November 2nd, 2020

Dear Editor,

Attached is a revised version of our manuscript, "Impacts of biogenic polyunsaturated aldehydes on metabolism and community composition of particle-attached bacteria in coastal hypoxia" by Zhengchao Wu et al.

We greatly appreciate the comments and suggestions provided by the reviewers and editor. They have been very constructive, contributing significantly to improve the overall quality of our paper. We have carefully addressed all their points in the revised manuscript, and have detailed our changes in the response to reviewers (below). We hope that you will now find our manuscript suitable for publication.

Sincerely yours,

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Response to Anonymous Referee #1

General comments

1. The manuscript is generally well written with proper language although some sentences need improvements. Results are well presented in the figures, but some complement of error bars are needed.

Response: Thanks to the reviewer for constructive comments. The manuscript has been proofread by a native English speaker to correct grammar errors and to improve the written language. We have redone these figures by including the error bars in the revised manuscript.

2. I lack a convincing motivation to the importance of PUAs compared to the multitude of other organic compounds, and the importance of particle-attached bacteria as compared to free-living. No direct comparison with other organic compounds or free-living bacteria is done in the study.

Response: The reviewer is right that there are other organic compounds that may also likely affect bacterial respiration, such as 2-n-pentyl-4-quinolinol (Long et al., 2003) and acylated homoserine lactones (Hmelo et al., 2011). However, a perennial bloom of PUA-producing diatoms in the PRE mouth (Wu and Li, 2016) should argue for the importance of PUAs for microbial activity here compared to many other organic compounds. The reviewer is also right that free-living bacteria are important for community respiration in the ocean. However, our focus here is on coastal zones. Our field measurements suggested that bacterial respiration in the hypoxic waters was largely contributed by particle-attached bacteria (>0.8 μ m) with FLB (0.2-0.8 μ m) only accounting for 25-30% of the total rates. We have provided the data of the size-fractionated bacterial respiration rates in Figure S1 of the revised manuscript.

3. A general importance of PUAs and particle attached bacteria should be tuned down in the discussion and conclusion. The effects of PUA on particle-attached bacteria is still of value as such.

Response: OK. We have rewritten the relevant sentences in the discussion and the conclusion sections as suggested by the reviewer.

4. A short-coming of the experimental design is a lack of true replication of the treatments.

Response: We actually have two replicates for each treatment. The original text was not well written. We have clarified this in the revised manuscript.

5. In addition, the fact that only one season has been investigated.

Response: The hypoxia could only occur during the summer in our study region. Therefore, a seasonality of hypoxia suggested by the reviewer may be unnecessary.

6. I also miss proper measurement of the abundance, acidity and taxonomy of free-living bacterial to put the claimed influence of particle attached bacteria in perspective.

Response: We have added data of the size-fractionated bacterial respiration rates (for both free-living and particle-attached bacteria) in the hypoxic waters of station Y1 to the revised manuscript along with the bulk bacteria taxonomy data, although we do not have measurements for free-living bacteria abundance and taxonomy.

7. A similar argument for the lack of other organic compounds in the study.

Response: The reviewer is right about that there are other organic compounds that may also likely affect bacterial respiration. In the revised manuscript, we have compared PUAs with other organic compounds that would potentially affect bacterial activities in the hypoxia, such as 2-n-pentyl-4-quinolinol (PQ) and acylated homoserine lactones (AHLs). "A perennial bloom of PUA-producing diatoms in the PRE mouth (Wu and Li, 2016) may indicate the importance of PUAs for microbial activity here compared to many other organic compounds, such as 2-n-pentyl-4-quinolinol (Long et al., 2003) and acylated homoserine lactones (Hmelo et al., 2011)."

8. The conclusions must therefore be made more cautious, specific and these shortcomings commented on. Some speculative statements in the discussion and conclusion section need to be removed or rephrased.

Response: We agree with the reviewer on this. In the revised manuscript, we have carefully rewritten the discussion and the conclusion sections.

9. There are parts of the method descriptions that need to be clarified, better specified or added. A major revision in this spirit is required to motivate publication.

Response: We thank the review for these comments. We have carefully rewritten all the parts of descriptions of the relevant methodology suggested by the reviewer.

Detailed comments

10. r. 13-14 There is an extensive literature on eutrophication driven hypoxia in e.g. the Baltic Sea since 4 decades (cf. Cloern 2001). Please rephrase sentence accordingly

Response: Agree. In the revised manuscript, the sentence has been rewritten as "Eutrophication-driven coastal hypoxia is of great interest for decades". 11. r. 15. Do you mean ": : : water mainly dominated: : :".

Response: Agree.

12. r. 21 Please change "activity" to "..e.g. bacterial respiration and growth: : :) and revise the sentence.

Response: Done.

13. r. 35 Change to ".. deoxygenation is also tightly: : :"

Response: Done.

14. r. 43-45 In most aquatic environments free living bacteria are dominating in numbers as well as biomass (e.g. Kirchman 2008). The reference is not convincingly showing that particle-associated bacterial dominate in terms of abundance or how the growth of particle associated bacteria was measured. Please revise the message. This also question the focus on particle associated bacteria in the manuscript.

Response: We agree with the reviewer that free-living bacteria are most dominant in many parts of the ocean. However, our focus is on the high turbid coastal transition zone where particle-attached bacteria can be relatively more important. We have added another reference of Lee et al (2015) to this sentence to show the importance of PAB in coastal regions. Lee, S., Lee, C., Bong, C., Narayanan, K., Sim, E.: The dynamics of attached and free-living bacterial population in tropical coastal waters, Mar. Freshwater Res., 66, 701-710, 2015.

15. r. 49-51 In many cases free-living bacteria dominate the respiration (Robinson and leB Williams 2005). Both types of bacteria is preferably studied. This question the general relevance of the study.

Response: The reviewer is right about that free-living bacteria are more important for community respiration in the open ocean. However, what we focused on is the coastal transition zone, where particle-attached bacteria could be important. Our field measurements suggested that bacterial respiration in the hypoxic waters was largely contributed by particle-attached bacteria (>0.8 μ m) with FLB (0.2-0.8 μ m) only accounting for 25-30% of the total rates. We have added these results to the revised manuscript. We have also rewritten these sentences to emphasize more on the general relevance of our study for both types of bacteria.

16. r. 64-65. However, many other organic compounds may drive the bacterial respiration. Please provide some reference showing to that extent PUF is contributing to bacterial respiration.

Response: We agree with the reviewer on this point. References for PUAs contribution to bacterial

metabolisms have been provided in the revised manuscript. "The strong effect of PUAs on bacterial growth, production, and respiration has been well demonstrated in the laboratory studies (Ribalet et al., 2008) and the field studies (Balestra et al., 2011; Edwards et al., 2015)."

17. r.68-76 I would prefer more explicit research questions to be addressed in this paragraph for clarity and coupling to performed experiments.

Response: Ok. We have rewritten this paragraph as "...There are three specific questions to address here: What are the relative roles of PAB and FLB on bacterial respiration in the hypoxic waters? What are the actual levels of PUAs in the hypoxic waters? What are the responses of PAB to PUAs in the hypoxic waters? For the first question, size-fractionated bacterial respiration rates were estimated for both FLB (0.2-0.8 μ m) and PAB (>0.8 μ m) in the hypoxic waters. For the second question, the concentrations of particulate and dissolved PUAs within the hypoxic waters were measured in the field. Besides, the hotspot PUAs concentration associated with the suspended particles within the hypoxic waters was directly quantified for the first time using large-volume filtration and subsequent on-site derivation and extraction. For the third question, field PUAs-amended incubation experiments were conducted for PAB (>25 μ m) retrieved from the low-oxygen waters. We focused on particles of >25 μ m to better explore the role of PUAs on PAB given the actual levels of PUAs hotspots, to assess the PAB responses (including bacterial abundance, respiration, production, and community composition) to the exogenous PUAs in the hypoxic waters..."

18. r. 84-85 Please define here what depths that were used for middle and bottom water categories (e.g. figure 6).

Response: The middle layer was at 12 m with the bottom layer at 25 m (4m above the seafloor) for station Y1 (Figure 6). We have clarified this in the revised manuscript.

19. r. 87-88 Please define the abbreviations pPUA and dPUA.

Response: Agree. The pPUAs and dPUAs have been defined in the revised manuscript.

20. r. 97 Filtration and freezing of nutrient samples may release nutrients from broken cells.

Response: The influences of filtration and freezing/thaw on nutrient concentration should be negligible due to high nutrient concentration in the coastal system. They will only affect the oligotrophic open ocean waters with nanomolar nutrients (Li QP and Hansell DA, Anal Chim Acta, 611, 68-72, 2008).

21. r.124-129 How is the centrifugation and resuspension of particles influencing their morphology, attachments of PUAs and PABs?

Response: We believe that particle morphology and the attachments of PUAs and PABs will not be influenced by low-speed centrifugation (3000 rpm for one minute) or by a gently shaking for resuspension. The same approaches have been used to study particle-attached bacteria and particle-related compounds on sinking particles (Hmelo et al., 2011)

22. r. 131-132 Pleas provide the recovery efficiency of particle attached PUAs after the preparation procedure described.

Response: The recovery for the particle-adsorbed PUAs should be 100% as the supernatants after centrifugations have all been added back to the final 50 ml centrifuge tube. We have clarified this in the revised manuscript.

23. r. 136-137. Please provide a reference where the method is validated.

Response: Done. References have been added to the revised manuscript. The protocol is modified from those of Edwards et al. (2015) and Wu and Li (2016).

24. r. 146-147 Freeze thawing may relate PUAs from living cells also.

Response: The reviewer is right that free-thawing would release PUAs from living cells. What we actually mean in the text is that we use the same determination method for undisrupted and disrupted PUAs although they are pre-treated differently (one with direct extraction method and the other with the freeze-thaw method). Anyway, we have clarified this in the revised manuscript.

25. r. 152 Should it be nmol per some volume or particle unit?

Response: Agree. It is the concentration of the undisturbed PUAs in the 50 mL sampling tube. We have clarified this in the revised manuscript.

26. r.160 Please specify what is meant by clean. What was the washing procedure?

Response: It means sterile. We have rewritten this in the revised manuscript. Generally, these bottles have been soaked in 10% HCl for 24 h, rinsed with deionized water for several times, and sterilized before use.

27. r.162 As presented here there was no true replication of the treatments?

Response: We actually have two replicates for each treatment. The sample in the 1-L Nalgene bottle had been transferred to two 0.5 L bottles for each treatment. We have clarified this in the revised manuscript.

28. r. 167 How does methanol included in the procedure affect bacterial abundance and activity? Any control or test for this? Please comment on relevance for natural conditions.

Response: The methanol has been added as a cosolvent for PUAs (Franze et al., 2018). The methanol concentration of 0.05% in our experiment should not have a large effect on bacteria, since the previous study suggested that bacterial strains would not be significantly affected by methanol at a level below 1% (Patterson and Ricke, 2015).

Patterson J.A., and S.C. Ricke, S.C., (2015) Effect of ethanol and methanol on growth of ruminal bacteria *Selenomonas ruminantium* and *Butyrivibrio fibrisolvens*, Journal of Environmental Science and Health, Part B, 50:1, 62-67.

29. r. 176-179 Give some information on how close to natural conditions these final concentration of PUAs are.

Response: The PUAs level was close to the hotspot PUAs concentration of 240 μ mol L⁻¹ found in a station near the PRE and was also comparable to the hotspot concentration of ~26 μ mol L⁻¹ found in the temperate west North Atlantic (Edwards et al., 2015). We should emphasize that the concentration of PUAs in the water-column is inhomogeneous due to the presence of particles. The hotspot concentration of PUAs associated with these particles should be the PUAs concentration in the volume of the water parcel displaced by the aggregation particles.

