Response letter 26th june 2020

We highly value the comments and suggestions of Reviewers and we have improved the manuscript accordingly. We hope you find the answers and the revised version appropriate. Here below are our answers to the comments and manuscript with changes

Associate Editor Decision: Publish subject to minor revisions (review by editor) (04 Jun 2020) by Michael Bahn

Comments to the Author:

Dear authors,

both reviewers think that their concerns have been largely addressed, even though some of their suggestions for additional relevant testing were not taken up by you. The reviewers have provided a number of excellent specific suggestions for further improvement, which I kindly ask you to consider during this next round of revisions. Please include also a more critical assessment of the limitations of the current model version.

Best regards,

Michael Bahn

-*We have now added a critical assessment of the limitations of the current model version to the end of the discussion*

Referee 1

This is my second review of this manuscript for Biogeosciences and I am pleased that the authors incorporated many of the suggestions made by me and the other reviewer. In brief, the strength of the study emerges from their combining an individual-based, two species competition model of Sphagnum community interactions with underlying functional models of water and carbon dynamics and linking it to a hydrology simulation model. Historically, most modeling efforts have focused on understanding and modeling the mechanisms of competition and function in Sphagnum. One of the highlights of this study is the connection to an existing hydrology and surface exchange model that was used to generate local environmental conditions that served as forcing variables for the simulation. Overall, I believe that this manuscript represents a valuable step in developing predictive models of peatland function.

I still have a couple of issues that I think the authors ought to address and then provide a list of minor edits.

A. There are still many places where “water retention” and “water content” are used

synonymously. In my mind, retention is when existing water is not released via evaporation or drainage. I don't think this is the meaning the authors intend. Most of the times, I think the authors mean capitulum "water content". This incorporates fluxes in and out, but also the capacity for storage. I think the language needs to be clarified and it begins with its usage in the title. Here is a list of places that I found places where this should be clarified: L2, L28, L35, L75, and L522; there may be others.

-*Changed as suggested.*

B. I remain surprised that the authors still did not show sample light response curves in the Appendix, which were the source of many of their photosynthetic parameters; both reviewers questioned their measurements, especially with regards to the long time for samples to desiccate within the chamber. In the appendix, data for the photosynthetic—water content relationships are shown (Fig B2). I think light response curves should be added. They may be very useful if this study is used for comparative purposes in the future.

-*The measured light response curves for S. magellanicum and S. fallax are now added to Appendix B as Figure B1.*

Minor edits

C. L15: what do you mean by "dynamic community structure"? There should be a better way to state this.

-*Clarified in lines 15-19 to "Current peatland models generally treat vegetation as a static community, although plant community structure is known to alter as a response to environmental change. Because vegetation structure and ecosystem functioning are tightly linked, realistic projections of peatland response to climate change requires including vegetation dynamics in ecosystem models."*

D. L25. The description of what PMS does should be more clear. Perhaps it would be better worded by placing competition up front. "PMS employs a, stochastic, individual-based approach to simulating competition based on species' differences in functional traits."

Changed in lines 25-30 into: In this study, we developed the Peatland Moss Simulator (PMS), simulating community dynamics of the peatland moss layer. PMS is a process-based model that employs a stochastic, individual-based approach simulating competition within peatland moss layer based on species differences in functional traits.

E. L35: replace "retention" with "relations"

-*changed*

F. L53: unclear what "dynamic community structure" relates to. "lack mechanisms that underlie and cause dynamic community processes"?

-*Clarified. Now says "Peatland models have generally considered vegetation structure unrealistically as static component*"

G. L55: delete "in order"

-deleted

H. L55: delete after “and the research community”
-deleted

I. L73: replace “the peers” with “its peers”.
-changed

J. L75: and water storage
-added

K. L87: I think the ability to carry out such a quantitative study is not new and don’t believe this has been a great hindrance.
-deleted

L. L93: occur to occurs
-changed

M. L98: rely to relies
-changed

N. L114: replace “locates” to “is located”*-changed*
O. L118: to “10% of the surface is occupied….”
-changed

P. L128: to “flow with…”
-changed

Q. L143: “race for space” to competition
-changed

R. L414: change PCS to PMS
-changed

S. L491: lawnss to lawns
-changed
T. L535: delete the
-deleted

U. L573 and 577: replace dynamical with dynamic
-changed

Referee 2

Dear authors,

Thank you for addressing the concerns raised by me and the other reviewer. Several points have now been clarified in the text (please do check the grammar, which is not flawless in the added sections). There are a few remarks that you have responded to only in your reply to us but not in the main text (e.g. my previous points 3b (horizontal water exchange), point C and D of reviewer 3 (autogenic processes leading to hummock formation and differences in moss hydraulic conductivity)). I think those are missed chances of improving the outlook section for your model. In general, I would appreciate a more explicit acknowledgment of the limitations of your model. After all, even if you managed to recreate some ´realistic´ patterns, several potentially important processes are still missing from the model, so you cannot be sure whether you produced these patterns for the right reasons. It is absolutely fine to start with a simple model (even if your main purpose is to `illustrate the reality´ L188 in your response) and to ´leave perfection for later´ (L70 in your response), but thereby it is helpful to line out the path to perfection (well, at least to a model in which the importance of additional processes has been tested) for the readers.

-*We have now included a section in the discussion that focus on the limitation of our model, in lines 588-617. We had that in the first version of the manuscript but is was accidentally left out when we combined the two manuscripts (empirical and modelling) into one and tried to make it concise*.

Also, I think there are two points (6 and 7) that I think you may have misunderstood, so that I will try to formulate them better here. I pasted my old remark and your reply in here to keep track of the context.

My previous point 6 and your reply:

As an important difference between your and previous models lies in the coupling to environmental fluctuations and stochasticity (L97-98), it would make sense to present a test of the importance of these processes to the model output. Would a simpler model provide similarly good results?

R: We believe that the main purpose of modelling is to illustrate the reality and serve as a tool for systematic assessment of the processes. Simple community models without individual-based processes implicitly weigh on generality and forgive outliers. However, environmental fluctuation and extremes are becoming more frequent and intensive with climate change, and this is likely to give advantage to an otherwise unlikely change in peatland community. To help with this situation, our modelling is able to populate outputs along a probability distribution and allows assessing individuals with different trait combinations as a part of the probabilities. As these models are fundamentally different in focuses and underlying mechanisms, simply comparing the goodness of results seems pointless.

My new comment: In this case I was not suggesting that you compare your model to previous, unrelated, models, but that you do tests with your own model, simplifying some processes and seeing if that degrades the results. E.g. instead of using realistic parameter distributions just use a

random number generator to e.g. modify the length growth of individual shoots (grid cells). Instead of using realistic environmental fluctuations just use the smoothed mean monthly climate. These are just some examples, I am sure you can think of better ones.

- *We added Test 9-10 following the suggestions. In Test 9, we used smoothed monthly mean meteorological data to drive the model simulations. In Test 10, we eliminated the stochasticity in all model parameters and used only the mean values for simulation.*

Continuation of my previous comment: I would also be interested in seeing the effects of the water retention and photosynthetic water-response parameters separately. Especially since the parameters for the latter may suffer from some measurement artefacts.

R: This is a very appreciated comment. Our future goal is also to make the picture clearer and understanding the factorial effects is a very important aspect. At the moment, our data and techniques are insufficient to separate the different effects. Therefore, model testing based on the parameters quantified by the "mixed" information could be less informative, unless we have had improved measurement data.

In addition, S. fallax and S. magellanicum are largely different in both water retention and photosynthetic response to water stress. Further testing on species either with similar water retention, or with similar photosynthetic response would be more informative to this question.

My new comment: In my mind, a model is the perfect opportunity to pretend that your species are not different in both but just in one or the other aspect and to test the individual effects of these parameters. You would not have real data to validate the result, but that is not the point here. The point is to understand what these parameters do and what would happen with hypothetical species with these parameter combinations. For me, this type of test is what would constitute a `systematic assessment of the processes´ (L189 in your response).

-*Based on the suggestion, we simulated the mean cover of the two moss species by setting photosynthetic water-response parameters to be the same for both species but keep the water retention effects different (Test 7-8)*

My new comment regarding my previous point 7: I did not mean to say that module III is unimportant, but it is not evaluated in this paper (just used to create input data for the presented model modules), so it seems too much (therefore unbalanced) to spend several pages on explaining the details of module III. I would recommend moving this information to the supplement.
- *Moved to the supplements (as a part of Supplement A)*

Additionally: Table 4, the sensitivity analysis: how do these 10% changes relate to the actual uncertainty in the parameter values?

-We used the 10% changes to test the parameter sensitivity, as we have limited information on the actual values of the parameters regarding to our site. Indeed, a change of 10% may not be comparable to the actual uncertainty of some parameter values (e.g. hydraulic conductivity of peat may vary by several orders, depending on the peat quality). However, the test for parameter sensitivity here aims to reveal the robustness of model at "current" states, rather than to investigate scenarios with different actual variations.

Modelling the habitat preference of two key *Sphagnum* species in a poor fen as controlled by capitulum water ~~retention~~content

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Abstract

Current peatland models generally treat ~~lack dynamic plant community structure~~vegetation as static, although plant community structure is known to alter as a response to environmental change. Because the vegetation ~~dynamics~~ structure and ecosystem functioning are tightly linked~~. R~~ realistic projections of peatland response to climate change require~~s~~ including vegetation dynamics in ecosystem models. In peatlands, *Sphagnum* mosses are key engineers. M~~The m~~oss community composition primarily follows habitat moisture conditions. The species known preference along the prevailing moisture gradient might not directly serve as a reliable predictor for future species compositions as water table fluctuation is likely to increase. Hence, modelling the mechanisms that control the habitat preference of *Sphagna* is a good first step for modelling ~~the~~ community dynamics in peatlands. In this study, we developed the Peatland Moss Simulator (PMS), simulating community dynamics of the peatland moss layer~~a process-based model, for simulating community dynamics of the peatland moss layer that results in habitat preferences of *Sphagnum* species along moisture gradient.~~ PMS is a process-based model that employs a stochastic, ~~n~~ individual-based approach simulating competition within peatland moss layer based on species differences in~~to describe the variation of~~ functional traits~~ among shoots and the stochastic base of competition~~. At the shoot-level, growth and competition were driven by net photosynthesis, which was regulated by hydrological processes via capitulum water ~~retention~~content. The model was tested by predicting the habitat preferences of *S. magellanicum* and *S. fallax*, two key species representing dry (hummock) and wet (lawn) habitats in a poor fen peatland (Lakkasuo, Finland). PMS successfully captured the habitat preferences of the two

Sphagnum species, based on observed variations in trait properties. Our model simulation further showed that the validity of PMS depended on the interspecific differences in capitulum water ~~retention~~ content being correctly specified. Neglecting the water ~~retention~~ content differences led to the failure of PMS to predict the habitat preferences of the species in stochastic simulations. Our work highlights the importance of capitulum water content ~~retention~~ to the dynamics and carbon functioning of *Sphagnum* communities in peatland ecosystems. Studies of peatland responses to changing environmental conditions thus need to include capitulum water processes as a control on ~~the~~ moss community ~~vegetation~~ dynamics. ~~For that o~~Our PMS model could be used as an elemental design for the future development of dynamic vegetation models for peatland ecosystems.

Keywords: *Sphagnum* moss; capitulum water content; competition; peatland community dynamics; process-based modelling; moss traits; Peatland Moss Simulator (PMS)

1.Introduction

Peatlands have important roles in the global carbon cycle as they store about 30% of the world's soil carbon (Gorham, 1991; Hugelius et al., 2013). Environmental changes, like climate warming and land-use changes, are expected to impact the carbon functioning of peatland ecosystems (Tahvanainen, 2011). Predicting the functioning of peatlands under environmental changes requires models to quantify the interactions among ecohydrological, ecophysiological and biogeochemical processes. These processes are known to be strongly regulated by vegetation (Riutta et al. 2007; Wu and Roulet, 2014), which can change ~~during~~ over decadal time scales~~frame~~ under changing hydrological conditions (Tahvanainen, 2011). ~~Current~~ Ppeatland models have generally considered vegetation structure unrealistically as static component ~~lack mechanisms for the dynamic plant community structure~~ (e.g. Frolking et al., 2002; Wania et al., 2009)~~ that is unrealistic~~. The recent regional-scale peatland model developed by Chaudhary et al. (2017) includes dynamic vegetation shifts among a single moss plant functional type (PFT) and four vascular PFTs~~.~~ ~~–~~but to support realistic predictions on peatland functioning and global biogeochemical cycles~~Therefore,~~ th~~os~~e mechanisms that drive changes in ~~plant~~moss community structure need to be identified and integrated with ecosystem processes~~.~~, ~~in order to support realistic predictions on peatland functioning and the research community working on global biogeochemical cycles.~~

A major fraction of peatland biomass is formed by *Sphagnum* mosses (Hayward and Clymo, 1983; Vitt, 2000). Although individual *Sphagnum* species often have narrow habitat niches (Johnson et al., 2015), different *Sphagnum* species replace each other along water table gradient and therefore, as a genus, spread across a wide range of water table conditions (Rydin and

McDonald, 1985; Andrus et al. 1986; Rydin, 1993; Laine et al. 2009). The species composition of the *Sphagnum* community strongly affects ecosystem processes such as carbon sequestration and peat formation through interspecific variability in species traits such as photosynthetic potential and ~~–~~litter quality (Clymo, 1970; O'Neill, 2000; Vitt, 2000; Turetsky, 2003). The *Sphagnum* ~~production of~~ biomass and litter production ~~from *Sphagna*, which~~ gradually raises the moss carpet, which ~~,~~ feeds back into ~~in turn affects~~ the species composition (Robroek et al. 2009). Hence, modelling the moss community dynamics is fundamental for predicting temporal changes of peatland vegetation. As the distribution of *Sphagnum* species primarily follows the variability in peatland water table ~~in a peatland~~ (Andrus 1986; Väliranta et al. 2007), modelling the habitat preference of *Sphagnum* species along a moisture gradient could be a good first step for predicting moss community dynamics ~~in peatland ecosystems~~ (Blois et al., 2013).

For a given *Sphagnum* species, the optimal habitat represents the environmental conditions for it to achieve higher rates of net photosynthesis and shoot elongation than ~~the~~ its peers (Titus & Wagner, 1984; Rydin & McDonald, 1985; Rydin, 1997; Robroek et al., 2007a; Keuper et al., 2011). Capitulum water content and water storage, which is determined by the balance between the evaporative loss and water gains from capillary rise and precipitation, represents one of the most important controls on net photosynthesis (Titus & Wagner, 1984; Murray et al. 1989; Van Gaalen et al. 2007; Robroek et al., 2009). To quantify the water processes in mosses, hydrological models have been developed to simulate the water movement between moss carpet and the peat underneath ~~(Price, 2008; Price and Waddington, 2010)~~, as regulated by the variations in meteorological conditions and energy balance (Price, 2008; Price and Waddington, 2010). On the other hand, experimental work has addressed the species-specific responses of net photosynthesis to changes in capitulum water content (Titus & Wagner, 1984; Hájek and Beckett, 2008; Schipperges and Rydin, 2009) and light intensity (Rice et al., 2008; Laine et al., 2011; Bengtsson et al., 2016). Net photosynthesis and hydrological processes are linked via capitulum water retention, which controls the response of capitulum water content to water potential changes (Jassey & Signarbieux, 2019). However, these mechanisms have not been integrated with ecosystem processes in modelling. ~~Due to the lack of quantitative tools, the hypothetical importance of capitulum water retention has not yet been verified.~~

Along with the capitulum water processes, modelling the habitat preference of *Sphagna* requires quantification of the competition among mosses, i.e., the "race for space" (Rydin, 1993; Rydin, 1997; Robroek et al., 2007a; Keuper et al., 2011): *Sphagnum* shoots can~~ould~~ form new capitula and spread laterally, if there is space available. This reduces or eliminates the light source for any plant that is buried by its peers (Robroek et al. 2009). As the competition occurs between neighboring shoots, its modelling requires downscaling water-energy processes from the ecosystem to the shoot level. For that, *Sphagnum* competition needs to be modelled as spatial processes, considering that spatial coexistence and the variations of functional traits among shoot individuals may impact the community dynamics (Bolker et al., 2003; Amarasekare, 2003).

