May 4, 2019

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Biogeosciences,

My coauthors and I would like to thank the two reviewers for their constructive comments and criticisms of our paper. They carefully read our paper and caught a number of things that our internal reviews missed, raise some good questions which we address and our manuscript is much improved as a result. Our responses to each of the points raised by the reviewer are given below in blue text to make it easily distinguishable from the comments. Additionally, we have included a "track changes" version with the submission of our revised document, as well as the clean new version. The line numbers in our responses below refer to the clean revised version. We hope that you will find our responses complete and that the paper is ready for publication.

Sincerely,

Jeff Drazen

Reviewer 1

Overall, I think this manuscript is very well-done and deserves to be published. The authors very carefully integrated previously-published and new data to give a comprehensive view of fish community recovery following experimental manganese nodule mining disturbance. The data are sound, and the manuscript is well-written. I have some minor and some medium comments, listed below.

Line 29: "relatively diverse" – in comparison to what? Is 16 taxa an average level of diversity for the abyssal Pacific?

Line 29 - Changed this to "The abyssal fish community included 16 taxa and was dominated by *Ipnops meadi.*"

Throughout: Please be consistent with English spelling. There are instances of "plow" and "plough" in the manuscript, so the authors should choose one and stick with it.

Corrected

Line 30: "Several years" is ambiguous, so please state the number of years.

Now specified as "at 6 months and 3 years"

Line 31: The authors state that fish density increased because of changes in regional environmental conditions, but they did not measure any environmental parameters in the present study. I expect this was a finding of the oft-cited Bluhm (2001) paper, but either way, it leaves the reader expecting environmental data throughout the paper. I would just remove the phrase "due to changes in regional environmental conditions" from the abstract.

This phrase now removed.

Line 31: This is incredibly nit-picky, but it is improper to begin a sentence with digits.

Converted to text instead of digits.

Line 33: I'm not convinced that a lower abundance of the dominant species in disturbed areas by itself means there was only partial fish community recovery. My experience with disturbance-recovery studies is largely based on benthic invertebrates, which have very distinct community-level changes over time post-disturbance. There is early dominance by opportunistic species followed by an increase in species richness and finally a return to the pre-disturbance state. Is there any similar framework for fish? If so, I would love to see the authors cite previous fish disturbance-recovery studies to back up their claim of partial recovery. This could go in the discussion but be mentioned in the abstract as well.

We respond to this comment below under line 310 – which reiterates this comment here.

Line 102: Please name the research vessel these cruises were on.

We have now added that they were on the RV Sonne

Line 115: The OFOS has been updated several times throughout the last 10 years, so please add in parentheses the manufacturer of the camera and lasers used. It would be also helpful to know which flash was used.

The camera is (iSiTEC, CANON EOS 5D Mark III) and the lasers (iSiTEC, custom built). These details are now added on lines 118-120.

Line 130: Please give a measure of the spread (i.e. standard deviation) of the photo area.

The standard deviation and interquartile range are now given on line 133-134.

Line 141: This transect-elimination method is confusing. It sounds like you assigned each photo a habitat type, then grouped all the photos from a single transect by habitat type, and if there weren't enough photos that you were likely to see at least one fish, you eliminated the transect-habitat type group? This paragraph could use some revision to be clearer.

The reviewer is correct in both their understanding of the paragraph and the recognition that the paragraph was poorly written! We have rewritten the paragraph and hopefully it makes much more sense now. We define "habitat transects" and our method for eliminating those that were too small to include statistically. Note this was also a comment for reviewer #2.

Line 162: Why were there so many variations in the baited camera specs?

We were not a part of the older cruises during which the baited cameras were deployed. As we understand it, the variation in times of deployment were simply due to scheduling around many other cruise activities. They targeted about 700 images per deployment (limits to film and batteries) so the image interval varied based on deployment time. The baiting protocol changed when they decided they wanted to capture some of the invertebrates (and it turns out a fish too!). Regardless of this variability the results are unique, the ISA is mandating scavenger studies in the CCZ and so these results are important to publish for general comparisons as we have made.

Line 163: Give a measure of spread (i.e. standard deviation) for the number of images per deployment.

This has been added and also in Table 1

Line 165: Time of first arrival is said to have been measured, but it is not presented in the text of the Results. What information does this metric give about each of the scavenging taxa?

We use the persistence values sparingly and now compare the time of first arrival values between the reference and disturbed areas (line 254-255, 299) finding that they are similar. Due to their sparing use we have moved them to a supplementary table.

Line 165 and throughout: The singular of "taxa" is "taxon."

Corrected throughout

Line 171: Why not use PERMANOVA to evaluate differences in community structure for the 1989 data as above for the 2015 data? PERMANOVA is more robust than ANOSIM.

We were comparing community data and initially thought that we didn't need the compare actual distances (PERMANOVA) but rather the rank order of differences (ANOSIM). Given the comments of both reviewers we have changed our statistical treatment to PERMANOVA. The result is unchanged. The test is described now on line 174-175 and the result has been amended on line 253.

Line 176: You should pick a consistent terminology for transect-habitat type groups. Define it in the paragraph at line 141 (along with a clear explanation of how some were eliminated), and use the term again here.

This sentence has now been corrected to be consistent with the language in the methods (original line 141). We also made some changes to the paragraph starting on line 197 to maintain consistent and clear terminology as requested.

Line 203: The first two sentences of this paragraph belong in the discussion

As written the reviewer is correct but this is also data that we analyze in this paper and want to include it in the results. Thus we have revised this paragraph so it reads clearly as results rather than discussion.

Line 212: A significant interaction means that change over time was only significant in one habitat type. This sentence should be revised.

Line 213 - This statement is not exactly correct. It means that the changes over time differed between the three habitat types in figure 5. We revised the sentence.

Line 212: Some information is missing from this sentence – which habitat had lower fish density?

Line 215 - The ploughed habitat. Revised to clearly specify this.

Line 218: Can Bluhm be contacted or original data accessed? 2001 was not that long ago.

Unfortunately Bluhm has left ocean science. He has not uploaded the data which was used for the 2001 paper. As part of the JPIO Oceans project the raw seafloor images collected during the cruise have been digitized in 2016. These images will be made available via https://portal.geomar.de/ in the near future.

Line 292: "Neighborhood" is an ambiguous measure of scale. Please replace it with something more concrete.

Range of attraction is baited camera studies is much debated hence our use of a general term. We don't want to indicate a specific or concrete value here but we have removed the offending word and made the important point that attraction of fish could have come from as far away as the reference area.

Line 310: The fish community at large appears to have recovered from the disturbance because there is similar density and community composition of fishes, so the only finding for "partial recovery" is the low density of I. meadi over disturbed tracks. Again, I'm not convinced that this one result is sufficient to justify the conclusion of "partial recovery" of the fish community. If anything, I think it indicates "partial" or lack of recovery in the benthic infaunal community, which is I. meadi's food source. The fish are there – they're just hanging out where the food is. I think the authors should add additional justification for their conclusion of "partial recovery," potentially framing their findings in the framework of other fish community disturbance-recovery studies.

We have evaluated this comment in some detail because these are semantics that really matter. Our use of the term "partial recovery" is a bit ambiguous and has likely led to the comment. Ipnops meadi is the numerically dominant fish in the system and it has not recolonized the tracks. We agree that they are found very close nearby and we don't think their population has declined regionally. It is our argument that their lack of occurrence in disturbed habitat is due to a lack of food. The dominant fishs' distribution is affected. So the fish community is still affected by the DISCOL experiment 26 years later, even if population abundance may be regionally unaffected. The importance of this finding is that the DEA is very small and a mosaic of disturbed and undisturbed habitat. At an industrial scale, mining will affect 100's km2 per year and potentially extirpate some of these fishes for decades even if they can move out of the way of mining vehicles. This will have population consequences due to the large spatial and temporal scale of effect. These consequences could extend beyond I. meadi. We have removed our term "partial recovery" from the abstract and discussion in relation to the fish community we measured. We also clarify the above point on lines 318-322 and 364-373, in the abstract and in the concluding paragraph.

We evaluated comparing our results with other disturbances of fish communities. There is a literature on trawling but in this case the fishes are targeted and removed so the comparisons are not very easy to make without delving into the subject beyond the scope of this paper.

Line 382: The authors should cite one or more CCZ studies to back up the statements in the previous 3 sentences.

These were cited in Table 3 which these sentences referred to. It is appropriate to also cite them in the text which we now do (line 397).

Section 4.3: What are the implications of the fish community overlap between the CCZ and DISCOL areas? Do the broad geographic ranges of fishes imply mined areas could be recolonized from elsewhere? Is it even known how far a given fish might swim to colonize a new area?

The first paragraph of section 4.3 discussed why such comparisons are important. Then we were purposely circumspect in this regard. Indeed, much of this section of the discussion talks about the fact that we can't speak that carefully about community overlap because often we are dealing with genera or higher taxa from photographic identification. But we understand the reviewer's point that readers will want some statement of conclusion. Thus, on line 403 we added the text "Rather it is likely that there are some species from this community, such as those that occur in both the DISCOL and CCZ regions, with broad distributions that could recolonize a mining license area if extirpated by mining. The extent of such conclusions must be made with caution because the overlap between the two areas may be artificially high."

Tables: consider adding a few horizontal lines to aid visualization of the data, for example between titles and data and between different table sections

Lines added as per suggestion

Table 1: It would be helpful if blank cells were filled with zeros so the reader could better keep track of the rows

Zeros added

Table 2: Persistence and time of first arrival data are not discussed at all in the manuscript. What is the importance of these metrics, and what information do they provide?

As indicated above we use these variables sparingly and have now moved them to a supplementary data table.

It would be helpful to have a table of PERMANOVA results.

We have included the PERMANOVA results for the two factor (habitat and time) evaluation of fish density in a supplementary table and cited this on line 213. The PERMANOVA test between reference and disturbed areas from the baited cameras was univariate and not significant so we have not included those results.

Fig. 1: Add an inset giving the location of the study site in the eastern Pacific.

Figure 1 had now been updated as suggested.

Fig. 3: Use colors consistent with Fig. 1 for the different habitat types

Colors now match

Fig. 3, 4, 5: Add tick marks on the outsides of axes for better visualization and consistency with Fig. 7

Done.

Reviewer #2. Kathy Dunlop.

The manuscript from Drazen et al. examines data from towed camera transects and baited camera surveys to resolve communities of deep-sea fishes and mobile scavengers at the DISCOL site (Peru Basin). Abyssal fish and scavenger communities were compared to the communities from the Clarion Clipperton Zone and between disturbed plough and reference sites.

I think this is a nicely put together manuscript, which presents a lot of new and interesting data that provides an important contribution to the evaluation and management of the environmental impact of deep-sea mining. I am therefore happy to recommend that the manuscript be accepted with some minor revisions. My suggested revisions are detailed below.

