

Dear Prof. Treude,

We would like to thank you and the two anonymous referees for reviewing our paper and the insightful and constructive feedback. Please find our replies below where we explain how we addressed each comment and the changes that were made in our manuscript. Unless otherwise indicated, the line numbers in our replies refer to the revised manuscript. We include a marked-up version of our manuscript that highlights all the relevant changes that were made, and we now uploaded the high-resolution version of the figures at the end of our manuscript.

Kind regards,

Martijn Hermans

Prof. Treude's comments:

Instead of 'bottom water' you could also use the term "supernatant" or 'bottom-near" water. I was in the past criticized by physical oceanographers for using the term 'bottom water' in my sediment studies, because it is a fixed term in oceanography and covers a large volume (an entire water mass).

Reply: We prefer to use “bottom water” since for chemical oceanographers this is a common term. We now made more explicit what bottom water refers to see line No. 32: “Depletion of oxygen (O₂) in bottom waters (i.e. water directly above the seafloor)”

- The term "suboxic" opposite to oxic and anoxic is not ideal, because suboxic is also anoxic. But I am aware that the term is frequently used together with cable bacteria biogeochemistry. I am OK with keeping the term, but please double check that your final manuscript has a proper definition of the terms at the beginning (oxic = with O₂, suboxic = anoxic but not sulfidic, anoxic = anoxic and sulfidic).

Reply: We explain this in more detail in the revised version of our manuscript, see line No. 51-52: “This spatial coupling of surficial O₂ reduction with H₂S oxidation at several centimetres depth creates a suboxic zone that is devoid of any O₂ and H₂S”

- In Fig. 2B a red and a green profile are shown in the same graphs. Please use a different color to allow green-red handicapped people to enjoy it.

Reply: We updated the figure and changed the colours to blue and orange, so it can be read by colour blind people as well.

Anonymous Referee #1

Received and published: 13 August 2020

The study documents the effect of cable bacteria on sediment geochemistry in repacked sediment cores on sediment collected from the Black Sea. The findings support current paradigms in the cable bacteria literature. This is a nice case study that has been well presented and written up. The dominance of cable bacteria in the oxygen budget is an interesting finding. There is also some additional data on element mapping that sheds further light on the effect of cable bacteria on sediment geochemistry at different sites. I only have a few relatively minor comments and suggestions for improvement.

Reply: We thank Anonymous Referee #1 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

Comment #1 The authors used the classic sediment repacking method to start a cable growth cycle. What effect might homogenisation have had on the final findings? For example, would siderite be likely to be so close to the surface under normal circumstances.

Reply: The homogenisation likely did not have a significant effect on our final findings. Homogenisation of the sediment is essential to obtain proper replicate cores and is known to induce growth of cable bacteria. While some of the FeS and siderite might have oxidised during sieving, homogenising and repacking, this does not affect the conclusions of our experiment. There is still ample FeS and siderite present in our sediment cores at the start of the experiment, and our time-series indicate that both FeS and siderite were dissolved by cable bacteria activity over time. Both FeS and siderite are frequently observed in marine surface sediments (e.g. Sulu-Gambari et al. 2016).

Comment #2 Line 163 – no need to say ‘bottom water’ just water samples?

Reply: Here, we specifically use the term bottom water samples, since it refers to the overlying water in the cores. Water samples could also refer to water column or pore water samples. Therefore, we think it is best to use the term bottom water samples, to avoid potential confusion.

Comment #3 Line 172 – please elaborate a little on exactly what you mean by salt corrections.

Reply: Freeze-drying removes the water from samples. However, the salt stays behind in the dried sediment. Hence, the weight of the freeze-dried sediment used for the sequential extractions needs to be corrected for this salt content if we wish to calculate elemental/mineral concentrations, such as Fe oxide and metal bound P contents per gram of sediment. Without the salt correction the absolute

elemental/mineral concentrations would be underestimated. Hence, we subtract the weight of the salt from the freeze-dried sediment to calculate the ‘real’ weight of the dry sediment. This is now explained in more detail in our methods, see line **No. 175-178**: “After freeze-drying, the salt from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried sediment.”

