

1 **Ideas and perspectives: A strategic assessment of methane and nitrous oxide measurements**
2 **in the marine environment**

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50 **Abstract.** In the current era of rapid climate change, accurate characterization of climate-
51 relevant gas dynamics - namely production, consumption and net emissions - is required for all
52 biomes, especially those ecosystems most susceptible to the impact of change. Marine
53 environments include regions that act as net sources or sinks for numerous climate-active trace
54 gases including methane (CH₄) and nitrous oxide (N₂O). The temporal and spatial distributions
55 of CH₄ and N₂O are controlled by the interaction of complex biogeochemical and physical
56 processes. To evaluate and quantify how these mechanisms affect marine CH₄ and N₂O cycling
57 requires a combination of traditional scientific disciplines including oceanography,
58 microbiology, and numerical modeling. Fundamental to these efforts is ensuring that the
59 datasets produced by independent scientists are comparable and interoperable. Equally critical is
60 transparent communication within the research community about the technical improvements
61 required to increase our collective understanding of marine CH₄ and N₂O. An Ocean Carbon &
62 Biogeochemistry (OCB) sponsored workshop was organized to enhance dialogue and
63 collaborations pertaining to marine CH₄ and N₂O. Here, we summarize the outcomes from the
64 workshop to describe the challenges and opportunities for near-future CH₄ and N₂O research in
65 the marine environment.

66

67

68 **1. Background**

69 The most abundant greenhouse gases in the troposphere, excluding water vapor, are carbon
70 dioxide (CO₂), methane (CH₄), and nitrous oxide (N₂O). Together they account for more than
71 80% of the total radiative forcing (IPCC, 2013) and their current tropospheric mole fractions and
72 rates of increase are unprecedented in recent Earth history (Ciais et al., 2013; Burke et al., 2020;
73 Fig. 1a and 1b). While CO₂ is the most abundant of the three greenhouse gases, CH₄ and N₂O
74 both have a higher warming potential than CO₂ (Montzka et al., 2011). To accurately constrain
75 the contribution of CH₄ and N₂O to Earth's radiation budget and their representation in
76 predictive models requires their sources and sinks to be quantified with high resolution at the
77 global scale.

78 The oceans are a fundamental component of the global climate system and are a net source of
79 tropospheric CH₄ and N₂O at the global scale, although local to regional budgets may include
80 both source and sink components. There are far fewer marine measurements of dissolved CH₄
81 and N₂O than of dissolved CO₂ and while there is substantial international coordination with
82 regard to CO₂ analysis, calibration and data reporting, no such coordination yet exists for CH₄
83 and N₂O (Wilson et al. 2018). Given the increasing prominence of climate change on scientific
84 and societal agendas, greater coordination among the marine CH₄ and N₂O scientific community
85 to provide more targeted measurements and increase the quality and interoperability of CH₄ and
86 N₂O observations is particularly timely.

87 Despite the lack of an international coordinating framework, there have been important
88 advances in our understanding of marine CH₄ and N₂O in numerous research disciplines, ranging
89 from cellular metabolism and model microbial systems to large-scale modeling. For example,
90 recent work identified novel microorganisms and metabolic pathways in the production of N₂O
91 (Trimmer et al., 2016; Caranto and Lancaster, 2017) and CH₄ (Repeta et al. 2016; Bižić et al.,
92 2020). Earth system models now incorporate improved N₂O parameterizations to better resolve
93 the ocean's role in the global N₂O cycle (Battaglia and Joos, 2018). New techniques enable the
94 discrimination of ancient and modern dissolved CH₄ (Sparrow et al., 2018) and the transfer of
95 CH₄-derived carbon to other carbon pools (Pohlman et al., 2011; Garcia-Tigreros and Kessler,
96 2018). Other technological and analytical advances include improved near-continuous

97 spectroscopic analysis that yield greater sampling resolution in surface waters (e.g. Gülzow et
98 al., 2011; Arévalo-Martínez et al., 2013; Erler et al., 2015) and the deployment of analytical
99 devices on robotic vehicles (Nicholson et al., 2018).

100 These scientific advances and an improvement in the quantity and quality of CH₄ and N₂O
101 observations are timely given that large areas of both the open and coastal ocean remain under-
102 sampled (Fig. 1c and 1d). Limited observations contribute to uncertainty in marine CH₄ and
103 N₂O inventories, their rates of production and consumption, and their emissions. The uncertainty
104 associated with CH₄ and N₂O inventories is particularly problematic given that the marine
105 environment is susceptible to an accelerating rate of anthropogenic change that will continue to
106 modify the global cycles of carbon and nitrogen into the future. Environmental impacts on
107 marine CH₄ and N₂O distributions include increasing seawater temperatures, decreasing
108 concentrations of dissolved oxygen (O₂), acidification, retreat of ice and mobilization of carbon
109 substrates from former permafrost, altering coastal run-off, and eutrophication (IPCC, 2019).
110 These impacts will undoubtedly alter future CH₄ and N₂O exchange with the atmosphere, but the
111 directions and magnitudes of these modified fluxes remains insufficiently understood.

112 The need to resolve the marine CH₄ and N₂O inventories prompted an evaluation of the
113 collective ability of the international scientific community to accurately determine the
114 distribution and emissions of CH₄ and N₂O, and the determining physical-biogeochemical
115 factors. This became the focus of a marine CH₄ and N₂O workshop hosted by the Ocean Carbon
116 and Biogeochemistry (OCB) program at Lake Arrowhead, California in October 2018. The
117 workshop considered CH₄ and N₂O equally on the same agenda, even though nearly all field,
118 laboratory, and modeling studies examine these trace gases separately. The rationale for this
119 dual approach is that CH₄ and N₂O share common considerations of the physical, chemical, and
120 microbial processes that dictate their water-column distributions (Bakker et al., 2014; Bodelier
121 and Steenbergh, 2014). In addition, many of the analytical procedures for quantifying CH₄ and
122 N₂O and the subsequent data quality assurances share many common requirements. The
123 opportunity to bring a large research community together to increase dialogue and encourage the
124 cross-fertilization of ideas was thus considered very valuable. This article articulates the
125 workshop outcomes framed in the context of current marine CH₄ and N₂O research and explores
126 future research opportunities and challenges.

127

128 2. Coordination of oceanic CH₄ and N₂O measurements

129 Our understanding of the temporal and spatial distributions of oceanic CH₄ and N₂O derives
130 from over five decades of open ocean and coastal observations, including targeted expeditions,
131 repeat hydrographic surveys, and time-series monitoring, each of which has been crucial to the
132 development of our current knowledge (Fig. 2). Targeted programs have enabled invaluable
133 insights into the role of oxygen deficient zones in N₂O cycling (Babbin et al., 2015; Bourbonnais
134 et al., 2017; Frey et al., 2020) and the exploration of CH₄-rich seeps and vents (Foucher et al.,
135 2009; Suess, 2010; Boetius and Wenzhöfer, 2013). Basin-scale repeat hydrographic surveys
136 (e.g. the international GO-SHIP program) have facilitated extensive water-column mapping to
137 identify relevant water masses and evaluate ventilation rates (Fig. 2d) (de la Paz et al., 2017).
138 Other oceanic surveys have focused exclusively on surface sampling, using continuous
139 equilibrator systems connected to various gas analyzers to yield high-resolution surface
140 concentration fields of CH₄ and N₂O (Gülzow et al., 2013; Erler et al., 2015; Kodovska et al.,
141 2016; Thornton et al., 2016a; Pohlman et al., 2017). In contrast, sustained long-term time-series
142 measurements of CH₄ and N₂O at fixed monitoring stations are relatively few, but they span a
143 range of latitudes and biogeochemical provinces (Fig. 2a and 2b). The time-series observations
144 provide the contextual background for seasonal and interannual variation that allow long-term
145 temporal trends and episodic events to be identified and evaluated (Farías et al., 2015; Wilson et
146 al., 2017; Ma et al., 2019). Overall, the majority of measurements enable the variability in
147 marine CH₄ and N₂O to be quantified at the mesoscale or greater (i.e. from hundreds of
148 kilometers to ocean basins), with monthly to annual resolution, but there are substantially fewer
149 datasets at the sub-mesoscale level (i.e. <10 km and hours to days) (Fig. 3). A major reason for
150 the limited sampling at the sub-mesoscale level is that it necessitates high-resolution
151 measurements to resolve the heterogeneous variability that exists at these time-space scales.
152 Such analyses have only recently become technically feasible (discussed in more detail in
153 Section 6).

