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2	A revised northern soil Hg pool, based on western Siberia permafrost peat Hg
3	and carbon observations
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28 Abstract

29	Natural and anthropogenic mercury (Hg) emissions are sequestered in terrestrial soils over short,
30	annual, to long, millennial time scales, before Hg mobilization and run-off impacts wetland and
31	coastal Ocean ecosystems. Recent studies have used Hg to carbon (C) ratios, R_{HgC} , measured in
32	Alaskan permafrost mineral and peat soils, together with a northern soil carbon inventory, to
33	estimate that these soils contain large amounts, 184 to 755 Gg of Hg in the upper 1 m. However,
34	measurements of R_{HgC} on Siberian permafrost peatlands are largely missing, leaving the size of
35	estimated northern soil Hg budget, and its fate under arctic warming scenarios uncertain. Here
36	we present Hg and carbon data for 6 peat cores, down to mineral horizons at 1.5 - 4 m depth,
37	across a 1700 km latitudinal (56 to 67°N) permafrost gradient in the Western Siberian lowlands
38	(WSL). Hg concentrations increase from south to north in all soil horizons, reflecting enhanced
39	net accumulation of atmospheric gaseous Hg by the vegetation Hg pump. The $R_{\rm HgC}$ in WSL peat
40	horizons decreases with depth from 0.38 Gg Pg^{-1} in the active layer to 0.23 Gg Pg^{-1} in
41	continuously frozen peat of the WSL. We estimate the Hg pool (0-1 m) in the permafrost-
42	affected part of WSL peatlands to be 9.3 \pm 2.7 Gg. We review and estimate pan-arctic organic
43	and mineral soil R_{HgC} to be 0.19 and 0.77 Gg Pg ⁻¹ , and use a soil carbon budget to revise the
44	northern soil Hg pool to be 67 Gg (37-88 Gg, interquartile range (IQR)) in the upper 30 cm, 225
45	Gg (102-320 Gg) in the upper 1 m, and 557 Gg (371-699 Gg) in the upper 3 m. Using the same
46	R_{HgC} approach, we revise the global upper 30 cm soil Hg pool to contain 1078 Gg of Hg (842-
47	1254 Gg, IQR), of which 6% (67 Gg) resides in northern permafrost soils. Additional soil and
48	river studies must be performed in Eastern and Northern Siberia to lower the uncertainty on
49	these estimates, and assess the timing of Hg release to atmosphere and rivers.

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53 **1. Introduction**

54 High-latitude organic-rich soils are key ecosystems controlling the transfer of carbon, 55 nutrients, and pollutants between the atmosphere, rivers, lakes and the Arctic Ocean. These soils 56 are most vulnerable to on-going climate change, due to the high mobility of carbon stored in the 57 form of peat deposits. Part of the peat layers are currently frozen but may be subjected to fast 58 thaw, especially in discontinuous and sporadic permafrost zones (Romanovsky et al., 2010). 59 Whilst the stock of C in Arctic and subarctic peat and mineral soils is fairly well quantified (472 Pg C (±27 Pg, 95% confidence interval, CI) in the upper 0-1m (Hugelius et al. 2014), this is not 60 61 true for pollutants such as mercury (Hg). Because of its strong bio-amplification in Arctic marine 62 biota (Morel et al., 1998), and exposure to native Arctic populations (AMAP, 2011), there is a 63 strong interest in understanding Hg biogeochemistry in Arctic environments (Outridge et al., 64 2008; Steffen et al., 2008; Stern et al., 2012).

Recent advances in quantifying Arctic Hg cycling show that Arctic Hg^{II} wet deposition is 65 generally low (Pearson et al., 2019), and that the vegetation Hg pump drives yearlong net gaseous 66 Hg⁰ (and CO₂) deposition, via foliar uptake to Arctic vegetation and litterfall to soils (Obrist et al. 67 68 2017; Jiskra et al. 2018; Jiskra et al., 2019). Soil core analyses in Alaska indicate that large 69 amounts of carbon and Hg have accumulated since the last glacial maximum, and two upscaling 70 approaches to Hg stocks in pan-Arctic permafrost soils resulted in differing estimates of 184 Gg 71 and 755 Gg for the upper 1 m (Schuster et al., 2018; Olson et al., 2018). Despite the overall net 72 atmospheric Hg deposition to soils, research has found that Arctic rivers export 44 Mg y⁻¹ of soil 73 Hg, bound to particulate and dissolved organic matter, to the Arctic Ocean (Fisher et al., 2012; 74 Dastoor et al., 2014; Zhang et al., 2015; Sonke et al., 2018). Together with coastal erosion of soils (30 Mg y^{-1}) , river Hg inputs constitute a terrestrial Hg flux of 74 Mg y⁻¹ to the Arctic Ocean that 75 76 is of similar magnitude to gross atmospheric deposition over the Arctic Ocean (80 Mg y⁻¹, Sonke 77 et al., 2018). Permafrost thawing has been shown to enhance river Hg export from soils to rivers





78 (St Pierre et al., 2018), is most pronounced in the discontinuous permafrost zone, and has been 79 suggested to potentially double over the next 50 years (Lim et al., 2019). The quantity of 80 atmospheric Hg deposition to northern peat soils that is presently re-emitted to the atmosphere is not well understood. Hg exchange studies indicate temporally limited Hg⁰ emission from the 81 Alaskan permafrost tundra at 68°N (Obrist et al., 2017), and strong year round net Hg⁰ emission 82 83 from Scandinavian peat at 64°N (Osterwalder et al., 2018). Other studies provide evidence for 84 vegetation type (Rydberg et al., 2010) and temperature and insolation control (Fahnestock et al., 2019) on net Hg⁰ deposition or emission. 85

86 All available data of Hg in permafrost soils originate from N-America or Scandinavia 87 (Jensen et al., 1991; Bailey et al., 2002; Talbot et al., 2017; Schuster et al., 2018; Olson et al., 88 2018). Except for two studies of Hg in a peat profile from a permafrost-free zone of western 89 Siberia (Golovatskaya et al., 2009; Lyapina et al., 2009), we did not find extensive measurements 90 of Hg in peat profiles from permafrost regions of the Russian Arctic and Siberia. Recent work 91 used a soil carbon GIS model to estimate the size of the northern permafrost soil Hg inventory to 92 be 755 \pm 427 Gg (95% CI) in the upper 1 m (Schuster et al., 2018). However, this estimate is 93 based on extrapolation of high Hg to organic carbon (C) ratios, R_{HgC}, of 1.6 Gg Hg per Pg of C (Gg Pg⁻¹) in Alaskan mineral soils to the entire N-American and Eurasian permafrost zone. A 94 second study used lower R_{HgC}, of 0.12 to 0.62 Gg Pg⁻¹, derived from observations on both Alaskan 95 96 organic and mineral soils and literature data, to estimate a lower northern soil 0-1 m Hg inventory 97 of 184 Gg (136-274 Gg, 37.5-62.5 percentiles) (Olson et al., 2018). Direct measurement of soil 98 Hg and carbon profiles in frozen peatlands of Siberia are needed to address these variable 99 estimates, and compare the size of permafrost soil Hg pool to the global soil Hg pool. This 100 constitutes the first and main objective of the present study. The second objective was assessing 101 the impact of permafrost type (absent, sporadic, discontinuous and continuous) on Hg 102 concentrations and pools in the active layer, frozen peat and mineral horizons. The third objective





was to relate Hg concentration in peat to that of other trace metals in order to reveal possible
mechanisms of Hg and other metal pollutant accumulation within the organic and mineral horizons
of frozen peatlands.

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107 **2.** Study Site and Methods

108 2.1.Sampling sites

109 Soil sampling was performed along a latitudinal transect of the western Siberia lowlands 110 (WSL) that comprised the southern taiga (Plotnikovo, 56°N), the middle taiga (Mukhrino, 60°N), 111 the northern part of the taiga zone (Kogalym, 62°N), forest-tundra (Khanymey 62°N and Pangody, 112 65°N) and tundra (Tazovsky, 67°N) biomes (Fig. 1). In the WSL, the permafrost zones follow the 113 temperature and vegetation distribution over the latitude at otherwise similar relief, lithology and 114 runoff, thus allowing to test the effect of permafrost by analyzing latitudinal features of Hg 115 distribution in soils. Key physico-geographical parameters of studied sites and soil types are listed 116 in Table S1 of the Supplementary Information. The WSL peat actively formed since the beginning 117 of the Holocene until freezing of bogs in the sub-Boreal period (11-4.5 ky, Kremenetski et al., 118 2003; Panova et al., 2010; Ponomareva et al., 2012; Loiko et al., 2019). Since 4.5 ky, the rate of 119 peat formation and bog extension in the permafrost-affected part of the WSL have decreased. In 120 the southern part of cryolithozone and permafrost-free part of WSL, peat accumulation and bog 121 extension remained active over the entire Holocene (Kurina and Veretennikova, 2015; Preis and 122 Karpenko, 2015; Kurina et al., 2018). The main mineral substrates underlying frozen peat layers 123 of the WSL are quaternary clays, sands, and alevrolites. In the southern part (sites Plotnikovo and 124 Mukhrino), the typical substrate is carbonate-bearing clays of lake-alluvium origin with rare layers 125 of sandstones (Table S1).

