



1 Elemental composition of invertebrates shells composed of different CaCO<sub>3</sub> polymorphs at  
2 different ontogenetic stages: a case study from the brackish Gulf of Gdansk (the Baltic Sea)

3

4 Anna Piwoni-Piórewicz<sup>1\*</sup>, Stanislav Strekopytov<sup>2†</sup>, Emma Humphreys-Williams<sup>2</sup>, Piotr  
5 Kukliński<sup>1,3</sup>

6

7 <sup>1</sup>Institute of Oceanology, Polish Academy of Sciences, Powstańców Warszawy 55, 81-712  
8 Sopot, Poland

9 <sup>2</sup>Imaging and Analysis Centre, Natural History Museum, Cromwell Road, London SW7 5BD,  
10 United Kingdom

11 <sup>3</sup>Department of Life Sciences, Natural History Museum, Cromwell Road, London SW7 5BD,  
12 United Kingdom

13

14 \*Corresponding author

15 E-mail: [apiwoni@iopan.pl](mailto:apiwoni@iopan.pl)

16 Tel.: (+ 48 58) 731 16 96

17 † Present address: Inorganic Analysis, LGC Ltd, Queens Road, Teddington, United Kingdom

18

19 Abstract

20 In this study, the concentrations of 12 metals: Ca, Na, Sr, Mg, Ba, Mn, Cu, Pb, V, Y, U and  
21 Cd in shells of bivalve molluscs (aragonitic: *Cerastoderma glaucum*, *Mya arenaria* and  
22 *Limecola balthica* and bimineralic: *Mytilus trossulus*) and arthropods (calcitic: *Amphibalanus*  
23 *improvisus*) were obtained. The main goal was to determine the incorporation patterns of  
24 shells built with different calcium carbonate polymorphs. The role of potential biological  
25 control on the shell chemistry was assessed by comparing the concentrations of trace elements  
26 between younger and older individuals (different size classes). The potential impact of  
27 environmental factors on the observed elemental concentrations in the studied shells is  
28 discussed. Specimens were collected from brackish waters of the Baltic Sea (the Gulf of  
29 Gdansk). For every species, 40 individuals (ten in each size class) were selected. Pre-cleaned  
30 shells were analysed by ICP-OES and ICP-MS to determine the concentrations of metals.

31 The distributions of elements both differ between species and exhibit high  
32 intraspecific variability. Calcitic shells preferentially incorporated Mg > Sr > Na, aragonitic  
33 shells incorporated Na > Sr > Mg, and bimineralic shells accumulated Na approximately two



34 times more intensively, than Mg and Sr which remained at similar levels. Among all species,  
35 the calcitic shells of *A. improvisus* most effectively concentrated the majority of the studied  
36 elements, especially  $Mg > Mn > Ba$ , which was contrary to the shells of aragonitic molluscs  
37 that contained the lowest levels of trace elements. The size-dependent distributions of  
38 elements in shells did not exhibit a consistent pattern. The highest significant differences were  
39 found for the bimineralic shells of *M. trossulus*, while the smallest were found for aragonitic  
40 shells; if any variability occurred, it was observed in heavy metals (Pb, Cd).

41 Our results indicate that elemental variability, especially that of Mg and Sr, is  
42 dominated by the properties of the crystal lattice. The inconsistent variability of trace element  
43 concentrations between species and within single populations supports the important role of  
44 species-specific biological control of the biomineralization process and indicates that  
45 environmental factors have a significant influence on the incorporation of trace elements into  
46 the shells.

47

48 Key words: calcite, aragonite, trace metals, ontogeny, shell, ICP-MS, ICP-OES

49

## 50 1. Introduction

51 Marine invertebrates such as molluscs (bivalves, gastropods), brachiopods, corals,  
52 echinoderms, bryozoans and some groups of protozoa (foraminifera) are able to precipitate  
53 shells and skeletons as composite materials consisting of calcium carbonate ( $CaCO_3$ )  
54 crystallised on a species-specific organic matrix (Cusack and Freer, 2008). In a living  
55 organism,  $CaCO_3$  is deposited mainly in the form of two polymorphs, calcite and aragonite,  
56 which are commonly found co-existing in the same specimens and appear to be precipitated in  
57 the same environments (Morse et al., 2007; Taylor et al., 2008). Trace elements are  
58 incorporated into the crystal lattice from both the ambient seawater and ingested food  
59 particles (Cubadda et al., 2001). Biological carbonates determine the largest component of the  
60 hydrosphere's carbon reservoir, and the biogenic production of  $CaCO_3$  has a strong  
61 interdependence on the ocean composition, biogeochemistry and the carbon cycle (Cohen and  
62 McConnaughey, 2003). In the last few decades, increasing attention has been paid to the  
63 relationship between the elemental compositions of shells and external environmental factors.  
64 It has been repeatedly observed that the crystal layers are precipitated successively at regular  
65 periodicities, recording the conditions in which the calcification occurred (Marchitto et al.,  
66 2000; Rodland et al., 2006). The elemental composition of the crystal lattice can provide  
67 records of seawater chemistry and has great potential for use in the interpretation of patterns



68 of ocean history (Freitas et al., 2006; Gillikin et al., 2006; Khim et al., 2003; Ponnurangam et  
69 al., 2016; Vander Putten et al., 2000). However, as recent studies have indicated (Dove,  
70 2010), the influence of environmental parameters on shell precipitation could be very  
71 complex. The mineralogy and chemistry of shells are likely to be not only linked to  
72 environmental conditions but are also simultaneously controlled by the organism itself. Even  
73 a single population or closely related species within the same habitat may exhibit different  
74 accumulation strategies (Rainbow et al. 2000).

75         Calcite and aragonite differ in structure, symmetry and properties. The calcium ion has  
76 9-fold coordination in aragonite and 6-fold coordination in calcite, and its coordination creates  
77 hexagonal and cubic packing (Putnis, 1992). Due to the spatial structure, larger cations, such  
78 as Sr, fit better in aragonite, while smaller cations, such as Mg, are energetically favoured in  
79 calcite. In natural systems, calcite commonly incorporates Mg (Morse et al., 2007; Reeder,  
80 1983; Wang and Xu, 2001), and increasing Mg concentrations are known to increase the  
81 calcite solubility (Kuklinski and Taylor, 2009; Smith et al., 2006). At the temperatures and  
82 pressures of the Earth's surface, low-Mg calcite is the most stable form of CaCO<sub>3</sub> (de Boer,  
83 1977). Nevertheless, in many marine organisms, aragonite and high-Mg calcite are the  
84 dominant phases precipitated from seawater (Dickson, 2004). The solution chemistry (Cusack  
85 and Freer, 2008), temperature (Balthasar and Cusack, 2015), pressure (Allison et al., 2001),  
86 CaCO<sub>3</sub> saturation state (Watson et al., 2012), pCO<sub>2</sub> (Lee and Morse, 2010) and phylogenesis  
87 (Kuklinski and Taylor, 2009; Smith et al., 1998; Smith and Girvan, 2010) are known to  
88 influence shell mineralogy. The main driving force controlling the mineralogy of precipitated  
89 CaCO<sub>3</sub> is the ratio of Mg to Ca ions in seawater (Cusack and Freer, 2008; Morse et al., 2007)..  
90 A Mg/Ca > 2 favours the precipitation of aragonite and high-Mg calcite. At high Mg/Ca  
91 ratios, such as those in modern seawater (Mg/Ca = 5.2), calcitic structures incorporate Mg,  
92 which is observed to inhibit calcite nucleation and growth, whereas aragonite nucleation is not  
93 affected by magnesium in solution (De ChoudensSanchez and Gonzalez, 2009; Morse et al.,  
94 2007). Most low-Mg calcite-producing taxa are known to actively control the amount of  
95 magnesium incorporated. Since the Mg/Ca ratio of the calcifying fluid is dissimilar to that of  
96 ambient seawater, organisms might be unaffected by changes in the environmental Mg/Ca  
97 ratio (Bentov and Erez, 2005; De Nooijer et al., 2014). Several studies of organisms that  
98 secrete calcareous skeletons have shown that lower seawater temperatures are correlated with  
99 the secretion of calcite skeletons with low Mg contents, rather than more soluble high-Mg  
100 calcite or aragonite skeletons (Taylor and Reid, 1990). The mineralogy of many calcareous  
101 structures changes with latitude, likely as a result of the temperature gradient from the poles



102 to the equator (Kuklinski and Taylor, 2009; Loxton et al., 2014; Taylor et al., 2014).  
103 However, the secretion of calcite by marine organisms inhabiting low-temperature areas and  
104 aragonite in warmer waters is not an absolute rule, as many polar molluscs form fully  
105 aragonitic shells (Cairns and Macintyre, 1992). Thermodynamics predicts that aragonite is the  
106 stable phase at pressures higher than 5000 hPa (roughly 40 m depth), and calcite is the stable  
107 phase at lower pressures. However, aragonite is the major constituent of shells or pearls,  
108 indicating its metastable formation in shallow waters (Sunagawa et al., 2007). The  
109 incorporation of Sr was suggested to play a significant role in the biomineralogical  
110 precipitation of aragonite (Allison et al., 2001). Many studies have demonstrated a clear  
111 correlation between the concentration of Sr in the hard parts and precipitation of the aragonite  
112 layer (Iglukowska et al., 2016; Reeder, 1983). The ionic radius of Sr is larger than that of Ca;  
113 thus, Sr is more likely to form 9-fold coordination, which triggers metastable aragonite  
114 nucleation (Sunagawa et al., 2007).

115         Although many studies have demonstrated that the biological control of shell  
116 composition is often more important than the environmental control (Carré et al., 2006;  
117 Freitas et al., 2005, 2006; Gillikin et al., 2005), there is evidence to the opposite trend (Klein  
118 et al., 1996; Pearce and Mann, 2006). In biologically controlled mineralization, the organism  
119 drives the process of nucleation and growth of the minerals independently from the  
120 environmental conditions. Basically, endogenous factors manifest through co-regulation of all  
121 the structures and functions of the organism, including its sex, growth rate, metabolism, and  
122 feeding strategy (Lowenstam and Weiner, 1989). Throughout the lifespan, the biogenic  
123 system experiences ontogenetic trends and seasonal variations in physiology, determining  
124 metabolic expenses on the basis of life's needs. Biological effects have been repeatedly used  
125 to explain any shifts of elemental concentrations in shells from a theoretical equilibrium  
126 (Davis et al., 2000; Watson et al., 1995). Ontogenetic fluctuations of the growth rate and  
127 metabolic activity affect the intensity of the metal uptake (Lee et al., 1998). Vander Putten et  
128 al. (2000) concluded that the seasonality of the accumulation of Mg, Sr and Pb in *Mytilus*  
129 *edulis* shells shows significant similarity across individuals, with a maximum during spring  
130 and early summer, and profiles cannot be explained by seasonal variations in the seawater  
131 composition. (Carré et al., 2006) developed a model of ion transport in bivalve shells that  
132 shows that Ca channels are less ion-selective when Ca<sup>2+</sup> fluxes are higher. Other studies have  
133 found that trace metal uptake rate increases as mussel filtration rate increases (Janssen and  
134 Scholtz, 1979).

135         It is still difficult to separate the degree to which the organism itself controls the



136 composition of the carbonate shell and how much this process is affected by the environment  
137 (Casella et al., 2017; Schöne and Krause, 2016; Weiner et al., 2001). Although calcareous  
138 parts serve as a powerful tool for the interpretation of the environmental conditions, detailed  
139 insight into the different factors controlling the composition of biological CaCO<sub>3</sub> is  
140 fundamental for understanding the chemical profiles. Tracking potential sources of variation  
141 within calcifying organisms of many species with diverse mineralogy should provide valuable  
142 insight into the patterns driving the biomineralization process (Smith et al., 1998).

143 The aim of this study is, by examining the trace element concentrations in calcitic,  
144 aragonitic and bimineralic shells, to assess the potential differences in free ion binding in two  
145 crystalline polymorphs in the low-salinity environmental system using the example of the  
146 southern Baltic Sea. Brackish waters influence the activities of metal ions, enhancing their  
147 bioavailability (Fritioff et al., 2005) and contributing to lower Ca<sup>2+</sup> concentrations, CaCO<sub>3</sub>  
148 saturation states, alkalinity and buffering capacity compared to those of more saline waters  
149 (Beldowski et al., 2010; Cai et al., 2010; Findlay et al., 2007). The seasonal changes (mainly  
150 the surface temperature variability, periodic lowering of salinity, irregular inflow of pollutants  
151 and intensity of biological processes) were found to affect both the metal concentration and  
152 the physiological processes of living organisms (Urey et al., 1951). For this study we selected  
153 dominant species of the calcifying benthic fauna in this region. This includes the bivalve  
154 molluscs: aragonitic *Cerastoderma glaucum*, *Mya arenaria*, and *Limecola balthica* and  
155 bimineralic *Mytilus trossulus*, as well as an arthropod, *Amphibalanus improvisus*, with a fully  
156 calcitic shell. The investigated species are deposit and filter feeders and have sessile or semi-  
157 sessile lifestyles. Their lifespan is long enough to include various seasons and years covering  
158 therefore a range of environmental conditions during their existence. The potential influence  
159 of local environmental conditions on the observed trace element concentrations in the shells  
160 of the studied organisms is discussed. A comparison of concentration levels between size  
161 classes is also performed, to demonstrate any potential biological control of the shell  
162 chemistry (assuming that the larger specimens were older than the smaller specimens of the  
163 same species).