154 Until recently there has been no formal coordination of observations across the CH₄ and N₂O
155 scientific community. In response to this, a Scientific Committee on Oceanic Research (SCOR)
156 Working Group was initiated in 2014 entitled: '*Dissolved N₂O and CH₄: Working towards a
157 global network of ocean time series measurements*'. A major goal of the SCOR Working Group
158 was to unite the international community in joint activities conceived to improve and inform

159 seagoing CH₄ and N₂O analyses. An important activity was the preparation and distribution of
160 common, combined gaseous CH₄ and N₂O standards to twelve international laboratories, with
161 the aim of improving and standardizing calibration (Bullister et al., 2017). A subsequent inter-
162 comparison of discrete seawater samples included the use of these standards and revealed the
163 variability between laboratories. While there were some encouraging results from the
164 intercomparison, such as the agreement between individual laboratories using contrasting
165 techniques, overall a large range was observed in CH₄ and N₂O concentration data generated by
166 the participating laboratories (Wilson et al., 2018). Such analytical discrepancies weaken our
167 collective ability as a community to evaluate temporal-spatial variability in marine CH₄ and N₂O.
168 The discrepancies also highlighted the need for Standard Operating Protocols (SOPs) for CH₄
169 and N₂O analyses to facilitate standardization of sampling, measurement, and calibration, as well
170 as the reporting of data and accompanying metadata in common repositories. The SOPs are
171 currently in preparation with intended publication on the Ocean Best Practices network.

172 A data repository for oceanic CH₄ and N₂O data known as the MarinE Methane and Nitrous
173 Oxide database (MEMENTO) was established in 2009 (Bange et al., 2009; Kock and Bange,
174 2015). MEMENTO is now sufficiently mature to support descriptions of the broad-scale surface
175 distributions of CH₄ and N₂O (e.g. Suntharalingam et al., 2012; Zamora and Oschlies, 2014;
176 Buitenhuis et al., 2018; Battaglia and Joos, 2018). Machine-learning mapping also recently
177 identified the various contributions of physical and biogeochemical predictor variables for CH₄
178 (e.g. depth, primary production; Weber et al., 2019; Fig 4b) and N₂O distributions (e.g.
179 chlorophyll, sea surface temperature, apparent oxygen utilization, and mixed-layer depth; Yang
180 et al., 2020; Fig. 4a). The application of gas transfer algorithms to the extrapolated oceanic CH₄
181 and N₂O distributions helped decrease the uncertainty in estimates of global air-sea exchange
182 fluxes (Fig. 4c), thereby fulfilling one of the key goals of MEMENTO (Bange et al., 2009). Net
183 global open ocean emissions of N₂O are now similarly estimated at 3–5 Tg N yr⁻¹ by both Yang
184 et al. (2020) and the Global Nitrous Oxide Project (Tian et al., 2020). In comparison, net global
185 ocean CH₄ emissions from machine-learning mapping were estimated at 6–12 Tg CH₄ yr⁻¹
186 (Weber et al., 2019), compared to 9–22 Tg CH₄ yr⁻¹ in the most up-to-date CH₄ synthesis
187 (Saunio et al., 2020). However, the narrower range for machine-learning derived CH₄
188 emissions retains high uncertainty in regions such as the Arctic, where emissions are highly
189 heterogeneous and compounded by seasonal ice cover. Identifying the causes for uncertainty in

190 high emission regions will greatly aid future sampling campaigns, as is discussed in the
191 following sections.

192

193 **3. Methane in marine environments**

194 In the surface waters of tropical and temperate oceans, a number of factors contribute to the
195 low supersaturation of CH₄ including direct aerobic production arising from the degradation of
196 methylated sulfur compounds by phytoplankton (Klitzsch et al., 2019) and methyl phosphonate
197 in phosphorus-depleted waters (Karl et al. 2008, Sosa et al., 2020), indirect production via
198 grazing (Schmale et al., 2018) and abiotic photoproduction (Li et al., 2020). A recent study
199 demonstrated that CH₄ production by cyanobacteria is linked to general cell metabolism and
200 does not rely on the presence of methylated precursor compounds (Bižić et al., 2020). Deep
201 within the ocean's pelagic interior, CH₄ is weakly undersaturated reflecting depletion via
202 microbial oxidation (Reeburgh 2007; Weber et al., 2019). Towards the coastline, CH₄
203 supersaturation increases by orders of magnitude (Figure 5b), reflecting terrestrial inputs (e.g.
204 river and groundwater), increased organic matter loading (Borges et al., 2018), and CH₄
205 diffusion and ebullition from shallow anoxic methane rich sediments (Zhang et al., 2008; Borges
206 et al., 2016; Upstill-Goddard and Barnes, 2016). Supersaturation of CH₄ occurs frequently in the
207 Arctic Ocean and its relatively shallow marginal seas with the most extreme values observed in
208 the Eurasian Arctic (e.g. Shakhova et al., 2010; Damm et al., 2015; Kosmach et al., 2015;
209 Thornton et al, 2016a; Lorenzen et al., 2016; Fenwick et al., 2017; Lapham et al., 2017).
210 Terrestrial and subsea permafrost are potential CH₄ sources to shelf waters in addition to CH₄
211 hydrates that are found in marginal shelves globally (Ruppel and Kessler, 2017). Large point
212 source CH₄ emissions, such as seafloor gas seeps can be large sources to the atmosphere in small
213 localized areas (e.g. Thornton et al., 2020), but these sites remain particularly difficult to
214 parameterize in models. This reflects limited observations and a poor understanding of their
215 spatial distributions, the driving mechanisms, and the wider context within the carbon cycle. For
216 example, the upwelling of cold, nutrient-rich water that accompanies CH₄ ascending the water
217 column stimulates CO₂ consumption by photosynthesizing phytoplankton, rendering such CH₄
218 seeps an overall net sink for climate-forcing gases (Pohlman et al., 2017). Recent work using
219 thermal infrared satellite retrievals indicates increased high-latitude oceanic CH₄ release in late
220 autumn, coincident with pycnocline breakdown and a deepening of the ocean mixed layer depth

221 thereby bringing deep CH₄ to the surface (Yurganov et al., 2019). This is especially notable in
222 the Kara and Barents Seas, but the remote observations have not yet been confirmed by surface
223 ocean measurements which are difficult and therefore rare, except during the Arctic summer.