30. r.180 it is not obvious that turbidity will be detected if cells remain below about 10^9 cells cm³. Were the cell concentration measured by direct microscopy?

Response: We did not measure cell concentration during the experiments. The experiment is designed to only qualitatively assess the PAB response to different types of carbon sources. The culture duration of over 30 days should be long enough for significant bacterial growth (say with cell concentration well exceed the detection limit) to show up if the organic substrate could be used as a carbon source (Dong et al., 2015, doi:10.5194/bg-12-2163-2015).

31. r. 181 Please provide a description on incubation conditions and length.

Response: Done. These experiments were performed in dark at room temperature for over 30 days. We have clarified this in the revised manuscript.

32. r. 185-196 The description of methodology is unclear. Please make clear if and how free-living bacterial abundance was measured? How is the methanol treatment accounting for bacteria from breaking particles? In addition, please provide a reference validating this method. Give some measure of the precision of the flow cytometer analysis and a relevant reference.

Response: We did not measure the free-living bacteria abundance. We only measure the abundance of the bulk water bacteria (>0.2 μ m) that includes both FLB and PAB. The method for bulk-water bacterial abundance has been added to the revised manuscript. We have also provided a reference for the flow cytometry method (Marie et al., 1997) as well as the relevant precision (CV%). The original text about methanol treatment was not well written, we have rewritten the sentence as "To break up particles and attached bacteria, 0.2 mL pure methanol was added to the 2 mL sample and vortexed".

Marie, D., Partensky, F., Jacquet, S. and Vaulot, D.: Enumeration and cell cycle analysis of natural populations of marine picoplankton by flow cytometry using the nucleic acid stain SYBR Green I, Appl. Environ. Microbiol. 63, 186-193, 1997

33. r.198-202 As no relevant pre-filtration is used you may include other organisms than bacteria in the respiration estimate. Please clarify and rephrase as needed.

Response: The reviewer is right about this. We did not perform any pre-filtration at this step. Besides the PAB, the particle aggregates of >25 μ m would likely consist of some phytoplankton and microzooplankton. So, the BR could be overestimated in our experiment. However, this effect could be relatively small, given that the chlorophyll-a of the raw seawater (hypoxic waters in the subsurface layer) was very low and there was not much microzooplankton in the sample (confirmed by FlowCAM). We have clarified this in the revised manuscript.

34. r. 208-209 What was the final TCA concentration in the sample. This does not follow the common procedure. Neither use of ethanol.

Response: Done. The final TCA concentration is 5%. The procedure of TCA step, as well as the use of ethanol, is based on the previous publication (Huang et al, 2018, doi: 10.1016/j.scitotenv.2018.03.222). We have clarified this in the revised manuscript.

35. r. 241 However, the lack of true replicates for the treatment (i.e. replicate 1-L Nalgene bottles per treatment) question a reliable result from the *t*-test.

Response: As we have responded to the previous point of this reviewer on No. 27, we have two replicates for each treatment. The sample in the 1-L Nalgene bottle had been transferred to two 0.5 L bottles for each treatment.

36. r. 251, Figure 1. Consider to reverse the colour palette. More logical to have blue for well oxygenated and red for hypoxia. Change "constant" to "similar".

Response: Done. The figure has been revised as suggested by the reviewer. The word "constant" has been replaced by "similar" as well.

37. r. 261-263. Pleas provide a statistical test for the claimed difference and confidence intervals (error bars) in figure 4.

Response: Agree. In the revised manuscript, we have added statistical information to the difference for BR and BP. Error bars have also been provided in the revised figure.

38. r.265. Proved statistical results for the claimed difference between bacterial phyla. Do the same for other differences claimed throughout the manuscript.

Response: It is a typo. A statistical test is not doable for comparing different bacterial compositions. In the revised manuscript, we have corrected the sentence as "... was <u>substantially</u> different from those of X2 and X3". In addition, statistical information has been checked for each comparison throughout the manuscript.

39. r. 273 Should it be Bacteroidetes also here?

Response: Yes. We have revised as 4% of Bacteroidetes.

39-2. Figure 5. Please present the type of error bars used. Same for all figures with error bars.

Response: Error bars are the standard deviations. We have clarified these in the figure legends of the revised manuscript.

40. r. 320-329 What can be considered significant differences as opposed to random variation in this analysis. Please motivate convincingly.

Response: Statistical information (t-value, n, and p-value) for comparing γ -pro percentage between control and treatments have been added to the revised manuscript. However, a statistical test is not doable for comparing the difference of the bacterial community compositions. We have replaced the word "significant" with "substantial" in the revised manuscript.

41. r. 335 One month is an extremely long incubation. How relevant is this for the application to the natural environment?

Response: Bacterial utilization of organic carbons may depend on the nature of the organic compound. Bacteria may need a longer period to utilize refractory organic matters (ALK and PAH). On the other hand, our experiment goal is to qualitatively assess the possibility of PUA as a carbon source for PAB growth. The color change can be more easily appreciated after one-month for both PAH and ALK and thus allow us to compare the bacterial responses to different organic

compounds (PUA, PAH, and ALK). There was no bacterial growth in the PUA medium throughout the one month should provide strong evidence that PUA was not used as a carbon source.

42. r. 360 Would be more informative to use a unit per particle or mass of particles? Litre of particles is unclear.

Response: The concentration of PUAs in the water-column is inhomogeneous due to the presence of particles. The hotspot concentration of PUAs should be the PUAs concentration in the volume of the water parcel displaced by the aggregation particles. Therefore, particle volume is more informative and allows a better comparison of the hotspot concentration with the bulk water concentration.

43. r. 365 This assumes that PUA is a major substrate among all other organic compounds. Please provide some references on this matter and discuss it critically.

Response: We should emphasize that our focus here is to explore its role as a signal substance for PAB metabolism rather than as an organic carbon substrate for PAB growth. Actually, PUAs accounts for only a small part of the particulate organic carbon (1-16%, Edwards et al., 2015). The specific arrangement of two double bonds and carbonyl chain makes PUAs not a group of labile organic carbon for bacterial utilization. Anyway, we have rewritten the sentence to clarify this in the revised manuscript.

44. r. 368 This should be compared with the biomass of free living bacteria. They may also be elevated in the hypoxic water. I find the lack of measurement of free-living bacteria a short coming in the context of claiming importance of PABs.

Response: The reviewer is right about that free-living bacteria (FLB) may also be elevated in the hypoxic water. Our field data suggested that FLB respiration accounts for only 25-30% of the total bacterial community respiration in the hypoxic waters. In the revised manuscript, we have provided the data of the size-fractionated respiration rates for FLB and PAB (Figure S1), as well as the data of the community composition of bulk bacteria (Figure S2).

45. r. 372-373. How is respiration by particle-attached bacteria distinguished from protozoa, phytoplankton and larger zooplankton? This is typically difficult to achieve. Comment in a critical manner.

Response: We cannot distinguish BR from the respirations of phytoplankton and microzooplankton (larger zooplankton has been picked off already). However, this effect could be relatively small, since the raw sweater in the hypoxic zone had very low chlorophyll-a and there was virtually not much microzooplankton in the sample (confirmed by FlowCAM). We have

clarified these in the revised manuscript.

46. r. 376-390. Given the apparent lack of true replication of the treatments (i.e. replicate 1 L Nalgene bottles) the conclusions regarding treatment effects is highly uncertain. This needs a discussion.

Response: As we have responded to the previous point of this reviewer on No. 27, we have two replicates for each treatment. We have clarified this in the revised manuscript.

47. r. 389-390 Relevance in the natural environment assumes that the applied concentrations are relevant for those occurring in the natural environment. Please consider, discuss and modify the conclusion accordingly.

Response: We should emphasize that the concentration of PUAs in the water-column is inhomogeneous due to the presence of particles. The micromolar level of PUAs for incubation was chosen to represent the actual hotspot concentration of PUAs (the PUAs concentration in the volume of the water parcel displaced by the aggregation particles) not the mean PUAs concentration (nanomolar level) in the background seawater. Anyway, we have clarified this and discussed them properly in the revised manuscript.

48. r. 391-392 I find it valuable to know if PUA stimulates bacterial activity whether as an organic substrate of metabolic signal substance. Please explain why only the latter would be ecologically important.

Response: We should note that PUAs accounts only for a small percentage of the organic carbon (<16%, Edwards et al., 2015). Also, PUAs can be toxic to some bacteria precluding its use as a carbon source. In addition, the specific arrangement of two double bonds and carbonyl chain make PUAs not a group of labile organic carbon for bacterial utilization. Therefore, it has less ecological importance as a carbon substrate.

49. r. 392-394 Please use the same concentration unit for comparability of levels. Heptadienal alone used for the test may not be comparable to a mixture of different PUAs (i.e. concentration more than twice used in the combined concentration). Why was not the same mixture used for this experiment? Other methods like using labelled PUA and analyse for metabolism of those would better test the mechanism of PUA effect.

Response: Agree. We have changed the unit (200 μ mol L⁻¹) in the revised manuscript. One reason for using heptadienal alone (C7) in the experiment is its lower toxicity compared to the other two (C8 and C10). Thus, C7 may be more likely used by bacteria if it can serve as a carbon source. Also, C7 is generally the dominant PUAs over the large area of our study regions (Wu and Li 2016). Therefore, we focus on C7 alone rather than the mixture of various PUAs to qualitatively assess the bioavailability of PUAs to bacteria. 50. r. 405-407 Please consider that a few species within the *Alteromonas phylum* may be responsible for the observed response. PUA metabolism might not be a function attributed to the whole phyla. Rephrase the discussion accordingly.

Response: We agree with the reviewer that various bacterial species within the genus *Altermonas* may respond differently to the PUAs treatments. We have revised the sentence as "...Our result was well consistent with the previous finding of the significant promotion effect of 13 or 106 μ mol L⁻¹ PUAs on *Alteromonas hispanica* from the pure culture experiment (Ribalet et al., 2008). An increase of PUAs could thus confer some of the γ -Pro (mainly special species within the genus *Alteromonas*, such as *A. hispanica*, Figure S2B) a competitive advantage over other bacteria ..."

51. r. 411-412 Please refer to what figure and test that show a difference between particle-attached and bulk bacteria.

Response: Ok. A figure has been provided in the supplement material to compare the community difference between particle-attached bacteria and the bulk bacteria (Figure S2A).

52. r. 427 How strongly are PUAs adsorbed to particles (i.e. chemical bonding)? How may this influence their potential to act as signalling molecules?

Response: It is still not well known about the mechanisms for PUAs adsorption on particles. PUAs may form a robust microzone around the particle, which would persist in the boundary layer and remain stable for some time (Juttner 2005).

Juttner, F. (2005) Evidence that Polyunsaturated Aldehydes of Diatoms are Repellents for Pelagic Crustacean Grazers, Aquatic Ecology. 39, 271-282.

53. r. 439-441 I have not seen any analyses of the lipoxygenase hypothesis in the study? It is thus speculative and should be removed. Focus on conclusion that can be derived from the performed study.

Response: Agree. The related sentences have been deleted in the revised manuscript.

54. r. 442-445 As there was no true replicates this conclusion should be made more cautious.

Response: As we have mentioned in our response to the previous point of this reviewer in No. 27, we did have two replicates for each treatment although more replicates are limited by the labor intensity of the experiment.

55. r.455-460 If this section should remain it need to be moved to the discussion section.

Response: Agree. The related sentences have been moved to the last part of the discussion section

in the revised manuscript.

56. r. 460-464 The sudden appearance of PUFA is not connected to the previous sentence?. Again, this part is highly speculative and not part of conclusion from the study. Remove or move parts to the discussion.

