



1 **How deep do we dig for surface soil? A comparison of patterns of**
2 **microbial C:N:P stoichiometry between topsoil and subsoil along an**
3 **aridity gradient**

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22 **Abstract**

23 Microbial stoichiometry and its driving factors play crucial roles in understanding the
24 balance of chemical elements in ecological interactions and nutrient limitations along
25 aridity gradients. However, little is known about the variation in these features along
26 aridity gradients due to the lack of comprehensive field investigations. Additionally,
27 previous studies focused on the surface soil (0-10 or 0-20 cm); however, the minimum
28 sampling depth for surface soil could impact the results of the vertical distribution of
29 microbial stoichiometry. In the present study, we measured the variation in microbial
30 stoichiometry, examined the major influential factors (climatic, edaphic and biotic
31 factors) along an aridity gradient and determined whether the sampling depth affected
32 microbial C:N:P stoichiometry. We found that the microbial C:N (topsoil: 6.59; subsoil:
33 6.83), C:P (topsoil: 60.2; subsoil: 60.5) and N:P ratios (topsoil: 9.29; subsoil: 8.91)
34 varied with soil depth and that the microbial C:N ratio significantly increased with soil
35 depth. The microbial C:N ratio significantly increased with increasing aridity for both
36 topsoil and subsoil, while the microbial N:P ratio decreased along the aridity gradient
37 only for the topsoil, which implied that drought-stimulated microbes tend to be more
38 N conservative, especially for the topsoil. Among all the factors, the soil organic carbon
39 (SOC) content and fungi to bacteria ratio exerted the largest influence on the microbial
40 C:N, C:P and N:P ratios at both soil depths, implying that the substrate supply and
41 microbial structure together controlled the microbial stoichiometry. The results also
42 revealed that the aridity index (AI) and plant aboveground biomass (AGB) influenced



43 the C:N ratio in microbial biomass at both soil depths, and the effects of those factors
44 decreased in the subsoil. The results of this study suggested that the flexibility of the
45 microbial N:P ratio should be considered when establishing the minimum sampling
46 depth in a vertical study. The present study also represented the first attempt to examine
47 the patterns of soil microbial stoichiometry for different soil depth along an aridity
48 gradient.

49 **Keywords:** grassland ecosystem, C:N:P stoichiometry, soil microbial biomass, aridity
50 gradient

51 **1 Introduction**

52 Ecological stoichiometry is a powerful tool for understanding the balance of chemical
53 elements required by organisms and the functions of ecosystems (Elser et al., 2000a;
54 Sterner and Elser, 2002). C, N and P are regarded as critical elements in global
55 biogeochemical cycling, and C:N:P stoichiometry in soil microorganisms offers
56 essential insight into the nutrient limitations of the organisms and communities in an
57 ecosystem (Manzoni et al., 2010; McGroddy et al., 2004).

58 Few studies have addressed the pattern of microbial stoichiometry along latitudinal
59 (Cleveland and Liptzin, 2007; Li et al., 2015; Xu et al., 2013) or environmental
60 gradients (Li et al., 2015; Li and Chen, 2004; Li et al., 2012). For example, Cleveland
61 and Liptzin (2007) analyzed microbial stoichiometry at the global scale and showed an
62 increasing trend in the microbial N:P ratio with latitude. However, Li et al. (2015)



63 summarized the data and found that the microbial N:P ratio decreased with latitude.
64 Undoubtedly, there is uncertainty in the values of microbial C:N, C:P and N:P ratios in
65 global studies, and the variations in these values might have been partially caused by
66 the numerous methods that were used in the various studies (Chen et al., 2016; Xu et
67 al., 2013). Furthermore, less exploration of soil microbial stoichiometry along an aridity
68 gradient at the regional scale impedes our ability to disentangle the trend of the changes
69 in microbial stoichiometry amid climate changes. Climate change, such as global
70 warming, is increasing the degree of aridity in drylands owing to the decreased
71 precipitation and increased evaporation (Li et al., 2017; Wang et al., 2014). Under this
72 background, key ecosystem processes that are regulated by soil microbes, such as soil
73 respiration and nutrient mineralization, may be dramatically impacted by the increased
74 degree of aridity, especially in fragile areas of arid and semiarid ecosystems (Chen et
75 al., 2014; Delgadobaquerizo et al., 2013). Therefore, we conducted a field investigation
76 across a 2100-km climatic transect in the Inner Mongolian grasslands to determine how
77 the microbial C:N, C:P and N:P ratios were affected by changing environmental
78 conditions.

79 Previous studies have also shown that a variety of abiotic factors impact microbial
80 C:N:P stoichiometry (Cleveland and Liptzin, 2007; Hartman and Richardson, 2013;
81 Manzoni et al., 2010). For instance, control experiments have found that warming can
82 indirectly affect the turnover of microbial biomass N by stimulating soil respiration
83 (Veraart et al., 2011; Butterbach-Bahl et al., 2013). Climate could exert an influence on