Comment #4 Line 205 – could you add a sentence on how the embedding was achieved?

Reply: The resin embedding process is described in more detail, see line **No. 209-214**: “On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013). Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled glovebox and subsequently embedded with Spurr’s epoxy resin as described in Jilbert et al. (2008). After curing, the epoxy-embedded core was split vertically using a rock saw.”

Comment #5 Line 210 onwards – consider adding this to methods.

Reply: The methods used to obtain these elemental maps from the epoxy resin embedded surface sediments from the Gulf of Finland and Lake Grevelingen are described in detail in the other studies cited (Sulu-Gambari et al. 2016; Sulu-Gambari et al. 2018; Hermans et al. Submitted). Therefore, we prefer not to describe the sampling process of those resin embedded cores in section 2.4 of our methodology.

Comment #6 also Line 261 – Only Ca and Si fluxes are presented, I couldn’t see them?

Reply: These Ca and Si fluxes are presented in Fig. S10 and Table S4 in the Supplementary Information, see line **No. 480** in the manuscript.

Comment #7 Line 384 not clear what you mean here, please elaborate.

Reply: We have used Fick’s law for the calculation of the diffusive fluxes, which not does take the effect of the electric field generated by cable bacteria on the diffusion potential into account. The Nernst-Planck equation, however, extends Fick’s law, because solutes can also be moved with respect to the fluid by electrostatic forces. By using Fick’s law, we underestimate the SO_4^{2-} reduction rate by a most ~10-20%. We will describe this in more detail, see line **No. 393-397**: “Solutes can also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the relatively low strength of the electric field in the cores ($<0.073\text{V m}^{-1}$ at day 18; as estimated from Fig 2B), including the contribution of ionic drift to the sulphate flux would lead to SO_4^{2-} reduction rates that are at most 10-20% higher.”

Comment #8 Line 415 – could it be that the nitrate is just denitrified? Also on this point, it seems that not flux measurements were made for nitrate. It seems likely that some nitrate is released to the water. It might be worth a brief discussion of a few scenarios here. All the nitrate is released to the water column, all the nitrate is denitrified by sediment bacteria and all the nitrate is denitrified by cable bacteria.

Reply: Unfortunately, we do not have data for NO_3^- . The problem of the abovementioned scenarios is that, if we would include the role of other groups of bacteria, and the potential release of NO_3^- to the water column, the mass balance for O_2 would have an even greater mismatch. When looking at the stoichiometry of NO_3^- by cable bacteria and the conversion of N to N_2 , we cannot close the budget fully that way. The text is now modified and explains this in more detail, see line **No. 425-430**: “These findings can be explained, however, if we assume that at least part of the NO_3^- that is being formed near the sediment-water interface is also used for the metabolic activity of cable bacteria. It has been shown that cable bacteria can couple the oxidation of $\sum\text{H}_2\text{S}$ to NO_3^- in the absence of O_2 (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments where O_2 is present in concert with NO_3^- near the sediment-water interface. However, we cannot exclude release of NO_3^- to the water column or denitrification by other bacteria in the sediment.”

Comment #9 Line 485 – very interesting!

Reply: Thank you, this potential niche for vivianite formation is indeed an interesting finding.

Comment #10 Line 522 – not obvious to me from Fig 8A, it is interesting, can you make this clearer?

Reply: This is now more explicit in the text, see line **No. 524-527**: “While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P is strongly correlated with Ca.”

Comment #11 Line 546 I agree this is likely driven by cables, but how is this different from a straight reaction diffusion scenario (given ubiquity of cables, such a scenario does seem unlikely though). I think this idea needs a little more development and explanation as to how it might actually be applied.

