
4 show relatively constant values while salinity values ranged about 5‰. The oxic/anoxic interface
5 progressively decreased from a depth of 18m in 1951 to 4m in 2003-2004. During the summer period
6 of the years 2006 and 2007, the oxic/anoxic interface developed in a depth of 17-18m. In the winter
7 periods of the same years low dissolved oxygen concentrations were determined in Aitoliko basin's
8 monimolimnio. This fact was ascribed to the anthropogenic deepening of the sill that connects
9 Aitoliko and its source basin (Messolonghi lagoon). Greater amounts of dense-salty water flowed
10 from Messolonghi lagoon, causing limited vertical mixing of the Aitoliko basin. The hydrodynamic
11 processes controlled the small scale mixing of the basin's water column, introduced oxygen into the
12 halocline and the bottom waters, without destroying stratification (Gianni et al., 2011; Gianni and
13 Zacharias, 2012).

14 Quite a few times in the past, anoxic conditions reached the water surface and the whole
15 water column of Aitoliko basin was anoxic for several days. Usually, during autumn or winter months
16 after intense southern or south-eastern winds, the balance that keeps the anoxic water under the
17 lighter well oxygenated water is disturbed. Wind enforces large quantities of salty water to enter from
18 Messolonghi into Aitoliko basin. This dense water mass is passing under the relative lighter anoxic
19 water of the Aitoliko basin, helping it to emerge. During these events, hydrogen sulfide, which is
20 accumulated in the lagoon's bottom layer, is released in the upper layers killing all the aquatic
21 organisms, while the atmosphere becomes stuffy for the local residents. The first reported mass
22 mortality event in the Aitoliko lagoon is referred to in 1881. Since then a numerous of such events
23 have been recorded in 1963, 1990 (Dassenakis, 1994; Leonardos and Sinis, 1997), 1992, 1995 and
24 2001. Such events have a decisive influence into the basin's physical, chemical and biological
25 characteristics and processes. The most recent holomictic event in Aitoliko basin is the one described
26 in this study and is occurred on the 4th of December 2008.



1 Regarding the Aitoliko basin's nutrient budget and productivity the most comprehensive
2 overview is presented in Daneilides (1991) study. Orthophosphate, total phosphorus, ammonium,
3 nitrite, nitrate and chlorophyll-a (chl-a) profiles were identified in the deepest part of Aitoliko basin
4 during the period April 1984-April 1985. Nutrient concentrations in the Aitoliko basin's surface layer
5 were not significantly high, indicating their consumption of the planktonic organisms and their
6 removal to greater depths through dead cells. The year average concentrations of TP, PO_4^{3-} , NH_4^+ ,
7 NO_3^- and NO_2^- in this layer were approximately 0.03mg/l, 0.02mg/l, 0.2mg/l, 0.1mg/l and 0.012mg/l,
8 respectively (Daneilidis, 1991). After the spring plankton bloom, nitrogen concentrations were
9 significantly reduced in the epilimnion, resulting in the inorganic N to inorganic P ratio reduction.
10 Increase of all the inorganic nitrogen forms in the winter period, rendered phosphorus the growth
11 limiting factor. A general reduction of all nutrient concentrations and increases in chl-a values
12 occurred, with the spring bloom onset. During the 1984-1985 period the reported spring chl-a
13 maximum was about 10 $\mu\text{g/l}$. The basin's deep layer was characterized by high ammonium,
14 orthophosphate and total phosphorus concentrations with constant values throughout the sampling
15 period. Specifically, about 15mg/l, 2mg/l and 2.5mg/l NH_4^+ , PO_4^{3-} and TP were determined in
16 Aitoliko basin bottom layer during the 1984-1985 period.

17 Under the typical meromictic conditions prevailed in Aitoliko basin, throughout the two
18 years of monitoring (2006-2008), low ammonium, nitrate, nitrite and orthophosphate concentration
19 were determined in the surface 10m. At greater depths where, anoxic conditions prevailed, a sharp
20 increase in ammonium and orthophosphate concentrations were recorded. Just after the Aitoliko
21 basin total mixing (December 2008), nitrate and nitrite were below the detection limits throughout
22 the water column, while the measured ammonia concentrations below 15m were approximately 4
23 times lower than that measured under the typical meromictic conditions that commonly prevail
24 during the winter time. This reduction is probably due to ammonium oxidation during water column
25 mixing, as large amounts of oxygenated water flows into basin's bottom. Changes in orthophosphate
26 concentrations after the storm event were observed as well. The orthophosphate deficiency during
27 this time was probably caused by its removal from the water column by absorption of orthophosphate



1 on re-oxidized mineral phases of Mn- and Fe- oxides (Sundy et al., 1992) and by precipitation of
2 high insoluble calcium phosphate (Bjerum et al., 1958; Drever, 1982).