84 microbial stoichiometry through changes to the microenvironment, such as soil
85 moisture and temperature, and it could also impact the availability of substrates in the
86 soil (Nielsen et al., 2009). Moreover, edaphic variables, such as SOC (Maria et al., 2014;
87 Chen et al., 2016) and soil texture (Li et al., 2015), might be associated with nutrient
88 mineralization and availability, thus influencing the C:N:P stoichiometry in microbial
89 biomass (Griffiths et al., 2012). A labeled incubation experiment showed that the
90 mineralization of organic P was mainly driven by the microbial C demands in P-poor
91 soils (Aponte et al., 2010; Heuck et al., 2015). In addition, microbial C:N, C:P and N:P
92 ratios were also affected by biotic factors such as plant productivity and the composition
93 of the microbial community (Fanin et al., 2013; Chen et al., 2016). Generally, fungi
94 exhibit a higher C:N ratio than bacteria (Michaels and Johannes, 2010); thus, a shift in
95 the fungi to bacteria ratio is expected to result in microbial stoichiometry changes (Li
96 et al., 2012; Heuck et al., 2015). However, those findings were based on literature
97 analyses or small-scale experiments, and the variations in microbial C:N, C:P and N:P
98 ratios at the regional scale have rarely been assessed systematically, and the drivers of
99 these variations need to be addressed more specifically with appropriate experimental
100 designs. Moreover, most research has focused on the top 10 cm of soil, which often has
101 high C availability and nutrient contents. It can be assumed that the effects of driving
102 factors exhibit minimal differentiation at deeper soil depths. However, soil at a deeper
103 depth might contain microbial communities that are specialized for their environment,
104 and their functions may differ from the functions of the communities in the surface soil



105 (Ghiorse and Wilson, 1988; Zvyagintsev, 1994; Fritze et al., 2000; Blume et al., 2002).
106 Certainly, the drivers that are responsible for the variations in microbial C:N, C:P and
107 N:P ratios in deeper soil remain poorly understood. Such knowledge of the nature of
108 soil microbial stoichiometry is fundamental for understanding ecosystem function,
109 especially in soil at deeper depths, which remains highly uncertain under climate
110 change background.

111 Substrates for microorganisms, such as available nutrients and water, decline
112 exponentially with depth, and the top 20 cm of soil accumulates the greatest amount of
113 microbial biomass, thereby attracting the attention of most researchers (Fierer et al.,
114 2003; Xu et al., 2013). Soil at a 0-20 cm depth was regarded as the surface soil in some
115 studies, while other researchers divided the soil from 0-20 cm into different soil depths
116 to explore the vertical differences between these depths (Aponte et al., 2010; Peng and
117 Wang, 2016). However, most studies used 0-10 cm as the surface soil to facilitate
118 sampling and comparative research (Cleveland and Liptzin, 2007; Li and Chen, 2004;
119 Chen et al., 2016). The depth of surface soil varies among studies, and sampling depth
120 can therefore have impacts on the study of the vertical patterns in soil microbial
121 stoichiometry (Tischer et al., 2014). Given that soil represents a highly heterogeneous
122 environment, especially in terms of site-specific soil development history, it is difficult
123 to draw general conclusions (Xu et al., 2013; Camenzind et al., 2018). In addition, if a
124 large difference existed between the soil at 0-10 cm and that at 10-20 cm, microbial
125 stoichiometry would be underestimated due to the ambiguous limitation of surface soil



126 (Tischer et al., 2014). To identify the soil depth that is appropriate for sampling and to
127 improve the understanding of surface soil research at a global scale, we designed a
128 study that divided the surface soil into 0-10 cm and 10-20 cm depths to compare the
129 differences in microbial stoichiometry at the regional scale.

130 In Inner Mongolia grasslands, the aridity exhibits a gradient that increases from
131 northeast to southwest (aridity index ranges from 0.16 to 0.54), and the vegetation types
132 vary from meadows to typical and desert steppes, thus providing an ideal platform to
133 better estimate the patterns and drivers of microbial C:N:P stoichiometry along an
134 aridity gradient (Chen et al., 2014; Li et al., 2017). Moreover, we aim to compare the
135 differences in microbial stoichiometry at different soil depths, especially in the surface
136 soil, to identify the optimal soil depth for vertical studies on microbial stoichiometry.
137 In particular, we tested the following hypotheses: (i) microbial C:N and C:P ratios
138 increase and the microbial N:P ratio decreases across an aridity gradient because of
139 differences in nutrient-use efficiency. (ii) Due to variations in resource supply among
140 different soil depths, the effects of driving factors on microbial C:N, C:P and N:P ratios
141 might decrease with soil depth. Finally, (iii) to adapt to the imbalance of resources,
142 microbial C:N, C:P and N:P ratios vary between soil depths and at a depth of 10 cm,
143 which could influence the research on the vertical patterns of microbial stoichiometry.

144 **2 Materials and methods**

145 **2.1 Study area**



146 This study was performed across the Inner Mongolian temperate grassland, which is a
147 central part of the Eurasian steppe. The study area is located at 39.2-49.6 °N latitude and
148 107.8-120.1 °E longitude and covers an area of 440,000 km². From northeast to
149 southwest, the mean annual temperature increases from -1.7 to 7.7 °C, and the mean
150 annual precipitation decreases from 402 mm to 154 mm (Table. 1). Three grassland
151 types, meadow steppe, typical steppe and desert steppe, are distributed along the
152 northeastern to southwestern gradient and are dominated by *Stipa baicalensis* and
153 *Leymus chinensis*, *S. grandis*, and *S. klemenzii*, respectively (Ma et al., 2010). The soil
154 types corresponding to the three grassland types are categorized as chernozems,
155 kastanozems, and calcisols, respectively, according to the soil classification system of
156 the Food and Agriculture Organization of the United Nations.