126 Mean annual atmospheric temperature (MAAT) increases from south to north, being equal 127 to -0.4, -1.2, -4.0, -5.6, -6.4, and -9.1°C at Plotnikovo, Mukhrino, Kogalym, Khanymey,





Pandogy and Tazovsky, respectively (Trofimova and Balybina, 2014). The permafrost is absent in Plotnikovo but present at all other sites and ranges from relict to isolated (Mukhrino), isolated to sporadic (Kogalym) in the south, to discontinuous (Khanymey, Pangody) and continuous (Tazovsky) in the north. At permafrost-affected sites, the average active (unfrozen) layer thickness (ALT) at the time of sampling of peat mounds (hummocks) ranged from 90 cm in the south to 45 cm in the north. The peat mounds of ombrotrophic bogs probed in this work are present across the full latitudinal gradient.

135 The vegetation of three studied types of bogs (polygonal, flat-mound and ridge-hollow) is 136 essentially oligotrophic (poor in nutrients) which indicates the ombrotrophic (rain and snow water 137 fed) conditions, i.e., the lack of groundwater input and lateral surface influx of nutrients. The flat-138 mound palsa is covered by dwarf shrubs (Ledum decumbens, Betula nana, Andromeda polifolia, 139 Vaccinium ssp., Empetrum nigrum), lichens (Cladonia ssp., Cetraria, Ochrolechia) and mosses 140 (Dicranum ssp., Polytrichum ssp., Sphagnum angustifolium, S. lenense). At southern sites, the 141 pine *Pinus sylvestris* is abundant on ridges (Peregon et al., 2008, 2009) whereas the two taiga sites 142 are dominated by Pinus sylvestris f. uliginosa with minor but permanently present Betula 143 pubescens and Pinus sibirica. Dwarf shrubs are dominated by Ledum palustre, Chamaedaphne 144 calyculata, Vaccinium vitis-idaea. The moss layer is dominated by Sphagnum fuscum, S. 145 angustifolium with the presence of Sphagnum magellanicum, S. capillifolium and boreal forest 146 moss species like Pleurozium schreberi.

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2.2. Sampling procedure, analyses and data treatment

Peat core samples were collected in August, when the depth of unfrozen layer was at its maximum (i.e., see Raudina et al., 2017). Based on measurements by temperature loggers over the summer period, the in-situ temperature of studied soil profile ranged from $15\pm5^{\circ}$ C in the top soil (0-20 cm) to $4\pm2^{\circ}$ C at the permafrost boundary (40-80 cm). The physical, chemical and





botanical properties of several peat cores collected in the homogeneous palsa region in the north
and ridgre-ryam complex in the south are highly similar among different peat mounds (Velichko
et al., 2011; Stepanova et al., 2015).

156 Peat cores were extracted using a Russian sediment/peat corer and the frozen part was 157 sampled using a motorized Russian peat corer (UKB-12/25 I, Russia) with a 4-cm diameter corer 158 sterilized with 40% ethanol prior to each extraction. We collected the full active (unfrozen) layer 159 peat column, the frozen peat column and some 10 to 30 cm of frozen mineral horizons using clean 160 powder-free vinyl gloves. Peat or mineral soil samples were divided in 5-10 cm segments using a 161 sharp sterile single-use plastic knife. Soil samples were placed in sterile PVC doubled-zipped bags 162 and kept at -20°C during transport and storage. To avoid contamination of peat from external 163 surroundings, we separated the part of the core for geochemical analysis exclusively from the 164 interior of the core (> 1 cm from the core liner) following conventional procedures (Wilhelm et 165 al., 2011).

166 Total Hg concentration, THg, in freeze-dried and ground slices of peat cores was 167 determined using a direct mercury analyzer (DMA-80 - Milestone, Italy). Analysis of reference 168 material BCR-482 (lichen, 480±28 ng g⁻¹), MESS3 (sediment, 91±9 ng g⁻¹) and NIST 1632d (coal, 93±3 ng g⁻¹) showed good reproducibility (mean±1 σ) of 467±28 ng g⁻¹, 80±6 ng g⁻¹ and 98±8 ng 169 170 g^{-1} , respectively. The average uncertainty on duplicate sample analysis did not exceed 5% (1 σ). 171 The carbon (C) and nitrogen (N) concentration were measured using catalytic combustion with 172 Cu-O at 900°C with an uncertainty of $\leq 0.5\%$ using Thermo Flash 2000 CN Analyzer, and aspartic 173 acid (C 36.09%±1.5%; N 10.52%±0.5%) and soil SRM (C 2.29%±0.07%; N 0.21%±0.01%) as 174 standards. Analyses of total C before and after sample treatment with HCl did not yield more than 175 1 % of inorganic C; therefore our total C-determination represents organic carbon. For trace and 176 major element analysis, soil samples were subjected to full acid digestion in the clean room 177 following ICP-MS (Agilent 7500 ce) analyses as described previously (Morgalev et al., 2017).





178	The Shapiro Wilk normality test was used to asses THg, elemental and R_{HgC} distributions,
179	and statistical data descriptors adjusted accordingly. All statistical tests used a significance level
180	of 95% ($\alpha = 0.05$). Spearman rank order correlations (significant at p < 0.05) were performed to
181	characterize the link of Hg with C, N and other major and trace elements. The differences in Hg
182	concentration between the active- and frozen peat layer were tested using the Mann-Whitney U
183	test for paired data at a significance level of 0.05.

184 C pools of different soil classes reported by Hugelius et al. (2014) were divided into two 185 categories, organic and mineral soils. Histosols and Histels were defined as organic soils. Turbels 186 and Orthels were considered as organic soils for the 0 - 0.3 m interval and as mineral soils for the 187 0.3 - 3m interval. All other soils were considered as mineral soils. To estimate the northern soil 188 Hg pool, C pools were multiplied with the respective R_{HgC} derived for organic (>20% C) and 189 mineral (<20% C) soil data from north America (excluding Alaska) and Eurasia. To calculate the 190 global Hg pool, a simpler approach was used and one singe R_{HgC} was considered for 5 climate 191 zones which were defined by latitude (arctic: > 67° , boreal: 50° - 67° , temperate: 35° - 50° , 192 subtropical: 23.45° - 35°, tropical: <23.45°) according to FAO and ITPS (2018). The uncertainty 193 was assessed with a Monte carlo approach using the rnorm and rlnorm function of R (version 3.6.1.) and is reported as the interquartile range (25th and 75th percentile) of 100,000 simulations. 194 195 For the northern soil Hg pool, final uncertainties incorporate the uncertainties on the C stock from 196 Hugelius et al. (2014) assuming normal distribution and uncertainties of R_{HgC} assuming log-197 normal distribution.

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3. Results

204 *3.1. Depth (vertical) distribution of Hg in peat profiles*

205 Hg concentration in peat cores of the WSL are illustrated in Fig. 2 and primary data on 206 soil chemical composition and Hg concentration are listed in Table S2 of the Supplement. The 207 upper 0-20 cm layer is 2 to 3 times enriched in Hg compared to the rest of the peat core in 208 permafrost-affected sites (Khanymey, Pangody and Tazovsky). This is not the case, however, for 209 the sporadic permafrost zone (Kogalym) and isolated permafrost zone (Mukhrino), where a local 210 maximum at ca. 35 cm depth was detected but no enrichment of upper 10-20 cm horizons 211 occurred. In the most southern, permafrost-free site of the WSL (Plotnikovo, southern taiga), the 212 Hg concentration profile in the peat was fairly constant with a local minimum at 100 cm depth. 213 The mean, depth-integrated Hg concentrations in active layer, permafrost and mineral horizons 214 are illustrated in Fig. 3 and summarized in Table 1. The latitudinal trend of Hg concentration in 215 peat consists of a systematic increase northward, both in permafrost and active peat layers. The 216 dominant ground vegetation (lichens) analyzed at 5 sites out of 6 (Plotnikovo, Kogalym, 217 Khanymey, Pangody and Tazovsky) did not show significantly different (U test Mann-Whitney) 218 Hg concentrations relative to the peat cores (Fig. 3). The typical concentrations of Hg in studied peat cores ranged from 7 to 284 ng g⁻¹ with a median (\pm IQR) of 67 \pm 57 ng g⁻¹. The Hg 219 220 concentration in the thawed, active layer was generally comparable to that in the frozen layer, 221 supported by a Mann-Whitney test, which did not show significant difference in Hg concentration 222 between frozen and thawed peat in all permafrost-affected sites. Within the latitudinal transect 223 from south to north, the Hg concentrations in peat are higher (Plotnikovo, Kogalym, Khanymey, 224 Pangody) or comparable (Tazovsky) to those in the mineral horizons.