## 164 2. Material and methods

### 165 2.1. Study area

166 The study area is located in the Gulf of Gdansk in the southern region of the Baltic Sea (Fig.  
167 1). The Gulf is partially sheltered from the northwest by the Hel Peninsula and from the west  
168 and south by the Polish coastline (Kruk-Dowgiałło and Szaniawska, 2008; Rainbow et al.,



169 2004). This location makes the seawater the most turbulent in January and the calmest in  
170 June, with weak bottom currents and minimal tidal amplitudes. The hydrophysical parameters  
171 of the Gulf are mostly driven by the temperate climate and the following seasonal changes.  
172 Differences in air temperature and water mixing cause seasonal fluctuations of the surface  
173 water temperature, ranging from approximately 4 to 22°C (Uścińowicz, 2011). The Gulf of  
174 Gdansk is a low-salinity system under the influence of brackish water from the open southern  
175 Baltic Sea and fresh waters from rivers, mainly the Vistula River; the Vistula is the largest  
176 river in Poland and has an average annual inflow in the estuary of 1080 m<sup>3</sup> s<sup>-1</sup>, which varies  
177 seasonally from 250 to 8000 m<sup>3</sup> s<sup>-1</sup> and has a maximum in spring (Cyberski et al., 2006).  
178 Thus, the average water salinity in the gulf is 7, varying from approximately 5.5 in summer to  
179 8.4 in winter (Bulnheim and Gosling, 1988; Szefer et al., 2002).

180 The Gulf is known as an area highly exposed to human impact. This is due to intensive  
181 usage of its resources and to anthropogenic emissions originating from various coastal  
182 sources, river inflows and atmospheric deposition. The most significant input of industrial and  
183 municipal pollution into the Gulf of Gdansk is derived from the Vistula River, which  
184 transports pollutants from a catchment area of 194,000 km<sup>2</sup> (Pruszek et al., 2005). Both the  
185 water discharge and sediment load into the Gulf are strongly seasonally dependent. The  
186 maximum river flow typically occurs in spring, with the minimum in autumn and winter,  
187 mainly due to periods of snow and ice melting (Cyberski et al., 2006; Pruszek et al., 2005;  
188 Szefer et al., 1996). Winter storms cause the re-deposition of sediments further into the Gulf  
189 of Gdansk (Damrat et al., 2013). Because of the natural conditions, mainly, the limited water  
190 exchange, contaminants brought by the river remain in the ecosystem for decades,  
191 accumulating in the sediments and in living organisms (Glasby et al. 2004; Szumiło-Pilarska  
192 et al. 2016).

193 The benthic community of the Gulf of Gdansk is characterized by the dominance of  
194 calcifying filter feeders and deposit feeders (Kruk-Dowgiałło and Dubrawski, 1998). The low  
195 salinity (~7) and alkalinity, which, in most regions of the Baltic Sea, are lower than those in  
196 oceanic surface waters, cause reduced CO<sub>3</sub><sup>2-</sup> and Ca<sup>2+</sup> concentrations in seawater, thus  
197 lowering the CaCO<sub>3</sub> saturation state ( $\Omega$ ). Furthermore, due to the seasonality of the  
198 temperature and water chemistry, the amplitude of  $\Omega$  in the Baltic Sea is high in comparison  
199 to that in the open ocean. It alternates between approximately 1 to 5 for calcite and 0.5 to 2.5  
200 for aragonite (Findlay et al., 2007). Seasonal changes also affect the feeding behaviour. The  
201 capacity for food processing is evolutionarily adapted to the concentrations of suspended  
202 matter, primarily phytoplankton, that prevail in the biotope during the productive seasons of



203 the year. During these periods, organisms exploit their growth potential (Pierscieniak et al.,  
204 2010; Staniszewska et al., 2016).

205

## 206 2.2. Species

207 The barnacle *Amphibalanus improvisus* (Darwin, 1854) (Arthropoda, Maxillopoda),  
208 commonly named the bay barnacle, is a small sessile crustacean that typically exists in  
209 shallow coastal zones that are less than 10 m deep. It is widespread around the Atlantic and  
210 has been dispersed by shipping to many parts of the world, now having a worldwide  
211 distribution. It is a euryhaline and eurythermal species that is absent only from the Arctic and  
212 Antarctic seas (Kerckhof, 2002). *A. improvisus* is a filter feeder that inhabits hard substrates.  
213 In the Baltic Sea, the reproduction of barnacles starts in spring with temperatures over 10°C  
214 and ends in autumn. The species grows to approximately 10 mm in diameter with a maximum  
215 height of approximately 6 mm, and generally, it has longevity of one year (Żmudziński,  
216 1990). It has a conical shell composed of six fused calcite plates (Weidema, 2000).

217 The clam *Cerastoderma glaucum* (Bruguère, 1789) (Mollusca, Bivalvia), commonly  
218 known as the lagoon cockle, is a saltwater clam found along the coasts of Europe and North  
219 Africa, including in the Mediterranean and Black Seas, the Caspian Lake, and the low-salinity  
220 Baltic Sea. It is a euryhaline species living in salinities of 4–84. *C. glaucum* can tolerate  
221 habitats with wide range of temperatures, from periodically freezing to above 30°C. It is a  
222 filter feeder that burrows 2–4 cm below the surface in soft sediments at shallow depths of  
223 secluded bays. It actively lives near the sediment surface, acting as a biodiffuser (Urban-  
224 Malinga et al., 2013). The clam is surrounded by a ribbed aragonite shell, which is externally  
225 yellowish to greenish brown (Jelnes et al., 1971). In the brackish environment of the Gulf of  
226 Gdansk, *C. glaucum* spawns in May-July and typically lives up to 4 years, achieving a height  
227 of 27 mm (Żmudziński, 1990).

228 The soft-shell clam *Mya arenaria* (Linnaeus, 1758) (Mollusca, Bivalvia) is a marine  
229 invasive species introduced into European waters from the Atlantic coasts of North America  
230 (Behrends et al., 2005). It has a wide global distribution, mainly due to its adaptability to  
231 varying environments with salinities between 4 and 35 and temperatures between –2 and  
232 28°C (Gofas, 2004; Strasser et al., 1999). *M. arenaria* burrows into the sediment up to 20–30  
233 cm below the surface of the sea floor. *M. arenaria* is a filter feeder, filtering organic particles  
234 and microinvertebrates using long fused siphons, and a deposit feeder. In the Gulf of Gdansk,  
235 *M. arenaria* is a common inhabitant of shallow waters up to a depth of 30 m. It spawns once





236 or twice a year in spring or summer, at temperatures of 10 – 15°C. Individuals live 10 – 12  
237 years. They have aragonitic shells and grow up to 70 mm (Żmudziński, 1990).

238 The clam *Limecola balthica* (Linnaeus, 1758) (Mollusca, Bivalvia) lives in the  
239 northern parts of the Atlantic and Pacific oceans, in Subarctic and European waters from  
240 southern France to the White Sea and Pechora Sea, including the Baltic Sea (Strelkov et al.,  
241 2007). It is a euryhaline clam capable of living in a wide range of water salinities from 3 to 40  
242 and at temperatures from –2 to more than 30°C (Sartori and Gofas, 2016). *L. balthica* is a  
243 filter feeder and deposit feeder that lives buried a few centimetres below the surface. It has a  
244 semi-sessile lifestyle, with the ability to undertake periodic migrations (Hiddink et al., 2002).  
245 In the Baltic Sea, *L. balthica* lives at depths of up to 40 m and grows to 24 mm. Adults  
246 reproduce in spring when the water temperature reaches 10°C and live 12 years (Żmudziński,  
247 1990). They have aragonitic shells varying in colour between individuals and locations,  
248 mainly exhibiting white, pink, yellow and orange (Sartori and Gofas, 2016).

249 The mussel *Mytilus trossulus* (Gould, 1850) (Mollusca, Bivalvia) is one of three  
250 closely related taxa in the *Mytilus edulis* complex of blue mussels, which, collectively, are  
251 widely distributed in the temperate and cold-water coasts of the Northern Hemisphere and are  
252 often dominant organisms on hard substrates of shallow nearshore habitats (Rainbow et al.,  
253 1999; Wenne et al., 2016). Generally, *M. trossulus* has a life span of approximately 12 years  
254 and grows to 100 mm, yet in the estuarine environment of the Gulf of Gdansk, it reaches a  
255 maximum length of approximately 50 mm (Gofas, 2004; Żmudziński, 1990). Mussels are  
256 sessile filter feeders, mainly depending on phytoplankton. They reproduce from late spring to  
257 early autumn, depending on the temperature and food abundance (Larsson et al., 2017;  
258 Lauringson et al., 2014; Rainbow et al., 2004). The shell of *M. trossulus* is bimineralic and  
259 consists of two calcium carbonate layers: an outer calcite and inner aragonite layer in variable  
260 proportions between individuals (Dalbeck, 2008; Piwoni-Piórewicz et al., 2017).

261

### 262 2.3. Sample collection and preparation

263 Samples of shells were gathered by a Van Veen grab sampler from four stations, GN, MA,  
264 M2 and MW, located in the Gulf of Gdansk (Table 1, Fig. 1). No single location that would  
265 be a source of several species presenting different forms of CaCO<sub>3</sub> was found. Three species  
266 were found at MA, M2 and MW (one at each station), while two species were collected at  
267 station GN (see Table 1). To ensure that the samples were not contaminated or modified by  
268 solutions of preservatives, the collected material was transported alive in tanks filled with  
269 seawater to the laboratory, where sample preparation took place.





270 By measuring the shell heights of clams and shell diameters of barnacles using a  
271 calliper with an accuracy of  $\pm 1$  mm, the shells were classified into four size classes. The  
272 division into size classes was performed on the basis of the size reached by each species in the  
273 southern Baltic Sea environment (Table 2). Forty shells (ten in each class) were selected for  
274 *A. improvisus*, *C. glaucum*, *M. arenaria* and *L. balthica*. For *M. trossulus*, 20 shells were  
275 selected, with five in each class, while the results for the rest were obtained from Piwoni-  
276 Piorewicz et al. (2017). After the removal of soft tissues, each shell was viewed under a  
277 stereoscopic microscope to examine the epibiotic flora and fauna, which could contaminate  
278 the sample and bias the chemical analysis. If any organisms were present on the shell, they  
279 were carefully removed. To remove the biofilm, the periostracum was scraped with a scalpel,  
280 and pre-cleaned shells were placed in an ultrasonic bath (InterSonic IS-7S) in ultra-pure water  
281 for 30 minutes and then dried at 70°C for 24 hours. The shells were then crushed and ground  
282 into a fine powder with an agate mortar and pestle. Aliquots (2.8 – 849 mg, mean = 132 mg)  
283 of the powdered samples were weighed using a 5-digit analytical balance, placed into a 15 ml  
284 plastic tube (Sarstedt®) and dissolved in a mixture of 1.5 ml concentrated nitric acid (HNO<sub>3</sub>,  
285 Sigma Aldrich®, Trace SELECT for trace analysis), 1.5 ml ultra-pure water and 0.3 ml 30%  
286 hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>, Merck® Suprapure grade). After 24 hours at 70°C, the liquid  
287 samples were diluted to c. 15 ml by weight with ultra-pure water.

288

#### 289 2.4. Elemental analysis

290 Concentrations of chemical elements in the digested samples were determined at the Natural  
291 History Museum, London, using a Thermo iCap 6500 Duo inductively coupled plasma optical  
292 emission spectrometer (ICP-OES) for Ca, Mg, Na and Sr and an Agilent 7700x inductively  
293 coupled plasma mass spectrometer (ICP-MS) for Ba, Cd, Cu, Pb, U, V, Mn and Y.  
294 Calibration of the ICP-OES analysis was performed using solutions that were matrix-matched  
295 to the high calcium concentrations in the samples. Multiple wavelengths for each element  
296 were recorded, and line selection was performed by accounting for the suitability of the  
297 wavelength to the concentrations in the samples. The accuracy and reproducibility of the  
298 analyses were checked using two calcium-rich certified reference materials (CRMs), JLs-1  
299 Limestone and JDo-1 Dolomite (both from the Geological Survey of Japan), which were  
300 dissolved by the total digestion method using hydrofluoric acid. The reference materials were  
301 diluted to match the concentrations of Ca in the sample solutions (Imai et al., 1996).

302 The limits of quantification (LOQ) of the ICP-MS analysis were generally determined  
303 as the concentration corresponding to ten times the standard deviation of the signal obtained



304 by analysing 5% HNO<sub>3</sub> solution (6–7 times) in each individual run. ICP-MS was run in  
305 helium (He) mode (5 ml min<sup>-1</sup> He, 99.9995% purity) for most trace elements to minimize the  
306 molecular interference from plasma and solution components and Ca from the samples. The  
307 accuracy and reproducibility was checked by periodic analyses of JLS-1 and JDo-1 (Imai et  
308 al., 1996).

