

1 Microbial functional signature in the atmospheric boundary layer

2
3 Romie Tignat-Perrier^{1,2*}, Aurélien Dommergue¹, Alban Thollot¹, Olivier Magand¹, Timothy
4 M. Vogel², Catherine Larose²

5
6 ¹Institut des Géosciences de l'Environnement, Université Grenoble Alpes, CNRS, IRD,
7 Grenoble INP, Grenoble, France

8 ²Environmental Microbial Genomics, Laboratoire Ampère, École Centrale de Lyon, Université
9 de Lyon, Écully, France

10
11 *Correspondence to:* Romie Tignat-Perrier (romie.tignat-perrier@univ-grenoble-alpes.fr)

12 **Abstract**

13
14 Microorganisms are ubiquitous in the atmosphere and some airborne microbial cells were
15 shown to be particularly resistant to atmospheric physical and chemical conditions (*e.g.*, UV
16 radiation, desiccation, presence of radicals). In addition to surviving, some cultivable
17 microorganisms of airborne origin were shown to be able to grow on atmospheric chemicals in
18 laboratory experiments. Metagenomic investigations have been used to identify specific
19 signatures of microbial functional potential in different ecosystems. We conducted a
20 comparative metagenomic study on the overall microbial functional potential and specific
21 metabolic and stress-related microbial functions of atmospheric microorganisms in order to
22 determine whether airborne microbial communities possess an atmosphere-specific functional
23 potential signature as compared to other ecosystems (*i.e.* soil, sediment, snow, feces, surface
24 seawater *etc.*). In absence of a specific atmospheric signature, the atmospheric samples
25 collected at nine sites around the world were similar to their underlying ecosystems. In addition,
26 atmospheric samples were characterized by a relatively high proportion of fungi. The higher
27 proportion of sequences annotated as genes involved in stress-related functions (*i.e.* functions
28 related to the response to desiccation, UV radiation, oxidative stress *etc.*) resulted in part from
29 the high concentrations of fungi that might resist and survive atmospheric physical stress better
30 than bacteria.

31
32 **Keywords:** atmospheric microorganisms, airborne microbial communities, planetary boundary
33 layer, metagenomic sequencing, comparative metagenomics, selective processes

34 **1 Introduction**

35
36 Microorganisms are ubiquitous in the atmosphere and reach concentrations of up to 10⁶
37 microbial cells per cubic meter of air (Tignat-Perrier et al., 2019). Due to their important roles
38 in public health and meteorological processes (Ariya et al., 2009; Aylor, 2003; Brown and
39 Hovmøller, 2002; Delort et al., 2010; Griffin, 2007), understanding how airborne microbial
40 communities are distributed over time and space is critical. While the concentration and
41 taxonomic diversity of airborne microbial communities in the planetary boundary layer have
42 recently been described (Els et al., 2019; Innocente et al., 2017; Tignat-Perrier et al., 2019), the
43 functional potential of airborne microbial communities remains unknown. Most studies have
44 focused on laboratory cultivation to identify possible metabolic functions of microbial strains
45 of atmospheric origin, mainly from cloud water (Amato et al., 2007; Ariya et al., 2002; Hill et
46 al., 2007; Vařtilingom et al., 2010, 2013). Given that cultivatable organisms represent about 1
47 % of the entire microbial community (Vartoukian et al., 2010), culture-independent techniques
48 and especially metagenomic studies applied to atmospheric microbiology have the potential to
49 provide additional information on the selection and genetic adaptation of airborne

50 microorganisms. However, to our knowledge, only five metagenomic studies on airborne
51 microbial communities at one or two specific sites per study exist (Aalismail et al., 2019; Amato
52 et al., 2019; Cao et al., 2014; Gusareva et al., 2019; Yooseph et al., 2013). Metagenomic
53 investigations of complex microbial communities in many ecosystems (for example, soil,
54 seawater, lakes, feces, sludge) have provided evidence that microorganism functional
55 signatures reflect the abiotic conditions of their environment, with different relative abundances
56 of specific microbial functional classes (Delmont et al., 2011; Li et al., 2019; Tringe et al., 2005;
57 Xie et al., 2011). This observed correlation of microbial community functional potential and
58 the physical and chemical characteristics of their environments could have resulted from genetic
59 modifications (microbial adaptation) (Brune et al., 2000; Hindré et al., 2012; Rey et al., 2016;
60 Yooseph et al., 2010) and/or physical selection. The latter refers to the death of sensitive cells
61 and the survival of resistant or previously adapted cells. This physical selection can occur when
62 microorganisms are exposed to physiologically adverse conditions.

63 The presence of a specific microbial functional signature in the atmosphere has not been
64 investigated yet. Microbial strains of airborne origin have been shown to survive and develop
65 under conditions typically found in cloud water (*i.e.* high concentrations of H₂O₂, typical cloud
66 carbonaceous sources, UV radiation *etc.*) (Amato et al., 2007; Joly et al., 2015; Vaitilingom et
67 al., 2013). While atmospheric chemicals might lead to some microbial adaptation, physical and
68 unfavorable conditions of the atmosphere such as UV radiation, low water content and cold
69 temperatures might select which microorganisms can survive in the atmosphere. From the pool
70 of microbial cells being aerosolized from Earth's surfaces, these adverse conditions might act
71 as a filter in selecting cells already resistant to unfavorable physical conditions. Fungal cells
72 and especially fungal spores might be particularly adapted to survive in the atmosphere due to
73 their innate resistance (Huang and Hull, 2017) and might behave differently than bacterial cells.
74 Still, the proportion and nature (*i.e.* fungi versus bacteria) of microbial cells that are resistant to
75 the harsh atmospheric conditions within airborne microbial communities are unknown.

76 Our objective was to determine whether airborne microorganisms in the planetary boundary
77 layer possess a specific functional signature as compared to other ecosystems since this might
78 indicate that microorganisms with specific functions tend to be more aerosolized and/or
79 undergo a higher survival in this environment. Our previous study showed that airborne
80 microbial taxonomy mainly depends on the underlying ecosystems, indicating that the local
81 environments are the main source of airborne microorganisms (Tignat-Perrier et al., 2019). Still,
82 we do not know if airborne microbial communities result from random or specific
83 aerosolization of the underlying ecosystems' microorganisms. We used a metagenomic
84 approach to compare the differences and similarities of both the overall functional potential and
85 specific microbial functions (metabolic and stress-related functions) between microbial
86 communities from the atmosphere and other ecosystems (soil, sediment, surface seawater, river
87 water, snow, human feces, phyllosphere and hydrothermal vent). We sampled airborne
88 microbial communities at nine different locations around the world during several weeks to get
89 a global-scale view and to capture the between and within-site variability in atmospheric
90 microbial functional potential.

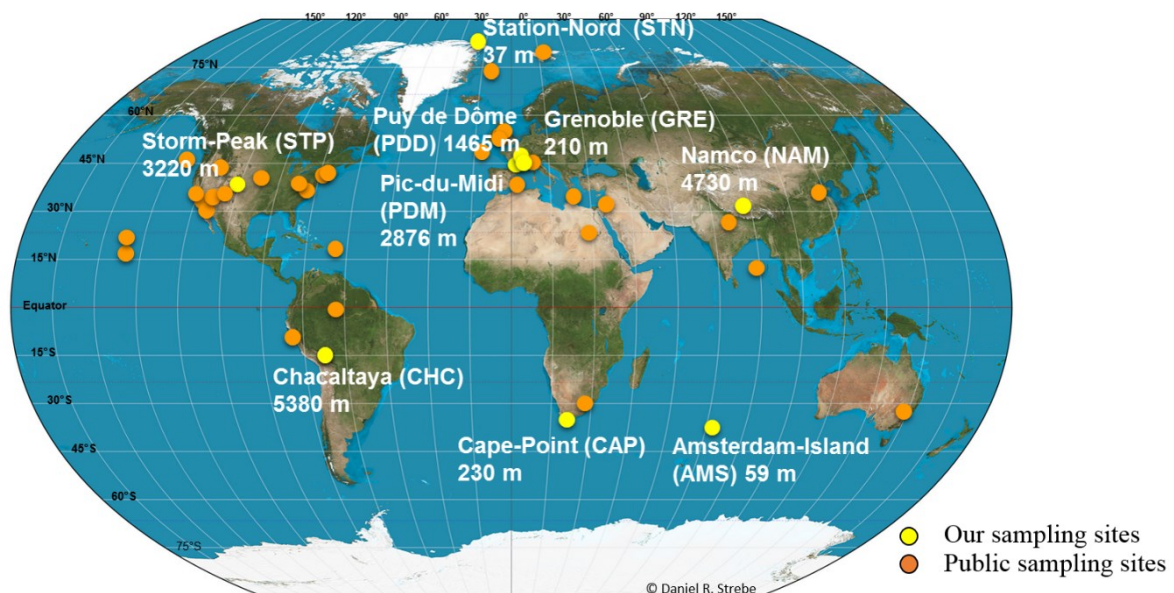
91

92 **2 Material and Methods**

93 **2.1 Sites and sampling**

94 Air samples were collected at nine sites in 2016 and 2017. Sites were characterized by different
95 latitudes (from the Arctic to the sub-Antarctica; **Fig 1**), elevations from sea level (from 59 m to
96 5230 m; **Fig 1**) and environment type (from marine for Amsterdam-Island or AMS, to coastal
97 for Cape Point or CAP, polar for Station Nord or STN and terrestrial for Grenoble or GRE,
98 Chacaltaya or CHC, puy de Dôme or PDD, Pic-du-Midi or PDM, Storm-Peak or STP and
99 Namco or NAM - **Table S1**). The number of samples collected per site varied from seven to

100 sixteen (**Table S1**). We collected particulate matter smaller than 10 μm (PM10) on quartz fiber
 101 filters (5.9'' round filter and 8'' \times 10'' rectangular types) using high volume air samplers
 102 installed on roof tops or terraces (roughly 10 m above ground level). A TISCH TE-5170V
 103 sampler, DIGITEL DA77/DA80 sampler, Chinese 2131 Laowin sampler or a custom made
 104 sampler (*i.e.* high-volume pump connected to a DIGITEL PM10 head and airflow meter) was
 105 used depending on the site (**Table S1**). To avoid contamination, quartz fiber filters as well as
 106 all the material in contact with the filters (*i.e.* filter holders, aluminium foils and plastic bags in
 107 which the filters were transported) were sterilized using strong heating (500 $^{\circ}\text{C}$ for 8 h) and UV
 108 radiation, respectively as detailed in Dommergue et al., 2019. The collection time per sample
 109 lasted one week, and the collected volumes ranged from 2000 m^3 to 10000 m^3 after
 110 standardization using SATP standards (Standard Ambient Pressure and Temperature). Detailed
 111 sampling protocols including negative control filters are presented in Dommergue et al. 2019.
 112 MODIS (Moderate resolution imaging spectroradiometer) land cover approach (5' x 5'
 113 resolution) (Friedl et al., 2002; Shannan et al., 2014) was used to quantify landscapes in the 50
 114 km diameter area of our nine sampling sites (**Fig S1**).
 115
 116



117
 118 **Fig 1. Sample collection locations.** Map showing the geographical location and elevation from
 119 sea level of our nine sampling sites (in yellow), and the geographical position of whose public
 120 metagenomes come from (in orange). Abbreviations of our nine sampling sites are indicated in
 121 brackets.