Response: Agree. The sentences have been moved to the discussion section in the revised manuscript. To avoid disconnection between them, we have revised the sentences as "...Eutrophication causes intense phytoplankton blooms in the coastal ocean. Sedimentation of the phytoplankton carbons will lead to their accumulation in the surficial sediment <u>(Cloern, 2001)</u>, <u>including PUFA compounds derived from the lipid production</u>. Resuspension and oxidation of these PUFA-rich organic particles during ..."

57. Literature cited

Cloern, J. E. 2001. Our evolving conceptual model of the coastal eutrophication problem. Mar. Ecol.-Prog. Ser. 210: 223-253, doi.

Robinson, C., and P. J. Le B Williams. 2005. Respiration and its measurement in surface marine waters, p. 147-180. In P. A. del Giorgio and P. J. Williams, le B [eds.], Respiration in aquatic ecosystems. Oxford University Press.

Response: Agree. The mentioned references have been cited in the revised manuscript.

Response to Anonymous Referee #2

1. However, there appeared to be some problems about the experimental design of this study and the manuscript fails to provide convincing results.

Response: We have carefully addressed the reviewer's comments on our experimental design and the related data and results. Please refer to our detailed responses to each of these specific comments below.

2. More details about the motivation and experiment procedure should be included and clarified and the conclusions should be carefully justified.

Response: We thank the reviewer for this suggestion. We have carefully rewritten these parts of the manuscript in the introduction and the method sections to clarify the motivation of our study and provide the details of the experiment setup, operational procedure, and relevant methodology. We have also rewritten the conclusion section as the review suggested. Please refer to the specific comments of this reviewer below for details of our revisions in each section.

3. My major concerns about this manuscript is that the authors consider PAB as bacteria attached to particles with a size >25 μ m in the microcosm incubation. A great variety of bacteria would be lost, which would affect the major conclusion of the manuscript. The abundance of bacteria attached to >25 μ m particles would be significantly lower than that of free-living bacteria.

Response: We completely agree with the reviewer that a complete PAB community should be acquired using a smaller filtration such as 0.8 μ m. Actually, in the high turbid estuarine waters of the PRE, PAB on the particle size of > 25 μ m could be only about 20 percent of the PAB on the particle size of > 2 μ m (Ge et al., 2020). However, we should emphasize that our primary goal is to explore the mechanism for PUAs affecting PAB variation and associated oxygen consumption in high turbidity and low oxygen regions of the PRE. Although the concentration of PUAs was nanomolar in the bulk water, it can reach a micromolar level on the surface of the particles where they are produced. The hotspot PUAs concentration associated with particles is defined as the PUAs concentration in the volume of the water parcel displaced by these particles. As PUAs can be accumulated on larger particles (Edwards et al., 2015), we chose particle aggregates of > 25 μ m to perform the PUA-amended experiments, in order to better explore the PUAs effects on PAB in the hypoxic waters. Future study may need to investigate PUAs impacts on PAB associated with the particle size of 0.8-25 μ m.

4. Also, is this particle size proper for measurement of polyunsaturated aldehydes?

Response: It is specific for particle-adsorbed PUAs on particles of > 25 μ m. A previous study by Edwards et al (2015) uses an even larger size of 50 μ m for collecting sinking particles for PAB and

the associated estimation of hotspot PUAs concentration.

5. In addition, it is not clear why the authors choose 1 or 100 μ mol L⁻¹ but not the background value for the incubation.

Response: We should emphasize that the concentration of PUAs in the water-column is inhomogeneous due to the presence of particles. Although the concentration of PUAs was nanomolar in the bulk water, it can reach a micromolar level on the surface of the particles where they are produced. The micromolar level of PUAs for incubation was chosen to represent the actual hotspot concentration of PUAs (the PUAs concentration in the volume of the water parcel displaced by the aggregation particles) not the mean PUAs concentration (nanomolar level) in the background seawater.

6. Moreover, bacterial community of the initial inoculates was lacking.

Response: In the revised manuscript, we have provided the initial bacterial community data (T=0) for the experiment in the supplementary material (Figure S2).

7. Personally, the most interesting part of this study is the role of polyunsaturated aldehydes-enhanced bacterial oxygen demand for the seasonal hypoxia. Thus, it is important to know to what extent different concentration of polyunsaturated aldehydes affect bacterial growth and respiration. Although the authors provide discussion on this, more analyses including the selection of background concentration of polyunsaturated aldehydes and testing on pure isolates are needed.

Response: We thank the reviewer for this suggestion. In the revised manuscript, we have provided a discussion on the impact of the background nanomolar level of PUAs on bacteria activity. "It should be mentioned that it remains controversial on the effect of background nanomolar PUAs on free-living bacteria, which is not our focus in this study. Previous studies suggested that 7.5 nmol L^{-1} PUAs would have a different effect on the metabolic activities of distinct bacterial groups in the NW Mediterranean Sea, although bulk bacterial abundance remained unchanged (Balestra et al., 2011). In particular, the metabolic activity of γ -Pro was least affected by nanomolar PUAs, although those of Bacteroidetes and Rhodobacteraceae were markedly depressed (Balestra et al., 2011). However, the daily addition of 1 nmol L^{-1} PUAs was found to not affect bacterial abundance and community composition during a mesocosm experiment in the Bothnian Sea (Paul et al., 2012)."

8. It is unclear why the authors studied the effect of polyunsaturated aldehydes of bacterial communities. The importance of polyunsaturated aldehydes was not properly and clearly presented. For examples, although the authors mentioned the effect of polyunsaturated aldehyde on marine microorganisms, detailed processes and mechanisms are not provided.

Response: We agree with the reviewer on this. In the revised manuscript, we have carefully rewritten this part by emphasizing the importance of PUAs on the microbial community and the associated mechanisms. "Phytoplankton-derived polyunsaturated aldehydes (PUAs) are known to affect marine microorganisms over various trophic levels by acting as infochemicals and/or by chemical defenses (Ribalet et al., 2008; Ianora and Miralto, 2010; Edwards et al., 2015; Franzè et al., 2018). PUAs are produced by stressed phytoplankters during the oxidation of membrane polyunsaturated fatty acids (PUFA) by lipoxygenase (Pohnert 2000) and are released from the surface of particles to the seawater by diffusion. The level of PUAs in the water-column are inhomogeneous, varying from sub-nanomolar offshore to nanomolar nearshore (Vidoudez et al., 2011; Wu and Li, 2016; Bartual et al., 2018), and to micromolar associated with particle hotspots (Edwards et al., 2015). The strong effect of PUAs on bacterial growth, production, and respiration has been well demonstrated in laboratory studies (Ribalet et al., 2008) and field studies (Balestra et al., 2011; Edwards et al., 2015). A perennial bloom of PUA-producing diatoms in the PRE mouth (Wu and Li, 2016) may indicate the relative importance of PUAs for microbial activity here compared to many other organic compounds, such as 2-n-pentyl-4-quinolinol (Long et al., 2003) and acylated homoserine lactones (Hmelo et al., 2011). A nanomolar level of PUAs recently reported in the coastal waters outside the PRE was hypothesized to affect oxygen depletion by promoting microbial utilization of organic matters in the bottom waters (Wu and Li, 2016). Meanwhile, the actual role of PUAs on bacterial metabolism within the bottom hypoxia remains largely unexplored."

9. Line 43: Please provide reference for the higher abundance of PAB compared to FLB.

Response: Done. The revised sentence is written as "In some coastal waters, PAB could be more abundant than the FLB with a higher metabolic activity to affect the coastal carbon cycle through organic matter remineralization (Garneau et al., 2009; Lee et al., 2015)."

10. Line 59-60: How polyunsaturated aldehyde affects marine microorganisms? Please provide more details.

Response: Done. We have rewritten the sentence as "Phytoplankton-derived polyunsaturated aldehydes (PUAs) are known to affect marine microorganisms over various trophic levels by acting as infochemicals and/or by chemical defenses (Ribalet et al., 2008; Ianora and Miralto, 2010; Edwards et al., 2015; Franzè et al., 2018). ".

11. Line 63: What is the meaning of affect oxygen depletion? Is this a promoting or inhibiting process?

Response: Done. The sentence has been rewritten as "... affect oxygen depletion by promoting microbial utilization of organic matters ...".

12. Line 72-73: It is strange to place this sentence here. Why PUAs did not serve as carbon source?

Response: Agree. We have moved the sentence to the method and result sections. Firstly, PUAs can be toxic to some bacteria precluding its use as a carbon source. Secondly, the specific arrangement of two double bonds and carbonyl chain make PUAs not a group of labile organic carbon for bacterial utilization. There were other studies supporting that PUAs could not serve as a carbon source for bacterial growth (Ribalet., 2008; Edwards et al., 2015).

13. Line 81: Is that "July 2^{nd} "?

Response: It is July 2nd. We have corrected this typo in the revised manuscript.

14. Line 88: what do you mean by "pPUAs and dPUAs"?

Response: Done. The abbreviations of pPUAs and dPUAs have been defined in the revised manuscript.

15. Line 265: I did not see description of methods about the bacterial community analysis in bottom waters of X1, X2, X3 and PAB on particles of >25 μ m.

Response: Done. We have added the method descriptions of these data to the section of 2.7.4 in the revised manuscript. "DNA samples for the bulk bacteria (>0.2 μ m) and PAB on particles of >25 μ m at station Y1 were also collected for bacterial community analysis using the same method described above. Methods for the bulk water bacterial community analyses at station X1, X2, and X3 during the 2016 cruise can be found in the published paper of Xu et al. (2018)."

16. Line 276: Please provide the concentration of pPUAs and dPUAs.

Response: Done. The mean concentrations of pPUAs and dPUAs have been provided in the revised manuscript.

17. Line 278: shown Line 391: According to the results, low dose (1 μ mol) of PUAs can stimulate the growth of PAB, significantly different from that of high dose (100 μ mol) treatment. However, the test of PUAs as organic carbon source was conducted with 200 μ mol of PUAs. I guess such a high concentration would adversely affect bacteria growth, while the low dose PUAs is likely to be used as organic sources.

Response: The 200 µM PUAs used in the test of carbon source possibility was to assure the same level of organic carbon substrate as those for ALK and PAHs. Bacteria may need a longer time and a higher substrate concentration to utilize these refractory organic matters (ALK and PAHs).

Although we have no test for the low-dose PUAs, the previous study has suggested that low-level PUAs (1 μ M and 10 μ M) were not used as a carbon source by bacteria (Edwards et al., 2015).

18. Line 686: data of panel D are reproduced from Ribalet et al., 2008. Is this panel E? No methods were provided for growth of *Alteromonas hispanica* MOLA151.

Response: Agree. It should be panel E. We have corrected this in the revised manuscript. We have also provided the growth method of A. hispanica MOLA151.

19. Line 448: Since bacteria on >25 μ m particles can be low, hypothesis on signaling molecules may be tuned down.

Response: Agree. We have rewritten this sentence as "... we hypothesize that PUAs may <u>potentially</u> act as signaling molecules for coordination among the high-density PAB below the salt-wedge, which would <u>likely</u> allow bacteria such as *Alteromonas* to thrive in degrading particulate organic matters. <u>Very possibly</u>, this process by changing community compositions and metabolic rates of PAB would lead to an increase of microbial oxygen utilization that <u>might</u> eventually contribute to the formation of coastal hypoxia."