309

### 310 2.5. Statistical analyses

311 To evaluate the effect of the shell size (ontogenesis stage) on the concentrations of trace  
312 elements in calcareous parts of *A. improvisus*, *C. glaucum*, *M. arenaria*, *L. balthica* and *M.*  
313 *trossulus*, the concentrations of Ca, Na, Sr, Mg, Ba, Mn, Cu, Pb, V, Y, U and Cd were  
314 examined in the four size classes separately for each species. The data were not normally  
315 distributed (Shapiro-Wilk test); therefore, significant differences between the mean  
316 concentrations of the selected elements in the size classes were identified by one-way  
317 Kruskal-Wallis nonparametric ANOVA ( $p$ -value = 0.05) and post-hoc Dunn's tests for  
318 multiple independent groups. Statistical computing and graphical visualizations were  
319 performed in RStudio.

320

### 321 3. Results

322 The species exhibited similar accumulation orders since they had the highest concentration of  
323 Na, Sr and Mg and the lowest concentrations of U and Cd (Table 3, Fig. 2). However, the  
324 concentrations of metals were different between the calcitic shells of *A. improvisus*, the  
325 aragonitic shells of *C. glaucum*, *M. arenaria*, *L. balthica* and the bimineralic shells of *M.*  
326 *trossulus*, showing high variability (Fig. 3). Manganese was the most variable and increased  
327 from  $1.4 \pm 1.0$  mg kg<sup>-1</sup> in aragonitic *M. arenaria* to  $620 \pm 155$  mg kg<sup>-1</sup> in calcitic *A.*  
328 *improvisus*. Trace elements (V, Cu, Y, Cd, Ba, Pb, U) were generally present at higher  
329 concentrations in calcitic shells than in aragonitic ones. Shells of *M. arenaria* had the lowest  
330 concentrations of impurities. The smallest variability between species was found for Sr, Na  
331 and Ca (Fig. 3).

332 The results of the Kruskal-Wallis nonparametric ANOVA test, which was used to  
333 compare the shell chemistry between the four size classes in each population, revealed the  
334 lack of ontogenetic changes within aragonitic *M. arenaria*. The smallest variability was found  
335 in aragonitic *L. balthica*, for which only the concentration of Na ( $p = 0.014$ ) decreased with  
336 shell growth. The third of the aragonite species, *C. glaucum*, showed high variability of the  
337 four elements. Throughout the lifespan, shells were built with increasing concentrations of Sr



338 ( $p = 0.002$ ), which was contrary to the decreasing concentration of Na ( $p = 0.015$ ), Mn ( $p =$   
339  $0.014$ ) and Cd ( $p = 0.009$ ). Similarly, the calcitic shells of *A. improvisus* also showed  
340 variability of four elements between size classes, namely, Mg ( $p = 0.007$ ), V ( $p = 0.011$ ), Cu  
341 ( $p = 0.027$ ) and Pb ( $p = 0.004$ ). However, in this case, the sequences of the ontogenetic  
342 changes were not straightforward but, rather, had a tendency to fluctuate between the  
343 youngest and the oldest individuals. The highest variability was found within the bimineralic  
344 shells of *M. trossulus*. The size classes differed in terms of five elements. The incorporation  
345 of V ( $p < 0.001$ ), Cu ( $p < 0.001$ ), Y ( $p = 0.013$ ), Cd ( $p = 0.002$ ) and U ( $p < 0.001$ ) into the  
346 crystal lattice decreased in larger mussels (Fig. 3, Fig. 4).

347 Detailed analyses of the significant differences in the studied elements between the  
348 size classes based on post-hoc Dunn's tests for multiple independent groups indicated that the  
349 variations were not linear (Fig. 5). In *L. balthica* aragonitic shells, sodium concentration  
350 decreased with ageing, showing differences between the size classes I and III and I and IV. In  
351 the shells of *C. glaucum*, Sr concentration increased gradually, reaching a peak in the size  
352 class IV, while statistically significant differences were observed between the size classes I  
353 and IV and the III and IV. An inverse pattern was observed for sodium, yet its concentration  
354 differed only between the youngest and oldest individuals. The aragonitic shells were  
355 characterized by a common trend of the Mn and Cd concentrations, which decreased from I  
356 size class and later reached a plateau. In the calcitic shells of *A. improvisus*, the concentrations  
357 of Mg, V, Cu and Pb decreased in larger individuals. The levels of magnesium in shells  
358 statically differed between the size classes III and IV; likewise, shells from the size class III  
359 had the highest concentrations. Vanadium and lead occurred at the highest concentrations in  
360 the shells of the youngest individuals, and later, they oscillated around a similar average. Co  
361 decreased in growing shells of barnacles, reaching the minimum in the size class III. There  
362 were no statistically significant differences between the classes III and IV, and the oldest half  
363 of the group maintained the downward trend. However, it is worth noting that Cu  
364 concentrations in many of the oldest shells are several times higher. The bimineralic shells of  
365 *M. trossulus* were characterized by the highest variability of trace element concentrations  
366 between size classes. The trend of decreasing concentrations was clearly marked for V, Co,  
367 Cd and U. The concentration of Y decreased in larger shells, but the oldest shells showed an  
368 upward trend; therefore, significant differences were found between the size classes I and III  
369 (Fig. 5).

370

371 4. Discussion



372 Magnesium and strontium are some of the most commonly studied elements in the skeletons  
373 of marine calcifiers due to their important role in the biological precipitation of  $\text{CaCO}_3$   
374 (Allison et al., 2001; De Choudens Sanchez and Gonzalez, 2009; Morse et al., 2007). We  
375 found statistically significant differences in the concentrations of Mg and Sr between the  
376 calcitic, aragonitic and bimineralic species. This confirms that the crystal structure of the shell  
377 plays a key role in the incorporation of major elements. In the calcitic *A. improvisus*, Mg was  
378 the dominant impurity, while aragonitic *C. glaucum*, *M. arenaria* and *L. balthica* had  
379 concentrations of Sr over 15 times higher than those of Mg. The shell of *M. trossulus* contains  
380 layers of calcite and aragonite, thus favouring both Mg and Sr. Not surprisingly, in this study,  
381 mussels were characterized by similar concentrations of Mg and Sr, reaching  $0.106 \pm 0.017$   
382  $\text{mg kg}^{-1}$  and  $0.117 \pm 0.010 \text{ mg kg}^{-1}$ , respectively (Table 3, Fig. 3). These observations are in  
383 agreement with results obtained by other authors for different calcareous species (Iglukowska  
384 et al., 2016; Morse et al., 1997; Reeder, 1983; Wang and Xu, 2001). Such results confirm the  
385 importance of the Mg and Sr ionic radius in terms of the aragonite and calcite lattice  
386 properties (England, 2005). The studied calcitic *A. improvisus*, without a doubt, has the  
387 highest Mg concentration in its shells ( $0.394 \pm 0.032 \text{ mg kg}^{-1}$ ); however, the Sr concentration  
388 was also higher in barnacles ( $0.225 \pm 0.019 \text{ mg kg}^{-1}$ ) compared to that in aragonitic clams  
389 ( $0.170 - 0.216 \text{ mg kg}^{-1}$ ) (Table 3, Fig. 3). Kinetic and biological effects (Urey et al., 1951)  
390 influence the partitioning of Sr between the shell and seawater, and Sr in shells is known to  
391 significantly exceed its concentration expected at the thermodynamic equilibrium (Schöne et  
392 al., 2010). An important consideration for Sr concentrations in calcite is that Mg incorporation  
393 into the shells may increase the Sr affinity through changes in the calcite crystal morphology  
394 (Carpenter and Arbor, 1992). Greater inclusion of minor ions into the shell structure may  
395 result in crystal defects, contributing to a higher incorporation rate of less compatible ions into  
396 the crystal lattice (Dalbeck, 2008; Davis et al., 2000). Furthermore, it was found that  
397 barnacles differ from other calcitic species, having an uncommonly high Sr level (Carpenter  
398 and Arbor, 1992; Ullmann et al., 2018), and the highest concentration in *A. improvisus* in the  
399 Gulf of Gdansk could act as species-specific adaptation (Fig. 3).

400 Sodium belongs to the group of proxies used in the robust reconstruction of salinity,  
401 and this relationship seems to be related mainly to the increase of Na activity in seawater with  
402 increasing salinity (Findlater et al., 2014; Wit et al., 2013). Although we observe that for  
403 invertebrates collected from stations with similar salinities between 6.9 and 7.3, the aragonitic  
404 species incorporated more Na than did the calcitic and bimineralic species (Figs. 3 and 4). The  
405 concentration of Na ranged, on average, from 0.2 wt% in *M. trossulus* to 0.32 wt% in *C.*



406 *glaucum* (Table 3). This finding could mean that the polymorph orientation affects Na  
407 concentration in shells. Following the orientation of the crystal lattice, Na, as a cation larger  
408 than Ca, performs better in the aragonite structure (Davis et al., 2000). Dalbeck (2008)  
409 compared the minor element variations between the polymorphs of *Mytilus edulis* bimineralic  
410 shells, and found that Na occurred in higher concentrations in the aragonitic layer. Although  
411 the Na/Ca ratio is frequently discussed as a climate proxy, the degree to which Na  
412 concentration in biogenic carbonate represents the environmental salinity varies from one  
413 species to another (Vander Putten et al., 2000; Schöne and Krause Jr, 2016). This study shows  
414 lack of pattern in the distribution of Na between individuals from one population and between  
415 species.

416 The concentrations of manganese and barium in the calcitic shells of *A. improvisus*  
417 were several orders of magnitude larger than those in other species, reaching  $625 \pm 160 \text{ mg}$   
418  $\text{kg}^{-1}$  and  $73.2 \pm 19.3 \text{ mg kg}^{-1}$ , respectively (Table 3). The bimineralic shells of the mussel *M.*  
419 *trossulus* collected from the same location as *A. improvisus* (GN) contained  $56.6 \pm 20.0 \text{ mg}$   
420  $\text{kg}^{-1}$  of Mn and  $17.0 \pm 6.7 \text{ mg kg}^{-1}$  of Ba, while the aragonitic shells of clams incorporated  
421 from  $1.419 \pm 1.01 \text{ mg kg}^{-1}$  to  $11.933 \pm 10.47 \text{ mg kg}^{-1}$  of Mn and from  $2.63 \pm 1.58 \text{ mg kg}^{-1}$  to  
422  $15.12 \pm 10.15 \text{ mg kg}^{-1}$  of Ba (Table 4, Fig. 3). This trend does not seem to be strongly linked  
423 to the crystal lattice orientation because smaller Mn cations are energetically favoured in  
424 calcitic species, while larger cations, such as Ba, better fit into aragonite (Dalbeck, 2008;  
425 Davis et al., 2000; Wang and Xu, 2001). It was expected that aragonite shells would  
426 incorporate Ba more intensively than would calcitic shells (Findlater et al., 2014; Gillikin et  
427 al., 2006), yet this trend was not observed (Figs. 3 and 4). The organisms derive metal ions  
428 from both the surrounding water and food (Freitas et al. 2006; Gillikin et al. 2005; Poulain et  
429 al. 2015). It is highly probable that intra-species variabilities reflect the seasonal variability of  
430 Mn and Ba concentrations in seawater, while differences between species partly depend on  
431 feeding strategies. Findlater et al. (2014), based on shells of fossil freshwater bivalves and  
432 gastropods from the same location, concluded that a clear biochemical vital effect was evident  
433 for Mn and Ba. Dissolved Mn is intensively delivered to the oceans by riverine input, and in  
434 the Gulf of Gdansk, the geographical range of the Vistula River inflow plays a very important  
435 role in the Mn distribution (Uścińowicz, 2011). However, much lower concentrations of Mn  
436 were found in the shells of molluscs than in barnacles from the same location (Figs. 3 and 4).  
437 A similar relationship was found in the soft tissues of *M. trossulus* and *A. improvisus*  
438 collected from different locations in the Gulf of Gdansk in May 1998 (Rainbow et al., 2000).  
439 The range of Mn in the soft tissues of *M. trossulus* varied from 19.0 to 41.0, while that in *A.*



440 *improvisus* ranged from 187 to 307 mg kg<sup>-1</sup>. Therefore, *A. improvisus* is a very effective  
441 accumulator of Mn. The disproportionality of Mn and Ba concentrations between molluscs  
442 and barnacles could reflect the ability of molluscs to limit vital processes and to reduce the  
443 calcification rate during stressful conditions (Berge et al., 2006; Hiebenthal et al., 2012). A  
444 limited food base, low temperatures, saturation states during colder months and even the high  
445 river inflow in early spring (Cyberski et al., 2006; Pruszek et al., 2005; Szefer et al., 1996) can  
446 reduce the precipitation rates of *C. glaucum*, *M. arenaria*, *L. balthica* and *M. trossulus*. In  
447 contrast, the calcified shells of barnacles grow more or less continuously (Bourget and Crisp,  
448 1975) and represent yearly calcification. The Mn and Ba incorporation from seawater into  
449 calcite and aragonite has been shown to correlate with changes in primary production and  
450 phytoplankton blooms (Freitas et al., 2006; Lazareth et al., 2003; Vander Putten et al., 2000).  
451 As many clams and barnacles are filter-feeding organisms, Mn- or Ba-rich particles are  
452 ingested with food and from the water column. Schoemann et al. (1998) recorded a seasonal  
453 increase of dissolved Mn in the water column after peaks in phytoplankton blooms. This  
454 partly explains why *M. arenaria* had the smallest concentrations of Mn and Ba. This clam  
455 represents a deposit-feeding mode and thus may have limited access to suspended particles.  
456 Uptake and accumulation in filter-feeders correlates with metal concentrations in seawater,  
457 whereas accumulation in deposit-feeders most likely reflects the metal concentrations in the  
458 sediments (Newman and McIntosh, 1982). Luoma and Jenne (1976) concluded that the degree  
459 of metal accumulation by deposit-feeders was directly related to the characteristic rate of  
460 metal desorption from a particular type of sediment.