3 Bottom overturn increases the nutrient content of surface layers, enhancing algal blooms.
4 Chlorophyll-a and algal cell abundances in surface layers are correlated with changes of the
5 oxic/anoxic interface depth. Chlorophyll-a in surface waters decreases sharply as the oxic/anoxic
6 interface moves deeper, while the depth of oxic/anoxic interface affects the prevalence of
7 phytoplankton species (Ma et al., 2006). Chlorophyll-a is the most appropriate parameter to follow
8 the growth of both algae and *Cyanobacteria*, in a coastal environment. Chlorophyll-a is a good
9 indicator of the biomass of algae present in waters (Räike et al., 2003). The increase of nutrients in
10 surface layers can also amplify the growth and prevalence of harmful algal species in the water
11 column which grow quickly with division rates of 1 to 5d⁻¹ (Kreiberg, 1999). Higher harmful algal
12 cell density occurred when more H₂S developed in the bottom waters and the interface moves to
13 shallower depths. Large blooms of dinoflagellates were recorded in the surface waters of Torquay
14 Canal when the interface migrates upward, while diatoms prevail when anoxic conditions limited in
15 great depths (Ma et al., 2006).

16 The total mixing event in Aitoliko basin in the first days of December 2008 led to
17 chlorophyll-a concentration increases throughout the water column, comparing with the measured
18 values under the typical meromictic conditions prevail commonly during the winter time. Changes
19 observed in the vertical distribution of chlorophyll-b as well. Unlike the typical winter profiles, chl-
20 b maximum concentrations were recorded at the basin's surface while a sharp decrease was observed
21 with increasing depth. Additionally, the vertical profile of chlorophyll-c four days after the storm
22 event was remarkable. Chl-c mainly occurs in algae which are absent from the surface layer under
23 the typical winter conditions in Aitoliko basin. Values equal to 3µg/l were measured on 8th of
24 December 2008. The abrupt increase in chlorophyll-c, immediately after the total mixing event, could
25 be associated with a bloom, for example, of dinoflagellates which encouraged under these conditions.
26 Certainly the observed changes indicate change in the composition of the lagoon's phytoplankton



1 community, can only confirmed by analytical determination of the prevailing algal types. In this
2 study algal community structure was not determined.

3 The changes induced in the physical, chemical and biological characteristics of an anoxic
4 meromictic basin as Aitoliko, following total mixing, highlight the impact of accumulated nutrients
5 and sulfides in the bottom layer.

6 However, questions arising are related to the importance of hypolimnetic nutrient and sulfide
7 load under meromictic conditions, and its relation with the surface water quality. Trying to address
8 out this concern, the focus was on the temporal variations of specific nutrients in the Aitoliko surface
9 layer and their relation with; a) changes on the physicochemical characteristics of the basin's water
10 column, b) changes on the bottom water phosphorus and nitrogen concentration, and c) their effect
11 on the basin's primary productivity.

12 The relative high concentrations of ammonium characterized the surface layer of the Aitoliko
13 basin during the summer and autumn months of 2006 and that was sharply reduced during the winter.
14 At the same period, increase of epilimnetic nitrites concentration, (ammonium oxidation intermediate
15 product) and decrease of the bottom layer ammonium and orthophosphate values were recorded.
16 These changes coincide with the Aitoliko basin's hypolimnion oxygenation due to weak/small scale
17 mixing of the water column caused by the denser water inflow from the Messolonghi lagoon in the
18 summer months (Gianni et al., 2011). Dissolved oxygen increase in Aitoliko basin's water column
19 may be the answer to the observed nutrient variations during that period.