122
 123 **2.2 Molecular biology analyses**
 124 **2.2.1 DNA extraction**

125 DNA was extracted from three circular pieces (punches) from the quartz fiber filters (diameter
 126 of one punch: 38 mm) using the DNeasy PowerWater kit with some modifications as detailed
 127 in Dommergue et al., 2019. During cell lysis, the PowerBead tube containing the three punches
 128 and the pre-heated lysis solution were heated at 65 $^{\circ}\text{C}$ during one hour after a 10-min vortex
 129 treatment at maximum speed. We then separated the filter debris from the lysate by
 130 centrifugation at 1000 rcf for 4 min. From this step on, we followed the DNeasy PowerWater
 131 protocol. We conducted additional extractions on French agricultural soil samples collected at
 132 the Côte Saint André (that is part of the sample collection locations). We used 250 mg of soil

133 on which the same DNA extraction methodology as for air samples was applied. DNA
134 concentration eluted in 100 μ L of buffer was measured using the Qubit Fluorometric
135 Quantification kit (Thermo Fisher Scientific). DNA was stored at -20 $^{\circ}$ C.
136

137 **2.2.2 16S and 18rRNA gene qPCR analyses**

138 The 16S and 18SS rRNA gene copy numbers were calculated per cubic meter of air (for air
139 samples) and per gram of soil (for soil samples). Standards, primers and methodology are
140 presented in Tignat-Perrier et al., 2019.
141

142 **2.2.3 MiSeq Illumina metagenomic sequencing**

143 **Metagenomic library preparation.** Metagenomic libraries were prepared from 1 ng of DNA
144 using the Nextera XT Library Prep Kit and indexes following the protocol in Illumina's
145 "Nextera XT DNA Library Prep Kit" reference guide with some modifications for samples with
146 DNA concentrations below 1 ng as follows. The tagmented DNA was amplified over 13 PCR
147 cycles instead of 12 PCR cycles, and the libraries (after indexing) were resuspended in 30 μ L
148 of RBS buffer instead of 52.5 μ L. Metagenomic sequencing was performed using the MiSeq
149 and V2 technology of Illumina with 2 x 250 cycles. At the end of the sequencing, the adapter
150 sequences were removed by internal Illumina software.

151 **Reads quality filtering.** Reads 1 and reads 2 per sample were not paired but merged in a
152 common file before filtering them based on read quality using the tool FASTX-Toolkit
153 (http://hannonlab.cshl.edu/fastx_toolkit/) using a minimum read quality of Q20, minimum read
154 length of 120 bp and one maximum number of N per read. Samples with less than 6000 filtered
155 sequences were removed from the dataset.
156

157 **2.2.4 Downloading of public metagenomes**

158 Public metagenomes were downloaded from the MGRAST (Keegan et al., 2016) and SRA
159 (Leinonen et al., 2011; NCBI Resource Coordinators, 2018) databases as quality filtered read-
160 containing fasta files and raw read containing fastq files, respectively. The fastq files containing
161 raw reads underwent the same quality filtering as our metagenomes (as discussed above). The
162 list of the metagenomes, type of ecosystem, number of sequences and sequencing technology
163 (*i.e.* MiSeq, HiSeq or 454) are summarized in **Table S2**. The sampling sites are positioned on
164 the map in **Fig 1**.
165

166 **2.3 Data analyses**

167 All graphical and multivariate statistical analyses were carried out using the vegan (Oksanen et
168 al., 2019), ggplot2 (Hadley and Winston, 2019) and reshape2 (Wickham, 2017) packages in the
169 R environment (version 3.5.1).
170

171 **2.3.1 Annotation of the metagenomic reads**

172 Firstly, to access the overall functional potential of each sample, all the filtered sequences per
173 sample were functionally annotated using DIAMOND (Buchfink et al., 2015) and the nr
174 database, then the gene-annotated sequences were grouped in the different SEED functional
175 classes (around 7000 functional classes, referred simply to as functions) using MEGAN6
176 (Huson et al., 2016). Functional classes that were present ≤ 2 times in a sample were removed
177 of this sample. In parallel, the Kraken software (Wood and Salzberg, 2014) was used to retrieve
178 the bacterial and fungal sequences separately from the filtered sequences using the Kraken
179 bacterial database and FindFungi (Donovan et al., 2018) fungal database (both databases
180 included complete genomes), respectively (and using two different runs of Kraken). Separately,
181 both the bacterial and fungal sequences were also functionally annotated using DIAMOND and
182 MEGAN6 (number of sequences functionally annotated in **Table S3**).

183 Secondly, for specific metabolic and stress-related functions, we annotated the sequences using
184 eggNOG-Mapper version 1 (DIAMOND option; Huerta-Cepas et al., 2017), then examined
185 specific GO (Gene Ontology; Gene Ontology Consortium, 2019) terms chosen based on their
186 importance for microbial resistance to atmospheric-like conditions. The different GO terms
187 used were the following: GO:0042744 (hydrogen peroxide catabolic activity), GO:0015049
188 (methane monooxygenase activity) as specific metabolic functions and GO:0043934
189 (sporulation), GO:0009650 (response to UV), GO:0034599 (cell response to oxidative stress),
190 GO:0009269 (response to desiccation) as stress-related functions. The number of hits of each
191 GO term was normalized per 10000 annotated sequences and calculated from all sequences,
192 bacterial sequences and fungal sequences for each sample. The ratio was given per 10000
193 annotated sequences and not 100 annotated sequences to get hit numbers superior to 1. The
194 number of sequences annotated by eggNOG-Mapper was also evaluated (**Table S3**). The
195 putative concentration of a specific function or functional class in the samples is determined as
196 the concentration of sequences annotated as one of the functional proteins associated to this
197 function (or functional class).

198

199 **2.3.2 Statistical analyses**

200 Observed functional richness and evenness were calculated per sample after rarefaction on all
201 sequences (rarefaction at 2000 sequences), bacterial sequences (rarefaction at 500 sequences)
202 and fungal sequences (rarefaction at 500 sequences). Rarefaction was used only for the purpose
203 of calculating the diversity metrics (richness and evenness). For the following analyses, no
204 rarefaction was applied on the number of reads per sample. The distribution of the samples was
205 analyzed based on the SEED functional classes (using all sequences). PCoA and hierarchical
206 clustering analysis (average method) were carried out on the Bray-Curtis dissimilarity matrix
207 based on the relative abundances of the different SEED functional classes. SIMPER analyses
208 were used to identify the functions responsible for the clustering of samples in groups. Because
209 of the non-normality of the data, Kruskal-Wallis analyses (non-parametric version of ANOVA)
210 and Dunn's post-hoc tests were used to test the difference between the percentage of fungal
211 sequences as well as the number of hits of each Gene Ontology term (normalized per 10000
212 annotated sequences) among the different sites and the different ecosystems.

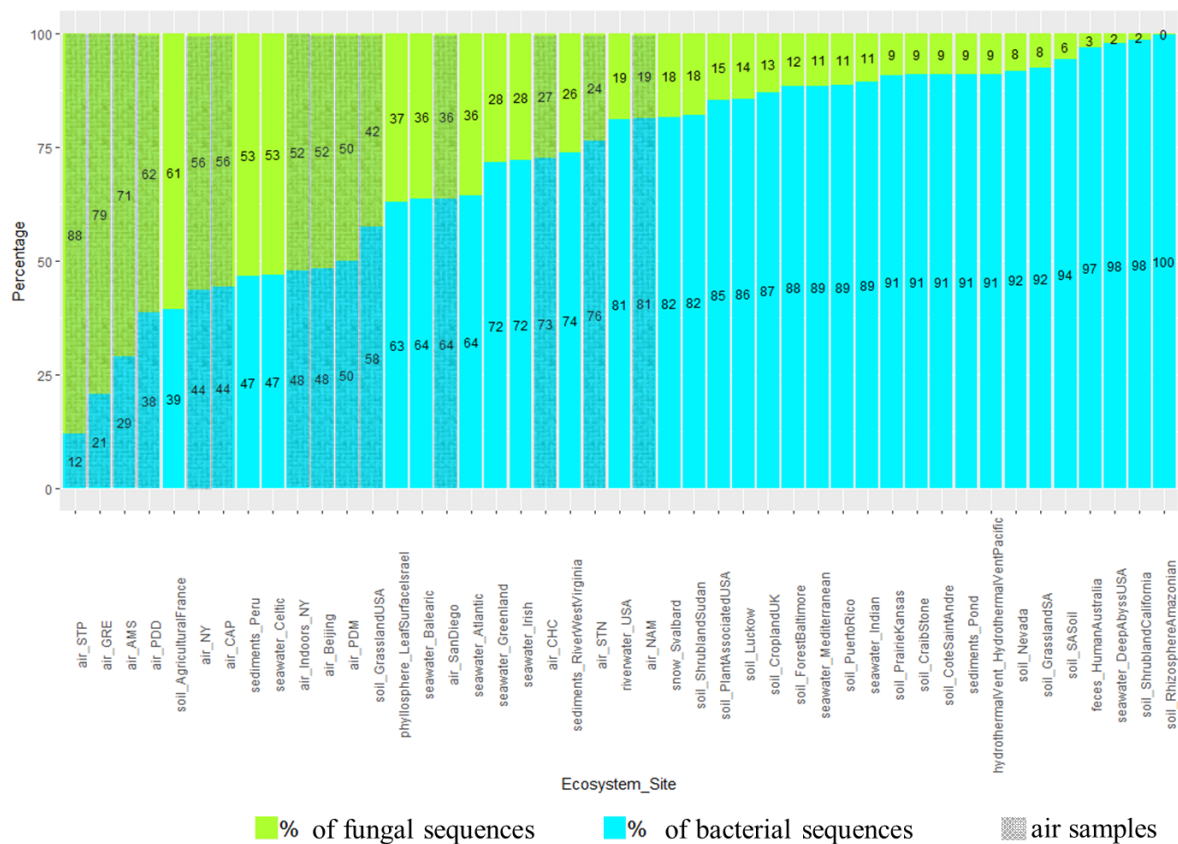
213

214 **3 Results**

215 **3.1 Percentage of fungal sequences**

216 The percentage of sequences annotated as belonging to fungal genomes (or fungal sequences,
217 as opposed to bacterial sequences) was on average higher in air samples compared to soil ($P < 10^{-5}$),
218 snow ($P = 10^{-3}$), seawater ($P = 0.03$) and sediment samples ($P = 10^{-3}$; **Fig 2** and **Table S4**).
219 Among the air samples, NAM (19%), STN (24%) and CHC (27%) showed the lowest
220 percentages of fungal sequences on average while STP (88%), GRE (79%), AMS (71%) and
221 PDD (62%) showed the highest percentages. For the ecosystems that were only represented by
222 one sample, and therefore, were not evaluated by the Kruskal-Wallis test, we observed average
223 percentages of fungal reads of 3% in feces, 9% in hydrothermal vents, 19% in river water
224 samples and 37% in the phyllosphere. Some samples from soil, sediments and seawater such as
225 French agricultural soil (61%), Peru sediments (53%) and Celtic seawater (53%) had relatively
226 high percentages of fungal sequences while other samples had less than 50%. The approximated
227 number of fungal and bacterial cells in air and soil was also estimated using 16S rRNA and 18S
228 rRNA gene copy numbers per cubic meter of air and gram of soil, respectively. Air samples
229 showed ratios between 16S and 18S rRNA gene copy numbers from around 4.5 times up to 160
230 times lower than soil samples (**Table S4**; some qPCR data have already been published in
231 Tignat-Perrier et al., 2020). The qPCR data were used to see if similar results, or the same trend,