1	Impacts of biogenic polyunsaturated aldehydes on metabolism and community
2	composition of particle-attached bacteria in coastal hypoxia
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12	
13	Abstract. Eutrophication-driven coastal hypoxia is of great interest recently for decades, though its
14	mechanisms are-remain not fully understood. Here, we showed elevated concentrations of particulate and
15	dissolved polyunsaturated aldehydes (PUAs) associated with the hypoxic waters meanly dominated by-
16	particle-attached bacteria (PAB) in the bottom water layer of a salt-wedge estuary. Bacterial respiration
17	within the hypoxic waters was mainly contributed by particle-attached bacteria (PAB) (>0.8 µm), with
18	free-living bacteria (0.2-0.8 µm) only accounting for 25-30 % of the total rate. The concentration of
19	particle-adsorbed PUAs (~10 μ mol L ⁻¹) in the hypoxic waters were directly quantified for the first time
20	based on large-volume-filtration and subsequent on-site PUAs derivation and extraction. PUAs-amended
21	incubation experiments for PAB (>25 μ m) retrieved from the low-oxygen waters were also performed to
22	explore the impacts of PUAs on the growth and metabolism of PAB and associated oxygen utilization. We
23	found an increase in cell growth of PAB in response to low-dose PUAs (1 μ mol L ⁻¹) but an enhanced
24	cell-specific bacterial respiration and production metabolic activity in response to high-dose PUAs (100

25	μ mol L ⁻¹)-including bacterial respiration and production. Improved cell-specific metabolism of PAB in
26	response to high-dose PUAs was also accompanied by a significantshift of PAB community structure
27	with increased dominance of genus Alteromonas within the Gammaproteobacteria. We thus conclude that a
28	high_PUAs concentration associated with aggregate particles within the bottom layer may be important
29	crucial for some species within Alteromonas to regulate PAB community structure. The change of bacteria
30	community could lead to an enhancement of oxygen utilization during the degradation of particulate
31	organic matters and thus likely contribute to the formation of coastal hypoxia. These findings are
32	potentially important for coastal systems with large river inputs, intense phytoplankton blooms driven by
33	eutrophication, as well as strong hypoxia developed below the salt-wedge front.

1. Introduction

35	Coastal hypoxia, defined as dissolved oxygen levels $< 62.5 \mu mol kg^{-1}$, has become a worldwide problem in
36	recent decades (Diaz and Rosenberg, 2008; Helm et al., 2011). It could affect diverse life processes from
37	genes to ecosystems, resulting in the spatial and temporal change of marine food-web structures (Breitburg
38	et al., 2018). Coastal deoxygenation <u>iswas</u> also tightly coupled with other global issues, such as <u>global</u>
39	warming and ocean acidification (Doney et al., 2012). Formation and maintenance of
40	eutrophication-derived hypoxia in the coastal waters should reflect the interaction between physical and
41	biogeochemical processes (Kemp et al., 2009). Generally, seasonal hypoxia occurs in the coastal ocean
42	when strong oxygen sinks are coupled with restricted resupply during periods of strong density
43	stratification. Termination of the event occurs with oxygen resupply when stratification is eroded by
44	vertical mixing (Fennel and Testa, 2019).
45	Bacterial respiration accounts for the largest portion of aquatic oxygen consumption and is thus pivotal
46	for the development of hypoxia and oxygen minimum zones (Williams and del Giorgio, 2005; Diaz and
47	Rosenberg, 2008). Generally, free-living bacteria (FLB) dominate the community respiration in many parts
48	of the ocean (Robinson and Williams, 2005; Kirchman, 2008). Compared to the FLB, the role of
49	particle-attached bacteria (PAB) on community respiration is less addressed, particularly in the coastal
50	oceans. In some coastal waters, PAB could be more abundant than the FLB with a higher metabolic activity
51	and mayto affect the coastal carbon cycle through organic matter remineralization (Garneau et al., 2009;
52	Lee et al., 2015). An increased contribution of PAB to respiration relative to FLB can occur during the
53	development of coastal phytoplankton bloom (Huang et al., 2018). In the Columbia River estuary, the
54	particle-attached bacterial activity could be 10-100 folds higher than that of its free-living counterparts
55	leading to its dominant role in organic detritus remineralization (Crump et al., 1998). Therefore, it is crucial
56	to assess the respiration process associated with PAB and its controlling factors in these regions, to fully
57	understand oxygen utilization in the hypoxic area with an intense supply of particulate organic matters.
58	There is an increasing area of seasonal hypoxia in the nearshore bottom waters of the Pearl River

59	Estuary (PRE) and the adjacent northern South China Sea (NSCS) (Yin et al., 2004; Zhang and Li 2010; Su
60	et al., 2017). The hypoxia is generally developed at the bottom of the salt-wedge where downward mixing
61	of oxygen is restrained due to increased stratification and where there is an accumulation of
62	eutrophication-derived organic matter due to flow convergence driven by local hydrodynamics (Lu et al.,
63	2018). Besides physical and biogeochemical conditions, aerobic respiration is believed the ultimate cause
64	of hypoxia here (Su et al., 2017). Thus, microbial respiration had been strongly related to the consumption
65	of bulk dissolved organic carbon in the PRE hypoxia (He et al., 2014).
66	Phytoplankton-derived polyunsaturated aldehydes (PUAs) are known to affect marine microorganisms
67	over various trophic levels by acting as infochemicals and/or by chemical defenses (Ribalet et al., 2008;
68	Ianora and Miralto, 2010; Edwards et al., 2015; Franzè et al., 2018). PUAs are produced by stressed
69	phytoplankters during the oxidation of membrane polyunsaturated fatty acids (PUFA) by lipoxygenase
70	(Pohnert 2000) and are released from the surface of particles to the seawater by diffusion. The level of
71	PUAs in the water-column are inhomogeneous, varying from sub-nanomolar offshore to nanomolar
72	nearshore (Vidoudez et al., 2011; Wu and Li, 2016; Bartual et al., 2018), and to micromolar associated with
73	particle hotspots (Edwards et al., 2015). The strong effect of PUAs on bacterial growth, production, and
74	respiration has been well demonstrated in laboratory studies (Ribalet et al., 2008) and field studies (Balestra
75	et al., 2011; Edwards et al., 2015). A perennial bloom of PUA-producing diatoms in the PRE mouth (Wu
76	and Li, 2016) may indicate the relative importance of PUAs for microbial activity here compared to many
77	other organic compounds, such as 2-n-pentyl-4-quinolinol (Long et al., 2003) and acylated homoserine
78	lactones (Hmelo et al., 2011). A nanomolar level of PUAs recently reported in the coastal waters outside
79	the PRE was hypothesized to affect oxygen depletion by promoting controlling microbial utilization of
80	organic matters in the bottom waters (Wu and Li, 2016). Meanwhile, the actual role of PUAs on bacterial
81	metabolism within the bottom hypoxia remains largely unexplored.
82	In this study, we <u>investigate</u> the particle-attached bacteria within the core of the hypoxic waters by
83	exploring the linkage between PUAs and bacterial oxygen utilization on the suspended organic particles.

84	There are three specific questions to address here: What are the relative roles of PAB and FLB on bacterial
85	respiration in the hypoxic waters? What are the actual levels of PUAs in the hypoxic waters? What are the
86	responses of PAB to PUAs in the hypoxic waters? For the first question, size-fractionated bacterial
87	respiration rates were estimated for both FLB (0.2-0.8 µm) and PAB (>0.8 µm) in the hypoxic waters. For
88	the second question, the concentrations of particulate and dissolved PUAs within the hypoxic waters were
89	measured in the field. Besides, the hotspot PUAs concentration associated with the suspended particles
90	within the hypoxic waters were was directly quantified for the first time using large-volume filtration and
91	subsequent on-site derivation and extraction. For the third question, field PUAs-amended incubation
92	experiments were conducted for PAB (>25 μ m) retrieved from the low-oxygen waters. We focused on
93	particles of >25 µm to better explore the role of PUAs on PAB given the accumulations of PUAs on large
94	aggregates. The doses of PUAs treatments were selected to represent the actual levels of PUAs hotspots, to
95	assess the PAB responses (including bacterial abundance, respiration, production, and community
96	composition) to the exogenous PUAs in the hypoxic waters. An additional experiment was also performed
97	to verify that the observed effects of PUAs on PAB were not due to an increase of carbon source. By
98	synthesizing these-fieldexperimental results with the change of water-column biogeochemistry-of the-
99	hypoxic zone, we hope to explore the underlying mechanism for particle-adsorbed PUAs influencing on
100	community structure and metabolism of PAB in the low-oxygen waters, as well as to understand its
101	contribution to coastal deoxygenation of the NSCS shelf-sea.
102	

103 **2. Methods**

104 **2.1 Descriptions of field campaigns and sampling approaches**

Field survey cruises were conducted in the PRE and the adjacent NSCS during June 17th-28th, 2016 and June 18st-JuneJuly 2nd, 2019 (Figure 1). Briefly, vertical profiles of temperature, salinity, dissolved oxygen, and turbidity were acquired from a Seabird 911 rosette sampling system. The oxygen sensor data were corrected by field titration measurements during the cruise. Water samples at various depths were collected

109	using 6 or 12 liters (12 or 24 positions) Niskin bottles attached to the Rosette sampler. Surface water
110	samples were collected at ~1m or 5 m depth, while bottom water samples were obtained at depths ~4 m
111	above the bottom. Chlorophyll-a (Chl-a) samples were taken at all depths at all stations and nutrients were
112	also sampled except at a few discrete stations. For the 2016 cruise, samples for pPUAs were collected at all
113	depths close to station X1 (Figure 1A). During the summer of 2019, vertical profiles of particulate PUAs
114	(pPUAs) and dissolved PUAs (dPUAs) were determined at Y1 in the hypoxic zone and Y2 outside the
115	hypoxic zone with field PUAs-amended experiments conducted at Y1 (Figure 1B). For station Y1, the
116	middle layer was defined as 12 m with the bottom layer as 25 m. At this station, samples at different depths
117	were collected for determining the size-fractionated respiration rates and the whole water bacterial
118	taxonomy.

120 **2.2 Determination of chlorophyll-***a*, dissolved nutrients

For Chl-*a* analyses, 500 mL of water sample was gently filtered through a 0.7 μm Whatman GF/F filter.
The filter was then wrapped by a piece of aluminum foil and stored at -20 °C on board. Chl-*a* was extracted
at 4 °C in the dark for 24 h using 5 mL of 90% acetone. After centrifuged at 4000 rpm for 10 min, Chl-*a*was measured using a standard fluorometric method with a Turner Designs fluorometer (Parsons et al.,
1984). Water samples for nutrients were filtered through 0.45 μm Nucleopore filters and stored at -20 °C.
Nutrient concentrations including nitrate plus nitrite, phosphate, and silicate were measured using a
segmented-flow nutrient autoanalyzer (Seal AA3, Bran-Luebbe, GmbH).

128

129 **2.3 Sampling and measurements of particulate and dissolved PUAs in one-liter seawater**

130 We used a similar protocol of Wu and Li (2016) for pPUAs and dPUAs collection, pretreatment, and

determination. Briefly, 2-4 liters of water sample went through a GF/C filtration with both the filter and the

132 filtrate collected separately. The filter was rinsed by the derivative solution with the suspended particle

133 samples collected in a glass vial. After adding internal standard, the samples in the vial were frozen and

thawed three times to mechanically break the cells for pPUAs. The filtrate from the GF/C filtration was also added with internal standard and transferred to a C18 solid-phase extraction cartridge. The elute from the cartridge with the derivative solution was saved in a glass vial for dPUAs. Both pPUAs and dPUAs samples were frozen and stored at -20 °C.

In the laboratory, the pPUAs sample was thawed with the organic phase extracted. After the solvent was evaporated with the sample concentrated and re-dissolved in hexane, pPUAs was determined using gas chromatography and mass spectrometry (Agilent Technologies Inc., USA). Standards series were prepared by adding certain amounts of three major PUAs to the derivative solution and went through the same pretreatment and extraction steps as samples. Derivatives of dPUAs were extracted and measured by similar methods as pPUAs, except that the calibration curves of dPUAs were constructed separately. The units of pPUAs and dPUAs are nmol L⁻¹ (nmol PUA in one-liter seawater).