461 Shells of *A. improvisus*, characterized by the highest concentrations of Mg, Sr, Mn and  
462 Ba, were simultaneously the richest in most other trace metals, which were likewise more  
463 concentrated in bimineralic *M. trossulus* than in aragonitic clams (Table 3, Fig. 2). It was  
464 mentioned previously that impurities arise in deformed crystal lattices (Dalbeck, 2008; Davis  
465 et al., 2000), and these lattices could be changed through the intensive incorporation of Ca  
466 substitutions. However, it is also important to mention that the ionic radii of smaller cations,  
467 such as V, Cd, Y, Pb, U and Cu, are more energetically favoured in calcite than in aragonite  
468 (Morse et al., 1997; Reeder, 1983; Wang and Xu, 2001).

469 Trace metal concentrations in shells are usually linked to local environmental  
470 conditions. If heavy metals in the shells of calcifying invertebrates reflect the environmental  
471 levels of bioavailable ions, the highest concentrations of most trace metals in the shells of *A.*  
472 *improvisus* and *M. trossulus* from GN station (Table 4, Fig. 3) are probably related to local  
473 contamination. Rainbow et al. (2004) compared the differences between the bioavailabilities



474 of seven metals (Cu, Zn, Fe, Pb, Ni, Cd and Mn) to *A. improvisus* and *M. trossulus* across five  
475 localities in the Gulf of Gdansk, including GN station, which was often the richest metal  
476 source for marine organisms. Among the stations selected in this study, GN is closer than  
477 MA, MW and M2 to the mouth of the Vistula River, which is a major source of metal  
478 contaminants in the marine system (Fig. 1).

479 In this study, individuals were collected over a wide range of sizes from each station  
480 (Table 2), representing different ages and various periods of time, living under the influence  
481 of seasonal changes. The southern Baltic Sea is driven by cyclical environmental dynamics,  
482 which evoke physiological stress, reduce the food base and determine the biogeochemical  
483 cycle (Elder and Collins, 1991). It is certain that the biological processes of an animal  
484 precipitating carbonate complicate the calcification course (Vander Putten et al., 2000;  
485 Zacherl et al., 2003). However, the physiology of an organism is likely dependent on the  
486 environmental conditions it experiences (Gillikin et al., 2005). The recorded concentrations of  
487 metals in all populations exhibited marked variability among individuals within a single  
488 location (Table 3, Fig. 3), an attribute which was previously noticed by several researchers  
489 (Gillikin et al., 2005; Vander Putten et al., 2000). The wide range of inter-species variability  
490 suggests that the uptake of metals, and the subsequent bioavailability, are highly dependent on  
491 biological factors such as the growth rate, feeding strategy, genotype and phylogeny. Within  
492 species, organisms from juveniles to adults experience morphological and functional changes  
493 related to sex, metabolic rate or reproductive stage, which affect the biomineralization process  
494 (Carré et al., 2006; Freitas et al., 2006; Gillikin et al., 2005; Schöne et al., 2010, 2011; Warter  
495 et al., 2018). In shells of the studied *A. improvisus*, *C. glaucum*, *M. arenaria*, *L. balthica* and  
496 *M. trossulus*, differences in elemental concentrations over the lifespan did not produce a  
497 consistent pattern. From the chemical profiles, it is evident that the size-dependent  
498 distributions of elements in the shells were only marked in a few cases (Figs. 5 and 6). The  
499 studied *A. improvisus* and *M. trossulus* showed the greatest chemical variability between size  
500 classes (Fig. 4). However, in this case, the varied elements were not uniform for *A. improvisus*  
501 and *M. trossulus*, even though the organisms came from one location (Table 1). No size-  
502 related trend was observed for Y, U, and Cd in barnacles or for Pb in molluscs (Fig. 3). It was  
503 previously found that different species in the same habitat might show different patterns of  
504 metal accumulation, which are caused by specific routes of uptake or differences in the  
505 physiological pathway of metals (Rainbow, 2002, 1995). Rainbow et al. (1999) tested *A.*  
506 *improvisus* and *M. trossulus* from the Gulf of Gdansk as environmental biomonitors by  
507 measuring the concentrations of Co, Zn, Fe, Cd, Pb, Mn and Ni in soft tissues. They revealed





508 that the mussels and barnacles gathered at one location did not show the same variation in  
509 metal bioavailabilities.

510 In this study, generally, when statistical differences between size classes were  
511 recorded, the concentrations of metals decreased with the size of the shell. The reverse trend  
512 was found only for Sr in the shells of *C. glaucum*, in which the concentration of Sr increased  
513 with size (Fig. 4). It is possible that the younger specimens had a greater calcium carbonate  
514 precipitation rate and uptake of trace elements due to a higher metabolism in the earlier stages  
515 of the animal life (Dalbeck, 2008). Rosenberg and Hughes (1991) suggested that areas of  
516 higher shell curvature, such as the umbo, require greater metabolic expenditure, resulting in  
517 an increase of metals uptake. This supports the idea of a higher metabolic rate and metal  
518 concentrations in younger specimens (Szefer et al., 2002). This study also indicates that the  
519 smallest individuals seem to be the most efficient accumulators of trace elements. This is  
520 mostly expressed in the cases of a few trace elements, in which the concentration decreased  
521 across the four size classes in the shells of *A. improvisus*, *M. trussulus* and *C. glaucum* (Fig.  
522 4). There are several studies confirming a negative relationship between metal concentrations  
523 and shell size (Martincic et al., 1992; Piwoni-Piórewicz et al., 2017; Ritz et al., 1982). Catsiki  
524 and Gialamas (1994) suggested that apart from metabolic processes, an active detoxification  
525 mechanism is responsible for this trend, and its efficiency is higher in older and larger  
526 individuals.

527 Nevertheless, many of the elements studied herein showed a lack of statistically  
528 significant relationships between the shell sizes. Few trace metals in the shells of *A.*  
529 *improvisus* (Y, U, Cd), *C. glaucum* (V, Cu, Y, Pb, U) and *M. trussulus* (Pb) and all trace  
530 metals in the shells of *L. balthica* and *M. arenaria* (V, Cu, Y, Pb, U, Cd) showed no  
531 significant variability related with the growth of organisms (Fig. 3). This is not unusual in  
532 biomineralization studies. Saavedra et al. (2004) observed no differences between Cd, Pb, Cr,  
533 Ni, As, Cu and Zn concentrations for different shell lengths of the raft *Mytilus*  
534 *galloprovincialis* separated into four size classes. Protasowicki, Dural, & Jaremek (2008)  
535 similarly found that the concentrations of Hg, Pb, Cd, Cu, Zn, Cr, Ni, Fe, Mn, V, Li and Al in  
536 the shells of the mussel *Mytilus edulis* from the Polish coast of the Baltic Sea did not vary  
537 between shell sizes. This inconsistent pattern proves that biological effects modify the way in  
538 which environmental data are recorded in the biogenic hard parts.

539

540 5. Summary



541 The shells of the calcitic *A. improvisus*, the aragonitic *C. glaucum*, *L. balthica*, *M. arenaria*  
542 and the bimineralic *M. trossulus* from the Gulf of Gdansk are accumulators of a wide  
543 spectrum of metals from the surrounding environment. The selected taxa differed in terms of  
544 the chemical compositions of shells, both between taxa and within a single population.  
545 Elemental variability is strongly associated with the properties of the crystal lattice, which are  
546 different between calcite and aragonite. Remarkably high variability of elemental  
547 concentrations between species, especially for manganese, supports the importance of  
548 species-specific biological control of the biomineralization process. The high variability of  
549 elemental values within a single population indicates that both physiological and  
550 environmental factors have an influence on the incorporation of these trace elements into the  
551 shell. Given that the species were obtained from various sites, an impact from local  
552 environmental factors cannot be excluded.

553 This study has shown that interpreting the chemical profiles of calcium carbonate  
554 shells to determine the marine environment conditions is recommended to investigate how  
555 metal concentrations vary among large numbers of species from contrasting habitats.  
556 Nevertheless, we recommend using shells of similar sizes to reduce the ontogenetic  
557 variability.

558

#### 559 Author contribution

560 Anna Piwoni-Piórewicz and Piotr Kukliński designed the study and carried them out.  
561 Stanislav Strekopytov and Emma Humphreys-Williams conducted elemental analysis. Anna  
562 Piwoni-Piórewicz prepared the manuscript with contributions from all co-authors.

563

#### 564 Competing interests

565 The authors declare that they have no conflict of interest.

566

#### 567 Data availability

568 The underlying research data can be accessed at Institute of Oceanology, Polish Academy of  
569 Sciences, Powstańców Warszawy 55, 81-712 Sopot, Poland.

570

#### 571 Acknowledgements

572 We thank Dr. Halina Kendzierska from the University of Gdansk for cooperation during  
573 sampling. The research leading to these results received funding from the Polish National



- 574 Science Centre in the frame of project contracts PANIC/2016/23/B/ST10/01936 and  
575 LOGGER/2017/25/N/ST10/02305.  
576
- 577 References
- 578 Allison, N., Finch, A. A., Sutton, S. R. and Newville, M.: Strontium heterogeneity and  
579 speciation in coral aragonite: implications for the strontium paleothermometer, *Geochim.*  
580 *Cosmochim. Acta*, 65(16), 2669–2676, doi:10.1016/S0016-7037(01)00628-7, 2001.
- 581 Balthasar, U. and Cusack, M.: Aragonite-calcite seas—Quantifying the gray area, *Geology*,  
582 43(2), 99–102, doi:10.1130/G36293.1, 2015.
- 583 Behrends, B., Hertweck, G., Liebezeit, G. and Goodfriend, G.: Earliest Holocene occurrence  
584 of the soft-shell clam, *Mya arenaria*, in the Greifswalder Bodden, Southern Baltic, *Mar. Geol.*,  
585 216(1–2), 79–82, doi:10.1016/j.margeo.2005.01.002, 2005.
- 586 Beldowski, J., Löffler, A., Schneider, B. and Joensuu, L.: Distribution and biogeochemical  
587 control of total CO<sub>2</sub> and total alkalinity in the Baltic Sea, *J. Mar. Syst.*, 81(3), 252–259,  
588 doi:https://doi.org/10.1016/j.jmarsys.2009.12.020, 2010.
- 589 Bentov, S. and Erez, J.: Novel observations on biomineralization processes in foraminifera  
590 and implications for Mg/Ca ratio in the shells, *Geology*, 33(11), 841, doi:10.1130/G21800.1,  
591 2005.
- 592 Berge, J. A., Bjerkeng, B., Pettersen, O., Schaanning, M. T. and Øxnevad, S.: Effects of  
593 increased sea water concentrations of CO<sub>2</sub> on growth of the bivalve *Mytilus edulis* L.,  
594 *Chemosphere*, 62(4), 681–687, doi:10.1016/j.chemosphere.2005.04.111, 2006.
- 595 de Boer, R. B.: Stability of Mg-Ca carbonates, *Geochim. Cosmochim. Acta*, 41(2), 265–270,  
596 doi:10.1016/0016-7037(77)90234-4, 1977.
- 597 Bourget, E. and Crisp, D. J.: Factors affecting deposition of the shell in *Balanus balanoides*  
598 (L.), *J. Mar. Biol. Assoc. United Kingdom*, 55(1), 231–249,  
599 doi:https://doi.org/10.1017/S0025315400015873, 1975.
- 600 Bulnheim, H. P. and Gosling, E.: Population genetic structure of mussels from the Baltic Sea,  
601 *Helgoländer Meeresuntersuchungen*, 42, 113–129, 1988.
- 602 Cai, W.-J., Hu, X., Huang, W.-J., Jiang, L.-Q., Wang, Y., Peng, T.-H. and Zhang, X.:  
603 Alkalinity distribution in the western North Atlantic Ocean margins, *J. Geophys. Res.*,  
604 115(C8), C08014, doi:10.1029/2009JC005482, 2010.
- 605 Cairns, S. D. and Macintyre, I. A. N. G.: Phylogenetic Implications of Calcium Carbonate  
606 Mineralogy in the Stylasteridae (Cnidaria : Hydrozoa), *Sediment. Geol.*, 7(1), 96–107, 1992.
- 607 Carpenter, J. and Arbor, A.: Sr/Mg ratios of modern marine calcite: Empirical indicators of