However, coexistence generally reliesy on simple coefficients to describe the interactions among individuals (e.g. Czárán and Iwasa, 1998; Anderson and Neuhauser, 2000; Gassmann et al., 2003; Boulangeat et al., 2018), thus being decoupled from environmental fluctuation or the stochasticity of biophysiological processes.

This study aims to develop and test a model, the Peatland Moss Simulator (PMS), to simulate community dynamics within the peatland moss layer that results in realistic habitat preference of *Sphagnum* species along a moisture gradient. In PMS, community dynamics is driven by *Sphagnum* photosynthesis. –Photosynthesis in turn is regulated by capitulum water retention through capitulum moisture content–. Therefore, we hypothesize that water retention of the capitula is the mechanism driving moss community dynamics. We test the model validity using data from an experiment based on two *Sphagnum* species with different positions along moisture gradient in the same peatland site. If our hypothesis holds, the model will (1) correctly predict the competitiveness of the two species in wet and dry habitats; and (2) fail to predict competitiveness if the capitulum water retention and water content of the two species are not correctly specified.

2. Materials and methods

2.1 Study site

The peatland site being modelled is locatedds in Lakkasuo, Orivesi, Finland (61° 47’ N; 24° 18’ E). The site is a poor fen fed by mineral inflows from a nearby esker (Laine et al 2004). Most of the site is formed by lawns dominated by *Sphagnum recurvum* complex (*Sphagnum fallax*, accompanied by *Sphagnum flexuosum* and *Sphagnum angustifolium*) and *Sphagnum papillosum*. Less than 10% of surface areis occupied by hummocks, with *Sphagnum magellanicum* and *Sphagnum fuscum*, which arebeing 15-25 cm higher than the lawn surfaces with *Sphagnum magellanicum* and *Sphagnum fuscum*. Both microforms are covered by continuous *Sphagnum* carpet with a sparse cover of vascular plants (projection cover of *Carex* 12% on average), which spread homogeneously over the topography. The annual mean water table was 15.6 ± 5.0 cm deep from at lawn surface (Kokkonen et al., 2019). More information about the site can be found in Kokkonen et al. (2019).

2.2 Model outline

The Peatland Moss Simulator (PMS) is a process-based, stochastic model, which simulates the temporal dynamics of *Sphagnum* community as driven by variations in precipitation, irradiation, and energy flow andwith individual-based interactions (Fig. 1). In PMS, the studied ecosystem is seen as a dual-column system consisting of hydrologically connected habitats of hummocks and

lawns (community environment in Fig. 1). For each habitat type, the community area is downscaled to two-dimensional cells representing the scale of individual shoots (i.e. 1 cm^2). Each grid cell can be occupied by one capitulum from a single *Sphagnum* species. The community dynamics, i.e. the changes in species abundances, are driven by the growth and competition of *Sphagnum* shoots at the grid-cell level (Module I in Fig. 1). These processes were regulated by the grid-cell-specific conditions of water and energy (Module II in Fig. 1), which are derived from the community environment (Module III in Fig. 1).

In this study, we focused on developing Module I and II (Section 2.3) and employed an available soil-vegetation-atmosphere transport (SVAT) model (Gong et al., 2013a, 2016) to describe the water-energy processes for Module III (Appendix A). We assumed that the temporal variation in water table was similar in lawns and hummocks, and the hummock-lawn differences in water table (*dWT* in Fig. 1) followed their difference in surface elevations (Wilson, 2012). At the grid cell level, the photosynthesis of capitula drove the biomass growth and elongation of shoots, which led to the ~~"race for space~~competition~~"~~ between adjacent grid cells. The net photosynthesis rate was controlled by capitulum water content (W_{cap}), which was defined by the capitulum water retention in relation to water potential (h) (Section 2.4). The values for functional traits that regulate the growth and competition processes were randomly selected within their normal distribution measured in the field (Section 2.4). Unknown parameters that related the lateral water flows of the site are estimated using a machine-learning approach (Section 2.5). Finally, Monte-Carlo simulation was used to support the analysis on the habitat preferences of *Sphagnum* species and hypothesis tests (Section 2.6). The list of used symbols is given in Table 1.

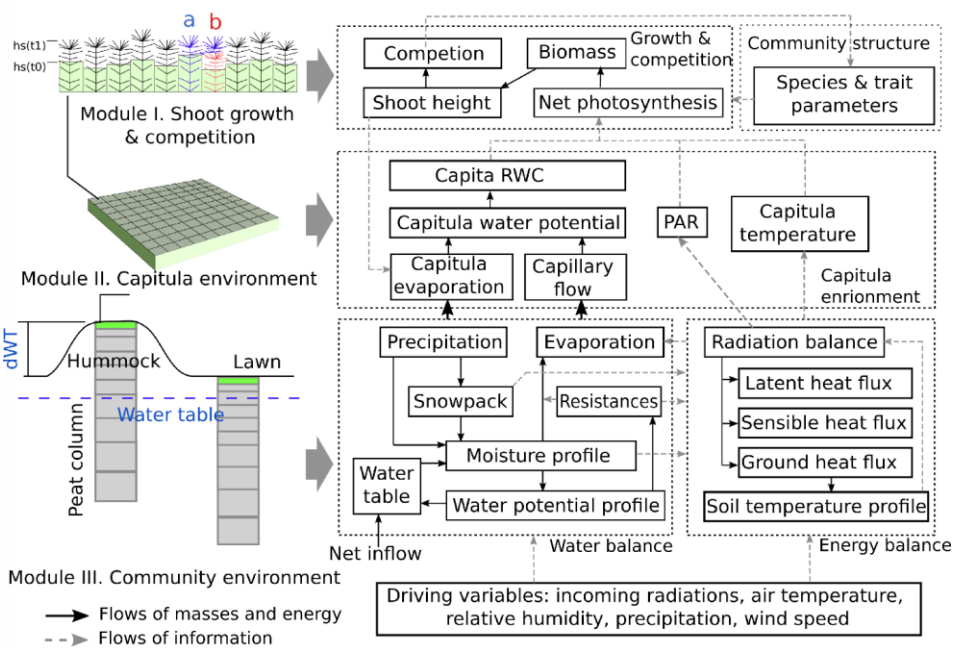


Fig. 1 Framework of Peatland Moss Simulator (PMS).

2.3 Model development

2.3.1 Calculating shoot growth and competition of Sphagnum mosses (Module I)

Calculation of Sphagnum growth

To model grid cell biomass production and height increment, we assumed that capitula were the main parts of shoots responsible for photosynthesis and production of new tissues, instead of the stem sections underneath. We employed a hyperbolic light-saturation function (Larcher, 2003) to calculate the net photosynthesis, which was parameterized based on empirical measurements made from the target species collected from the study site (see Appendix B for materials and methods):

$$A_{20} = \left(\frac{Pm_{20} * PPFD}{\alpha_{PPFD} + PPFD} - Rs_{20}\right) * B_{cap} \quad (1)$$

where subscript 20 denotes the variable value measured at 20 °C; *Rs* is the mass-based respiration rate (μmol g^{-1} s^{-1}); *Pm* is the mass-based rate of maximal gross photosynthesis (μmol g^{-1} s^{-1}); *PPFD* is the photosynthetic photon flux density (μmol m^{-2} s^{-1}); B_{cap} is the capitulum

biomass; and α_{PPFD} is the half-saturation point (μmol m^{-2} s^{-1}) for photosynthesis.

By adding multipliers for capitula water content (f_W) and temperature (f_T) to Eq. (1), the net photosynthesis rate A (μmol m^{-2} s^{-1}) was calculated as following:

$$A = \left[\frac{Pm_{20} * PPFD}{\alpha_{PPFD} + PPFD} f_T(T) - Rs_{20} f_R(T)\right] * B_{cap} * f_W(W_{cap}) \quad (2)$$

where $f_W(W_{cap})$ describes the responses of A to capitulum water content, W_{cap}; $f_T(T)$ describes the responses of Pm to capitulum temperature T (Korrensalo et al., 2017). $f_W(W_{cap})$ was estimated based on the empirical measurements (Appendix B; see Section 2.4). The temperature response $f_R(T)$ is a Q_{10} function that describes the temperature sensitivity of Rs (Frolking et al., 2002):

$$f_R(T) = Q_{10}^{(T - T_{opt})/10} \quad (3)$$

where Q_{10} is the sensitivity coefficient; T is the capitulum temperature (°C); T_{opt} (20 °C) is the reference temperature of respiration.

The response of A to W_{cap} ($f_W(W_{cap})$, Eq. 2) was described as a second-order polynomial function):

$$f_W(W_{cap}) = a_{W0} + a_{W1} * W_{cap} + a_{W2} * W_{cap}^2 \quad (4)$$

where a_{W0}, a_{W1} and a_{W2} are coefficients.

Plants can store carbohydrates as nonstructural carbon (NSC, e.g. starch and soluble sugar) to support fast growth in spring or post-stress periods, like after drought events (Smirnoff et al., 1992; Martínez-Vilalta et al., 2016; Hartmann and Trumbore, 2016). We linked the production of shoot biomass to the immobilization of NSC storage (modified from Eq. 10 in Asaeda and Karunaratne, 2000). The change in NSC storage depends on the balance between net photosynthesis and immobilization:

$$M_B = s_{imm} * NSC * k_{imm} \alpha_{imm}^{T-20} \quad (5)$$

$$\partial NSC / \partial t = A - M_B, NSC \epsilon [0, NSC_{max}] \quad (6)$$

where M_B is the immobilized NSC to biomass production during a time step (g); k_{imm} is the specific immobilization rate (g g^{-1}) (Asaeda and Karunaratne 2000); α_{imm} is the temperature constant; s_{imm} is the multiplier for temperature threshold, where s_{imm} = 1 when T > 5 ºC but s_{imm} = 0 if $T \leq 5$ ºC. NSC_{max} is the maximal NSC concentration in *Sphagnum* biomass (Turetsky et al., 2008). Timing of growth is controlled by a temperature threshold and NSC availability. Growth occurs when T > 5 ºC and NSC is above zero. The dynamics of NSC storage are related to WC through net photosynthesis.

The increase in shoot biomass drove the shoot elongation:

$$\partial Hc/\partial t = \frac{M_B}{H_{spc} S_c} \quad (7)$$

where Hc is the shoot height (cm); H_{spc} is the biomass density of *Sphagnum* stems (g m^{-2} cm^{-1}) and S_c is the area of a cell (m^2).

Calculation of Sphagnum competition and community dynamics

To simulate the competition among *Sphagnum* shoots, we first compared *Hc* of each grid cell (source grid cell, i.e. grid cell *a* in Fig. 1) to its four neighboring cells and marked the one with lowest position (e.g. grid cell *b* in Fig. 1) as the target of spreading. The spreading of shoots from a source to a target grid cell occurred when the following criteria were fulfilled: i) the height difference between source and target grid cells exceeded a threshold value; ii) NSC accumulation in the source grid cell was large enough to support the growth of new capitula in the target grid cell; iii) the capitula in the source grid cell can split at most once per year.

The threshold of height difference in rule i) was set equal to the mean diameter of capitula in the source cell, based on the assumption that the shape of a capitulum was spherical. When shoots spread, the species type and model parameters in the target grid cell were overwritten by those in the source grid cell, assuming the mortality of shoots originally in the target cell. During the spreading, NSC storage was transferred from the source cell to the target cell to form new capitula. In cases where spreading did not take place, establishment of new shoots from spores ~~was allowed to~~could maintain the continuity of *Sphagnum* carpet at the site. During the establishment from spores, which was rare and occurred during the first years of simulation, the ~~properties~~ traits of *Sphagnum* species were randomized within their normal distribution measured in the field.

2.3.2 Calculating grid cell-level dynamics of environmental factors (Module II)

Module II computes grid-cell values of W_{cap}, *PPFD* and *T* for Module I. The cell-level *PPFD* and *T* were assumed to be equal to the community means, which were solved by the SVAT scheme in Module III (Appendix A.). The community level evaporation rate (*E*) was partitioned to cell-level (E_i) as following:

$$E_i = E * \left(\frac{Sv_i}{r_{bulk,i}}\right) \Big/ \sum \left(\frac{Sv_i}{r_{bulk,i}}\right) \quad (8)$$

where $r_{bulk,i}$ is the bulk surface resistance of cell *i*, which is as a function ($r_{bulk,i} = fr(h_i)$) of grid-cell-based water potential h_i, capitulum biomass (B_{cap}) and shoot density (D_S) based on the empirical measurements (Appendix B); Sv_i was the evaporative area, which was related to the height differences among adjacent grid cells:

$$Sv_i = Sc_i + lc \sum_j \left(Hc_i - Hc_j\right) \quad (9)$$

where lc is the width of a grid cell (cm); and subscript j denotes the four-nearest neighbouring grid cells. In this way, changes in the height difference between the neighboring shoots feeds back to affect the water conditions of the grid cells, via alteration of the evaporative surface area.

The grid cell-level changes in capitula water potential (h_i) were~~as~~ driven by the balance between the evaporation (E_i) and the upward capillary flow to capitula:

$$\partial h_i = \frac{K_m}{C_i}\left[\frac{(h_i - h_m)}{0.5 z_m} - 1 - E_i\right] \quad (10)$$

where h_m is the water potential of the living moss layer, solved in Module III (Appendix A.); z_m is the thickness of the living moss layer (z_m=5 cm); K_m is the hydraulic conductivity of the moss layer and that is set to be the same for each grid cell; C_i is the cell-level specific water uptake capacity ($C_i = \partial W_{cap,i}/\partial h_i$). $\partial W_{cap,i}/\partial h_i$ could be derived from the capitulum water retention function $h_i = f_h(W_{cap})$. W_{cap} can be then calculated from the estimated from h_i and affect the calculation of net photosynthesis through $f_W(W_{cap})$ (Eq. 2).

2.4 Model parameterization

Selection of Sphagnum species

We chose *S. fallax* and *S. magellanicum*, which form 63% of total plant cover at the study site at Lakkasuo (Kokkonen et al., 2019), as the target species representing the lawn and hummock habitats respectively. These species share a similar ~~a~~ niche along the gradients of soil pH and nutrient richness (Wojtuń et al., 2003), but are discriminated by their preferences of water table level (Laine et al., 2004). While *S. fallax* is commonly found close to the water table (Wojtuń et al., 2003), *S. magellanicum* can occur along a wider range of a dry-wet gradient, from intermediately wet lawns up to dry hummocks (Rice et al., 2008; Kyrkjeeide, et al., 2016; Korresalo et al., 2017). The transition from *S. fallax* to *S. magellanicum* along the wet-dry gradient thus indicates the decreasing competitiveness of *S. fallax* against *S. magellanicum* with a lowering water table.