224 Seabed CH₄ emissions are hypothesized to increase in a warming ocean through the
225 decomposition of gas hydrates, the degradation of subsea permafrost under some high-latitude
226 seas, and the increased biodegradation of sediment carbon (Romanovskii et al., 2005; Biastoch et
227 al., 2011; Lapham et al., 2013; Ruppel and Kessler, 2017; Borges et al., 2019). Effort is thus
228 focused on quantifying the fraction of CH₄ generated in or released from marine sediments that
229 ultimately enters the atmosphere, particularly on shallow continental shelves and in coastal
230 ecosystems. Natural stable isotopes have been used to inform spatial and temporal changes in
231 dissolved CH₄ concentrations (e.g. Pack et al., 2011; Mau et al., 2012; Weinstein et al., 2016;
232 Leonte et al., 2017; Chan et al., 2019) and incubation experiments with added stable isotopes and
233 radiotracers have helped elucidate how oxidation (anaerobically in sediments and aerobically in
234 the water column), ebullition (where CH₄ pore water partial pressure exceeds sediment
235 hydrostatic pressure), and subsequent bubble dissolution in the water column interact to mitigate
236 CH₄ emissions to air (Steinle et al., 2015; Jordan et al., 2020). The information deriving from
237 these various approaches is inherently different but complementary. Isotope tracer incubations
238 provide snapshots of rates specific to the methanotrophic community and CH₄ concentration at
239 the time of sampling, whereas concentrations and isotopic gradients are used to infer *in situ* rates
240 integrated over space and time. A recent study deployed a remotely operated vehicle to examine
241 the isotopic fractionation of CH₄ during bubble ascent and used this to constrain the extent of
242 bubble dissolution (Leonte et al., 2018). This work demonstrated an experimental approach
243 established for broadly constraining water column CH₄ cycling directly from a surface research
244 vessel.

245 Despite the range of analytical and experimental approaches available, determining whether
246 the origin of the emitted CH₄ is seafloor release or aerobic production in the upper water column
247 remains problematic. To date there is no straightforward way to routinely distinguish between
248 seafloor derived and water column generated CH₄ for all locations. Even so, stable carbon and
249 hydrogen isotope measurements (i.e. $\delta^{13}\text{C-CH}_4$ and $\delta^2\text{H-CH}_4$) combined with ancillary data may
250 provide valuable source information. For example, combining these measurements with the ratio
251 of CH₄ to higher order hydrocarbons (e.g. ethene (C₂H₄) and ethane (C₂H₆)) can be used to infer

252 for example, whether the origin of the CH₄ is thermogenic, sub-seafloor, or biogenic within the
253 water column (Whiticar, 1999; Pohlman et al., 2009; Lan et al., 2019). Continuous shipboard
254 measurement of CH₄ isotopes in surface water (e.g. Pohlman et al., 2017) and in the atmospheric
255 boundary layer (Pankratova et al., 2019; Berchet et al., 2020) are now possible and they have
256 been used in combination with atmospheric inversion models to characterize and discriminate
257 marine-emitted CH₄ from other sources (Berchet et al., 2020). Application of this method to
258 land-based monitoring stations appears promising for apportioning CH₄ emissions from various
259 marine regions and sources (Thonat et al., 2019). Additionally, in regions where aerobic CH₄
260 oxidation is substantial, the resulting isotopic fractionation generates measurable vertical and/or
261 horizontal seawater gradients that can also be used to identify contrasting biogenic CH₄ sources
262 (Leonte et al., 2020). However, the general overlap in isotope compositions of sediment CH₄
263 (e.g. Thornton et al., 2016b; Sapart et al., 2017) can complicate purely isotope-based
264 determinations of sources.

265 Measurements of the natural radiocarbon content of dissolved oceanic CH₄, while being
266 highly specialized and requiring substantial amounts of ship time and processing (Kessler and
267 Reeburgh, 2005; Sparrow and Kessler, 2017), provide valuable source information because the
268 ¹⁴C-CH₄ measurements are normalized to the same δ¹³C value and are unaffected by the extent
269 of oxidation. The bubbles sampled from hydrate and active seafloor seeps are largely devoid of
270 radiocarbon (Pohlman et al., 2009; Kessler et al., 2008; Douglas et al., 2016). However, CH₄ in
271 sediments can also be derived from more modern or recently deposited organic material and an
272 exact determination of individual contributions is hard to achieve (Kessler et al., 2008; Sparrow
273 et al., 2018). The powerful insights made by radiocarbon-CH₄ investigations would be further
274 strengthened by concurrent sampling of other analytes that offer CH₄ source information, such as
275 clumped isotopes. Isotope clumping, the co-occurrence of two or more of the less abundant
276 isotopes in a molecule (e.g. ¹³C and ²H or ¹H and ²H), provides unique information on marine
277 CH₄ sources (Stolper et al., 2014; Wang et al., 2015; Douglas et al., 2017; Young et al., 2017;
278 Labidi et al., 2020). In this approach, the isotopic deviations in samples from their random
279 probability distributions can give insight into formation temperature and the extent of
280 biochemical disequilibrium. However, the sample size required for a clumped isotope analysis
281 in the oceanic environment away from areas of seafloor emission is large and exceeds the
282 already demanding volume requirements for ¹⁴C analyses by 1–2 orders of magnitude (Douglas

283 et al., 2017). While the requirement of large sample size and lengthy measurement time
284 currently preclude their more widespread application, clumped isotope measurements offer
285 future promise in refining our understanding of the processes of marine CH₄ production and
286 consumption.

287

288 **4. Nitrous oxide in marine environments**

289 The large-scale spatial distribution of N₂O in the global ocean is reasonably well-established.
290 The highest open ocean N₂O values are in upwelling environments, where concentrations extend
291 up to micromolar levels (Arévalo-Martínez et al., 2015) and production rates can be as high as
292 120 nM d⁻¹ (Frey et al., 2020). The highly elevated N₂O concentrations can be proximal to
293 regions with some of the lowest recorded N₂O concentrations, in the cores of O₂ deficient zones.
294 This coexistence of the highest and lowest observed N₂O concentrations over vertical distances
295 of tens of meters make upwelling regions a focal point for N₂O research, particularly since O₂
296 deficient ocean zones are increasing in size (Stramma et al., 2011). In contrast, in the surface
297 waters of the expansive oligotrophic ocean gyres, N₂O is weakly supersaturated (103-105%)
298 with respect to atmospheric equilibrium (Weiss et al., 1992; Wilson et al., 2017, Charpentier et
299 al., 2010). Nitrous oxide becomes more highly saturated in the surface waters of equatorial
300 upwelling regions due to the upward advection of N₂O-rich waters (Arévalo-Martínez et al.,
301 2017). For the Arctic Ocean, the data indicate low net N₂O emissions, with some areas acting as
302 net N₂O sources and others as N₂O sinks (Fenwick et al., 2017, Zhang et al., 2015).

303 Several parameters control net N₂O emissions from the ocean, including temperature,
304 salinity, dissolved O₂, apparent oxygen utilization (AOU), nutrients, and microbial community
305 abundance and composition. A recent modeling study trained with just three of these variables
306 (chlorophyll, O₂, and AOU) accounted for 60% of the observed variability in oceanic N₂O
307 concentrations (Yang et al., 2020; Fig. 5a), highlighting the importance of N₂O in productive
308 upwelling systems. Correlations between N₂O and environmental variables provide some insight
309 into the factors controlling its distribution, but they provide no information about the
310 microorganisms or metabolic pathways involved. Microbial production of N₂O occurs during
311 the metabolic processes of nitrification and denitrification (Stein and Yung, 2003). To determine
312 which process dominates N₂O production at any given location requires the application of
313 multiple methodological approaches, ideally in parallel.