232 on the estimated ratio between fungi and bacteria in air compared to the planetary-bound
 233 ecosystems could be obtained from metagenomic data and qPCR.
 234

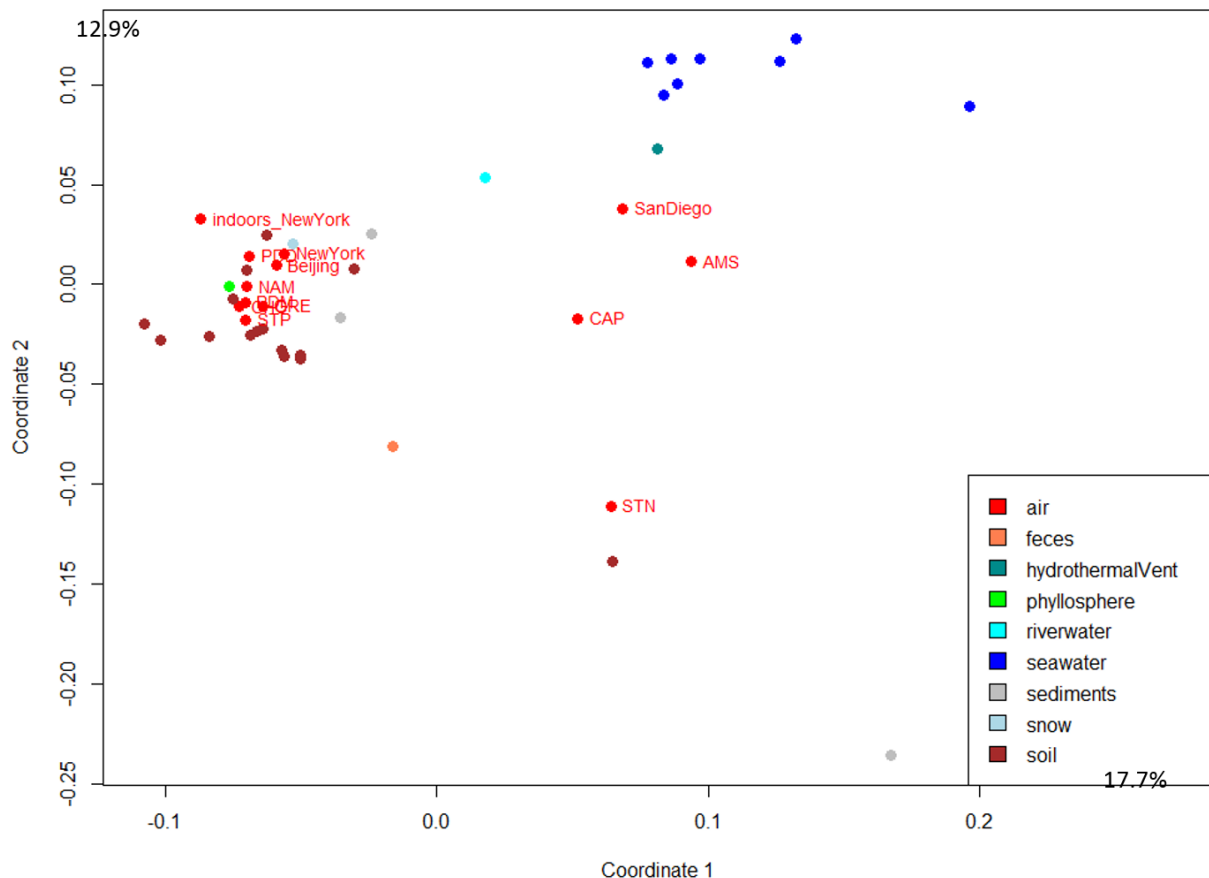


235
 236 **Fig 2. Percentage of fungal and bacterial sequences in the metagenomes.** The percentages
 237 are established as the number of sequences annotated as belonging to fungal and bacterial
 238 genomes over the sum of bacterial and fungal sequences in the metagenomes. The mean was
 239 calculated for the sampling sites including several metagenomes. Air sites (*i.e.* our 9 sites + 5
 240 sites where public air metagenomes come from) are distinguished by grey hatching lines.

241
 242 **3.2 Airborne microbial functional profiles**

243 The fifty most abundant SEED functional classes represented in atmospheric samples are listed
 244 in **Table S5**. The 5-FCL-like protein, the long chain fatty acid CoA ligase and the TonB-
 245 dependent receptor were the top three functions based on number of annotated reads observed
 246 when including all the sequences (**Table S5**). The atmospheric microbial functional profiles
 247 based on the SEED functions were compared between samples from the different weeks of
 248 sampling and between different locations. The profiles were graphed using PCo multivariate
 249 analysis to visualize differences and similarities. The different samples (sampled during
 250 sequential weeks) from the same site did not cluster tightly together on the PCo multivariate
 251 analysis. In order to incorporate weekly variation when comparing sites, we used the microbial
 252 functional profile averaged per site in the subsequent multivariate analyses done with the data
 253 from other ecosystems (**Fig 3**). The PCo multivariate analysis showed that terrestrial
 254 atmospheric sites (GRE, NAM, STP, PDD, PDM, CHC, New York) grouped with the soil,
 255 sediment and snow samples while the marine and coastal atmospheric sites (AMS, CAP, San
 256 Diego) were situated between the datasets from soil, seawater and river water (**Fig 3**). The polar
 257 site STN did not group with the other sites. When considering only the bacterial sequences (*i.e.*,

258 excluding the fungal sequences), the distribution of the terrestrial atmospheric sites did not
 259 change, while the marine Amsterdam-Island, coastal Cape Point and polar Station Nord
 260 atmospheric sites were further from the seawater and river water datasets than when the fungal
 261 sequences were included (**Fig S2**). The distribution of the different datasets underwent further
 262 changes when considering only the fungal sequences. We observed an absence of a clear
 263 separation between soil and seawater since they (for the majority) grouped closely together, and
 264 terrestrial atmospheric datasets did not group with the other non-atmospheric datasets from soil,
 265 sediment and snow (**Fig S2**).
 266



267 **Fig 3. Distribution of the samples based on the microbial functional profile.** The PCo
 268 analysis of the Bray-Curtis dissimilarity matrix is based on the functional potential structure of
 269 each site. For the site including several metagenomes, the average profile was calculated. Colors
 270 indicate the ecosystems in which the sites belong to. The percentages (17.7 % and 12.9 %)
 271 indicate the part of the variance explained by the first and second axis, respectively.
 272

273
 274 **3.3 Airborne microbial functional richness and evenness**

275 Functional richness and evenness were evaluated using the relative abundance of sequences in
 276 the different SEED categories. The average richness in SEED functional classes (or functions)
 277 in the PBL was lower than the average functional class richness in soil, surface seawater,
 278 hydrothermal vents, river water, phyllosphere and feces ($P < 0.05$) (**Table S3**). Among the
 279 different atmospheric samples, the functional class richness was highest in Beijing (4060 +/-
 280 112 functional classes) and New York indoor air samples (3302 +/- 299 functional classes)
 281 ($P < 0.05$), and lowest in Station Nord (956 +/- 547 functional classes). When looking at the
 282 bacteria-annotated sequences, almost the same trend was observed, *i.e.* the functional class

283 richness in air was lower than in soil, hydrothermal vents, river water, phyllosphere and feces,
284 and not different from the other ecosystems ($P < 0.05$ and > 0.05 , respectively) (**Table S3**). The
285 functional class richness was higher in Beijing (2835 +/- 59 functional classes) and New York
286 indoor air samples (2183 +/- 387 functional classes) compared to the other air samples whose
287 values ranged between 270 +/- 197 functional classes in Amsterdam-Island and 1142 +/- 461
288 functional classes in Chacaltaya. For fungal sequences, the functional class richness in the
289 atmosphere was lower than the functional class richness in soil, surface seawater, feces,
290 hydrothermal vents, river water and phyllosphere ($P < 0.05$) (**Table S3**). Within air samples, the
291 functional class richness based on fungal sequences was higher in Beijing (1129 +/- 92
292 functional classes) and New York indoor air samples (687 +/- 206 functional classes) than in
293 the other air sites ($P < 10^{-5}$) whose values ranged from 66 +/- 58 functional classes in
294 Amsterdam-Island and 392 +/- 131 functional classes in Storm Peak (**Table S3**). The functional
295 class evenness in air was on average higher than in soil ($P = 0.03$), and not different to the
296 functional class evenness observed in the other ecosystems (sediment, seawater, snow). When
297 looking at the bacterial and fungal sequences separately, the functional class evenness in air
298 was on average higher than in soil, feces, phyllosphere and riverwater ($P < 0.05$) (**Table S3**).
299