20 In addition, during the 2006-2007 winter period, a PO_4^{3-} concentration increase in the surface
21 layer of the Aitoliko basin, was recorded. This increase could be attributed to external sourcing with
22 inorganic phosphorus. This scenario could be strengthened, considering that the basin's main
23 freshwater source is the D₆ pumping station. D₆ is located near the basin's entrance and supplies
24 Aitoliko with water from an extensive drainage network which extends to the adjacent cultivated
25 land. The seasonal PO_4^{3-} maximum surface concentrations coincide with the minimum flow of the
26 D₆ pumping station (Table 1), and is lagging 3-4 months behind the main fertilization period,
27 weakening the external loading scenario. In a second scenario, the observed epilimnetic PO_4^{3-}



1 increase could be attributed to the surface layers enrichment from the deeper waters. The
2 hydrodynamic processes that control the small scale mixing of Aitoliko basin water column,
3 introducing oxygen into the halocline and the bottom waters, without destroying the stratification
4 (Gianni et al., 2011; Gianni and Zacharias, 2012), probably also governs the nutrients transport from
5 the deep pool.

6 The phosphorus concentration increases in the Aitoliko epilimnion during winter and spring
7 months of 2007 had an immediate impact on the basin's primary productivity. A particularly high
8 chlorophyll-a concentration was determined the spring of 2007 and that was significantly higher than
9 those typically characterize Aitoliko basin (see spring 2006 values; Daneilides, 1991). Moreover
10 during this period, a maximum of chlorophyll-c in the surface of the Aitoliko basin was observed
11 providing evidence that the spring-summer phytoplankton bloom accompanied changes in the
12 dominant planktonic types.

13 During the 2007-2008 winter period an additional reduction on the epilimnetic NH_4^+
14 concentration was recorded and accompanied increased NO_2^- concentrations. The intensities of both
15 changes were milder than that reported for the previous sampling year. No changes were observed in
16 the surface PO_4^{3-} concentrations while hypolimnetic NH_4^+ and PO_4^{3-} values increased during that
17 period. The spring 2007 algal bloom increased the organic load and was enhanced in the anoxic
18 layers of the Aitoliko basin. The decomposition of this organic matter can be linked the bottom layer
19 nitrogen and phosphorus content rise. It becomes obvious that the Aitoliko bottom water oxygenation
20 episode in the winter months of 2008 had different impact on the basin's water column and this is
21 reflected in nutrient and chlorophyll profiles.

22

23

24 **5. Conclusions**

25 Storm events in coastal environments can result in stratification destruction and water total
26 column mixing. In systems where anoxic/sulfidic bottom waters exist, vertical mixing leads to a
27 basin's anoxic crisis. The massive fish kills is the result of an intense disturbance of a natural system



1 which led to the dissolved oxygen consumption in the surface water by hydrogen sulfide the anoxic
2 water layer and direct H₂S toxicity. Hydrogen sulfide release is accompanied by PO₄³⁻ and NH₄⁺
3 release from bottom waters to the interface and surface waters promoting algal blooms. This
4 disturbance affects almost all the physical, chemical and biological parameters in a water basin.

5 A few days after the total mixing the physical rehabilitation of the basin's water column
6 starts. The surface layer is gradually oxygenated; indicating water column balance restoration, under
7 the influence of the local meteorological and hydrological processes. Geochemical processes
8 equilibrium starts by dissolved oxygen concentrations resetting to the normal levels and this takes
9 longer but not more than a few months. Environmental recolonization starts soon after the
10 physicochemical conditions return to normal, but requires several years before the ecosystem be
11 restored.

12 The induced changes in the physical, chemical and biological characteristics of an anoxic
13 meromictic basin, after its total mixing, highlight the impact of the accumulated nutrients and sulfides
14 in the bottom layer.

15 Just as significant is the role of hypolimnetic nutrient and sulfide load in meromictic periods.
16 The H₂S, NH₄⁺, PO₄³⁻ rich bottom layer can supply nutrients to the surface waters when the
17 physicochemical and hydrological basin's water column characteristics permit it. Nutrient and
18 organic matter flux water column turn over affects both quantitative nature of and qualitative the
19 basin's primary productivity and shifts the ecosystems quality character.

20 **Author contribution:** Areti Gianni and Ierotheos Zacharias designed and carried out the samplings
21 while the experiments were designed and carried out by Areti Gianni. Areti Gianni prepared the
22 manuscript with contributions from Zacharias.