Some general points are that there are several long running sentences that I think could be improved by separating. Also in the discussion I think it is important to discuss some of the limitations more clearer, i.e. site classification, identification of species in images (perhaps it was more difficult to identify species in the 1989 images compared to those more recently taken) and also the uneven sampling effort. One primary aspect I was uncertain about was whether it is possible to detect differences in mobile fauna between some of the habitat treatments that are quite close together.

We address these comments below under each more specific point.

Abstract

Line 22. Replace "were performed beginning" with "were began".

Sentence modified

Line 28. Please state which years that abyssal fish surveys were conducted. The abstract would benefit from a clearer distinct of the DISCOL data analysed and when it was collected if possible.

Date ranges have been added.

Line 33: I found that the reference to "regional environmental conditions" didn't fit as I could find no analysis of the data in relation to environmental parameters.

Same comment as reviewer 1 and this phrase has been eliminated.

Lines 39 – 42: Separate the final sentence.

Sentence separated

Introduction

Lines 72 – 74. Long sentence. Consider revising and mentioning loss of prey items.

Sentence separated.

Line 77 – 79. "constructing a biogeography" I found the meaning unclear. Consider revising sentence.

We tried to clarify this sentence as follows "…it is important to characterize the fish community in regions that will likely experience mining in the near future and to begin constructing species ranges and community biogeographies,…"

Line 85. Include in the DISCOL region

Added.

Lines 84 – 89. Consider separating.

We separated the last phrase referring to the CCZ.

Methods

Line 115. Can the details of the camera and the laser systems be briefly described.

This is mentioned in the comments made to the other reviewer.

Lines 118 – 125. It would seem more correct to me to classifiy the level of physical disturbance by the known position in relation to the ploughs and EBS (if possible) than being manually assessment in the images.

We disagree with this assessment. In most parts of the DEA the plough tracks are so dense that the accuracy of the acoustic positioning is not good enough to calculate the distance of the images to the plough and EBS. Visual confirmation was considered more reliable. With the images, it's clear whether we're looking at ploughs, EBS or undisturbed, or a combination of all. The disturbed regions are very distinct, but as very little is known about the plume generated during the disturbance, we cannot comment on how 'exposed' the locations just away from the tracks actually were. If currents were in a uniform direction, the seafloor areas downstream and upstream of the track would be differently exposed. Given the large number of tracks (78) across the DEA, probably all regions were exposed to suspended sediments to some degree. Thus we believe our categories and visual selection is the most efficient and appropriate.

Line 134. Missing open brackets

The brackets are not missing but we realize this parenthetical phrase was a bit long so we have reworeded the sentence and added one which was in parentheses before. It now reads "Images were manually annotated for fishes using a variety of published keys. For data on octopi see Purser et al., (2017) and for all invertebrates and benthic fauna see Marcon et al. submitted.

Lines 143 - 144. Repetition from earlier on how fish density was estimated.

This repeated sentence was deleted.

Lines 145 - 150. I wasn't clear on the information on the transect areas with few images. I was unsure how you can tell if an image is likely or unlikely to detect a fish. Sounds like these transects were not used in the analysis so maybe not necessary to discuss them in detail.

Reviewer 1 had a similar comment listed under line 141 and line 176 and we address this issue in those responses above.

Line 160. Please specify which type of bait was used.

A single fish of ~500-1000g was used. We do not have records of which species was utilized but from the images the bait fish was a Carangid or Lutjanid. This is now indicated on line 163.

Line 168 - 169. I would recommend using a clearer definition of the criteria on how it was determined whether species were included on eliminated in the analysis.

We had thought this was a clear definition – species that reside in the field of view are considered but those that crawl or drift through are not. To clarify we have provide some examples of species that were omitted from our annotations. These include medusa and holothurians, for instance.

Line 171: I was not clear to me why PERMANOVA was used for the first analysis but ANOSIM for the second.

See our response to this questions which reviewer 1 also raised.

Results

Line 176. 46 habitats sampled – consider re-wording as sounds like there is 46 different habitat types.

Reviewer 1 had a similar comment at the same place. See our response to reviewer 1's comment.

Table 1. C. leptolepis? (remove ?), unided fish (spell out in full)

We have not removed the "?" from C. leptolepis as it represents some uncertainty in the identification of this taxon. We did spell out "unidentified fish."

Table 3. Some improvement can be made on the spacing of the headers.

Adjusted.

Figure 6. I can recommend improvement be made on the presentation of some of the overexposed images in the plate.

These images are not of the best resolution or quality due to the age of the photos (taken in 1982 and 1992) and the digitization process. However, we have attempted to increase image quality by sharpening the images and adjusting contrast/brightness.

Discussion

Line 286. Our results, 26 years

Comma added

The difference in the effects on sedentary and more mobile bentho-pelagic fauna is touched upon in the discussion. Something that was highlighted to me was whether it is possible to detect differences in mobile fauna (especially with baited underwater camera surveys) between some of the habitat treatments that are close together (i.e undisturbed, EBS, Plough and Transition). I would think that some species could travel between the habitat types to reach the bait.

Indeed, we agree! On lines 297 we state "It seems likely that the scavengers were attracted from a larger area that could have included the proximate reference or undisturbed areas. This could occur even if these animals were not commonly residing in the disturbance area due to habitat or prey community alteration."

Observations of deep-sea fishes and mobile scavengers from the abyssal DISCOL experimental mining area

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Keywords – DISCOL; manganese nodule; teleost; scavenger; Macrouridae, Ipnopidae, Ophidiidae, Zoarcidae, crustacea

Abstract

Industrial interest in deep-sea mineral extraction began decades ago and today it is at an all-time high, accelerated by global demand for metals. Several seafloor ecosystem disturbance experiments ~~were performed beginning~~began in the 1970's, including the DISturbance and reCOLonization experiment (DISCOL) conducted in the Peru Basin in 1989. A large seafloor disturbance was created by repeatedly plough~~w~~ing the seafloor over an area of ~10.8 km^2. Though a number of studies in abyssal mining regions have evaluated megafaunal biodiversity and ecosystem responses, few have included quantitative and detailed data on fishes or scavengers despite their ecological importance as top predators. We used towed camera transects (1989-1996, 2015) and baited camera data (1989-1992) to evaluate the fish community at the DISCOL site. The abyssal fish community ~~was relatively diverse with~~included 16 taxa and was dominated by *Ipnops meadi*. Fish density was lower in ploughed habitat ~~during the several~~at 6 months and 3 years following disturbance but thereafter increased over time~~ in part due to~~

changes in regional environmental conditions. 26 Twenty-six years post disturbance there were no differences in overall total fish densities between reference and experimental areas, but the dominant fish, *I. meadi*, still exhibited much lower densities in ploughed habitat, likely avoiding these areas and suggesting that the fish community remains by disturbance affected after decades. At the scale of industrial mining, these results could translate to population level effects. only partial fish community recovery. The scavenging community was dominated by eelpouts (*Pachycara* spp), hermit crabs (*Probeebei mirabilis*) and shrimp. The large contribution of hermit crabs appears unique amongst abyssal scavenger studies worldwide. The abyssal fish community at DISCOL was similar to that in the more northerly Clarion Clipperton Zone, though some species have only been observed at DISCOL thus far. Also, further species level identifications are required to refine this assessment. Additional studies across the polymetallic nodule provinces of the Pacific are required to further evaluate the environmental drivers of fish density, , and diversity and species biogeographies. This information , which will be important for the development of appropriate management plans aimed at minimizing human impact from deep-sea mining.

1. Introduction

The world's oceans are becoming increasingly exploited for their resources, and anthropogenic effects now reach the farthest corners and depths of ocean ecosystems (Ramirez-Llodra et al., 2011). New uses of our oceans are emerging. Industrial interest in deep-sea mineral extraction is at an all-time high, accelerated by global demand for minerals such as cobalt, zinc, copper, nickel, and rare-earth elements, which are enriched in seamount crusts as well as manganese nodules and deposited at hydrothermal vents. Currently, the International Seabed Authority has granted 29 exploration contracts to companies to explore for metals and rare-earth minerals in areas totaling >1,200,000 km^2 of seafloor in the Pacific, Atlantic, and Indian Oceans (www.isa.org.jm). Though the current intensity of commercial interest combined with technological innovations will soon lead to exploitation, this idea has a long history. Thus several seafloor ecosystem disturbance experiments were performed beginning in the 1970's (reviewed in Jones et al., 2017).

One of these, the DISturbance and reCOLonization experiment (DISCOL) was conducted in the Peru Basin in 1989. A large experimental seafloor disturbance was created by repeatedly plough~~w~~ing the seafloor. Biological surveys were conducted prior to the disturbance and several times thereafter to monitor seafloor ecosystem recovery (Thiel et al., 2001). Studies of the site seven years after disturbance showed only partial recovery (Thiel et al., 2001;Bluhm, 2001). Similar studies carried out in the north Pacific have also given indications that seafloor communities have not recovered or only partially recovered in periods of 26-37 years following disturbance (Miljutin et al., 2011;Jones et al., 2017;Gollner et al., 2017). This is not surprising given low rates of recruitment and growth common in these ecosystems, and the removal of the hard substrate upon which a large portion of the fauna depends (Amon et al., 2016;Vanreusel et al., 2016;Purser et al., 2017).

Though a number of studies in abyssal mining regions have evaluated megafaunal biodiversity and ecosystem responses, few have included quantitative and detailed data on fishes or scavengers (Leitner et al., 2017). However, many fishes are top predators that can have important influences on communities and ecosystems (Estes et al., 2011;Drazen and Sutton, 2017). Though fishes are mobile and may not suffer immediate mortality from mining, they will be affected by the large sediment plumes created (Oebius et al., 2001) and by the loss of foraging habitat or prey resources. Thus, ~~so~~ they may suffer regionally from local mining activities.

Also, top predators can bioaccumulate metals and other contaminants (Chouvelon et al., 2012;Choy et al., 2009;Bonito et al., 2016) that may be released from the activities of mining. Thus, it is important to characterize the fish community in regions that will likely experience mining in the near future and to begin constructing ~~a~~ species ranges and community biogeographies~~y~~, so that scientists and managers can evaluate potential mining impacts and appropriately locate protected no-mining zones (Wedding et al., 2013).

In 2015 a survey was performed of the DISCOL area using photo and video transecting techniques in a similar manner to the historical surveys of the area conducted into the late 1990s. In addition, archived analogue baited camera images collected shortly after the 1989 disturbance (1989-1992) were digitized and analyzed for fishes and other mobile scavengers, some of which may avoid transecting vehicles (Trenkel et al., 2004;Colton and Swearer, 2010). Our goal was to a) describe the fish and scavenger community in the DISCOL region in detail for the first time, b) evaluate the fish community response to disturbance and potential recovery, and c) compare the fish and scavenger community to that observed to the north of the equator in the Clarion Clipperton Zone (CCZ). ~~where t~~The majority of abyssal mining exploration licenses have been thus far granted in the CCZ, and this is where initial pilot mining activities are likely to commence.