Parameterization of morphological traits, net photosynthesis and capitulum water retention

We empirically quantified the morphological traits capitulum density (D_S, shoots cm^{-2}), biomass of capitula (B_{cap}, g m^{-2}), biomass density of living stems (H_{spc}, g cm^{-1} m^{-2}), net photosynthesis parameters (Pm_{20}, Rs_{20} and α_{PPFD}) and the water retention properties (i.e., $f_h(W_{cap})$ and $fr(h)$, Eqs. 8 and 10) for the two *Sphagnum*~~*selected*~~ species~~ from the same site~~ (see Appendix B for methods). The values (mean ± SD) of the morphological parameters, the photosynthetic parameters and polynomial coefficients (a_{W0}, a_{W1} and a_{W2}, Eq. 3) are listed in Table 2. For each parameter, a random value was initialized for each cell based on the measured means and SD,

assuming the variation of parameter values is normally distributed.

We noticed that the fitted $f_W(W_{cap})$ was meaningful when W_{cap} was below the optimal water content for photosynthesis (W_{opt} = -0.5 a_{W1}/ a_{W2}). If $W_{cap} > W_{opt}$, photosynthesis decreased linearly with increasing W_{cap}, as being limited by the diffusion of CO_2 (Schipperges and Rydin, 1998). In that case, $f_W(W_{cap})$ was calculated following Frolking et al. (2002):

$$f_W(W_{cap}) = 1 - 0.5\frac{W_{cap}-W_{opt}}{W_{max}-W_{opt}} \quad (11)$$

where W_{max} is the maximum water content of capitula.

It is known that W_{max} is around 25-30 g g^{-1} (e.g. Schipperges and Rydin, 1998), or about 0.31 - 0.37 cm^3 cm^{-3} in term of volumetric water content (assuming 75 g m^{-2} capitula biomass and 0.6 cm height of capitula layer). This range is broadly lower than the saturated water content of moss carpet (> 0.9 cm^3 cm^{-3}, McCarter and Price, 2014). Consequently, we used the following equation to convert volumetric water content to capitula RWC, when h_i was higher than the boundary value of -10^4 cm:

$$W_{cap} = min(W_{max}, \theta_m/(H_{cap} * B_{cap} * 10^{-4})) \quad (12)$$

where W_{max} is the maximum water content that set to 25 g g^{-1} for both species; θ_m is the volumetric water content of moss layer; H_{cap} is the height of capitula and is set to 0.6 cm (Hájek and Beckett, 2008).

~~*Parameterization of SVAT processes*~~

~~For the calculation of surface energy balance, we set the height and leaf area of vascular canopy to 0.4 m and 0.1 m^2 m^{-2}, consistent with the scarcity of vascular canopies at the site. The aerodynamic resistance (r_{aero}, Eq. A14, Appendix A) for surface energy fluxes was calculated following Gong et al. (2013a). The bulk surface resistance of community (r_{ss}, Eq. A13, Appendix A) was summarized from the cell-level values of $r_{bulk,i}$, that $1/r_{ss} = \sum(1/r_{bulk,i})$. To calculate the peat hydrology and water table, peat profiles of hummock and lawn communities were set to 150 cm deep and stratified into horizontal layers of depths varying from 5cm (topmost) to 30cm (deepest). For each peat layer, the thermal conductivity (K_T) of fractional components, i.e. peat, water and ice, were evaluated following Gong et al. (2013a). The bulk density of peat (ρ_{bulk}) was set to 0.06 g cm^{-3} below acrotelm (40 cm depth, Laine et al., 2004), and decreased linearly toward the living moss layer. The saturated hydraulic conductivity (K_{sat}, Eq. A6, Appendix A) and water retention parameters (i.e. α and n, Eq. A5, Appendix A) of water retention curves were calculated as functions of ρ_{bulk} and the depth of peat layer following Päivänen (1973). K_{sat}, α and n for the living moss layer were adopted from the values measured by McCarter and Price (2014)~~

from *S. magellanicum* carpet. The parameter values for SVAT processes are listed in Table 3.

Calculation of snow dynamics

In boreal and arctic regions, the amount and timing of snow melt has crucial impact on moisture conditions, especially at fen peatlands. Therefore, to have realistic spring conditions we introduced a snow pack model, SURFEX v7.2 (Vionnet et al., 2007), into the SVAT modelling. The snow pack model simulates snow accumulation, wind drifting, compaction and changes in metamorphism and density. These processes influenced the heat transport and freezing-melting processes (i.e. S_h and S_T, see Eq. A1-A2, Appendix A). In this modelling, we calculate the snow dynamics on a daily basis in parallel to the SVAT simulation. Daily snowfall was converted into a snow layer and added to ground surface. For each of the day-based snow layers (D-layers), we calculated the changes in snow density, particle morphology and layer thicknesses. At each time step, D-layers were binned into layers of 5-10 cm depths (S-layers) and placed on top of the peat column for SVAT modelling. With a snow layer present, surface albedos (i.e. a_s, a_i) were modified to match those of the topmost snow layer (see Table 4 in Vionnet et al., 2007). If the total thickness of snow was less than 5 cm, all D-layers were binned into one S-layer. The thermal conductivity (K_T), specific heat (C_T), snow density, thickness and water content of each S-layer were calculated as the mass-weighted means from the values of D-layers. Melting and refreezing tended to increase the density and K_T of a snow layer but decrease its thickness (see Eq. 18 in Vionnet et al., 2007). The fraction of melted water that exceeded the water holding capacity of a D-layer (see Eq. 19 in Vionnet et al., 2007) was removed immediately as infiltration water. If the peat layer underneath was saturated, the infiltration water was removed from the system as lateral discharge.

Boundary conditions and driving variables

A zero-flow boundary was set at the bottom of peat. At peat surface the boundary conditions of water and energy were defined by the ground surface temperature (T_0, see Eq. A10-A15 in Appendix A) and the net precipitation (*P* minus *E*). The profiles of layer thicknesses, ρ_{bulk} and hydraulic parameters were assumed to be constant during simulation. Lateral boundary conditions were used to calculate the spreading of *Sphagnum* shoots among cells along the edge of the model domain so that shoots can spread across the edge of simulation area and invade into the grid cell at the boarder of the opposite side.

The model simulation was driven by climatic variables of air temperature (*Ta*), precipitation (P), relative humidity (*Rh*), wind speed (*u*), incoming shortwave radiation (*Rs*) and longwave radiation (*Rl*). To support the stochastic parameterization of the model and Monte Carlo simulations, Weather Generator (Strandman et al., 1993) was used to generate randomized scenarios based on long-term weather statistics (period of 1981-2010) from 4 closest weather stations of Finnish Meteorological Institute. This generator had been intensively tested and applied under Finnish conditions (Kellomäki and Väisänen, 1997; Venäläinen et al., 2001; Alm

~~et al., 2007). We also compared the simulated meteorological variables against 2-year data measured from Siikaneva peatland site (61°50 N; 24°10 E), located 10 km away from our study site (Appendix C).~~

2.5 Model calibration for lateral water influence

We used a machine-learning approach to estimate the influence of upstream area on the water balance of the site. The rate of net inflow (*I*, see Eq. A18 in Appendix A.) was described as a function of Julian day (*JD*), assuming the inflow was maximum after spring thawing and then decreased linearly with time:

$$I_j = (a_N * JD + b_N) * Ks_j, JD > JD_{thaw} \quad (11)$$

where subscript *j* denotes the peat layers under water table; *Ks* is the saturated hydraulic conductivity; JD_{thaw} is the Julian day that thawing completed; and a_N and b_N are parameters.

We simulated water table changes using climatic scenarios from the Weather Generator (~~Section 2.4~~Appendix A). During the calibration, the community compositions were set constant, such that *S. magellanicum* fully occupied the hummock habitat whereas *S. fallax* fully occupied the lawn habitat. The simulated multi-year means of weekly water table values were compared to the weekly mean water table obtained observed at the site during years 2001, 2002, 2004 and 2016. The cost function for the learning process was based on the sum of squared error (*SE*) of the simulated water table:

$$SE = \Sigma(WTs_k - WTm_k)^2 \quad (12)$$

where *WTm* is the measured multi-year weekly mean of water table; *WTs* is the simulated multi-year weekly mean of water table; and subscript *k* denotes the week of year when the water table was sampled.

The values of a_N and b_N were estimated using the Gradient Descent approach (Ruder, 2016), by minimizing *SE* in above Eq. (19):

$$X_N(j) := X_N(j) - \Gamma \frac{\partial SE}{\partial X_N(j)} \quad (13)$$

where Γ is the learning rate (Γ = 0.1). Appendix D shows the simulated water table with the calibrated inflow term *I*, as compared against the measured values from the site.

2.6 Model-based analysis

First, we examined the ability of model to capture the preference of *S. magellanicum* for the hummock environment and *S. fallax* for the lawn environment (Test 1). For both species, the probability of occupation was initialized as 50% in a cell, and the distribution of species in the communities were randomly patterned. Monte-Carlo simulations (40 replicates) were carried out, with a time step of 30 minutes. A simulation length of 15 years was selected based on preliminary studies, in order to cover the major ~~part~~ interval of change and to ease ~~the~~ computational demand. Biomass growth, stem elongation and the spreading of shoots were simulated on a daily basis. The establishment of new shoots in deactivated cells was calculated at the end of each simulation year. We then assessed if the model could capture the dominance of *S. magellanicum* in the hummock communities and the dominance of *S. fallax* in lawn communities. The simulated annual height increments of mosses were compared to the values measured for each community type. To measure moss height growth in the field, we deployed 20 cranked wires on *S. magellanicum* dominated hummocks and 15 on *S. fallax* dominated lawns in 2016. Each cranked wire was a piece of metal wire attached with plastic brushes at the side anchored into the moss carpet (e.g. Clymo 1970, Holmgren et al., 2015). Annual height growth (*dH*) was determined by measuring the change in the exposed wire length above moss surface from the beginning to the end of growing season.

Second, we tested the robustness of the model to the uncertainties in a set of parameters (Test 2-4). In test 2, ~~W~~we focused on parameters that closely linked to hydrology and growth calculations, but were roughly parameterized (e.g., k_{imm}, r_{aero}) or adopted as a prior from other studies (e.g., K_{sat}, α, n, NSC_{max}; see Table 3). One at a time, each parameter value was adjusted by +10 % or -10 ~~%, and species cover was simulated using the same runtime settings as Test 1 with 40 Monte-Carlo runs. The simulated means of cover were then compared to those calculated without the parameter adjustment~~. 40 Monte-Carlo simulations were run using the same runtime settings as in Test 1. The simulated means of cover were then compared to those calculated without the parameter adjustment.

Tests 3-4 were then carried out to test whether the model could correctly predict competitiveness of the species in dry and wet habitats, if the species-specific trends of capitulum water ~~retention~~ content were not correctly specified. For both species, we set the values of parameters controlling the water retention (i.e. B_{cap} and D_S, Appendix B) and the water-stress effects on net photosynthesis (i.e. W_{cap}, Eq. 4) to be the same as those in *S. magellanicum* (Test 3) or same as those in *S. fallax* (Test 4). Our hypothesis would be supported if removing the interspecific differences in *RWC* responses led to the failure to predict the habitat preferences of the species.

We implemented Tests 5-6 to test the importance of parameters that directly control the species ability to overgrow another species with more rapid height increment (i.e. Pm_{20}, Rs_{20}, α_{PPFD} and H_{spec}) in lawn and hummock conditions. We eliminated the species differences in the parameter

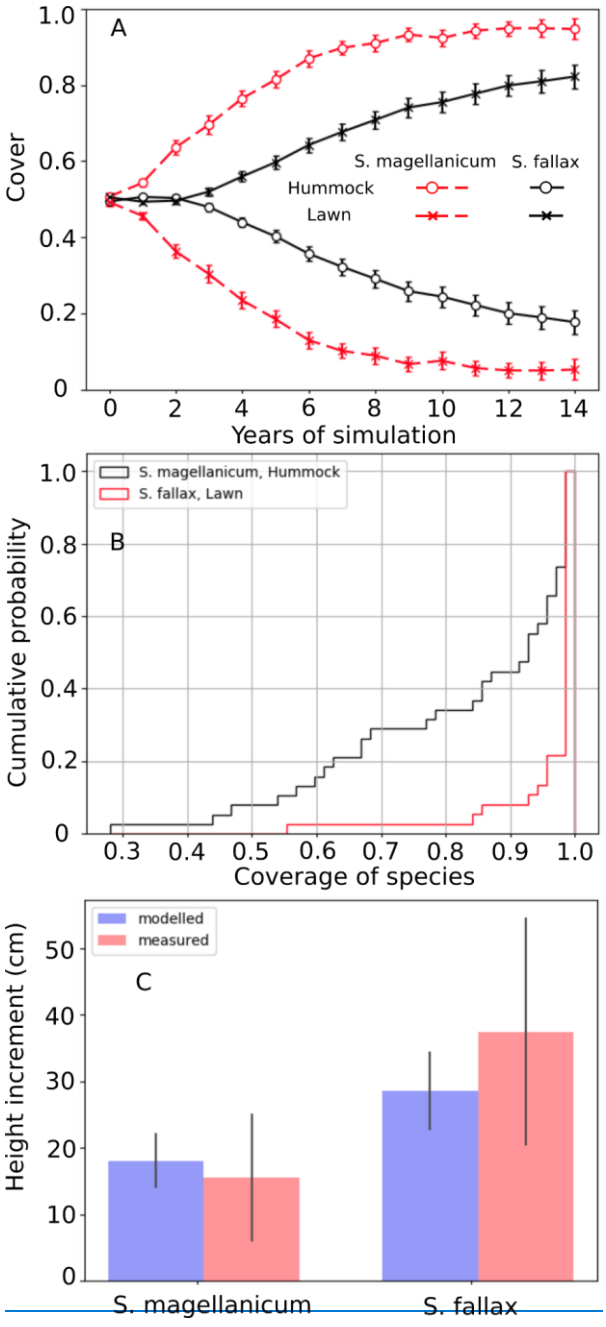
values to be same as those in *S. magellanicum* (Test 5) and same as those in *S. fallax* (Test 6). The effects of the manipulation were compared against those from Tests 3-4. For each of Tests 3-6, 80 Monte-Carlo simulations were run using the setups described in Test 1.

Test 7-8 were implemented to separate the effects of photosynthetic water-response parameters from the effects of the water retention of capitula. We set the photosynthetic water-response parameters to be the same as those in *S. magellanicum* (Test 7) and same as those in *S. fallax* (Test 8). As our model aimed to couple the environmental fluctuations and stochasticity of ecosystem processes, we further tested the model responses to the absences of environmental fluctuations (Test 9) or the absence of stochasticity in model parameters (Test 10). In Test 9, monthly mean values of meteorological variables were used to drive the model simulation. In Test 10, we removed the stochasticity of model parameters, and assigned average value to each parameter of grid cells. For each of Tests 7-10, 40 Monte-Carlo simulations were run using the setups described in Test 1.