23 **Competing interests:** The authors declare that they have no conflict of interest.

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1 References

- 2 APHA: Standard Methods for the Examination of Water and Wastewater, American Public Health
3 Association, Washington, DC, 1998.
- 4 Astor, Y., Muller-Karger, F., and Scranton, M.I.: Seasonal and interannual variation in the
5 hydrography of the Cariaco Basin: implications for basin ventilation, *Cont. Shelf Res.*, 23,
6 125–144, [https://doi.org/10.1016/S0278-4343\(02\)00130-9](https://doi.org/10.1016/S0278-4343(02)00130-9), 2003.
- 7 Baric, A., Grbec, B., Kuspilic, G., Marasovic, I., Nincevic, Z., and Grubelic, I.: Mass mortality event
8 in a small saline lake (lake Rogoznica) caused unusual holomictic conditions, *Sci. Mar.*, 67,
9 129-141, <https://doi.org/10.3989/scimar.2003.67n2129>, 2003.
- 10 Bjerum, J., Schwarzenbach, G., and Sillen, L.G.: Stability Constants, in: *Inorganic Ligands Part II*,
11 Special Publication No. 7, The Chemical Society of London: London, UK, pp 125-129, 1958.
- 12 Brandi, K. R., Michael, A. A., and Christopher, A.: Hydrogen sulfide production and volatilization
13 in a polymictic eutrophic saline lake, Salton Sea, California, *Sci. Total Environ.*, 406, 205-
14 218, <https://doi.org/10.1016/j.scitotenv.2008.07.021>, 2008.
- 15 Conley, D.J., Bonsdorff, E., Carstensen, J., Destouni, G., Gustafsson, B.G., Hansson, L.A., Rabalais,
16 N.N., Voss, and M., Zillen, L.: Tackling hypoxia in the Baltic Sea. Is engineering a solution?
17 *Environ. Sci. Technol.*, 43, 3407–3411, <https://doi.org/10.1021/es8027633>, 2009.
- 18 Daneilidis, D.: Systematic and ecological study of the diatoms of Messolonghi-Aetoliko-Klisova
19 lagoons, PhD Thesis, University of Athens, 1991.
- 20 Dassenakis, M., Krasakopoulou, E., and Matzara, B.: Chemical characteristics of Aetoliko lagoon,
21 Greece, after an ecological shock, *Mar. Pollut. Bull.*, 28, 427-433,
22 [https://doi.org/10.1016/0025-326X\(94\)90128-7](https://doi.org/10.1016/0025-326X(94)90128-7), 1994.
- 23 Diaz, R.J., and Rosenberg, R.: Marine benthic hypoxia: A review of its ecological effects and the
24 behavioural responses of benthic macrofauna, *Oceanogr. Mar. Biol.*, 33, 245–303, 1995.
- 25 Diaz, R.J., and Rosenberg, R.: Spreading dead zones and consequences for marine ecosystems,
26 *Science*, 321, 926–929, <https://doi.org/10.1126/science.1156401>, 2008.