2. Methods

In 1989 a ~10.8 km^2 circular region of the Peru basin in the Pacific, the DISCOL experimental area (the DEA), was artificially ploughed, in an effort to simulate the effects of deep-sea mining (Thiel et al., 2001). The study site (7º 04.4’ S, 88º 27.60’ W) ranges in depth from 4120-4200 m. Sediments are fine grained clays overlain with heterogeneous cover of manganese nodules, sometimes in high density. The plough-harrow device was 8 m wide and when deployed, overturned the first 10-15cm of seafloor sediment, ploughing the nodules into the seafloor and removing this hard substrate from the sediment / water interface. The plough was towed in 78 radial transects through the disturbance area with ˜20% of the seafloor directly disturbed by the plough. The most central region of the DEA was the most highly disturbed area crosscut by the majority of plough tows (Fig. 1; Foell et al., 1992).

In 2015 the DISCOL site was revisited and sampled twice (RV Sonne cruises SO242-1 and 2). The initial cruise was conducted in the summer and primarily conducted detailed acoustic and image-based mapping of the plough tracks using Autonomous Underwater Vehicles and ship-based sensors. This initial cruise also towed an epibenthic sled (EBS) several times across the seafloor, removing the top 20 cm of seafloor in trenches of ~2m x 500 m. These sled deployments were conducted to more accurately simulate the upper sediment removal envisioned as a likely consequence of mining. The second of these cruises focused on the detailed photographic study of the historic and recent disturbances mapped during the first cruise.

For investigation of megafauna, including fishes, the Alfred Wegner Institute (AWI) OFOS LAUNCHER towed camera system was used to conduct photographic transects of the seafloor. The OFOS LAUNCHER is identical to the OFOBS system described in Purser et al. (2018), with the exceptions that the OFOS was not equipped with INS, side scan or forward facing sonar systems. OFOS was flown at a height of ~1.7m above the seafloor and used a 23 megapixel downward looking still camera (iSiTEC, CANON EOS 5D Mark III) to take images every 15 seconds, each of which also captured the laser points projected by a tri-laser (50 cm spacing) sizing device (iSiTEC, custom built). Ship speed was maintained at 0.2-0.4 knots.

Given the high heterogeneity of the seafloor area studied, each image was manually assessed to represent one of a range of disturbance categories. These were 1) ‘Reference’ areas, not directly within the target circle of seafloor ploughed in 1989 (DEA), 2) ‘Undisturbed’ areas within the central DEA circle, but not actually impacted by the plough harrow directly, 3) ‘Transition’ images, within which both the edge of a plough track was visible as well as surrounding seafloor, 4) ‘Ploughed’ images within which only ploughed seafloor was visible and 5) ‘EBS’ areas, disturbed a month prior to SO242-2 by the towed epibenthic sled deployed by SO242-1. These five disturbance categories represent increasing levels of physical disturbance.

Image area captured within each image was determined by measuring the spacing of the laser points in a subset of 3663 images using the PAPARA(ZZ)I software application (Marcon and Purser, 2017). The image area of all remaining images was calculated from the camera altitude (distance to seafloor) using a second order polynomial regression of the laser-based measurements. The average seafloor image area was 5.71 ± 3.44 m^2. (interquartile range 4.45-6.25 m^2). In some instances, the camera was manually triggered to capture images of fishes that would have been missed in between timed images, or to capture a fish at a more suitable angle for identification. Images were manually annotated for fishes (for octopi see Purser et al., (2017) and for all invertebrates and benthic fauna see Marcon et al. submitted) using a variety of published keys. ~~(~~F~~f~~or data on octopi see Purser et al., (2017) and for all invertebrates and benthic fauna see Marcon et al. submitted. ~~)~~Fish density was estimated by dividing the number of fish viewed in regular timed images by the area photographed. Manually triggered images were not included in density estimates as these would present a positive bias towards images with fish in them. Diversity was evaluated using rarefaction curves (on all images, timed and manually triggered, because this approach only requires positive occurrences) to enable comparisons between habitat types that were not sampled at the same intensity.

OFOS transects often crossed several habitat types, so for fish density estimates, the images from each transect were divided into habitat type ~~subsets~~transects.~~ Fish density was estimated for each of these by dividing the number of fish viewed in the regularly timed images by the area photographed.~~ For some habitat ~~categories~~transects, there were very few images collected~~ during a transect~~. ~~In this case, w~~We eliminated ~~all the subsets/samples~~ ~~these~~those habitat transects ~~if they~~which were so small that they were unlikely to contain ~~have seen~~ at least one fish. We based this assessment ~~based~~ on the mean density of all habitat transects (both large and small) ~~samples~~ of 30.6 fish ha^{-1}, translating to a threshold sample area of 330 m^2. If used in the analysis, these small habitat transects ~~image sets~~ would either bias the results towards zero estimates if no fish were present in the small image set, or towards incorrectly high estimates if a few fish happened to be present~~in the small set of collected images~~. Fish density was compared between habitat types using a permutational ANOVA on a Euclidean distance matrix to account for uneven sample sizes and non-normal data distribution.

Baited cameras are now a widely used tool to census marine fishes (Bailey et al., 2007) because they can attract often sparsely distributed animals to within the census view, including some that might avoid active camera survey tools. Thus, for fully describing diversity and species abundances within a regional fish assemblage, they are indispensable. However, in contrast to transect methods, they are more difficult to use for estimations of accurate animal densities (Priede and Merrett, 1998;Yeh and Drazen, 2011).

During the first post disturbance cruise in 1989 and three years later in 1992 (Sonne cruises 61 and 77), free fall baited cameras (freefall baited observing systems - FBOS) were deployed (Brandt et al., 2004). These utilized a Benthos 35mm survey camera and strobe. Bait (a single 500-1000g Carangid or Lutjanid) was attached to a rod or placed in a small clear plastic tube ~1m from the camera, resting on the seafloor. Oblique images of ~1.7m^2 of the seafloor were taken every 2 to 5.5 min for ~24 to 55 hours, averaging 725±43 images per deployment. Animals were counted in each image. Metrics extracted from the imagery include the maximum number of each taxon~~a~~ visible in any one image over the camera deployment (MaxN), the time of first arrival for each taxon~~a~~ (T_{arr}), and the proportion of images in which a taxon~~a~~ was present for a camera deployment (Yeh and Drazen, 2011;Linley et al., 2017;Leitner et al., 2017). Only species that were clearly attracted to the bait were enumerated. This eliminated species that were photographed as they were simply drifting or crawling through the field of view, such as medusa and holothurians. Further, many small amphipods were often present at the bait but could not be reliably counted and so are not included. Deployments in 1989 were made within both the reference and disturbance areas, and a PERMANOVA ~~n analysis of similarity test (ANOSIM)~~test was used to compare community compositions on a Bray-Curtis similarity matrix based on square root transformed MaxN data.

3. Results

3.1 Photographic transects

20 OFOS transects ~~samples~~ were performed resulting in 46 habitat transects ~~samples~~ (Fig. 1). From these a total of 16733 images were examined with 306 fishes observed in 300 images (Table 1). Fishes were represented by 14 taxa (not including the category “unidentified fishes”; Fig. 2). Several groups were distinct but could not be identified to species whereas others were only identifiable to genus or family. The most common species observed was the benthic *Ipnops* cf *meadi* representing 61% of the fish observations. The Ophidiids were the most speciose family observed with 6 operational taxonomic units (OTU), some of which were distinct but could not be identified conclusively.

Across the five different habitat types, sampling effort was very uneven. Within the full data set, images taken of reference area habitat and in unploughed habitat within the experimental area were most abundant (Table 1). Seafloor images showing the disturbed habitat types (transient, ploughed and epibenthic sled (EBS) tracks) were less numerous. For all the data combined, as well as for the unploughed habitat type alone, rarefaction curves suggested adequate sampling as an asymptote was beginning to be reached in both cases (Fig. 3). However, within the other habitat types, rarefaction curves suggested more sampling was required to fully capture the fish diversity. Thus, the use of estimated species richness was needed for diversity comparisons. Interestingly, the disturbed habitat types had higher rarified diversity (ES 26) than the reference area or neighboring unploughed habitat (Fig. 3).

Fish densities were highly variable. Across all ~~sample areas surveyed~~habitat transects, seafloor areas imaged ranged from 355 to 7798 m^2 and fish density ranged from 0 to 71.4 fish ha^{-1}. ~~Across all samples a~~Average fish density was 30.2 ± 18.2 fish ha^{-1} (Fig. 4). Across the habitat types, density did not vary significantly (PERMANOVA, p>0.05). The density of the most common fish, *I. meadi*, could also be estimated and ranged from 0 to 68 fish ha^{-1}, averaging 18.4 ± 17.5 fish ha^{-1} across all ~~samples~~ habitat types (Fig. 4). Its density was significantly lower in the ploughed habitat type compared to undisturbed and reference habitats. Only a single *I. meadi* was found in the EBS habitat type (Table 1), but this individual did not occur in a habitat ~~sample~~ transect of sufficient length for density estimation. *Ipnops meadi* density in the two ~~samples~~ habitat transects available for analysis was zero.

Since the time of initial disturbance fish density in the DEA changed a great deal.~~Our fish density estimates can be compared to those published in Bluhm (2001). Bluhm's time series of densities suggests that there were n~~No fish were observed 6 months post disturbance, then fish density increased at year 3 and had returned to pre-disturbance density levels after 7 years (Fig. 5; Bluhm 2001). At this time, ophiuroids, holothurians, fish and hermit crabs were observed in the plough tracks. We examined this data (1989-1996) and the 2015 data for the reference, ploughed and unploughed habitat types~~, in addition to those presented in Bluhm's original work~~ using a two factor PERMANOVA. Habitat type and time were significant predictors of fish density with lower fish densities in the ploughed habitat (p<0.01; Supplementary Table 1). Also, the differences in the densities of fish across the three habitat types changed significantly with time since the disturbance (habitat x time, p<0.05). Fish density in the ploughed habitat was significantly (p<0.05) lower than the other habitat types right after the disturbance, marginally lower at 6 months post disturbance (p=0.057), and significantly lower at 3 years post disturbance~~, and marginally lower at 6 months post disturbance (p=0.057)~~. At 7 years the undisturbed habitat type in the DEA had higher fish density than the reference area. At 26 years, as already mentioned, there was no difference between habitats. Fish densities were similar to levels found in the undisturbed habitats and the reference area at 3 years post disturbance but higher than other times (Fig. 5). It was not possible to evaluate the times series data for *I. meadi* as Bluhm (2001) did not publish species specific results.