314 One of the most commonly used approaches is the incubation of discrete water samples
315 under *in situ* conditions with stable isotope (^{15}N) addition such as ^{15}N enriched NH_4^+ , NO_2^- or
316 NO_3^- to measure N_2O production rates from nitrification and denitrification (e.g. Ji et al., 2017).
317 These approaches also provide insight into the microorganisms involved. For example, N_2O
318 resulting from archaeal NH_4^+ oxidation is mostly formed from a combination of NH_4^+ and
319 another N compound (e.g. NO_2^-) whereas bacteria produce N_2O from NH_4^+ alone (Santoro et al.,
320 2011, Stieglmeier et al., 2014; Carini et al. 2018; Lancaster et al., 2018; Frey et al. 2020).
321 Unfortunately, as with all incubation-based approaches ^{15}N techniques are subject to bottle
322 artifacts, and the strong dependence of N_2O production and consumption on ambient O_2
323 increases the potential for contamination during the collection and manipulation of anoxic deep
324 seawaters. Incubation based rate measurements are also compromised by abiotic N_2O
325 production via chemodenitrification, specifically the reduction of NO_2^- coupled to Fe^{2+}
326 oxidation, as observed in high Fe environments (Ostrom et al., 2016; Buchwald et al., 2016;
327 Wankel et al., 2017). These issues highlight the need for incubation techniques that mitigate the
328 effect of experimental artifacts (Stewart et al., 2012).

329 In addition to isotope addition and incubation, natural abundance water-column
330 measurements of N_2O concentrations, isotopes, and isotopomers yield valuable rate and process
331 information. These measurements are free from experimental artifacts and can be used to
332 integrate over appropriate temporal and spatial scales. For example, nitrification in sunlit waters
333 has been inferred from N_2O distributions (Dore and Karl, 1996), and N_2O production close to the
334 ocean surface is a large contributor to the uncertainty in oceanic N_2O emissions (Ward et al.,
335 1982; Zamora and Oeschlies, 2014). Isotopomers are isomers having the same number of each
336 isotope of each element but differing in their structural positions. Nitrous oxide isotopomers are
337 increasingly used, sometimes in combination with box models, to estimate the rates of different
338 N_2O production pathways, in the upwelling systems off southern Africa (Frame et al., 2014) and
339 Peru (Bourbonnais et al., 2017). There is however some disagreement about whether isotopomer
340 signatures are robust indicators of the formation pathway (Yoshida and Toyoda, 2000; Sutka et
341 al., 2006) or whether there is fractionation during production (Schmidt et al., 2004; Casciotti et
342 al., 2018). Greater clarity is therefore required in the use of N_2O isotopes and isotopomers to
343 infer metabolic pathways of N_2O formation. Notwithstanding this issue, field measurements of
344 N_2O isotopes and/or isotopomers have the potential to greatly increase current experimental

345 capabilities and robustness (Yu et al., 2020). However, the development of spectroscopic gas
346 analysis systems that have been so advantageous to CH₄ research has been slower for N₂O. This
347 is due to the higher costs and the increased complexity of the laser systems, although progress is
348 being made to improve instrumental precision, and to decrease matrix effects and spectral
349 interferences (e.g. Harris et al., 2019).

350 A better understanding of the microorganisms responsible for N₂O production and
351 consumption is fundamental to deriving more accurate estimates of process rates. For example,
352 the metabolic activity of ammonia oxidizing archaea can exceed that of ammonia oxidizing
353 bacteria in the ocean (Santoro et al., 2010; Löscher et al., 2012; Fuchsman et al., 2017). The
354 differing sensitivities of these archaea and bacteria to dissolved O₂ (Stahl and de la Torre, 2012;
355 Hink et al., 2017) are a critical factor in evaluating the microbial response to changing
356 environmental conditions, as shown for the terrestrial environment (Prosser et al., 2020).
357 Therefore, to understand the impact of deoxygenation on oceanic N₂O emission requires a better
358 understanding of both archaeal and bacterial metabolisms and their environmental niches. Field-
359 based sequencing not only characterizes the community but can highlight potential metabolic
360 pathways when they might not otherwise be inferred. For example, transcripts encoding for N₂O
361 consumption (*nosZ*) have repeatedly been identified in the oxic water column, despite
362 denitrification being an anaerobic metabolic process (Wyman et al., 2013; Sun et al., 2017). The
363 transcription of *nosZ* has been also located in highly dynamic O₂ permeable coastal sediments
364 (Marchant et al., 2017). Denitrification under aerobic conditions is attributed to fluctuations in
365 O₂, NO₃⁻, organic matter and other parameters that affect the availability of electron donors and
366 acceptors which ultimately influences whether a coastal environment is a net source or sink of
367 N₂O, as discussed in the next section.

368

369 **5. CH₄ and N₂O in shallow marine environments**

370 Coastal and other shallow (<50 m) marine systems are globally relevant CH₄ and N₂O source
371 regions. However, their emission rates to the atmosphere are weakly constrained in comparison
372 with the open ocean. Several factors contribute to the uncertainty, including the high diversity of
373 coastal and shallow marine ecosystems and lack of consistency in adequately defining them,
374 locally heterogeneous conditions causing strong spatial and temporal concentration gradients,
375 highly uncertain spatial distribution of CH₄ seeps, a bias towards studies in the northern

376 hemisphere, and incomplete or sometimes inappropriate sampling strategies (Al-Haj and
377 Fulweiler, 2020). Until these issues are resolved it will remain difficult to adequately define the
378 contribution from shallow marine systems to global CH₄ and N₂O budgets. An important
379 illustration of this is reflected in the prevailing view that large geological sources (e.g. seeps,
380 mud volcanoes, and hydrates) are the main contributors to marine CH₄ emissions (Ciais et al.,
381 2013). The most recent modeled estimate of global marine CH₄ emissions (6–12 Tg CH₄ yr⁻¹)
382 reported that near-shore environments (depths of 0–50 m) contribute a large and highly uncertain
383 diffusive flux (Weber et al., 2019). A study of coastal ecosystems, in this case defined as shelf,
384 estuarine, and tidally influenced rivers, estimated them to contribute 7 Tg CH₄ yr⁻¹ (Anderson et
385 al., 2010) while another estimated 1–7 Tg CH₄ yr⁻¹ for estuaries alone (Borges and Abril, 2011).
386 Similar uncertainties exist for N₂O. Estimates of coastal N₂O emissions (which include coastal,
387 estuarine, and riverine sources) range from 0.1–2.9 Tg N yr⁻¹ (Ciais et al., 2013), although a
388 recent review of N₂O production across a range of estuarine habitats placed N₂O fluxes at the
389 lower end of these estimates (0.17–0.95 Tg N yr⁻¹) (Murray et al., 2015). Based on these data,
390 coastal systems account for around one third of total marine N₂O emissions (Yang et al., 2020).

391 The direct quantification of CH₄ and N₂O emissions from shallow coastal ecosystems has
392 historically involved using gas concentrations measured in discrete water and air samples
393 combined with a gas transfer velocity (k_w). For the coastal and open ocean, the dominant driver
394 of gas exchange is wind speed (e.g. Nightingale et al., 2000; Wanninkhof, 2014) whereas in
395 nearshore, shallow water environments the interaction of water, depth, and tidal current speeds
396 may be a major contributor to near surface turbulence. Several k_w parameterizations are now in
397 use for coastal waters (e.g. Raymond and Cole 2001; Kremer et al., 2003; Zappa et al., 2003;
398 Borges and Abril, 2011; Ho et al. 2011; Rosentreter et al., 2017; Jeffrey et al., 2018) which
399 increases the uncertainties associated with CH₄ and N₂O emissions. For example, a fivefold
400 variation in CH₄ emissions from a single system occurred when applying different
401 parameterizations to the measured gradients in CH₄ (Ferrón et al., 2007).

402 To constrain emissions over small areas, continuous air-sea fluxes can be measured using
403 free-floating chambers (e.g. Bahlmann et al., 2015; Rosentreter et al., 2018; Yang et al., 2018;
404 Murray et al., 2020), but issues related to turbulence modification may still generate flux artifacts
405 (Upstill-Goddard, 2006). To overcome these problems in the future, a greater reliance on direct
406 and robust continuous techniques for air-sea flux measurement, such as eddy covariance (e.g.