145

146 **2.4 Particle collections by large-volume filtrations in hypoxia waters.**

147 Large volumes (~300 L) of the middle (12 m) and the bottom (25 m) waters within the hypoxia zone were collected by Niskin bottles at station Y1. For each layer, the water sample was quickly filtered through a 148 sterile fabric screen (25 µm filter) on a disk filter equipped with a peristaltic pump to qualitatively obtain 149 particles of $>25 \,\mu\text{m}$. Larger zooplankters were picked off immediately. The particle samples were gently 150 back-flushed three times off the fabric screen using particle-free seawater (obtained using a 0.2 µm 151 152 filtration of the same local seawater) into a sterile 50-mL sampling tube. The volume of total particles from large-volume-filtration was measured as follows: The collected 153 particle in the 50 mL tube was centrifuged for one minute at a speed of 3000 revolutions per minute (r.p.m) 154 155 with the supernatant removed saved (Hmelo et al., 2011). The particle sample was resuspended as slurry by 156 gently shaking and transferred into a sterile 5 mL graduated centrifuge tube. The sample was centrifuged 157 again by the same centrifuging speed with the final volume of the total particles recorded. The unit for the

158 total particle <u>volume</u> is mL.

- All the particles were transferred back to the sterile 50 mL centrifuge tube (so as all the supernatants)
 with 0.2-µm-filtered seawater, which was used for subsequent measurements of particle-adsorbed PUAs as
 well as for PUAs-amended incubation experiments of particle-attached bacteria.
- 162

163 2.5 Measurements of particle-adsorbed PUAs

After gently shaking, 3 mL of sample in the 50 mL sampling tube (see section 2.4) was used for the 164 analyses of particle-adsorbed PUAs concentration (two replicates) according to the procedure shown in 165 Figure 2 (modified from the protocols of Edwards et al. 2015 and Wu and Li 2016). The sample (3 mL) 166 was transferred to 50 mL centrifuge tubes for PUAs derivatization on board. An internal standard of 167 benzaldehyde was added to obtain a final concentration of 10 µM. The aldehydes in the samples were 168 derivatized by the addition of O-(2,3,4,5,6-pentafluorobenzyl) hydroxylamine hydrochloride solution in 169 deionized water (pH=7.5). The reaction was performed at room temperature for 15 min (shaking slightly 170 for mix every 5 min). Then 2 mL sulfuric acid (0.1%) solution was added to a final concentration of 0.01% 171 172 acid (pH of 2-3) to avoid new PUAs induced by enzymatic cascade reactions. The derivate samples were subsequently sonicated for 3 min before the addition of 20 mL hexane, and the upper organic phase of the 173 extraction was transferred to a clean tube and stored at -20 °C. 174 Upon returning to the laboratory, the adsorbed PUAs on these particles (undisrupted PUAs) were 175

determined with the same analytical methods as those for the disrupted pPUAs (freeze-thaw methods to 176 include the portion of PUAs eventually produced as cells die, Wu and Li 2016) except for the freeze-thaw 177 step. A separate calibration curve was made for the undisrupted PUAs derivates. A standard series of 178 heptadienal, octadienal, and decadienal $(0, 0.1, 0.5, 1.0, 2.5, 5.0, 10.0, 25.0 \text{ nmol } \text{L}^{-1})$ was prepared before 179 each analysis by diluting a relevant amount of the PUA stock solution (methanolic solution) with deionized 180 181 water. These standard solutions were processed through all the same experimental steps as those mentioned above for derivation, extraction, and measurement of the undisrupted PUAs sample. The unit for the 182 undisrupted PUAs is nmol L⁻¹. The total amount of the undisrupted PUAs in the 50 mL sampling tube was 183

184 <u>the product of the measured concentration and the total volume of the sample.</u>

185 The hotspot PUAs concentration associated with the aggregate particles is defined as the PUAs

186 <u>concentration in the volume of the water parcel displaced by these particles. Therefore, the</u> final

187 <u>concentration of particle-adsorbed PUAs in the water column</u>, defined as PUAs [µmol L⁻¹], <u>should be</u> equal
 188 to the moles of particle-adsorbed PUAs (nmol, <u>the undisrupted PUAs</u>) divided by the volume of particles
 189 (mL).

190

191 **2.6 Incubation of particle-attached bacteria with PUAs treatments.**

192 Impact The impact of PUAs on microbial growth and metabolisms in the hypoxia zone was assessed by field incubation of particle-attached bacteria on particles of $> 25 \,\mu m$ collected from large-volume filtration 193 with direct additions of low or high doses of PUAs (1 or 100 μ mol L⁻¹) on June, 29th, 2019 (Figure 2). 194 A sample volume of ~32 mL in the centrifuge tube (section 2.4) was transferred to a sterile elean-195 Nalgene bottle before being diluted by particle-free seawater to a final volume of 4 L. About 3.2 L of the 196 197 sample solution was transferred into four sterileclean 1-L Nalgene bottles (each with 800 mL). One 1-L bottle was used for determining the initial conditions: after gentle shaking, the solution was transferred into 198 six biological oxygen demand (BOD) bottles with three for initial oxygen concentration (fixed immediately 199 by Winkler reagents) and the other three for initial bacterial abundance, production, and community 200 structure. The other three 1-L bottles were used for three different treatments (each with two replicates in 201 202 two 0.5-L bottles): the first one served as the control with the addition of 200 μ L methanol, the second one with 200 μ L low-dose PUAs solution, and the third one with 200 μ L high-dose PUAs solution (Table 1). 203 The solution in each of the three treatments (0.5-L bottles) was transferred to six parallel replicates by 204 205 60-mL BOD bottles. These BOD bottles were incubated at *in situ* temperature in the dark for 12 hours. At 206 the end of each incubation experiment, three of the six BOD bottles were used for determining the final 207 oxygen concentrations with the other three for the final bacterial abundance, production, and community 208 structure.

209	To test the possibility of PUAs as carbon sources for bacterial utilization, a minimal medium was
210	prepared with only sterile artificial seawater but not any organic carbons (Dyksterhouse et al., 1995). A
211	volume of 375 μ L sample (from the above 4 L sample solution) was inoculated in the minimal medium
212	amended with heptadienal in a final concentration of about <u>200 μ</u> mol L ⁻¹ . <u>This PUA level was close to the</u>
213	hotspot PUAs of 240 μ mol L ⁻¹ found in the suspended particles of a station near the PRE. It was also
214	comparable to the hotspot PUAs of 25.7 μ mol L ⁻¹ in the temperate west North Atlantic (Edwards et al.,
215	2015). For comparisons, the same amount of sample was also inoculated in the minimal medium (75 mL)
216	amended with an alkane mixture (ALK, n-pentadecane and n-heptadecane) at a final concentration of 0.25
217	g L^{-1} , or with a mixture of polycyclic aromatic hydrocarbons (PAH, naphthalene and phenanthrene) at a
218	final concentration of 200 ppm. These experiments were performed in dark at room temperature for over 30
219	days. Significant turbidity changes in the cell culture bottle over incubation time will be observed if there is
220	a carbon source for bacterial growth.
221	
222	2.7 Measurements of bacteria-related parameters
223	(1) Bacterial abundance
224	At the end of the 12-h incubation period, a 2 mL sample from each BOD bottle was preserved in 0.5%
225	glutaraldehyde. The fixation lasted for half of an hour at room temperature before being frozen in liquid N_2
226	
227	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously
	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved
228	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved effective for samples with high particle concentrations. To <u>break up particles and account for detaching</u> -
228 229	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved effective for samples with high particle concentrations. To <u>break up particles and account for detaching</u> <u>attached bacteria due to breaking up particles</u> , 0.2 mL pure methanol was added to the 2 mL sample and
228 229 230	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved effective for samples with high particle concentrations. To <u>break up particles and account for detaching</u> - <u>attached bacteria due to breaking up particles</u> , 0.2 mL pure methanol was added to the 2 mL sample and vortexed. The sample was then incubated in an ultrasonic bath (35 kHz, 2 x 320W per period) at 35 °C for
228229230231	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved effective for samples with high particle concentrations. To <u>break up particles and account for detaching</u> - <u>attached bacteria due to breaking up particles</u> , 0.2 mL pure methanol was added to the 2 mL sample and vortexed. The sample was then incubated in an ultrasonic bath (35 kHz, 2 x 320W per period) at 35 °C for 15 min. Subsequently, the tube sample was filtered with <u>a</u> 50 µm-filter to remove large detrital particles.
 228 229 230 231 232 	and stored in a -80 °C freezer. In the laboratory, the samples were performed through a previously published procedure for detaching particle-attached bacteria (Lunau et al., 2005), which had been proved effective for samples with high particle concentrations. To break up particles and account for detaching- attached bacteria due to breaking up particles, 0.2 mL pure methanol was added to the 2 mL sample and vortexed. The sample was then incubated in an ultrasonic bath (35 kHz, 2 x 320W per period) at 35 °C for 15 min. Subsequently, the tube sample was filtered with <u>a</u> 50 µm-filter to remove large detrital particles. The filtrate samples for surface-associated bacteria cells were diluted by 5-10 folds using TE buffer

234	addition of 1-µm beads, bacterial abundance (BA) of the samples was counted by a flow cytometer
235	(Beckman Coulter CytoFlex S) with bacteria detected on a plot of green fluorescence versus side scatter_
236	(Marie et al., 1997). The precision of the method estimated by the coefficient of variation (CV%) was
237	generally less than 5%.
238	For bulk-water bacteria abundance (including both FLB and PAB), 1.8 mL of seawater sample was
239	collected after a 20-µm prefiltration. The sample was transferred to a 2 mL centrifuge tube and fixed by
240	adding 20 μ L of 20% paraformaldehyde before storage in a -80 °C freezer. In the laboratory, 300 μ L of the
241	sample after thawing was used for staining with SYBR Green and analyzed using the same flow cytometry
242	method as above (Marie, et al, 1997).
243	
244	(2) Bacterial respiration
245	For BOD samples, bacterial respiration (BR) was calculated based on the oxygen decline during the 12-h
246	incubation and was converted to carbon units with the respiratory quotient assumed equal to 1 (Hopkinson,
247	1985). Dissolved oxygen was determined by a high-precision Winkler titration apparatus (Metrohm-848,
248	Switzerland) based on the classic method (Oudot et al., 1988). We should mention that BR could be
249	overestimated if phytoplankton and microzooplankton were present in the particle aggregates of $> 25 \ \mu m$.
250	However, this effect could be relatively small because the raw seawater in the hypoxic zone had very low
251	chlorophyll-a and because there was virtually not much microzooplankton in the sample (confirmed by
252	FlowCAM).
253	Method for the estimation of the bulk water bacterial respiration at stations X1, X2, and X3 can be
254	found in Xu et al (2018). For the bulk water at station Y1, the size-fractionated respiration rates, including
255	free-living bacteria of 0.2-0.8 µm and particle-associated community of >0.8 µm (we assumed that they
256	were mostly PAB given the low phytoplankton chlorophyll- <i>a</i> of the sample and the absence of zooplankton
257	during the filtration), were estimated based on the method of García-Martín et al (2019). Four 100 mL
258	polypropylene bottles were filled with seawater. One bottle was immediately fixed by formaldehyde. After

- 15 min, the sample in each bottle was incubated in the dark at the *in situ* temperature after the addition of
 the Iodo-Nitro-Tetrazolium (INT) salt at a final concentration of 0.8 mmol L⁻¹. The incubation reaction
 lasted for 1.5 h before being stopped by formaldehyde. After 15 min, all the samples were sequentially
 filtered through 0.8 and 0.2 µm pore size polycarbonate filters and stored frozen until further measurements
 by spectrophotometry.
- 264