3.2 Baited camera observations

Six baited camera deployments were conducted, 5 in 1989 and 1 in 1992 (Table 2). Six taxa of fishes were identified (Fig. 6). The most abundant (MaxN) taxon in the deployments was the eelpout *Pachycara nazca*. This species occurred in all 6 deployments, reached a MaxN of 9 in two of the deployments and on average was present in 55% of the images. Individuals of the rattail *Coryphaenoides* sp. were either *C. armatus* or *C. yaquinae*, or both were present but, we could not differentiate them in the photographs. This taxona was present in all of the deployments but was observed on average in only 2.1% of images, and MaxN was never more than 2. Several ophidiids and a synaphobranchid eel were also observed. Ophidiids were generally rare and seen infrequently though *Bassozetus nasus* did generally persist at the bait when observed.

The baited camera also attracted 9 taxa of invertebrates (Table 2). The small shrimp *Hymenopeneus nereus* was present in all of the deployments in relatively large numbers (average MaxN = 9), with up to 15 visible at one time and was present on average in 63% of the images. The hermit crab *Probeebei mirabilis*, was also observed in every deployment but in varying numbers (from 1 to 9) and in 29% of the images. Penaeid shrimp were also observed in every deployment and were the third most abundant and common scavenging species. Two species were identified, *Cerataspis monstrosus* (identified as *Plesiopeneus armatus* in earlier papers; Leitner et al 2017) and *Benthiscymus* sp. Frequently, these could not be distinguished as they differ in the shape of the antennal scale and rostrum which were not always clearly visible. Large Munnopsid isopods were seen in all but one deployment but did not remain in the field of view for long. Ophiuroids were not abundant or common, being observed in three deployments as single individuals, but they stayed in the field of view for a long time (high persistence values).

Two of the camera deployments in 1989 were made in the disturbance area 6 months post event. In one of these deployments there was no obvious sign of disturbance in the limited field of view. In the other, a plough harrow track was clearly visible (FBOS006; Table 2). Low numbers of the benthic eelpout, *P. nazca*, were observed during this deployment. This deployment also had the lowest numbers of the benthic shrimp, *H. nereus*. However, the community composition did not vary significantly between the 1989 deployments inside the disturbance experiment area~~ed~~ and reference areas (~~ANOSIM~~PERMANOVA, p>0.05). Further, the times of first arrival of the scavengers was variable between deployments and not consistently longer at the disturbance area except for *B. iris* (Supplementary Table 2).

Overall, the diversity observed with the small number of camera deployments was fairly uniform, as evident from the plateau reached in both rarefaction and species accumulation curves (Fig. 7). This was the case for all scavengers and for the fishes alone. The baited cameras observed fewer taxa of fishes compared to the photo transects (Table 1, 2). Many of the fishes observed in the photo transects included less mobile benthic species such as members of the Ipnopidae, Bathysauridae and numerous unidentified ophidiids. However, the baited camera deployments identified two fish species that were not observed in the photo transects, *Barathrites iris* and a Synaphobranchid eel, both mobile scavengers.

4. Discussion

4.1 A description of the fish and scavenging community and relationship to past DISCOL studies

We present some of the first detailed fish assemblage information for the abyssal eastern Pacific~~ where seafloor mining will likely occur~~. Earlier studies at the DISCOL site presented limited fish assemblage results from the first few years of the experiment and report finding 8 fish taxa with *Ipnops* sp. being the most abundant (Bluhm, 1994). All of the taxa that were observed in these initial investigations were also present in our 2015 survey results, with the exception of *Halosaurus* sp. Moreover, we observed 6 additional taxa in 2015, and together with analysis of the 1989-1992 baited camera deployments, we have observed a total of 16 taxa. Interestingly, the earlier camera transect surveys flew the camera system higher off the bottom (3-3.5m vs 1.7m) which is perhaps more appropriate for the survey of larger, mobile fishes. Advances in photographic identification of abyssal fishes across the Pacific and improvements in photographic quality have resulted in the greater detail in the present analysis.

The baited camera deployments provided additional information on the DISCOL fish community and also provided data on scavenging invertebrate fauna. Past taxonomic works have used trapped specimens to document the presence of the eelpouts *P. nasca* and *P. bulbiceps* (Anderson and Bluhm, 1997) and the ophidiid *B. iris* (specimen deposited at the Senckenberg Museum). The physical specimens provide some vouchers for taxa that were identified from photographs. Two taxonomic studies used the baited camera imagery to tentatively identify the ophidiid *Bassozetus nasus* (Nielsen and Merrett, 2000) and large Munnopsid isopods which were thought to belong to the genus *Paropsurus* (Brandt et al., 2004). Bluhm et al (1995) briefly states that *P. mirabilis* and ophiuroids were commonly seen in the baited camera photos, but these results were not given in any detail. We show the eelpouts, the shrimp *H. nereus*, and hermit crabs are indeed common and regular bait attending fauna at this site (see below for comparisons to other abyssal regions).

4.2 Evaluation of the fish community response to disturbance and potential recovery

Our results, 26 years post disturbance, when compared to earlier sampling, provide some insight into the recovery potential of the fish fauna. The striking result found by Bluhm (2001) was that no fishes were observed in the disturbance area within 6 months of the disturbance; however, we show the presence of fish and scavenging invertebrates at this time from baited camera deployments. Samples sizes were low, but the community seems comparable to that in the reference areas at the same time. It seems likely that the scavengers were attracted from ~~the~~ a larger area that could have included ~~neighborhood, some possibly from~~ the proximate reference or undisturbed areas as suggested by ~~.~~the similar arrival times of the scavengers in reference and disturbance deployments. This could occur even if these animals were not commonly residing in the disturbance area due to habitat or prey community alteration because the DISCOL experiment created a patchwork of habitats over ~~the~~a scale ~~of~~as small as 10's of meters (Fig. 1).

~~Only partial recovery of t~~The fish community remains affected by the DISCOL experiment even after ~~has occurred~~ 26 years ~~post disturbance~~. Total fish density in the ploughed habitat of the DEA increased over time and in relation to the reference and undisturbed habitat suggesting recovery. It should be noted that large interannual changes were evident at the reference site with fish densities peaking 3 years post disturbance and at high levels again at 26 years (Fig. 5). An increase in megafaunal density over the first 7 years of the experiment was already documented and hypothesized to be the result of increased phytodetrital food flux and growing populations regionally (Bluhm, 2001). Such variation in megafaunal abundance is a regular feature of abyssal communities (Kuhnz et al., 2014;Ruhl and Smith, 2004). Comparisons between habitats at a point in time can provide a more robust means to assess recovery after plough disturbance (Miljutin et al., 2011). We found no differences in total fish density between the disturbed and undisturbed habitats at 26 years. Further, diversity (ES 26) was slightly higher in the disturbed habitat areas, although with relatively small sample sizes. However, the most common fish *I. meadi*, that makes up more than half of all the fish observations, had only a third of the density in 26-year-old plough tracks compared to undisturbed and reference areas, and only one individual was seen in the fresh EBS tracks (Fig. 4). This fish is found in the undisturbed habitat which occurs in a patchwork with the disturbance tracks, so regional reductions in population density are unlikely. Rather it seems that *I. meadi* actively avoids ~~The avoidance of *I. meadi* over~~ the plough tracks, showing~~s~~ that even the mobile fish community remains affected ~~has not fully recovered from~~ by the disturbance after more than two decades. This species' response likely relates to its biology as a rather sedentary, small benthic fish that, based on limited data, feeds on polychaetes, small bivalves, and crustaceans (Nielsen, 1966;Crabtree et al., 1991). Its prey may not have recovered in the tracks (Jones et al., 2017;Borowski, 2001). Most of the other fishes observed are benthopelagic and when swimming across a habitat mosaic might as easily be seen over an old plough track as over other habitat. Even if benthopelagic species tend to favor undisturbed habitat, this would be difficult to see in ~~the~~ data at such a small scale. Our other benthic species include the lizardfish *B. mollis* which preys on mobile fishes and shrimps and *B. sewelli*, which is a larger member of the Ipnopidae, but was too infrequently observed to assess habitat preferences (Table 1).

Conclusions about fish community recovery over time must be taken with caution. With a sparsely distributed fauna and the high variability in density, there are limits on statistical power and thus our confidence. The earlier DISCOL surveys differed in methodology to the current surveys including average altitude of the camera above bottom, image quality, and attention to the fishes. Our diversity estimates may well be higher as a result. Density estimates could also be affected by these same factors. The most common fish in the surveys, *I. meadi*, is relatively small and despite reflective eyes (Fig. 2) may have been more visible in our 2015 surveys in closer proximity to the seafloor. The influence many of these parameters have had on abundance estimations of fauna in the DISCOL region has been investigated in detail for a region of the DEA which was surveyed several times during the initial 7-year period and again in 2015. In 2015, the OFOS was deployed at 1.7 and 4 m in this region, and additionally an AUV was flown at 5 m to image the same region of seafloor. The results from these comparative studies (Purser et al. submitted for this special issue) show the sensitivity of density and diversity indices in the DISCOL area to changes in flight height, illumination, and camera type. Larger megafauna, such as fish, were clearly visible in images collected from higher altitudes, therefore resulting in both higher diversity and abundance estimates for a given transect length than achieved with lower flying camera systems. Certainly, methodology plays a very important role in determining the accuracy of sampling strategies in this ecosystem for determination of these parameters.

Our results add to a growing body of literature that generally finds little or partial recovery of faunal communities, even decades after simulated mining disturbances. Epifaunal megafauna density was considerably lower in disturbance tracks made 20 and 37 years prior to re-survey during the OMCO experiment in the CCZ (Vanreusel et al., 2016). Meta-analyses of abyssal disturbance experiments in the CCZ suggest that recovery of density and diversity is faster in mobile than sedentary fauna (Gollner et al., 2017;Jones et al., 2017). For instance, the mobile holothurian community appears to have recovered from disturbance in terms of density and community composition at the DISCOL site after 26 years (Stratmann et al., 2018). Most holothurians are detrital deposit feeders and their food source settling from above may not be greatly affected by the plough disturbance, whereas some fishes, such as *I. meadi*, likely rely upon epifaunal and infaunal macrofauna for food. The meiofauna and macrofauna have not recovered completely after 26 years in the CCZ (Miljutin et al., 2011), or after 7 years at the DISCOL site (Borowski, 2001). Some of the variation in the recovery potential observed between studies is undoubtedly derived from the variation in disturbance type and intensity. The direct benthic scale of actual nodule mining activities is suggested to be from 300-600 km^2 y^{-1} for a single mining license (Oebius et al., 2001;Levin et al., 2016). While it may seem that a local disruption in *I. meadi*'s distribution is a mild fish community effect 26 years post disturbance, it must be kept in mind that the DISCOL experiment did not completely disturb the DEA and that the scale of this experiment is very small in relation to industrial scale deep-sea mining. We argue that at industrial scales of seafloor disturbance that *I. meadi* could exhibit major regional reductions in population density that would last for decades and that such effects could extend to other species as well. Fishes may avoid direct mining activities but experience long term habitat losses at spatial scales that seem very likely to result in regional population consequences. Plumes of sediment from collectors or from discharge of the ore dewatering plume (Rolinski et al., 2001) will greatly expand this area and magnitude of effect. Therefore, it seems unlikely that the small-scale disturbance experiments, such as DISCOL (~10.8 km^2), will be adequate for evaluating the potential effects of full scale nodule mining. Further, the physical disturbance made in all experimental studies to date have not been directly reminiscent of the impacts actual mining will make in terms of volumes of surface sediment removed or displaced, subsequent sediment compaction, or generation of the high resolution topographical changes associated with the ridges and troughs likely to result from tracked mining vehicle movement (Jones et al., 2017;Doya et al., 2017;Jones et al., 2018).