157 2.2 Sampling and data collection

158 Along this transect, a total of 58 sites that were slightly disturbed by humans and
159 domestic animals were sampled, including 10 in the meadow steppe, 28 in the typical
160 steppe, and 20 in the desert steppe (Fig. A1). At each site, three replicate soil samples
161 at depths of 0-10 cm and 10-20 cm were collected from three 1 × 1 m subplots arranged
162 diagonally in a 10 × 10 m plot. The samples were mixed to form one composite sample.
163 After gentle homogenization and removal of roots, the soil was sieved through a 2-mm
164 mesh and stored to conduct the following experiments. The total carbon (TC)
165 concentrations were measured using an elemental analyzer (Vario EL III, Elementar,
166 Germany). The soil inorganic carbon (SIC) content was determined with a carbonate



167 content analyzer (Eijelkamp 08.53, Netherlands). The SOC content was calculated by
168 subtracting the SIC from the TC. The soil elemental contents were reported in mmol kg^{-1} .
169 ¹. Soil pH was measured in a suspension with a soil:water ratio of 1:2.5. After the
170 removal of organic matter and carbonates, the soil texture was determined using a
171 particle size analyzer (Malvern Masterizer 2000, UK). The AI was extracted from the
172 CGIAR-CSI Global-Aridity and GlobalPET database (Zorner, Trabucco, Bossio, &
173 Verchot, 2008).

174 **2.3 Soil microbial analyses**

175 Microbial biomass carbon (MBC) and microbial biomass nitrogen (MBN) were
176 determined following the chloroform fumigation- K_2SO_4 extraction method, according
177 to Vance et al. (1987) and Wu et al. (1990). The soil was preincubated at 25 °C for two
178 weeks at a field water capacity of 40%. Then, the soil was fumigated with chloroform
179 for 24 h in a vacuum. The fumigated and nonfumigated samples were extracted using
180 0.5 M K_2SO_4 with a soil:solution ratio of 1:4. The C and N contents were measured
181 with a multi N/C analyzer (Anaytik Jena, Germany). Using a universal conversion
182 factor of 0.45, the amounts of MBC and MBN were calculated by subtracting the
183 amounts of extractable C and N in the nonfumigated samples from those in the
184 fumigated samples. Microbial biomass phosphorus (MBP) was estimated according to
185 the method described in Hedley and Stewart (1982) and modified by Wu et al. (1990).
186 The fumigation procedure was the same as that for MBC and MBN. The fumigated and
187 nonfumigated samples were extracted using $0.5 \text{ mol L}^{-1} \text{ NaHCO}_3$ and were analyzed to



188 determine the TP concentration using a colorimetric method. Using a universal
189 conversion factor of 0.40, the amount of MBP was calculated by subtracting the amount
190 of extractable P in the nonfumigated samples from that in the fumigated samples.
191 Phospholipid fatty acids (PLFAs) were extracted from the soil using the method
192 described by Bossio and Scow (1998). Briefly, 8 g of soil (dry weight) was used for
193 PLFA analysis. After the different phases were dried, the sample was analyzed by
194 capillary gas chromatography following transesterification for quantitative analysis
195 relative to an internal standard. The following PLFAs were used as markers for each of
196 the specific groups: for fungi, 18:1 ω 9c, 18:2 ω 6c, 18:3 ω 6c; for bacteria, i13:0, a13:0,
197 i14:0, i15:0, a15:0, 15:1 ω 6, 2OH16:0, i16:0, 16:1 ω 7c, 16:1 ω 9c, a17:0, i17:0, 17:1 ω 8c,
198 cy17:0, i18:0, 18:1 ω 7, 18:1 ω 5 and cy19:0.

199 **2.4 Statistical analyses**

200 The C:N, C:P and N:P ratios in the soil microbial biomass were log₁₀ transformed
201 before analysis to improve their normality. Paired samples t-tests were used to
202 determine the differences in the soil microbial biomass C, N and P between the topsoil
203 and subsoil and the differences in the C:N:P stoichiometry ratios in the soil microbial
204 biomass. Ordinary least squares regression analyses were conducted to evaluate the
205 relationship between the C:N:P ratios in the soil microbial biomass and latitude, aridity
206 index, AGB, SOC, sand percentage and fungi to bacteria ratio (F:B ratio). The analyses
207 were performed with SPSS 19.0 software (IBM Corporation, Armonk, NY, USA). A
208 structural equation model (SEM) was used to test the multivariate effects (direct and



209 indirect) on the C:N:P ratios in the microbial biomass through hypothetical factor
210 pathways (Fig. S3). The SEM was constructed using the Amos 17.0 software package
211 (Smallwaters Corporation, Chicago, IL, USA).

212 **3 Results**

213 **3.1 The variation in microbial C:N:P stoichiometry between soil depths along** 214 **the environmental gradient**

215 Across all data, the soil microbial biomass C:N, C:P and N:P ratios varied by orders of
216 magnitude and were well constrained (Fig. A2). Distinct water contents, soil bulk
217 density, sand percentages and SOC contents were found between soil depths ($P < 0.05$,
218 Fig. 1a, 1b, 1c, 1f). The microbial biomass C, N and P concentrations in the topsoil
219 were significantly larger than that in the subsoil ($P < 0.05$, Table. 2). The C:N, C:P and
220 N:P ratios in the microbial biomass of the topsoil were 6.59, 60.2, and 9.29, respectively,
221 while those values in the subsoil were 6.83, 60.5 and 8.91, respectively (Table. 2).
222 Moreover, the microbial C:N ratio in the subsoil was significantly higher than that in
223 the topsoil (Fig. 2b).

