

1 Microbial functional signature in the atmospheric boundary layer

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12 Abstract

13
14 Microorganisms are ubiquitous in the atmosphere and some airborne microbial cells were
15 shown to be particularly resistant to atmospheric physical and chemical conditions (*e.g.*, UV
16 radiation, desiccation, presence of radicals). In addition to surviving, some cultivable
17 microorganisms of airborne origin were shown to be able to grow on atmospheric chemicals in
18 laboratory experiments. Metagenomic investigations have been used to identify specific
19 signatures of microbial functional potential in different ecosystems. We conducted a
20 comparative metagenomic study on the overall microbial functional potential and specific
21 metabolic and stress-related microbial functions of atmospheric microorganisms in order to
22 determine whether airborne microbial communities possess an atmosphere-specific functional
23 potential signature as compared to other ecosystems (*i.e.* soil, sediment, snow, feces, surface
24 seawater *etc.*). In absence of a specific atmospheric signature, the atmospheric samples
25 collected at nine sites around the world were similar to their underlying ecosystems. In addition,
26 atmospheric samples were characterized by a relatively high proportion of fungi. The higher
27 proportion of sequences annotated as genes involved in stress-related functions (*i.e.* functions
28 related to the response to desiccation, UV radiation, oxidative stress *etc.*) resulted in part from
29 the high concentrations of fungi that might resist and survive atmospheric physical stress better
30 than bacteria.

31
32 **Keywords:** atmospheric microorganisms, airborne microbial communities, planetary boundary
33 layer, metagenomic sequencing, comparative metagenomics, selective processes

34 1 Introduction

35
36 Microorganisms are ubiquitous in the atmosphere and reach concentrations of up to 10⁶
37 microbial cells per cubic meter of air (Tignat-Perrier et al., 2019). Due to their important roles
38 in public health and meteorological processes (Ariya et al., 2009; Aylor, 2003; Brown and
39 Hovmøller, 2002; Delort et al., 2010; Griffin, 2007), understanding how airborne microbial
40 communities are distributed over time and space is critical. While the concentration and
41 taxonomic diversity of airborne microbial communities in the planetary boundary layer have
42 recently been described (Els et al., 2019; Innocente et al., 2017; Tignat-Perrier et al., 2019), the
43 functional potential of airborne microbial communities remains unknown. Most studies have
44 focused on laboratory cultivation to identify possible metabolic functions of microbial strains
45 of atmospheric origin, mainly from cloud water (Amato et al., 2007; Ariya et al., 2002; Hill et
46 al., 2007; Väitilingom et al., 2010, 2013). Given that cultivatable organisms represent about 1
47 % of the entire microbial community (Vartoukian et al., 2010), culture-independent techniques
48 and especially metagenomic studies applied to atmospheric microbiology have the potential to
49 provide additional information on the selection and genetic adaptation of airborne

50 microorganisms. However, to our knowledge, only five metagenomic studies on airborne
51 microbial communities at one or two specific sites per study exist (Aalismail et al., 2019; Amato
52 et al., 2019; Cao et al., 2014; Gusareva et al., 2019; Yooseph et al., 2013). Metagenomic
53 investigations of complex microbial communities in many ecosystems (for example, soil,
54 seawater, lakes, feces, sludge) have provided evidence that microorganism functional
55 signatures reflect the abiotic conditions of their environment, with different relative abundances
56 of specific microbial functional classes (Delmont et al., 2011; Li et al., 2019; Tringe et al., 2005;
57 Xie et al., 2011). This observed correlation of microbial community functional potential and
58 the physical and chemical characteristics of their environments could have resulted from genetic
59 modifications (microbial adaptation) (Brune et al., 2000; Hindré et al., 2012; Rey et al., 2016;
60 Yooseph et al., 2010) and/or physical selection. The latter refers to the death of sensitive cells
61 and the survival of resistant or previously adapted cells. This physical selection can occur when
62 microorganisms are exposed to physiologically adverse conditions.

63 The presence of a specific microbial functional signature in the atmosphere has not been
64 investigated yet. Microbial strains of airborne origin have been shown to survive and develop
65 under conditions typically found in cloud water (*i.e.* high concentrations of H₂O₂, typical cloud
66 carbonaceous sources, UV radiation *etc.*) (Amato et al., 2007; Joly et al., 2015; Vaïtilingom et
67 al., 2013). While atmospheric chemicals might lead to some microbial adaptation, physical and
68 unfavorable conditions of the atmosphere such as UV radiation, low water content and cold
69 temperatures might select which microorganisms can survive in the atmosphere. From the pool
70 of microbial cells being aerosolized from Earth's surfaces, these adverse conditions might act
71 as a filter in selecting cells already resistant to unfavorable physical conditions. Fungal cells
72 and especially fungal spores might be particularly adapted to survive in the atmosphere due to
73 their innate resistance (Huang and Hull, 2017) and might behave differently than bacterial cells.
74 Still, the proportion and nature (*i.e.* fungi versus bacteria) of microbial cells that are resistant to
75 the harsh atmospheric conditions within airborne microbial communities are unknown.

76 Our objective was to determine whether airborne microorganisms in the planetary boundary
77 layer possess a specific functional signature as compared to other ecosystems since this might
78 indicate that microorganisms with specific functions tend to be more aerosolized and/or
79 undergo a higher survival in this environment. Our previous study showed that airborne
80 microbial taxonomy mainly depends on the underlying ecosystems, indicating that the local
81 environments are the main source of airborne microorganisms (Tignat-Perrier et al., 2019). Still,
82 we do not know if airborne microbial communities result from random or specific
83 aerosolization of the underlying ecosystems' microorganisms. We used a metagenomic
84 approach to compare the differences and similarities of both the overall functional potential and
85 specific microbial functions (metabolic and stress-related functions) between microbial
86 communities from the atmosphere and other ecosystems (soil, sediment, surface seawater, river
87 water, snow, human feces, phyllosphere and hydrothermal vent). We sampled airborne
88 microbial communities at nine different locations around the world during several weeks to get
89 a global-scale view and to capture the between and within-site variability in atmospheric
90 microbial functional potential.

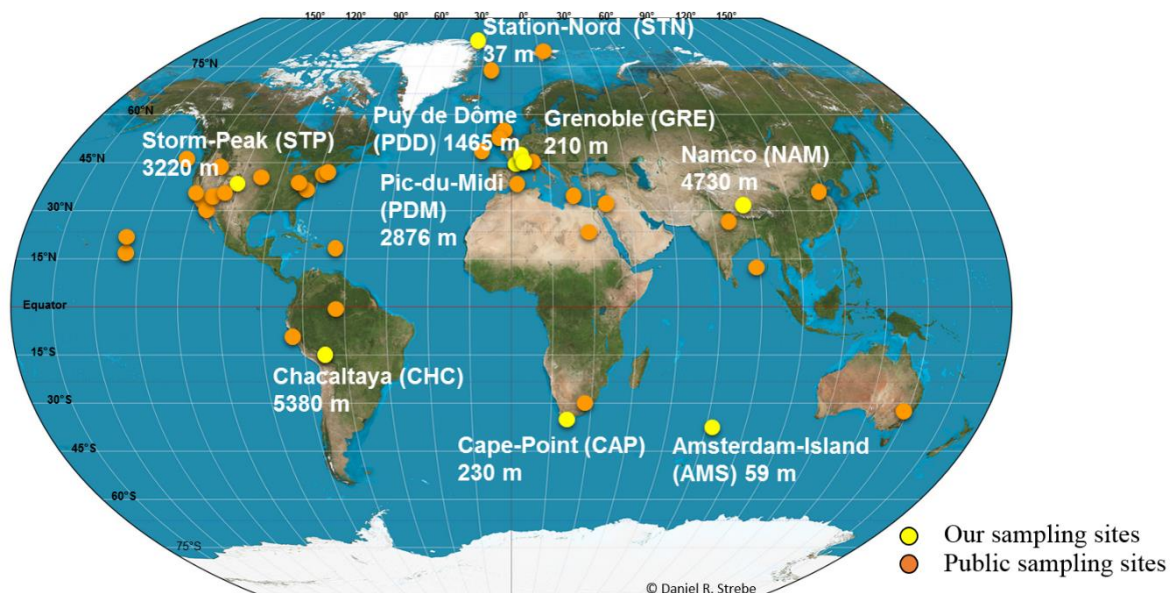
91

92 **2 Material and Methods**

93 **2.1 Sites and sampling**

94 Air samples were collected at nine sites in 2016 and 2017. Sites were characterized by different
95 latitudes (from the Arctic to the sub-Antarctica; **Fig 1**), elevations from sea level (from 59 m to
96 5230 m; **Fig 1**) and environment type (from marine for Amsterdam-Island or AMS, to coastal
97 for Cape Point or CAP, polar for Station Nord or STN and terrestrial for Grenoble or GRE,
98 Chacaltaya or CHC, puy de Dôme or PDD, Pic-du-Midi or PDM, Storm-Peak or STP and
99 Namco or NAM - **Table S1**). The number of samples collected per site varied from seven to