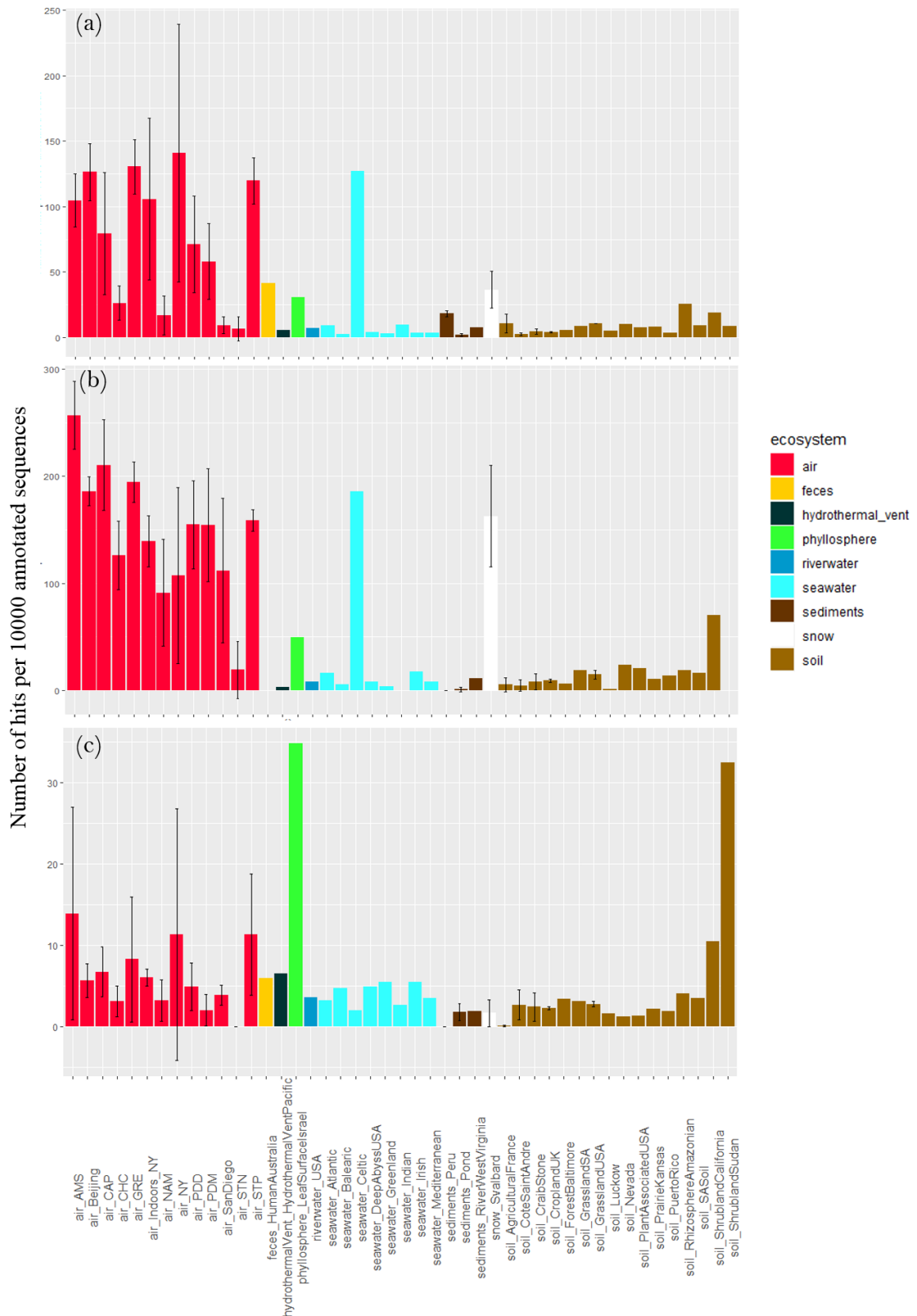
300 **3.4 Concentration of specific microbial functions that might have a role under** 301 **atmospheric conditions**

302 Two metabolic functions associated with abundant atmospheric chemicals (H_2O_2 and CH_4)
303 were examined, hydrogen catabolism and methane monooxygenase activity. The concentration
304 of sequences annotated as hydrogen peroxide catabolic related functional proteins per 10000
305 sequences varied between air sites ($P = 2 \times 10^{-5}$) with highest values for Amsterdam-Island (27
306 +/- 1) and Grenoble (27 +/- 1) (**Fig S3**). It was on average higher in air compared to soil ($P = 10^{-4}$)
307 and surface seawater ($P = 10^{-4}$). The French agricultural soil showed the highest relative
308 abundance (133 +/- 4). When considering the fungal and bacterial sequences separately, this
309 concentration was not different between air and the other ecosystems ($P > 0.05$) (**Fig S3**). The
310 number of sequences annotated as methane monooxygenase-related functional proteins per
311 10000 sequences was only detectable when considering all the sequences (*i.e.* bacterial and
312 fungal sequences). The number of sequences annotated as methane monooxygenase-related
313 functional proteins did not vary between air sites ($P > 0.05$) while we observed a high variability
314 between sampling periods within sites, but on average it was not different from the ecosystems
315 ($P > 0.05$).

316 Different stress response functions (sporulation, UV response, oxidative stress cell response,
317 desiccation response, chromosome plasmid partitioning protein ParA and lipoate synthase)
318 were examined. The concentration of sequences annotated as sporulation-related functional
319 proteins per 10000 annotated sequences largely varied between air sites ($P = 2 \times 10^{-9}$), with the
320 lowest values observed for Station Nord (7 +/- 9), San Diego (9 +/- 6), Namco (17 +/- 15) and
321 Chacaltaya (26 +/- 13), and the highest values observed for Storm Peak (120 +/- 18), Beijing
322 (126 +/- 22), Grenoble (131 +/- 21) and New York (141 +/- 98) (**Fig 4**). It was on average higher
323 in air compared to soil ($P < 10^{-5}$), sediments ($P < 10^{-5}$) and surface seawater ($P = 4 \times 10^{-4}$) although
324 the Celtic seawater sample presented a very high concentration (127). Snow showed a relatively
325 high average concentration (*i.e.* 36) which was not different from air concentration ($P > 0.05$).
326 For the ecosystems including one value (*i.e.* one sample, so not integrated in the Kruskal-Wallis
327 tests), feces showed a relatively high concentration of sequences annotated as sporulation-
328 related functional proteins (*i.e.* 41) while hydrothermal vent, phyllosphere and river water
329 showed relatively low concentrations compared to air (< 10). When considering the fungal
330 sequences separately from the bacterial sequences, the same trend was observed, *i.e.* the
331 concentration of sequences annotated as sporulation-related functional proteins in air was on
332 average higher compared to soil ($P < 10^{-5}$), sediments ($P < 10^{-5}$), surface seawater ($P = 7 \times 10^{-4}$) as

333 well as phyllosphere, hydrothermal vent and river water. The concentration was relatively high
334 in the Celtic seawater (186) and the snow samples (163 +/- 47). We also observed a large
335 variability within air sites ($P=3\times 10^{-5}$). When considering the bacterial sequences only, this
336 concentration in air was on average higher compared to soil ($P=0.02$), sediments ($P=4\times 10^{-3}$)
337 and snow ($P=0.01$), and showed a smaller variability between air sites. Two samples, the
338 phyllosphere (*i.e.* 35) and the shrubland soil from Sudan (*i.e.* 32) showed high numbers of
339 sequences annotated as sporulation-related functional proteins per 10000 annotated sequences
340 (**Fig 4**).

341



342
343
344
345

Fig 4. Proportion of sequences annotated as sporulation related functional proteins in the metagenomes. Average number of sequences annotated as proteins implicated in sporulation per 10000 annotated sequences from (a) all sequences, (b) fungal sequences and (c) bacterial

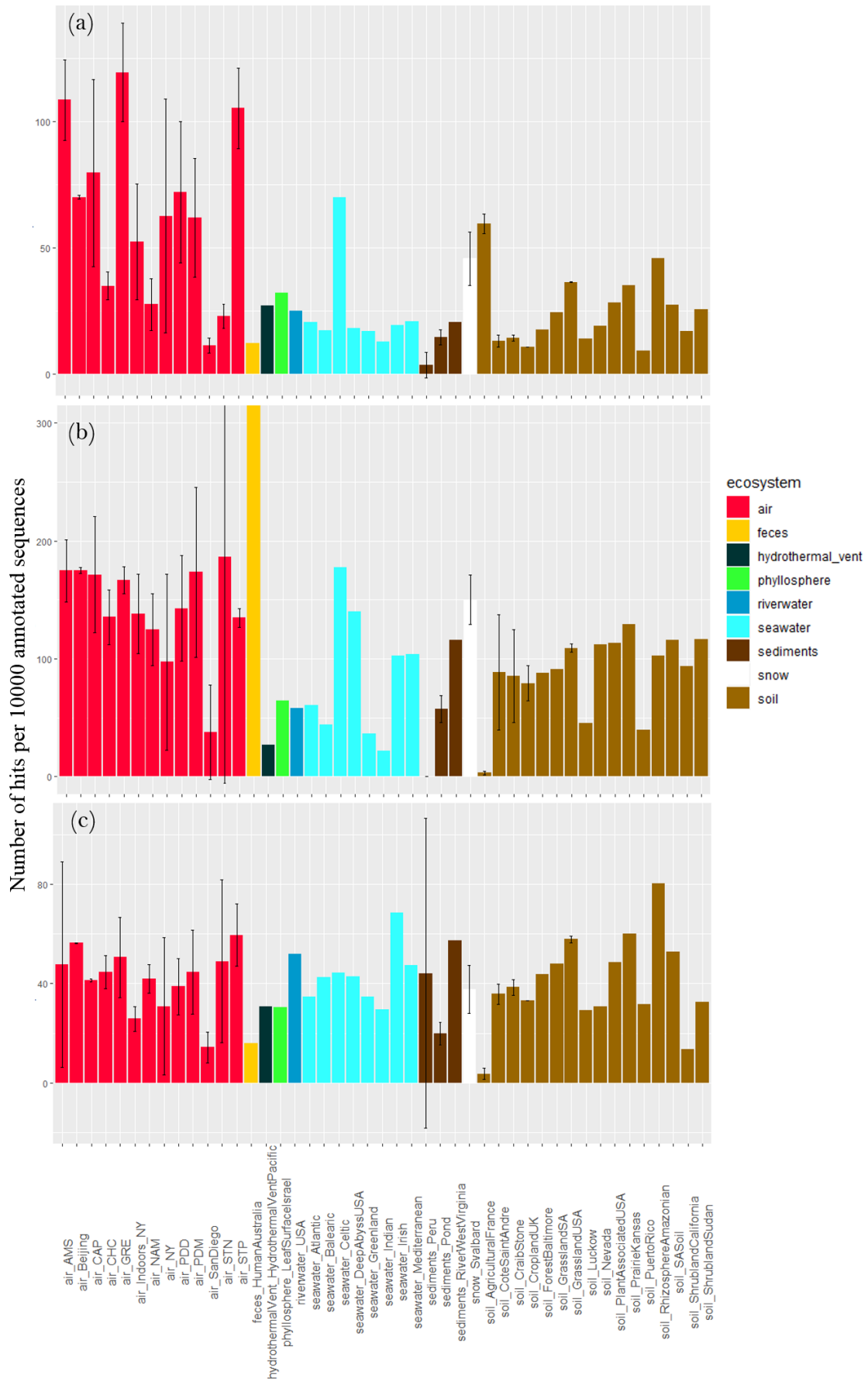
346 sequences per site. Colors indicate the ecosystems in which the sites belong to. For the sites
347 including several metagenomes, the standard deviation was added.