224 Significant positive relationships were found between the microbial C:N ratio and
225 latitude (Topsoil, $R^2 = 0.14$, $P < 0.01$; Subsoil, $R^2 = 0.12$, $P < 0.05$, Fig. 2a), while a
226 negative relationship was found between the microbial N:P ratio and latitude (Topsoil,
227 $R^2 = 0.18$, $P < 0.001$; Fig. 2c). The results revealed a significant positive relationship
228 between the AI and the microbial C:N ratio (Topsoil, $R^2 = 0.10$, $P < 0.05$; Subsoil, R^2



229 =0.09, $P < 0.05$, Fig. 2d) and significant negative relationships between the AI and the
230 microbial N:P ratio (Topsoil, $R^2 = 0.10$, $P < 0.05$; Fig. 2f). The microbial C:N ratio was
231 positively related to AGB (Topsoil, $R^2 = 0.06$, $P < 0.05$, Fig. 3a), SOC (Topsoil, $R^2 = 0.12$,
232 $P < 0.01$; Subsoil, $R^2 = 0.09$, $P < 0.05$, Fig. 3d) and was negatively related to the sand
233 percentage (Topsoil, $R^2 = 0.11$, $P < 0.01$; Subsoil, $R^2 = 0.11$, $P < 0.01$, Fig. 3g). A
234 significant positive relationship was found between the microbial C:P ratio and the
235 content of soil organic matter (Subsoil, $R^2 = 0.08$, $P < 0.06$, Fig. 3e). No or weak
236 associations were found between the microbial C:N, C:P and N:P ratios and the
237 aforementioned ecological factors in the subsoil (Fig. 3).

238 **3.2 Effects of driving factors on the microbial C:N:P stoichiometry at different soil** 239 **depths**

240 The final SEM adequately fit the data, as shown by several robust goodness-of-fit
241 measures (P value and minimum discrepancy). The model explained 38% (topsoil) and
242 27% (subsoil) of the variation in the microbial C:N ratio, 17% and 19% of that in the
243 microbial C:P ratio, and 29% and 16% of that in the microbial N:P ratio (Fig. 4a, b, c,
244 d). Effects of AI, AGB, SOC and the F:B ratio on the microbial C:N ratio were found
245 at both soil depths (Fig. 4a, 4b). The SOC made the largest positive contribution to the
246 variation in the microbial C:N ratio in the topsoil (Fig. 4a, 4b). We found direct effects
247 of the sand percentage, SOC and F:B ratio on the microbial C:P ratio at both soil depths,
248 and the SOC content made the largest contribution to the variation in the microbial C:P
249 ratio in the topsoil, which was higher than that in the subsoil (Fig. 4e, 4f). Influences of



250 sand% and the SOC content on the microbial N:P ratio were found in the topsoil, while
251 the F:B ratio and the SOC explained most of the variation in the microbial N:P ratio in
252 the subsoil (Fig. 4e, 4f).

253 **4 Discussion**

254 **4.1 The pattern of microbial C:N, C:P and N:P ratios along latitude and the** 255 **aridity gradient**

256 The microbial C:N ratio demonstrated a significant increase with increasing latitude, in
257 contrast to the decrease that was demonstrated for the microbial N:P ratio. Such results
258 paralleled the results of studies on ecological stoichiometry, which revealed that the
259 C:N ratio of microorganisms increased with latitude, while the N:P ratio decreased with
260 latitude, suggesting increasing N limitations in microorganism ecosystems in high-
261 latitude areas (Li et al., 2015; Chen et al., 2016). The regional-scale microbial
262 stoichiometry followed the global-scale stoichiometry patterns that were observed for
263 plant leaves (Reich and Oleksyn, 2004; Yuan et al., 2011), litter (McGroddy et al.,
264 2004), and soil (Sardans et al., 2012), conforming to the substrate age hypothesis, which
265 predicts young soils to be N-limited, whereas old soils tend to be P limited (Walker and
266 Syers, 1976; Vitousek et al., 2010). As stated in our hypothesis, we observed that the
267 microbial C:N ratio significantly increased with increasing aridity, while the microbial
268 N:P ratio decreased along the aridity gradient, indicating that increasing aridity affects
269 ecological stoichiometry by mediating the growth rate of microorganisms in semiarid



270 regions (Elser et al., 2000b; Peng and Wang, 2016). Dan and Wang noted that
271 increasing aridity reduced the soil microbial abundance in drylands, and a decreased
272 growth rate in dry areas might result in decreased allocation to P-rich ribosomal RNA
273 (and thus higher C:P and N:P ratios) (Wang et al., 2014; Maestre et al., 2015).
274 Additionally, drought decreased the C:N ratio in microbial biomass, which serves as a
275 protective mechanism as microbes decrease their nitrogen use efficiency (NUE, the
276 ratio of N invested in growth over total N uptake) and tend to be more N conservative
277 under dry climatic conditions (Mooshammer et al., 2014; Delgado-Baquerizo et al.,
278 2017). Moreover, under dry climate conditions, the soil microbial communities shift
279 from r-strategists (fast-growing copiotrophs) to K-strategists (slow-growing
280 oligotrophs), as microorganisms with K-strategies have lower nutrient demands (N and
281 P) and growth rates, invest more nutrients into extracellular enzymes to gain limited
282 nutrients and thus have higher cellular C:N:P ratios than r-strategists (Fierer et al., 2007;
283 Fierer et al., 2010). Our study further illustrated the latitudinal pattern of microbial
284 stoichiometry and first attempted to examine the variation in microbial stoichiometry
285 along an aridity gradient at the regional scale.