4.3 Comparison of the DISCOL fish and scavenger communities to those within the CCZ

Nodule mining is likely to affect very large areas of the seafloor over decades (Wedding et al., 2015). Mobile fishes and other scavengers likely have the greatest ability to migrate away from mining disturbances, but they may be affected regionally through the redistribution of prey resources and sublethal effects from toxic metals or sediment plumes. Consequently, the biogeographies of taxa, even mobile species, are an important input to spatial management approaches (Watling et al., 2013). The scale of species distributions will help determine where and how large reserve areas should be in order to protect species. Comparison of the present findings in the south Pacific to those in the CCZ polymetallic nodule province to the north, across the equatorial upwelling, provide some insight into the ranges of abyssal fishes and scavengers in this mining relevant region. Past studies frequently combined fish and scavenger taxa into larger functional groups such as megafauna (Jones et al., 2017), but some studies have presented lists of species, which are the focus of the comparison here.

A number of the fish taxa observed with camera transects in the CCZ (Pawson and Foell, 1983;Radziejewska and Stoyanova, 2000;Tilot, 2006;Amon et al., 2017) have also been identified in the DISCOL area suggesting large species distributions (Table 3). 10 of the 14 taxa in the DISCOL region are shared with the CCZ. Four taxa were identified from DISCOL that were not previously identified from the CCZ region, none of which were abundant. Four fishes were observed in the various CCZ studies but not at the DISCOL site. A number of abyssal species have pan-Pacific and even global distributions (Priede, 2017). However, we are not suggesting that there is only a single community of fishes and scavengers integrated over 1000's of kilometers. Rather it is likely that there are some species from this community, such as those that occur in both the DISCOL and CCZ regions, with broad distributions that could recolonize a mining license area if extirpated by mining. The extent of such conclusions must be made with caution because tThe overlap between the two areas may be artificially high. Unrealistic overlap could arise due to the difficulty in identifying species from photographs, particularly those taken from high altitudes, and hence the use of genera and higher taxonomic categories. Further there are some taxa which can easily be confused depending upon image quality. For instance in the DISCOL site we identified the ophidiid, *Porogadus* sp. which has a long whip like tail and narrow body similar to Halosaurs which have been observed in the CCZ (Amon et al., 2017) and in an earlier study at the DISCOL site (Bluhm, 1994). We suspect that with increasing camera resolution and better taxonomic experience, photographic data and its analysis will improve greatly. Also, taxa are much more easily identified in oblique imagery. For instance, Halosaurs have prominent high pectoral fins and a single short dorsal whereas *Porogadus* has a long low dorsal fin all of which are seen in oblique imagery. We suggest the use of both oblique and vertical cameras on the same platforms in future studies. There has been some suggestion that oblique imagery would also alleviate avoidance issues with mobile taxa, but in the one abyssal study that used both oblique and vertical cameras, greater fish density was found in the vertical imagery (Milligan et al., 2016). Finally, collecting physical specimens and genetic data would be a great complement to the camera-based approach. Trawling for fish samples in mining claim areas will be challenging due to the great depth and the abundance of nodules, which can break nets and greatly damage specimens. Baited traps are effective for some of the fauna (Leitner et al., 2017;Linley et al., 2016).

The scavenging communities exhibit some interesting differences to those described from the eastern CCZ region and other abyssal Pacific locations. The dominant DISCOL scavengers were the shrimp *H. nereus*, eelpouts *Pachycara* spp., and the hermit crab *P. mirabilis*. The presence of large numbers of hermit crabs at the DISCOL site has been noted in earlier transect studies (Bluhm, 2001), and their large contribution to the scavenging community seems unique amongst abyssal scavenger studies. The most similar finding was a few hermit crabs (*Sympagurus birkenroadi*, MaxN= 2) attending bait from 2000 – 3000m depths off Hawaii (Yeh and Drazen, 2009). The large numbers of *H. nereus* is similar to the community in the eastern CCZ (Leitner et al., 2017). However, the eastern CCZ fishes were dominated by *Coryphaenoides* spp., which were not abundant at the DISCOL site. Overall the DISCOL scavenging community appears more similar to that observed in the western CCZ, which hosted lower numbers of *Coryphaenoides* spp. and greater numbers of ophidiids and shrimp (Leitner et al., 2017). The differences from east to west in the CCZ have been postulated to be related to the lower surface productivity in the west. Indeed, more oligotrophic regions have been shown to shift the dominance of the scavenging fishes from Macrourids to Ophidiids (Linley et al., 2017;Fleury and Drazen, 2013). However, the average long-~~-~~term chlorophyll concentration at the DISCOL site estimated from the MODIS satellite (30x30km box from 2006-2016) is about 1.5 times higher (0.22 mg chl-a m^{-3}) than that reported by Leitner et al (2017) in the eastern CCZ. Whether the community differences observed between the DISCOL and CCZ regions are the result of variations in overlying productivity, species distributions, or other habitat factors cannot be discerned until a greater number of baited camera studies are conducted across the region.

In conclusion, the DISCOL site has a relatively diverse abyssal fish community dominated by *Ipnops meadi*. Fish density increased in the ploughed habitat type over time and became similar to undisturbed habitat types at 26 years post disturbance, but the density of *I. meadi* is still only a third of the undisturbed habitat types indicating ~~only partial recovery~~that the DISCOL experiment continues to affect ~~of~~ the fish fauna through altered distributions. At the temporal and spatial scales of industrial mining, changes in habitat availability could lead to population reductions even if fishes can avoid the direct activities of mining. The abyssal fish communities observed in the central eastern Pacific at DISCOL and the more northerly CCZ are similar with many shared taxa. However, further species level identifications are required which requires the collection of physical specimens through trawling or baited traps. The scavenging community in the DISCOL site is unique in the prevalence of the hermit crab, *P. mirabilis*, which does not appear in the CCZ in either camera transects or baited camera deployments. Not surprisingly, fishes and mobile scavengers appear generally to have large ranges but also large shifts in community composition across the CCZ (Leitner et al., 2017) and across the equator. As commercial mining of polymetallic nodule provinces rapidly progresses, with commercial field trials commencing in the Belgian and German claim areas of the CCZ in the first months of 2019, gaining a better understanding of these remote ecosystems is of paramount importance. Until key fauna, such as the various benthic fish species utilizing these habitats are better known, ensuring that appropriate management plans are developed to best minimize human impact during mining will be extremely problematic.

5. Author Contributions

JCD and ABL analyzed the data and wrote the manuscript. SM annotated the baited camera images and assembled the data. AP and YM designed and conducted the camera transect experiments, quantified image coverage, helped write the manuscript, and generated the map figure. JG digitized and archived the original baited camera images. All authors read and commented on the manuscript.

6. Competing interests

The authors declare that they have no conflict of interest.

7. Acknowledgements

We thank the many DISCOL participants past and present who worked diligently to collect data over a 26-year study. We thank Kathy Dunlop and an anonymous reviewer for their careful comments on a draft of this manuscript. The Moore foundation provided funding for JCD, SM, and ABL to participate in this study. The SO242 cruises and accompanying work was funded by the German Ministry of Education and Science BMBF (grant number 03F0707A-G) through the project Mining Impact of the Joint Programming Initiative Healthy and Productive Seas and Oceans (JPIO).

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Table 1. Numbers of photo transect (OFOS system) observations (all images/ timed images only) for fishes in the DISCOL area by habitat type 26 years after initial experiment. The percent of images with fishes are calculated from the timed images only.

			Habitat type				
OTU	**Family**	**total**	**reference**	**undisturbed**	**transition**	**ploughed**	**ebs**
Bathysaurus mollis	Bathysauridae	13/11	2/1	5/4	2	2	2
Bathytyphlops cf *sewelli*	Ipnopidae	5	0	3/3	0	2	0
Ipnops cf *meadi*	Ipnopidae	188/178	68/64	97/91	11	11	1
Liparidae	Liparidae	4/3	1	3/2	0	0	0
Coryphaenoides armatus/yaquinae	Macrouridae	6/5	0	3/3	3/2	0	0
Coryphaenoides leptolepis?	Macrouridae	1/0	0	1/0	0	0	0
Bassozetus cf *nasus*	Ophidiidae	6	2	1	2	1	
Bassozetus sp. B	Ophidiidae	2	0	1	1	0	0
Bathyonus caudalis	Ophidiidae	30/26	8	15/12	2	3/2	2
Leucicorus sp.	Ophidiidae	3/2	3/2	0	0	0	0
Ophidiid sp. 3	Ophidiidae	6	1	2	1	2	0
Ophidiidae unided	Ophidiidae	16/14	2	8/6	1	5	0
Porogadus sp.	Ophidiidae	11	4	3	3	1	0
Pachycara spp.	Zoarcidae	4/2	2/1	2/1	0	0	0
unidentified fish		11/10	4/3	4	0	2	1
	#fish	306/281	97/89	148/133	26/25	29/28	6
	# OTUs	14	10	13/12	9	8	3
	# images	16733	5964	7155	1209	2055	350
	# images with fish	300/275	97/89	145/130	23/22	29/28	6
	% images with fish	1.6%	1.5%	1.8%	1.8%	1.4%	1.7%

Table 2. Deployment MaxN~~, persistence (pers.) and T_{arr}~~ for each bait-attending species by camera deployment (FBOS system). DEA – DISCOL experimental area, *deployment filmed a plough harrow track.