348

349 The concentration of sequences annotated as UV response related functional proteins per 10000
350 annotated sequences varied between air sites ($P=10^{-5}$), with values ranging from 16 +/- 2 in
351 Namco and 19 +/- 4 in STN to 29 +/- 3 in Storm Peak and 36 +/- 6 in Amsterdam-Island (**Fig**
352 **S4**). The concentration was on average higher in air compared to sediments ($P<10^{-5}$), soil
353 ($P<10^{-5}$) and comparable to snow and surface seawater ($P>0.05$). The other ecosystems showed
354 lower ratios (feces, phyllosphere) or comparable concentrations (hydrothermal vent, river
355 water) compared to air. Within the soil samples, the French agricultural soil samples showed a
356 high average concentration (56 +/- 8), which increased the average ratio observed in soil
357 samples. When considering fungal sequences separately, the concentration of sequences
358 annotated as UV response related functional proteins was higher in air compared to soil
359 ($P=9\times 10^{-4}$), and comparable to the other ecosystems ($P>0.05$). When considering the bacterial
360 sequences only, this concentration in air was on average higher compared to seawater ($P=3\times 10^{-3}$)
361 and sediments ($P=6\times 10^{-3}$).

362 The concentration of sequences annotated as oxidative stress cell response related functional
363 proteins per 10000 annotated sequences varied largely between air sites ($P=5\times 10^{-7}$), with the
364 lowest values observed for Station Nord (23 +/- 5), San Diego (11 +/- 3) and Namco (28 +/-
365 10), and the highest values observed for Storm Peak (105 +/- 16), Amsterdam-Island (108 +/-
366 16) and Grenoble (119 +/- 19) (**Fig 5**). The concentration was on average higher in air compared
367 to soil ($P<10^{-5}$), sediments ($P<10^{-5}$) and surface seawater ($P=2\times 10^{-3}$). Snow showed a relatively
368 high average value (46 +/- 11), not different from air ($P>0.05$). The other ecosystems (feces,
369 river water, hydrothermal vent, phyllosphere) showed lower ratios compared to air. When
370 considering fungal sequences separately, the concentration of sequences annotated as oxidative
371 stress related functional proteins per 10000 sequences was on average higher in air compared
372 to soil ($P<10^{-5}$), sediments ($P<10^{-5}$) and surface seawater ($P=10^{-3}$). Feces showed a very high
373 average value (2237). When considering bacterial sequences separately, this concentration was
374 not different between air and the other ecosystems ($P>0.05$). When considering both fungal and
375 bacterial sequences separately, the variability in the concentration of sequences annotated as
376 oxidative stress cell response related functional proteins between air sites diminished and their
377 difference was not detected anymore ($P>0.05$).

378



380 **Fig 5. Proportion of sequences as oxidative stress cell response related functional proteins**
381 **in the metagenomes.** Average number of sequences annotated as proteins implicated in
382 oxidative stress cell response per 10000 annotated sequences from (a) all sequences, (b) fungal
383 sequences and (c) bacterial sequences per site. Colors indicate the ecosystems in which the sites
384 belong to. For the sites including several metagenomes, the standard deviation was added.

385

386

387 The concentration of sequences annotated as desiccation response related functional proteins
388 per 10000 sequences varied between air sites ($P=2\times 10^{-5}$), with the highest values in Grenoble
389 (4 +/- 1), Storm Peak (4 +/- 1) and Amsterdam-Island (3 +/- 3), and the lowest values in Station
390 Nord (0.5 +/- 1) and San Diego (0.1 +/- 0.1) (**Fig S4**). It was on average higher in air compared
391 to the other ecosystems ($P=4\times 10^{-9}$). Still Svalbard snow and French agricultural soil showed
392 high values (2 +/- 1 and 3 +/- 1, respectively) (**Fig S4**). When considering fungal sequences
393 only, the concentration in air was higher compared to soil ($P>10^{-5}$), sediments ($P>10^{-5}$) and
394 surface seawater ($P=10^{-3}$). No difference between the ecosystems was observed when
395 considering bacterial sequences separately ($P=0.62$).

396 Two proteins (lipoate synthase and chromosome plasmid partitioning protein ParA) related to
397 stress response showed high relative concentrations in bacterial sequences of a few air samples
398 compared to the other ecosystems (**Fig S3**), although the number of sequences related to these
399 proteins was on average not higher in the atmosphere than other ecosystems ($P>0.05$).

400

401 **4 Discussion**

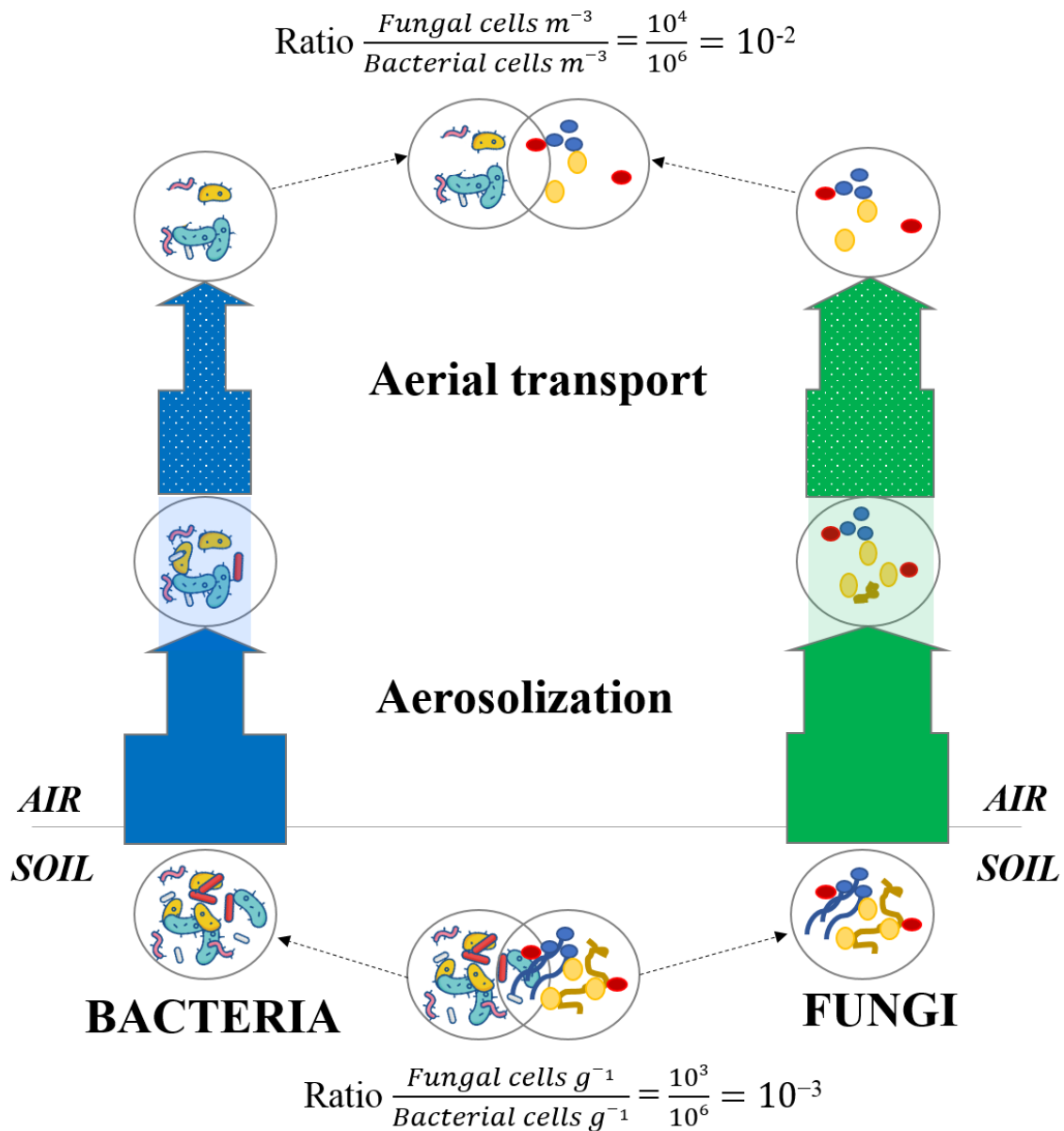
402 Metagenomic investigations of different ecosystems revealed a specific functional potential
403 signature of their associated microbial communities (Delmont et al., 2011; Tringe et al., 2005).
404 These specific signatures are thought to result from microbial adaptation and/or physical
405 selection to the environmental abiotic conditions (Hindré et al., 2012; Li et al., 2019; Rey et al.,
406 2016) and are a reflection of the high relative abundances of genes coding for specific functions
407 essential for microorganisms to survive and develop in these environments. For example,
408 microbial metagenomes of human feces were characterized by high relative abundances of
409 sequences annotated as beta-glucosidases that are associated with high intestinal concentrations
410 of complex glycosides; and microbial metagenomes of oceans were enriched in sequences
411 annotated as enzymes catalyzing DMSP (dimethylsulfoniopropionate), that is an organosulfur
412 compound produced by phytoplankton (Delmont et al., 2011). Our results showed a clear
413 separation between surface seawater, river water, human feces and almost all the soil samples
414 (which grouped with the sediment and snow samples at the scale used here) on the PCo analysis
415 based on the microbial functional potential (**Fig 3**). For air microbiomes, the PCo analyses
416 showed that the individual air samples did not group for each site and that they did not form a
417 cluster separated from the other ecosystems based on the overall microbial functional potential
418 averaged per site (**Fig 3**). Air samples seemed to group with their underlying ecosystems (**Fig**
419 **S1**). While terrestrial air samples (GRE, NAM, CHC, STP, PDD, PDM) grouped with snow,
420 soil and sediment samples, the marine (Amsterdam-Island), coastal (Cape Point) and arctic
421 (Station Nord) air samples were closer to surface seawater and river water samples. Airborne
422 microbial functional potential (and especially metabolic functional potential as SEED
423 functional classes included mainly metabolic functions and few stress response related
424 functions) might be dependent on the ecosystems from which microorganisms are aerosolized.
425 Moreover, it seems that bacterial sequences are mainly responsible for the distribution of the
426 samples on the PCo analysis (as observed when comparing the PCoA to that carried out with
427 the fungal sequences only) although they were in smaller numbers compared to fungal
428 sequences for many of the air samples (*i.e.* STP, GRE, AMS, PDD, CAP, Beijing *etc.*). The
429 low statistical weight of fungal sequences relative to the overall sequences might be related to

430 their low richness in terms of functional genes that might have resulted in the spreading of the
431 samples on the PCoA based on the fungal sequences (**Table S3**).