286 **4.2 Direct effects of ecological factors on controlling microbial C:N, C:P and N:P** 287 **ratios at different soil depths**

288 Among the ecological factors examined, our study found that the patterns of microbial
289 C:N and C:P ratios were associated with SOC and the F:B ratio, suggesting that the
290 available C and microbial community structure together regulated the shift in microbial



291 stoichiometry. If the environmental parameters were considered individually, SOC was
292 found to be significantly positively related to the microbial C:N and C:P ratios, which
293 is consistent with the results observed from global data analysis, suggesting that SOC
294 may control microbial stoichiometry by mediating the substrate stoichiometry, e.g., the
295 soil C:N and C:P ratios (Hartman and Richardson, 2013; Maria et al., 2014;
296 Mooshammer et al., 2014). In deeper soil, microbial metabolic processes are limited by
297 C availability and energy (C), such as denitrification and P mineralization (Fierer et al.,
298 2003; Peng and Wang, 2016; Camenzind et al., 2018). SEM also illustrated that the
299 microbial community structure is an important feature in determining microbial
300 stoichiometry. The F:B ratio has recently been found to have a vital influence on the
301 patterns of microbial C:N and N:P ratios in soil at a large scale (Chen et al., 2016). An
302 experiment indicated that fungi have lower resource requirements and higher C:N and
303 C:P ratios than bacteria, and thus microbial C:N:P stoichiometry impacted the microbial
304 community structure as a result of the F:B ratio (Mouginot et al., 2014). In our study,
305 the lower F:B ratio might have led to a shift in the microbial nutrient stoichiometry at
306 deeper soil depths (Tischer et al., 2014). Overall, the SEM highlighted the important
307 role of the C availability and microbial community structure in driving the variations in
308 microbial C:N, C:P and N:P ratios at both soil depths.

309 Moreover, AGB and AI also exerted direct influences on microbial C:N or C:P ratios,
310 and those impacts mainly acted in the topsoil but were weaker in the subsoil, supporting
311 our hypothesis. The climate imposes important controls on both the plant community



312 and the microbial taxa along with their interactions with soil nutrients (Chadwick et al.
313 1999; Vitousek 2004; Oleksyn 2004). At the same time, the meadow steppe ecosystem
314 with high productivity maintained relatively high soil C and N contents, which resulted
315 in high C:P and N:P ratios in these regions; thus, plant productivity exerted a positive
316 influence on microbial C:N (Aponte et al., 2010; Manzoni et al., 2010). Because of the
317 vertical distribution of those influences, the effects might decrease with soil depth.

318 Interestingly, our results revealed that the microbial N:P ratio was mainly impacted by
319 the F:B ratio and SOC, while the sand percentage and SOC had direct negative effects
320 on the ratio in the subsoil, suggesting the flexibility of microbial stoichiometry in
321 response to distinct resource supplies between soil depths (Peng and Wang, 2016). The
322 soil depth affected the microbial biomass N and P, which decreased nearly twofold from
323 the topsoil to the subsoil (Table 2). However, the results showed that the N and P cycles
324 responded asymmetrically to soil depth, which might be attributed to the high
325 variability in P availability (Li et al., 2015; Zechmeister-Boltenstern et al., 2016).
326 Generally, P is mostly derived from parent material, while N is mainly a biological
327 element (Vitousek and Farrington, 1997; Vitousek et al., 2010). Therefore, it is believed
328 that P variations regulate large-scale patterns in microbial N:P stoichiometry and
329 nutrient-use strategies (Camenzind et al., 2018; Heuck et al., 2015). With a high
330 proportion of sand, the soil becomes porous, which may lead to increased leaching of
331 available P to deeper soil depths (Otten et al., 1999; Achbergerová and Nahálka, 2011).
332 Similarly, P leaching caused by weathering led to a shift in the N:P ratio in the soil, and



333 a vertical study found a high variation in the N:P ratio between soil depths across a
334 large scale. The high spatial heterogeneity of the N:P ratio in soil and soil microbial
335 biomass therefore indicates that the N:P ratio could be an indicator of the ecosystem
336 nutrient status at deeper soil depths.

337 **4.3 How deep should we dig to evaluate the microbial stoichiometry in surface**
338 **soil?**

339 The results showed significant differences in the sand percentage, SOC content and F:B
340 ratio between soil depths, suggesting that the resource supplies between topsoil and
341 subsoil were distinct. We also observed that the microbial C:N, C:P and N:P ratios
342 varied between soil depths, indicating the flexibility of the microbial community in
343 response to distinct resource supplies between soil depths (Tian et al., 2010; Peng and
344 Wang, 2016). Similar findings were found in the top 16 cm of soil in a Mediterranean
345 oak forest (0-8 cm and 8-16 cm), where the microbial nutrient ratios (C:N, C:P and N:P)
346 varied with soil depth (Aponte et al., 2010). Tischer et al. (2014) sampled the top 20
347 cm of soil (0-5 cm, 5-10 cm, 10-20 cm) and observed that the microbial C:N ratio
348 changed with soil depth. Moreover, sampling to a depth of 10 cm showed a significant
349 difference in the microbial N:P ratio (Tischer et al., 2014). The detection of the
350 differences in the microbial N:P ratio in our study depended strongly on the sampling
351 depth, suggesting that the microbial N:P ratio might provide insight into the nature of
352 ecosystem nutrient limitations in a vertical study (Cleveland and Liptzin, 2007; Fierer
353 et al., 2010). In addition, SEM also showed that the microbial N:P ratio was controlled



354 by multiple driving factors at different soil depths, indicating that a 0-10 cm or
355 shallower sampling interval should be used when studying the vertical patterns in the
356 microbial N:P ratio.

357 **5 Uncertainties and perspectives**

358 The first uncertainty was related to the determination of fungal and bacterial biomasses
359 by PLFA markers, which have limited targets for fungi and bacteria. This uncertainty
360 should be noted when interpreting the results in the present study. Methodological
361 advances in sequencing approaches might be used to more accurately index the
362 microbial community and reveal insights into the regulation of microbial C:N:P
363 stoichiometry in distinct soil microbial taxa or functional groups. Second, the theory
364 used to construct the model is another source of uncertainty; the theory related to the
365 drivers of microbial stoichiometry used in this study was mostly derived from a
366 literature review and summarized data. In future research, more control experiments
367 with the manipulation of C availability, especially at deeper soil depths, would further
368 improve our understanding of the changes in microbial stoichiometry and nutrient
369 limitations under the impacts of global change.