Deployment	FBOS003	FBOS004	FBOS005	FBOS006*	FBOS007	FBOS013	average
Date	2/20/1989	3/3/1989	3/16/1989	3/21/1989	3/22/1989	2/16/1992	
Image interval (min)	5.5	3.5	5	2	3.5	3.5	
# images	729	791	681	683	718	734	724±43~~3~~
Latitude	7° 2.12' S	7° 1.97' S	7° 4.83' S	7° 4.53' S	7° 4.55' S	7° 4.72' S	
Longitude	88° 26.53' W	88° 28.57' W	88° 21.33' W	88° 26.25' W	88° 27.92' W	88° 27.63' W	
General location	Reference area	Reference area	Reference area	~~Disturbance area~~DEA	~~Disturbance area~~DEA	~~Disturbance area~~DEA	
Depth (m)	4057	4167	4076	4220	4159	4170	
Fishes							
Barathrites iris	1	1		1	1		1
Bassozetus cf *nasus*	2	1		1		1	1
Coryphaenoides armatus/yaquinae	1	2	2	1	1	1	1
Leucicorus sp			1				1
Pachycara spp.	9	3	5	3	9	4	6
Synaphobranchidae	1	1	2		1	1	1
Crustaceans							
Hymenopeneus nereus	8	10	8	5	9	15	9
Total Penaeid shrimp	3	4	2	3	2	3	3
Cerataspis monstrosus	1	1		1	1		1
Benthiscymus sp.	2	2	1	2	1	2	2
Munnidopsis sp.		2	1			1	1
Munnopsidae	2	1	2	1	1		1
Mysidae		1	2	1			1
Probeebei mirabilis	1	9	4	3	4	6	5
Other taxa							
Octopoda	1		1				1
Ophiuroidea	1				1	1	1

Table 3. Fish taxa occurrences from DISCOL and abyssal sites of the CCZ. * listed in Bluhm (1994), bc – observed by baited camera only, #only these taxa out of 17 are given in the original reference

Taxa	**Family**	This study	(Amon et al., 2017;Amon et al., 2016)	(Pawson and Foell, 1983)	(Radziejewska and Stoyanova, 2000)	(Tilot, 2006)#
Bathysaurus mollis	Bathysauridae	x	x	x		x
Halosauridae	Halosauridae	*	x			
Bathytyphlops sewelli	Ipnopidae	x				
Ipnops meadi	Ipnopidae	x	x	x	x	x
Liparidae	Liparidae	x				x
Coryphaenoides armatus/yaquinae	Macrouridae	x	x	x	x	x
Coryphaenoides leptolepis?	Macrouridae	x				
Barathrites iris	Ophidiidae	bc	bc			x
Bassozetus sp.	Ophidiidae	x	x	x		
Bassozetus sp. B (sp 4 in Amon et al 2017)	Ophidiidae	x	x			
Bathyonus caudalis (sp 5 in Amon et al 2017)	Ophidiidae	x	x			
Leucicorus sp.	Ophidiidae	x				
Ophidiid sp. 1	Ophidiidae		x			
Ophidiid sp. 2	Ophidiidae		bc			
Ophidiid sp. 3	Ophidiidae	x	x			
Ophidiidae	Ophidiidae	x		x		x
Porogadus sp.	Ophidiidae	x				
Typhlonus nasus	Ophidiidae			x		x
Histiobranchus bathybius	Synaphobranchidae		x			
Synaphobranchidae	Synaphobranchidae	bc				x
Pachycara spp.	Zoarcidae	x	x			
Zoarcidae	Zoarcidae		x	x		

Figure Captions

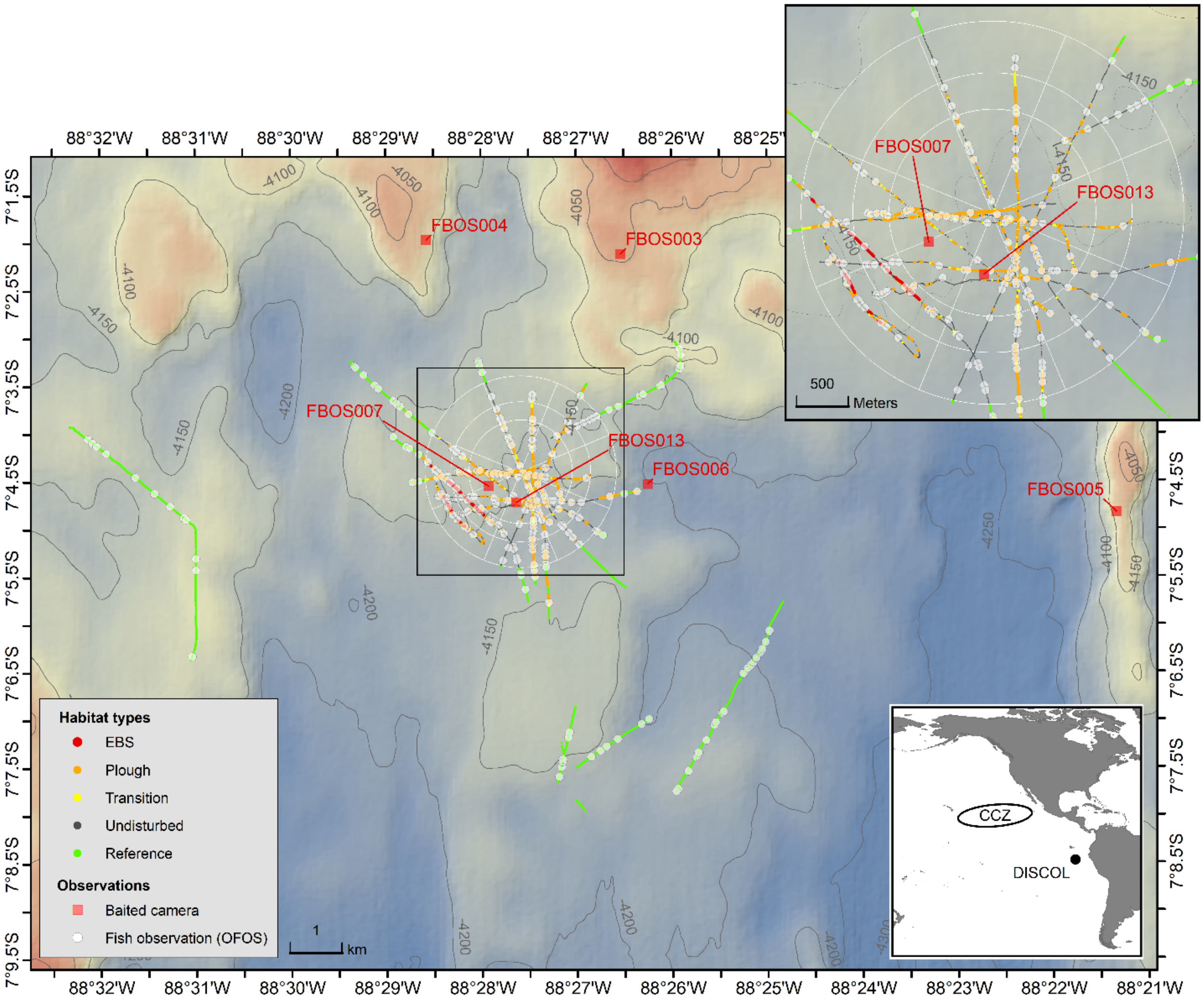

Figure 1. Map of the DISCOL study site showing the distribution of OFOS camera transects (colors indicate the 5 habitat types), the OFOS-based fish observations (white circles), and the location of the baited camera deployments (red squares). The white circular pattern and spokes shows the location and extent of the DEA.

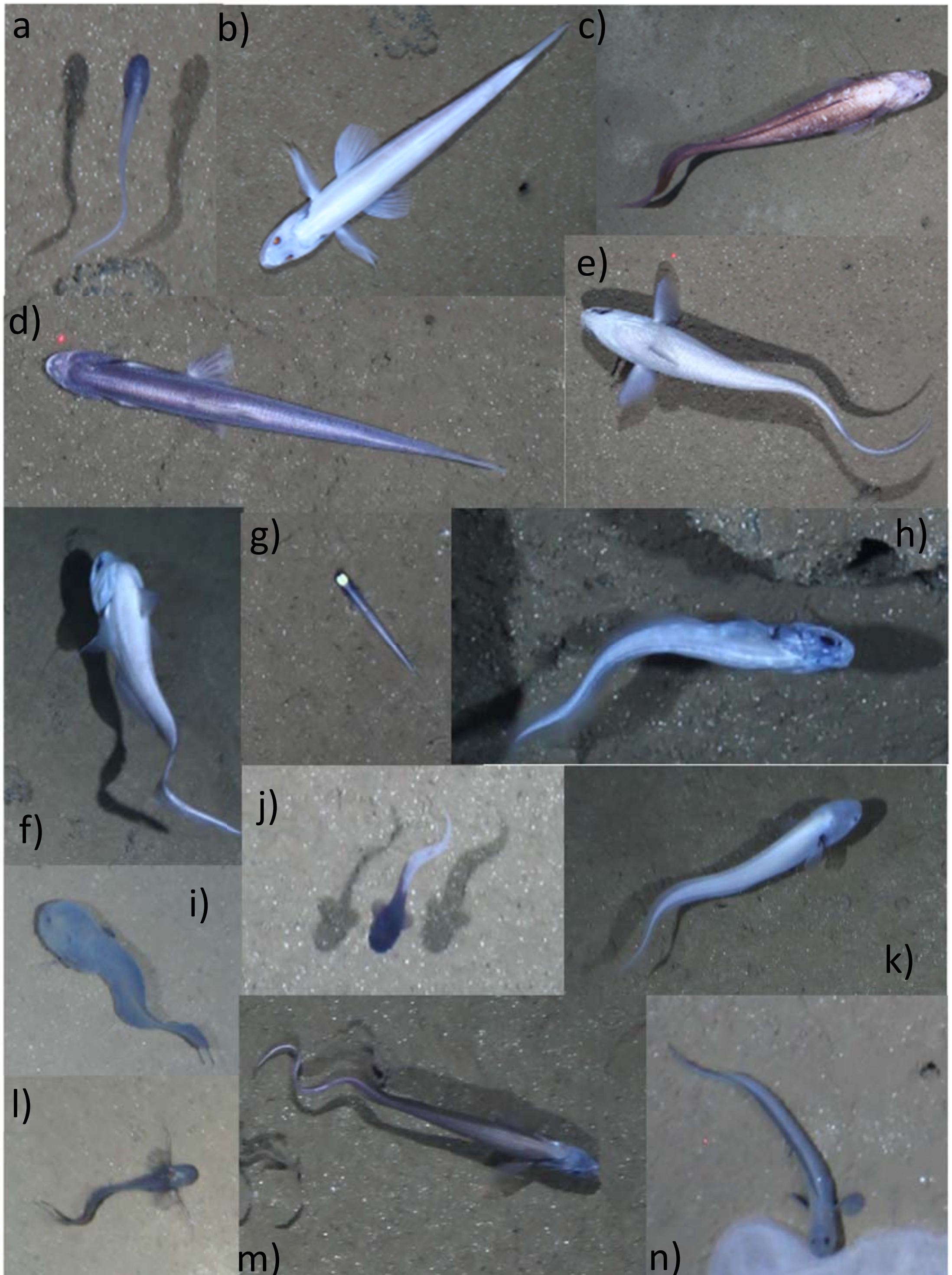

Figure 2. Representative images of OTUs identified in the DISCOL region during the 2015 survey. A) *Bassozetus* cf. *nasus* b) *Bathysaurus mollis* c) *Bathyonus* cf. *caudalis* d) *Bathytyphlops* cf. *sewelli* e) *Coryphaenoides armatus*/*yaquinae* f) *Coryphaenoides leptolepis* g)

Ipnops cf. *meadi* h) *Leucicorus* sp. i) Liparidae grey morphotype h) Liparidae bicolor morphotype k) *Bassozetus* sp. B l) Ophidiid sp. 3 m) *Porogadus* sp. n) *Pachycara* cf. *nazca*.