265 (3) Bacterial production

Bacterial production (BP) was determined using a modified protocol of the ³H-leucine incorporation 266 method (Kirchman, 1993). Four 1.8-mL aliquots of the sample were collected by pipet from each BOD 267 incubation and added to 2-mL sterile microcentrifuge tubes, which were incubated with ³H-leucine (in a 268 final concentration of 4.65 μ mol Leu L⁻¹, Perkin Elmer, USA). One tube served as the control was fixed by 269 adding 100% trichloroacetic acid (TCA) immediately (in a final concentration of 5%). The other three were 270 terminated with TCA at the end of the 2-h dark incubation. Samples were filtered onto 0.2-µm 271 272 polycarbonate filters and then rinsed twice with 5% TCA and three times with 80% ethanol (Huang et al., 2018) before being stored at -80 °C. In the laboratory, the filters were transferred to scintillation vials with 273 5 mL of Ultima Gold scintillation cocktail. The incorporated ³H was determined using a Tri-Carb 2800TR 274 275 liquid scintillation counter. Bacterial production was calculated with the previous published leucine-to-carbon empirical conversion factors of 0.37 kg C mol leucine⁻¹ in the study area (Wang et al., 276 2014). Bacterial carbon demand (BCD) was calculated as the sum of BP and BR. Bacterial growth 277 efficiency (BGE) was equated to BP/BCD. 278

279

280 (4) Bacterial community structure

At the end of incubation, <u>the DNA sample was obtained by filtering 30 mL of each BOD water via a</u>
0.22-µm Millipore filter, which was preserved in a cryovial with the DNA protector buffer and stored at
-80 °C. DNA was extracted using the DNeasy PowerWater Kit with genomic amplification by Polymerase

Chain Reaction (PCR). The V3 and V4 fragments of bacterial 16S rRNA were amplified at 94 °C for 2 min 284 and followed by 27 cycles of amplification (94 °C for 30 s, 55° C for 30 s, and 72 °C for 60 s) before a 285 286 final step of 72 °C for 10 min. Primers for amplification included 341F (CCTACGGGNGGCWGCAG) and 805R (GACTACHVGGGTATCTAATCC). Reactions were performed in a 10-uL mixture containing 1 uL 287 Toptag Buffer, 0.8 μ L dNTPs, 10 μ M primers, 0.2 μ L Tag DNA polymerase, and 1 μ L Template DNA. 288 289 Three parallel amplification products for each sample were purified by an equal volume of AMpure XP magnetic beads. Sample libraries were pooled in equimolar and paired-end sequenced (2×250 bp) on an 290 291 Illumina MiSeq platform.

High-quality sequencing data was obtained by filtering on the original off-line data. Briefly, the raw 292 data was pre-processed using TrimGalore to remove reads with qualities of less than 20 and FLASH2 to 293 294 merge paired-end reads. Besides, the data were also processed using Usearch to remove reads with a total base error rate of greater than 2 and short reads with a length of less than 100 bp and using Mothur to 295 296 remove reads containing more than 6 bp of N bases. We further used UPARSE to remove the singleton 297 sequence to reduce the redundant calculation during the data processing. Sequences with similarity greater than 97% were clustered into the same operational taxonomic units (OTUs). R software was used for 298 community composition analysis. 299

300 DNA samples for the bulk bacteria (>0.2 μm) and PAB on particles of > 25 μm at station Y1 were also
 301 collected for bacterial community analysis using the same method described above. Methods for the bulk
 302 water bacterial community analyses at stations X1, X2, and X3 during the 2016 cruise can be found in the
 303 published paper of Xu et al. (2018).

304

305 2.8 Statistical Analysis

306 All statistical analyses were performed using the statistical software SPSS (Version 13.0, SPSS Inc.,

307 Chicago, IL, USA). A student's t-test with a 2-tailed hypothesis was used when comparing PUAs-amended

308 treatments with the control or comparing stations inside and outside the hypoxic zone, with the null

309 hypothesis being rejected if the probability (p) is less than 0.05. We consider p of <0.05 as significant and p

of <0.01 as strong significant. Ocean Data View with the extrapolation model "DIVA Gridding" method

311 was used to contour the spatial distributions of physical and biogeochemical parameters.

312

313 **3. Results**

314 **3.1** Characteristics of hydrography, biogeochemistry, and bulk bacteria community in the hypoxic

315 **zone**

During our study periods, there was a large body of low oxygen bottom water with the strongest hypoxia (< 316 62.5 μ mol kg⁻¹) on the western shelf of the PRE (Figure 1), which was relatively similar constant among 317 different summers of 2016 and 2019 (Figure 1). For vertical distribution, a strong salt-wedge structure was 318 319 found over the inner shelf (Figures 3A, 3D) with freshwater on the shore side due to intense river discharge. Bottom waters with oxygen deficiency (< 93.5 µmol kg⁻¹) occurred below the lower boundary of the 320 salt-wedge and expanded ~60 km offshore (Figure 3E). In contrast, a surface high Chl-a patch (6.3 μ g L⁻¹) 321 322 showed up near the upper boundary of the front, where there was enhanced water-column stability, low turbidity, and high nutrients (Figures 3B, 3C). Therefore, there was a spatial mismatch between the 323 subsurface hypoxic zone (Figure 3E) and the surface chlorophyll-bloom (Figure 3F) during the 324 estuary-to-shelf transect, as both the surface Chl-a and oxygen right above the hypoxic zones at the bottom 325 boundary of the salt-wedge were not themselves maxima. 326

There were much higher rates of respiration (BR) (t=7.8, n=9, p<0.01) and production (BP) (t=13.0, n=9, p<0.01) for the bulk bacterial community (including FLB and PAB) in the bottom waters of X1 within the hypoxic core than those of X2 and X3 outside the hypoxic zone during June 2016 (Figure 4, modified from data of Xu et al., 2018). The size-fractionated respiration rates were quantified at station Y1 during the 2019 cruise (Figure S1) to distinguish the different roles of FLB and PAB on bacterial respiration in the hypoxic waters. Our results suggested that bacterial respiration within the hypoxic waters was largely. contributed by PAB (>0.8 µm), which was about 2.3-3 folds of that by FLB (0.2-0.8 µm).

334	<u>The bulk bacterial composition of the bottom water of X1 during the 2016 cruise with 78% of</u>
335	α -Proteobacteria (α -Pro), 15% of γ -Proteobacteria (γ -Pro), and 6% of Bacteroidetes was significantly
336	different from those of X2 and X3 (91% α -Pro, 5% γ -Pro, and 2% Bacteroidetes), although their bacterial
337	abundances were about the same (Figure 4). These pointed to the importance of γ -Pro (mainly genus-
338	Alteromonadaceae and Pseudoalteromonaceae) and Bacteroidetes (mainly genus Flavobacteriaceae) in the
339	low-oxygen waters (genus data not shown). Compared to that of the 2016 cruise, there was a different
340	taxonomic composition of <u>the</u> bulk bacterial community <u>in</u> the hypoxic waters <u>of</u> the 2019 cruise with on
341	average 33% of α -Pro, 25% of γ -Pro, and 14% of Bacteroidetes. <u>Furthermore</u> , relative to the bulk bacterial-
342	community of the hypoxic waters, there was <u>a</u> substantially different taxonomic composition <u>for</u> PAB on -
343	particles of (>25 μ m) with on average 66% of γ -Pro, 22% of α -Pro, and 4% of BacteroidetesFlavobacteria_
344	(Figure S2A). In particular, there was an increase of γ -Pro, but a decrease of α -Pro and Bacteroidetes, in the
345	PAB (>25 μm) relative to the bulk bacterial community. On the genus level, the PAB (>25 μm) was largely
346	dominated by the Alteromonas group in both the middle and bottom waters (Figure S2B).

348 **3.2 PUAs concentrations in the hypoxic zone**

Generally, there were significantly higher pPUAs of 0.18 nmolL⁻¹ (t=3.20, n=10, p<0.01) and dPUAs of 349 0.12 nm<u>ol L⁻¹ (t=7.61, n=8, p<0.01)</u> in the hypoxic waters than in the nearby bottom waters without 350 hypoxia (0.02 nmol L^{-1} and 0.01 nmol L^{-1}). Vertical distributions of pPUAs and dPUAs in the bulk seawater 351 were showed for two stations (Y1 and Y2) inside and outside the hypoxic zone (Figure 1). Nanomolar 352 levels of pPUAs and dPUAs were found in the water column in both stations (Figures 5E, 5F). There were 353 354 high pPUAs and dPUAs in the bottom hypoxic waters of station Y1 (Figure 5E, 5F) together with locally elevated turbidity (Figure 3B) when compared to the bottom waters outside, which likely a result of particle 355 resuspension. For station Y2 outside the hypoxia, we found negligible pPUAs and dPUAs at depths below 356 357 the mixed layer (Figure 5E, 5F), which could be due to PUAs dilution by the intruded subsurface seawater. Particle-adsorbed PUAs in the low-oxygen waters were quantified for the first time with the direct 358

particle volume estimated by large-volume-filtration (see the method section), which would reduce the uncertainty associated with particle volume calculated by empirical equations derived for marine-snow particles (Edward et al., 2015). We found high levels of particle-adsorbed PUAs (~10 μ mol L⁻¹) in these waters (Figure 6), which were orders of magnitude higher than the bulk water pPUAs or dPUAs concentrations (<0.3 nmol L⁻¹, Figure 5E, 5F). Particle-adsorbed PUAs of the low-oxygen waters mainly consisted of heptadienal (C7_PUA) and octadienal (C8_PUA), with decadienal (C10_PUA) making uponly a small percentage.

366

367 3.3 Particle-attached bacterial growth and metabolism in the hypoxic zone

Incubation of the PAB acquired from the low-oxygen waters with direct additions of different doses of 368 exogenous PUAs over a period of 12 hours was carried out to examine the change of bacterial growth and 369 370 metabolism activities in response to PUA-enrichments. At the end of the incubation experiments, BA was not different from the control for the PH treatment (Figure 7A). However, for the PL treatment, there were 371 372 substantial increases of BA in both the middle and the bottom waters compared to the initial conditions (Figure 7A). In particular, BA of $\sim 3.2 \pm 0.04 \times 10^9$ cells L⁻¹ in the bottom water for the PL treatment was 373 significantly higher (<u>t=12.26</u>, <u>n=12</u>, p<0.01) than the control of $2.5 \pm 0.07 \times 10^9$ cells L⁻¹. 374 375 BR was significantly promoted by the low-dose PUAs with a 21.6% increase in the middle layer

(t=11.91, n=8, p<0.01) and a 25.8% increase in the bottom layer (t=11.50, n=8, p<0.01) compared to the

377 controls. Stimulating effect of high-dose PUAs on BR was even stronger with 47.0% increase in the middle

378 layer (t=30.56, n=8, p<0.01) and 39.8% increase in the bottom layer (t=9.40, n=8, p<0.01) (Figure 7B).

379 Meanwhile, the cell-specific BR was significantly improved for both layers with high-dose of PUAs

380 (t=15.13, n=8, p<0.01 and t=4.77, n=8, p<0.01), but not with low-dose of PUAs (Figure 7C) due to

- increase of BA (Figure 7A). BGE was generally very low (<1.5%) during all the experiments (Figure 7D)
- due to substantially high rates of BR (Figure 7B) than BP (Figure 7E). Also, there was no significant
- 383 difference in BGE between controls and PUAs treatments for both layers (Figure 7D).