- 1 Dimitriou, E., Rousi, A., Tasioulis, J., and Koutsikopoulos, C.: Environmental parameters recorded
2 during and after an anoxic crisis in the Etoliko lagoon (Western Greece), in: Proceedings of
3 14th Hellenic Congress of Ichthyologists, Piraeus, p 78, 2010.
- 4 Drever J.I.: The Geochemistry of Natural Waters. Surface and Groundwater Environments, Third
5 Edition, Prentice Hall, New Jersey, US, 1982.
- 6 Druon, J.N., Schrimpf, W., Dobricic, S., and Stips, A.: Comparative assessment of large-scale marine
7 eutrophication: North Sea area and Adriatic Sea as case studies, *Mar. Ecol.-Prog. Ser.*, 272,
8 1-23, <https://doi.org/10.3354/meps272001>, 2004.
- 9 Fallesen, G., Andersen, F., and Larsen, B.: Life, death and revival of the hypertrophic Mariager Fjord,
10 Denmark, *J. Marine Syst.*, 25, 313-321, [https://doi.org/10.1016/S0924-7963\(00\)00024-5](https://doi.org/10.1016/S0924-7963(00)00024-5),
11 2000.
- 12 Gianni, A., Kehayias, G., and Zacharias, I.: Geomorphology modification and its impact to anoxic
13 lagoons, *Ecol. Eng.*, 37, 1869-1877, <https://doi.org/10.1016/j.ecoleng.2011.06.006>, 2011.
- 14 Gianni, A., and Zacharias, I.: Modeling the hydrodynamic interactions of deep anoxic lagoons with
15 their source basins, *Estuar. Coast. Shelf S.*, 110, 157-167,
16 <https://doi.org/10.1016/j.ecss.2012.04.030>, 2012.
- 17 Hatzikakidis, A.: Seasonal hydrological study in Messolonghi-Aitoliko lagoon, In 'Proceedings of
18 the Hellenic Hydrobiological Institute, Volume V, Athens. p 85, 1951.
- 19 Kemp, W.M., Testa, J.M., Conley, D.J., Gilbert, D., and Hagy, J.D.: Temporal responses of coastal
20 hypoxia to nutrient loading and physical controls, *Biogeosciences*, 6, 2985–3008,
21 <https://doi.org/10.5194/bg-6-2985-2009>, 2009.
- 22 Kreiberg, H.: Analysis of harmful blooms of the dinoflagellate *Heterosigma* at the Pacific Biological
23 Station Mariculture Facility in 1993 and 1997, in: Clarke C, Eds Aquaculture update 85,
24 Pacific Biological Station, Nanaimo, BC, pp. 93-99, 1999.
- 25 Kršinić, F., Carić, M., Viličić, D., and Ciglencečki, I.: The calanoid copepod *Acartia italica* Steuer,
26 phenomenon in the small saline Lake Rogoznica (Eastern Adriatic Coast), *J. Plankton Res.*,
27 22, 1441-1464, <https://doi.org/10.1093/plankt/22.8.1441>, 2000.



- 1 Leonardos, I., and Sinis, A.: Fish mass mortality in the Etolikon lagoon, Greece: The role of local
2 geology, *Cybiurn*, 21, 201-206, 1997.
- 3 Luther, III, G.W., Ma, S., Trouwborst, R., Grazer, B., Blickley, M., Scarborough, R., and Mensinger,
4 M.: The role of anoxia, H₂S, storm events in fish kills of dead-end canals of Delaware Inland
5 Bays, *Estuaries*, 27, 551-560, <https://doi.org/10.1007/BF02803546>, 2004.
- 6 Ma, S., Whereat, E.B., and Luther III, G.W.: Shift of algal community structure in dead end lagoons
7 of the Delaware Inland Bays during seasonal anoxia, *Aquat. Microb. Ecol.*, 44, 279-290,
8 <https://doi.org/10.3354/ame044279>, 2006.
- 9 Njiru, M., Mkumbo O.C., and Van der Knaap, M.: Some possible factors leading to decline in fish
10 species in Lake Victoria, *Aquat. Ecosyst. Health*, 13, 3-10,
11 <https://doi.org/10.1080/14634980903566253>, 2010.
- 12 Petricoli, D., Balkan-Petricoli, T., Vilicic, D., and Pozar-Domac, A.: Freshwater phytoplankton
13 bloom in Visovac lake-A possible cause of benthic mortality in Krka estuary (Adriatic Sea,
14 Croatia), *Mar. Ecol.*, 17, 373-382, <https://doi.org/10.1111/j.1439-0485.1996.tb00515.x>,
15 1996.
- 16 Rabalais, N. N.: Nitrogen in aquatic ecosystems. Royal Swedish Academy of Sciences, *Ambio*, 31,
17 102-112, <https://doi.org/10.1579/0044-7447-31.2.102>, 2002.
- 18 Rabalais, N.N., Diaz, R.J., Levin, L.A., Turner, R.E., Gilbert, D., and Zhang, J.: Dynamics and
19 distribution of natural and human-caused hypoxia, *Biogeosciences*, 7, 585-619,
20 <https://doi.org/10.5194/bg-7-585-2010>, 2010.
- 21 Rake, A., Pietilainen, O.P., Rekolainen, S., Kauppila, P., Pitkanen, H., Niemi, J., Raateland, A., and
22 Vuorenmaa, J.: Trends of phosphorus, nitrogen and chlorophyll a concentrations in Finnish
23 rivers and lakes in 1975-2000, *Sci. Total Environ.*, 310, 47-59,
24 [https://doi.org/10.1016/S0048-9697\(02\)00622-8](https://doi.org/10.1016/S0048-9697(02)00622-8), 2003.
- 25 Ram, A., Jaiswar, Jaiswar, J.R.M., Rokade M.A. Bharti, S., Vishwasrao, C., and Majithiya, D.:
26 Nutrients, Hypoxia and Mass Fishkill Events in Tapi Estuary, India, *Estuar. Coast. Shelf S.*,
27 148, 48-58, <https://doi.org/10.1016/j.ecss.2014.06.013>, 2014.