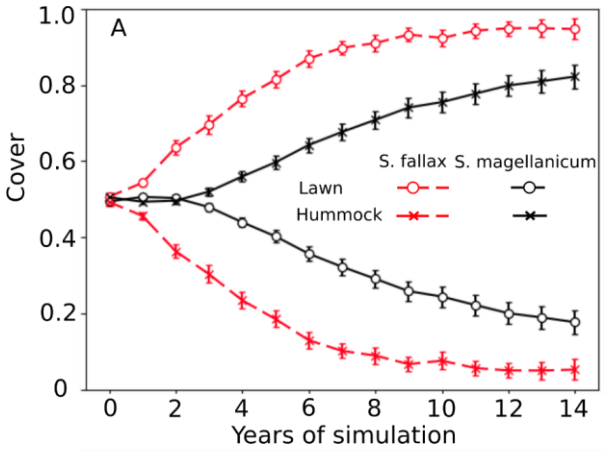
3 Results

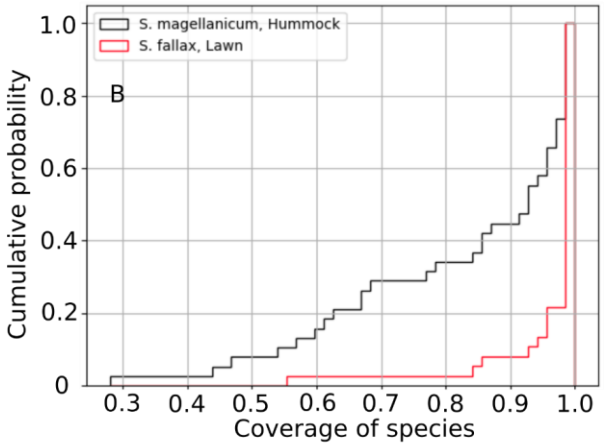
3.1 Simulating the habitat preferences of *Sphagnum* species as affected by water ~~retention~~ content traits of capitulum

Test 1 ~~showed~~ deemonstrated the ability of model to capture the preference of *S. magellanicum* for the hummock environment and *S. fallax* for the lawn environment (Fig. 2A). The simulated annual changes in species covers were greater in lawn than in hummock habitats during the first 5 simulation years. The changes in lawn habitats slowed down around year 10 and the cover of *S. fallax* plateaued at around 95±2.8% (mean ±standard error). In contrast, the cover of *S. magellanicum* on hummocks continued to grow until the end of simulation and reached 83±3.1%. In the lawn habitats, the cover of *S. fallax* increased in all Monte-Carlo simulations and the species occupied all grid cells in 70% of the simulations. In the hummock habitats, the cover of *S. magellanicum* increased in 91% of Monte-Carlo simulations, and formed monocultural community in 16% of simulations (Fig. 2B). The height growth of *Sphagnum* mosses was significantly greater at lawns than at hummocks ($P<0.01$). The ranges of simulated height growths agreed well with the observed values from field measurement for both species (Fig. 2C).


A
Cover
Years of simulation
S. magellanicum
S. fallax
Hummock
Lawn
B
Cumulative probability
Coverage of species
S. magellanicum, Hummock
S. fallax, Lawn
C
Height increment (cm)
modelled
measured
S. magellanicum
S. fallax

485


A
Cover
Years of simulation
S. fallax
S. magellanicum
Lawn
Hummock


B
S. magellanicum, Hummock
S. fallax, Lawn
Cumulative probability
Coverage of species

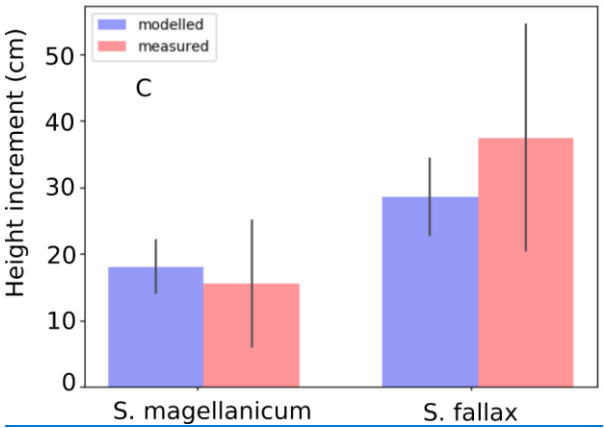

C
modelled
measured
Height increment (cm)
S. magellanicum
S. fallax

Figure 2. Testing the ability of P~~CS~~MS to predict habitat preference of *Sphagnum magellanicum* and *S. fallax* (Test 1). The hummock and lawn habitats were differentiated by water table depth, surface energy balances and capitulum water potential in modelling. In the beginning of simulation, the cover of the two species was set equal and it was allowed to develop with time. (A) Annual development of the relative cover (mean and standard error) of the two species in hummock and lawn habitats, (B) the cumulative probability distribution of the cover of the two species at the end of the 15-year period based on 40~~80~~ Monte-Carlo simulations, and (C) the simulated and measured means of annual height growth of *Sphagnum* surfaces in their natural habitats in hummock and lawn habitats.

3.2 Testing model robustness

Test 2 addressed the model robustness to the uncertainties in several parameters that closely linked to hydrology and growth calculations. Modifying most of the parameter values by +10% or -10% yielded marginal changes in the mean cover of species in either hummock or hollow habitat (Table 4). Reducing the moss carpet and peat hydraulic parameter *n* had stronger impacts on *S. fallax* cover in hummocks than in lawns. Nevertheless, changes in simulated cover that were caused by parameter manipulations were generally smaller than the standard deviations of the means i.e. fitting into the random variation.

3.3 Testing the controlling role of capitulum water ~~retention~~ content for community dynamics

In Tests 3 and 4, the model incorrectly predicted the competitiveness of two species when the interspecific differences of capitulum water ~~retention~~ content were eliminated. In both tests, *S. fallax* became dominant in all habitats. The use of water responses characteristic to *S. magellanicum* for both species (Test 3) led to faster development of *S. fallax* cover and higher coverage at the end of simulation (Fig. 3A), as compared with the simulation results where the water responses characteristic to *S. fallax* were used for both species (Test 4, Fig. 3B). The pattern was more pronounced in hummock than in lawn habitats.

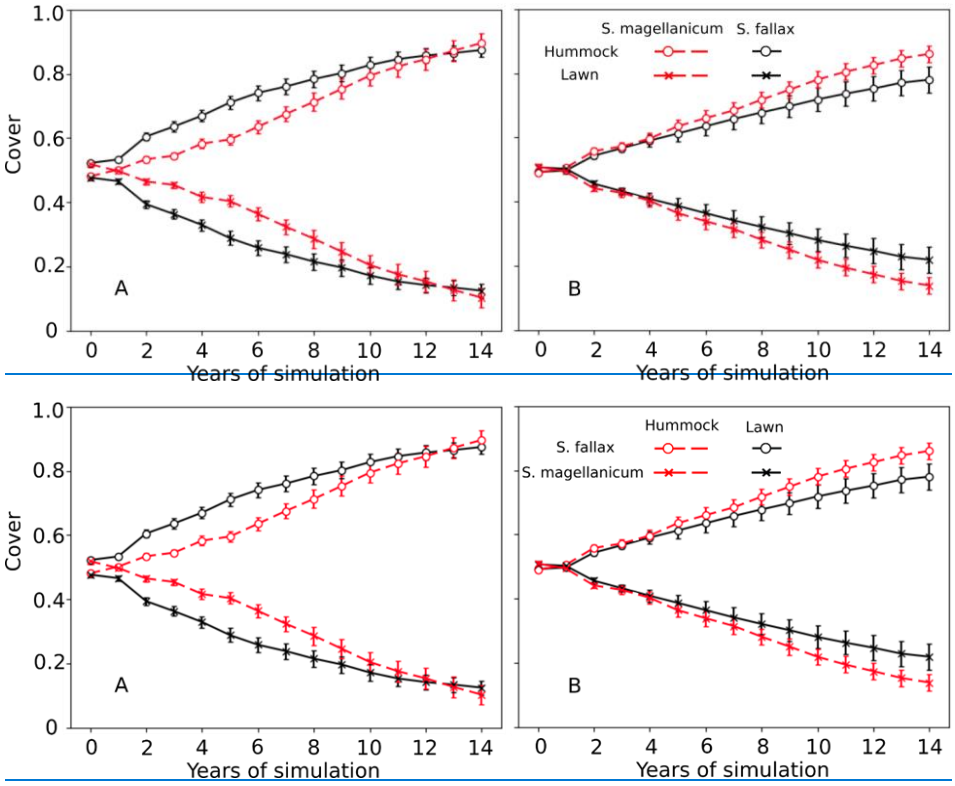


Figure 3. Testing the importance of capitulum water ~~retention~~ content to the habitat preference of *S. magellanicum* and *S. fallax*. The development of the relative cover (mean and standard error) were simulated in hummock and lawn habitats over a 15-year time frame for the two species. For both species, parameter values for the capitulum water ~~retention~~content, capitulum biomass (B_{cap}) and density (D_S) were set to be the same as those from (A) *S. magellanicum* (Test 3) or (B) *S. fallax* (Test 4).

In Tests 5 and 6, the species differences in the growth-related parameters were eliminated. However, the model still predicted the dominances of *S. fallax* and *S. magellanicum* in lawn and hummock habitats, respectively (Fig. 4). The increase in the mean cover of *S. magellanicum* was especially fast in hummock habitat in comparison to the results of the unchanged model from Test 1 (Fig. 2A). In lawns, the use of *S. fallax* growth parameters for both species gave stronger competitiveness to *S. magellanicum* (Fig. 4B) than using the *S. magellanicum* parameters (Fig. 4A). In Test 7 and 8, ignoring the interspecific differences in the photosynthetic water-response

parameters did not change the simulated habitat preferences of *S. fallax* and *S. magellanicum* (Table 5). Using the water response parameters of *S. fallax* decreased the mean cover of *S. fallax* in lawns but increased the cover of *S. magellanicum* on hummocks. In contrast, using the water response parameters of *S. magellanicum* increased the mean cover of *S. fallax* in lawns but decreased the cover of *S. magellanicum* on hummocks.

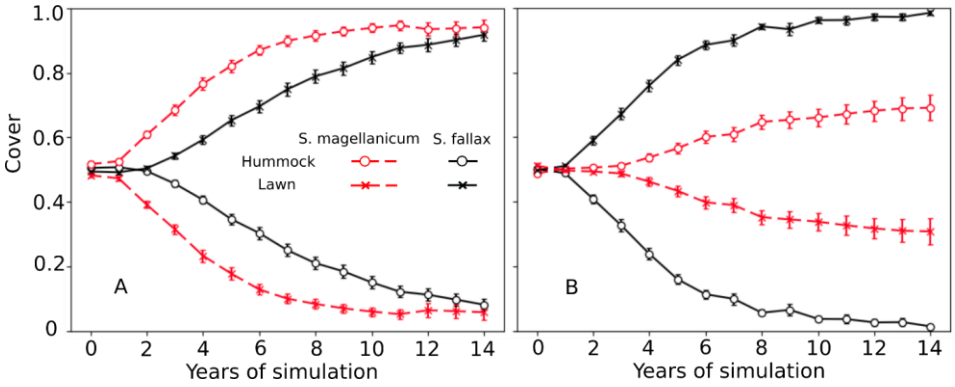


Figure 4. Testing the importance of parameters regulating net photosynthesis and shoot elongation to the habitat preference of *S. magellanicum* and *S. fallax*. Annual development of the relative cover (mean and standard error) of the two species were simulated for hummock and lawn habitats over a 15-year time frame. For both species, the parameter values (i.e. Pm_{20}, Rs_{20}, α_{PPFD} and H_{spec}) were set to be the same as those from (A) *S. magellanicum* (Test 5) or (B) *S. fallax* (Test 6).

3.4 Testing the effects of environmental fluctuations and stochasticity of ecosystem processes on community dynamics

In Tests 9, the model failed to simulate the preference of *S. magellanicum* to hummocks (Table 5) if the environmental fluctuation was ignored. However, the simulated cover of *S. fallax* in lawns was higher as compared to unchanged condition (i.e. Test 1). Using mean value for each model parameters led to mono output, i.e. *S. magellanicum* occupied 100% hummock area whereas *S. fallax* took over lawns completely.

4 Discussion

In peatland ecosystems, ~~*Sphagnum*~~ *Sphagna* are keystone species ~~differentially~~ distributed primarily along the hydrological gradient (e.g. Andrus et al. 1986; Rydin, 1986). In a context where substantial change in peatland hydrology is expected under a changing climate in northern areas (e.g. longer snow-free season, lower summer water table and more frequent droughts), there is a pressing need to understand how peatland plant communities could react and how *Sphagnum* species could redistribute under habitat changes. In this work, we developed the Peatland Moss Simulator (PMS), a process-based stochastic model, to simulate the competition between *S. magellanicum* and *S. fallax*, two key species representing dry (hummock) and wet (lawn) habitats in a poor fen peatland. We empirically showed that these two species differed in characteristics that likely affect their competitiveness along a moisture gradient.

~~The~~ Capitulum water retention for the lawn-preferring species (*S. fallax*) was weaker than that for the hummock-preferring species (*S. magellanicum*). Compared to *S. magellanicum*, the capitula of *S. fallax* held less water at saturation and water content decreased more rapidly with dropping water potential. Hence, *S. fallax* would dry faster than *S. magellanicum* under the same rate of water loss. Moreover, ~~the~~ water content in *S. fallax* capitula was less resistant to evaporation. These differences indicated that it is harder for *S. fallax* capitula to buffer evaporative water loss ~~of water~~ and thereby avoid or delay desiccation. Similar differences between hummock and hollow species have been found ~~also~~ earlier (Titus & Wagner, 1984; Rydin & McDonald, 1985). In addition, the net photosynthesis of *S. fallax* is more sensitive to changes in capitulum water content than *S.* ~~*magellanicum*~~ *magellanicum*, as seen in a steeper decline in photosynthesis with decreasing water content (Fig. B2C). Consequently, the growth of *S. fallax* is more likely to be ~~constrained~~ slowed down by dry periods, when the capillary water cannot fully compensate the evaporative loss (Robroek et al., 2007b) making it less competitive in habitats prone to desiccation.

The PMS successfully captured the habitat preferences of the two *Sphagnum* species (Test 1): starting from a mixed community with equal probabilities for both species, the lawn habitats with shallower water table were eventually dominated by the typical lawn species *S. fallax*, whereas hummock habitats, which were 15 cm higher than the lawn surface, were taken over by *S. magellanicum*. The low final cover of *S. magellanicum* simulated in lawn habitats agreed well with field observation from our study site, where *S. magellanicum* cover was less than 1% in lawns~~s~~ (Kokkonen et al., 2019). On the other hand, *S. fallax* was outcompeted by *S. magellanicum* in the hummock habitats. This result is consistent with previous findings that hollow-preferring *Sphagna* are less likely to survive in hummock environments with greater drought pressure (see Rydin 1985; Rydin et al. 2006; Johnson et al., 2015). The simulated annual height increments of mosses also agreed well with the observed values for both habitat types. ~~As was the case in o~~Our simulation for lawn habitat shows that~~,~~ the looser stem structure of *S.*

fallax~~,~~ allows it to allocate more of ~~the~~ its produced biomass into height growth, and thereby~~in~~ overgrow *S. magellanicum*, in which new biomass forms a ~~that allocates the produced biomass to form~~ compact stem, packed with thick fascicles. This finding indicates~~ed~~ that PMS can capture key mechanisms in controlling the growth and competitive interactions of the *Sphagnum* species.