The ratio of Hg:C (R_{HgC} , $\mu g g^{-1}$, corresponding to Gg Pg⁻¹) ranged between 0.05 and 2.0 over the peat columns, and was 5 to 10 times higher in mineral horizons compared to frozen peat and active layers (**Fig. 4**). The R_{HgC} in the active layer and in the mineral horizons increased 3-





- fold from the south (56°N) to the north (67°N). In the frozen peat horizon, the R_{HgC} ratio increased
- two-fold from sporadic and isolated to continuous permafrost zone.
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231 3.2. Regional and total pools of Hg in the WSL peat and mineral layers

232 The mass of Hg per area of soil in the active- and frozen peat layer as well as in the top 30 233 cm of frozen mineral horizons of the six studied WSL peat profiles was calculated by multiplying 234 bulk soil peat and mineral layer densities (range from 0.01 to 0.38 g/cm³, Table S2) by Hg 235 concentration and integrating over the corresponding depths. The surface area - normalized Hg stock systematically increased from south to north (ca. 0.3 to 6.0 mg Hg m⁻² and ca. 0.8 to 13.7 236 237 mg Hg m⁻², in the 0-30 and the 0-100 cm peat layer, respectively (Fig. 5 A). This northward 238 increase was most pronounced for the active layer, was less evident for frozen peat, and 239 insignificant for the upper 30 cm of mineral horizon located under the peat (Fig. 5 B). Taking into 240 account the proportion of bogs (peatlands) in each zone (1° latitudinal grid) from Sheng et al. 241 (2004), we calculated the pool of Hg in permafrost-free and permafrost-affected WSL peatlands 242 (Fig. 6). The total pool of Hg in the 0-100 cm layer of peat bogs exhibits a maximum (356-580 243 Mg) in the discontinuous permafrost zone.

We estimate the total organic soil Hg pool in the WSL from the Hg stock (mg Hg m⁻² over 244 245 0-100 cm depth) for permafrost and permafrost-free zones (Fig. 7 A), extrapolated to the full 246 average thickness of peat in the WSL (280 cm, Sheng et al., 2004), assuming that Hg concentration 247 in the upper 0-100 cm peat layer is the same as in 100-280 cm of peat and multiplied by the area 248 of bogs in each latitudinal grid (S, m²) as shown in Fig 7 B. This yields 1.7 Gg Hg in the 249 permafrost-free zone and 7.6 Gg Hg in the permafrost-bearing zone with a total Hg pool of 9.3 250 Gg in the WSL. For this calculation we did not take into account the mineral horizons and we used 251 variable active layer thickness across the latitudinal gradient of WSL, as estimated at our sampling 252 sites (Table S1). The amount of Hg in permafrost-bearing zone within the active (unfrozen) peat





- layer (0-160 cm in the south and 0-20 cm in the north) of the WSL is 2.0 Gg, and that in the frozen
 (160-280 cm in the south and 20-280 cm in the north) layer) is 5.6 Gg.
 Alternatively, to calculate the total pool of Hg in WSL bogs, we used the R_{HgC} inferred
 from our data across the gradient of permafrost and biomes (Table 1). Taking into account the C
 pool in the WSL (70.2 Pg C of 0-280 cm depth layer, Sheng et al., 2004) and the median R_{HgC} of
 0.133 µg/g in the WSL, we calculated Hg for the full depth of the peat layer in each zone. This
 also gives 9.3 Gg Hg for a total area of 592,440 km².
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261 *3.3. Correlation of Hg with other elements in the peat cores*

262 Spearman rank order correlations of Hg with other elements demonstrated significant 263 positive relationships (R > 0.60; p < 0.05) with K, Rb, Cs, P, As, W, V, Cr, Cu in the active 264 (unfrozen) layer (**Table S3** of Supplment). However, these relationships were less pronounced in 265 the frozen peat, where only Mg, Ca, Sr, Mn, N, P, As, Cu, Ni, Sb and some REE demonstrated 266 minor (0.40 < R < 0.55) positive correlations with Hg. Finally, in the mineral layer, significant 267 (R > 0.70) positive correlations of Hg were observed with Li, Ca, Sr, P, N, Mn, Ni, Co, Cr, Cd. A 268 positive (R = 0.60) relationship between Hg and C was observed in mineral horizons, whereas no 269 correlation was detected in both frozen and thawed peat. This is consistent with some studies of 270 peat soil in Brazil (Roulet et al., 1998) and Arctic tundra soils (Olson et al., 2018). At the same 271 time, there was a positive correlation of Hg with N in the active layer, frozen peat and mineral 272 horizons (R = 0.50, 0.47 and 0.75, respectively). Stronger and more stable correlation of Hg with 273 N compared to C was also noted by Roulet (2000). 274

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278 4. Discussion

279 4.1. Hg association with other elements in peat

280 Stronger accumulation of Hg relative to C in mineral horizons in the north (Tazovsky, Fig. 281 4) may be linked to the clay nature of mineral layers (Roulet et al., 1998; Baptista-Salazar et al., 282 2017) in these regions (Table S1) but also to the presence of specific host phases of Hg (see 283 examples of peat minerals in Rudmin et al., 2018). Dissolved oxygen measurements in soil 284 porewaters at the Tazovsky site indicate that mineral gleysoils and peat histosols, which often 285 overlay former lake sediments, are anoxic (Raudina et al., 2017; Loiko et al., 2019). The Hg host 286 phases in these soils are therefore likely sulfide minerals. Indeed, known Hg carriers in peat 287 deposits are Fe and Zn sulfide minerals or organic-bound sulfide functional groups (Smieja-Król 288 et al., 2010; 2014; Prietzel et al., 2009; Skyllberg et al., 2003, Bates et al., 1998; Steinmann and 289 Shotyk, 1997).

290 In the peat active layer, Hg was positively correlated with K, Rb, Cs, P, As, V, Cr, Cu 291 (Table S3). In the frozen part of the peat core, Hg was positively correlated with Ca, N, Mn, Sr, 292 Mg, P (Table S3). Indeed, atmospheric particles in snow across the WSL exhibit strong 293 enrichment in Mo, W, As, Sb, Ni, Cu, Zn, Cd, Pb, Mg, Ca, and Na (Shevchenko et al., 2017). The 294 strong positive correlation of Hg with these elements in peat soils of WSL suggests a common 295 atmospheric origin. Note however, that the cited elements deposit with particles, rainfall and 296 snowfall, whereas atmospheric Hg transfer to peat occurs mainly via the vegetation pump, with tundra and taiga vegetation actively taking up atmospheric gaseous Hg⁰ through foliage (Obrist 297 298 et al., 2017; Jiskra et al., 2018).

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300 *4.2. Estimating the northern soil Hg pool*

301 A recent study used a median R_{HgC} value of 1.6 µg g⁻¹, observed mainly in mineral soil 302 samples (median SOC of 3%, IQR=1.7 to 8.7 %) along a transect in Alaska, to estimate a northern





303 permafrost soil Hg pool of 755±427 Gg in the upper 0-100 cm, and 1656±962 Gg in the upper 0-304 300 cm (Schuster et al. 2018). In the case of western Siberia, this high R_{HgC} value overestimates 305 the Hg pool 12-fold, given that the median R_{HgC} in WSL peat is only 0.13±0.12 (median±IQR) 306 (Table 1, Fig. 4). The extrapolation based on Alaskan R_{HgC} for the whole Northern Hemisphere 307 permafrost region also suggests that the WSL contains large amounts of Hg in the upper 0-30 cm 308 $(20-40 \text{ mg Hg m}^{-2})$ and in the upper 0-100 cm $(40-80 \text{ mg Hg m}^{-2})$. These numbers are much higher than the direct measurements in this study: 0.3 mg Hg m⁻² in 0-30 cm and 0.8-1.3 mg Hg m⁻² in 309 310 0-100 cm layer in the permafrost-free zone (Plotnikovo and Mukhrino sites); 0.5 mg Hg m⁻² in 0-30 cm and 3.0 mg Hg m⁻² in 0-100 cm layer in the sporadic zone (Kogalym site); 1.8-4.0 mg Hg 311 312 m^{-2} in 0-30 cm and 9.6-11.9 mg Hg m^{-2} in 0-100 cm layer in the continuous to discontinuous permafrost zone (Khanymey and Pangody sites), and 6.0 mg Hg m⁻² in 0-30 cm and 13.7 mg Hg 313 314 m^{-2} in 0-100 cm layer in continuous permafrost zone (Tazovsky). It is worth noting that the recent 315 data of Talbot et al. (2017) for Ontario (Canada) bogs (Sarea = 1,133,990 km²; 18.8 Gg Hg for the 316 277 ± 123 cm depth) are consistent with the results of the present study in the WSL (9.3 Gg Hg 317 for 592,440 km² for the 280 ± 100 cm depth).