370 **6 Conclusion**

371 The ratios of C, N, and P in the microbial biomass were 6.59:60.2:9.29 in the topsoil,
372 which deviated from the 6.83:60.5:8.91 ratio in the subsoil. Moreover, significant
373 difference was found in the microbial C:N ratio between topsoil and subsoil, indicating



374 that the flexibility of microbial stoichiometry should be considered for vertical study.

375 In addition, drought decreased the C:N ratio in microbial biomass, consistent with the

376 perspective that microbes mediate their nitrogen use efficiency and tend to be more N

377 conservative under drier climatic conditions. The microbial N:P ratio trend along the

378 aridity gradient was consistent with the growth rate hypothesis that a decreased growth

379 rate in dry areas results in decreased allocation to P-rich ribosomal RNA and thus a

380 higher N:P ratio. These findings confirmed the importance of SOC, the microbial

381 structure and soil texture in shaping the pattern of microbial stoichiometry in semiarid

382 grassland systems. The influence of ecological factors decreased from topsoil to subsoil,

383 as well as the decline in climatic, edaphic and biotic factors. Overall, these results

384 illustrated N and P limitation in microbial biomass at deeper soil depths along aridity

385 gradient and limited responses to ecological factors in the subsoil.

386 *Author contributions.* HH, WM and YY devised the study. YL carried out the

387 experiment and data analyses. DK, YC and DC assisted with the data analyses and

388 interpretation. XN, TW, XZ, MZ and HB assisted with the experiment. All authors

389 contributed to the preparation of the paper.

390 *Competing interests.* The authors declare that they have no conflict of interest.

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396

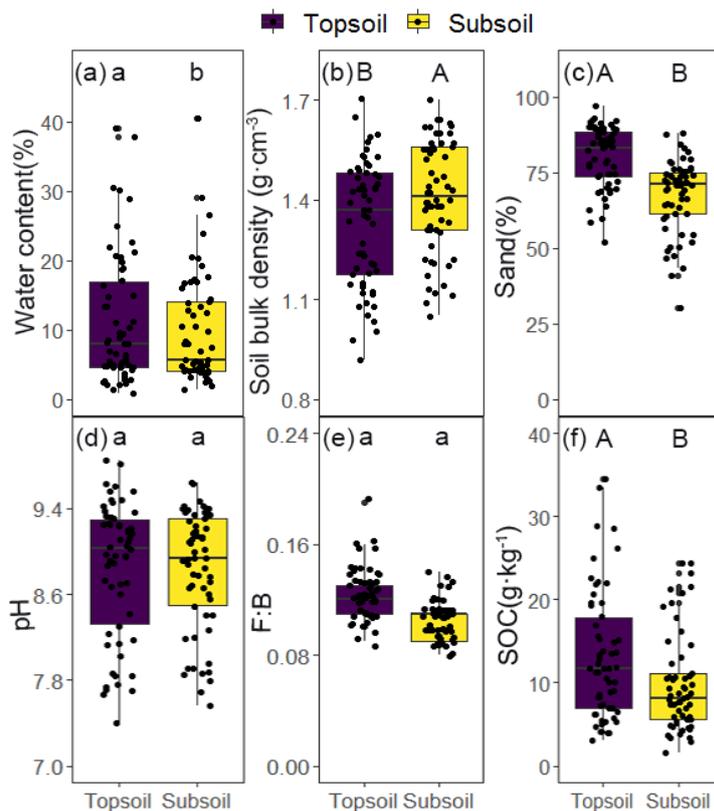
Table 1. Basic information of study sites

Biome	Latitude (°N)	Longitude (°E)	MAP (mm)	MAT (°C)	Aridity Index	AGB (g · cm ⁻³)	Dominant species
MS	48.1(43.9–49.6)	119(116–120)	353(262–381)	–0.45(–1.81–1.71)	0.48 (0.38–0.54)	136(88–168)	<i>Stipa baicalensis</i>
							<i>Leymus chinensis</i>
							<i>Stipa grandis</i>
TS	45.6(43.5–49.5)	117(114–119)	304 (205–402)	1.11(–2.09–3.29)	0.37 (0.25–0.50)	102(49.4–159.8)	<i>Stipa klemenzii</i>
							<i>Stipa breviflora</i>
DS	41.7(39.2–43.6)	115(108–113)	223(154–293)	5.63(4.13–7.67)	0.23 (0.16–0.32)	43.4(24.6–76.5)	

397

Note: Data represent the means, with minimum and maximum values in parentheses. MS, meadow steppe; TS, typical steppe; DS, desert steppe. ☺

398



399

400 **Figure 1.** Basic characteristics of study sites across the inner Mongolia grassland at
401 different soil depths. Different letters indicate significant differences between soil
402 depths on log₁₀-transformed data (paired t-test, lower letter, P<0.05; upper case letter,
403 P<0.001). MS, meadow steppe; TS, typical steppe; DS, desert steppe.