~~The testing of~~P~~ p~~arameter sensitivity testing showed the robustness of PMS regarding the uncertainties in parameterization, as the simulated changes in the mean species cover, under 10% changes in several parameters, ~~-~~were generally less than the standard deviations of the means ~~under 10% changes in several parameters~~. ~~We found that~~D ~~d~~ecreasing the value of the hydraulic parameter *n* (Table 3, Eq. A5) increased the presence of *S. fallax* in the hummock habitats. This was expected as *n* is a scaling factor and therefore its changes get magnified~~-~~: a lower *n* value will lead to higher water content in the unsaturated layers above the water table (van Genuchten, 1978), which allows ~~is important to~~ wet-adapted *Sphagna* ~~in order~~ to survive dry conditions (Hayward and Clymo, 1982; Robroek et al., 2007b; Rice et al., 2008). In contrast, the response of *Sphagnum* cover to the changes in other hydraulic parameters (i.e. α, *n*, K_h) was~~ere~~ limited in lawn habitats. This could be due to the relatively shallow water table in lawns, which was able to maintain sufficient capillary rise to the moss carpet and capitula. Decreasing the values of the specific immobilization rate (*kimm*~~-~~) and maximal NSC concentration in *Sphagnum* biomass (*NSC*$_{max}$~~-~~) mainly decreased the cover of *S. fallax* in lawn habitats, consistent with the importance of biomass production to *Sphagna* in high moisture environment (e.g. Rice et al., 2008; Laine et al., 2011). In addition, the SVAT modelling for hummocks and lawns (Module III, Fig. 1) employed same hydraulic parameter values obtained from *S. magellanicum* hummocks (McCarter and Price, 2014). For lawns, t~~T~~his could overestimate K_m but underestimate *n* ~~for lawns~~, as the lawn peat would be'~~could~~ be less efficient in holding high water ~~retention~~ content and generating capillary-flow ~~generation, as compared to~~than hummock peat (Robroek et al., 2007b; Branham and Strack, 2014). As the decrease in K_m and increase in *n* showed counteracting effects on the simulated species covers (Table 4), the biases in the parameterization of K_m and *n* may not critically impact model performance.

Both our empirical measurements and PMS simulations indicate the importance of capitulum water ~~retention~~content~~-~~ as a mechanism controlling the moss community dynamics in peatlands. It has long been hypothesized and experimentally studied that *Sphagnum* niche is defined by two processes. Firstly, dry, high elevation habitats such as hummocks physically select species with ability to remain moist (Rydin, 1993).~~—~~ If the interspecific differences in water retention and water-stress effects were correctly specified (Test 1, Fig. 2) our model predicted this phenomena of stronger competitiveness of *S. magellanicum* against *S. fallax* in hummock habitats correctly. Alternatively, the model failed to predict the distribution of *S. magellanicum* on hummocks, if these interspecific differences in the water processes were neglected (Test 3 and Test 4, Fig. 3). During low water table periods in summer ~~the~~ capillary rise may not fully compensate for ~~the~~

high evaporation (Robroek et al., 2007b; Nijp et al., 2014). In such circumstances, capitulum water potential could drop rapidly towards the pressure defined by the relative humidity of air (Hayward and Clymo, 1982). Consequently, the ability of capitula to retain cytoplasmic water ~~would be~~is particularly important for the hummock-preferring species, as was also shown by Titus & Wagner (1984).

Secondly, ~~,~~ in habitats with more persistently high moisture content such as lawns and hollows, ~~the~~ interspecific competition becomes important: ~~-~~it is well acknowledged that species from such habitats generally have higher growth rates and photosynthetic capacity compared to hummock species (e.g. Laing et al., 2014; Bengtsson et al., 2016). Our results also agreed on this, as setting the growth-related parameters (i.e. Pm_{20}, Rs_{20}, α_{PPFD} and H_{spec}) of *S. magellanicum* to be the same as those of *S. fallax* decreased the *S. fallax* cover in both hummock and lawn habitats (Test 6, Fig. 4B). However, ~~the model still captured the~~ such changes didn't impact the simulated habitat preferences for the tested species~~ without including the interspecific differences in those growth-related parameters~~. Based on this, the growth-related parameters ~~could~~ seem to be less important than those water-related ones. Further on, our Tests 7 and 8 showed that when interspecific differences in the water-stress effects on photosynthesis were removed, the model still predicted the correct habitat preferences of *S. magellanicum* and *S. fallax*. Therefore, the interspecific differences in capitulum water retention could be the main determinant on the habitat preferences of the tested species.

There have been growing concerns ~~on~~ about the shift of peatland communities from *Sphagnum*-dominated towards more vascular-abundant under a drier and warmer climate (Wullschleger et al., 2014; Munir et al. 2015; Dieleman et al. 2015). Nevertheless, the potential of *Sphagnum* species composition to adjust to this forcing remains poorly understood. Particularly in oligotrophic fens, where the vegetation is substantially shaped by lateral hydrology (Tahvanainen, 2011; Turetsky et al., 2012), plant communities can be highly vulnerable to hydrological changes (Gunnarsson et al. 2002; Tahvanainen, 2011). Based on the validity and robustness of PMS, we believe PMS could serve as one of the first mechanistic tools to investigate the direction and rate of change in *Sphagnum* communities ~~to change~~ under environmental forcing. The hummock-lawn differences showed by Test 1 imply~~ied~~ that *S. magellanicum* could outcompete *S. fallax* within a decad~~cal time frame~~ in a poor fen community, if the water table of habitats like lawns was lowered by 15 cm (Test 1). Although this was derived from a simplified system with only the two species, it highlighted the potential of rapid turnover of *Sphagnum* species: the hummock-lawn difference of water table in simulation was comparable to the expected water table drawdown in fens under the warming climate (Whittington and Price, 2006; Gong et al., 2013b). The effect traits of mosses, while studied less than those of vascular plant traits, have far reaching impacts on biogeochemistry of ecosystems

such as peatlands, where mosses form the most significant plant group (Cornelissen et al. 2007). Because of the large interspecific differences of traits such as photosynthetic potential, hydraulic properties and litter chemistry (Laiho 2006; Straková et al., 2011; Korrensalo et al., 2017; Jassey & Signarbieux, 2019), change in *Sphagnum* community composition is likely to impact long-term peatland stability and functioning (Waddington et al., 2015). Turnover between hummock and wetter habitat species would feedback to climate as they differ in their decomposability (Straková et al. 2012; Bengtsson et al. 2016). As hummock species produces more calcitrant litter the carbon bind into the system would take longer to get released back to atmosphere. In addition, the replacement of wet adapted moss species with hummock species is likely to result in higher ability to maintain carbon sink under periods of drought (Jassey, & Signarbieux, 2019).

Although efforts have been made on analytical modelling to obtain boundary conditions for equilibrium states of moss and vascular communities in peatland ecosystems (Pastor et al., 2002), the dynamic~~al~~ process of peatland vegetation has not been well-described or included in earth system models (ESMs). Existing ecosystem models usually consider the features of peatland moss cover as "fixed" (Sato et al., 2007; Wania et al., 2009; Euskirchen et al., 2014), or change directionally following a projected trajectory (Wu and Roulet, 2014). Chaudhury et al. (2017) have a dynamic peatland vegetation model, with a single moss PFT and four vascular PFTs, so moss productivity relative to vascular plants can vary, however moss characteristics are fixed to a single set of values. Our modelling approach provided a way to incorporate the environmental fluctuation and the mechanisms of dynamic~~al~~ moss cover into peatland carbon modelling. PMS employed an individual-based approach where each grid cell carries a unique set of trait properties, so that shoots with favorable trait combinations in prevailing environment are thus able to replace those whose trait combinations are less favorable. Moreover, the model included the spatial interactions of individuals, which can impact the sensitivity of coexistence pattern to environmental changes (Bolker et al., 2003; Sato et al., 2007; Tatsumi et al., 2019). This mimics the stochasticity in plant responses to environmental fluctuations, which ~~are~~ is essential to community assembly and trait filtering under environmental forcing (Clark et al., 2010). The importance of incorporating environmental fluctuations with the stochasticity of biophysiological processes is supported by our Test 9 and 10. If the monthly mean climate conditions were used as input, our model failed to predict the dominance of *S. magellanicum* on hummocks. If the stochasticity of model parameters were omitted and only mean values were used, the model generated only single output disregarding the randomness of environmental conditions. ~~Moreover, the model includesd the spatial interactions of individuals, which can impact the sensitivity of coexistence pattern to environmental changes (Bolker et al., 2003; Sato et al., 2007; Tatsumi et al., 2019). Because~~ As these features are considered essential to the "next generation" DVMs (Scheiter et al., 2013), our PMS ~~with competition based on growth rates~~ could be considered as an elemental design for future DVM development.

~~To conclude, our PMS could successfully capture the habitat preferences of the modelled~~

~~*Sphagnum*. In this respect, our PMS model could provide fundamental support for the future development of dynamic vegetation models for peatland ecosystems. Based on our findings, the capitulum water processes should be considered as a control on the vegetation dynamics in future impact studies on peatlands under changing environmental conditions.~~

We see PMS as an elemental design for the future development of dynamic vegetation models for peatland ecosystems, yet there are certain uncertainties and features that should be developed further. We used a gas-exchange-based method to quantify the simultaneous changes in capitula water potential, water content and carbon uptake of *Sphagnum* moss capitula. It should be noted that, the measurements mainly covered the changes from RWC_{opt} towards RWC_{cmp} (Table~~.~~ 1 and Fig. 3). However, capitula water content could be higher than RWC_{opt} at saturation (e.g. about 25-30 g g^{-1}; Schipperges and Rydin, 1998). When RWC is high, vapor diffusion may occur mainly from the capitula surface or macropores, instead of the inside capitula. Hence, our methodology may not be suitable to reflect the water potential changes under near-saturation conditions. In our modelling, we used the volumetric water content of moss carpet to estimate RWC as an approximation for wet conditions (Eq. 17). The accuracy of such approximation for high RWC conditions remains ambiguous and more information is still required.

We assumed that tissue structure did not change during the measurement process, and that the aerodynamic resistance (r_a, Eq. 3) for vapor to diffuse from the inner capitula to the headspace was constant. However, capitula drying may change leaf curvature, especially in species with slim and sparsely spread leaves (Laine et al., 2018). Such changes in the branch-leaf structure could expose the more of the leaf surface to evaporation and reduce the value of r_a. Consequently, PMS could underestimate capitula water potential towards the drying end for those species, if a constant r_a is derived from the maximal evaporation rate (E_m, Eq. 5; Fig 3C).

The water-retention relationship in PCM may not sufficiently capture water potential changes at wet and dry extremes (e.g., *S. magellanicum* in Fig. 4C). Water retention functions developed for mineral soils (e.g., Clapp and Hornberge, 1978; van Genuchten, 1980) may not be well parameterized for peat soils and moss (non-vacular) vegetation, particularly under very dry or wet conditions. Hence, further studies are needed to improve the description o~~n~~f the nonlinearity of capitula water content, as influenced by capitula morphology (e.g. capitula biomass and shoot density) and structural changes of branch leaves.

PMC lacks horizontal (lateral) water transport that may allow individuals of lawn species to be present in hummocks (Rydin 1985). With additional experimental data, such as species-specific hydraulic conductivity, the current model could be improved to also quantify the horizontal water transport among neighboring grid cells.

To conclude, PMS could successfully capture the habitat preferences of the modelled *Sphagnum* species. In this respect, PMS could provide fundamental support for the future development of dynamic vegetation models for peatland ecosystems. Based on our findings, capitulum water

processes should be considered as a control on vegetation dynamics in future impact studies on peatlands under changing environmental conditions.

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Code and data availability. The data and the code to reproduce the analysis is available upon request to the corresponding author.

Author contributions. JG and EST designed the study. JG, AML and NK conducted the experiment and analysis. JG, EST, NR and SF designed the model. JG coded the model and conducted the model simulation and data analysis. JG and EST wrote the manuscript with contributions from all co-authors.

Competing interests. The authors declare that they have no conflict of interest.

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1032 Table. 1 List of symbols and abbreviations

Symbol	Description	Unit
A	Net photosynthesis rate	μmol m^{-2} s^{-1}
A_m	Maximal net photosynthesis rate	μmol m^{-2} s^{-1}
α_{imm}	Temperature constant for NSC immobilization	
α_{PPFD}	Half-saturation point of PPFD for photosynthesis.	μmol m^{-2} s^{-1}
B_{cap}	Capitulum biomass	g m^{-2}
C_T	Specific heat	J K^{-1} kg^{-1}
D_S	Capitulum density	shoots cm^{-2}
dH	Annual height growth of *Sphagnum* mosses	cm
dWT	Hummock-lawn differences in water table	cm
E	Rate of evaporation	cm $timestep^{-1}$
f_W	Water content multiplier on photosynthesis rate	
f_T	Temperature multiplier on photosynthesis rate	
h	Water potential	cm
Hc	Shoot height of *Sphagnum* mosses	cm
H_{cap}	Height of capitula	cm
H_{spc}	Biomass density of living *Sphagnum* stems	g m^{-2} cm^{-1}
I	Rate of net inflow water	cm
k_{imm}	Specific immobilization rate	g g^{-1}
JD_{thaw}	Julian day when thawing completed	
K_h	Hydraulic conductivity of peat layer	cm s^{-1}
K_m	Hydraulic conductivity of moss layer	cm s^{-1}

K_{sat}	Saturated hydraulic conductivity	cm s^{-1}
K_T	Thermal conductivity	W m^{-1} K^{-1}
lc	Width of a grid cell in simulation	cm
M_B	Immobilized NSC to biomass production	g
NSC_{max}	Maximal NSC concentration in *Sphagnum* biomass	g g^{-1}
P	Precipitation	cm
Pm	Mass-based rate of maximal gross photosynthesis	µmol g^{-1} s^{-1}
$PPFD$	Photosynthetic photon flux density	µmol m^{-2} s^{-1}
ρ_{bulk}	Bulk density of peat	g cm^{-3}
r_{aero}	Aerodynamic resistance	s m^{-1}
r_{bulk}	Cell-level bulk surface resistance	s m^{-1}
r_{ss}	Bulk surface resistance of community	s m^{-1}
Rh	Relative humidity	%
Rs	Mass-based respiration rate	µmol g^{-1} s^{-1}
R_s	Incoming shortwave radiation	W m^{-2}
R_l	Incoming longwave radiation	W m^{-2}
S_c	Area of a cell in model simulation	m^2
s_{imm}	Multiplier for temperature threshold	
Sv_i	Evaporative area of a cell i	cm^2
T	Capitulum temperature	ºC
Ta	Air temperature	ºC
T_{opt}	reference temperature of respiration (20 ºC)	ºC
u	Wind speed	m s^{-1}

W_{cap}	Capitulum water content	g g^{-1}
W_{cmp}	Capitulum water content at the compensation point	g g^{-1}
W_{max}	Maximum water content of capitula	g g^{-1}
W_{opt}	Optimal capitulum water content for photosynthesis	g g^{-1}
W_{cf}	field-water contents of *Sphagnum* capitulum	g g^{-1}
W_{sf}	field-water contents of *Sphagnum* stem	g g^{-1}
WTm	Measured multi-year mean of weekly water table	cm
WTs	Simulated multi-year mean of weekly water table	cm
z_m	Thickness of the living moss layer	cm
θ_m	Volumetric water content of moss layer	
θ_r	permanent wilting point water content	
θ_s	saturated water content	

Abbreviations:

Γ	Learning rate of gradient decedent algorithms
D-layer	Daily-based snow layer
ICOS	Integrated Carbon Observation System
JD	Julian day
NSC	Nonstructural carbon
PMS	Peatland Moss Simulator
RWC	Capitulum water content
SD	Standard deviation
SE	Standard error
SSE	Sum of squared error

SVAT	Soil-vegetation-atmosphere transport
WT	Water table

Table. 2 Species-specific values of morphological and photosynthetic parameters for *S. magellanicum* and *S. fallax*. The parameters include: capitulum density (D_S), capitulum biomass (B_{cap}), specific height of stem (H_{spc}), maximal gross photosynthesis rate at 20 ºC (Pm_{20}), respiration rate at 20 ºC (Rs_{20}), half-saturation point of photosynthesis (α_{PPFD}), and polynomial coefficients (a_{W0}, a_{W1} and a_{W2}) for the responses of net photosynthesis to capitulum water content. Parameter values are given as mean ± standard deviation.