318 A revised value of the Hg pool in the northern soils was recently provided by Olson et al. 319 (2018) who combined measured R_{HgC} values for Alaskan tundra soils with literature data, and derived R_{HgC} of 0.12 µg g⁻¹ for 0-30 cm (organic) and 0.62 µg g⁻¹ for 30-100 cm (mineral) layers. 320 321 Olson et al. (2018) estimate northern permafrost soil Hg pools of 26 Gg (0-30 cm) and 158 Gg 322 (30-100 cm), which combined (184 Gg) is 4 times lower than the number of 755 Gg (0-100 cm) 323 by Schuster et al. (2014). Both studies rely on R_{HgC} measurements from Alaskan soils, due to 324 relatively easy road access to the sampling sites along the Dalton Highway. Bedrock along the 325 Dalton Highway contains relatively high geogenic Hg levels (mean concentration: 32 ng/g), 326 resulting in a high geogenic contribution in mineral soils (39% for B horizons and 20% for A 327 horizons, Obrist et al., 2017). As a result, R_{HgC} in mineral soils along the Dalton highway are





328 higher (median = 1.6, Schuster et al. 2018) than for other mineral soils in North Amercia and 329 Eurasia (median = 0.64, Figure 8). It is clear that any upscaling calculation of pan-Arctic 330 permafrost Hg depends critically on the R_{HgC} of the 0-30 and 30-100 cm peat layers, as Eurasian 331 sporadic to continuous permafrost represents 54% of the northern soil C inventory (Table 2). 332 Compared to the previously assumed R_{HgC} of 1.6 (Schuster et al. 2018), and 0.12 - 0.62 $\mu g g^{-1}$ (Olson et al., 2018), we observe lower R_{HgC} ranging from 0.065 to 0.38 μ g g⁻¹ at 0-100 cm depth. 333 334 Setting Alaska aside as a geographic region, we find that North American and Eurasian mineral (<20% SOC) soil R_{HgC} was lower ($\mu = 0.77 \ \mu g \ g^{-1}$, median= 0.63 $\mu g \ g^{-1}$ (IQR = 0.32 to 0.80 $\mu g \ g^{-1}$) 335 ¹), n=131) than R_{HgC} reported for Alaska (median= 1.64 µg g⁻¹ (IQR = 0.91 to 2.93 µg g⁻¹), n=589) 336 337 (Figure 8). The R_{HgC} in organic soils (>20% SOC, including data from Alaska) was approximately 338 4 times lower (μ = 0.19 μ g g⁻¹, median= 0.15 μ g g⁻¹ (IQR = 0.09 to 0.24 μ g g⁻¹), n=449) than that 339 in mineral soils of North America and Eurasia (Figure 8), consistent with the observed difference 340 in WSL mineral and organic soils. Higher R_{HgC} observed in mineral soils may originate from a 341 contribution of geogenic Hg from the weathered bedrock (independent of C stock) and/or a higher 342 mineralization rate of C (preferential C over Hg loss) in predominantly oxic mineral soils 343 compared to anoxic peat soils.

344 In Table 2 we revisit the full 0-300 cm northern permafrost soil Hg inventory, based on 345 N-American (excluding Alaska) and Eurasian R_{HgC} based on the literature data compilations of 346 Olson et al. (2018) and Schuster et al. (2018), and our observed WSL R_{HgC} for Eurasia, multiplied 347 by estimated northern tundra soil organic C pools for 0-300 cm from and Hugelius et al. (2014). 348 The error made by neglecting high R_{HeC} in Alaskan mineral soils is small, on the order of 2.5 Gg 349 Hg, as estimated from the relatively small Alaskan C pool of 2.6 Pg C (Tarnocai et al., 2009). We 350 estimate the northern soil Hg pool to be 67 Gg (37-88 Gg, IQR) in the upper 30 cm, 225 Gg (102-351 320 Gg) in the upper 1 m, and 557 Gg (371-699 Gg) in the upper 3 m (Table 3). Note that our 352 revised values in the 0 - 1m range (225 Gg) is similar to that of Olson et al. (184 Gg), but lower





- than that of Schuster et al. (755 Gg). We find that Hg stocks in organic soils (>20% SOC) represent
 56% and 21% of the total Hg stock in the 0-30 cm and 0-100 cm depth range, respectively (Table
 2). The rest of the pan-arctic Hg is associated with C in mineral soils (<20% SOC) for which
 relatively sparse data exists (n=131). In particular, turbel and orthel mineral soils, which are
 estimated to contain 49 to 62% of total arctic C (Hugelius et al., 2014) and 36 to 85% of Hg at the
 various depth intervals need to be further investigated.
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4.3. Estimating the global soil (0-30 cm) Hg pool

361 To estimate the global soil Hg pool, we combined the more detailed Arctic pool estimate 362 (separating organic and mineral soils) with a more basic approach for the other climate zones, 363 where we derived bulk R_{HgC} for the 0-30 cm surface soils based on published literature data and 364 multiplied it with global C stock estimates for each climate zone (arctic, boreal, temperate, 365 subtropical and tropical) from the global soil organic carbon map (FAO and ITPS, 2018). The R_{HgC} increases from cold climate zones to warmer climate, from 0.15 $\mu g~\text{g}^{\text{-1}}$ for Arctic organic 366 367 soils, to 1.8 μ g g⁻¹ in subtropical and tropical soils (Figure 9, Table 4). This latitudinal trend in 368 R_{HgC} likely reflects a combination of low C mineralization rates in colder north and additional Hg 369 sorption to Fe(oxy)hidroxides in old tropical soils. Taking into account the variation in R_{HgC} and 370 C stocks across the climate zones, we estimate a global Hg stock of 1084 Gg (848 – 1258 Gg, 371 IQR) for the top 0 - 30 cm (Table 4). Previous global Hg soil pool estimates vary between 232 372 and 1150 Gg (Selin et al., 2008; Smith-Downey et al., 2010; Amos et al., 2013, 2015; Hararuk et 373 al., 2013; Wang et al., 2019). Schuster et al. (2018) concluded that Arctic permafrost soils store 374 nearly twice the amount of Hg as all other soils, the Ocean and atmosphere combined, but in doing 375 so they compared different global (0 - 30cm) and Arctic soil depth ranges (0 - 300cm). Our revised 376 estimate of the pan-Arctic permafrost and global soil pool suggests that, for a similar depth range 377 of 0-30 cm, permafrost soils contain 6% (67 Gg) of the global soil Hg pool (1084 Gg).





378 *4.4. Northern soil Hg sequestration and Hg loss*

379 Olson et al. (2018) recognized that the large 0-100 cm northern soil Hg pool is the result 380 of thousands of years of net atmospheric Hg deposition. The latitudinal trend of northward 381 increasing peat Hg concentration in the WSL (Fig. 4, 5) illustrates that this net Hg deposition is a fine balance between the vegetation Hg pump, which sequesters Hg⁰ in soils via foliar uptake and 382 litterfall, and Hg⁰ emission during biomass decay of vegetation debris. Annual gross Hg 383 384 sequestration by vegetation, via the vegetation pump, likely scales with primary productivity and 385 therefore decreases northward as insolation and growing season decrease. However, in the north, 386 degradation rates of vegetation biomass are lower than in the south: the moss biomass losses 387 during decomposition in the forest tundra zone (5-6% over 1st year and 10-12% over 2 years) is lower than that in the southern taiga (10-20% over 1st year and 20-40% over 2 years), based on in-388 389 situ biomass degradation experiments across the WSL gradient of biomes (Vishnyakova and 390 Mironycheva-Tokareva, 2018). The net result is a higher preservation of soil Hg in the north, 391 where less emission of Hg⁰ during plant decay occurs. The dependence of this balance, between Hg⁰ sequestration and Hg⁰ re-emission, on climate explains qualitatively the contrasting 392 393 observations made in Toolik (AK, USA, 68°N, MAAT = -7°C, Obrist et al., 2017) and Degerö Stormyr (Sweden, 64° N, MAAT = 2° C, Osterwalder et al., 2018). At Toolik, net Hg⁰ deposition 394 395 by vegetation and soil uptake occurs on an annual basis, whereas at Degerö Stormyr higher temperatures result in net annual Hg⁰ emission. More research is needed to quantify the climate 396 dependence of Hg⁰ sequestration (as soil Hg^{II}) and Hg⁰ re-emission before we can predict and 397 398 model northern soil Hg loss to the atmosphere due to global warming trajectories.

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404 **5. Conclusions**

405 Western Siberian peatlands contain a large amount of Hg in frozen and thawed peat; the lateral pools of peat palsa bogs range from 1-2 mg Hg m⁻² in the south to 10-15 mg Hg m⁻² in the 406 407 north. This northward increase of Hg concentration and pools can be explained by better 408 preservation of organic-bound Hg due to colder temperatures and shorter active period in the 409 continuous permafrost zone compared to the discontinuous and sporadic zones. We revisited the 410 full 0-300 cm northern permafrost soil Hg inventory, based on published R_{HgC} and our observed 411 WSL R_{HgC} for Eurasia, together with estimated northern tundra soil organic C pools for 0-300 cm 412 from Hugelius et al. (2014). We estimate the 0-300 cm northern permafrost soil Hg inventory to 413 be 557 Gg (371-699 Gg, IQR), which is three times lower than a previous estimate of 1656 ± 962 414 Gg Hg for the same depth range (Schuster et al., 2018). We estimate the global soil Hg pool to be 415 1084 Gg for the 0-30cm depth range. The permafrost Hg pool for the same 0 - 30cm depth range 416 is 67 Gg, and while large compared to the 3 Gg of Hg residing in the Arctic Ocean (Soerensen et 417 al., 2016), it represents only 6% of the global soil Hg pool.

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419 **Data availability.** Hg and C concentration data of the WSL soil samples are available in the 420 supplement. The permafrost data from Schuster et al. 2018 and a global compilation of R_{HgC} data 421 is available as supplementary information

422 (https://agupubs.onlinelibrary.wiley.com/doi/full/10.1002/2017GL075571, last access: 6
423 December 2019). The data from the Olson et al. 2018 study is available from the corresponding
424 author upon request. The data from the tropical climate zone was compiled from original
425 publications of Almeida (2005); Almeida et al. (2005); Campbell et al. (2003); Melendez-Perez
426 et al. (2014).