- 608 ocean chemistry and precipitation rate, *Geochim. Cosmochim. Acta*, 56, 1837–1849,  
609 doi:[https://doi.org/10.1016/0016-7037\(92\)90314-9](https://doi.org/10.1016/0016-7037(92)90314-9), 1992.
- 610 Carré, M., Bentaleb, I., Bruguier, O., Ordinola, E., Barrett, N. T. and Fontugne, M.:  
611 Calcification rate influence on trace element concentrations in aragonitic bivalve shells:  
612 Evidences and mechanisms, *Geochim. Cosmochim. Acta*, 70(19), 4906–4920,  
613 doi:<https://doi.org/10.1016/j.gca.2006.07.019>, 2006.
- 614 Casella, L. A., Griesshaber, E., Yin, X., Ziegler, A., Mavromatis, V., Müller, D., Ritter, A. C.,  
615 Hippler, D., Harper, E. M., Dietzel, M. and Immenhauser, A.: Experimental diagenesis:  
616 insights into aragonite to calcite transformation of *Arctica islandica* shells by hydrothermal  
617 treatment, *Biogeosciences*, 14(6), 1461–1492, doi:10.5194/bg-14-1461-2017, 2017.
- 618 Catsiki, V. A., Katsilieri, C. and Gialamas, V.: Chromium distribution in benthic species from  
619 a gulf receiving tannery wastes (Gulf of Geras - Lesbos island, Greece), *Sci. Total Environ.*,  
620 9697(93), 173–185, doi:[https://doi.org/10.1016/0048-9697\(94\)90308-5](https://doi.org/10.1016/0048-9697(94)90308-5), 1994.
- 621 De Choudens-Sanchez, V. and Gonzalez, L. A.: Calcite and Aragonite Precipitation Under  
622 Controlled Instantaneous Supersaturation: Elucidating the Role of CaCO<sub>3</sub> Saturation State  
623 and Mg/Ca Ratio on Calcium Carbonate Polymorphism, *J. Sediment. Res.*, 79(6), 363–376,  
624 doi:<https://doi.org/10.2110/jsr.2009.043>, 2009.
- 625 Cohen, A. L. and McConnaughey, T. A.: Geochemical perspectives on coral  
626 mineralization, *Rev. Mineral. geochemistry*, 354, 151–187,  
627 doi:<https://doi.org/10.2113/0540151>, 2003.
- 628 Cubadda, F., Superiore, I., Conti, M. E., Cubadda, F., Enrique, M. and Campanella, L.: Size-  
629 dependent concentrations of trace metals in four Mediterranean gastropods, *Chemosphere*,  
630 45(4–5), 561–569, doi:[https://doi.org/10.1016/S0045-6535\(01\)00013-3](https://doi.org/10.1016/S0045-6535(01)00013-3), 2001.
- 631 Cusack, M. and Freer, A.: Biomineralization: elemental and organic influence in carbonate  
632 systems, *Chem. Rev.*, 108(11), 4433–54, doi:<https://doi.org/10.1021/cr078270o>, 2008.
- 633 Cyberski, J., Grześ, M., Gurty-Korycka, M., Nachlik, E. and Kundziewicz, W. W.: History of  
634 floods on the river Vistula, *Hydrol. Sci. J.*, 51(5), 799–817,  
635 doi:<https://doi.org/10.1623/hysj.51.5.799>, 2006.
- 636 Dalbeck, P.: Crystallography, stable isotope and trace element analysis of *Mytilus edulis*  
637 shells in the context of ontogeny, University of Glasgow. [online] Available from:  
638 <http://theses.gla.ac.uk/>, 2008.
- 639 Damrat, M., Zaborska, A. and Zajączkowski, M.: Sedimentation from suspension and  
640 sediment accumulation rate in the Vistula River prodelta, Gulf of Gdańsk (Baltic Sea),  
641 *Oceanologia*, 55(4), 937–950, doi:<https://doi.org/10.5697/oc.55-4.937>, 2013.



- 642 Davis, K. J., Dove, P. M. and De Yoreo, J. J.: The role of Mg<sup>2+</sup> as an impurity in calcite  
643 growth, *Science* (80-. ), 290(5494), 1134–1138, doi:10.1126/science.290.5494.1134, 2000.
- 644 Dickson, J. A. D.: Echinoderm skeletal preservation; calcite-aragonite seas and the Mg/Ca  
645 ratio of Phanerozoic oceans, *J. Sediment. Res.*, 74(3), 355–365,  
646 doi:https://doi.org/10.1306/112203740355, 2004.
- 647 Dove, P. M.: The rise of skeletal biominerals, *Elements*, 6(1), 37–42,  
648 doi:https://doi.org/10.2113/gselements.6.1.37, 2010.
- 649 Elder, J. F. and Collins, J.: Freshwater Molluscs as Indicators of Bioavailability and Toxicity  
650 of Metals in Surface-Water Systems, *Rev. Environ. Contam. Toxicol.*, 122, 37–69, 1991.
- 651 Findlater, G., Shelton, A., Rolin, T. and Andrews, J.: Sodium and strontium in mollusc shells:  
652 Preservation, palaeosalinity and palaeotemperature of the Middle Pleistocene of eastern  
653 England, *Proc. Geol. Assoc.*, 125(1), 14–19, doi:https://doi.org/10.1016/j.pgeola.2013.10.005,  
654 2014.
- 655 Findlay, H., Tyrrell, T., Bellerby, R., Merico, A. and Skjelvan, I.: Ecosystem modelling of the  
656 Norwegian Sea: investigating carbon and nutrients dynamics as a consequence of biological  
657 and physical processes, *Biogeosciences Discuss.*, 4, 3229–3265, 2007.
- 658 Freitas, P., Clarke, L. J., Kennedy, H., Richardson, C. and Abrantes, F.: Mg/Ca, Sr/Ca, and  
659 stable-isotope ( $\delta^{18}\text{O}$  and  $\delta^{13}\text{C}$ ) ratio profiles from the fan mussel *Pinna nobilis*: Seasonal  
660 records and temperature relationships, *Geochemistry, Geophys. Geosystems*, 6(Q04D14),  
661 1525–2027, doi:10.1029/2004GC000872, 2005.
- 662 Freitas, P. S., Clarke, L. J., Kennedy, H., Richardson, C. a. and Abrantes, F.: Environmental  
663 and biological controls on elemental (Mg/Ca, Sr/Ca and Mn/Ca) ratios in shells of the king  
664 scallop *Pecten maximus*, *Geochim. Cosmochim. Acta*, 70(20), 5119–5133,  
665 doi:https://doi.org/10.1016/j.gca.2006.07.029, 2006.
- 666 Fritioff, A., Kautsky, L. and Greger, M.: Influence of temperature and salinity on heavy metal  
667 uptake by submersed plants, *Environ. Pollut.*, 133(2), 265–274,  
668 doi:https://doi.org/10.1016/j.envpol.2004.05.036, 2005.
- 669 Gillikin, D. P., Lorrain, A., Navez, J., Taylor, J. W., Andre, L., Keppens, E., Baeyens, W. and  
670 Dehairs, F.: Strong biological controls on Sr/Ca ratios in aragonitic marine bivalve shells,  
671 *Geochemistry, Geophys. Geosystems*, 6(5), doi:https://doi.org/10.1029/2004GC000874,  
672 2005.
- 673 Gillikin, D. P., Dehairs, F., Lorrain, A., Steenmans, D., Baeyens, W. and André, L.: Barium  
674 uptake into the shells of the common mussel (*Mytilus edulis*) and the potential for estuarine  
675 paleo-chemistry reconstruction, *Geochim. Cosmochim. Acta*, 70(2), 395–407,



- 676 doi:https://doi.org/10.1016/j.gca.2005.09.015, 2006.
- 677 Glasby, G. P., Szefer, P., Geldon, J. and Warzocha, J.: Heavy-metal pollution of sediments  
678 from Szczecin Lagoon and the Gdansk Basin, Poland, *Sci. Total Environ.*, 330, 249–269,  
679 doi:https://doi.org/10.1016/j.scitotenv.2004.04.004, 2004.
- 680 Gofas, S.: *Mytilus trossulus* Gould, 1850, MolluscaBase [online] Available from:  
681 <http://www.marinespecies.org/aphia.php?p=taxdetails&id=140482> (Accessed 6 February  
682 2017), 2004.
- 683 Hiddink, J. G., Marijnissen, S. A., Troost, K. and Wolff, W. J.: Predation on 0-group and  
684 older year classes of the bivalve *Macoma balthica*: interaction of size selection and intertidal  
685 distribution of epibenthic predators, *J. Exp. Mar. Bio. Ecol.*, 269(2), 223–248,  
686 doi:https://doi.org/10.1016/S0022-0981(02)00002-3, 2002.
- 687 Hiebenthal, C., Philipp, E., Eisenhauer, A. and Wahl, M.: Interactive effects of temperature  
688 and salinity on shell formation and general condition in Baltic Sea *Mytilus edulis* and *Arctica*  
689 *islandica*, *Aquat. Biol.*, 14(3), 289–298, doi:10.3354/ab00405, 2012.
- 690 Iglukowska, A., Beldowski, J., Chelchowski, M., Chierici, M., Kedra, M., Przytarska, J.,  
691 Sowa, A. and Kuklinski, P.: Chemical composition of two mineralogically contrasting Arctic  
692 bivalves' shells and their relationships to environmental variables, *Mar. Pollut. Bull.*, 114,  
693 903–916, doi:https://doi.org/10.1016/j.marpolbul.2016.10.071, 2016.
- 694 Imai, N., Terashima, S., Itoh, S. and Ando, A.: Compilation of analytical data on nine GSJ  
695 geochemical reference samples, Sedimentary rock series, *Geostand. Newsl.*, 20(2), 165–216,  
696 doi:https://doi.org/10.1111/j.1751-908X.1996.tb00184.x, 1996.
- 697 Janssen, H. H. and Scholtz, N.: Uptake and cellular distribution of cadmium in *Mytilus edulis*,  
698 *Mar. Biol.*, 55, 133–141, 1979.
- 699 Jelnes, J. E., Petersen, G. H. and Russell, P. C.: Isoenzyme taxonomy applied on four species  
700 of *Cardium* from Danish and British waters with a short description of the distribution of the  
701 species (*Bivalvia*), *Ophelia*, 9(1), 15–19,  
702 doi:https://doi.org/10.1080/00785326.1971.10430087, 1971.
- 703 Kerekhof, F.: Barnacles (Cirripedia, Balanomorpha) in Belgian waters, an overview of the  
704 species and recent evolutions, with emphasis on exotic species, *Bull. kon. Belg. Inst.*  
705 *Natuurw., Biol.*, 72(Suppl.), 93–104, 2002.
- 706 Khim, B., Krantz, D. E., Cooper, L. W. and Grebmeie, J. M.: Seasonal discharge of estuarine  
707 freshwater to the western Chukchi Sea shelf identified in stable isotope profiles of mollusk  
708 shells, *J. Geophys. Res. Ocean.*, 108(C9), doi:https://doi.org/10.1029/2003JC001816, 2003.
- 709 Klein, R. T., Lohmann, K. C. and Thayer, C. W.: Sr/Ca and <sup>13</sup>C/<sup>12</sup>C ratios in skeletal calcite



- 710 of *Mytilus trossulus*: Covariation with metabolic rate, salinity, and carbon isotopic  
711 composition of seawater, *Geochim. Cosmochim. Acta*, 60(21), 4207–4221,  
712 doi:[https://doi.org/10.1016/S0016-7037\(96\)00232-3](https://doi.org/10.1016/S0016-7037(96)00232-3), 1996.
- 713 Kruk-Dowgiałło, L. and Dubrawski, R.: The state of environment of the Gulf of Gdańsk  
714 coastal zone in autumn 1994 and summer 1995, *Oceanol. Stud.*, 27(4), 137–159, 1998.
- 715 Kruk-Dowgiałło, L. and Szaniawska, A.: Gulf of Gdańsk and Puck Bay, in *Ecology of Baltic*  
716 *Coastal Waters*, pp. 139–165., 2008.
- 717 Kuklinski, P. and Taylor, P. D.: Mineralogy of Arctic bryozoan skeletons in a global context,  
718 *Facies*, 55(4), 489–500, doi:10.1007/s10347-009-0179-3, 2009.
- 719 Larsson, J., Lind, E. E., Corell, H., Grahn, M., Smolarz, K. and Lönn, M.: Regional genetic  
720 differentiation in the blue mussel from the Baltic Sea area, *Estuar. Coast. Shelf Sci.*, 195, 98–  
721 109, doi:<https://doi.org/10.1016/j.ecss.2016.06.016>, 2017.
- 722 Laurantson, V., Kotta, J., Orav-Kotta, H. and Kaljurand, K.: Diet of mussels *Mytilus trossulus*  
723 and *Dreissena polymorpha* in a brackish nontidal environment, *Mar. Ecol.*, 35(Suppl 1), 56–  
724 66, doi:<https://doi.org/10.1111/maec.12120>, 2014.
- 725 Lazareth, C. E., Vander Putten, E., André, L. and Dehairs, F.: High-resolution trace element  
726 profiles in shells of the mangrove bivalve *Isognomon ehippium*: A record of environmental  
727 spatio-temporal variations?, *Estuar. Coast. Shelf Sci.*, 57(5–6), 1103–1114,  
728 doi:[https://doi.org/10.1016/S0272-7714\(03\)00013-1](https://doi.org/10.1016/S0272-7714(03)00013-1), 2003.
- 729 Lee, B. G., Wallace, W. G. and Luoma, S. N.: Uptake and loss kinetics of Cd, Cr and Zn in  
730 the bivalves *Potamocorbula amurensis* and *Macoma balthica*: effects of size and salinity, *Mar.*  
731 *Ecol. Prog. Ser.*, 175, 177–189, doi:10.3354/meps175177, 1998.
- 732 Lee, J. and Morse, J. W.: Influences of alkalinity and p CO<sub>2</sub> on CaCO<sub>3</sub> nucleation from  
733 estimated Cretaceous composition seawater representative of “calcite seas”, *Geology*, 38(2),  
734 115–118, doi:<https://doi.org/10.1130/G30537.1>, 2010.
- 735 Lowenstam, H. A. and Weiner, S.: *On Biomineralization*, Oxford University Press, New  
736 York., 1989.
- 737 Loxton, J., Kuklinski, P., Barnes, D., Najorka, J., Spencer Jones, M. and Porter, J.: Variability  
738 of Mg-calcite in Antarctic bryozoan skeletons across spatial scales, *Mar. Ecol. Prog. Ser.*,  
739 507, 169–180, doi:10.3354/meps10826, 2014.
- 740 Luoma, S. N. and Jenne, E. A.: Factors affecting the availability of sediment-bound cadmium  
741 to the estuarine, deposit-feeding clam, *Macoma balthica*, in *Radioecology and energy*  
742 *resources.*, 1976.
- 743 Marchitto, T. M., Jones, G. A., Goodfriend, G. A. and Christopher, R. W.: Precise temporal