100 sixteen (**Table S1**). We collected particulate matter smaller than 10 μm (PM10) on quartz fiber
 101 filters (5.9'' round filter and 8'' \times 10'' rectangular types) using high volume air samplers
 102 installed on roof tops or terraces (roughly 10 m above ground level). A TISCH TE-5170V
 103 sampler, DIGITEL DA77/DA80 sampler, Chinese 2131 Laowin sampler or a custom made
 104 sampler (*i.e.* high-volume pump connected to a DIGITEL PM10 head and airflow meter) was
 105 used depending on the site (**Table S1**). To avoid contamination, quartz fiber filters as well as
 106 all the material in contact with the filters (*i.e.* filter holders, aluminium foils and plastic bags in
 107 which the filters were transported) were sterilized using strong heating (500 $^{\circ}\text{C}$ for 8 h) and UV
 108 radiation, respectively as detailed in (Dommergue et al., 2019). The collection time per sample
 109 lasted one week, and the collected volumes ranged from 2000 m^3 to 10000 m^3 after
 110 standardization using SATP standards (Standard Ambient Pressure and Temperature). Detailed
 111 sampling protocols including negative control filters are presented in Dommergue et al. 2019.
 112 MODIS (Moderate resolution imaging spectroradiometer) land cover approach (5' \times 5'
 113 resolution) (Friedl et al., 2002; Shannan et al., 2014) was used to quantify landscapes in the 50
 114 km diameter area of our nine sampling sites (**Fig S1**).
 115
 116



117
 118 **Fig 1. Sample collection locations.** Map showing the geographical location and elevation from
 119 sea level of our nine sampling sites (in yellow), and the geographical position of whose public
 120 metagenomes come from (in orange). Abbreviations of our nine sampling sites are indicated in
 121 brackets.

122
 123 **2.2 Molecular biology analyses**
 124 **2.2.1 DNA extraction**

125 DNA was extracted from three circular pieces (punches) from the quartz fiber filters (diameter
 126 of one punch: 38 mm) using the DNeasy PowerWater kit with some modifications as detailed
 127 in (Dommergue et al., 2019). During cell lysis, the PowerBead tube containing the three
 128 punches and the pre-heated lysis solution were heated at 65 $^{\circ}\text{C}$ during one hour after a 10-min
 129 vortex treatment at maximum speed. We then separated the filter debris from the lysate by
 130 centrifugation at 1000 rcf for 4 min. From this step on, we followed the DNeasy PowerWater
 131 protocol. We conducted additional extractions on French agricultural soil samples collected at
 132 the Côte Saint André (that is part of the sample collection locations). We used 250 mg of soil

133 on which the same DNA extraction methodology as for air samples was applied. DNA
134 concentration eluted in 100 μ L of buffer was measured using the Qubit Fluorometric
135 Quantification kit (Thermo Fisher Scientific). DNA was stored at -20 $^{\circ}$ C.
136

137 **2.2.2 Real-Time qPCR analyses**

138 **16S rRNA gene and 18rRNA gene qPCR.** The 16S and 18SS rRNA gene copy numbers were
139 calculated per cubic meter of air (for air samples) and per gram of soil (for soil samples).
140 Standards, primers and methodology are presented in Tignat-Perrier et al., 2019.
141

142 **2.2.3 MiSeq Illumina metagenomic sequencing**

143 **Metagenomic library preparation.** Metagenomic libraries were prepared from 1 ng of DNA
144 using the Nextera XT Library Prep Kit and indexes following the protocol in Illumina's
145 "Nextera XT DNA Library Prep Kit" reference guide with some modifications for samples with
146 DNA concentrations below 1 ng as follows. The tagmented DNA was amplified over 13 PCR
147 cycles instead of 12 PCR cycles, and the libraries (after indexing) were resuspended in 30 μ L
148 of RBS buffer instead of 52.5 μ L. Metagenomic sequencing was performed using the MiSeq
149 and V2 technology of Illumina with 2 x 250 cycles. At the end of the sequencing, the adapter
150 sequences were removed by internal Illumina software.

151 **Reads quality filtering.** Reads 1 and reads 2 per sample were not paired but merged in a
152 common file before filtering them based on read quality using the tool FASTX-Toolkit
153 (http://hannonlab.cshl.edu/fastx_toolkit/) using a minimum read quality of Q20, minimum read
154 length of 120 bp and one maximum number of N per read. Samples with less than 6000 filtered
155 sequences were removed from the dataset.
156

157 **2.2.4 Downloading of public metagenomes**

158 Public metagenomes were downloaded from the MGRAST and SRA (NCBI) databases as
159 quality filtered read-containing fasta files and raw read containing fastq files, respectively. The
160 fastq files containing raw reads underwent the same quality filtering as our metagenomes (as
161 discussed above). The list of the metagenomes, type of ecosystem, number of sequences and
162 sequencing technology (*i.e.* MiSeq, HiSeq or 454) are summarized in **Table S2**. The sampling
163 sites are positioned on the map in **Fig 1**.
164

165 **2.3 Data analyses**

166 All graphical and multivariate statistical analyses were carried out using the vegan (Oksanen et
167 al., 2019), ggplot2 (Hadley and Winston, 2019) and reshape2 (Wickham, 2017) packages in the
168 R environment (version 3.5.1).
169

170 **2.3.1 Annotation of the metagenomic reads**

171 Firstly, to access the overall functional potential of each sample, all the filtered sequences per
172 sample were functionally annotated using Diamond and the nr database, then the gene-
173 annotated sequences were grouped in the different SEED functional classes (around 7000
174 functional classes, referred simply to as functions) using MEGAN version 6 (Huson et al.,
175 2009). Functional classes that were present ≤ 2 times in a sample were removed of this sample.
176 In parallel, the Kraken software (Wood and Salzberg, 2014) was used to retrieve the bacterial
177 and fungal sequences separately from the filtered sequences using the Kraken bacterial database
178 and FindFungi (Donovan et al., 2018) fungal database (both databases included complete
179 genomes), respectively (and using two different runs of Kraken). Separately, both the bacterial
180 and fungal sequences were also functionally annotated using Diamond and MEGAN version 6
181 (number of sequences functionally annotated in **Table S3**).

182 Secondly, for specific metabolic and stress-related functions, we annotated the sequences using
183 eggNOG-Mapper version 1 (Diamond option), then examined specific GO (Gene Ontology)
184 terms chosen based on their importance for microbial resistance to atmospheric-like conditions.
185 The different GO terms used were the following: GO:0042744 (hydrogen peroxide catabolic
186 activity), GO:0015049 (methane monooxygenase activity) as specific metabolic functions and
187 GO:0043934 (sporulation), GO:0009650 (response to UV), GO:0034599 (cell response to
188 oxidative stress), GO:0009269 (response to desiccation) as stress-related functions. The number
189 of hits of each GO term was normalized per 10000 annotated sequences and calculated from all
190 sequences, bacterial sequences and fungal sequences for each sample. The ratio was given per
191 10000 annotated sequences and not 100 annotated sequences to get hit numbers superior to 1.
192 The number of sequences annotated by eggNOG-Mapper (Huerta-Cepas et al., 2017) was also
193 evaluated (**Table S3**). The putative concentration of a specific function or functional class in
194 the samples is determined as the concentration of sequences annotated as one of the functional
195 proteins associated to this function (or functional class).