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- 429 sampling. AGL and OSP did all laboratory analysis. MJ and JES did the northern soil and global
- 430 soil Hg pool calculations. All authors contributed to writing of the manuscript.
- 431

432 We declare no competing interests

433

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- 442
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444 **References**

Almeida, M. D., Lacerda, L. D., Bastos, W. R., and Herrmann, J. C.: Mercury loss from soils
following conversion from forest to pasture in Rondônia, Western Amazon, Brazil, Environ. Pollut.,
137(2), 179–186, doi:10.1016/j.envpol.2005.02.026, 2005.

Almeida, M.D.: Biogeoquímica de mercúrio na interface solo-atmosfera na Amazônia, Ph.D.
thesis, Univsidade Federal Fluminense, Niterói, Brazil, 221 pp., 2005.

Amos, H. M., Jacob, D. J., Kocman, D., Horowitz, H. M., Zhang, Y., Dutkiewicz, S., Horvat,
M., Corbitt, E. S., Krabbenhoft, D. P., and Sunderland, E. M.: Global biogeochemical implications
of mercury discharges from rivers and sediment burial, Environ. Sci. Technol., 48(16), 9514–9522,
doi:10.1021/es502134t, 2014.

Amos, H. M., Sonke, J. E., Obrist, D., Robins, N., Hagan, N., Horowitz, H. M., Mason, R. P.,
Witt, M., Hedgecock, I. M., Corbitt, E. S., and Sunderland, E. M.: Observational and Modeling
Constraints on Global Anthropogenic Enrichment of Mercury, Environ. Sci. Technol., 49(7), 4036–
4047, doi:10.1021/es5058665, 2015.

Anisimov, O. A., and Sherstiukov, A. B.: Evaluating the effect of climatic and environmental
 factors on permafrost in Russia, Earth's Cryosphere, XX (2), 78-86, 2016.

Anisimov, O., and Reneva, S.: Permafrost and changing climate: the Russian
perspective, AMBIO: A Journal of the Human Environment. 35(4), 169176, https://doi.org/10.1579/0044-7447(2006)35[169:PACCTR]2.0.CO;2, 2006.

Bailey, E. A., Gray, J. E., and Theodorakos, P. M.: Mercury in vegetation and soils at
abandoned mercury mines in southwestern Alaska, USA, Geochemistry Explor. Environ. Anal.,
2(3), 275–285, doi:10.1144/1467-787302-032, 2002.





466 Baptista-Salazar, C., Richard, J. H., Horf, M., Rejc, M., Gosar, M., and Biester, H.: Grain-467 size dependence of mercury speciation in river suspended matter, sediments and soils in a mercury 468 mining area at varying hydrological conditions, Appl. Geochemistry, 81, 132-142, 469 doi:10.1016/j.apgeochem.2017.04.006, 2017. 470 Bates, A. L., Spiker, E. C., and Holmes, C. W.: Speciation and isotopic composition of 471 sedimentary sulfur in the Everglades, Florida, USA, Chem. Geol., 146(3-4), 155-170, 472 doi:10.1016/S0009-2541(98)00008-4, 1998. 473 Bedritsky, A.I.: Global climate and soil cover Russia: estimation of risks and 474 environmental and econoical consequences of land degradation. Adaptive systems and 475 technologies of agriculture and forestry, Moscow, National report, 2018. (In Russian). 476 Brown, J., Ferrians Jr, O. J., Heginbottom, J. A., and Melnikov, E. S.: Circum-arctic map 477 of permafrost and ground ice conditions, National Snow and Ice Data Center, Digital media, 478 Boulder, CO 80309-0449 USA, 1998, revised February 2001. 479 Campbell, L. M., Hecky, R. E., Muggide, R., Dixon, D. G., and Ramlal, P. S.: Variation 480 and distribution of total mercury in water, sediment and soil from northern Lake Victoria, East 481 Africa, Biogeochemistry, 65(2), 195–211, doi:10.1023/A:1026058417584, 2003. 482 Dastoor, A. P., and Durnford, D. A.: Arctic Ocean: Is It a Sink or a Source of Atmospheric 483 Mercury?, Environ. Sci. Technol., 48(3), 1707–1717, doi:10.1021/es404473e, 2014. 484 Fahnestock, M. F., Bryce, J. G., McCalley, C. K., Montesdeoca, M., Bai, S., Li, Y., 485 Driscoll, C. T., Crill, P. M., Rich, V. I., and Varner, R. K.: Mercury reallocation in thawing 486 subarctic peatlands, Geochemical Perspect. Lett., 6, doi:10.7185/geochemlet.1922, 2019. 487 FAO and ITPS: Global Soil Organic Carbon Map (GSOCmap), Technical Report, Rome, 488 162 pp. 2018. 489 Fisher, J. A., Jacob, D. J., Soerensen, A. L., Amos, H. M., Steffen, A., and Sunderland, E. 490 M.: Riverine source of Arctic Ocean mercury inferred from atmospheric observations, Nat. 491 Geosci., 5(7), 499–504, doi:10.1038/ngeo1478, 2012. 492 Frey, K. E., McClelland, J. W., Holmes, R. M., and Smith, L. G.: Impacts of climate 493 warming and permafrost thaw on the riverine transport of nitrogen and phosphorus to the Kara 494 Sea, J. Geophys. Res. Biogeosciences, 112(4), doi:10.1029/2006JG000369, 2007.a 495 Frey, K. E., Siegel, D. I., and Smith, L. C.: Geochemistry of west Siberian streams and 496 potential response to permafrost degradation, Water Resour. Res., 43(3), their 497 doi:10.1029/2006WR004902, 2007.b 498 Golovatskaya, E. A., and Lyapina, E. E.: Distribution of total mercury in peat soil profiles 499 in West Siberia, Contemp. Probl. Ecol., 2(2), 156–161, doi:10.1134/S199542550902012X, 2009. 500 Hararuk, O., Obrist, D., and Luo, Y.: Modelling the sensitivity of soil mercury storage to 501 climate-induced changes in soil carbon pools, Biogeosciences, 10(4), 2393–2407, doi:10.5194/bg-502 10-2393-2013, 2013. 503 Hugelius, G., Strauss, J., Zubrzycki, S., Harden, J. W., Schuur, E. A. G., Ping, C. L., 504 Schirrmeister, L., Grosse, G., Michaelson, G. J., Koven, C. D., O'Donnell, J. A., Elberling, B., 505 Mishra, U., Camill, P., Yu, Z., Palmtag, J., and Kuhry, P.: Estimated stocks of circumpolar 506 permafrost carbon with quantified uncertainty ranges and identified data gaps, Biogeosciences, 507 11(23), 6573-6593, doi:10.5194/bg-11-6573-2014, 2014. 508 Jensen, A., and Jensen, A.: Historical deposition rates of mercury in scandinavia estimated 509 by dating and measurement of mercury in cores of peat bogs, Water, Air, Soil Pollut., 56(1), 769-510 777, doi:10.1007/BF00342315, 1991. 511 Jiskra, M., Sonke, J. E., Agnan, Y., Helmig, D., and Obrist, D.: Insights from mercury 512 stable isotopes on terrestrial-atmosphere exchange of Hg(0) in the Arctic tundra, Biogeosciences, 513 16(20), 4051-4064, doi:10.5194/bg-16-4051-2019, 2019. 514 Jiskra, M., Sonke, J. E., Obrist, D., Bieser, J., Ebinghaus, R., Myhre, C. L., Pfaffhuber, K.

515 A., Wängberg, I., Kyllönen, K., Worthy, D., Martin, L. G., Labuschagne, C., Mkololo, T.,





Ramonet, M., Magand, O., and Dommergue, A.: A vegetation control on seasonal variations in
global atmospheric mercury concentrations, Nat. Geosci., 11(4), 244–250, doi:10.1038/s41561018-0078-8, 2018.

Kremenetski, K. V., Velichko, A. A., Borisova, O. K., MacDonald, G. M., Smith, L. C.,
Frey, K. E., and Orlova, L. A.: Peatlands of the Western Siberian lowlands: Current knowledge
on zonation, carbon content and Late Quaternary history, Quat. Sci. Rev., 22(5–7), 703–723,
doi:10.1016/S0277-3791(02)00196-8, 2003.

Kurina, I.V., and Veretennikova, E.E.: Impact of climate change of the Holocene on the
development of the ridge-hollow swamp complex of Western Siberia, Izvestiya Rossiiskoi
Akademii Nauk, Seriya Geograficheskaya, 2, 74-87, https://doi.org/10.15356/0373-2444-2015-274-87, 2015. (In Russian).

Kurina, I.V., Veretennikova, E.E., Il'ina, A.A., Dyukarev, E.A., Golovatskaya, E.A., and
Smirnov, S.V.: Reconstruction of conditions of formation of the eutrophic peatland deposits in
south of the taiga zone of Western Siberia, Izvestiya Rossiiskoi Akademii Nauk, Seriya
Geograficheskaya, 4, 66-76, https://doi.org/10.1134/S2587556618040106, 2018.