- 744 correlation of Holocene mollusk shells using sclerochronology, *Quat. Res.*, 53(2), 236–  
745 246, doi:https://doi.org/10.1006/qres.1999.2107, 2000.
- 746 Martincic, D., Kwokal, Z., Peharec, Z., Margus, D. and Branica, M.: Distribution of Zn, Pb,  
747 Cd and Cu between seawater and transplanted mussels (*Mytilus galloprovincialis*), *Sci. Total*  
748 *Environ.*, 119, 211–230, doi:https://doi.org/10.1016/0048-9697(92)90265-T, 1992.
- 749 Morse, J. W., Wang, Q. and Tsio, M. Y.: Influences of temperature and Mg:Ca ratio on  
750 CaCO<sub>3</sub> precipitates from seawater, *Geology*, 25(1), 85–87, doi:https://doi.org/10.1130/0091-  
751 7613(1997)025<0085:IOTAMC>2.3.CO;2, 1997.
- 752 Morse, J. W., Arvidson, R. S. and Lüttge, A.: Calcium Carbonate Formation and Dissolution,  
753 *Chem. Rev.*, 107(2), 342–381, doi:https://doi.org/10.1016/0166-445X(82)90002-9, 2007.
- 754 Newman, M. C. and McIntosh, A. W.: Influence of lead in components of a fresh water  
755 ecosystem on molluscan tissue lead concentrations, *Aquat. Toxicol.*, 2, 25–29,  
756 doi:https://doi.org/10.1016/0166-445X(82)90002-9, 1982.
- 757 De Nooijer, L. J., Spero, H. J., Erez, J., Bijma, J. and Reichart, G. J.: Biomineralization in  
758 perforate foraminifera, *Earth-Science Rev.*, 135, 48–58,  
759 doi:https://doi.org/10.1016/j.earscirev.2014.03.013, 2014.
- 760 Pearce, N. J. G. and Mann, V. L.: Trace metal variations in the shells of *Ensis siliqua* record  
761 pollution and environmental conditions in the sea to the west of mainland Britain, *Mar. Pollut.*  
762 *Bull.*, 52(7), 739–755, doi:https://doi.org/10.1016/j.marpolbul.2005.11.003, 2006.
- 763 Pierscieniak, K., Grzymała, J., Wolowicz, M., Pierścieniak, K., Grzymała, J. and Wołowicz,  
764 M.: Differences in reproduction and condition of *Macoma balthica* and *Mytilus trossulus* in  
765 the Gulf of Gdansk (Southern Baltic Sea) under anthropogenic influences, *Int. J. Oceanogr.*  
766 *Hydrol.*, 39(4), 17–32, doi:https://doi.org/10.2478/v10009-010-0054-0, 2010.
- 767 Piwoni-Piórewicz, A., Kukliński, P., Strekopytov, S., Humphreys-Williams, E., Najorka, J.  
768 and Iglukowska, A.: Size effect on the mineralogy and chemistry of *Mytilus trossulus* shells  
769 from the southern Baltic Sea: implications for environmental monitoring, *Environ. Monit.*  
770 *Assess.*, 189(4), 2017.
- 771 Ponnurangam, A., Bau, M., Brenner, M. and Koschinsky, A.: Mussel shells of *Mytilus edulis*  
772 as bioarchives of the distribution of rare earth elements and yttrium in seawater and the  
773 potential impact of pH and temperature on their partitioning behavior, *Biogeosciences*, 13(3),  
774 751–760, 2016.
- 775 Poulain, C., Gillikin, D. P., Thébault, J., Munaron, J. M., Bohn, M., Robert, R., Paulet, Y. M.  
776 and Lorrain, A.: An evaluation of Mg/Ca, Sr/Ca, and Ba/Ca ratios as environmental proxies in  
777 aragonite bivalve shells, *Chem. Geol.*, 396, 42–50, doi:10.1016/j.chemgeo.2014.12.019, 2015.



- 778 Protasowicki, M., Dural, M. and Jaremek, J.: Trace metals in the shells of blue mussels  
779 (*Mytilus edulis*) from the Poland coast of Baltic sea, *Environ. Monit. Assess.*, 141(1–3), 329–  
780 37, 2008.
- 781 Pruszek, Z., van Ninh, P., Szmytkiewicz, M., Hung, N. M. and Ostrowski, R.: Hydrology and  
782 morphology of two river mouth regions (temperate Vistula Delta and subtropical Red River  
783 Delta), *Oceanologia*, 47(3), 365–385, 2005.
- 784 Putnis, A.: An introduction to mineral sciences, Cambridge Univ. Press, 1992.
- 785 Vander Putten, E., Dehairs, F., Keppens, E. and Baeyens, W.: High resolution distribution of  
786 trace elements in the calcite shell layer of modern *Mytilus edulis*: environmental and  
787 biological controls, *Geochim. Cosmochim. Acta*, 64(6), 997–1011,  
788 doi:[https://doi.org/10.1016/S0016-7037\(99\)00380-4](https://doi.org/10.1016/S0016-7037(99)00380-4), 2000.
- 789 Rainbow, P.: Trace metal concentrations in aquatic invertebrates: why and so what?, *Environ.*  
790 *Pollut.*, 120(3), 497–507, doi:[https://doi.org/10.1016/S0269-7491\(02\)00238-5](https://doi.org/10.1016/S0269-7491(02)00238-5), 2002.
- 791 Rainbow, P. S.: Physiology, physicochemistry and metal uptake—A crustacean perspective,  
792 *Mar. Pollut. Bull.*, 31(1–3), 55–59, doi:[https://doi.org/10.1016/0025-326X\(95\)00005-8](https://doi.org/10.1016/0025-326X(95)00005-8), 1995.
- 793 Rainbow, P. S., Amiard-Triquet, C., Amiard, J. C., Smith, B. D., Best, S. L., Nassiri, Y. and  
794 Langston, W. J.: Trace metal uptake rates in crustaceans (amphipods and crabs) from coastal  
795 sites in NW Europe differentially enriched with trace metals, *Mar. Ecol. Prog. Ser.*, 183, 189–  
796 203, doi:[doi:10.3354/meps183189](https://doi.org/10.3354/meps183189), 1999.
- 797 Rainbow, P. S., Wolowicz, M., Fialkowski, W., Smith, B. D. and Sokolowski, A.:  
798 Biomonitoring of trace metals in the Gulf of Gdansk , using mussels (*Mytilus trossulus*) and  
799 barnacles (*Balanus improvisus*), *Water Res.*, 34(6), 1823–1829,  
800 doi:[https://doi.org/10.1016/S0043-1354\(99\)00345-0](https://doi.org/10.1016/S0043-1354(99)00345-0), 2000.
- 801 Rainbow, P. S., Fialkowski, W., Sokolowski, A., Smith, B. D. and Wolowicz, M.:  
802 Geographical and seasonal variation of trace metal bioavailabilities in the Gulf of Gdansk,  
803 Baltic Sea using mussels (*Mytilus trossulus*) and barnacles (*Balanus improvisus*) as  
804 biomonitors, *Mar. Biol.*, 144(2), 271–286, doi:<https://doi.org/10.1007/s00227-003-1197-2>,  
805 2004.
- 806 Reeder, R. J.: Crystal chemistry of the rhombohedral carbonates, *Rev. Mineral.*, 11(1), 1–47,  
807 1983.
- 808 Ritz, D. A., Swain, R. and Elliott, N. G.: Use of the Mussel *Mytilus edulis planulatus*  
809 (*Lamarck*) in Monitoring Heavy Metal Levels in Seawater, *Mar. Freshw. Res.*, (1973), 491–  
810 506, doi:<https://doi.org/10.1071/MF9820491>, 1982.
- 811 Rodland, D. L., Schöne, B. R., Helama, S., Nielsen, J. K. and Baier, S.: A clockwork mollusc:



- 812 Ultradian rhythms in bivalve activity revealed by digital photography, *J. Exp. Mar. Bio. Ecol.*,  
813 334(2), 316–323, doi:<https://doi.org/10.1016/j.jembe.2006.02.012>, 2006.
- 814 Rosenberg, G. D. and Hughes, W. W.: A metabolic model for the determination of shell  
815 composition in the bivalve mollusc, *Mytilus edulis*, *Lethaia*, 24, 83–96,  
816 doi:<https://doi.org/10.1111/j.1502-3931.1991.tb01182.x>, 1991.
- 817 Sartori, A. F. and Gofas, S.: *Macoma balthica* (Linnaeus, 1758), MolluscaBase [online]  
818 Available from: <http://www.marinespecies.org/aphia.php?p=taxdetails&id=141579> (Accessed  
819 5 February 2018), 2016.
- 820 Schoemann, V., DeBaar, H. J. W., DeJong, J. T. M. and Lancelot, C.: Effects of  
821 phytoplankton blooms on the cycling of manganese and iron in coastal waters, *Limnol.*  
822 *Oceanogr.*, 43(7), 1427–1441, doi:<https://doi.org/10.4319/lo.1998.43.7.1427>, 1998.
- 823 Schöne, B. R. and Krause Jr, R. A.: Retrospective environmental biomonitoring – Mussel  
824 Watch expanded, *Glob. Planet. Change*, 144, 228–251,  
825 doi:<https://doi.org/10.1016/j.gloplacha.2016.08.002>, 2016.
- 826 Schöne, B. R., Zhang, Z., Jacob, D., Gillikin, D. P., Tütken, T., Garbe-Schönberg, D.,  
827 McConnaughey, T. and Soldati, A.: Effect of organic matrices on the determination of the  
828 trace element chemistry (Mg, Sr, Mg/Ca, Sr/Ca) of aragonitic bivalve shells (*Arctica*  
829 *islandica*) - Comparison of ICP-OES and LA-ICP-MS data, *Geochem. J.*, 44(1), 23–37,  
830 doi:<https://doi.org/10.2343/geochemj.1.0045>, 2010.
- 831 Schöne, B. R., Zhang, Z., Radermacher, P., Thébault, J., Jacob, D. E., Nunn, E. V. and  
832 Maurer, A.: Sr/Ca and Mg/Ca ratios of ontogenetically old, long-lived bivalve shells (*Arctica*  
833 *islandica*) and their function as paleotemperature proxies, *Palaeogeogr. Palaeoclimatol.*  
834 *Palaeoecol.*, 302(1–2), 52–64, doi:<https://doi.org/10.1016/j.palaeo.2010.03.016>, 2011.
- 835 Smith, A. M. and Girvan, E.: Understanding a bimineralic bryozoan: Skeletal structure and  
836 carbonate mineralogy of *Odontionella cyclops* (Foveolariidae: Cheilostomata: Bryozoa) in  
837 New Zealand, *Palaeogeogr. Palaeoclimatol. Palaeoecol.*, 289(1–4), 113–122,  
838 doi:<https://doi.org/10.1016/j.palaeo.2010.02.022>, 2010.
- 839 Smith, A. M., Nelson, C. S. and Spencer, H. G.: Skeletal carbonate mineralogy of New  
840 Zealand bryozoans, *Mar. Geol.*, 151(1–4), 27–46, 1998.
- 841 Smith, A. M., Key, M. M. and Gordon, D. P.: Skeletal mineralogy of bryozoans: Taxonomic  
842 and temporal patterns, *Earth Sci. Rev.*, 78(3–4), 287–306,  
843 doi:<https://doi.org/10.1016/j.earscirev.2006.06.001>, 2006.
- 844 Staniszewska, M., Nehring, I. and Mudrak-Cegiołka, S.: Changes of concentrations and  
845 possibility of accumulation of bisphenol A and alkylphenols, depending on biomass and