Reply: Focusing of Fe and Mn oxides and associated P can indeed also occur in sediments overlain by oxic waters, where no cable bacteria are active. However, as demonstrated by our experiment (and various other studies (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016), in sediment populated by active cable bacteria, the upward fluxes of Fe^{2+} and Mn^{2+} are higher due to the dissolution of FeS, Fe- and Mn carbonates. This allows strong focusing of Fe and Mn oxides in a thin layer within a relatively short time

frame. We added the following section in our manuscript, see line No. 567-577: “Focussing of Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and can also occur in the absence of cable bacteria. However, the upward fluxes of Fe^{2+} and Mn^{2+} in sediments populated by cable bacteria are higher due to active dissolution of Fe and Mn minerals at depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame following an environmental perturbation (such as a transition to oxic bottom waters after a period of anoxia or mixing of the sediment), more Fe^{2+} and Mn^{2+} can oxidise upon contact with O_2 near the sediment-water interface and thus stronger enrichments of Fe and Mn minerals will be observed. Hence, focussing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in sediments populated by active cable bacteria compared to sediments where no cable bacteria are active under such conditions.”

Comment #12 Figure 3 not clear how this was generated. Based on the pictures?

Reply: The depth intervals of the oxic, suboxic and anoxic zone are based on the micro-electrode data. We made this more explicit in the caption of Fig. 3., see line No. 948-950: “Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two different cores.”

References

- Bockris, J. O. M., and Reddy, A. K. (1998). Ion-ion interactions. *Modern Electrochemistry 1: Ionics*: 225-359.
- Hermans, M., Astudillo Pascual, M., Behrends, T., Lenstra, W. K., Conley, D. J., and Slomp, C. P. (Submitted). Coupled dynamics of iron, manganese and phosphorus in brackish coastal sediments populated by cable bacteria.
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- Sulu-Gambari, F. and others (2018). Phosphorus cycling and burial in sediments of a seasonally hypoxic Marine Basin. *Estuaries and Coasts* **41**: 921-939.
- Sulu-Gambari, F., Seitaj, D., Behrends, T., Banerjee, D., Meysman, F. J., and Slomp, C. P. (2016). Impact of cable bacteria on sedimentary iron and manganese dynamics in a seasonally-hypoxic marine basin. *Geochimica et Cosmochimica Acta* **192**: 49-69.

Anonymous Referee #2

Received and published: 31 August 2020

The manuscript represents a very comprehensive study of potential processes and effects of cable bacteria in sediments. Investigations on cable bacteria and their influence on biogeochemical processes are still in the beginning, but more and more studies show their importance for the element cycling; importance of cable bacteria activity on the oxygen demand in coastal sediments. In the present study, the authors used sediment cores from the coastal area of the Black Sea, which they homogenized and freed from macrofauna. This probably increased the availability of labile organic material and its distribution in deeper sediment layers. Furthermore, the sediment was anoxically stored until the experiment, during the experiment the overlying bottom water was saturated with oxygen so that a steady state must be established at the beginning of the incubations. This fact does not reduce the results of the experiment or the quality of the manuscript.

Reply: We thank Anonymous Referee #2 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

Comment #1 However, the authors should consider the study presented here as potential processes and not directly related to a coastal region (in this case the Black Sea). Therefore, I would strongly suggest to rewrite the manuscript and change the focus of the manuscript by concentrating on the "potential processes and bio geochemical impacts" rather than to directly relate it to coastal sediments of the Black Sea.

Reply: We are aware that the outcomes from our experiment are potential processes, which cannot be directly translated to the field site. This is why, in our title we specifically used the term "on coastal Black Sea sediment" instead of "in coastal Black Sea sediments". Other examples of sentences in our manuscript (here these line numbers refer to the old version of our manuscript) that indicate that we are not directly relating our results to a coastal region are:

- **Line No. 100:** "In this study, we assess whether cable bacteria activity can establish in sediments that are relatively poor in FeS in an incubation experiment using siderite-bearing sediments from a coastal site in the Black Sea."
- **Line No. 552:** "We can only speculate about the possible *in-situ* relevance of cable bacteria at the coastal site in the western Black Sea where the sediment for our incubation was collected."
- **Line No. 565-566:** "Further field studies are required to assess the role of cable bacteria at our field site, preferably including an assessment of the burrow structures."