Parameter	Unit	*S. magellanicum*	*S. fallax*	Equation
D_S	cm^{-2}	0.922±0.289	1.46±0.323	-[a]
B_{cap}	$g\ m^{-2}$	75.4±21.5	69.2±19.6	-[a]
H_{spc}	$g^{-1}\ cm^{-1}$	45.4 ± 7.64	32.6±6.97	(7)
Pm_{20}	$\mu mol\ g^{-1}\ s^{-1}$	0.0189±0.00420	0.0140±0.00212	(2)
Rs_{20}	$\mu mol\ g^{-1}\ s^{-1}$	0.00729±0.00352	0.00651±0.00236	(2)
α_{PPFD}	$\mu mol\ m^{-2}\ s^{-1}$	101.4±14.1	143±51.2	(2)
a_{W0}	unitless	-1.354±0.623	-1.046±0.129	(4)
a_{W1}	unitless	0.431±0.197	0.755±0.128	(4)
a_{W2}	unitless	-0.0194±0.0119	-0.0751±0.0223	(4)

[a] the parameter was used in the linear models predicting the $\log_{10}$-transformed capitulum water potential (h) and bulk resistance (r_{bulk}) for *S. fallax* and *S. magellanicum*. The capitulum density and photosynthetic parameter values measured here are well within the range of those reported in literature for these species (McCarter & Price, 2014; Laing et al. 2014; Bengtsson et al. 2016; Korrensalo et al. 2016).

Table 3. Parameters values for SVAT simulations (Module III). The parameters include: saturated hydraulic conductivity (K_{sat}), water retention parameters of water retention curves (α and *n)*, saturated water content ($\theta_{s)}$, permanent wilting point water content ($\theta_{r,)}$, snow layer surface albedos (a_s, a_l), the thermal conductivity (K_T), specific heat (C_T), maximal nonstructural carbon (NSC) concentration (NSC_{max}).

Parameter	Value	Equation	Source
Ksat	162	A6	McCarter and Price, 2014
n	1.43	A5	McCarter and Price, 2014
α	2.66	A5	McCarter and Price, 2014
θ_s	0.95[a]	A5	Päivänen, 1973
θ_r	0.071[b]	A5	Weiss et al., 1998
a_s	0.15	A9	Runkle et al., 2014
a_l	0.02	A10	Thompson et al., 2015
$K_{T,water}$	0.57	A4	Letts et al., 2000
$K_{T,ice}$	2.20	A4	Letts et al., 2000
$K_{T,org}$	0.25	A4	Letts et al., 2000
$C_{T,water}$	4.18	A3	Letts et al., 2000
$C_{T,ice}$	2.10	A3	Letts et al., 2000
$C_{T,org}$	1.92	A3	Letts et al., 2000
NSC_{max}	0.045	6	Turetsky et al., 2008

[a] The value was calculated from bulk density (ρ_{bulk}) as θ_s= 97.95 - 79.72ρ_{bulk} following Päivänen (1973); [b] The value was calculated as θ_r = 4.3 + *67*ρ_{bulk} following Weiss et al. (1998).

Table 4. Results from the Test 2 addressing the robustness of the model to the uncertainties in a set of parameters. Each parameter was increased or decreased by 10%. Model was run for *S. magellanicum* and *S. fallax* in their preferential habitats. Difference in mean cover between simulations under changed and unchanged parameter values are given with the standard deviations (SD) of the means in brackets. The parameters include: specific immobilization rate (*kimm*), maximal nonstructural carbon (NSC) concentration (NSC_{max}), hydraulic conductivity of moss layer (K_m), hydraulic conductivity of peat layer (K_h), water retention parameters of water retention curves (α and *n)*, snow layer surface albedo (a_s) and aerodynamic resistance (r_{aero}).

Change in parameter value	Equation	Changes in simulated cover, % (SD)	
		S. magellanicum (hummock)	*S. fallax* (lawn)
kimm +10%	5	-1.2 (3.5)	-3.5 (3.8)
kimm -10%		+2.7 (0.4)	-5.0 (3.4)
NSC_{max} +10%	6	+4.5 (2.9)	+0.7 (3.0)
NSC_{max} -10%		-0.7 (4.0)	-4.8 (4.5)
K_m +10%	10	+1.0 (3.1)	-1.7 (2.3)
K_m -10%		-1.7 (2.7)	+4.1 (4.3)
K_h +10%	A1	-1.1 (3.0)	+1.1 (2.0)
K_h -10%		-1.8 (3.1)	-0.5 (2.7)
n +10%	A5	-1.6 (3.2)	-3.2 (3.2)
n -10%		-9.4 (3.6)	-0.3 (2.9)
α +10 %	A5	-0.5 (2.9)	-0.3 (2.3)
α -10 %		-1.3 (3.6)	+3.2 (1.0)
a_s +10%	A9	-2.2 (3.8)	+0.6 (2.1)
a_s -10%		+3.3 (3.4)	+1.2 (1.8)
r_{aero} +10%	A14, A15	-2.1 (3.4)	+0.3 (2.1)
r_{aero} -10%		-3.8 (4.4)	+2.3 (1.1)

Table 5. Result from the Test 7-10 addressing the importance of meteorological fluctuations, stochasticity of model parameters and the photosynthetic water-response. In Test 7, monthly mean values of meteorological variables were used to drive the model simulation. In Test 8, the stochasticity of model parameters was removed, and average values were used to parameters at grid cell level. In Test 9-10, the photosynthetic water-response parameters (i.e. a_{W0}, a_{W1} and a_{W2}. See Table 2) were set to be the same as those in *S. magellanicum* (Test 9) and same as those in *S. fallax* (Test 10). The mean cover of *S. magellanicum* on hummocks and *S. fallax* on lawns after the simulation of 15 year periods are listed in the table.

Test	*S. magellanicum* (hummock)	*S. fallax* (lawn)
7	73%	96%
8	90%	72%
9	14 %	100 %
10	100 %	100 %

Appendix A. Calculating community SVAT scheme (Module III)

Transport of water and heat in peat profile

Simulating the transport of water and heat in the peat profiles was based on Gong et al. (2012, 2013). Here we list the key algorithms and parameters. Ordinary differential equations governing the vertical transport of water and heat in peat profiles were given as:

$$C_h \frac{\partial h}{\partial t} = \frac{\partial}{\partial z}\left[K_h\left(\frac{\partial h}{\partial z} + 1\right)\right] + S_h \quad \text{(A1)}$$

$$C_T \frac{\partial T}{\partial t} = \frac{\partial}{\partial z}\left(K_T \frac{\partial T}{\partial z}\right) + S_T \quad \text{(A2)}$$

where t is the time step; z is the thickness of peat layer; h is the water potential; T is the temperature; C_h and C_T are the specific capacity of water (i.e. $\partial\theta/\partial h$) and heat; K_h and K_T are the hydraulic conductivity and thermal conductivity, respectively; and S_h and S_T are the sink terms for water and energy, respectively.

C_T and K_T were calculated as the volume-weighted sums from components of water, ice and organic matter:

$$C_T = C_{water}\theta_{water} + C_{ice}\theta_{ice} + C_{org}(1 - \theta_{water} - \theta_{ice}) \quad \text{(A3)}$$

$$K_T = K_{water}\theta_{water} + K_{ice}\theta_{ice} + K_{org}(1 - \theta_{water} - \theta_{ice}) \quad \text{(A4)}$$

where C_{water}, C_{ice} and C_{org} are the specific heats of water, ice and organic matter, respectively; K_{water}, K_{ice} and K_{org} are the thermal conductivities of water, ice and organic matter, respectively; and θ_{water} and θ_{ice} are the volumetric contents of water and ice, respectively.

For a given h, $C_h=\partial\theta(h)/\partial h$ was derived from the van Genuchten water retention model (van Genuchten, 1980) as:

$$\theta(h) = \theta_r + \frac{(\theta_s - \theta_r)}{[1+(\alpha|h^n|)^m]} \quad \text{(A5)}$$

where θ_s is the saturated water content; θ_r is the permanent wilting point water content; α is a scale parameter inversely proportional to mean pore diameter; n is a shape parameter; and m=1-1/n.

Hydraulic conductivity (K_h) in an unsaturated peat layer was calculated as a function of θ by combining the van Genuchten model with the Mualem model (Mualem, 1976):

$$K_h(\theta) = K_{sat} S_e^{L_e}\left[1 - \left(1 - S_e^{1/m}\right)^m\right] \quad \text{(A6)}$$

where K_{sat} is the saturated hydraulic conductivity; S_e is the saturation ratio and $S_e = (\theta-\theta_r)/(\theta_s-\theta_r)$; and L_e is the shape parameter (L_e=0.5; Mualem, 1976).

Boundary conditions and surface energy balance

A zero-flow condition was assumed at the lower boundary of the peat column. The upper boundary condition was defined by the surface energy balance, which was driven by net radiation (*Rn*). The dynamics of *Rn* at surface x (*x*=0 for vascular canopy and *x*=1 for moss surface) was determined by the balance between incoming and outgoing radiation components:

$$Rn_x = Rsn_{b,x} + Rsn_{d,x} + Rln_x \quad \text{(A7)}$$

where $Rsn_{b,x}$ and $Rsn_{d,x}$ are the absorbed energy from direct and diffuse radiation; Rln_x is the absorbed net longwave radiation.

Algorithms for calculating the net radiation components were detailed in Gong et al. (2013), as modified from the methods of Chen et al. (1999). Canopy light interception was determined by the light-extinction coefficient (k_{light}), leaf area index (Lc) and solar zenith angle. The partitioning of reflected and absorbed irradiances at ground surface was regulated by the surface albedos for the shortwave (a_s) and longwave (a_l) components, and the temperature of surface *x* (T_x) also affects net longwave radiation:

$$Rn_x = Rsn_{b,x} + Rsn_{d,x} + Rln_x \quad \text{(A8)}$$

$$Rsn_{d,x} = Rs_{id,x}(1 - a_s) \quad \text{(A9)}$$

$$Rln_x = Rl_{i,x}(1 - a_l) - \varepsilon\delta T_x^4 \quad \text{(A10)}$$

where Rs_{ib}, Rs_{id}, Rl_i are the incoming beam, diffusive and longwave radiations; ε is the emissivity ($\varepsilon = 1-a_l$); δ is the Stefan Boltzmann's constant (5.67×10^{-8} W m^{-2} K^{-4}).

Rn_x was partitioned into latent heat flux (λE_x), sensible heat flux (H_x) and ground heat flux (for canopy G_1=0):

$$Rn_x = H_x + \lambda E_x + G_x \quad \text{(A11)}$$

$$G_1 = K_T\,(T_x - Ts)/(0.5z) \quad \text{(A12)}$$

where *Ts* is the temperature of the moss carpet; *z* is the thickness of the moss layer (*z* = 5 cm).

The latent heat flux was calculated by the "interactive scheme" (Daamen and McNaughton, 2000; see also in Gong et al., 2016), which is a K-theory-based, multi-source model:

$$\lambda E_x = \frac{(\Delta/\gamma)A_x r_{sa,x} + \lambda VPD_b}{r_{b,x} + (\Delta/\gamma) r_{sa,x}} \quad \text{(A13)}$$

where Δ is the slope of the saturated vapor pressure curve against air temperature; λ is the latent heat of vaporization; E is the evaporation rate; VPD_b is the vapor pressure deficit at z_b; $r_{b,x}$ is the total resistance to water vapor flow, the sum of boundary layer resistance ($r_{sa,x}$) and surface resistance (r_{ss}); and A is the available energy for evapotranspiration and $A_x = Rn_x - G_x$.

The calculations of γ, λ and VPD_b require the temperature (T_b) and vapor pressure (e_b) at the mean source height (z_b). These variables were related to the total of latent heat ($\sum \lambda E_x$) and sensible heat ($\sum H_x$) from all surfaces using the Penman-type equations:

$$\Sigma \lambda E_x = \rho_a C_p \, (e_b - e_a)/(r_{aero}\gamma) \quad \text{(A14)}$$

$$\Sigma H_x = \rho_a C_p \, (T_b - T_a)/r_{aero} \quad \text{(A15)}$$

where $\rho_a C_p$ is the volumetric specific heat of air; r_{aero} is the aerodynamic resistance between z_b and the reference height z_a, and was a function of T_b accounting for the atmospheric stability (Choudhury and Monteith, 1988); and γ is the psychrometric constant ($\gamma = \rho_a Cp/\lambda$).

Changes in the energy balance affect the surface temperature (T_x) and vapor pressure (e_x), which further feed back to the energy availability (Eq. A10, A12), the source-height temperature, VPD and the resistance parameters (e.g., r_{aero}). The values of T_x and e_x were solved iteratively by coupling the energy balance equations (eqs. A11–A15) with the Penman-type equations (see also Appendix B in Gong et al., 2016):

$$\lambda E_x = \rho_a C_p \, (e_x - e_b)/\left(r_{sa,x}\gamma\right) \quad \text{(A16)}$$

$$H_x = \rho_a C_p \, (T_x - T_b)/r_{sa,x} \quad \text{(A17)}$$

where the boundary-layer resistance for ground surface ($r_{sa,1}$) and canopy ($r_{sa,0}$) were calculated following the approaches of Choudhury and Monteith (1988).

Sink terms of transport functions for water and heat

The sink term $S_{h,i}$ (see Eq. A11) for each soil layer i was calculated as:

$$S_{h,i} = E_i - P_i - W_{melt,i} - I_i \quad \text{(A18)}$$

where E_i is the evaporation loss of water from the layer; P_i is rainfall ($P_i = 0$ if the layer is not topmost, i.e. $i>1$); $W_{melt,i}$ is the amount of melt water added to the layer; I_i is the net water inflow and was calibrated in Section 2.5.

The value of E_i was calculated as:

$$E_i = f_{top}E_0 + f_{root}(i)E_1 \quad \text{(A19)}$$

where E_0 and E_1 are the evaporation rate from ground surface and canopy (Eq. A13); f_{top} is the location multiplier for the topmost layer (f_{top} = 0 in cases i>1); and $f_{root}(i)$ is the fraction of fine-root biomass in layer i.

The value of $W_{melt,i}$ was controlled by the freeze-thaw dynamics of soil water and snow pack, which were related to the heat diffusion in soil profile (Eq. A2). We set the freezing point temperature to 0 ºC, and the temperature of a soil layer was held constant (0 ºC) during freezing or melting. For the ith soil layer, the sink term (S_T) in heat transport equation (Eq. A2) was calculated as:

$$S_{T,i} = f_{phase}max\left(|T_i|C_{T,i}, W_{phase}\lambda_{melt}\right) \quad \text{(A20)}$$

where $C_{T,i}$ is specific heat of soil layer (Eq. A13); W_{phase} is the water content for freezing (W_{phase} = θ_w) or melting (W_{phase} = θ_{ice}); λ_{melt} is the latent heat of freezing; f_{phase} is binarial coefficient that denotes the existence of freezing or thawing. For each time step t, we computed $T_i(t)$ with a piror assumption that $S_{T,i}$=0. Then f_{phase} was determined by whether the temperature changed across the freezing point, i.e. f_{phase}=1 if $T_i(t)$*$T_i(t\text{-}1) \leq 0$, otherwise f_{phase}= 0.