407 Podgrajsek et al., 2016), that avoid any need for k_w , will be necessary. Eddy-covariance
408 measurements also capture both diffusive and ebullitive flux components (Thornton et al., 2020).
409 Combining this approach with new analytical techniques such as cavity enhanced absorption
410 spectroscopy (CEAS) and non-dispersive infrared (NDIR) should continue to improve the
411 quality of CH₄ and N₂O flux estimates (McDermitt et al., 2011; Nemitz et al., 2018; Maher et al.,
412 2019). Indeed, eddy flux towers aboard ships (Thornton et al., 2020) and in coastal locations
413 (Yang et al., 2016; Gutiérrez-Loza et al., 2019) are now being equipped with CH₄
414 instrumentation that enables the integration of CH₄ fluxes over large areas. There are fewer N₂O
415 flux estimates made with CEAS and NDIR and the implementation of N₂O sensors on eddy flux
416 towers remains limited. Recently, N₂O emissions from Eastern Boundary Upwelling Systems
417 were quantified using inversion modeling based on atmospheric measurements from coastal
418 monitoring stations highlighting the potential of this approach to constrain N₂O emissions from
419 remote oceanographic regions that have significant spatial and temporal heterogeneity (Ganesan
420 et al., 2020; Babbin et al., 2020). Inverse modeling of atmospheric measurements was also
421 recently used to constrain CH₄ emissions from the East Siberian Arctic Shelf (Tohjima et al.,
422 2020)

423 Coastal measurements of CH₄ and N₂O also require the collection of ancillary data such as
424 water-column depth, tidal motions (Rosentreter et al., 2018; Huang et al., 2019; Pfeiffer-Hebert
425 et al., 2019), and other information relating to diel processes (Maher et al., 2016). Such data are
426 important because for example, the magnitude of CH₄ and N₂O fluxes vary over a diel period
427 depending on the redox environment as a result of tidal effects and changes in inorganic N and
428 O₂ availability (Seitzinger and Kroeze, 1998; Call et al., 2015; Vieillard and Fulweiler, 2014;
429 Maher et al., 2015; Murray et al., 2015; Foster and Fulweiler, 2019). The magnitude of CH₄ and
430 N₂O fluxes also varies over longer temporal scales (seasonally to yearly) due to additional
431 factors such as groundwater inputs, adjacent land-use, dissolved O₂, organic matter content and
432 quality, and macrofaunal distributions (Barnes and Upstill-Goddard, 2011; Upstill-Goddard and
433 Barnes, 2016; Gelesh et al., 2016; Bonaglia et al., 2017; Borges et al., 2018; Wells et al., 2018;
434 Ray et al., 2019; Al-Haj and Fulweiler, 2020; Reading et al., 2020). To determine the
435 contributing factors and resolve the spatial distributions, mobile sampling platforms such as
436 small vessels (Müller et al., 2016; Brase et al., 2017; Tait et al., 2017), and autonomous vehicles
437 (Manning et al., 2019) are essential. Recent improvements in gas sensors and in technology such

438 as sonar and ebullition sensors will further increase our ability to measure dynamic fluxes
439 (Maher et al., 2019; Lohrberg et al., 2020). Improvements to the quality and quantity of CH₄ and
440 N₂O measurements in coastal systems will enable the development of iterative forecast models,
441 further improving estimates of global coastal CH₄ and N₂O fluxes.

442

443 **6. Leveraging culture studies to further our ecosystem understanding**

444 A more complete understanding of marine CH₄ and N₂O necessitates closer integration between
445 biogeochemistry, model requirements, and targeted microbiological studies involving both single
446 microorganism isolates and enrichment cultures. Marine CH₄ and N₂O budgets deriving from
447 both ‘bottom-up’ (e.g. emissions inventories, ocean and terrestrial process models) and ‘top-
448 down’ (e.g. inverse analyses of atmospheric trace-gas measurements) approaches would greatly
449 benefit from more highly constrained metabolic processes. Specifically, this includes rates of
450 CH₄ or N₂O production and consumption for key model microorganisms, and the kinetic
451 parameters associated with these metabolic rates. Reliable inventories of key microbially
452 mediated process rates will improve the robustness of Earth System models used for predicting
453 climate-mediated changes to marine CH₄ and N₂O emissions.

454 For N₂O, laboratory studies quantifying microbial process rates, such as for nitrification and
455 denitrification, are relatively few (e.g. Frame and Casciotti 2010; Santoro et al. 2011; Löscher et
456 al. 2012; Ji et al. 2015; Qin et al., 2017). Consequently, models largely continue to use process
457 rates optimized using water column concentrations of N₂O, O₂, and related nitrogen cycle
458 quantities (e.g. Battaglia and Joos, 2018; Buitenhuis et al., 2018; Landolfi et al., 2017). Future
459 model parameterizations for N₂O will require information on the variability of microbial process
460 yields derived from culture studies with controlled varying conditions of O₂ (Goreau et al. 1980,
461 Frame and Casciotti 2010, Löscher et al. 2012; Ji et al., 2018), pH (Breider et al., 2019; Hopkins
462 et al. 2020), temperature, and nutrients. Automated incubation systems have measured N₂O
463 production kinetics and yield as functions of the concentrations of O₂ and total ammonia nitrogen
464 (Molstad et al., 2007; Hink et al., 2017). Quantifying the physiology of relevant microorganisms
465 and connecting them to environmental characteristics will provide insights into why, for
466 example, some shallow marine habitats act as N₂O sinks while others are N₂O sources, or how
467 N₂O is produced in well oxygenated open-ocean waters, as compared to oxygen deficient zones.

468 For CH₄, a key requirement to relate *in situ* CH₄ production with transport to atmospheric
469 emissions is our ability to accurately determine rates of CH₄ oxidation. Fundamental issues
470 include the challenges of cultivating methanotrophs and of replicating environmental conditions
471 such as pressure and the chemistry of CH₄ gas bubbles. The increased emphasis on CH₄
472 dynamics in shallow water environments highlighted in Section 5, must be supported by culture-
473 based measurements of CH₄ oxidation that control for temperature, O₂ and other important
474 variables. In comparison to CH₄ oxidation, culture-based studies are used increasingly to
475 identify organisms capable of aerobic CH₄ production and their underlying metabolic pathways
476 (Carini et al., 2014; Klintzsch et al; 2019; Bižić et al., 2020).

477 Specific cellular yields and consumption rates of CH₄ and N₂O are not the sole objective of
478 culturing experiments. Cultivation of microorganisms involved in CH₄ and N₂O production and
479 consumption provides vital information into the physiology, metabolism, and interactions of
480 environmentally relevant clades. When combined with genomic approaches, insights can
481 therefore be gained into the diversity and global distribution of organisms involved in CH₄ and
482 N₂O cycling. For CH₄ some unexpected physiologies have been revealed (Ettwig et al., 2010;
483 Haroon et al., 2013; Ettwig et al., 2016), which has directed research into sources and sinks of
484 CH₄ in the natural environment. Similarly, our understanding of how and when ammonia
485 oxidizers produce N₂O has been facilitated by studies of cultured nitrifiers and detailed analysis
486 of their biochemistry (Stahl and de la Torre, 2012; Caranto and Lancaster, 2017). Recent
487 combinations of cultivation studies with environmental genomics, albeit largely for terrestrial
488 systems, have revealed a variety of denitrifiers, many of which are only involved in specific
489 denitrification steps (Ganesh et al., 2014; Lycus et al, 2017; Hallin et al, 2018; Marchant et al.,
490 2018; Conthe et al, 2019).