432 Metagenomes extracted from atmospheric samples taken around the planet were characterized
433 by a relatively high percentage of fungal sequences as compared to other ecosystems even
434 though bacterial sequences still dominated. This percentage varied across the different sites
435 with a higher percentage at terrestrial sites whose surrounding landscapes were vegetated like
436 Grenoble (GRE), puy de Dôme (PDD) and Pic-du-midi (PDM) (surrounding landscapes in **Fig**
437 **S1**). This percentage was also relatively high at the marine site Amsterdam-Island (AMS),
438 where fungi might come from the ocean and/or the vegetated surfaces of the small island. A
439 high percentage of fungal sequences was also reported for air samples from Beijing, New York
440 and San Diego and validates our DNA extraction method set-up specifically for quartz fiber
441 filter (Dommergue et al., 2019). Similarly, the sequencing technology (Illumina MiSeq) could
442 not have been responsible for the larger percentage of fungal sequences observed in our datasets
443 as the Beijing and New York/San Diego air sample datasets originated from Illumina HiSeq
444 and 454 sequencing technology, respectively. qPCR results on the 16S rRNA gene (bacterial
445 cell concentration estimation) and on the 18S rRNA gene (fungal cell concentration estimation)
446 on our air samples in comparison to agricultural soil samples evidenced that the ratio between
447 fungal and bacterial cell number might be much higher (from 4.5 to 160 times higher for the
448 most vegetated site Grenoble) in air than in soil (**Table S4**; see Tignat-Perrier et al., 2020 for
449 more qPCR data on air samples). The ratio between fungal and bacterial cell number might be
450 higher in the planetary boundary layer (PBL) than in other environments like soil (Malik et al.,
451 2016), and thus, would explain the relatively higher percentage of fungal sequences observed
452 in air metagenomes. High throughput sequencing allows the sequencing of a small part of the
453 metagenomic DNA (with large fungal genomes likely to be sequenced first) and might explain
454 why the values of the ratio between 16S and 18S rRNA gene copy numbers obtained by qPCR
455 does not match exactly those obtained by the metagenomic sequencing approach, while they
456 show the same trend. Our study is one of the very few metagenomic investigations of the air
457 environment existing so far. Further studies are needed to confirm our results based on a limited
458 number of sequences per sample.

459 Fungi in the atmosphere are expected to be found mostly as fungal spores. While some spore
460 and hyphae concentrations have been measured in air (Després et al., 2012), the relative
461 concentration of fungal spores and fungal hyphae fragments and its temporal dynamics at the
462 same site remains unknown. Our results showed that the number of sporulation-related
463 functions was higher in air than the other ecosystems (with the exception of snow and
464 phyllosphere). While fungal hyphae are not expected to be particularly resistant to extreme
465 conditions such as UV radiation, fungal spores are specifically produced to resist and survive
466 overall adverse atmospheric conditions (Huang and Hull, 2017). Their thick membrane and
467 dehydrated nature make them particularly resistant to abiotic atmospheric conditions such as
468 UV radiation, oxidative stress, desiccation as well as osmotic stress. **Fig 6** presents a conceptual
469 model that could explain the higher ratio between fungi and bacteria observed in air. During
470 aerosolization and aerial transport, bacteria and fungi might be under stress and might undergo
471 a physical selection with the survival of the most resistant cells to the adverse atmospheric
472 conditions (*i.e.* UV radiation, desiccation *etc.*) and the death of non-resistant cells. As fungi
473 (and especially fungal spores) might be naturally more resistant and adapted to atmospheric
474 conditions than bacteria, we expect a larger decline of bacterial cells compared to fungal cells
475 and spores in air. This might have as a consequence an increase in the ratio between fungi and
476 bacteria compared to their non-atmospheric origins (*i.e.* the surrounding ecosystems) (**Fig 6**).

477



478
 479
 480
 481
 482
 483
 484
 485
 486
 487

Fig 6. Microbial cell loss due to atmospheric physical stress. Conceptual model on the microbial cell loss occurring during the aerosolization and aerial transport steps due to physical selection. The thickness of the arrows represents the impact of the physical selection on both bacterial and fungal cell loss (the more microbial cells survive the physical selection, the thicker becomes the arrow). Approximate ratios are indicative and result from 16S rRNA and 18S rRNA gene qPCR data on Côte Saint André soil samples (crop soil, France) and puy de Dôme air samples (France; puy de Dôme landscape is mainly composed of croplands as shown in Fig S1).

488
 489
 490
 491
 492
 493

The high variability between the air sites and between air samples of the same site could be explained by the variability in the inputs from the different surrounding landscapes. Our previous paper showed that local inputs were the main sources of planetary boundary layer microorganisms and that local meteorology (especially the wind direction) had a major impact on the temporal variability of airborne microbial communities by affecting which of the

494 different local sources were upwind (Tignat-Perrier et al., 2019). Our results did not show a
495 specific (metabolic) functional potential signature for the atmosphere, which was rather mainly
496 driven by the surrounding landscapes. Our results are consistent with both a pre-metabolic
497 adaptation of airborne microorganisms to the chemicals of the sources (*i.e.* surrounding
498 landscapes) and a potential metabolic adaptation to these chemicals in the atmosphere.
499 Atmospheric chemistry is dependent on the underlying ecosystem chemistry since the main
500 sources of atmospheric chemicals are Earth surface emissions. Yet, the oxidizing conditions of
501 the atmosphere might lead to rapid transformations of atmospheric chemicals by photochemical
502 reactions. These specific atmospheric chemical reactions (*i.e.* photochemical) produce species
503 which, with the gases like CH₄, characterize the atmosphere (O₃, H₂O₂, OH *etc.*). Although
504 some microbial strains from cloud water origin have been shown to metabolize and grow on
505 culture medium in the presence of H₂O₂ (Vařtilingom et al., 2013), radical species and their
506 precursors are reactive compounds and might not easily serve as energy and carbon sources for
507 microorganisms (Imlay, 2013). Our results on specific metabolic related functions showed that
508 functions related to methane monooxygenase activity (CH₄ degradation) and hydrogen
509 peroxide catabolism (H₂O₂ degradation) were present in air but not in higher proportion than in
510 other ecosystems (**Fig S3**). Reactive compounds can cause oxidative stress to airborne
511 microorganisms. In association to adverse physical conditions like UV radiation and
512 desiccation, oxidative compounds might create more of a physical stress than provide a new
513 metabolic source for airborne microorganisms. Laboratory investigations of cultivable
514 microorganisms of an airborne origin showed the presence of particularly resistant strains under
515 stressful conditions similar to the atmospheric ones (*i.e.* similar UV radiation levels; different
516 oxidative conditions) (Joly et al., 2015; Yang et al., 2008). However, no study has shown
517 whether these apparently adapted cells represented the majority of airborne microorganisms.
518 Since the overall SEED functional classes included mainly metabolic functions, specific stress
519 related functions using GO (Gene Ontology) terms were also evaluated. We observed that on
520 average, air showed more stress-related functions (UV response, desiccation and oxidative
521 stress response related functions) than the other ecosystems due to the higher concentration of
522 fungi (relatively to bacteria) in air. Thus, when the annotated sequences were separated between
523 sequences belonging to fungal and bacterial genomes, the bacterial and fungal sequences from
524 air samples did not show a significantly higher concentration of stress-related functions
525 compared to the samples coming from other ecosystems (**Fig 4, 5, Fig S4**).
526 Fungal genomes are expected to carry genes associated to global stress-related functions (*i.e.*
527 UV radiation, desiccation, oxidative stress), because of the innate resistance of fungi especially
528 fungal spores. These genes associated to global stress-related functions are likely acquired
529 during sporulation formation and certainly do not result from adaptation of fungi in air. When
530 studying genes coding more specific proteins that are not associated to spore resistance, such
531 as lipoate synthase and chromosome plasmid partitioning protein ParA, that might play a role
532 in oxidative stress (Allary et al., 2007; Bunik, 2003) and are more generally found in stress
533 resistance and adaptability of microorganisms (Shoeb et al., 2012; Zhang et al., 2018), they
534 were occasionally found in relatively high concentration in air samples (**Fig S3**). The detection
535 of metagenomic sequences annotated as genes coding specific proteins in air samples remains
536 difficult because of the low microbial biomass recovered. That is why we examined the
537 presence and concentration of global functions (*i.e.* UV protection related functions, oxidative
538 stress response related functions *etc.*) rather than specific functional genes.
539 The constant and large input of microbial cells to the planetary boundary layer and their
540 relatively short residence time (a few hours to a few days based on a model assuming that
541 microbial cells behave like non biological aerosols (Jaenicke, 1980)) might have hindered the
542 observation of the potential adaptation (physical selection and/or microbial adaptation) of
543 airborne microorganisms to the stressful atmospheric conditions and to the atmospheric

544 chemicals as discussed above. This issue might be addressed by investigating microbial
545 functional potential in the free troposphere (preferentially high enough above the ground so as
546 not to be influenced by the surface) where the microbial fluxes are smaller than in the planetary
547 boundary layer and where microbial airborne residence time might last much longer than in the
548 planetary boundary layer. This troposphere approach might help in determining the role of
549 stress in the atmosphere and validate our conceptual model on the physical stress of microbial
550 cells taking place during aerosolization and aerial transport selecting the resistant cells (**Fig 6**).
551 Another explanation might be due to the metagenomic approach that allows to sample both
552 living and dead cells. Aerosolization has been shown to be particularly stressful and even lethal
553 for microorganisms (Alsved et al., 2018; Thomas et al., 2011). The functional potential from
554 the dead cells in air might have a greater weight on the overall functional potential observed
555 and lead to the dilution of the functional potential of the actual living cells that have adapted to
556 atmospheric conditions. This might apply for both the overall functional potential discussed
557 previously and the stress-related functions.