196

197 **2.3.2 Statistical analyses**

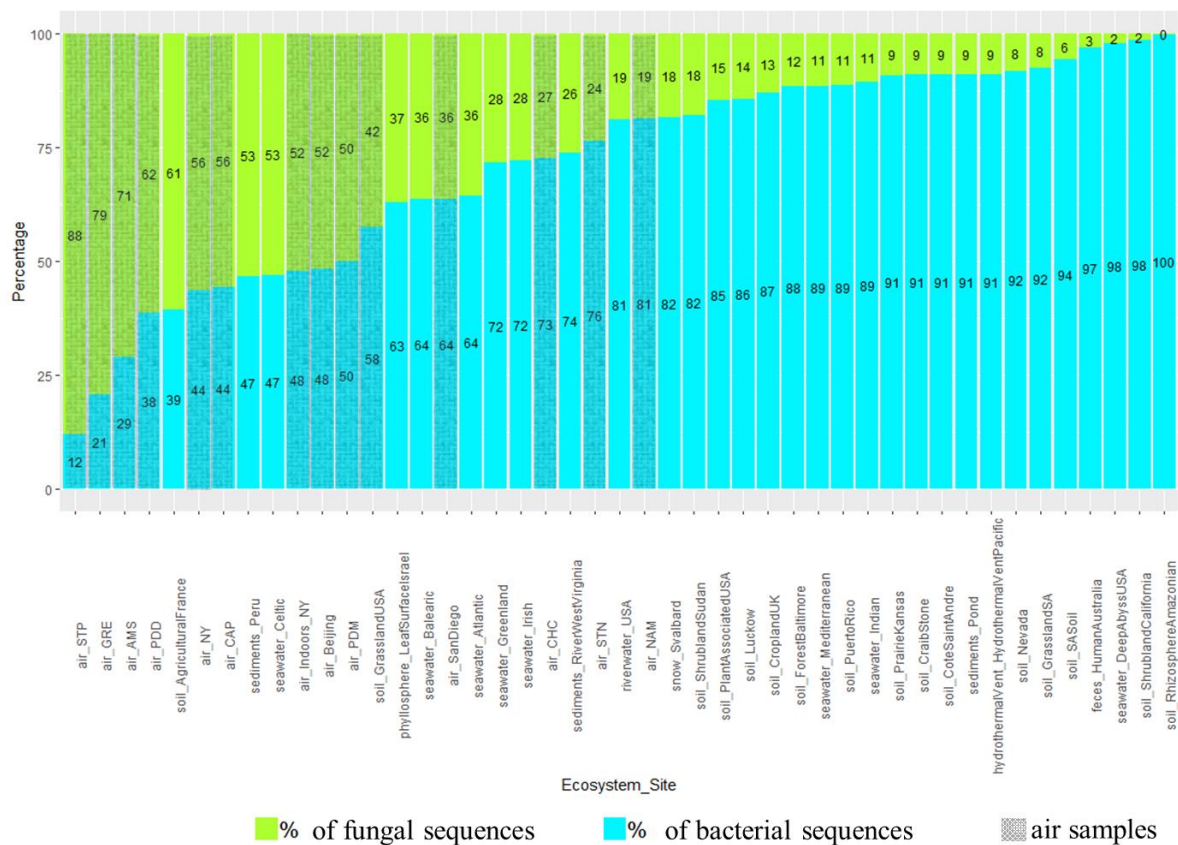
198 Observed functional richness and evenness were calculated per sample after rarefaction on all
199 sequences (rarefaction at 2000 sequences), bacterial sequences (rarefaction at 500 sequences)
200 and fungal sequences (rarefaction at 500 sequences). Rarefaction was used only for the purpose
201 of calculating the diversity metrics (richness and evenness). For the following analyses, no
202 rarefaction was applied on the number of reads per sample. The distribution of the samples was
203 analyzed based on the SEED functional classes (using all sequences). PCoA and hierarchical
204 clustering analysis (average method) were carried out on the Bray-Curtis dissimilarity matrix
205 based on the relative abundances of the different SEED functional classes. SIMPER analyses
206 were used to identify the functions responsible for the clustering of samples in groups. Because
207 of the non-normality of the data, Kruskal-Wallis analyses (non-parametric version of ANOVA)
208 and Dunn's post-hoc tests were used to test the difference between the percentage of fungal
209 sequences as well as the number of hits of each Gene Ontology term (normalized per 10000
210 annotated sequences) among the different sites and the different ecosystems.

211

212 **3 Results**

213 **3.1 Percentage of fungal sequences**

214 The percentage of sequences annotated as belonging to fungal genomes (or fungal sequences,
215 as opposed to bacterial sequences) was on average higher in air samples compared to soil ($P < 10^{-5}$),
216 snow ($P = 10^{-3}$), seawater ($P = 0.03$) and sediment samples ($P = 10^{-3}$; **Fig 2** and **Table S4**).
217 Among the air samples, NAM (19%), STN (24%) and CHC (27%) showed the lowest
218 percentages of fungal sequences on average while STP (88%), GRE (79%), AMS (71%) and
219 PDD (62%) showed the highest percentages. For the ecosystems that were only represented by
220 one sample, and therefore, were not evaluated by the Kruskal-Wallis test, we observed average
221 percentages of fungal reads of 3% in feces, 9% in hydrothermal vents, 19% in river water
222 samples and 37% in the phyllosphere. Some samples from soil, sediments and seawater such as
223 French agricultural soil (61%), Peru sediments (53%) and Celtic seawater (53%) had relatively
224 high percentages of fungal sequences while other samples had less than 50%. The approximated
225 number of fungal and bacterial cells in air and soil was also estimated using 16S rRNA and 18S
226 rRNA gene copy numbers per cubic meter of air and gram of soil, respectively. Air samples
227 showed ratios between 16S and 18S rRNA gene copy numbers from around 4.5 times up to 160
228 times lower than soil samples (**Table S4**; some qPCR data have already been published in
229 Tignat-Perrier et al., 2020). The qPCR data were used to see if similar results, or the same trend,
230 on the estimated ratio between fungi and bacteria in air compared to the planetary-bound
231 ecosystems could be obtained from metagenomic data and qPCR.