404 **Table 2.** The microbial biomass C, N and P concentrations and microbial C:N:P stoichiometric ratios across the Inner Mongolian grassland at

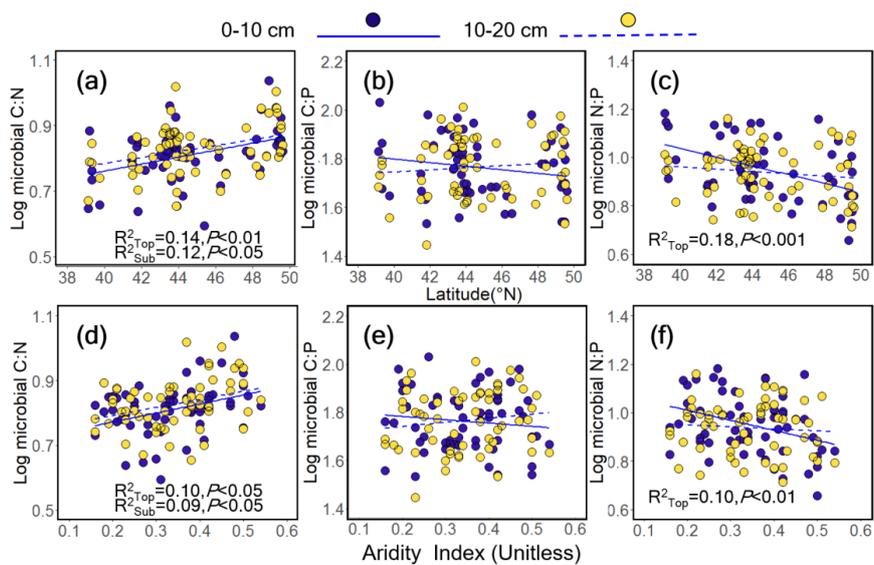
405

Biome	MBC	MBN	MBP	C:N	Microbial biomass	
	(mmol · kg ⁻¹)	(mmol · kg ⁻¹)	(mmol · kg ⁻¹)		C:P	N:P
0-10 cm	21.8(18.5-25.1)A	3.23(2.80-3.67)A	0.38(0.32-0.44)A	6.59(6.26-6.91)a	60.2(55.6-64.8)a	9.29(8.0-9.97)a
10-20 cm	14.5(12.4-16.6)B	2.08(1.81-2.35)B	0.24(0.21-0.27)B	6.83(6.50-7.15)b	60.5(56.0-65.1)a	8.91(8.35-9.49)a

407 Note *: Different letters indicate significant differences between soil depths based on log10-transformed data (paired t-test, lower-case letters,

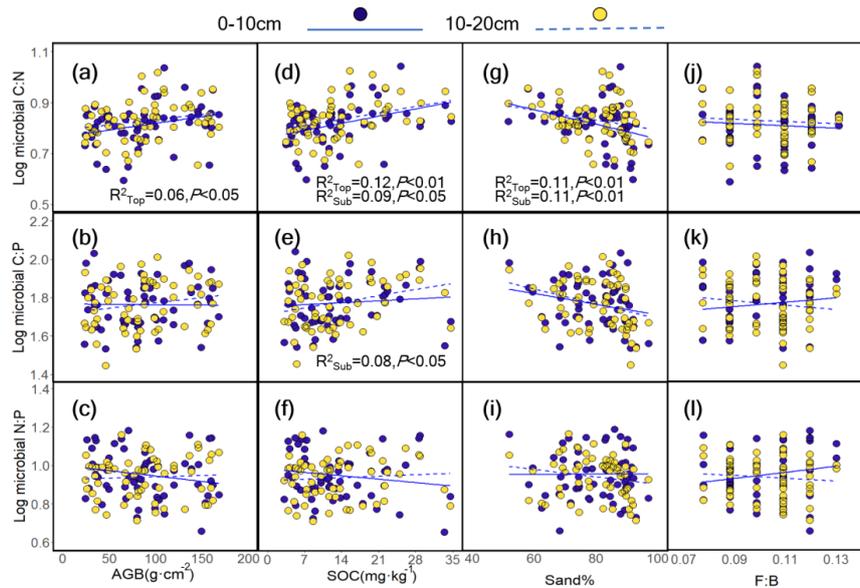
408 P<0.05; upper-case letters, P<0.001). MS, meadow steppe; TS, typical steppe; DS, desert steppe.