- 1 Reissmann J.H., Burchhard, H., Feistel, R., Hagen E., Lass, H.U., Mohrholz, V., Nausch, G., Umlauf,
2 L., and Wicczorek, G.: Vertical mixing in the Baltic Sea and consequences for eutrophication
3 - A review, *Prog. Oceanogr.*, 82, 47-80, <https://doi.org/10.1016/j.pocean.2007.10.004>, 2009.
- 4 Rouso, A., Coluccelli, A., Iermano, I., Falcieri, F., Ravaioli, M., Bortoluzzi, G., Focaccia, P.,
5 Stanghellini, G., Ferrari, C.R., Chiggiato, J., and Deserti, M.: An operational system for
6 forecasting hypoxic events in the northern Adriatic Sea, *Geofizika*, 26, 191-213, 2009.
- 7 Rozan, T.F., Taillefert, M., Trouwborst, R.E., Glazer, B.T., Ma, S., Herszage, J., Valdes, L.M., Price,
8 K.S., and Luther III, G.W.: Iron, sulfur and phosphorus cycling in the sediments of a shallow
9 coastal bay: implications for sediment nutrient release and benthic macroalgal blooms,
10 *Limnol. Oceanogr.*, 47, 1346-1354, <https://doi.org/10.4319/lo.2002.47.5.1346>, 2002.
- 11 Sorokin, Y.I., Sorokin, P.Y., and Pavagnan, G.: On the “bloom” dinoflagellate *Conyaulax tamarensis*
12 in the lagoons of CA Pisani, Veneto, Italy, *Journal of General Biology*, 57, 387-388, 1996a.
- 13 Sorokin, Y.I., Sorokin, P.Y., Pavagnan, G.: On an extremely dense bloom of the dinoflagellate
14 *Alexandrium tamarensis* in lagoons of the Po river delta: Impact on the environment, *J. Sea
15 Res.*, 35, 251-255, [https://doi.org/10.1016/S1385-1101\(96\)90752-2](https://doi.org/10.1016/S1385-1101(96)90752-2), 1996b.
- 16 Sundy, B., Gobeil, Ch., Silverberg, N., and Mucci, A.: The phosphorus cycle in coastal marine
17 sediments, *Limnol. Oceanogr.*, 37, 1129-1145, <https://doi.org/10.4319/lo.1992.37.6.1129>,
18 1992.
- 19 Vollenweider, R.A., Rinaldi, A., and Montanari, G.: Eutrophication, structure and dynamics of a
20 marine coastal system: Results of ten-year monitoring along the Emilia-Romagna coast
21 (Northwest Adriatic Sea), *Sci. Total Environ., Supplement*, 63-106,
22 <https://doi.org/10.1016/B978-0-444-89990-3.50014-6>, 1992.
- 23 Zhang, J., Gilbert, D., Gooday, A.J., Levin, L., Naqvi, S. W. A., Middelburg, J.J. Scranton, M., W.
24 Ekau, W., Pena, A., Dewitte, B., Oguz, T., Monteiro, P.M.S, Urban, E., Rabalais, N.N.,
25 Ittekkot, V., Kemp, W.M., Ulloa, O., Elmgren, R., Escobar-Briones, E., and Van der Plas,
26 A.K.: Natural and human-induced hypoxia and consequences for coastal areas: synthesis and



- 1 future development, Biogeosciences, 7, 1443–1467, [https://doi.org/10.5194/bg-7-1443-](https://doi.org/10.5194/bg-7-1443-2010)
- 2 2010, 2010.