491

492 **7. Outlook and priorities for marine CH₄ and N₂O measurements**

493 This perspectives article has assessed the collective ability of the scientific community to
494 determine the spatial variability of marine CH₄ and N₂O distributions, the underlying
495 mechanisms that determine this variability, and the resulting sea-to-air emissions. Shallow
496 marine environments and oxygen deficient zones are widely recognized as deserving of greater
497 attention because they have high CH₄ and N₂O concentrations with inherently high uncertainties
498 that complicate any assessment of their emissions to air (Bange et al., 1994; Bange et al., 1996;

499 Bakker et al., 2014; James et al., 2016; Borges et al., 2016; Tian et al., 2020). Fortunately, recent
500 technological advances that have increased our ability to conduct high-resolution measurements
501 allow an optimistic outlook for making substantial progress in quantifying the CH₄ and N₂O
502 budgets of these ecosystems. Even so, the inherent complexity of shallow marine environments
503 clearly warrants a strategically coordinated approach to optimize the value of future studies.
504 Issues to consider include identifying the locations of complementary sampling sites,
505 standardizing sampling strategies and techniques, and agreeing the use of common ancillary
506 measurements that set the broad biogeochemical context (Bange et al., 2019). In contrast to the
507 open ocean, measurement campaigns in shallow water environments are amenable to the use of
508 eddy covariance flux towers, and they have the potential to lever resources from existing
509 observation networks, which in North America include the Long-Term Ecological Research
510 network (LTER) and the National Estuarine Research Reserve (NERR) System (Novick et al.,
511 2018). Indeed, such activities are already underway; an increasing number of flux towers are
512 being equipped for CH₄ measurements (Torn et al., 2019) and future efforts should focus on the
513 inclusion of N₂O (see Section 5).

514 We are encouraged that the Global Carbon Project with its objective of developing a
515 complete picture of the global carbon cycle including interactions and feedbacks has expanded to
516 include CH₄ (Saunois et al., 2020) and is now incorporating N₂O (Tian et al., 2020). These
517 Projects compile the most recent data from peer-reviewed analyses of the sources and sinks of
518 atmospheric CH₄ and N₂O from both natural and human activities. For example, the aquatic
519 components of the recent Global Carbon Project N₂O budget reported emissions from the open
520 ocean, inland waters, estuaries and coastal zones. Low-oxygen oceanic regions associated with
521 eastern-boundary upwelling zones, and the coastal ocean were identified as key regions with
522 significant N₂O variability requiring more detailed assessment via measurement campaigns and
523 model analyses (Tian et al., 2020). Contribution to the Global Carbon Project and similar
524 initiatives will identify areas of synergistic CH₄ and N₂O research for oceanographers and other
525 Earth observation scientists (Ganesan et al., 2019). Furthermore, as highlighted in Section 6,
526 field observations alone are insufficient to improve the robustness of Earth System models and
527 leveraging laboratory-based microbial process studies is highly recommended.

528 The success of any coordinated CH₄ and N₂O research program relies heavily on having
529 uniformly high confidence in the various resulting datasets and their interoperability, and we
530 identify three key initiatives that are paramount to ensuring this:

531 (i) The first is to develop and adopt Standard Operating Protocols (SOPs) to help obtain
532 intercomparable CH₄ and N₂O datasets of the highest possible accuracy and precision.
533 Currently, there is no community-defined level of analytical uncertainty to characterize high
534 quality CH₄ and N₂O measurements. However, attaining an analytical agreement between
535 multiple laboratories of $\leq 1\%$ is considered achievable for the repeat oceanographic surveys and
536 time-series observations (Fig. 3). For context, an analytical agreement of $\leq 1\%$ would permit the
537 ocean's response to the increasing tropospheric CH₄ and N₂O mole fractions to be resolved on
538 timescales of 10 and 5 years, respectively. These values are based on the changes in surface
539 ocean CH₄ and N₂O concentrations that are predicted to occur due to the ongoing increase in
540 tropospheric CH₄ and N₂O mole fractions at a seawater temperature of 20°C and a salinity of 35
541 g kg⁻¹, and assuming all sources and sinks remaining constant. In our recent marine CH₄ and
542 N₂O inter comparison exercise we concluded that the diversity of analytical procedures
543 employed by the participants was a major cause of high variability between the reported
544 concentrations, highlighting an urgent requirement for CH₄ and N₂O SOPs (Wilson et al., 2018).
545 Consequently, these SOPs are now being compiled, and they will be freely available via the
546 Ocean Best Practices System.

547 (ii) The second is increased regularity of intercomparison exercises through the periodic
548 distribution of consensus material, i.e. water samples in which CH₄ and N₂O concentrations are
549 known with high confidence, obtained by pooling analyses from several laboratories with
550 demonstrated analytical capability. These will help the scientific community to monitor data
551 comparability and accuracy, particularly in the case of highly elevated concentrations of CH₄ and
552 N₂O, i.e. those exceeding atmospheric equilibrium concentrations by at least an order of
553 magnitude.

554 (iii) The third activity builds on the previous initiative and calls for the production of Global
555 Data Products for dissolved CH₄ and N₂O measurements. To date, individual CH₄ and N₂O
556 measurements are represented at the global scale by the MEMENTO database which has been
557 very successful at compiling CH₄ and N₂O datasets and making them readily accessible to the
558 modeling community. However, the MEMENTO database does not currently include a Global

559 Data Product that includes publicly accessible quality controlled dissolved CH₄ and N₂O
560 datasets. The international marine carbon science community has widely embraced such an
561 approach for fCO₂, by submitting data to the Surface Ocean CO₂ Atlas (SOCAT), which was
562 initiated in response to the need for a quality controlled, publicly available, global surface CO₂
563 dataset (e.g. Bakker et al., 2016). Due to the fewer measurements, a similar data product for
564 marine CH₄ and N₂O would be needed every ~5 years. We consider the production of Global
565 Data Products for dissolved CH₄ and N₂O to be essential for supporting future global modeling
566 efforts and to enhance field observations.