409



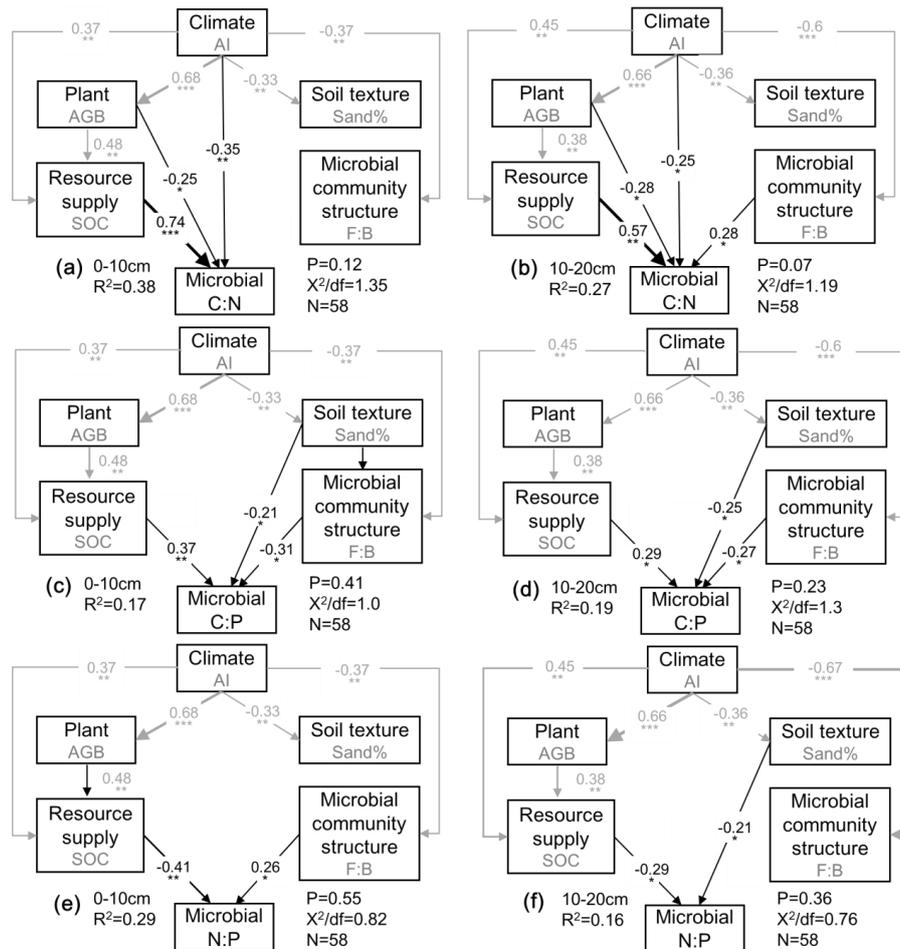
410

411 **Figure 2.** Relationships between the C:N, C:P and N:P ratios in soil microbial
412 biomass and latitude (a-c) and aridity index (d-f) in the Inner Mongolian grassland.



413

414 **Figure 3.** Relationships between the C:N, C:P and N:P ratios in the soil microbial
415 biomass and AGB (a-c), SOC (d-f), sand percentage (g-i) and F:B ratio (j-l). AGB,
416 above ground biomass; SOC, soil organic carbon; F:B ratio, fungi to bacteria ratio.

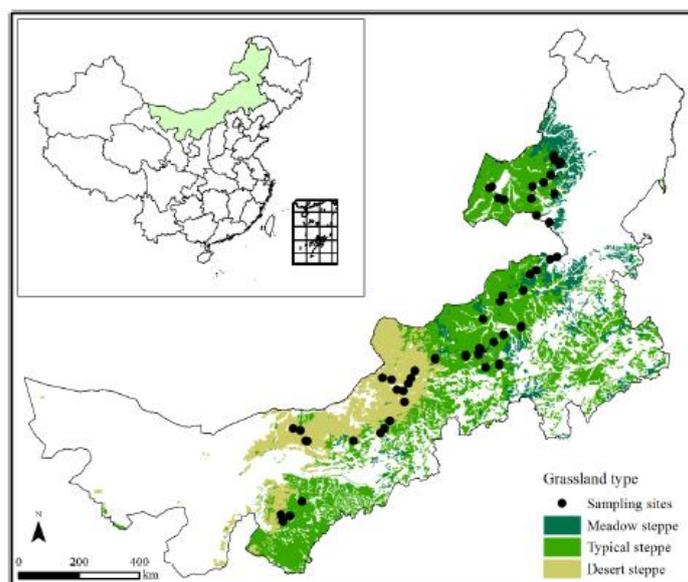


417

418 **Figure 4.** The structural equation model (SEM) shows the direct and indirect influences
 419 of various ecological factors on the microbial C:N (a,b), C:P (c,d) and N:P (e-f) ratios
 420 in the topsoil and subsoil. Black and gray arrows indicate direct and indirect pathways,
 421 respectively. Numbers on the arrows indicate standardized path coefficients,
 422 proportional to the arrow width. R^2 indicates the variation of the microbial C:N and C:P
 423 ratios explained by the model. *, $P<0.05$; **, $P<0.01$; ***, $P<0.001$.

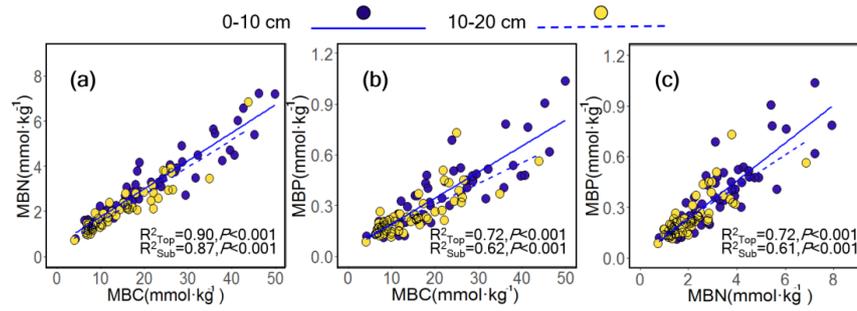


424 **Appendix**



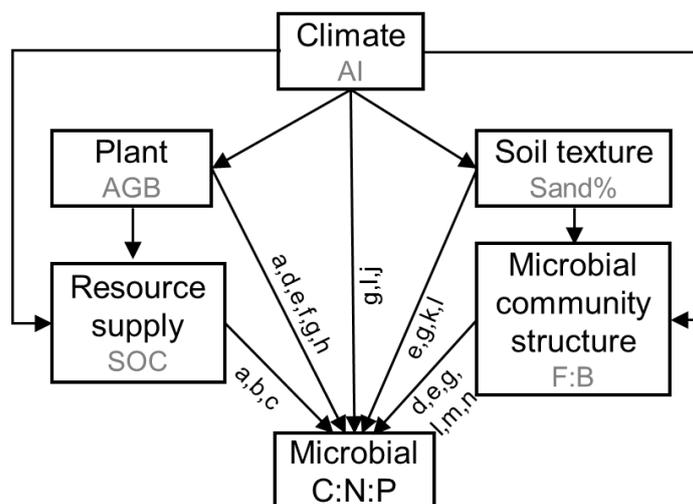
425

426 **Figure A1.** Geographic locations of the sampling sites in the Inner Mongolian grassland



427

428 **Figure A2.** Relationships between the soil microbial biomass C, N and P concentrations



429

430 **Figure A3.** Hypothetical model showing how ecological factors affect microbial C:N:P
 431 stoichiometry

432 **Table A1.** The references to support hypothetical models

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