Lim, A. G., Sonke, J. E., Krickov, I. V., Manasypov, R. M., Loiko, S. V., and Pokrovsky,
O. S.: Enhanced particulate Hg export at the permafrost boundary, western Siberia, Environ.
Pollut., 254, doi:10.1016/j.envpol.2019.113083, 2019.

Loiko, S., Raudina, T., Lim, A., Kuzmina, D., Kulizhskiy, S., and Pokrovsky, O.:
Microtopography Controls of Carbon and Related Elements Distribution in the West Siberian
Frozen Bogs, Geosciences, 9(7), 291, doi:10.3390/geosciences9070291, 2019.

Lyapina, E. E., Golovatskaya, E. A., and Ippolitov, I. I.: Mercury concentration in natural
objects of west Siberia, Contemp. Probl. Ecol., 2(1), 1–5, doi:10.1134/S1995425509010019,
2009.

Melendez-Perez, J. J., Fostier, A. H., Carvalho, J. A., Windmöller, C. C., Santos, J. C., and
Carpi, A.: Soil and biomass mercury emissions during a prescribed fire in the Amazonian rain
forest, Atmos. Environ., 96, 415–422, doi:10.1016/j.atmosenv.2014.06.032, 2014.

543 Morel, F. M. M., Kraepiel, A. M. L., and Amyot, M.: The chemical cycle and 544 bioaccumulation of mercury, Annu. Rev. Ecol. Syst., 29, 543–566, 545 doi:10.1146/annurev.ecolsys.29.1.543, 1998.

Morgalev, Y. N., Lushchaeva, I. V., Morgaleva, T. G., Kolesnichenko, L. G., Loiko, S. V.,
Krickov, I. V., Lim, A., Raudina, T. V., Volkova, I. I., Shirokova, L. S., Morgalev, S. Y.,
Vorobyev, S. N., Kirpotin, S. N., and Pokrovsky, O. S.: Bacteria primarily metabolize at the active
layer/permafrost border in the peat core from a permafrost region in western Siberia, Polar Biol.,
40(8), 1645–1659, doi:10.1007/s00300-017-2088-1, 2017.

Nadyozhina, E.D., Shkolnik, I.M., Pavlova, T.V., Molkentin, E.K., and Semioshina, A.A.:
Permafrost response to the climate warming as simulated by the regional climate model of the
main geophysical observatory, Earth's Cryosphere, XII 3, 3–11, 2008. (In Russian).

Obrist, D., Agnan, Y., Jiskra, M., Olson, C. L., Colegrove, D. P., Hueber, J., Moore, C.
W., Sonke, J. E., and Helmig, D.: Tundra uptake of atmospheric elemental mercury drives Arctic
mercury pollution, Nature, 547(7662), 201–204, doi:10.1038/nature22997, 2017.

Olson, C., Jiskra, M., Biester, H., Chow, J., and Obrist, D.: Mercury in Active-Layer
Tundra Soils of Alaska: Concentrations, Pools, Origins, and Spatial Distribution, Global
Biogeochem. Cycles, 32(7), 1058–1073, doi:10.1029/2017GB005840, 2018.

Osterwalder, S., Sommar, J., Åkerblom, S., Jocher, G., Fritsche, J., Nilsson, M. B., Bishop,
K., and Alewell, C.: Comparative study of elemental mercury flux measurement techniques over
a Fennoscandian boreal peatland, Atmos. Environ., 172, 16–25,
doi:10.1016/j.atmosenv.2017.10.025, 2018.





Outridge, P. M., Macdonald, E. R. W., Wang, G. F., Stern, G. A., and Dastoor, A. P.: A
mass balance inventory of mercury in the Arctic Ocean, Environ. Chem, 5, 89–111,
doi:10.1071/EN08002, 2008.

Panova, N. K., Trofimova, S. S., Antipina, T. G., Zinoviev, E. V., Gilev, A. V., and
Erokhin, N. G.: Holocene dynamics of vegetation and ecological conditions in the southern Yamal
Peninsula according to the results of comprehensive analysis of a relict peat bog deposit, Russ. J.
Ecol., 41(1), 20–27, doi:10.1134/S1067413610010042, 2010.

Pavlov, A.V., and Malkova, G.V.: Mapping of trends of the contemporary ground
temperature changes in the Russian north, Earth's Cryosphere, XIII 4, 32–39, 2009. (In Russian).
Pearson, C., Howard, D., Moore, C., and Obrist, D.: Mercury and trace metal wet
deposition across five stations in Alaska: controlling factors, spatial patterns, and source regions,
Atmos. Chem. Phys., 19(10), 6913–6929, doi:10.5194/acp-19-6913-2019, 2019.

Peregon, A., Maksyutov, S., and Yamagata, Y.: An image-based inventory of the spatial
structure of West Siberian wetlands, Environ. Res. Lett., 4(4), doi:10.1088/17489326/4/4/045014, 2009.

Peregon, A., Maksyutov, S., Kosykh, N. P., and Mironycheva-Tokareva, N. P.: Map-based
inventory of wetland biomass and net primary production in western Siberia, J. Geophys. Res.
Biogeosciences, 113(1), doi:10.1029/2007JG000441, 2008.

Pokrovsky, O. S., Manasypov, R. M., Loiko, S. V., Krickov, I. A., Kopysov, S. G.,
Kolesnichenko, L. G., Vorobyev, S. N., and Kirpotin, S. N.: Trace element transport in western
Siberian rivers across a permafrost gradient, Biogeosciences, 13(6), 1877–1900, doi:10.5194/bg13-1877-2016, 2016.

Pokrovsky, O. S., Manasypov, R. M., Loiko, S., Shirokova, L. S., Krickov, I. A.,
Pokrovsky, B. G., Kolesnichenko, L. G., Kopysov, S. G., Zemtzov, V. A., Kulizhsky, S. P.,
Vorobyev, S. N., and Kirpotin, S. N.: Permafrost coverage, watershed area and season control of
dissolved carbon and major elements in western Siberian rivers, Biogeosciences, 12(21), 6301–
6320, doi:10.5194/bg-12-6301-2015, 2015.

Ponomareva, O.E., Gravis, A.G., and Berdnikov, N.M.: Contemporary dynamics of frost
mounds and flat peatlands in north taiga of West Siberia (on the example of Nadym site), Earth's
Cryosphere, 16, 21–30, 2012. (In Russian).

Preis, Y., and Karpenko, L.V.: Detailed reconstruction of bog functional state as a response
to continental climate changes in Holocene (the middle taiga of Western Siberia), Bulletin of the
Tomsk Polytechnic University, Geo Assets Engineering, 326 (2), 90-102, 2015. (In Russian).

597 Prietzel, J., Tyufekchieva, N., Eusterhues, K., Kögel-Knabner, I., Thieme, J., Paterson, D., 598 McNulty, I., de Jonge, M., Eichert, D., and Salomé, M.: Anoxic versus oxic sample pretreatment: 599 Effects on the speciation of sulfur and iron in well-aerated and wetland soils as assessed by X-ray 600 318-330, absorption near-edge spectroscopy (XANES), Geoderma, 153(3-4), doi:10.1016/j.geoderma.2009.08.015, 2009. 601

Raudina, T. V., Loiko, S. V., Lim, A. G., Krickov, I. V., Shirokova, L. S., Istigechev, G.
I., Kuzmina, D. M., Kulizhsky, S. P., Vorobyev, S. N., and Pokrovsky, O. S.: Dissolved organic
carbon and major and trace elements in peat porewater of sporadic, discontinuous, and continuous
permafrost zones of western Siberia, Biogeosciences, 14(14), 3561–3584, doi:10.5194/bg-143561-2017, 2017.

Romanovsky, V. E., Smith, S. L., and Christiansen, H. H.: Permafrost thermal state in the
 polar Northern Hemisphere during the international polar year 2007–2009: a synthesis, Permafrost
 periglac, 21(2), 106-116, https://doi.org/10.1002/ppp.683, 2010.

Romanovsky, V.E., Kholodov, A.L., Marchenko, S.S., Oberman, N.G., Drozdov, D.S.,
Malkova, G.V., Moskalenko, N.G., Vasiliev, A. A., Sergeev, D. O., and Zheleznyak, M. N.:
Thermal State and Fate of Permafrost in Russia: First Results of IPY, in: Proceedings of the 9th





International Conference on Permafrost, edited by: Kane, D.L., and Hinkel, K.M., University of
Alaska, Fairbanks, June 29 – July 3, 2008, vol. 2, 1511–1518, 2008.

Roulet, M., Lucotte, M., Canuel, R., Farella, N., Courcelles, M., Guimarães, J. R. D.,
Mergler, D., and Amorim, M.: Increase in mercury contamination recorded in lacustrine sediments
following deforestation in the central Amazon, Chem. Geol., 165(3–4), 243–266,
doi:10.1016/S0009-2541(99)00172-2, 2000.

Roulet, M., Lucotte, M., Saint-Aubin, A., Tran, S., Rhéault, I., Farella, N., De Jesus Da
Silva, E., Dezencourt, J., Sousa Passos, C. J., Santos Soares, G., Guimarães, J. R. D., Mergler, D.,
and Amorim, M.: The geochemistry of mercury in central Amazonian soils developed on the
Alter-do-Chao formation of the lower Tapajos River Valley, Para state, Brazil, Sci. Total Environ.,
223(1), 1–24, doi:10.1016/S0048-9697(98)00265-4, 1998.