233

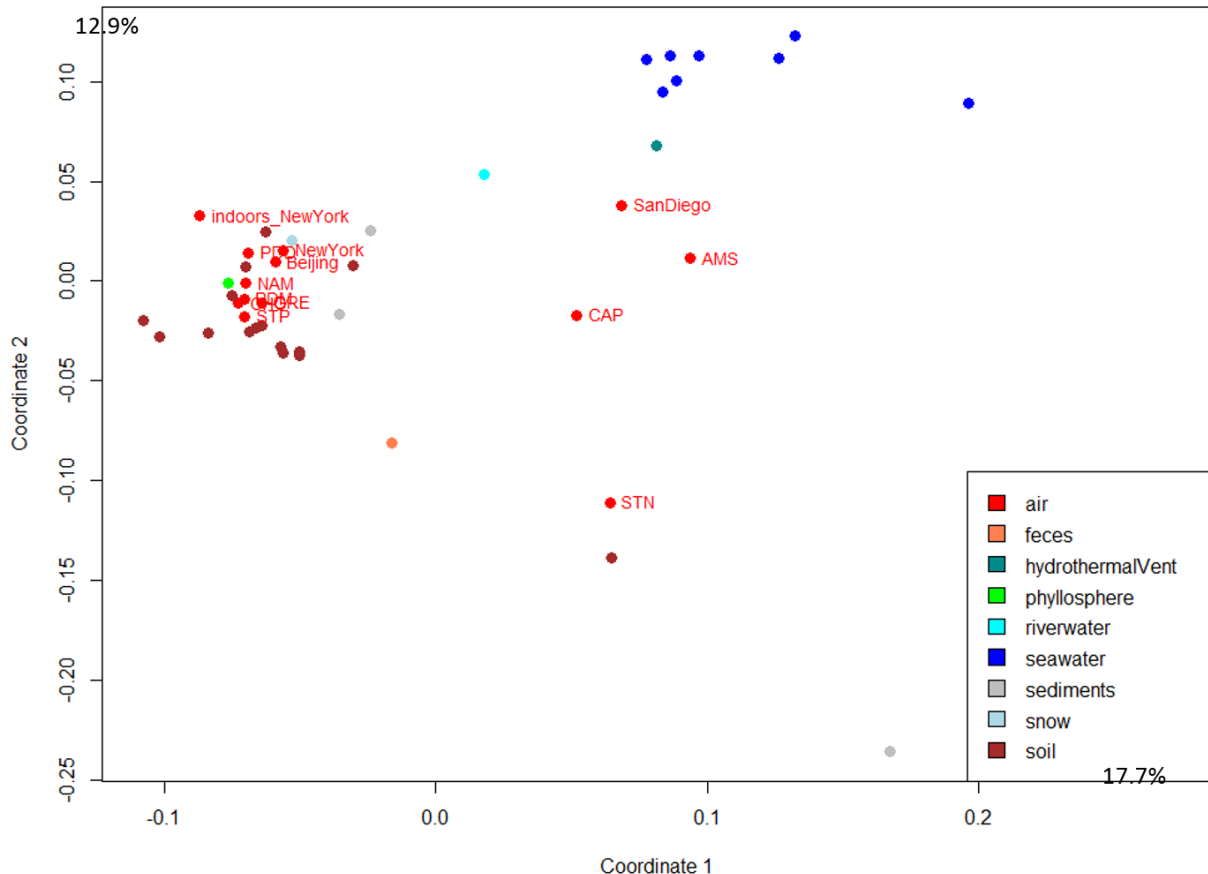
234 **Fig 2. Percentage of fungal and bacterial sequences in the metagenomes.** The percentages
 235 are established as the number of sequences annotated as belonging to fungal and bacterial
 236 genomes over the sum of bacterial and fungal sequences in the metagenomes. The mean was
 237 calculated for the sampling sites including several metagenomes. Air sites (*i.e.* our 9 sites + 5
 238 sites where public air metagenomes come from) are distinguished by grey hatching lines.

239

240 3.2 Airborne microbial functional profiles

241 The fifty most abundant SEED functional classes represented in atmospheric samples are listed
 242 in **Table S5**. The 5-FCL-like protein, the long chain fatty acid CoA ligase and the TonB-
 243 dependent receptor were the top three functions based on number of annotated reads observed
 244 when including all the sequences (**Table S5**). The atmospheric microbial functional profiles
 245 based on the SEED functions were compared between samples from the different weeks of
 246 sampling and between different locations. The profiles were graphed using PCo multivariate
 247 analysis to visualize differences and similarities. The different samples (sampled during
 248 sequential weeks) from the same site did not cluster tightly together on the PCo multivariate
 249 analysis. In order to incorporate weekly variation when comparing sites, we used the microbial
 250 functional profile averaged per site in the subsequent multivariate analyses done with the data
 251 from other ecosystems (**Fig 3**). The PCo multivariate analysis showed that terrestrial
 252 atmospheric sites (GRE, NAM, STP, PDD, PDM, CHC, New York) grouped with the soil,
 253 sediment and snow samples while the marine and coastal atmospheric sites (AMS, CAP, San
 254 Diego) were situated between the datasets from soil, seawater and river water (**Fig 3**). The polar
 255 site STN did not group with the other sites. When considering only the bacterial sequences (*i.e.*,
 256 excluding the fungal sequences), the distribution of the terrestrial atmospheric sites did not
 257 change, while the marine Amsterdam-Island, coastal Cape Point and polar Station Nord

258 atmospheric sites were further from the seawater and river water datasets than when the fungal
 259 sequences were included (**Fig S2**). The distribution of the different datasets underwent further
 260 changes when considering only the fungal sequences. We observed an absence of a clear
 261 separation between soil and seawater since they (for the majority) grouped closely together, and
 262 terrestrial atmospheric datasets did not group with the other non-atmospheric datasets from soil,
 263 sediment and snow (**Fig S2**).
 264



265 **Fig 3. Distribution of the samples based on the microbial functional profile.** The PCo
 266 analysis of the Bray-Curtis dissimilarity matrix is based on the functional potential structure of
 267 each site. For the site including several metagenomes, the average profile was calculated. Colors
 268 indicate the ecosystems in which the sites belong to. The percentages (17.7 % and 12.9 %)
 269 indicate the proportion of the variance explained by the first and second axis, respectively.
 270

271
 272 **3.3 Airborne microbial functional richness and evenness**

273 Functional richness and evenness were evaluated using the relative abundance of sequences in
 274 the different SEED categories. The average richness in SEED functional classes (or functions)
 275 in the PBL was lower than the average functional class richness in soil, surface seawater,
 276 hydrothermal vents, river water, phyllosphere and feces ($P < 0.05$) (**Table S3**). Among the
 277 different atmospheric samples, the functional class richness was highest in Beijing (4060 +/-
 278 112 functional classes) and New York indoor air samples (3302 +/- 299 functional classes)
 279 ($P < 0.05$), and lowest in Station Nord (956 +/- 547 functional classes). When looking at the
 280 bacteria-annotated sequences, almost the same trend was observed, *i.e.* the functional class
 281 richness in air was lower than in soil, hydrothermal vents, river water, phyllosphere and feces,
 282 and not different from the other ecosystems ($P < 0.05$ and > 0.05 , respectively) (**Table S3**). The

283 functional class richness was higher in Beijing (2835 +/- 59 functional classes) and New York
284 indoor air samples (2183 +/- 387 functional classes) compared to the other air samples whose
285 values ranged between 270 +/- 197 functional classes in Amsterdam-Island and 1142 +/- 461
286 functional classes in Chacaltaya. For fungal sequences, the functional class richness in the
287 atmosphere was lower than the functional class richness in soil, surface seawater, feces,
288 hydrothermal vents, river water and phyllosphere ($P < 0.05$) (**Table S3**). Within air samples, the
289 functional class richness based on fungal sequences was higher in Beijing (1129 +/- 92
290 functional classes) and New York indoor air samples (687 +/- 206 functional classes) than in
291 the other air sites ($P < 10^{-5}$) whose values ranged from 66 +/- 58 functional classes in
292 Amsterdam-Island and 392 +/- 131 functional classes in Storm Peak (**Table S3**). The functional
293 class evenness in air was on average higher than in soil ($P = 0.03$), and not different to the
294 functional class evenness observed in the other ecosystems (sediment, seawater, snow). When
295 looking at the bacterial and fungal sequences separately, the functional class evenness in air
296 was on average higher than in soil, feces, phyllosphere and riverwater ($P < 0.05$) (**Table S3**).