384	For the bottom layer, BP was $12.6 \pm 0.8 \ \mu g \ C \ L^{-1} \ d^{-1}$ for low-dose PUAs and $16.4 \pm 0.6 \ \mu g \ C \ L^{-1} \ d^{-1}$
385	for high-dose PUAs, which were both significantly ($\underline{t=2.98}, \underline{n=8}, p < 0.05$ and $\underline{t=10.41}, \underline{n=8}, p < 0.01$) higher
386	than the control of $10.6 \pm 0.6 \ \mu g \ C \ L^{-1} \ d^{-1}$. Meanwhile, BP in the middle layer was significantly higher
387	(<u>t=2.52</u> , <u>n=8</u> , <u>p<0.05</u>) than the control for high-dose PUAs ($13.4 \pm 0.9 \ \mu g \ C \ L^{-1} \ d^{-1}$) but not for low-dose
388	PUAs $(12.6 \pm 0.9 \ \mu g \ C \ L^{-1} \ d^{-1})$ (Figure 7E). The cell-specific BP (sBP, 7.9 \pm 0.5 and 6.9 \pm 0.2 fg C cell ⁻¹ d ⁻¹)
389	for high-dose PUAs were significantly ($\underline{t=2.62}, \underline{n=8}, \underline{p<0.05}$ and $\underline{t=11.26}, \underline{n=8}, \underline{p<0.01}$) higher than the
390	control in both layers (Figure 7F). Meanwhile, for low-dose PUAs, the sBP in both layers were not
391	significantly different from the controls.

393 3.4 Particle-attached bacterial community change during incubations

Generally, γ -Pro dominated (>68%) the bacterial community at the class level for all experiments, followed by the second largest bacterial group of α -Pro. There was a <u>large-significant</u> increase of γ -Pro by high-dose PUAs with increments of 17.2% (*t*=9.25, *n*=8, *p*<0.01) and 19.5% (*t*=6.32, *n*=8, *p*<0.01) for the middle and the bottom layers, respectively (Figure 8A). However, there was no significant substantial change of bacterial community composition by low-dose PUAs for both layers (Figure 8A).

399 On the genus level, there was also a large difference in the responses of various bacterial subgroups to the exposure of PUAs (Figure 8B). The main contributing genus for the promotion effect by high-dose 400 PUAs was the group of Alteromonas spp., which showed a large increase in abundance by 73.9% and 401 402 69.7% in the middle and the bottom layers. For low-dose PUAs, the promotion effect of PUAs on 403 Alteromonas spp. was still found although with a much lower intensity (5.4% in the middle and 19.4% in the bottom). The promotion effect of γ -Pro by high-dose PUAs was also contributed by bacteria Halomonas 404 405 spp. (percentage increase from 1.7% to 7.4%). Meanwhile, some bacterial genus, such as Marinobacter and 406 *Methylophaga* from γ -Pro, or *Nautella* and *Sulfitobacter* from α -Pro, showed decreased percentages by 407 high-dose PUAs (Figure 8B).

408

409 **3. 5 Carbon source preclusion experiments for PUAs**

After one month of incubation, PAB inoculated from the low-oxygen waters showed dramatic responses to both PAH and ALK (Figure 9). In particular, the mediums of PAH addition became turbid brown (bottles on the left) with the medium of ALK addition turning into milky white (bottles in the middle) (Figures 9B and 9D). For comparison, they were both clear and transparent at the beginning of the experiments (Figures 9A and 9C). These results should reflect the growth of bacteria in these bottles with the enrichments of organic carbons. Meanwhile, the minimal medium with the addition of heptadienal (C7_PUA) remained clear and transparent as it was originally, which would indicate that PAB did not grow in the treatment of C7_PUA.

417

418 **4. Discussion**

Hypoxia occurs if the rate of oxygen consumption exceeds that of oxygen replenishment by diffusion,
mixing, and advection (Rabouille et al., 2008). The spatial mismatch between the surface chlorophyll-*a*maxima and the subsurface hypoxia during our estuary-to-shelf transect should indicate that the
low-oxygen feature may not be directly connected to particle export by the surface phytoplankton bloom.
This outcome can be a combined result of riverine nutrient input in the surface, water-column stability
driven by wind and buoyancy forcing, and flow convergence for an accumulation of organic matters in the
bottom (Lu et al., 2018).

Elevated concentrations of pPUAs and dPUAs near the bottom boundary of the salt-wedge should 426 427 reflect a sediment source of PUAs, as the surface phytoplankton above them was very low. PUA-precursors such as polyunsaturated fatty acids (PUFA) could be accumulated as detritus in the surface sediment near 428 the PRE mouth during the spring blooms (Hu et al., 2006). Strong convergence at the bottom of the 429 430 salt-wedge could be driven by shear vorticity and topography (Lu et al., 2018). This would allow for the 431 resuspension of small detrital particles. Improved PUAs production by oxidation of the resuspended PUFA 432 could occur below the salt-wedge as a result of enhanced lipoxygenase activity (in the resuspended organic 433 detritus) in response to salinity increase by the intruded bottom seawater (Galeron et al., 2018).

434	Direct measurement of the adsorbed PUAs concentration associated with the suspended particles
435	<u>of >25 μm</u> by <u>the method of combined</u> large-volume filtration and on-site derivation and extraction_yield a
436	high level of ~10 μ mol L ⁻¹ within the hypoxic zone. This value is comparable to those previously reported
437	in sinking particles (>50 μ m) of the open ocean using particle-volume calculated from diatom-derived
438	marine snow particles (Edward et al., 2015). Note that there was also a higher level of $\frac{240}{\mu}$ µmol L ⁻¹ found
439	in <u>another</u> station outside the PRE. Compared to the nanomolar levels of dPUAs and pPUAs in the waters,
440	a <u>A</u> micromolar level of particle-adsorbed PUAs could act as a hotspot for bacteria likely exerting important
441	impacts as signaling molecules on microbial utilization of particulate organic matters and subsequent
442	oxygen consumption.
443	The hypoxic waters below the salt-wedge have high turbidity probably due to particle resuspension.
444	High particle concentration here may indicate the important role of PAB, which could be more abundant
445	than_FLB in the turbid waters near the mouth of the PRE (Ge et al., 2020), similar to those found in the
446	Columbia River estuary (Crump et al., 1998). Also, anaerobic bacteria and taxa preferring low-oxygen
447	conditions were found more enriched in the particle-attached communities than their free-living
448	counterparts in the PRE (Zhang et al., 2016). Our field measurements suggested that bacterial respiration
449	within the hypoxic waters was largely contributed by PAB (>0.8 µm) with FLB (0.2-0.8 µm) playing a
450	relatively small role. Therefore, it is crucial to address the linkage between the high-density PAB and the
451	high level of particle-adsorbed PUAs associated with the suspended particles in the low-oxygen waters.
452	Interestingly, our PUA-amended experiments for PAB (>25 μ m) retrieved from the low-oxygen waters
453	revealed distinct responses of PAB to different doses of PUAs treatments with an increase in cell growth in
454	response to low-dose PUAs (1 μ mol L ⁻¹) but an elevated cell-specific metabolic activity including bacterial
455	respiration and production in response to high-dose PUAs (100 μ mol L ⁻¹). An increase in cell density of
456	PAB by low-dose PUAs could likely reflect the stimulating effect of PUAs on PAB growth. This finding
457	was consistent with the previous report of a PUAs level of 0-10 μ mol L ⁻¹ stimulating respiration and cell
458	growth of PAB in sinking particles of the open ocean (Edwards et al., 2015). The negligible effect of

459	low-dose PUAs on bacterial community structure in our experiments was also in good agreement with
460	those found for PAB from sinking particles (Edwards et al., 2015). However, we do not see the inhibitory
461	effect of 100 μ mol L ⁻¹ PUAs on PAB respiration and production previously found in the open ocean
462	(Edward et al., 2015). Instead, the stimulating effect for high-dose PUAs on bacterial respiration and
463	production was even stronger with ~50%-of increments. The bioactivity of PUAs on bacterial strains could
464	likely arise from its specific arrangement of two double bonds and carbonyl chain (Ribalet et al., 2008).
465	Our findings-stronglysupport the important role of PUAs in enhancing bacterial oxygen utilization in the-
466	low-oxygen waters.
467	It should be mentioned that it remains controversial the effect of background nanomolar PUAs on
468	free-living bacteria, which is not our focus in this study. Previous studies suggested that 7.5 nmol L ⁻¹ PUAs
469	would have a different effect on the metabolic activities of distinct bacterial groups in the NW
470	Mediterranean Sea, although bulk bacterial abundance remained unchanged (Balestra et al., 2011). In
471	particular, the metabolic activity of γ -Pro was least affected by nanomolar PUAs, although those of
472	Bacteroidetes and Rhodobacteraceae were markedly depressed (Balestra et al., 2011). Meanwhile, the daily
473	addition of 1 nmol L ⁻¹ PUAs was found to not affect the bacterial abundance and community composition
474	during a mesocosm experiment in the Bothnian Sea (Paul et al., 2012).
475	It is important to verify that the PUAs are not an organic carbon source but a stimulator for PAB
476	growth and metabolism. This was supported by the fact that the inoculated PAB could not grow in the
477	medium with $\frac{200}{\mu}$ mol L ⁻¹ of PUAs although they grew pretty well in the mediums with a similar amount
478	of ALK or PAH. Our results support the previous findings that the density of Alteromonas hispanica was
479	not significantly affected by 100 μ mol L ⁻¹ of PUAs in the minimal medium (without any organic carbons)
480	during laboratory experiments (Figure 9E), where PUAs were considered to act as cofactors for bacterial
481	growth (Ribalet et al., 2008).
482	Improved cell-specific metabolism of PAB in response to high-dose PUAs was accompanied by a

483 significant shift of bacterial community structure. The group of PAB with the greatest positive responses to

exogenous PUAs was genus Alteromonas within the γ -Pro, which is well-known to have a particle-attached 484 lifestyle with rapid growth response to organic matters (Ivars-Martinez et al., 2008). This result is 485 486 contradicted to by the previous finding of a reduced percentage of the γ -Pro class by high-dose PUAs in the PAB of open ocean sinking particles (Edward et al., 2015). Meanwhile, previous studies suggested that 487 different genus groups within the γ -Pro may respond distinctly to PUAs (Ribalet et al., 2008). Our result 488 was well consistent with the previous finding of the significant promotion effect of 13 or 106 µmol L⁻¹ 489 490 PUAs on Alteromonas hispanica from the pure culture experiment (Ribalet et al., 2008). An increase of 491 PUAs could thus confer some of the γ -Pro (mainly special species within the genus Alteromonas, such as A. 492 *hispanica*, Figure S2B) a competitive advantage over other bacteria, leading to their population dominance 493 on particles in the low-oxygen waters. These results provide evidences for a previous hypothesis that PUAs 494 could shape the bacterioplankton community composition by driving the metabolic activity of bacteria with 495 neutral, positive, or negative responses (Balestra et al., 2011).