Rudmin, M., Ruban, A., Savichev, O., Mazurov, A., Dauletova, A., and Savinova, O.:
Authigenic and detrital minerals in peat environment of vasyugan swamp, western Siberia,
Minerals, 8(11), doi:10.3390/min8110500, 2018.

Rydberg, J., Klaminder, J., Rosén, P., and Bindler, R.: Climate driven release of carbon
and mercury from permafrost mires increases mercury loading to sub-arctic lakes, Sci. Total
Environ., 408(20), 4778–4783, doi:10.1016/j.scitotenv.2010.06.056, 2010.

Schuster, P. F., Schaefer, K. M., Aiken, G. R., Antweiler, R. C., Dewild, J. F., Gryziec, J.
D., Gusmeroli, A., Hugelius, G., Jafarov, E., Krabbenhoft, D. P., Liu, L., Herman-Mercer, N., Mu,
C., Roth, D. A., Schaefer, T., Striegl, R. G., Wickland, K. P., and Zhang, T.: Permafrost Stores a
Globally Significant Amount of Mercury, Geophys. Res. Lett., 45(3), 1463–1471,
doi:10.1002/2017GL075571, 2018.

Selin, N. E., Jacob, D. J., Yantosca, R. M., Strode, S., Jaeglé, L., and Sunderland, E. M.:
Global 3-D land-ocean-atmosphere model for mercury: Present-day versus preindustrial cycles
and anthropogenic enrichment factors for deposition, Global Biogeochem. Cycles, 22(2),
doi:10.1029/2007GB003040, 2008.

Sheng, Y., Smith, L. C., MacDonald, G. M., Kremenetski, K. V., Frey, K. E., Velichko,
A. A., Lee, M., Beilman, D. W., and Dubinin, P.: A high-resolution GIS-based inventory of the
west Siberian peat carbon pool, Global Biogeochem. Cycles, 18(3), doi:10.1029/2003GB002190,
2004.

Shevchenko, V. P., Pokrovsky, O. S., Vorobyev, S. N., Krickov, I. V., Manasypov, R. M.,
Politova, N. V., Kopysov, S. G., Dara, O. M., Auda, Y., Shirokova, L. S., Kolesnichenko, L. G.,
Zemtsov, V. A., and Kirpotin, S. N.: Impact of snow deposition on major and trace element
concentrations and elementary fluxes in surface waters of the Western Siberian Lowland across a
1700'km latitudinal gradient, Hydrol. Earth Syst. Sci., 21(11), 5725–5746, doi:10.5194/hess-215725-2017, 2017.

Skyllberg, U., Qian, J., Frech, W., Xia, K., and Bleam, W. F.: Distribution of mercury,
methyl mercury and organic sulphur species in soil, soil solution and stream of a boreal forest
catchment, Biogeochemistry, 64(1), 53–76, doi:10.1023/A:1024904502633, 2003.

Smieja-Król, B., Fiałkiewicz-Kozieł, B., Sikorski, J., and Palowski, B.: Heavy metal
behaviour in peat - A mineralogical perspective, Sci. Total Environ., 408(23), 5924–5931,
doi:10.1016/j.scitotenv.2010.08.032, 2010.

Smith-Downey, N. V, Sunderland, E. M., and Jacob, D. J.: Anthropogenic impacts on
global storage and emissions of mercury from terrestrial soils: Insights from a new global model,
J. Geophys. Res., 115(G3), G03008, doi:10.1029/2009JG001124, 2010.

Soerensen, A. L., Jacob, D. J., Schartup, A. T., Fisher, J. A., Lehnherr, I., St Louis, V. L.,
Heimbürger, L. E., Sonke, J. E., Krabbenhoft, D. P., and Sunderland, E. M.: A mass budget for
mercury and methylmercury in the Arctic Ocean, Global Biogeochem. Cycles, 30(4), 560–575,
doi:10.1002/2015GB005280, 2016.





Sonke, J. E., Teisserenc, R., Heimbürger-Boavida, L. E., Petrova, M. V., Marusczak, N.,
Le Dantec, T., Chupakov, A. V., Li, C., Thackray, C. P., Sunderland, E. M., Tananaev, N., and
Pokrovsky, O. S.: Eurasian river spring flood observations support net Arctic Ocean mercury
export to the atmosphere and Atlantic Ocean, Proc. Natl. Acad. Sci. U. S. A., 115(50), E11586–
E11594, doi:10.1073/pnas.1811957115, 2018.

St. Pierre, K. A., St. Louis, V. L., Lehnherr, I., Gardner, A. S., Serbu, J. A., Mortimer, C.
A., Muir, D. C. G., Wiklund, J. A., Lemire, D., Szostek, L., and Talbot, C.: Drivers of Mercury
Cycling in the Rapidly Changing Glacierized Watershed of the High Arctic's Largest Lake by
Volume (Lake Hazen, Nunavut, Canada), Environ. Sci. Technol., 53(3), 1175–1185,
doi:10.1021/acs.est.8b05926, 2019.

Steffen, A., Douglas, T., Amyot, M., Ariya, P., Aspmo, K., Berg, T., Bottenheim, J.,
Brooks, S., Cobbett, F., Dastoor, A., Dommergue, A., Ebinghaus, R., Ferrari, C., Gardfeldt, K.,
Goodsite, M. E., Lean, D., Poulain, A. J., Scherz, C., Skov, H., Sommar, J., and Temme, C.: A
synthesis of atmospheric mercury depletion event chemistry in the atmosphere and snow, Atmos.
Chem. Phys., 8(6), 1445–1482, doi:10.5194/acp-8-1445-2008, 2008.

Steinmann, P., and Shotyk, W.: Chemical composition, pH, and redox state of sulfur and
iron in complete vertical porewater profiles from two Sphagnum peat bogs, Jura Mountains,
Switzerland, Geochim. Cosmochim. Acta, 61(6), 1143–1163, doi:10.1016/S00167037(96)00401-2, 1997.

Stepanova, V. A., Pokrovsky, O. S., Viers, J., Mironycheva-Tokareva, N. P., Kosykh, N.
P., and Vishnyakova, E. K.: Elemental composition of peat profiles in western Siberia: Effect of
the micro-landscape, latitude position and permafrost coverage, Appl. Geochemistry, 53, 53–70,
doi:10.1016/j.apgeochem.2014.12.004, 2015.

685 Stern, G. A., Macdonald, R. W., Outridge, P. M., Wilson, S., Chételat, J., Cole, A., 686 Hintelmann, H., Loseto, L. L., Steffen, A., Wang, F., and Zdanowicz, C.: How does climate arctic Total 687 change influence mercury?, Sci. Environ., 414, 22-42, 688 doi:10.1016/j.scitotenv.2011.10.039, 2012.

Talbot, J., Moore, T. R., Wang, M., Ouellet Dallaire, C., and Riley, J. L.: Distribution of
lead and mercury in Ontario peatlands, Environ. Pollut., 231, 890–898,
doi:10.1016/j.envpol.2017.08.095, 2017.

Tamocai, C., Canadell, J. G., Schuur, E. A. G., Kuhry, P., Mazhitova, G., and Zimov, S.:
Soil organic carbon pools in the northern circumpolar permafrost region, Global Biogeochem.
Cycles, 23(2), 1–11, doi:10.1029/2008GB003327, 2009.

Trofimova, I. E., and Balybina, A. S.: Classification of climates and climatic
regionalization of the West-Siberian plain, Geogr. Nat. Resour., 35(2), 114–122,
doi:10.1134/S1875372814020024, 2014.

Velichko, A. A., Timireva, S. N., Kremenetski, K. V., MacDonald, G. M., and Smith, L.
C.: West Siberian Plain as a late glacial desert, Quat. Int., 237(1–2), 45–53, doi:10.1016/j.quaint.2011.01.013, 2011.

Vishnyakova, E.K., and Mironycheva-Tokareva, N.P.: Moss decomposition in Western
Siberian mires, in: Mosses: Ecology, Life Cycle and Significance, edited by: Pokrovsky, O.,
Volkova, I., Kosykh, N., and Shevchenko, V., 4th ed. Nova Science Publishers Inc. New York,
pp. 217–241. 2018

Vorobyev, S. N., Pokrovsky, O. S., Serikova, S., Manasypov, R. M., Krickov, I. V.,
Shirokova, L. S., Lim, A., Kolesnichenko, L. G., Kirpotin, S. N., and Karlsson, J.: Permafrost
boundary shift in Western Siberia may not modify dissolved nutrient concentrations in rivers,
Water (Switzerland), 9(12), doi:10.3390/w9120985, 2017.

Wang, X., Yuan, W., Lin, C.-J., Zhang, L., Zhang, H., and Feng, X.: Climate and
Vegetation As Primary Drivers for Global Mercury Storage in Surface Soil, Environ. Sci.
Technol., 53(18), 10665–10675, doi:10.1021/acs.est.9b02386, 2019.