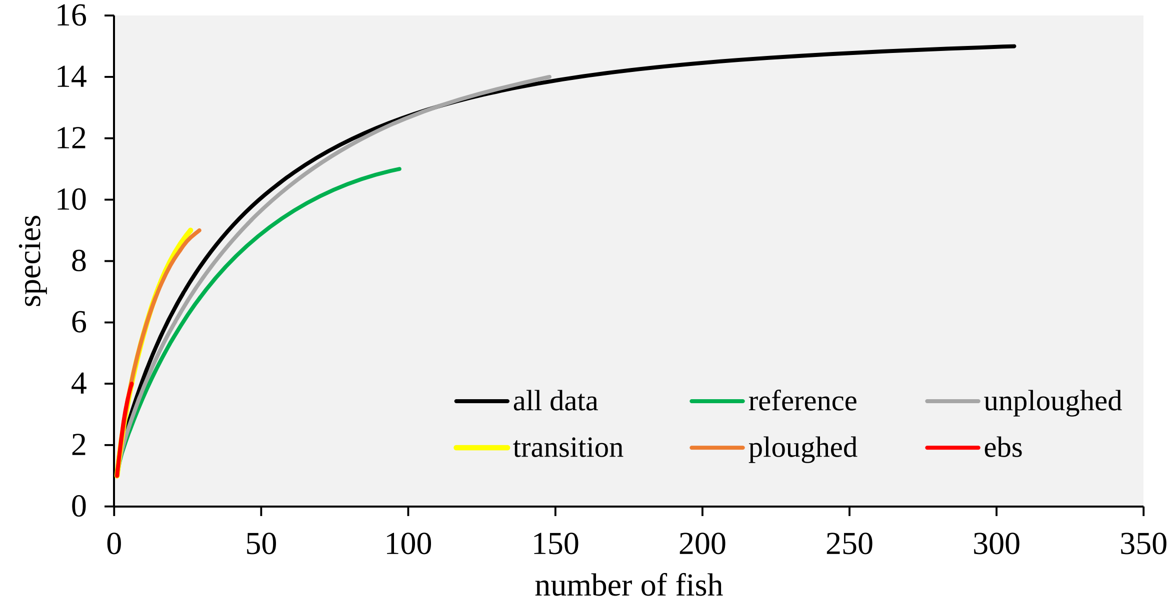

Figure 3. Rarefaction curves, estimated species richness as a function of the number of fish observations, for OFOS transects across habitat types.

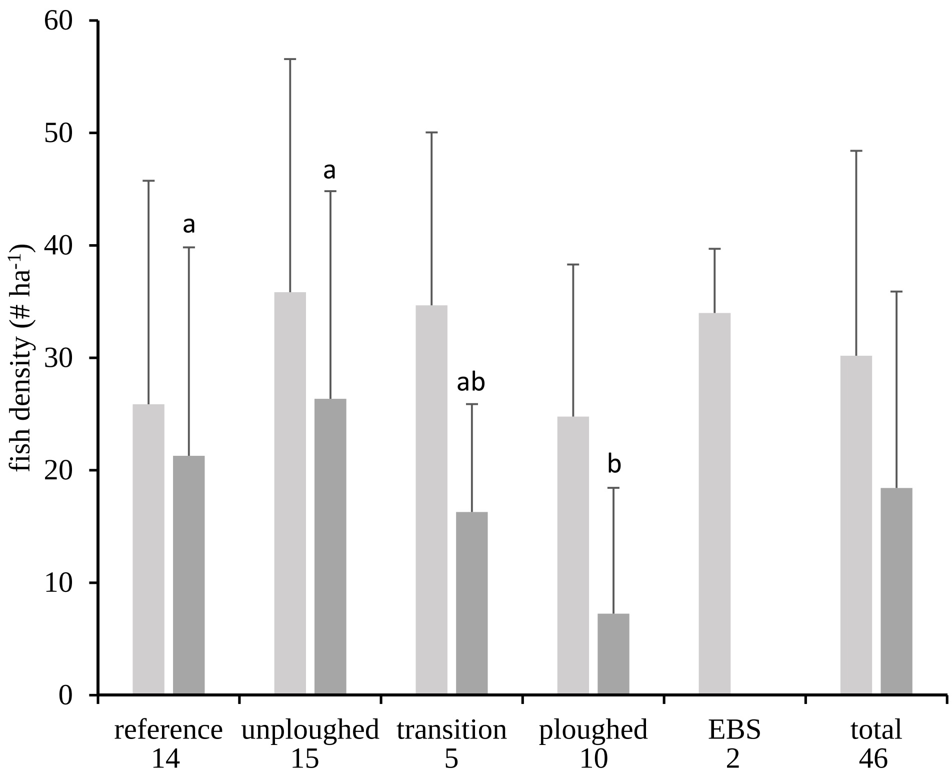

Figure 4. Total fish (light grey) and *I. meadi* (dark gray) density (mean and standard deviation) from the 2015 OFOS transects by habitat type (timed images only) and for the entire dataset.

The number of separate transects for each habitat type is given under its name. Letter symbols for each habitat indicate significant differences in *I. meadi* density ($p<0.05$).

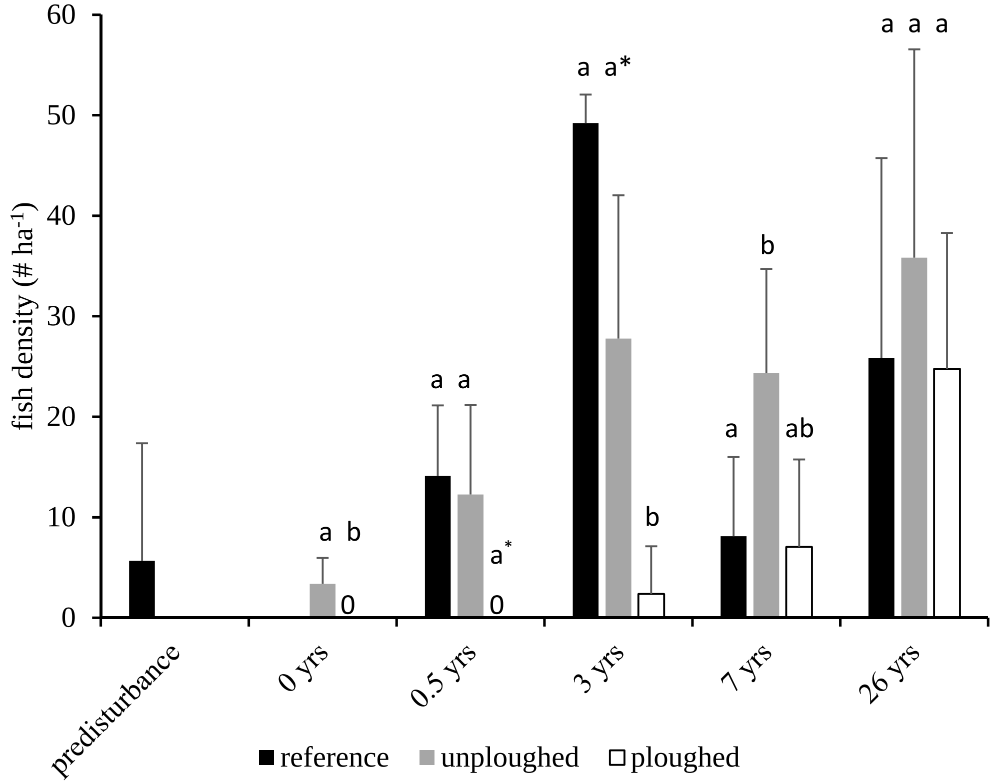

Figure 5. Fish density (mean and standard deviation) from predisturbance (1989) to 26 years post disturbance (2015) in the reference area and in the ploughed and unploughed habitats of the DEA. Data from predisturbance to 7 years post disturbance are from Bluhm (2001). Letter symbols for each time indicate significant differences between habitat types ($p<0.05$). At 0.5 yrs the asterisk indicates a marginal significant difference ($p = 0.057$).