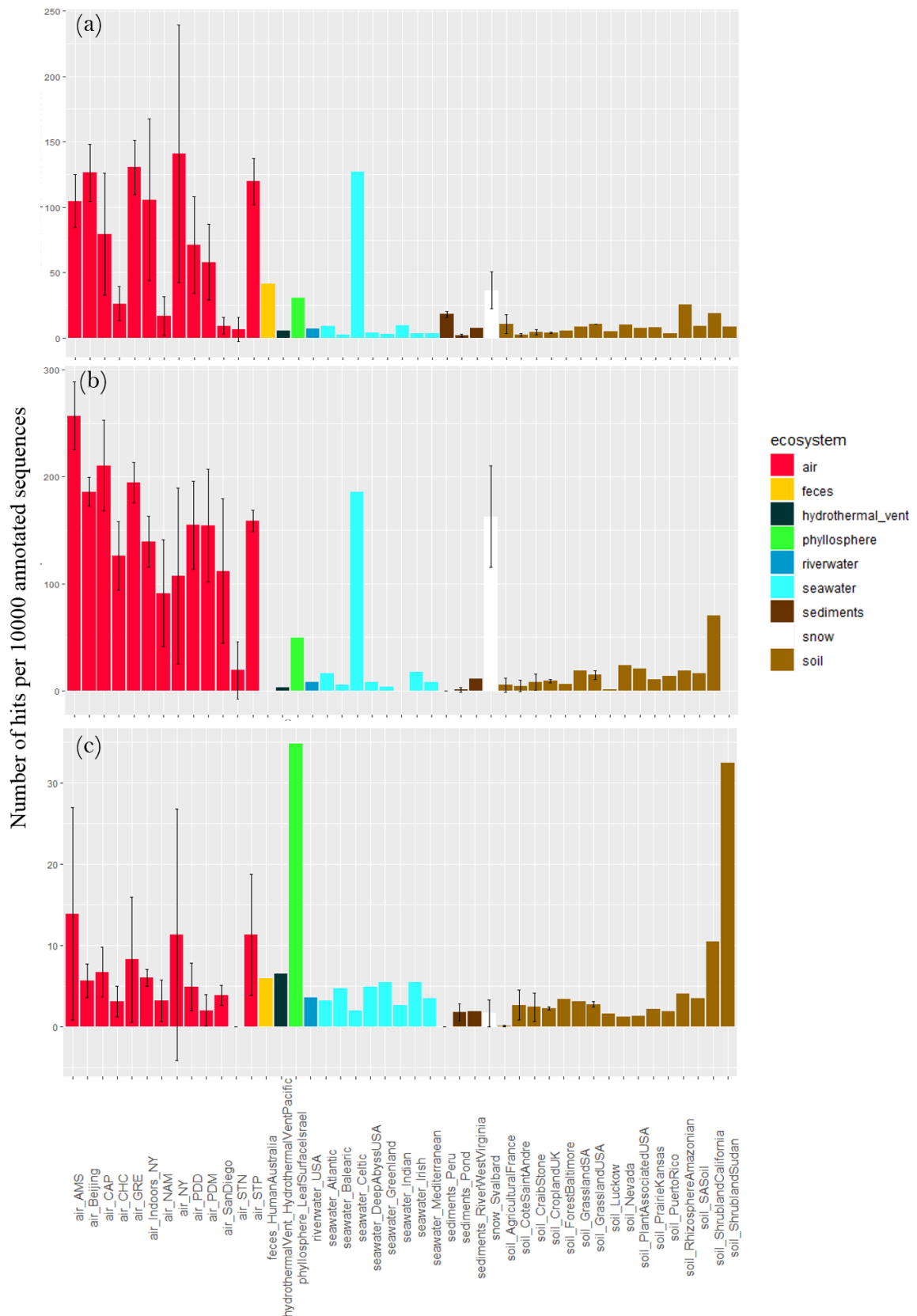
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298 **3.4 Concentration of specific microbial functions that might have a role under** 299 **atmospheric conditions**

300 Two metabolic functions associated with abundant atmospheric chemicals (H_2O_2 and CH_4)
301 were examined, hydrogen catabolism and methane monooxygenase activity. The concentration
302 of sequences annotated as hydrogen peroxide catabolic related functional proteins per 10000
303 sequences varied between air sites ($P = 2 \times 10^{-5}$) with highest values for Amsterdam-Island (27
304 +/- 1) and Grenoble (27 +/- 1) (**Fig S3**). It was on average higher in air compared to soil ($P = 10^{-4}$)
305 and surface seawater ($P = 10^{-4}$). The French agricultural soil showed the highest relative
306 abundance (133 +/- 4). When considering the fungal and bacterial sequences separately, this
307 concentration was not different between air and the other ecosystems ($P > 0.05$) (**Fig S3**). The
308 number of sequences annotated as methane monooxygenase-related functional proteins per
309 10000 sequences was only detectable when considering all the sequences (*i.e.* bacterial and
310 fungal sequences). The number of sequences annotated as methane monooxygenase-related
311 functional proteins did not vary between air sites ($P > 0.05$) while we observed a high variability
312 between sampling periods within sites, but on average it was not different from the ecosystems
313 ($P > 0.05$).

314 Different stress response functions (sporulation, UV response, oxidative stress cell response,
315 desiccation response, chromosome plasmid partitioning protein ParA and lipoate synthase)
316 were examined. The concentration of sequences annotated as sporulation-related functional
317 proteins per 10000 annotated sequences largely varied between air sites ($P = 2 \times 10^{-9}$), with the
318 lowest values observed for Station Nord (7 +/- 9), San Diego (9 +/- 6), Namco (17 +/- 15) and
319 Chacaltaya (26 +/- 13), and the highest values observed for Storm Peak (120 +/- 18), Beijing
320 (126 +/- 22), Grenoble (131 +/- 21) and New York (141 +/- 98) (**Fig 4**). It was on average higher
321 in air compared to soil ($P < 10^{-5}$), sediments ($P < 10^{-5}$) and surface seawater ($P = 4 \times 10^{-4}$) although
322 the Celtic seawater sample presented a very high concentration (127). Snow showed a relatively
323 high average concentration (*i.e.* 36) which was not different from air concentration ($P > 0.05$).
324 For the ecosystems including one value (*i.e.* one sample, so not integrated in the Kruskal-Wallis
325 tests), feces showed a relatively high concentration of sequences annotated as sporulation-
326 related functional proteins (*i.e.* 41) while hydrothermal vent, phyllosphere and river water
327 showed relatively low concentrations compared to air (<10). When considering the fungal
328 sequences separately from the bacterial sequences, the same trend was observed, *i.e.* the
329 concentration of sequences annotated as sporulation-related functional proteins in air was on
330 average higher compared to soil ($P < 10^{-5}$), sediments ($P < 10^{-5}$), surface seawater ($P = 7 \times 10^{-4}$) as
331 well as phyllosphere, hydrothermal vent and river water. The concentration was relatively high
332 in the Celtic seawater (186) and the snow samples (163 +/- 47). We also observed a large