- 846 composition, in zooplankton of the Southern Baltic (Gulf of Gdansk), Environ. Pollut.,  
847 213(June), 489–501, doi:<https://doi.org/10.1016/j.envpol.2016.03.004>, 2016.
- 848 Strasser, M., Walensky, M. and Reise, K.: Juvenile-adult distribution of the bivalve *Mya*  
849 *arenaria* on intertidal flats in the Wadden Sea: why are there so few year classes?, Helgol.  
850 Mar. Res., 33(1), 45, doi:DOI <https://doi.org/10.1007/PL00012137>, 1999.
- 851 Strelkov, P., Nikula, R. and Väinölä, R.: *Macoma balthica* in the White and Barents Seas:  
852 properties of a widespread marine hybrid swarm (Mollusca: Bivalvia), Mol. Ecol., 16(19),  
853 4110–4127, doi:<https://doi.org/10.1111/j.1365-294X.2007.03463.x>, 2007.
- 854 Sunagawa, I., Takahashi, Y. and Imai, H.: Strontium and aragonite-calcite precipitation, J.  
855 Mineral. Petrol. Sci., 102(3), 174–181, doi:<https://doi.org/10.2465/jmps.060327a>, 2007.
- 856 Szefer, P., Szefer, K., Glasby, G. P., Pempkowiak, J. and Kaliszan, R.: Heavy-metal pollution  
857 in surficial sediments from the Southern Baltic sea off Poland, J. Environ. Sci. Heal. Part A,  
858 31(10), 2723–2754, doi:<https://doi.org/10.1080/10934529609376520>, 1996.
- 859 Szefer, P., Frelek, K., Szefer, K., Lee, C. B., Kim, B. S., Warzocha, J., Zdrojewska, I. and  
860 Ciesielski, T.: Distribution and relationships of trace metals in soft tissue, byssus and shells of  
861 *Mytilus edulis trossulus* from the southern Baltic, Environ. Pollut., 120(2), 423–444,  
862 doi:[https://doi.org/10.1016/S0269-7491\(02\)00111-2](https://doi.org/10.1016/S0269-7491(02)00111-2), 2002.
- 863 Szumiło-Pilarska, E., Grajewska, A., Falkowska, L. and Hajdrych, J.: Species differences in  
864 total mercury concentration in gulls from the Gulf of Gdansk (Southern Baltic), J. Trace  
865 Elem. Med. Biol., 33, 100–109, doi:<https://doi.org/10.1016/j.jtemb.2015.09.005>, 2016.
- 866 Taylor, J. D. and Reid, D. G.: Shell microstructure and mineralogy of the Littorinidae:  
867 ecological and evolutionary significance, in Progress in Littorinid and Muricid Biology, pp.  
868 199–215., 1990.
- 869 Taylor, P. D., James, N. P., Bone, Y., Kuklinski, P., Kyser, T. K., Taylor, P. D., James, N. P.,  
870 Bone, Y., Kuklinski, P. and Kyser, T. K.: Evolving mineralogy of cheilostome bryozoans,  
871 Palaios, 24(7), 440–452, doi:<https://doi.org/10.2110/palo.2008.p08-124r>, 2008.
- 872 Taylor, P. D., Kudryavtsev, U. A. B. and Schopf, J. W.: Calcite and aragonite distributions in  
873 the skeletons of bimineralic bryozoans as revealed by Raman spectroscopy, Invertebr. Biol.,  
874 127(1), 87–97, doi:<https://doi.org/10.1111/j.1744-7410.2007.00106.x>, 2014.
- 875 Ullmann, C. V., Gale, A. S., Huggett, J., Wray, D., Frei, R., Korte, C., Broom-Fendley, S.,  
876 Littler, K. and Hesselbo, S. P.: The geochemistry of modern calcareous barnacle shells and  
877 applications for palaeoenvironmental studies, Geochim. Cosmochim. Acta, 243, 149–168,  
878 doi:<https://doi.org/10.1016/j.gca.2018.09.010>, 2018.
- 879 Urban-Malinga, B., Warzocha, J. and Zalewski, M.: Effects of the invasive polychaete



- 880 Marenzelleria spp. on benthic processes and meiobenthos of a species-poor brackish system,  
881 J. Sea Res., 80, 25–34, doi:https://doi.org/10.1016/j.seares.2013.02.005, 2013.
- 882 Urey, H. C., Lowenstam, H. A., Epstein, S. and McKinney, C. R.: Measurement of  
883 paleotemperatures and temperatures of the Upper Cretaceous of England, Denmark, and the  
884 southeastern United States, Geol. Soc. Am. Bull., 62, 399–416,  
885 doi:https://doi.org/10.1130/0016-7606(1951)62[399:MOPATO]2.0.CO;2, 1951.
- 886 Uścinowicz, S., Ed.: Geochemia osadów powierzchniowych Morza Bałtyckiego, Państwowy  
887 Instytut Geologiczny – Państwowy Instytut Badawczy, Warszawa., 2011.
- 888 Wang, Y. and Xu, H.: Prediction of trace metal partitioning between minerals and aqueous  
889 solutions: a linear free energy correlation approach, Geochim. Cosmochim. Acta, 65(10),  
890 1529–1543, doi:https://doi.org/10.1016/S0016-7037(01)00551-8, 2001.
- 891 Warter, V., Erez, J. and Müller, W.: Environmental and physiological controls on daily trace  
892 element incorporation in *Tridacna crocea* from combined laboratory culturing and ultra-high  
893 resolution LA-ICP-MS analysis, Palaeogeogr. Palaeoclimatol. Palaeoecol., 496, 32–47  
894 [online] Available from: https://doi.org/10.1016/j.palaeo.2017.12.038, 2018.
- 895 Watson, D., Foster, P. and Walker, G.: Barnacle shells as biomonitoring material, Mar. Pollut.  
896 Bull., 31(1–3), 111–115, doi:https://doi.org/10.1016/0025-326X(95)00021-E, 1995.
- 897 Watson, S. A., Peck, L. S., Tyler, P. A., Southgate, P. C., Tan, K. S., Day, R. W. and Morley,  
898 S. A.: Marine invertebrate skeleton size varies with latitude, temperature and carbonate  
899 saturation: Implications for global change and ocean acidification, Glob. Chang. Biol., 18(10),  
900 3026–3038, doi:https://doi.org/10.1111/j.1365-2486.2012.02755.x, 2012.
- 901 Weidema, I. R.: Introduced species in the Nordic countries, Nord. Counc. Minist., 13, 242,  
902 2000.
- 903 Weiner, S., Israel, R., Dove, P. M., Tech, V. and Virginia, U. S. A.: An Overview of  
904 Biomineralization Processes and the Problem of the Vital Effect, Rev. Mineral. geochemistry,  
905 54(1), 1–29, doi:https://doi.org/10.2113/0540001, 2001.
- 906 Wenne, R., Bach, L., Zbawicka, M., Strand, J. and McDonald, J. H.: A first report on  
907 coexistence and hybridization of *Mytilus trossulus* and *M. edulis* mussels in Greenland, Polar  
908 Biol., 39(2), 343–355, doi:10.1007/s00300-015-1785-x, 2016.
- 909 Wit, J. C., Nooijer, L. J. D., Wolthers, M. and Reichart, G. J.: A novel salinity proxy based on  
910 Na incorporation into foraminiferal calcite, Biogeosciences, 10, 6375–6387,  
911 doi:https://doi.org/10.5194/bg-10-6375-2013, 2013.
- 912 Zacherl, D. C., Paradis, G. and Lea, D. W.: Barium and strontium uptake into larval  
913 protoconchs and statoliths of the marine neogastropod *Kelletia kelletii*, Geochim.



914 Cosmochim. Acta, 67(21), 4091–4099, doi:https://doi.org/10.1016/S0016-7037(03)00384-3,  
 915 2003.

916 Żmudziński, L.: Świat zwierzęcy Bałtyku – atlas makrofauny, WSiP, Warszawa., 1990.

917

918 Tables and Figures

919 **Table 1** Details of the research stations utilized for the collection of organisms. The  
 920 environmental parameters T (temperature) and salinity were measured near the bottom during  
 921 sample collection

Station	Latitude N	Longitude E	Collected species	Depth (m)	T (°C)	Salinity	Date
GN	54°32.0'	18°48.1'	<i>Amphibalanus improvisus</i>	36	3.1	7.3	May 2013
GN	54°32.0'	18°48.1'	<i>Mytilus trossulus</i>	36	3.1	7.3	May 2013
M2	54°38.9'	18°33.0'	<i>Cerastoderma glaucum</i>	10	16.9	6.9	June 2014
MA	54°37.2'	18°32.8'	<i>Mya arenaria</i>	10	19.8	6.9	June 2014
MW	54°37.4'	18°37.5'	<i>Limecola balthica</i>	31	4.7	7.0	June 2014

922

923 **Table 2** The range of shell sizes for each studied species within a given size class

Species	Mineral type	Size classes (mm)				Maximum size in the Gulf of Gdansk (mm)*
		1	2	3	4	
<i>Amphibalanus improvisus</i>	Calcite	3–4	5–6	7–8	9–10	10
<i>Mytilus trossulus</i>	Bimineralic	6–5	16–25	26–35	36–44	50
<i>Cerastoderma glaucum</i>	Aragonite	4–8	9–12	12–16	16–20	27
<i>Mya arenaria</i>	Aragonite	10–20	20–30	30–40	40–49	70
<i>Limecola balthica</i>	Aragonite	4–7	8–11	12–15	15–18	24

\*Based on the literature data (Zmudziński, 1990)

924

925

926 **Table 3** The concentrations of the studied elements in shells of the collected organisms

<i>Amphibalanus improvisus</i>							
Elements	Mineral type	Unit	N	Mean	±1 SD	Min	Max
Ca	Calcite	wt %	37	31.6	2.60	27.6	39.0
Mg			37	0.394	0.032	0.337	0.472
Na			37	0.285	0.047	0.222	0.508
Sr			37	0.225	0.019	0.198	0.283
V		mg kg <sup>-1</sup>	36	0.466	0.156	0.155	0.78
Mn			37	625	160	414	1161
Cu			37	3.90	2.20	1.20	8.30
Y			37	0.21	0.12	0.070	0.78
Cd			37	0.062	0.020	0.029	0.125
Ba			37	73.0	20.1	59.0	145
Pb	37	1.34	0.79	0.456	3.86		
U	37	0.108	0.043	0.035	0.220		
<i>Cerastoderma glaucum</i>							
Elements	Mineral type	Unit	N	Mean	±1 SD	Min	Max
Ca			40	34.8	1.60	30.4	38.1



Mg			40	0.008	0.003	0.005	0.023
Na			40	0.322	0.021	0.277	0.361
Sr			40	0.170	0.014	0.143	0.199
-----							
V			37	0.069	0.061	0.003	0.239
Mn			40	11.9	10.5	0.614	48.8
Cu			40	4.08	6.29	0.188	30.6
Y			40	0.070	0.074	0.006	0.401
Cd			37	0.011	0.009	0.001	0.043
Ba			40	15.0	10.0	1.22	35.4
Pb			34	0.14	0.10	0.006	0.298
U			40	0.096	0.084	0.003	0.339

*Limecola balthica*

Elements	Mineral type	Unit	N	Mean	±1 SD	Min	Max
Ca		wt %	40	34.5	1.70	29.0	37.5
Mg			40	0.008	0.003	0.003	0.018
Na			40	0.31	0.03	0.25	0.367
Sr			40	0.177	0.021	0.135	0.244
-----							
V	Aragonite	mg kg <sup>-1</sup>	35	0.035	0.031	0.002	0.102
Mn			40	7.90	7.10	0.264	28.5
Cu			40	3.00	3.40	0.047	11.5
Y			40	0.08	0.11	0.002	0.421
Cd			37	0.007	0.008	0.0004	0.031
Ba			40	13.0	10.0	0.638	27.6
Pb			39	0.14	0.16	0.007	0.585
U			40	0.034	0.024	0.004	0.088

*Mya arenaria*

Elements	Mineral type	Unit	N	Mean	±1 SD	Min	Max
Ca		wt %	40	36.3	1.8	30.5	39.1
Mg			40	0.015	0.002	0.011	0.021
Na			40	0.308	0.021	0.26	0.345
Sr			40	0.216	0.028	0.168	0.295
-----							
V	Aragonite	mg kg <sup>-1</sup>	40	0.016	0.014	0.001	0.062
Mn			40	1.40	1.00	0.276	5.32
Cu			40	0.031	0.024	0.006	0.114
Y			40	0.020	0.018	0.003	0.089
Cd			37	0.002	0.003	0.0002	0.016
Ba			40	2.60	1.60	0.574	6.26
Pb			40	0.015	0.014	0.002	0.077
U			40	0.006	0.005	0.001	0.022