558

559 **Conclusion**

560 We conducted the first global comparative metagenomic analysis to characterize the microbial
561 functional potential signature in the planetary boundary layer. Air samples showed no specific
562 signature of microbial functional potential which was mainly correlated to the surrounding
563 landscapes. However, air samples were characterized by a relatively high percentage of fungal
564 sequences compared to the source ecosystems (soil, surface seawater *etc.*). The relatively higher
565 concentrations of fungi in air drove the higher proportions of stress-related functions observed
566 in air metagenomes. Fungal cells and specifically fungal spores are innately resistant entities
567 well adapted to atmospheric conditions and which might survive better aerosolization and aerial
568 transport than bacterial cells. Stress-related functions were present in airborne bacteria but
569 rarely in higher concentrations compared to the bacterial communities in other ecosystems.
570 However, the constant flux of microbial cells to the planetary boundary layer might have
571 complicated the determination of a physical selection and/or microbial adaptation of airborne
572 microorganisms, especially bacterial communities. Meta-omics investigations on air with a
573 deeper sequencing are needed to confirm our results and explore the functionality of
574 atmospheric microorganisms further.

575

576 **References**

577 Aalismail, N. A., Ngugi, D. K., Díaz-Rúa, R., Alam, I., Cusack, M. and Duarte, C. M.: Functional
578 metagenomic analysis of dust-associated microbiomes above the Red Sea, *Sci Rep*, 9(1), 1–12,
579 doi:10.1038/s41598-019-50194-0, 2019.

580 Allary, M., Lu, J. Z., Zhu, L. and Prigge, S. T.: Scavenging of the cofactor lipoate is essential for the
581 survival of the malaria parasite *Plasmodium falciparum*, *Mol Microbiol*, 63(5), 1331–1344,
582 doi:10.1111/j.1365-2958.2007.05592.x, 2007.

583 Alsved, M., Holm, S., Christiansen, S., Smidt, M., Ling, M., Boesen, T., Finster, K., Bilde, M.,
584 Löndahl, J. and Šantl-Temkiv, T.: Effect of Aerosolization and Drying on the Viability of
585 *Pseudomonas syringae* Cells, *Front Microbiol*, 9, 3086, doi:10.3389/fmicb.2018.03086, 2018.

586 Amato, P., Demeer, F., Melaouhi, A., Fontanella, S., Martin-Biesse, A.-S., Sancelme, M., Laj, P. and
587 Delort, A.-M.: A fate for organic acids, formaldehyde and methanol in cloud water: their
588 biotransformation by micro-organisms, *Atmospheric Chemistry and Physics*, 7(15), 4159–4169,
589 doi:https://doi.org/10.5194/acp-7-4159-2007, 2007.

590 Amato, P., Besaury, L., Joly, M., Penaud, B., Deguillaume, L. and Delort, A.-M.: Metatranscriptomic
591 exploration of microbial functioning in clouds, *Sci Rep*, 9(1), 1–12, doi:10.1038/s41598-019-41032-4,
592 2019.

- 593 Ariya, P., Sun, J., Eltouny, N., Hudson, E., Hayes, C. and Kos, G.: Physical and chemical
594 characterization of bioaerosols--Implications for nucleation processes, *International Reviews in*
595 *Physical Chemistry*, 28, 1–32, doi:10.1080/01442350802597438, 2009.
- 596 Ariya, P. A., Nepotchatykh, O., Ignatova, O. and Amyot, M.: Microbiological degradation of
597 atmospheric organic compounds, *Geophysical Research Letters*, 29(22), 34–1–34–4,
598 doi:10.1029/2002GL015637, 2002.
- 599 Aylor, D. E.: Spread of Plant Disease on a Continental Scale: Role of Aerial Dispersal of Pathogens,
600 *Ecology*, 84(8), 1989–1997, 2003.
- 601 Brown, J. K. M. and Hovmöller, M. S.: Aerial dispersal of pathogens on the global and continental
602 scales and its impact on plant disease, *Science*, 297(5581), 537–541, doi:10.1126/science.1072678,
603 2002.
- 604 Brune, A., Frenzel, P. and Cypionka, H.: Life at the oxic–anoxic interface: microbial activities and
605 adaptations, *FEMS Microbiol Rev*, 24(5), 691–710, doi:10.1111/j.1574-6976.2000.tb00567.x, 2000.
- 606 Buchfink, B., Xie, C. and Huson, D. H.: Fast and sensitive protein alignment using DIAMOND, *Nat*
607 *Methods*, 12(1), 59–60, doi:10.1038/nmeth.3176, 2015.
- 608 Bunik, V. I.: 2-Oxo acid dehydrogenase complexes in redox regulation, *Eur. J. Biochem.*, 270(6),
609 1036–1042, doi:10.1046/j.1432-1033.2003.03470.x, 2003.
- 610 Cao, C., Jiang, W., Wang, B., Fang, J., Lang, J., Tian, G., Jiang, J. and Zhu, T. F.: Inhalable
611 Microorganisms in Beijing's PM_{2.5} and PM₁₀ Pollutants during a Severe Smog Event, *Environ. Sci.*
612 *Technol.*, 48(3), 1499–1507, doi:10.1021/es4048472, 2014.
- 613 Delmont, T. O., Malandain, C., Prestat, E., Larose, C., Monier, J.-M., Simonet, P. and Vogel, T. M.:
614 Metagenomic mining for microbiologists, *ISME J*, 5(12), 1837–1843, doi:10.1038/ismej.2011.61,
615 2011.
- 616 Delort, A.-M., Vařtilingom, M., Amato, P., Sancelme, M., Parazols, M., Mailhot, G., Laj, P. and
617 Deguillaume, L.: A short overview of the microbial population in clouds: Potential roles in
618 atmospheric chemistry and nucleation processes, *Atmospheric Research*, 98(2), 249–260,
619 doi:10.1016/j.atmosres.2010.07.004, 2010.
- 620 Després, V., Huffman, J. A., Burrows, S. M., Hoose, C., Safatov, A., Buryak, G., Fröhlich-Nowoisky,
621 J., Elbert, W., Andreae, M., Pöschl, U. and Jaenicke, R.: Primary biological aerosol particles in the
622 atmosphere: a review, *Tellus B: Chemical and Physical Meteorology*, 64(1), 15598,
623 doi:10.3402/tellusb.v64i0.15598, 2012.
- 624 Dommergue, A., Amato, P., Tignat-Perrier, R., Magand, O., Thollot, A., Joly, M., Bouvier, L.,
625 Sellegri, K., Vogel, T., Sonke, J. E., Jaffrezo, J.-L., Andrade, M., Moreno, I., Labuschagne, C., Martin,
626 L., Zhang, Q. and Larose, C.: Methods to investigate the global atmospheric microbiome, *Front.*
627 *Microbiol.*, 10, doi:10.3389/fmicb.2019.00243, 2019.
- 628 Donovan, P. D., Gonzalez, G., Higgins, D. G., Butler, G. and Ito, K.: Identification of fungi in
629 shotgun metagenomics datasets, *PLOS ONE*, 13(2), e0192898, doi:10.1371/journal.pone.0192898,
630 2018.
- 631 Els, N., Larose, C., Baumann-Stanzer, K., Tignat-Perrier, R., Keuschnig, C., Vogel, T. M. and Sattler,
632 B.: Microbial composition in seasonal time series of free tropospheric air and precipitation reveals
633 community separation, *Aerobiologia*, doi:10.1007/s10453-019-09606-x, 2019.
- 634 Friedl, M. A., McIver, D. K., Hodges, J. C. F., Zhang, X. Y., Muchoney, D., Strahler, A. H.,
635 Woodcock, C. E., Gopal, S., Schneider, A., Cooper, A., Baccini, A., Gao, F. and Schaaf, C.: Global land

636 cover mapping from MODIS: algorithms and early results, *Remote Sensing of Environment*, 83(1),
637 287–302, doi:10.1016/S0034-4257(02)00078-0, 2002.

638 Gene Ontology Consortium, T.: The Gene Ontology Resource: 20 years and still GOing strong,
639 *Nucleic Acids Res*, 47(D1), D330–D338, doi:10.1093/nar/gky1055, 2019.

640 Griffin, D. W.: Atmospheric Movement of Microorganisms in Clouds of Desert Dust and
641 Implications for Human Health, *Clin Microbiol Rev*, 20(3), 459–477, doi:10.1128/CMR.00039-06,
642 2007.

643 Gusareva, E. S., Acerbi, E., Lau, K. J. X., Luhung, I., Premkrishnan, B. N. V., Kolundžija, S.,
644 Purbojati, R. W., Wong, A., Houghton, J. N. I., Miller, D., Gaultier, N. E., Heinle, C. E., Clare, M. E.,
645 Vettath, V. K., Kee, C., Lim, S. B. Y., Chénard, C., Phung, W. J., Kushwaha, K. K., Nee, A. P., Putra,
646 A., Panicker, D., Yanqing, K., Hwee, Y. Z., Lohar, S. R., Kuwata, M., Kim, H. L., Yang, L., Uchida, A.,
647 Drautz-Moses, D. I., Junqueira, A. C. M. and Schuster, S. C.: Microbial communities in the tropical
648 air ecosystem follow a precise diel cycle, *PNAS*, 116(46), 23299–23308,
649 doi:10.1073/pnas.1908493116, 2019.

650 Hadley, W. and Winston, C.: Create Elegant Data Visualisations Using the Grammar of Graphics,
651 [online] Available from: <https://cran.r-project.org/web/packages/ggplot2/ggplot2.pdf>, 2019.

652 Hill, K. A., Shepson, P. B., Galbavy, E. S., Anastasio, C., Kourtev, P. S., Konopka, A. and Stirm, B. H.:
653 Processing of atmospheric nitrogen by clouds above a forest environment, *Journal of Geophysical*
654 *Research: Atmospheres*, 112(D11), doi:10.1029/2006JD008002, 2007.

655 Hindré, T., Knibbe, C., Beslon, G. and Schneider, D.: New insights into bacterial adaptation through
656 *in vivo* and *in silico* experimental evolution, *Nature Reviews Microbiology*, 10(5), 352–365,
657 doi:10.1038/nrmicro2750, 2012.

658 Huang, M. and Hull, C. M.: Sporulation: How to survive on planet Earth (and beyond), *Curr Genet*,
659 63(5), 831–838, doi:10.1007/s00294-017-0694-7, 2017.

660 Huerta-Cepas, J., Forslund, K., Coelho, L. P., Szklarczyk, D., Jensen, L. J., von Mering, C. and Bork,
661 P.: Fast Genome-Wide Functional Annotation through Orthology Assignment by eggNOG-Mapper,
662 *Mol. Biol. Evol.*, 34(8), 2115–2122, doi:10.1093/molbev/msx148, 2017.

663 Huson, D. H., Beier, S., Flade, I., Górska, A., El-Hadidi, M., Mitra, S., Ruscheweyh, H.-J. and Tappu,
664 R.: MEGAN Community Edition - Interactive Exploration and Analysis of Large-Scale Microbiome
665 Sequencing Data, *PLoS Comput Biol*, 12(6), e1004957, doi:10.1371/journal.pcbi.1004957, 2016.