333 variability within air sites ($P=3\times 10^{-5}$). When considering the bacterial sequences only, this
334 concentration in air was on average higher compared to soil ($P=0.02$), sediments ($P=4\times 10^{-3}$)
335 and snow ($P=0.01$), and showed a smaller variability between air sites. Two samples, the
336 phyllosphere (*i.e.* 35) and the shrubland soil from Sudan (*i.e.* 32) showed high numbers of
337 sequences annotated as sporulation-related functional proteins per 10000 annotated sequences
338 (**Fig 4**).
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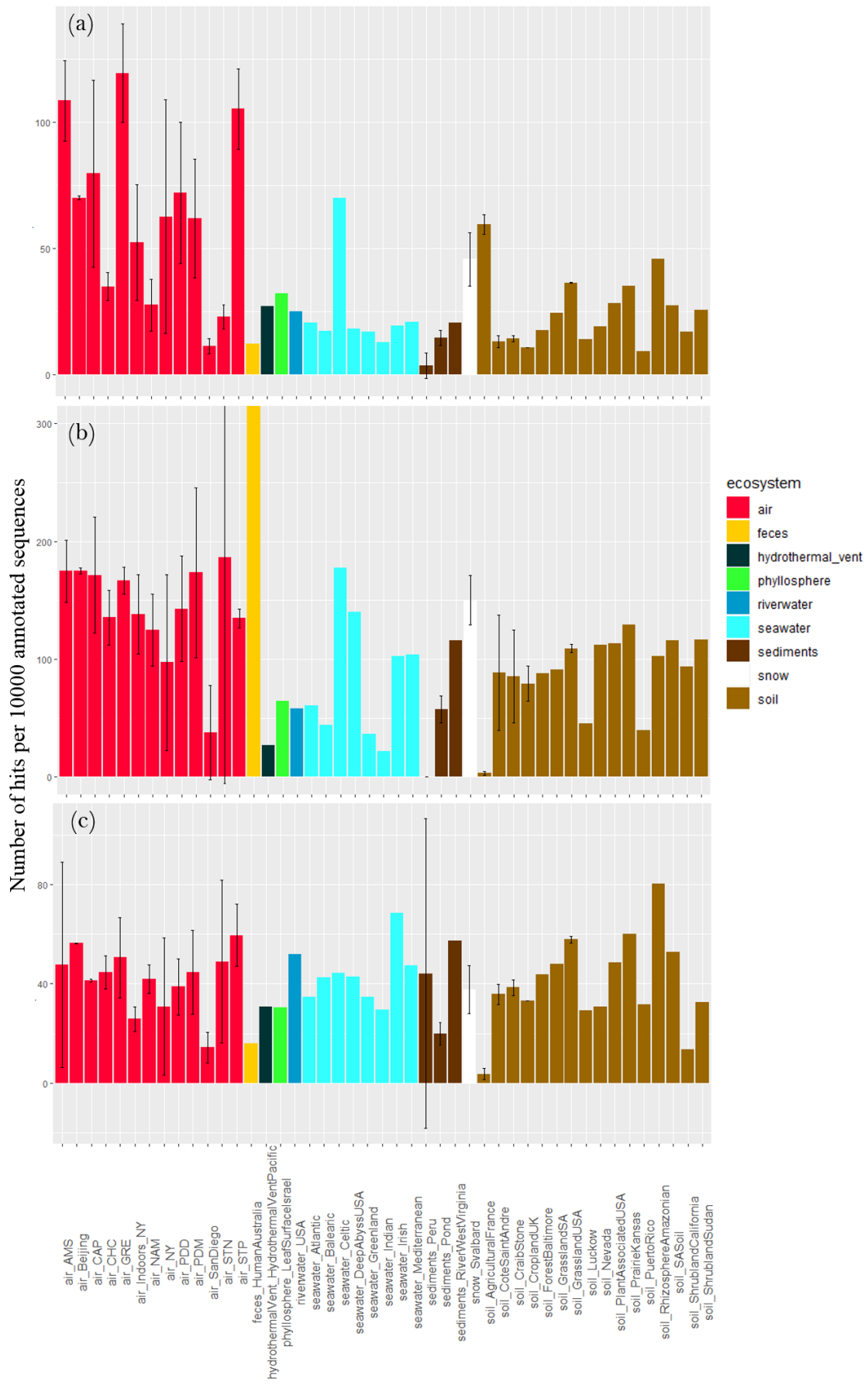
Fig 4. Proportion of sequences annotated as sporulation related functional proteins in the metagenomes. Average number of sequences annotated as proteins implicated in sporulation per 10000 annotated sequences from (a) all sequences, (b) fungal sequences and (c) bacterial

344 sequences per site. Colors indicate the ecosystems in which the sites belong to. For the sites
345 including several metagenomes, the standard deviation was added.

346
347 The concentration of sequences annotated as UV response related functional proteins per 10000
348 annotated sequences varied between air sites ($P=10^{-5}$), with values ranging from 16 +/- 2 in
349 Namco and 19 +/- 4 in STN to 29 +/- 3 in Storm Peak and 36 +/- 6 in Amsterdam-Island (**Fig**
350 **S4**). The concentration was on average higher in air compared to sediments ($P<10^{-5}$), soil
351 ($P<10^{-5}$) and comparable to snow and surface seawater ($P>0.05$). The other ecosystems showed
352 lower ratios (feces, phyllosphere) or comparable concentrations (hydrothermal vent, river
353 water) compared to air. Within the soil samples, the French agricultural soil samples showed a
354 high average concentration (56 +/- 8), which increased the average ratio observed in soil
355 samples. When considering fungal sequences separately, the concentration of sequences
356 annotated as UV response related functional proteins was higher in air compared to soil
357 ($P=9\times 10^{-4}$), and comparable to the other ecosystems ($P>0.05$). When considering the bacterial
358 sequences only, this concentration in air was on average higher compared to seawater ($P=3\times 10^{-3}$)
359 and sediments ($P=6\times 10^{-3}$).

360 The concentration of sequences annotated as oxidative stress cell response related functional
361 proteins per 10000 annotated sequences varied largely between air sites ($P=5\times 10^{-7}$), with the
362 lowest values observed for Station Nord (23 +/- 5), San Diego (11 +/- 3) and Namco (28 +/-
363 10), and the highest values observed for Storm Peak (105 +/- 16), Amsterdam-Island (108 +/-
364 16) and Grenoble (119 +/- 19) (**Fig 5**). The concentration was on average higher in air compared
365 to soil ($P<10^{-5}$), sediments ($P<10^{-5}$) and surface seawater ($P=2\times 10^{-3}$). Snow showed a relatively
366 high average value (46 +/- 11), not different from air ($P>0.05$). The other ecosystems (feces,
367 river water, hydrothermal vent, phyllosphere) showed lower ratios compared to air. When
368 considering fungal sequences separately, the concentration of sequences annotated as oxidative
369 stress related functional proteins per 10000 sequences was on average higher in air compared
370 to soil ($P<10^{-5}$), sediments ($P<10^{-5}$) and surface seawater ($P=10^{-3}$). Feces showed a very high
371 average value (2237). When considering bacterial sequences separately, this concentration was
372 not different between air and the other ecosystems ($P>0.05$). When considering both fungal and
373 bacterial sequences separately, the variability in the concentration of sequences annotated as
374 oxidative stress cell response related functional proteins between air sites diminished and their
375 difference was not detected anymore ($P>0.05$).

376



378 **Fig 5. Proportion of sequences as oxidative stress cell response related functional proteins**
379 **in the metagenomes.** Average number of sequences annotated as proteins implicated in
380 oxidative stress cell response per 10000 annotated sequences from (a) all sequences, (b) fungal
381 sequences and (c) bacterial sequences per site. Colors indicate the ecosystems in which the sites
382 belong to. For the sites including several metagenomes, the standard deviation was added.

383

384

385 The concentration of sequences annotated as desiccation response related functional proteins
386 per 10000 sequences varied between air sites ($P=2\times 10^{-5}$), with the highest values in Grenoble
387 (4 +/- 1), Storm Peak (4 +/- 1) and Amsterdam-Island (3 +/- 3), and the lowest values in Station
388 Nord (0.5 +/- 1) and San Diego (0.1 +/- 0.1) (**Fig S4**). It was on average higher in air compared
389 to the other ecosystems ($P=4\times 10^{-9}$). Still Svalbard snow and French agricultural soil showed
390 high values (2 +/- 1 and 3 +/- 1, respectively) (**Fig S4**). When considering fungal sequences
391 only, the concentration in air was higher compared to soil ($P>10^{-5}$), sediments ($P>10^{-5}$) and
392 surface seawater ($P=10^{-3}$). No difference between the ecosystems was observed when
393 considering bacterial sequences separately ($P=0.62$).

394 Two proteins (lipoate synthase and chromosome plasmid partitioning protein ParA) related to
395 stress response showed high relative concentrations in bacterial sequences of a few air samples
396 compared to the other ecosystems (**Fig S3**), although the number of sequences related to these
397 proteins was on average not higher in the atmosphere than other ecosystems ($P>0.05$).