567 The benefits of pursuing the three activities described above have already been clearly
568 demonstrated for carbon system measurements in the ocean. The intercomparability and high
569 accuracy and precision of carbon system measurements was achieved by streamlining
570 methodological approaches, universally adopting agreed SOPs, production of reference material,
571 and following community-driven quality control procedures (Dickson et al., 2007, Dickson et al.,
572 2010). It is encouraging to see the marine CH₄ and N₂O community beginning to move in a
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574
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587
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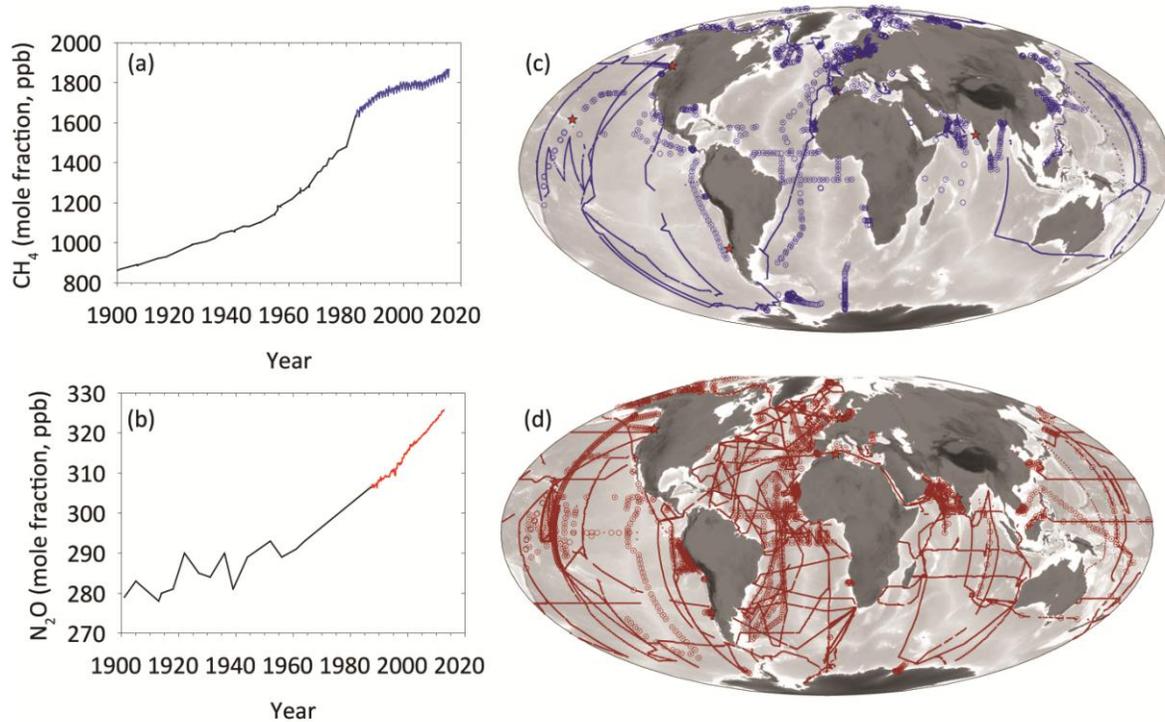
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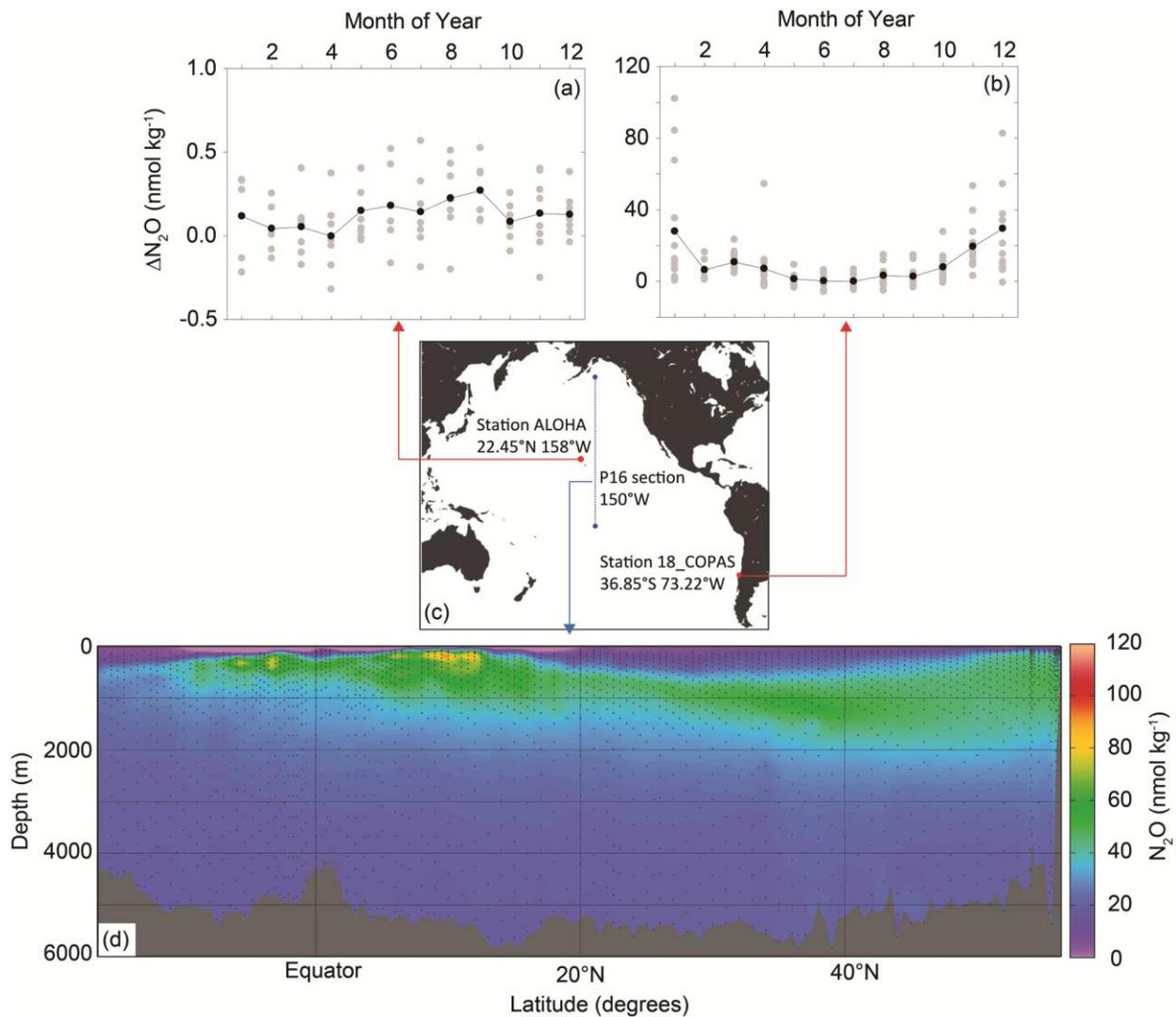
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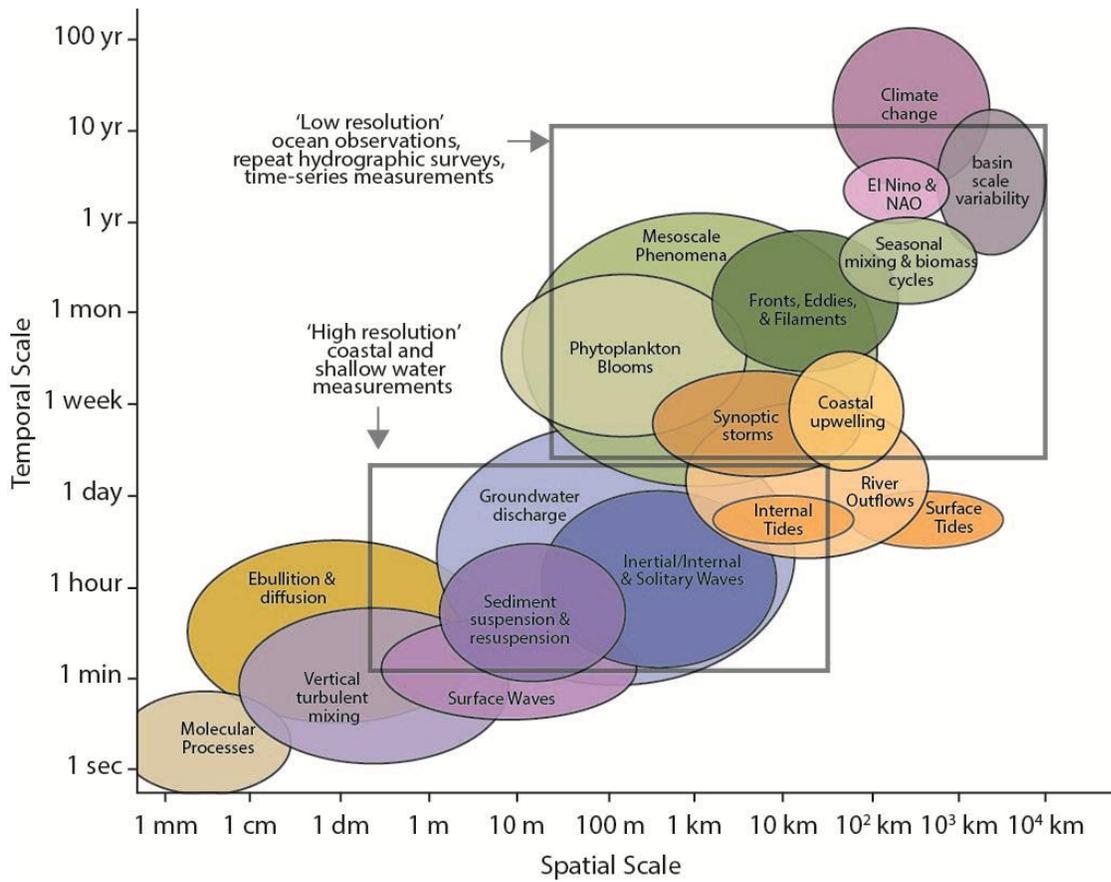