705

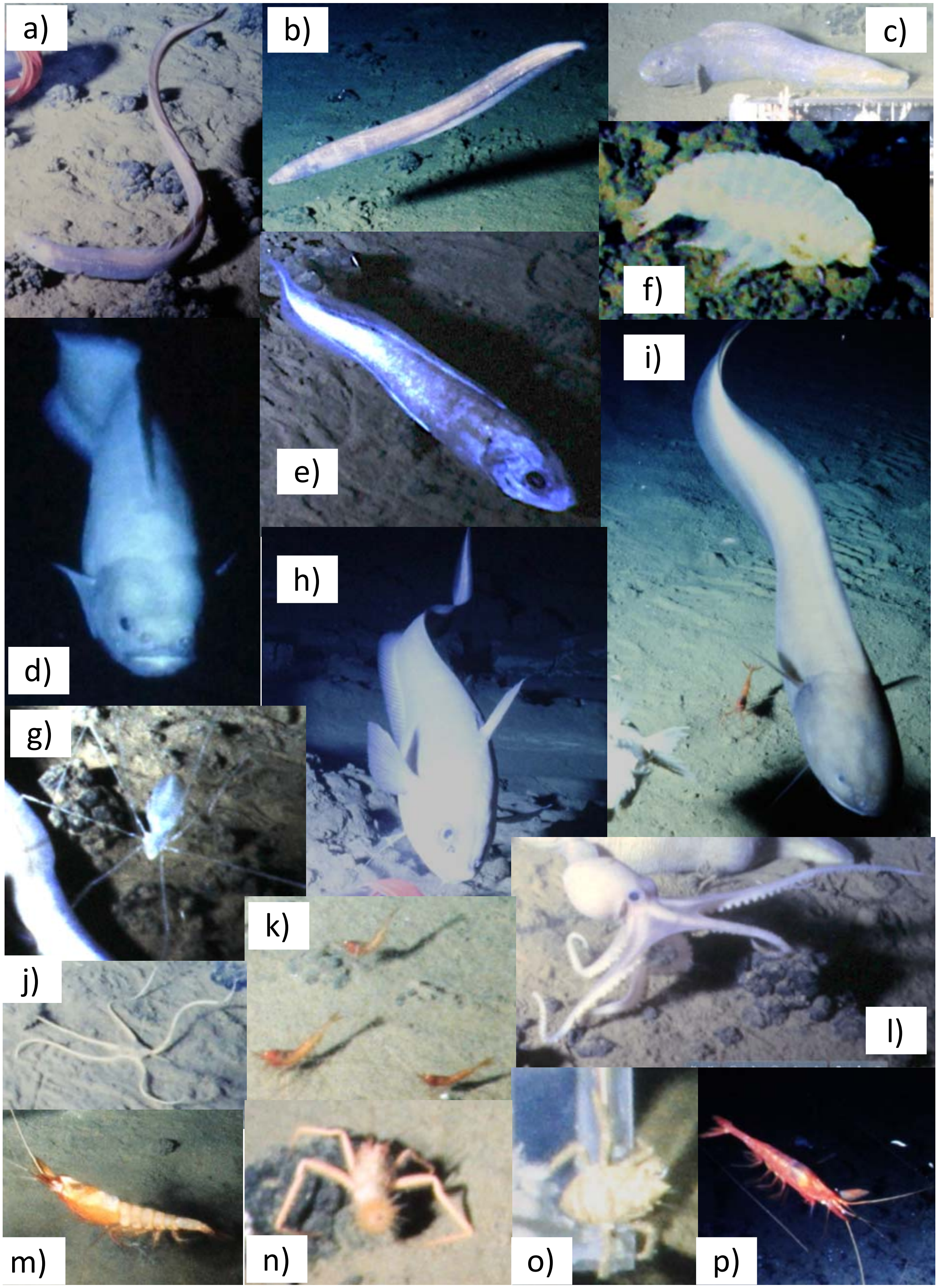

Figure 6. Representative images of OTUs identified using baited cameras in the DISCOL region. Photos taken in 1989 and 1992. A) *Illypohis* sp. B) Synaphobranchidae C) *Pachycara nazca* D) *Barathrites iris* E) *Leucicorus* sp. F) Large amphipod likely *Eurythenes* sp. G) Munnopsidae H) *Coryphaenoides* sp. I) *Bassozetus* c.f. *nasus* J) Ophiuroidea K) *Hymenopeneus nereus* L) Octopoda (*Vulcanoctopus* sp.) M) *Benthiscymus* sp. N) *Probeebei mirabilis* O) *Munnidopsis* sp P) *Cerataspis monstrosus*

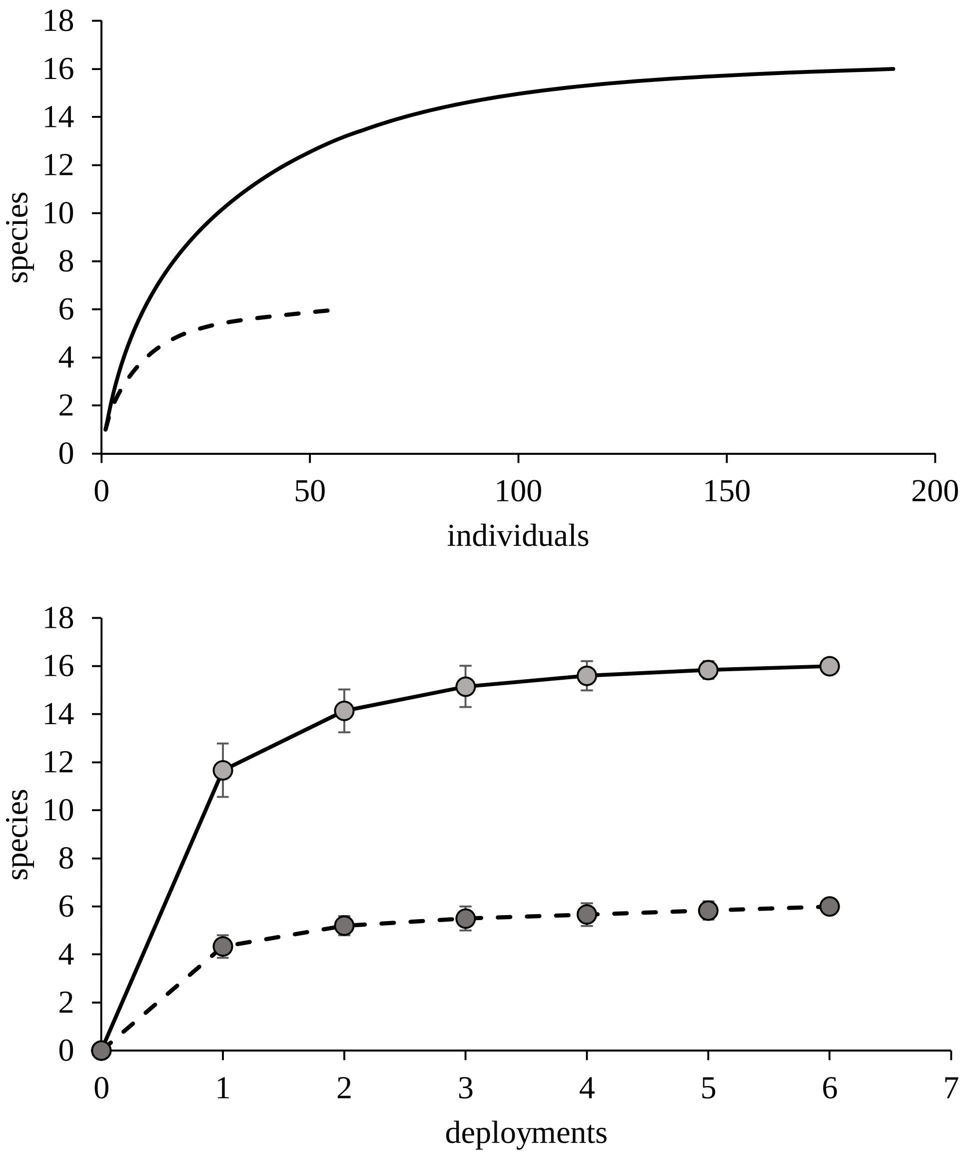

Figure 7. a) Rarefaction and b) species accumulation curves for baited camera observations. Solid lines represent all data and dashed lines are fishes only (both based on MaxN data).

Supplementary Table 1. Results of two way PERMANOVA test of fish density across time (0, 0.5, 3, 7 and 26 years) and habitat types (reference, unploughed and ploughed) and results for the pairwise tests of the interaction between habitat and time (tests between habitats within the factor time). Significant results are in bold text.

Source	df	SS	MS	Pseudo-F	p value	unique permutations
Habitat type	**2**	**2708**	**1354**	**6.3758**	**0.0031**	**9950**
Time	**5**	**9479**	**1896**	**8.9272**	**0.0001**	**9950**
Habitat x time	**7**	**3515**	**502.1**	**2.3645**	**0.028**	**9931**
Residual	72	15290	212.4			
Total	86	30970				

time (yrs)	Habitat type compared	t	p value	unique permutations
0	**Plough, Unploughed**	**2.5707**	**0.0473**	**16**
0.5	Plough, Reference	3.8252	0.1067	4
0.5	**Plough, Unploughed**	**2.3292**	**0.0576**	**15**
0.5	Reference, Unploughed	0.2528	0.7364	15
3	**Plough, Reference**	**14.998**	**0.0294**	**15**
3	**Plough, Unploughed**	**3.3832**	**0.0279**	**15**
3	**Reference, Unploughed**	**2.5093**	**0.0520**	**35**
7	Plough, Reference	0.19262	0.9034	31
7	**Plough, Unploughed**	**2.5538**	**0.0573**	**25**
7	**Reference, Unploughed**	**2.6771**	**0.0394**	**91**
26	Plough, Reference	0.15018	0.8839	9724
26	Plough, Unploughed	1.4837	0.1554	9803
26	Reference, Unploughed	1.3193	0.1910	9832

Supplementary Table 2. Time of first arrival (Tarr) and persistence (Pers.) or the proportion of images in which a taxon was observed for each baited camera deployment.

Deployment	FBOS003		FBOS004		FBOS005		FBOS006		FBOS007		FBOS013		average
	Pers.	T_{arr}	Pers.	T_{arr}	Pers.	T_{arr}	Pers.	T_{arr}	Pers.	T_{arr}	Pers.	T_{arr}	Pers.
Fishes													
Barathrites iris	0.4%	5:38	3.0%	4:54			0.6%	9:34	0.3%	15:38			1.1%
Bassozetus cf *nasus*	13%	39:20	4.9%	10:48			1.3%	18:16			21%	2:31	10%
Coryphaenoides armatus/yaquinae	1.0%	10:11	7.4%	1:49	2.3%	8:05	1.2%	12:40	0.1%	2:17	0.8%	2:52	2.1%
Leucicorus sp					0.4%	41:40							0.4%
Pachycara spp.	87%	2:07	21%	4:23	35%	11:55	32%	3:00	74%	5:53	80%	6:08	55%
Synaphobranchidae	2.5%	37:02	2.9%	3:37	11%	23:15			0.1%	17:23	7.4%	3:13	4.8%
Crustaceans													
Hymenopeneus nereus	65%	1:23	85%	1:07	40%	4:00	62%	0:54	39%	0:39	89%	0:21	63%
Total Penaeid shrimp	6.9%	0:11	20%	1:35	6.0%	5:55	21%	1:28	16%	0:18	24%	0:21	16%
Cerataspis monstrosus	0.7%	4:57	0.4%	15:38			1.0%	1:28	0.4%	40:01			0.6%
Benthiscymus sp.	2.2%	15:02	3.3%	2:17	0.7%	8:00	4.4%	6:16	4.0%	0:46	3.5%	2:17	3.0%
Munnidopsis sp.			11%	10:16	9.4%	35:10					7.2%	36:35	9.3%
Munnopsidae	4.4%	11:44	1.6%	39:02	4.4%	0:45	0.4%	12:12	1.0%	2:20			2.4%
Mysidae			0.4%	3:41	1.0%	10:55	0.3%	19:46					0.6%
Probeebei mirabilis	1.8%	0:22	32%	0:00	27%	2:20	12%	0:20	30%	0:11	69%	0:04	29%
Other taxa													
Octopoda	0.3%	10:55			2.3%	23:25							1.3%
Ophiuroidea	0.7%	21:55							47%	4:37	7.8%	1:31	19%