- 712 Wilhelm, R. C., Niederberger, T. D., Greer, C., and Whyte, L. G.: Microbial diversity of
- 713 active layer and permafrost in an acidic wetland from the Canadian high arctic, Can. J. Microbiol.,
- 714 57(4), 303–315, doi:10.1139/w11-004, 2011.
- 715 Zhang, Y., Jacob, D. J., Dutkiewicz, S., Amos, H. M., Long, M. S., and Sunderland, E. M.:
- 716 Biogeochemical drivers of the fate of riverine mercury discharged to the global and Arctic oceans,
- 717 Global Biogeochem. Cycles, 29(6), 854–864, doi:10.1002/2015GB005124, 2015.





719 **Table 1.** Mean (± SD) concentrations and stocks of C and Hg in 6 studied sites of WSL

720 peatbogs.

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Horizons	C, %	Hg, ng g ⁻¹	kg C m ⁻²	mg Hg m ⁻²	R_{HgC} (µg g ⁻¹)					
Plotnikovo (Pl), Southern taiga, 56.9°N*										
ALT (0-140 cm)	45±2	36±12	24	2.8	0.08±0.03					
Mineral (140-150 cm)	13	53	23	9.0	0.40					
Total (150 cm)	43±8	37±12	47	11.8	0.10±0.09					
0-30 cm	44±2	44±5	3	0.3	0.10±0.01					
0-100 cm	45±2	36±10	9	0.8	0.08±0.02					
Mukhrino (Mh), Middle taiga, 60.9°N*										
ALT (0-360 cm)	53±7	26±13	67	4.3	0.05±0.03					
Mineral (360-380 cm)	15±19	24±19	51	8.3	0.46 ± 0.46					
Total (380 cm)	51±10	26±19	118	12.6	0.07±0.12					
0-30 cm	50±0.4	29±7	4	0.3	0.06±0.02					
0-100 cm	52±5	32±17	18	1.3	0.06±0.04					
Kogalym (Kg), Northern taiga, 62.3°N*										
ALT (0-175 cm)	48±4	48±30	93	8.7	0.10±0.06					
Mineral (175-190 cm)	10±13	12±10	17	2.3	0.34±0.36					
Total (190 cm)	45±12	45±30	110	11	0.12±0.12					
0-30 cm	45±1	65±19	3	0.5	0.14±0.04					
0-100 cm	47±2	49±34	38	3 0.11±0.0						
	Khanyme	y (Kh), North	ern taiga, 63	.8°N						
ALT (0-34 cm)	44±2	64±43	17	2.1	0.15±0.10					
PF1 (34-100 cm)	50±2	47±13	78	7.6	0.09±0.02					
PF2 (34-138 cm)	48±6	47±11	119	11.8	0.10±0.02					
Mineral (138-147 cm)	1±1	4±1	2	0.5	0.31±0.13					
Total (147 cm)	42±16	47±28	138	14.4	0.13±0.09					
0-30 cm	43±1	71±46	13	1.8	0.17±0.11					
0-100 cm	47±4	54±29	95	9.6	0.12±0.07					
	Pangod	y (Pg), Forest	tundra, 65.9	°N						
ALT (0-40 cm)	50±4	78±25	38	5.3	0.16±0.07					
PF1 (40-100 cm)	53±4	61 ±25	54	6.6	0.11±0.04					
PF2 (40-155 cm)	48±10	67±23	78	11.0	0.15±0.07					
Mineral (155-185 cm)	3±1	24±11	15	12.5	0.88±0.28					
Total (185 cm)	41±19	62±28	130	28.8	0.27±0.30					
0-30 cm	50±5	83±26	26	4.0	0.17±0.07					
0-100 cm	52±4	68±25	92	11.9	0.13±0.06					
Tazovsky (Tz), Southern tundra, 67.4°N										
ALT (0-40 cm)	49±3	186±110	22	7.4	0.38±0.20					
PF1 (40-100 cm)	46±3	109±28	27	6.3	0.23±0.06					
PF2 (40-380 cm)	47±4	104±39	156	35.0	0.22±0.08					
Mineral (380-405 cm)	14±5	152±65	60	64.7	1.24±0.66					
Total (405 cm)	45±9	115±57	238	107.0	0.30±0.31					
0-30 cm	49±4	209±120	16	6.0	0.42±0.21					
0-100 cm	47±3	140±80	48	13.7	0.29±0.15					

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723 Footnote: ALT is Active Layer Thickness; PF1 is frozen peat, (ALT-100 см); PF2 is frozen peat

724 (ALT to mineral layer); 'Mineral' is mineral layer; 'Total' is total Hg content averaged over full

sampled depth. *In permafrost-free zone, the ALT extends from the surface to the mineral layer.

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- 729 Table 2. Estimated northern permafrost soil Hg inventory (Gg) for different depth ranges down
- to 300 cm. Hg pool uncertainties are reported as the interquartile range (IQR), i.e. the 25th to 75th
- 731 percentiles of the Hg pool distribution estimates by a Monte Carlo method. Soil organic carbon
- 732 (SOC) pools are from Hugelius et al. (2014).

depth range	soils	SOC	Hg Pool	IQR		Hg % of total per depth range
		Pg	Gg	G	ig	
0-30 cm	organic (>20% SOC)	172	32	12	42	48
	mineral (<20% SOC)	45	35	16	45	52
	total	217	67	37	88	
0-100 cm	organic (>20% SOC)	253	47	24	61	21
	mineral (<20% SOC)	219	178	58	271	79
	total	472	225	102	320	
0-200 cm	organic (>20% SOC)	366	68	42	86	16
	mineral (<20% SOC)	461	364	192	493	84
	total	827	433	257	564	
0-300 cm	organic (>20% SOC)	427	79	52	99	14
	mineral (<20% SOC)	607	477	292	621	86
	total	1035	557	371	699	

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735 **Table 3.** Comparison of estimated northern soil Hg pools by different studies.

			Schuster et al. 2018		Olson et al. 2018		this study	
depth range	SOC	95% CI	Hg Pool	95% CI	Hg Pool	25% CI	Hg Pool	IQR ¹
	Pg	Pg	Gg	Gg	Gg	Gg	Gg	Gg
0-30 cm	217	12	347	196	26	21-42 ¹	67	37-88
0-100 cm	472	27	755	427	184	115-232 ¹	225	102-320
0-200 cm	827	108	1323	764			433	257-564
0-300 cm	1035	150	1656	962			557	371-699

¹ Confidence interval (CI) corresponding to the 37.5th to 62.5th percentile

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Table 4. Estimated Hg pool for different climate zones based on reported R_{HgC} and carbon

pools. Hg pool uncertainties are reported as the interquartile range (IQR), i.e. the 25th to 75th

percentiles of the Hg pool distribution estimates by a Monte Carlo method.

Climate zone	C pool	median RHgC	mean RHgC	Hg pool	fraction	IQR		
	Pg	(µg/g)	(µg/g)	Gg	(%)	(Gg	
Tropics ¹	208	1.85	2.14	446	41.2	268	556	
Subtropics ¹	102	1.83	2.13	217	20.0	128	271	
Temperate ¹	191	1.35	1.55	297	27.4	186	367	
Boreal ¹	140	0.36	0.42	57	5.2	38	69	
Arctic ²	217 ³	(0.15, 0.64)	(0.19, 0.77)	67	6.2	37	88	
total	858			1085		848	1258	

¹Carbon pools are from FAO and ITPS (2018).

742 ² The arctic R_{HgC} and Hg pool are from Table 2.

³The arctic carbon pool is from Hugelius et al. (2014)







747 Fig. 1. Sampling sites and permafrost boundaries (modified after Brown et al., 2001) of WSL

- 748 territory investigated in this work. The climate and soil parameters of 6 sampling sites
- 749 (Tazovsky Tz, Pangody Pg, Khanymey Kh, Kogalym Kg, Mukhrino Mh, and Plotnikovo Pl) are
- 750 listed in Supplementary Table S1.
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Fig. 2. Vertical depth profile distribution of total Hg (THg) in 6 peat cores across a 1700 km
latitudinal transect of the WSL. Site location and physio-geographical parameters are shown in
Fig. 1 and Supplementary Table S1.

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Fig. 3. Mean (±SD), depth-integrated Hg concentrations in peat columns and mineral layers of 6
studied sites. Red asterisk represents the data from Lyapina et al. (2009).







Fig. 4. The ratio Hg:C (μg:g), median±IQR, in the active layer, frozen peat and mineral horizons

across the WSL latitudinal transect.







Fig. 5. Latitudinal variation in WSL soil Hg storage (mg Hg m⁻²) in the 0-30 and 0-100 cm peat layer (A) and in the active, frozen and mineral layers (B). In the permafrost-free zone, the first 40 cm were used to calculate Hg storage in the active layer. The permafrost peat layer is fixed from the lower boundary of the active (unfrozen) layer down to 100 cm. Finally, for the mineral layer we considered only the first 10 cm below peat deposits across the full latitudinal gradient of the WSL peatland.







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813 Fig. 6. Total, depth-integrated pools of Hg mass (Mg) in the upper 0-30 and 0-100 cm (red and

814 blue columns, respectively) of WSL frozen peatlands in each permafrost zone. The stocks are

815 calculated assuming the areal proportion of bogs from the landscape inventory across the WSL

816 (Sheng et al., 2004).



















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 $855 \qquad \mbox{Fig. 9. Histograms and median R_{HgC} for different global climate zones.}$