398

399 **4 Discussion**

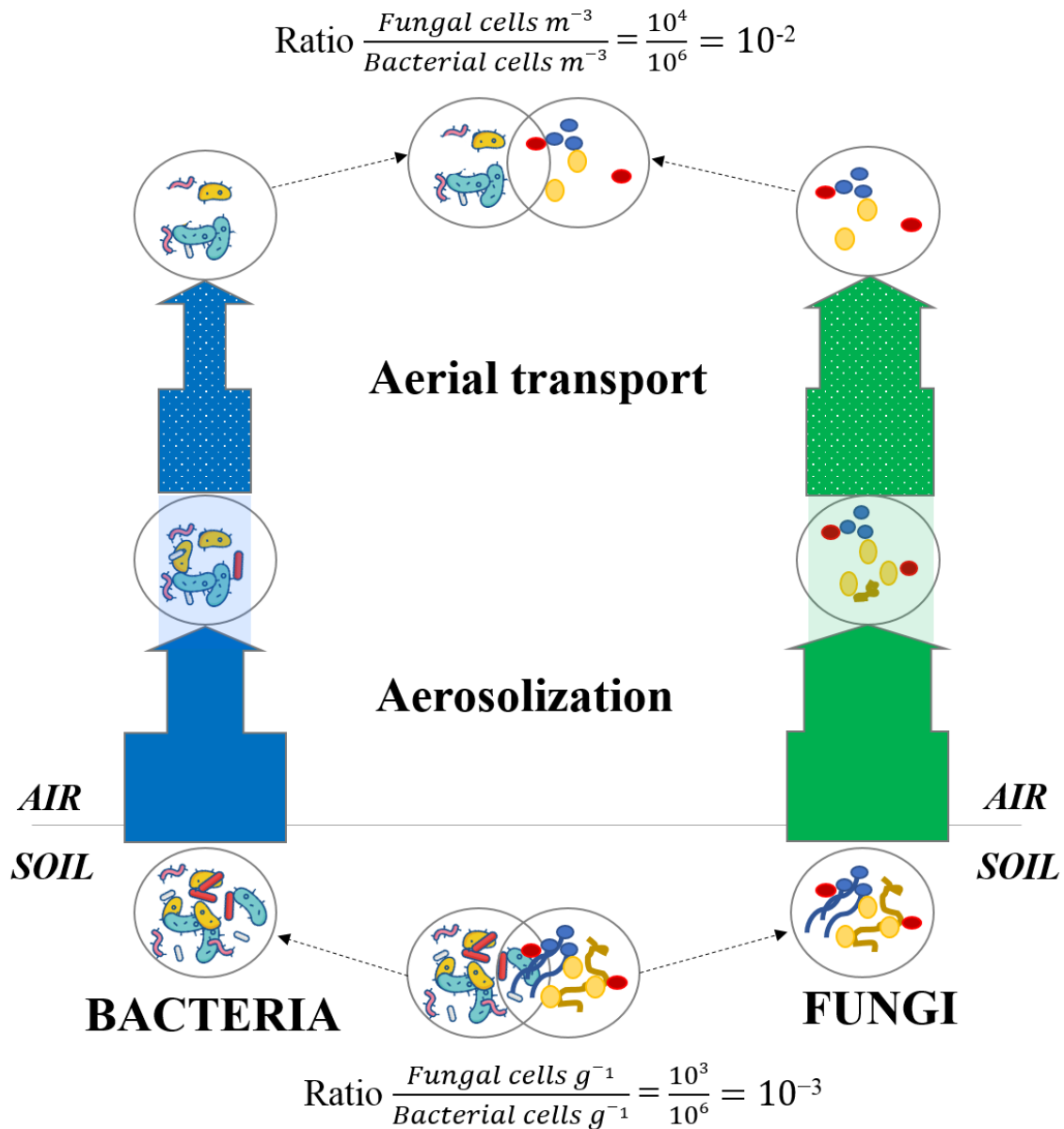
400 Metagenomic investigations of different ecosystems revealed a specific functional potential
401 signature of their associated microbial communities (Delmont et al., 2011; Tringe et al., 2005).
402 These specific signatures are thought to result from microbial adaptation and/or physical
403 selection to the environmental abiotic conditions (Hindré et al., 2012; Li et al., 2019; Rey et al.,
404 2016) and are a reflection of the high relative abundances of genes coding for specific functions
405 essential for microorganisms to survive and develop in these environments. For example,
406 microbial metagenomes of human feces were characterized by high relative abundances of
407 sequences annotated as beta-glucosidases that are associated with high intestinal concentrations
408 of complex glycosides; and microbial metagenomes of oceans were enriched in sequences
409 annotated as enzymes catalyzing DMSP (dimethylsulfoniopropionate), that is an organosulfur
410 compound produced by phytoplankton (Delmont et al., 2011). Our results showed a clear
411 separation between surface seawater, river water, human feces and almost all the soil samples
412 (which grouped with the sediment and snow samples at the scale used here) on the PCo analysis
413 based on the microbial functional potential (**Fig 3**). For air microbiomes, the PCo analyses
414 showed that the individual air samples did not group for each site and that they did not form a
415 cluster separated from the other ecosystems based on the overall microbial functional potential
416 averaged per site (**Fig 3**). Air samples seemed to group with their underlying ecosystems (**Fig**
417 **S1**). While terrestrial air samples (GRE, NAM, CHC, STP, PDD, PDM) grouped with snow,
418 soil and sediment samples, the marine (Amsterdam-Island), coastal (Cape Point) and arctic
419 (Station Nord) air samples were closer to surface seawater and river water samples. Airborne
420 microbial functional potential (and especially metabolic functional potential as SEED
421 functional classes included mainly metabolic functions and few stress response related
422 functions) might be dependent on the ecosystems from which microorganisms are aerosolized.
423 Moreover, it seems that bacterial sequences are mainly responsible for the distribution of the
424 samples on the PCo analysis (as observed when comparing the PCoA to that carried out with
425 the fungal sequences only) although they were in smaller numbers compared to fungal
426 sequences for many of the air samples (*i.e.* STP, GRE, AMS, PDD, CAP, Beijing *etc.*). The
427 low statistical weight of fungal sequences relative to the overall sequences might be related to

428 their low richness in terms of functional genes that might have resulted in the spreading of the
429 samples on the PCoA based on the fungal sequences (**Table S3**).

430 Metagenomes extracted from atmospheric samples taken around the planet were characterized
431 by a relatively high percentage of fungal sequences as compared to other ecosystems even
432 though bacterial sequences still dominated. This percentage varied across the different sites
433 with a higher percentage at terrestrial sites whose surrounding landscapes were vegetated like
434 Grenoble (GRE), puy de Dôme (PDD) and Pic-du-midi (PDM) (surrounding landscapes in **Fig**
435 **S1**). This percentage was also relatively high at the marine site Amsterdam-Island (AMS),
436 where fungi might come from the ocean and/or the vegetated surfaces of the small island. A
437 high percentage of fungal sequences was also reported for air samples from Beijing, New York
438 and San Diego and validates our DNA extraction method set-up specifically for quartz fiber
439 filter (Dommergue et al., 2019). Similarly, the sequencing technology (Illumina MiSeq) could
440 not have been responsible for the larger percentage of fungal sequences observed in our datasets
441 as the Beijing and New York/San Diego air sample datasets originated from Illumina HiSeq
442 and 454 sequencing technology, respectively. qPCR results on the 16S rRNA gene (bacterial
443 cell concentration estimation) and on the 18S rRNA gene (fungal cell concentration estimation)
444 on our air samples in comparison to agricultural soil samples evidenced that the ratio between
445 fungal and bacterial cell number might be much higher (from 4.5 to 160 times higher for the
446 most vegetated site Grenoble) in air than in soil (**Table S4**; see Tignat-Perrier et al., 2020 for
447 more qPCR data on air samples). The ratio between fungal and bacterial cell number might be
448 higher in the planetary boundary layer (PBL) than in other environments like soil (Malik et al.,
449 2016), and thus, would explain the relatively higher percentage of fungal sequences observed
450 in air metagenomes. High throughput sequencing allows the sequencing of a small part of the
451 metagenomic DNA (with large fungal genomes likely to be sequenced first) and might explain
452 why the values of the ratio between 16S and 18S rRNA gene copy numbers obtained by qPCR
453 does not match exactly those obtained by the metagenomic sequencing approach, while they
454 show the same trend. Our study is a preliminary metagenomic investigation of the air
455 environment with a limited number of sequences per sample, and further studies are needed to
456 confirm our results.