The taxonomic composition of PAB (>25 μ m) was substantially different from that of the bulk 496 bacteria community in the hypoxic zone (with a large increase of γ -Pro associated with particles, Figure 497 498 S2A). This result supports the previous report of γ -Pro being the most dominant clades attached to sinking particles in the ocean (DeLong et al., 1993). A broad range of species associated with γ -Pro were-was 499 known to be important for quorum sensing processes due to their high population density (Doberva et al., 500 2015) associated with sinking or suspended aggregates (Krupke et al., 2016). In particular, the genus of 501 γ -Pro such as *Alteromonas* and *Pseudomonas*, are well-known guorum-sensing bacteria that can rely on 502 503 diverse signaling molecules to affect particle-associated bacterial communities by coordinating gene expression within the bacterial populations (Long et al., 2003; Fletcher et al., 2007). 504

It has been reported that the growths of some bacterial strains of the γ -Pro such as *Alteromonas* spp. and *Pseudomonas* spp. could be stimulated and regulated by oxylipins like PUAs (Ribalet et al., 2008; Pepi et al., 2017). Oxylipins were found to promote biofilm formation of *Pseudomonas* spp. (Martinez et al., 2016) and could serve as signaling molecules mediating cell-to-cell communication of *Pseudomonas* spp. 509 by an oxylipin-dependent quorum sensing system (Martinez et al., 2019). As PUAs are an important group of chemical cues belonging to oxylipins (Franzè et al., 2018), it is thus reasonable to expect that PUAs may 510 511 also participate as potential signaling molecules for the quorum sensing among a high-density Alteromonas or Pseudomonas. A high level of particle-adsorbed PUAs occurring on organic particles in the low-oxygen 512 water would thus likely allow particle specialists such as *Alteromonas* to regulate bacterial community 513 structure, which could alter species richness and diversity of PAB as well as their metabolic functions such 514 as respiration and production when interacting with particulate organic matter in the hypoxic zone. Various 515 bacterial assemblages may have different rates and efficiencies of particulate organic matter degradation 516 517 (Ebrahimi et al., 2019). Coordination amongst these PAB could be critical in their ability to thrive on the recycling of POC (Krupke et al., 2016) and thus directly likely contribute to the acceleration of oxygen 518 519 utilizations in the hypoxic zone. Nevertheless, the molecular mechanism of the potential PUA-dependent quorum sensing of PAB may be an important topic for future study. 520 Our findings may likely apply to other coastal systems where there are large river inputs, intense 521 522 phytoplankton blooms driven by eutrophication, and strong hypoxia, such as the Chesapeake Bay, the

523 Adriatic Sea, and the Baltic Sea. For example, Chesapeake Bay is largely influenced by river runoff with

524 <u>strong eutrophication-driven hypoxia during the summer as a result of increased water stratification (Fennel</u>

525 and Testa, 2019) and enhanced microbial respiration fueled by organic carbons produced during spring

526 diatom blooms (Harding et al., 2015). Similar to the PRE, there was also a high abundance of γ -Pro in the

527 <u>low-oxygen waters of the Chesapeake Bay associated with the respiration of resuspended organic carbon</u>

528 (Crump et al., 2007). Eutrophication causes intense phytoplankton blooms in the coastal ocean.

529 <u>Sedimentation of the phytoplankton carbons will lead to their accumulation in the surficial sediment</u>

530 (Cloern, 2001), including PUFA compounds derived from the lipid production. Resuspension and oxidation

531 of these PUFA-rich organic particles during summer salt-wedge intrusion might lead to high

532 particle-adsorbed PUAs in the water column. These PUAs could likely shift the particle-attached bacterial

533 <u>community to consume more oxygen when degrading particulate organic matter and thus potentially</u>

534	contribute to the formation of seasonal hypoxia. In this sense, the possible role of PUAs on coastal hypoxia	
535	may be a byproduct of eutrophication driven by anthropogenic nutrient loading. Further studies are required	
536	to quantify the contributions from PUAs-mediated oxygen loss by aerobic respiration to total	
537	deoxygenation in the coastal ocean.	
538		
539	5. Conclusions	
540	In summary, we found elevated concentrations of pPUAs and dPUAs in the hypoxic waters-dominated by-	
541	PABbelow the salt-wedge. We also found high particle-adsorbed PUAs associated with particles of >25	
542	μm in the hypoxic waters based on the large-volume filtration method, which could generate a hotspot	
543	<u>PUAs concentration of >10 μmol L⁻¹ in the water column. In the hypoxic waters, bacterial respiration was</u>	
544	largely controlled by PAB (>0.8 μm) with FLB (0.2-0.8 μm) only accounting for 25-30% of the total	
545	respiration. The increase of PUAs in the bottom waters could be due to enhanced oxidation of resuspended	
546	PUFA by lipoxygenase in response to increased salinity driven by seawater intrusion at the bottom of the	
547	salt-wedge. Field PUA-amended experiments were conducted for PAB associated with particles of >25 µm	
548	<u>retrieved from the low-oxygen waters.</u> We found distinct responses of PAB (>25 μ m) to different doses of	
549	PUAs treatments with an increase of cell growth in response to low-dose PUAs (1 μ mol L ⁻¹) but an	
550	elevated cell-specific metabolic activity including bacterial respiration and production in response to	
551	high-dose PUAs (100 µmol L ⁻¹). Improved cell-specific metabolism of PAB in response to high-dose PUAs	
552	was also accompanied by a significant substantial shift of bacterial community structure with increased	
553	dominance of genus Alteromonas within the γ -Pro.	
554	Based on these observations, we hypothesize that PUAs may potentially act as signaling molecules for	
555	coordination among the high-density PAB below the salt-wedge, which would likely allow bacteria such as	
556	Alteromonas to thrive in degrading particulate organic matters. Very possibly, this process by changing	
557	community compositions and metabolic rates of PAB would lead to an increase of microbial oxygen	
558	utilization that <u>might eventually</u> contribute to the formation of coastal hypoxia. 23	

560	Data availability. Some of the data used in the present study are available in the Supplement. Other data
561	analyzed in this article are tabulated herein. For any additional data please request from the corresponding
562	author.
563	
564	Supplement. The supplement related to this article is available online at: bg-2020-243-supplement.
565	
566	Author Contributions. Q.P.L designed the project. Z.W. performed the experiments. Q.P.L and Z.W. wrote
567	the paper with inputs from all co-authors. All authors have given approval to the final version of the
568	manuscript.
569	
570	Competing interests. The authors declare no competing financial interest.
571	
572	Acknowledgements. We are grateful to the captains and the staff of R/V Haike68 and R/V Tan Kah Kee for
573	help during the cruises. We thank Profs Dongxiao Wang (SCSIO) and Xin Liu (XMU) for organizing the
574	cruises, Mr. Yuchen Zhang (XMU) for field assistances, Profs Changsheng Zhang (SCSIO) and Weimin
575	Zhang (GIM) for analytical assistance, as well as Prof. Dennis Hansell (RSMAS) for critical comments.
576	
577	Financial support. This work was supported by the National Natural Science Foundation of China
578	(41706181, 41676108), the National Key Research and Development Program of China
579	(2016YFA0601203), and the Key Special Project for Introduced Talents Team of Southern Marine Science
580	and Engineering Guangdong Laboratory (Guangzhou) (GML2019ZD0305). ZW also wants to
581	acknowledge a visiting fellowship (MELRS1936) from the State of Key Laboratory of Marine
582	Environmental Science (Xiamen University).

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- Table 1. Summary of treatments in the experiments of exogenous PUAs additions for the low-oxygen
 waters at station Y1 during June 2019. The PUAs solution includes heptadienal (C7_PUA), octadienal
- 740 (C8_PUA), and decadienal (C10_PUA) with the mole ratios of 10:1:10.

		Treatment
1	Control (methanol)	methanol
2	Low-dose PUAs (methanol)	2 mM PUAs in methanol
3	High-dose PUAs (methanol)	200 mM PUAs in methanol

Figures and Legends

Figure 1: Sampling map of the Pearl River Estuary and the adjacent northern South China Sea during (A) June $17^{\text{th}}-28^{\text{th}}$, 2016, (B) June 18^{st} -June 2^{nd} , 2019. Contour shows the bottom oxygen distribution with white lines highlighting the levels of 93.5 µmol kg⁻¹ (oxygen-deficient zone) and 62.5 µmol kg⁻¹ (hypoxic zone); dashed line in panel A is an estuary-to-shelf transect with <u>red-blue</u> dots for three stations with bacterial metabolic rate measurements; diamonds in panel B are two stations with vertical pPUAs and dPUAs measurements with Y1 the station for PUAs-amended experiments.

751

Figure 2: Procedure of large-volume filtration and subsequent experiments. A large volume of the 752 low-oxygen water was filtered through a 25-µm filter to obtain the particles-adsorbed PUAs and the 753 particle-attached bacteria (PAB). The carbon-source test of PUA for the inoculated PAB includes the 754 additions of PUA, alkanes (ALK), and polycyclic aromatic hydrocarbons (PAH). PUAs-amended 755 experiments for PAB include Control (CT), Low-dose (PL), and High-dose PUAs (PH). Samples in the 756 biological oxygen demand (BOD) bottles at the end of the experiment were analyses for bacterial 757 respiration (BR), abundances (BA), production (BP) as well as DNA. Note that pPUAs and dPUAs are 758 particulate and dissolved PUAs in the seawater. 759

760

Figure 3: Vertical distributions of (A) temperature, (B) turbidity, (C) nitrate, (D) salinity, (E) dissolved oxygen, and (F) chlorophyll-*a* from the estuary to the shelf of the NSCS during June 2016. Section locations are shown in Figure 1; the white line in panel D shows the area of oxygen deficiency zone (<93.5 μ mol kg⁻¹).

765

Figure 4: Comparisons of oxygen, bulk bacterial respiration (BR) and production (BP), as well as bulk bacterial abundances (BA) of α -Proteobacteria (α -Pro), γ -Proteobacteria (γ -Pro), Bacteroidetes (Bact), and other bacteria for the bottom waters between stations inside (X1) and outside (X2 and X3) the hypoxic zone during the 2016 cruise. Bulk bacteria community includes FLB and PAB of <20 µm. Locations of stations X1, X2, X3 are showed in Figure 1A. Error bars are the standard deviations.

771

Figure 5: Vertical distributions of (A) temperature, (B) salinity, (C) dissolved oxygen (DO), (D)
chlorophyll-*a* (Chl-*a*), (E) particulate PUAs (pPUAs) and (F) dissolved PUAs (dPUAs) inside (Y1) and
outside (Y2) the hypoxic zone during June 2019. Locations of station Y1 and Y2 are shown in Figure 1.
<u>Error bars are the standard deviations.</u>

Figure 6: Concentrations of particle-adsorbed PUAs (in micromoles per liter particle) in the <u>middle (12 m)</u>
and the bottom (<u>25 m)</u> waters of station Y1 during June 2019. Three different PUA components are also
shown including heptadienal (C7_PUA), octadienal (C8_PUA), and decadienal (C10_PUA). Error bars are
the standard deviations.

781

Figure 7: Responses of particle-attached bacterial parameters including (A) bacterial abundance ($BA_{particle}$), (B) bacterial respiration ($BR_{particle}$), (C) cell-specific bacterial respiration ($sBR_{particle}$), (D) bacterial growth efficiency ($BGE_{particle}$), (E) bacterial production ($BP_{particle}$), and (F) cell-specific bacterial production ($sBP_{particle}$) to different doses of PUAs additions at the end of the experiments for the middle (12 m) and the bottom waters (25 m) at station Y1. Error bars are standard deviations. The star represents a significant difference (p<0.05) with PL and PH the low and high dose PUA treatments and C the control.

788

Figure 8: Variation of particle-attached bacterial community compositions on (A) the phylum level and (B) the genus level in response to different doses of PUAs additions at the end of the experiments for the middle and the bottom waters at station Y1. Labels PL and PH are for the low- and high-dose PUAs with CT the control.

793

794 Figure 9: Carbon-source test of PUAs with cell culture of particle-attached bacteria inoculated from the low-oxygen waters of station Y1 including the initial conditions (Day0) at the beginning of the experiments 795 796 as well as results after 30 days of incubations (Day30) for (A, B) the middle and (C, D) the bottom waters, respectively. Bottles from left to right are the mediums (M) with the additions of polycyclic aromatic 797 hydrocarbons (M+PAH, 200 ppm), alkanes (M+ALK, 0.25 g L⁻¹), and heptadienal (M+C7 PUA, 0.2 mmol 798 L^{-1}); Note that a change of turbidity should indicate bacterial utilization of organic carbons. (E) the optical 799 density of bacterium Alteromonas hispanica MOLA151 growing in the minimal medium as well as in the 800 mediums with the additions of mannitol, pyruvate, and proline (M+MPP, 1% each,), heptadienal 801 (M+C7 PUA, 145µM), octadienal (M+C8 PUA, 130µM,), and decadienal (M+C10 PUA, 106µM). The 802 method for A. hispanica growth and the data in panel E are from Ribalet et al., 2008. 803



Figure 1





Figure 2



Figure 3







Figure 6



Figure 7



Figure 8



Figure 9