*Mytilus trossulus*

Elements	Mineral type	Unit	N	Mean	±1 SD	Min	Max
Ca		wt %	40	34.4	2.8	25.2	39.7
Mg			40	0.106	0.017	0.061	0.14
Na			21	0.201	0.031	0.147	0.291
Sr			41	0.117	0.01	0.101	0.143
-----							
V	Bimineralic	mg kg <sup>-1</sup>	40	0.78	0.52	0.206	2.30
Mn			35	54.0	15.0	31.8	92.6
Cu			40	14.0	10.0	3.36	58.3
Y			40	0.70	1.20	0.040	4.86
Cd			40	0.091	0.11	0.022	0.689



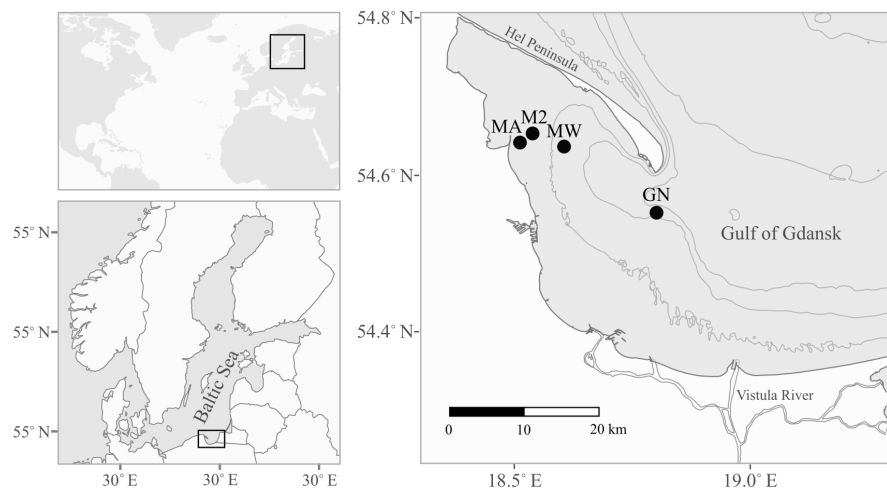


Ba			38	17.0	6.70	8.97	38.7
Pb			32	1.00	1.40	0.143	6.26
U			40	0.053	0.033	0.018	0.129

927

928

929

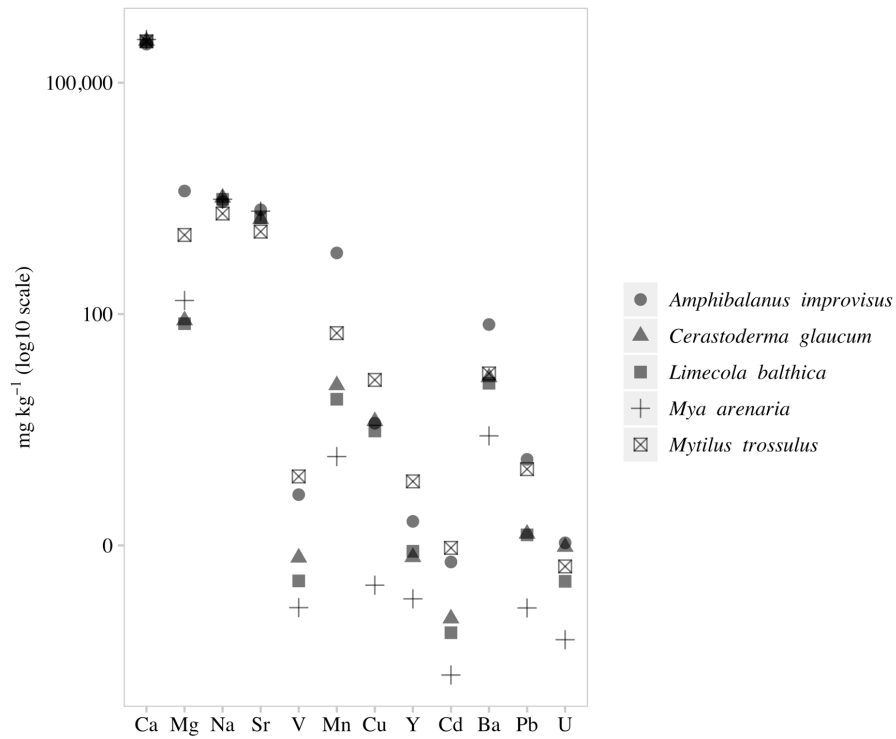


930

931 **Fig. 1** The study area, with sampling stations indicated by black dots (stations details see

932 Table 1)

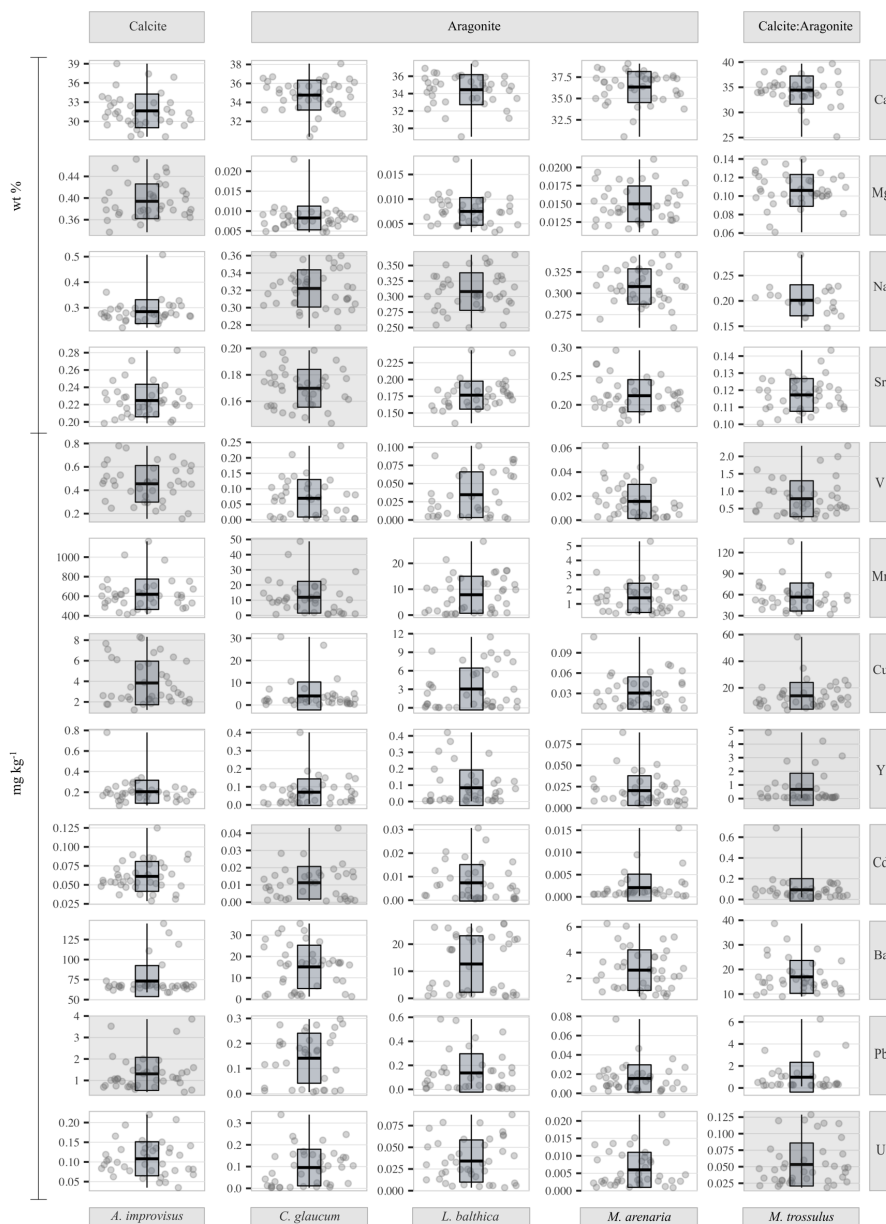
933



934

935 **Fig. 2** The distribution of logarithmically transformed mean metal concentrations in shells of  
936 *Amphibalanus improvisus*, *Cerastoderma glaucum*, *Limecola balthica*, *Mya arenaria* and  
937 *Mytilus trossulus*

938

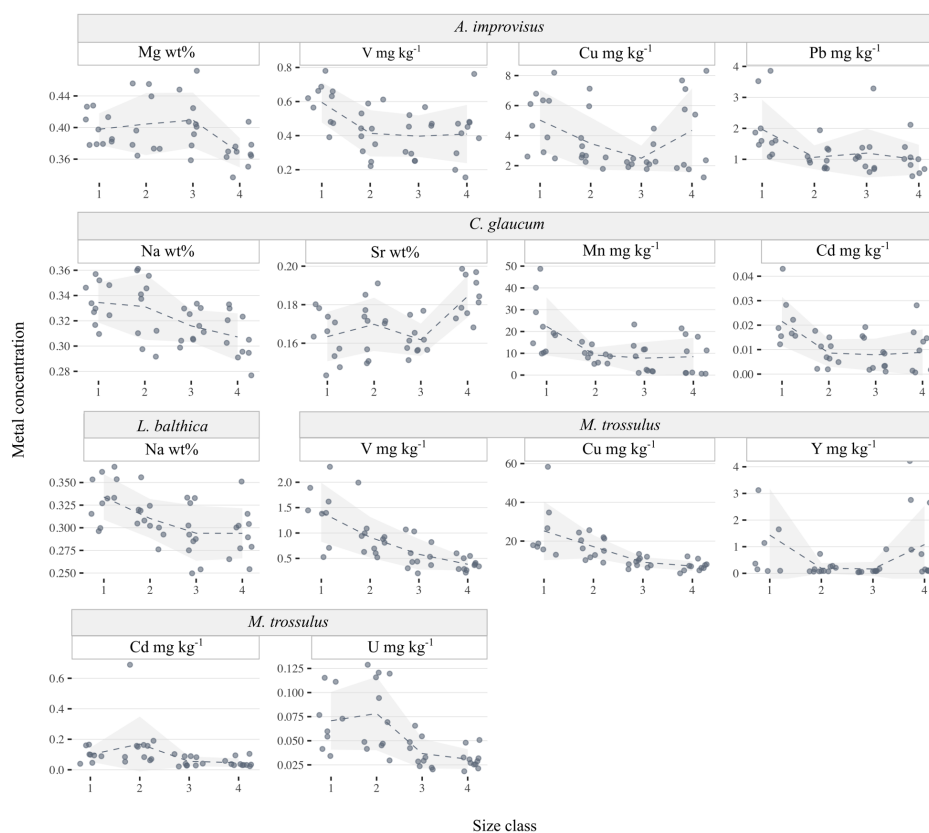


939

940 **Fig. 3** Metal concentrations in analysed shells of *Amphibalanus improvisus*, *Cerastoderma*  
 941 *glaucum*, *Limecola balthica*, *Mya arenaria* and *Mytilus trossulus*. The figure shows the raw  
 942 data as grey dots, with boxplots representing a  $\pm 1$  standard deviation around the means and  
 943 whiskers indicating the minimum and maximum concentrations. The grey areas show  
 944 elements with statistically significant differences between size classes



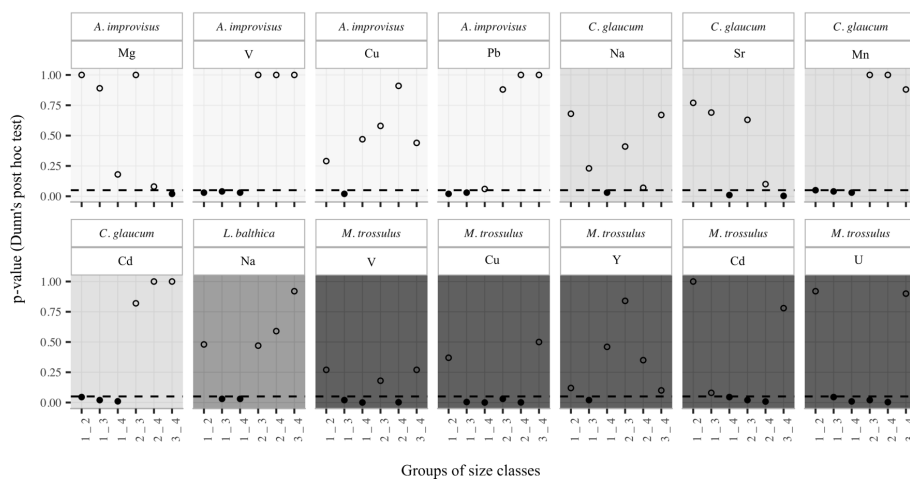
945



946

947 **Fig. 4** Concentrations of elements with statistically significant differences between the four  
948 size classes (for size class details see Table 2) in the shells of *Amphibalanus improvisus*,  
949 *Cerastoderma glaucum*, *Limecola balthica* and *Mytilus trossulus*. The figure shows the raw  
950 data as black dots, with the light grey areas representing the standard deviations around the  
951 means (middle line)

952



953

954 **Fig. 5** Pairwise comparisons of metal concentrations between size classes (for size class  
955 details see Table 2) in the shells of *Amphibalanus improvisus*, *Cerastoderma glaucum*,  
956 *Limecola balthica* and *Mytilus trossulus*. The broken line separates statistically significant  
957 dependence (black points) from statistically non-significant dependence (empty points)