457 Fungi in the atmosphere are expected to be found mostly as fungal spores. While some spore
458 and hyphae concentrations have been measured in air (Després et al., 2012), the relative
459 concentration of fungal spores and fungal hyphae fragments and its temporal dynamics at the
460 same site remains unknown. Our results showed that the number of sporulation-related
461 functions was higher in air than the other ecosystems (with the exception of snow and
462 phyllosphere). While fungal hyphae are not expected to be particularly resistant to extreme
463 conditions such as UV radiation, fungal spores are specifically produced to resist and survive
464 overall adverse atmospheric conditions (Huang and Hull, 2017). Their thick membrane and
465 dehydrated nature make them particularly resistant to abiotic atmospheric conditions such as
466 UV radiation, oxidative stress, desiccation as well as osmotic stress. **Fig 6** presents a conceptual
467 model that could explain the higher ratio between fungi and bacteria observed in air. During
468 aerosolization and aerial transport, bacteria and fungi might be under stress and might undergo
469 a physical selection with the survival of the most resistant cells to the adverse atmospheric
470 conditions (*i.e.* UV radiation, desiccation *etc.*) and the death of non-resistant cells. As fungi
471 (and especially fungal spores) might be naturally more resistant and adapted to atmospheric
472 conditions than bacteria, we expect a larger decline of bacterial cells compared to fungal cells
473 and spores in air. This might have as a consequence an increase in the ratio between fungi and
474 bacteria compared to their non-atmospheric origins (*i.e.* the surrounding ecosystems) (**Fig 6**).

475



476

477

478 **Fig 6. Microbial cell loss due to atmospheric physical stress.** Conceptual model on the
 479 microbial cell loss occurring during the aerosolization and aerial transport steps due to physical
 480 selection. The thickness of the arrows represents the impact of the physical selection on both
 481 bacterial and fungal cell loss (the more microbial cells survive the physical selection, the thicker
 482 becomes the arrow). Approximate ratios are indicative and result from 16S rRNA and 18S
 483 rRNA gene qPCR data on Côte Saint André soil samples (crop soil, France) and puy de Dôme
 484 air samples (France; puy de Dôme landscape is mainly composed of croplands as shown in **Fig**
 485 **S1**).

486

487 The high variability between the air sites and between air samples of the same site could be
 488 explained by the variability in the inputs from the different surrounding landscapes. Our
 489 previous paper showed that local inputs were the main sources of planetary boundary layer
 490 microorganisms and that local meteorology (especially the wind direction) had a major impact
 491 on the temporal variability of airborne microbial communities by affecting which of the

492 different local sources were upwind (Tignat-Perrier et al., 2019). Our results did not show a
493 specific (metabolic) functional potential signature for the atmosphere, which was rather mainly
494 driven by the surrounding landscapes. Our results are consistent with both a pre-metabolic
495 adaptation of airborne microorganisms to the chemicals of the sources (*i.e.* surrounding
496 landscapes) and a potential metabolic adaptation to these chemicals in the atmosphere.
497 Atmospheric chemistry is dependent on the underlying ecosystem chemistry since the main
498 sources of atmospheric chemicals are Earth surface emissions. Yet, the oxidizing conditions of
499 the atmosphere might lead to rapid transformations of atmospheric chemicals by photochemical
500 reactions. These specific atmospheric chemical reactions (*i.e.* photochemical) produce species
501 which, with the gases like CH₄, characterize the atmosphere (O₃, H₂O₂, OH *etc.*). Although
502 some microbial strains from cloud water origin have been shown to metabolize and grow on
503 culture medium in the presence of H₂O₂ (Vařtilingom et al., 2013), radical species and their
504 precursors are reactive compounds and might not easily serve as energy and carbon sources for
505 microorganisms (Imlay, 2013). Our results on specific metabolic related functions showed that
506 functions related to methane monooxygenase activity (CH₄ degradation) and hydrogen
507 peroxide catabolism (H₂O₂ degradation) were present in air but not in higher proportion than in
508 other ecosystems (**Fig S3**). Reactive compounds can cause oxidative stress to airborne
509 microorganisms. In association to adverse physical conditions like UV radiation and
510 desiccation, oxidative compounds might create more of a physical stress than provide a new
511 metabolic source for airborne microorganisms. Laboratory investigations of cultivable
512 microorganisms of an airborne origin showed the presence of particularly resistant strains under
513 stressful conditions similar to the atmospheric ones (*i.e.* similar UV radiation levels; different
514 oxidative conditions) (Joly et al., 2015; Yang et al., 2008). However, no study has shown
515 whether these apparently adapted cells represented the majority of airborne microorganisms.
516 Since the overall SEED functional classes included mainly metabolic functions, specific stress
517 related functions using GO (Gene Ontology) terms were also evaluated. We observed that on
518 average, air showed more stress-related functions (UV response, desiccation and oxidative
519 stress response related functions) than the other ecosystems due to the higher concentration of
520 fungi (relatively to bacteria) in air. Thus, when the annotated sequences were separated between
521 sequences belonging to fungal and bacterial genomes, the bacterial and fungal sequences from
522 air samples did not show a significantly higher concentration of stress-related functions
523 compared to the samples coming from other ecosystems (**Fig 4, 5, Fig S4**).

524 Fungal genomes are expected to carry genes associated to global stress-related functions (*i.e.*
525 UV radiation, desiccation, oxidative stress), because of the innate resistance of fungi especially
526 fungal spores. These genes associated to global stress-related functions are likely acquired
527 during sporulation formation and certainly do not result from adaptation of fungi in air. When
528 studying genes coding more specific proteins that are not associated to spore resistance, such
529 as lipoate synthase and chromosome plasmid partitioning protein ParA, that might play a role
530 in oxidative stress (Allary et al., 2007; Bunik, 2003) and are more generally found in stress
531 resistance and adaptability of microorganisms (Shoeb et al., 2012; Zhang et al., 2018), they
532 were occasionally found in relatively high concentration in air samples (**Fig S3**). The detection
533 of metagenomic sequences annotated as genes coding specific proteins in air samples remains
534 difficult because of the low microbial biomass recovered. That is why we examined the
535 presence and concentration of global functions (*i.e.* UV protection related functions, oxidative
536 stress response related functions *etc.*) rather than specific functional genes.

537 The constant and large input of microbial cells to the planetary boundary layer and their
538 relatively short residence time (a few hours to a few days based on a model assuming that
539 microbial cells behave like non biological aerosols (Jaenicke, 1980)) might have hindered the
540 observation of the potential adaptation (physical selection and/or microbial adaptation) of
541 airborne microorganisms to the stressful atmospheric conditions and to the atmospheric

542 chemicals as discussed above. This issue might be addressed by investigating microbial
543 functional potential in the free troposphere (preferentially high enough above the ground so as
544 not to be influenced by the surface) where the microbial fluxes are smaller than in the planetary
545 boundary layer and where microbial airborne residence time might last much longer than in the
546 planetary boundary layer. This troposphere approach might help in determining the role of
547 stress in the atmosphere and validate our conceptual model on the physical stress of microbial
548 cells taking place during aerosolization and aerial transport selecting the resistant cells (**Fig 6**).
549 Another explanation might be due to the metagenomic approach that allows to sample both
550 living and dead cells. Aerosolization has been shown to be particularly stressful and even lethal
551 for microorganisms (Alsved et al., 2018; Thomas et al., 2011). The functional potential from
552 the dead cells in air might have a greater weight on the overall functional potential observed
553 and lead to the dilution of the functional potential of the actual living cells that have adapted to
554 atmospheric conditions. This might apply for both the overall functional potential discussed
555 previously and the stress-related functions.

556

557 **Conclusion**

558 We conducted the first global comparative metagenomic analysis to characterize the microbial
559 functional potential signature in the planetary boundary layer. Air samples showed no specific
560 signature of microbial functional potential which was mainly correlated to the surrounding
561 landscapes. However, air samples were characterized by a relatively high percentage of fungal
562 sequences compared to the source ecosystems (soil, surface seawater *etc.*). The relatively higher
563 concentrations of fungi in air drove the higher proportions of stress-related functions observed
564 in air metagenomes. Fungal cells and specifically fungal spores are innately resistant entities
565 well adapted to atmospheric conditions and which might survive better aerosolization and aerial
566 transport than bacterial cells. Stress-related functions were present in airborne bacteria but
567 rarely in higher concentrations compared to the bacterial communities in other ecosystems.
568 However, the constant flux of microbial cells to the planetary boundary layer might have
569 complicated the determination of a physical selection and/or microbial adaptation of airborne
570 microorganisms, especially bacterial communities. Meta-omics investigations on air with a
571 deeper sequencing are needed to confirm our results and explore the functionality of
572 atmospheric microorganisms further.

573

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732

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