1214
 1215 Figure 1. Atmospheric values of (a) CH₄ and (b) N₂O with the black lines reconstructed from
 1216 ice-core measurements (Etheridge et al., 1998; Machida et al., 1995) and the colored lines from
 1217 Mauna Loa Observatory (<https://www.esrl.noaa.gov/gmd/dv/data/>). Global maps of marine (c)
 1218 CH₄ and (d) N₂O measurements available from the MEMENTO database
 1219 (<https://memento.geomar.de/>). The 2018 workshop focused on the marine contribution to
 1220 atmospheric CH₄ and N₂O and the underlying microbial and biogeochemical control
 1221 mechanisms.
 1222



1223
 1224 Figure 2. Repeat oceanic observations include both (a, b) fixed location time-series monitoring
 1225 observations and (c,d) hydrographic surveys. Together, such field observation programs helps
 1226 resolve temporal variability ranging from months to years and spatial variability at the ocean
 1227 basin scale (see Fig. 3). The Station ALOHA data derive from Wilson et al. (2018), the Station
 1228 18 data derive from Farías et al. (2015), and the P16 transect was conducted in 2015 by the
 1229 NOAA PMEL Tracer Group as part of the GO-SHIP program. The data shown in the plots are
 1230 N_2O concentrations, either as ΔN_2O (i.e. deviation from equilibrium value) or absolute values.

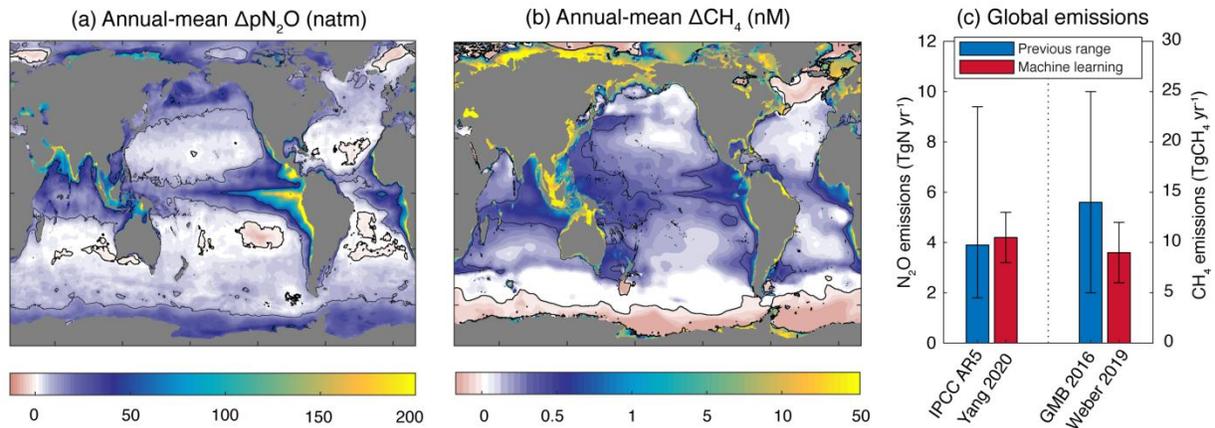
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1237 Figure 3. Time-space scale diagram illustrating various physical, biological, and climatological
 1238 processes relevant to marine CH₄ and N₂O (adapted from Dickey, 2003). To date, the majority
 1239 of marine CH₄ and N₂O measurements resolve variability at the mesoscale level or higher.

1240 Recent technological developments and the need to resolve concentrations and fluxes in shallow
 1241 water environments will increase the number of measurements conducted at the submesoscale
 1242 level (see Fig. 5). The low resolution oceanographic surveys are more likely to achieve a high
 1243 level of analytical accuracy compared to high resolution coastal measurements, however this is
 1244 compensated for by high temporal resolution achieved by underway sampling.



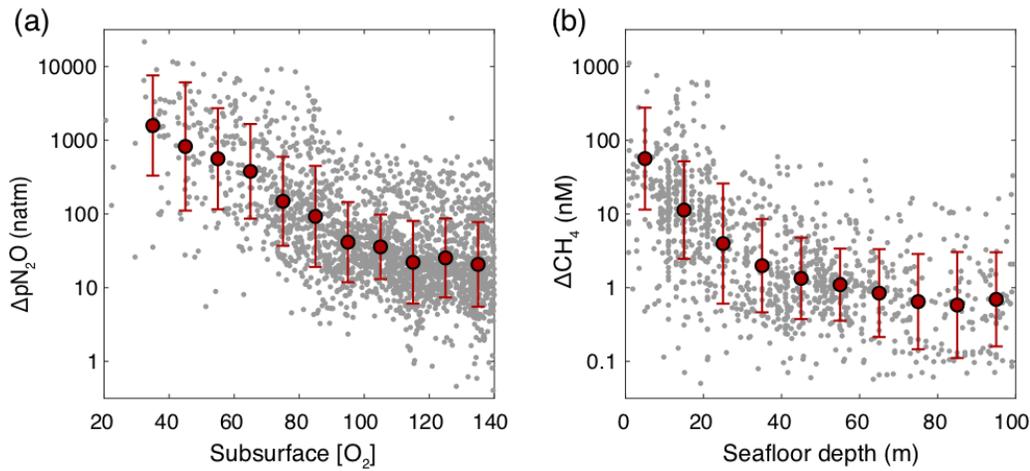
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1247 Figure 4. Distributions and emissions of marine CH_4 and N_2O , (a) Air-sea N_2O disequilibrium
 1248 mapped using a Regression Forest model (adapted from Yang et al., 2020), (b) Air-sea CH_4
 1249 disequilibrium mapped using an Artificial Neural Network model (adapted from Weber et al.,
 1250 2019). For consistency with the original publications, the air-sea disequilibrium is shown in
 1251 different units for N_2O (partial pressure) and CH_4 (concentration). (c) A summary of global
 1252 ocean CH_4 and N_2O emissions estimated by Yang et al. (2020) and Weber et al. (2019),
 1253 compared to the estimates of the IPCC 5th Annual Report (IPCC AR5) and the Global Methane
 1254 Budget (Saunois et al., 2016).

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1258 Figure 5. Key environmental predictors of surface ocean CH₄ and N₂O gradients. (a) Excess air-
1259 sea N₂O is best predicted by O₂ concentrations in the subsurface water-column (base of the
1260 mixed layer to a depth of 100 m) (adapted from Yang et al., 2020). (b) Excess CH₄ is best
1261 predicted by seafloor depth, reflecting the supply from anoxic sediments (adapted from Weber et
1262 al., 2019). The grey dots represent individual data points and the red dots with error bars
1263 represent mean ±1s.d. of binned data, using O₂ bins of 10 μM width and seafloor depth bins of
1264 10 m width.

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