Parameterization of SVAT processes

For the calculation of surface energy balance, we set the height and leaf area of vascular canopy to 0.4 m and 0.1 m^2 m^{-2}, consistent with the scarcity of vascular canopies at the site. The aerodynamic resistance (r_{aero}, Eq. A14, Appendix A) for surface energy fluxes was calculated following Gong et al. (2013a). The bulk surface resistance of community (r_{ss}, Eq. A13, Appendix A) was summarized from the cell-level values of $r_{bulk,i}$, that $1/r_{ss} = \sum(1/r_{bulk,i})$. To calculate the peat hydrology and water table, peat profiles of hummock and lawn communities were set to 150 cm deep and stratified into horizontal layers of depths varying from 5cm (topmost) to 30cm (deepest). For each peat layer, the thermal conductivity (K_T) of fractional components, i.e. peat, water and ice, were evaluated following Gong et al. (2013a). The bulk density of peat (ρ_{bulk}) was set to 0.06 g cm^{-3} below acrotelm (40 cm depth, Laine et al., 2004), and decreased linearly toward the living moss layer. The saturated hydraulic conductivity (K_{sat}, Eq. A6, Appendix A) and water retention parameters (i.e. α and n, Eq. A5, Appendix A) of water retention curves were calculated as functions of ρ_{bulk} and the depth of peat layer following Päivänen (1973). K_{sat}, α and n for the living moss layer were adopted from the values measured by McCarter and Price (2014) from *S. magellanicum* carpet. The parameter values for SVAT processes are listed in Table 3.

Calculation of snow dynamics

In boreal and arctic regions, the amount and timing of snow melt has crucial impact on moisture

conditions, especially at fen peatlands. Therefore, to have realistic spring conditions we introduced a snow-pack model, SURFEX v7.2 (Vionnet et al., 2007), into the SVAT modelling. The snow-pack model simulates snow accumulation, wind drifting, compaction and changes in metamorphism and density. These processes influenced the heat transport and freezing-melting processes (i.e. S_h and S_T, see Eq. A1-A2, Appendix A). In this modelling, we calculate the snow dynamics on a daily basis in parallel to the SVAT simulation. Daily snowfall was converted into a snow layer and added to ground surface. For each of the day-based snow layers (D-layers), we calculated the changes in snow density, particle morphology and layer thicknesses. At each time step, D-layers were binned into layers of 5-10 cm depths (S-layers) and placed on top of the peat column for SVAT modelling. With a snow layer present, surface albedos (i.e. a_s, a_l) were modified to match those of the topmost snow layer (see Table 4 in Vionnet et al., 2007). If the total thickness of snow was less than 5 cm, all D-layers were binned into one S-layer. The thermal conductivity (K_T), specific heat (C_T), snow density, thickness and water content of each S-layer were calculated as the mass-weighted means from the values of D-layers. Melting and refreezing tended to increase the density and K_T of a snow layer but decrease its thickness (see Eq. 18 in Vionnet et al., 2007). The fraction of melted water that exceeded the water holding capacity of a D-layer (see Eq. 19 in Vionnet et al., 2007) was removed immediately as infiltration water. If the peat layer underneath was saturated, the infiltration water was removed from the system as lateral discharge.

Boundary conditions and driving variables

A zero-flow boundary was set at the bottom of peat. At peat surface the boundary conditions of water and energy were defined by the ground surface temperature (T_0, see Eq. A10-A15 in Appendix A) and the net precipitation (*P* minus *E*). The profiles of layer thicknesses, ρ_{bulk} and hydraulic parameters were assumed to be constant during simulation. Lateral boundary conditions were used to calculate the spreading of *Sphagnum* shoots among cells along the edge of the model domain so that shoots can spread across the edge of simulation area and invade into the grid cell at the boarder of the opposite side.

The model simulation was driven by climatic variables of air temperature (*Ta*), precipitation (P), relative humidity (*Rh*), wind speed (*u*), incoming shortwave radiation (*Rs*) and longwave radiation (*Rl*). To support the stochastic parameterization of the model and Monte-Carlo simulations, Weather Generator (Strandman et al., 1993) was used to generate randomized scenarios based on long-term weather statistics (period of 1981-2010) from the four closest weather stations of the Finnish Meteorological Institute. This generator had been intensively tested and applied under Finnish conditions (Kellomäki and Väisänen, 1997; Venäläinen et al., 2001; Alm et al., 2007). We also compared the simulated meteorological variables against 2-year data measured from Siikaneva peatland site (61°50 N; 24°10 E), located 10 km away from our study site (Appendix C).

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Appendix B. Methods and results of the empirical study on *Sphagnum* capitula water retention as a controlling mechanism for peatland moss community dynamics

Measurement of morphological traits

To quantify morphological traits, samples of *S. fallax* and *S. magellanicum* were collected at the end of August 2016 with a core (size d 7cm, area 50 cm^2, height at least 8 cm) maintaining the natural density of the stand. Samples were stored in plastic bags at cool room (4 ºC) until measurements. Eight replicates were collected for each species. For each sample, capitulum density (D_S, shoots cm^{-2}) was measured and ten moss shoots were randomly selected and separated into capitula and stems (5 cm below capitula). The capitula and stems were moistened and placed on top of a tissue paper for 2 minutes to extract free-moving water, before weighing them for water-filled fresh weight. The samples were dried at 60 °C for at least 48h to measure the dry masses. The field-water contents of capitula (W_{cf}, g g^{-1}) and stems (W_{sf}, g g^{-1}) were then calculated as the ratio of water to dry mass for each sample. The biomass of capitula (B_{cap}, g m^{-2}) and living stems (B_{st}, g m^{-2}) were calculated by multiplying the dry masses with the capitulum density (D_S). Biomass density of living stems (H_{spc}, g cm^{-1} m^{-2}) was calculated by dividing B_{st} with the length of stems.

Measurement of photosynthetic traits

We measured the photosynthetic light response curves for *S. fallax* and *S. magellanicum* with fully controlled, flow-through gas-exchange fluorescence measurement systems (GFS-3000, Walz, Germany; Li-6400, Li-Cor, US) under varying light levels. In 2016, measurements on field-collected samples were done during May and early June, which is a peak growth period for *Sphagna* (Korrensalo et al. 2017). Samples were collected from the field site each morning and were measured the same day at Hyytiälä field station. Samples were stored in plastic containers and moistened with peatland water to avoid changes in plant status during the measurement. Right before the measurement we separated *Sphagnum* capitula from their stems and dried them lightly using tissue paper before placing an even layer of them in a custom-made cuvette by retaining the same density as naturally at field (Korrensalo et al. 2017). Net photosynthesis rate (A, µmol ~~mg~~$^{-21}$ s^{-1}) was measured at 1500, 250, 35, and 0 µmol m^{-2} s^{-1} photosynthetic photon flux density (*PPFD*) (Fig 1B). The light levels were chosen based on previous investigation by Laine et al. (2011, 2015), which showed increasing A until PPFD at 1500 and no photoinhibition even at high values of 2000 µmol m^{-2} s^{-1}. The samples were allowed to adjust to cuvette conditions before the first measurement and after each change in the PPFD level until the CO_2 rate had reached a steady level, otherwise the cuvette conditions were kept constant (temperature 20°C, CO_2 concentration 400 ppm, flow rate 500 umol s^{-1}, impeller at level 5 and relative humidity of inflow air 60%, yet the relative humidity remained on average 81% during the measurements). The time required for a full measurement cycle varied between 60 and 120

minutes. Each sample was weighed before and after the gas-exchange measurement, then dried at 40°C for 48 h to determine the biomass of capitula (B_{cap}). For each species, five~~our~~ samples were measured as replicates and were made to fit a hyperbolic light-saturation curve (Larcher, 2003):

$$A_{20} = \left(\frac{Pm_{20} * PPFD}{\alpha_{PPFD} + PPFD} - Rs_{20}\right) * B_{cap} \quad \text{(B1)}$$

where subscript 20 denotes the variable value measured at 20 °C; *Rs* is the mass-based dark respiration rate (µmol g^{-1} s^{-1}); *Pm* is the mass-based rate of maximal gross photosynthesis (µmol g^{-1} s^{-1}); and α_{PPFD} is the half-saturation point (µmol m^{-2} s^{-1}), i.e., PPFD level where half of *Pm* is reached. The measured morphological and photosynthetic traits are listed in Table 2.

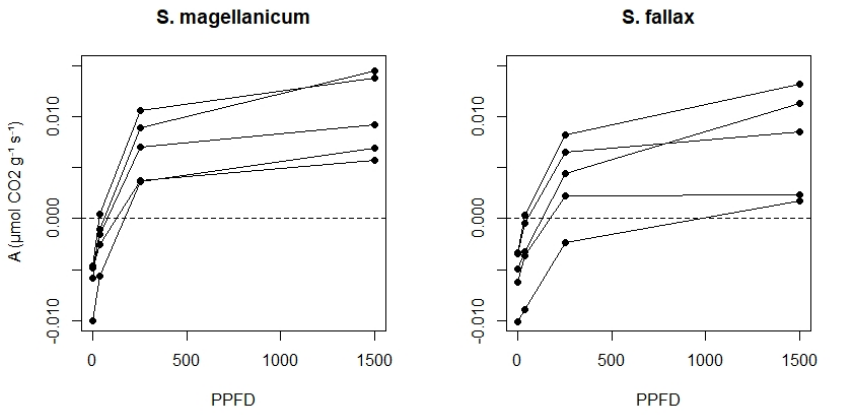

Figure B1. Measured light response curves for *S. magellanicum* and *S. fallax*.

Drying experiment

To link the water retention and photosynthesis of *Sphagnum* capitula, we performed a drying experiment using a GFS-3000 system to measure co-variations of capitulum water potential (*h*, cm water), water content (W_{cap}, g g^{-1}) and *A* (µmol ~~m~~$^{-2}$ g^{-1} s^{-1}). For both species, four mesocosms were collected in August 2018 and transported to laboratory in UEF Joensuu, Finland. Capitula were harvested and wetted by water from the mesocosms. The capitula were then placed gently on a piece of tissue paper for 2 minutes and then placed into the same cuvette as used in the previous photosynthesis measurement. The cuvette was then placed into GFS and measured under constant conditions of *PPFD* (1500 umol m-2 s-1), temperature (293.2K), inflow air (700 umol s-1), CO_2 concentration (400 ppm) and relative humidity (40%). Measurement was stopped

when A dropped to less than 10% of its maximum. Each measurement lasted between 120 and180 minutes. Each sample was weighed before and after the gas-exchange measurement, then dried at 40°C for 48 h to determine the biomass of capitula (B_{cap}).

The GFS-3000 records the vapor pressure (e_a, kPa) and the evaporation rate (E, g s^{-1}) simultaneously with A at every second (Heinz Walz GmbH, 2012). The changes in W_{cap} with time (t) was calculated as following:

$$RWC(t) = \left(W_{pre} - B_c - \sum_{t=0}^{t} E(t)\right)/B_c \quad \text{(B2)}$$

We assumed that the vapor pressure at the surface of water-filled cells equaled the saturation vapor pressure (e_s), and the vapor pressure in the headspace of cuvette equaled that in the outflow (e_a). The vapor pressure in capitula pores (e_i) thus can be calculated based on following gradient-transport function (Fig. B2A):

$$\lambda E(t) = \frac{\rho_a C_p}{\gamma} \frac{(e_i(t) - e_a(t))}{r_a(t)} = \frac{\rho_a C_p}{\gamma} \frac{(e_s - e_i(t))}{r_s(t)} \quad \text{(B3)}$$

where λ is the latent heat of vaporization; γ is the slope of the saturation vapor pressure - temperature relationship; r_a is the aerodynamic resistance (m s^{-1}) for vapor transport from inter-leaf volume to headspace; r_s is the surface resistance of vapor transport from wet leaf surface to inter-leaf volume. The bulk resistance for evaporation (r_{bulk}) was thus calculated as r_a+r_s.

We assumed that the structures of tissues and pores did not change during the drying process and assumed r_a to be constant during each measurement. A tended to increase with time t until it peaked (A_m) and then decreased (Fig. 2~~1~~B). The point A=A_m implied the water content where further evaporative loss would start to drain the cytoplasmic water, leading to the decrease in A. The response of A to W_{cap} was fitted as a second-order polynomial function (Robroek et al., 2009) using data from t_{Am} to t_n:

$$f_A(W_{cap}) = a_{W0} + a_{W1} * W_{cap} + a_{W2} * W_{cap}^2 \quad \text{(B4)}$$

where a_{W0}, a_{W1} and a_{W2} are parameters; and $f_A(W_{cap}) = A/A_m$. For each replicate, the optimal water content for photosynthesis (W_{opt}) was derived from the peak of fitted curve (Eq. 4). The capitulum water content at the compensation point W_{cmp}, where the rates of gross photosynthesis and respiration are equal, can be calculated from the point A=0.

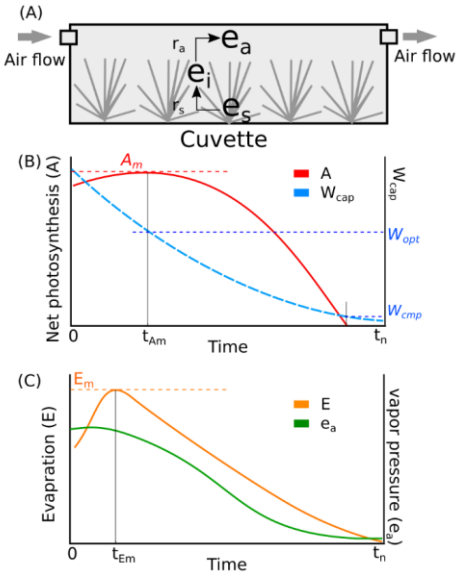

Figure B1~~2~~. Conceptual schemes of (A) cuvette setting and resistances, (B) the co-variations of net photosynthesis and W_{cap}, and (C) the co-variations of evaporation and vapor pressure in headspace during a measurement. Meanings of symbols: e_a, vapor pressure in headspace of cuvette (kPa); e_i, vapor pressure in branch-leaf structure of capitula; e_s, vapor pressure at the surface of wet tissues; r_a, aerodynamic resistance of vapor diffusion from inner capitula to headspace; r_s, surface resistance of vapor diffusion from wet tissue surface to inner capitula space; A, net photosynthesis rate (μmol m^{-2} s^{-1}); A_m, maximal net photosynthesis rate (μmol m^{-2} s^{-1}); W_{cap}, water content of capitula (g g^{-1}); W_{opt}, W_{cap} at A=A_m; W_{cmp}, W_{cap} at A=0; E, evaporation rate (mm s^{-1}).

Similarly, the evaporation rate (E) increased from the start of measurement until maximum evaporation E_m, and then decreased (Fig. B2~~1~~C). The point E=E_m implied the time when the wet capitulum tissues were maximally exposed to the air flow. Therefore, r_a was estimated as the minimum of bulk resistance using Eq. (B5), by assuming $e_i(t)$≈e_s when $E(t)$ = E_m:

$$r_a = \frac{\rho_a C_p}{\gamma}\frac{(e_s - e_a(t))}{\lambda E_m} \quad \text{(B5)}$$

Based on the calculated $e_i(t)$, we were able to derive the capitulum water potential (h) following the equilibrium vapor-pressure method (e.g. Price et al, 2008; Goetz and Price, 2015):

$$h = \frac{RT}{Mg} ln\left(\frac{e_i}{e_s}\right) + h_0 \quad \text{(B6)}$$

where R is the universal gas constant (8.314 J mol^{-1} K^{-1}); M the molar mass of water (0.018 kg mol^{-1}); g is the gravitational acceleration (9.8 N kg^{-1}); e_i/e_s is the relative humidity; h_0 is the water potential due to the emptying of free-moving water before measurement (set to 10 kPa according to Hayward and Clymo, 1982).