666 Imlay, J. A.: The molecular mechanisms and physiological consequences of oxidative stress: lessons
667 from a model bacterium, *Nat Rev Microbiol*, 11(7), 443–454, doi:10.1038/nrmicro3032, 2013.

668 Innocente, E., Squizzato, S., Visin, F., Facca, C., Rampazzo, G., Bertolini, V., Gandolfi, I., Franzetti,
669 A., Ambrosini, R. and Bestetti, G.: Influence of seasonality, air mass origin and particulate matter
670 chemical composition on airborne bacterial community structure in the Po Valley, Italy, *Sci. Total*
671 *Environ.*, 593–594, 677–687, doi:10.1016/j.scitotenv.2017.03.199, 2017.

672 Jaenicke, R.: Atmospheric aerosols and global climate, *Journal of Aerosol Science*, 11(5), 577–588,
673 doi:10.1016/0021-8502(80)90131-7, 1980.

674 Joly, M., Amato, P., Sancelme, M., Vinatier, V., Abrantes, M., Deguillaume, L. and Delort, A.-M.:
675 Survival of microbial isolates from clouds toward simulated atmospheric stress factors, *Atmospheric*
676 *Environment*, 117, 92–98, doi:10.1016/j.atmosenv.2015.07.009, 2015.

- 677 Keegan, K. P., Glass, E. M. and Meyer, F.: MG-RAST, a Metagenomics Service for Analysis of
678 Microbial Community Structure and Function, *Methods Mol Biol*, 1399, 207–233, doi:10.1007/978-
679 1-4939-3369-3_13, 2016.
- 680 Leinonen, R., Sugawara, H. and Shumway, M.: The Sequence Read Archive, *Nucleic Acids Res*,
681 39(Database issue), D19–D21, doi:10.1093/nar/gkq1019, 2011.
- 682 Li, Y., Zheng, L., Zhang, Y., Liu, H. and Jing, H.: Comparative metagenomics study reveals pollution
683 induced changes of microbial genes in mangrove sediments, *Scientific Reports*, 9(1), 5739,
684 doi:10.1038/s41598-019-42260-4, 2019.
- 685 Malik, A. A., Chowdhury, S., Schlager, V., Oliver, A., Puissant, J., Vazquez, P. G. M., Jehmlich, N.,
686 von Bergen, M., Griffiths, R. I. and Gleixner, G.: Soil Fungal:Bacterial Ratios Are Linked to Altered
687 Carbon Cycling, *Front. Microbiol.*, 7, doi:10.3389/fmicb.2016.01247, 2016.
- 688 NCBI Resource Coordinators: Database resources of the National Center for Biotechnology
689 Information, *Nucleic Acids Res*, 46(D1), D8–D13, doi:10.1093/nar/gkx1095, 2018.
- 690 Oksanen, J., Guillaume Blanchet, F., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P.
691 R., O'Hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E. and Wagner, H.:
692 Community Ecology Package, [online] Available from: <https://github.com/vegandevs/vegan>, 2019.
- 693 Rey, O., Danchin, E., Mirouze, M., Loot, C. and Blanchet, S.: Adaptation to Global Change: A
694 Transposable Element–Epigenetics Perspective, *Trends in Ecology & Evolution*, 31(7), 514–526,
695 doi:10.1016/j.tree.2016.03.013, 2016.
- 696 Shannan, S., Collins, K. and Emanuel, W. R.: Global mosaics of the standard MODIS land cover type
697 data, 2014.
- 698 Shoeb, E., Badar, U., Akhter, J., Shams, H., Sultana, M. and Ansari, M. A.: Horizontal gene transfer
699 of stress resistance genes through plasmid transport, *World J. Microbiol. Biotechnol.*, 28(3), 1021–
700 1025, doi:10.1007/s11274-011-0900-6, 2012.
- 701 Thomas, R. J., Webber, D., Hopkins, R., Frost, A., Laws, T., Jayasekera, P. N. and Atkins, T.: The
702 Cell Membrane as a Major Site of Damage during Aerosolization of *Escherichia coli*, *Appl. Environ.*
703 *Microbiol.*, 77(3), 920–925, doi:10.1128/AEM.01116-10, 2011.
- 704 Tignat-Perrier, R., Dommergue, A., Thollot, A., Keuschnig, C., Magand, O., Vogel, T. M. and
705 Larose, C.: Global airborne microbial communities controlled by surrounding landscapes and wind
706 conditions, *Sci Rep*, 9(1), 1–11, doi:10.1038/s41598-019-51073-4, 2019.
- 707 Tringe, S. G., von Mering, C., Kobayashi, A., Salamov, A. A., Chen, K., Chang, H. W., Podar, M.,
708 Short, J. M., Mathur, E. J., Detter, J. C., Bork, P., Hugenholtz, P. and Rubin, E. M.: Comparative
709 metagenomics of microbial communities, *Science*, 308(5721), 554–557, doi:10.1126/science.1107851,
710 2005.
- 711 Väitilingom, M., Amato, P., Sancelme, M., Laj, P., Leriche, M. and Delort, A.-M.: Contribution of
712 Microbial Activity to Carbon Chemistry in Clouds, *Appl Environ Microbiol*, 76(1), 23–29,
713 doi:10.1128/AEM.01127-09, 2010.
- 714 Väitilingom, M., Deguillaume, L., Vinatier, V., Sancelme, M., Amato, P., Chaumerliac, N. and Delort,
715 A.-M.: Potential impact of microbial activity on the oxidant capacity and organic carbon budget in
716 clouds, *PNAS*, 110(2), 559–564, doi:10.1073/pnas.1205743110, 2013.
- 717 Vartoukian, S. R., Palmer, R. M. and Wade, W. G.: Strategies for culture of 'unculturable' bacteria,
718 *FEMS Microbiology Letters*, 309(1), 1–7, doi:10.1111/j.1574-6968.2010.02000.x, 2010.

- 719 Wickham, H.: Flexibly Reshape Data: A Reboot of the Reshape Packa, [online] Available from:
720 <https://cran.r-project.org/web/packages/reshape2/reshape2.pdf>, 2017.
- 721 Wood, D. E. and Salzberg, S. L.: Kraken: ultrafast metagenomic sequence classification using exact
722 alignments, *Genome Biology*, 15(3), R46, doi:10.1186/gb-2014-15-3-r46, 2014.
- 723 Xie, W., Wang, F., Guo, L., Chen, Z., Sievert, S. M., Meng, J., Huang, G., Li, Y., Yan, Q., Wu, S.,
724 Wang, X., Chen, S., He, G., Xiao, X. and Xu, A.: Comparative metagenomics of microbial
725 communities inhabiting deep-sea hydrothermal vent chimneys with contrasting chemistries, *The*
726 *ISME Journal*, 5(3), 414–426, doi:10.1038/ismej.2010.144, 2011.
- 727 Yang, Y., Yokobori, S. and Yamagishi, A.: UV-resistant bacteria isolated from upper troposphere and
728 lower stratosphere, *Biol.Sci.Space*, 22, doi:10.2187/bss.22.18, 2008.
- 729 Yooseph, S., Nealon, K. H., Rusch, D. B., McCrow, J. P., Dupont, C. L., Kim, M., Johnson, J.,
730 Montgomery, R., Ferriera, S., Beeson, K., Williamson, S. J., Tovchigrechko, A., Allen, A. E., Zeigler,
731 L. A., Sutton, G., Eisenstadt, E., Rogers, Y.-H., Friedman, R., Frazier, M. and Venter, J. C.: Genomic
732 and functional adaptation in surface ocean planktonic prokaryotes, *Nature*, 468(7320), 60–66,
733 doi:10.1038/nature09530, 2010.
- 734 Yooseph, S., Andrews-Pfannkoch, C., Tenney, A., McQuaid, J., Williamson, S., Thiagarajan, M.,
735 Bami, D., Zeigler-Allen, L., Hoffman, J., Goll, J. B., Fadrosch, D., Glass, J., Adams, M. D., Friedman,
736 R. and Venter, J. C.: A Metagenomic Framework for the Study of Airborne Microbial Communities,
737 *PLOS ONE*, 8(12), e81862, doi:10.1371/journal.pone.0081862, 2013.
- 738 Zhang, H., Hu, Y., Zhou, C., Yang, Z., Wu, L., Zhu, M., Bao, H., Zhou, Y., Pang, M., Wang, R. and
739 Zhou, X.: Stress resistance, motility and biofilm formation mediated by a 25kb plasmid pLMSZ08 in
740 *Listeria monocytogenes*, *Food Control*, 94, 345–352, doi:10.1016/j.foodcont.2018.07.002, 2018.

741
742 **Competing interests.** The authors declare that they have no conflict of interest.

743
744 **Financial support.** This work was supported by the Agence Nationale de la Recherche
745 [ANR-15-CE01-0002–03 INHALE]; Région Auvergne-Rhône Alpes [ARC3 2016];
746 CAMPUS France [program XU GUANGQI] and the French Polar Institute IPEV [program
747 1028 and 399].

748
749 **Author contributions.** AD, CL and TMV designed the experiment. RTP, AD, AT and OM
750 conducted the sampling field campaign. RTP did the molecular biology, bioinformatics and
751 statistical analyses. RTP, AD, CL and TMV analyzed the results. RTP, TM, AD and CL wrote
752 the manuscript. All authors reviewed the manuscript.

753
754 **Acknowledgements.** The chemical analyses were performed at the IGE AirOSol platform. This
755 work was hosted by the following stations: Chacaltaya, Namco, puy de Dôme, Cape-Point, Pic-
756 du-Midi, Amsterdam-Island, Storm-Peak, Villum RS and we thank I.Jouvie, G.Hallar,
757 I.McCubbin, Benny and Jesper, B.Jensen, A.Nicosia, M.Ribeiro, L.Besaury, L.Bouvier,
758 M.Joly, I.Moreno, M.Rocca, F.Velarde for sampling and station management. Logistical
759 support for Amsterdam-Island and Villum field campaigns was provided by the French Polar
760 Institute IPEV (program 1028 and 399). We thank our project partners: K.Sellegrri, P.Amato,
761 M.Andrade, Q.Zhang, C.Labuschagne and L.Martin, J. Sonke. We thank R.Edwards, J. Schauer
762 and C.Worley for lending their HV sampler. We thank L.Pouilloux for computing assistance
763 and maintenance of the Newton server.

764

765 **Data availability.** Sequences reported in this paper have been deposited in [ftp://ftp-adn.ec-](ftp://ftp-adn.ec-lyon.fr/Tignat-Perrier_2020_air_metagen_INHALE/)
766 [lyon.fr/Tignat-Perrier_2020_air_metagen_INHALE/](ftp://ftp-adn.ec-lyon.fr/Tignat-Perrier_2020_air_metagen_INHALE/). A file has been attached explaining the
767 correspondence between file names and samples.