Statistical analysis

The light response curve (Eq. B1) and the response function of A/A_m to W_{cap} changes (Eq. B4) were fitted using nlme package in R (version 3.1). The obtained values of shape parameters a_{W0}, a_{W1} and a_{W2} (Eq. 4) were then used to calculate W_{opt} (W_{opt} = -0.5 a_{W1}/ a_{W2}) and W_{cmp} (W_{cmp} = 0.5 [-a_{W1} - (a_{W1}^2 - 4a_{W0} a_{W2})$^{0.5}$] / a_{W2}). We then applied ANOVA to compare *S. magellanicum* against *S. fallax* for the traits obtained from the field sampling (i.e. structural properties such as B_{cap}, D_S, H_{spc}, W_{cf}, W_{sf}) and from the gas-exchange measurements (i.e. Pm_{20}, Rs_{20}, W_{opt}, W_{cmp} and r_{bulk}), using R (version 3.1).

The measured values of capitulum water potential (h) were $\log_{10}$-transformed and related to the variations in W_{cap}, B_{cap} and D_S with a linear model. Similarly, a linear model was established to quantify the response of bulk resistance for evaporation (r_{bulk}) ($\log_{10}$-transformed) to the variations in h, B_{cap} and D_S. The linear regressions were based on statsmodels (version 0.9.0) in Python (version 2.7), as supported by Numpy (version 1.12.0) and Pandas (version 0.23.4) packages.

Results of the empirical measurements

The two *Sphagnum* species differed in their structural properties (Table B1). Lawn species S. *fallax* had looser structure than hummock species *S. magellanicum* as seen in lower capitulum density (D_S) and specific height (H_{spc}) in *S. fallax* than in *S. magellanicum* (P<0.05, Table. B1). Moreover, in conditions prevailing in the study site *S. fallax* mosses were dryer than *S. magellanicum*; the field-water contents of *S. fallax* capitulum (W_{cf}) and stem (W_{sf}) were 40% and 46% lower than *S. magellanicum* (P<0.01, Table. B1), respectively. The different density of capitulum of the two species differing in their capitulum size led to similar capitulum biomass (B_{cap}) (P=0.682) between *S. fallax* with small capitulum and *S. magellanicum* with large capitulum. Unlike the structural properties, maximal CO_2 exchange rates (Pm_{20} and Rs_{20}) did not differ between the two species (Table B1).

The drying experiment demonstrated how capitulum water content regulated capitulum processes in both studied *Sphagnum* species (Fig. B~~2~~3). Decreasing capitulum water content (W_{cap}) led to decrease in the water potential (h), the responses of h to W_{cap} varied among replicates (Fig. ~~2~~3A). The values of W_{cap} for *S. fallax* were generally lower than those for *S. magellanicum* under the same water potentials. The fitted linear models explained over 95% of the variations in the measured h for both species (Table. B2), although fitted responses of h to W_{cap} were slightly smoother than the measured ones, particularly for *S. magellanicum* (Fig. B~~2~~3A). The responses of h to W_{cap} was significantly affected by the capitulum density (D_S), capitulum biomass (B_{cap}) and their interactions with W_{cap} (Table. B2).

Decreasing capitulum water content (W_{cap}), and water potential (h), were associated with

increasing bulk resistance for evaporation (r_{bulk}, Fig. B~~2~~3B), although the sensitivity of r_{bulk} to h changes varied by replicates. The values of r_{bulk} from *S. fallax* were largely lower than those from *S. magellanicum* when the capitulum water content of the two species were similar. The fitted linear models explained the observed variations in the measured r_{bulk} well for both species (Fig. 2B and Table. B3). The variation in the response of r_{bulk} to h was significantly affected by capitulum density (D_S), capitulum biomass (B_{cap}) and their interactions with h (Table. B3).

Decreasing capitulum water content (W_{cap}) slowed down the net photosynthesis rate (Fig. B2C), as represented by the decreasing ratio of A/A_m. *S. fallax* required lower capitulum water content (W_{cap}) than *S. magellanicum* to reach photosynthetic maximum and photosynthetic compensation point. However, the ranges of capitulum water content from photosynthetic maximum (W_{opt}) or field capacity (W_{fc}) to that at compensation point (W_{cmp}) were smaller for *S. fallax* than *S. magellanicum*. Hence, *S. fallax* had narrower transition zone for photosynthesis to respond to drying, compared to *S. magellanicum*.

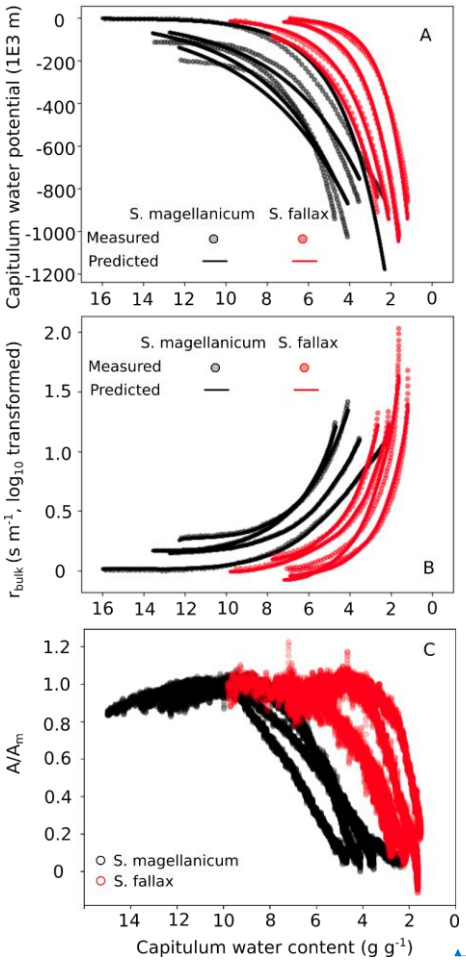

Figure B23. Responses of (A) capitulum water potential, (B) bulk resistance of evaporation, and (C) net photosynthesis to changes in capitulum water content (W_{cap}) of two *Sphagnum* species typical to hummocks (*S. magellanicum*, black) and lawns (*S. fallax*, red). As the measured results are based on the drying experiment starting with fully wetted capitula characteristic for both species, the X-axis is shown from high to low W_{cap}. The values predicted in (B) and (C) are based on linear models with parameter values listed in Tables B2 and B3 and predictor values from the drying experiment.

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Table. B1 Species-specific traits of morphological, photosynthetic and water-retention from *S. magellanicum* and *S. fallax*. Trait values (mean ± standard deviation) and ANOVA statistics F- and p-values are given for comparing the means of traits of the two species.

Trait	*S. magellanicum*	*S. fallax*	F	P (>F)
Capitulum density, D_S (capitula cm^{-2})	0.922±0.289	1.46±0.323	6.224[a]	0.037 *
Capitulum biomass, B_{cap} (g m^{-2})	75.4±21.5	69.2±19.6	0.181[a]	0.682
Specific height, H_{spc} (cm g^{-1} m^{-2})	45.4 ± 7.64	32.6±6.97	6.126[a]	0.038*
Field water content of capitula, W_{cf} (g g^{-1})	14.7±3.54	8.09±1.48	11.75[a]	0.009**
Field water content of stems, W_{sf} (g g^{-1})	18.4±1.92	10.2±1.50	45.81[a]	0.0001**
Maximal gross photosynthesis rate at 20 ºC, Pm_{20} (µmol g^{-1} s^{-1})	0.019±0.004	0.014±0.002	3.737[b]	0.101
Respiration rate at 20 ºC, Rs_{20} (µmol g^{-1} s^{-1})	0.007±0.004	0.007±0.002	0.012[b]	0.92
half-saturation point of photosynthesis, α_{PPFD} (µmol ~~g~~m^{-2} s^{-1})	101.4±14.1	143±51.2	2.856[b]	0.142
Optimal capitulum water content for photosynthesis, W_{opt} (g g^{-1})	9.41±0.73	5.81±1.68	11.57[b]	0.0145*
Capitulum water content at photosynthetic compensation point, W_{cmp} (g g^{-1})	3.67±0.83	1.78±0.43	12.35[b]	0.0126*
Minimal bulk resistance of evaporation, r_a (m s^{-1})	33.5±7.30	40.7±4.99	1.976[b]	0.2165

[a] soil-core measurement, sample n=5; [b] cuvette gas-exchange measurement, sample n=4; * the difference of means is significant (P<0.05); ** the difference of means is very significant (P<0.01).

Table B2. Parameter estimates of the linear model for the $\log_{10}$-transformed capitulum water potential (h) for *S. fallax* and *S. magellanicum*. Estimate value, standard error (SE), and test statistics p-values are given to the predictors of the models. Predictors are: capitulum biomass (B_{cap}), capitulum density (D_S), capitulum water content (W_{cap}), the interaction of capitulum biomass and water potential ($B_{cap}\times W_{cap}$), the interactions of capitulum biomass and capitulum density ($D_S\times W_{cap}$), the interactions of capitulum density and water potential ($D_S\times W_{cap}$), and the interaction of capitulum biomass, capitulum density and water potential ($B_{cap}\times D_S\times W_{cap}$). All coefficient values are significantly different from 0 ($p<0.001$).

Parameter	*S. magellanicum* (R^2=0.972)		*S. fallax* (R^2=0.984)	
	Value	SE	Value	SE
(Intercept)	25.30	0.253	-90.99	2.158
B_{cap}	-272.10	3.133	2294.67	52.342
W_{cap}	-9.50	0.031	-62.12	0.600
$B_{cap}\times W_{cap}$	114.61	0.387	1500.26	14.549
D_S	-21.76	0.253	104.11	2.376
$B_{cap}\times D_S$	268.95	3.112	-2422.79	55.251
$D_S\times W_{cap}$	9.33	0.031	68.35	0.661
$B_{cap}\times D_S\times W_{cap}$	-113.33	0.386	-1588.06	15.360

Table B3. Parameter estimates of the linear model for the $\log_{10}$-transformed capitulum evaporative resistance (r_{bulk}) for *S. fallax* and *S. magellanicum*. Estimate value, standard error (SE), and test statistics p-values are given to the predictors of the models. Predictors are: capitulum biomass (B_{cap}), capitulum density (D_S), water potential (h), the interaction of capitulum biomass and water potential ($B_{cap}\times h$), the interactions of capitulum biomass and capitulum density ($D_S\times h$), the interactions of capitulum density and water potential ($D_S\times h$), and the interaction of capitulum biomass, capitulum density and water potential ($B_{cap}\times D_S\times h$). All coefficient values are significantly different from 0 ($p<0.001$).

Parameter	*S. magellanicum* (R^2=0.998)		*S. fallax* (R^2=0.966)	
	Value	SE	Value	SE
(Intercept)	-1.13	0.027	55.07	2.225
B_{cap}	14.45	0.334	1334.55	53.968
h	0.0012	5.92e-05	-0.028	0.004
$B_{cap}\times h$	-0.0007	0.001	0.707	0.101
D_S	1.08	0.027	-60.53	2.450
$B_{cap}\times D_S$	-13.39	0.333	1406.36	56.968
$D_S\times h$	0.0002	5.89e-05	0.0317	0.005
$B_{cap}\times D_S\times h$	-0.0017	0.001	-0.733	0.106

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Appendix C. Comparisons of meteorological variables simulated by Weather Generator and those measured from Siikaneva peatland site (ICOS site located in 10 km distance from the study site Lakkasuo)

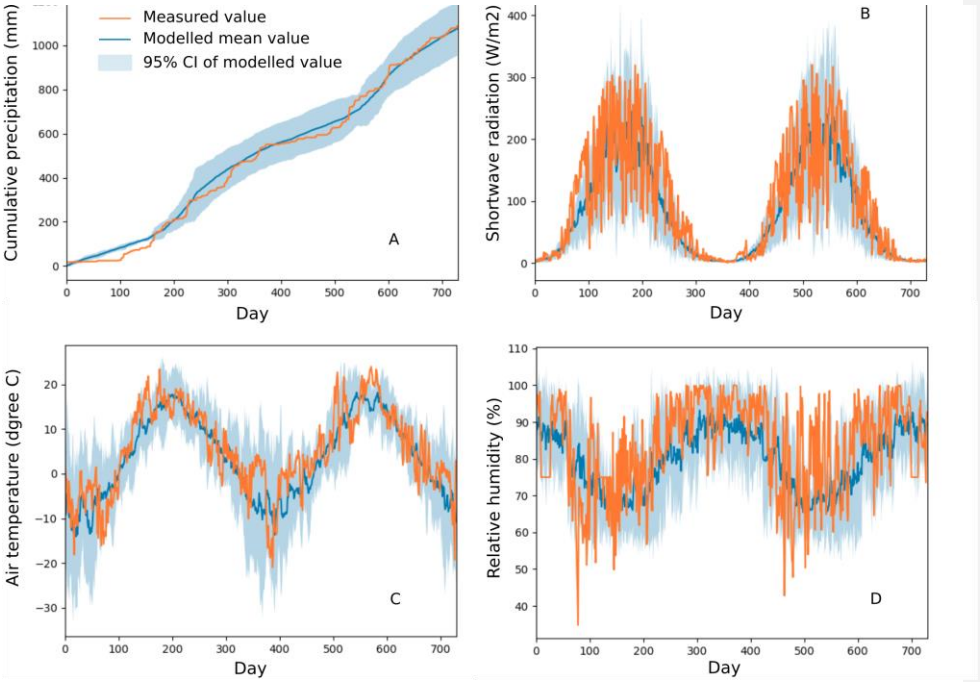

Fig. C1 Comparisons of meteorological variables simulated by Weather Generator and those measured from Siikaneva peatland site. The variables include (A) cumulative precipitation (mm), (B) incoming shortwave radiation (W m^{-2}), (C) air temperature (°C), and (D) relative humidity (%). These variables were measured and simulated at half-hourly timescale. The measurements were carried out during 2012-2013. Details about the site and measurements have been described by Alekseychik et al. (2018). The measured seasonal dynamics of the meteorological variables were generally in line with the 95% confidence intervals (CI) of the simulated values, which were calculated based on Monte-Carlo simulations (n=5).

Appendix D. Comparisons of seasonal water table measured from the study site and the values simulated based on calibrated net inflow

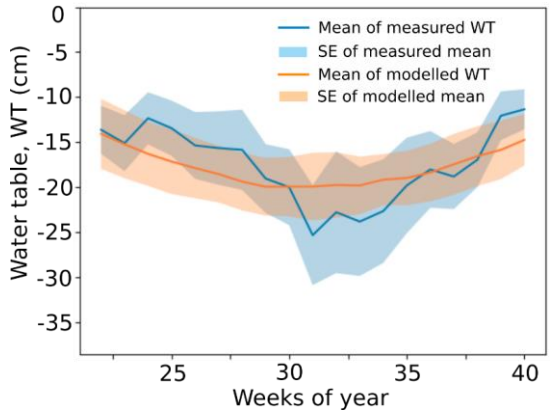

Fig. D1 Comparison of seasonal water table (WT) measured at the Lakkasuo study site and the values simulated by the calibrated PCS. WT values were sampled weekly from the lawn habitats both in field and in model output. The weekly mean WT was measured during 2001, 2002, 2004 and 2016. The modelled means and standard deviations (SD) of WT were based on 20 Monte-Carlo simulations. The simulated seasonality of mean WT generally followed the measured trends. The calibration reduced the sum of squared error (*SE*, Eq. 12) from 199.5 (a_N=b_N=0) to 117.3. The calibrated values for a_N and b_N were $-5.3575*10^{-4}$ and $4.7599*10^{-5}$, respectively (Eq. A18).