We added the following additional text in the abstract, introduction and conclusion sections to further emphasise that we are referring to potential processes.

Abstract; line No. 14-15: “to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and sulphur (S).”

Introduction; line No. 100-103: “In this study, we assess whether cable bacteria activity can establish and thrive in sediments that are relatively poor in FeS. Although, this will be done in a controlled incubation experiment with siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural environments populated by cable bacteria.”

Conclusion; line No. 599-605: “The results of our laboratory incubation (with a total duration of 621 days) show that cable bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The strong acidity of the pore water associated with the activity of cable bacteria, which was monitored using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a source of Fe that can form an Fe oxide-enriched surface layer.”

Comment #2 The difference between the natural distribution of cable bacteria and the experiment is also evident when looking at Fig. 1c. The authors can use their main results as shown here, but the focus should be on the conditions used in their experiment, which are rather artificial, but very nicely show the potential of cable bacteria in the biogeochemical cycling.

Reply: See our reply to comment #1 and the associated changes in the text. We are aware that the outcomes of our experiment regarding the impact of cable bacteria are amplified when compared to field conditions and that we cannot directly link this to the field site. This is because we optimized conditions to sustain metabolic activity of cable bacteria, and their growth:

- 1) The sediment was homogenised, which is known to induce growth of cable bacteria.
- 2) There was no bioturbation by meiofauna and macrofauna.
- 3) The bottom water was continuously oxygenated.

This optimisation was deliberately done to study the effects of cable bacteria on sediment biogeochemistry. We note that the approach used is common in studies of the biogeochemical impact of cable bacteria (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016).

Comment #3 In a second step the transfer to coastal sediments and their biogeochemical conditions can be done. Here the manuscript lacks the coherence (hypoxia and oxy-gen depletion as mentioned in the Introduction). In a final paragraph the transfer of the laboratory experiment to natural sediments and possible variations in biogeochemical processes as well as the influence of macrofauna (bioturbation and bioirrigation) can be discussed.

Reply: Sections 4.1 to 4.4 focus only on the experiment. In sections 4.5 and 4.6 we discuss the implications for the field. We added text to section 4.5 and 4.6 to clarify when we are referring to other laboratory experiments and results of field studies and the link with hypoxia.

Line No. 520-522: “This colour zonation is typical for sediments that have been geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016)”

Line No. 565-567: “may act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory experiments and at field sites, also in cases where visual observations are not conclusive.”

Line 577-578: “Macrofaunal activity within natural environments likely counteracts or prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer.”

Line 584-586: “At this site, which is in a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m⁻²) and polychaetes (up to ~1700 ind. m⁻²)...”

Comment #4 Is there any information about the organic carbon content of the sediment and how this changes over the incubation period? I would assume that this is the major driver for the development of biogeochemical zonation.

Reply: We now include the organic carbon content of the upper 0-0.5 cm of the sediment at the field site, as determined by Lenstra et al. (2019), in Table 1. We did not determine the change in organic carbon contents during the experiment because, at the typical rates of organic matter degradation expected here, only a small change in organic carbon content would be observed. Hence, we chose to focus on pore water NH₄⁺ profiles to obtain insight in rates of (anaerobic) organic matter degradation.

Comment #5 How does the development of the oxic zone, as shown in the experiment, relate to natural variations in coastal sediments ?

Reply: The range in O₂ penetration observed in the experiment is comparable to that observed in coastal systems with seasonal hypoxia (e.g. Seitaj et al. 2015). We included this in the manuscript: Line No. 531-534: “During the experiment, O₂ penetration varied within a narrow range and was initially fixed between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth of 2 mm (Fig. 8A). Such a range in O₂ penetration is in accordance with observations in coastal sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by...”

Comment #6 How does the experiment relate to the development of hypoxia and depletion of oxygen in coastal areas ? The experiment shows the opposite reaction (from anoxic surface layer to an oxygenated layer).

Reply: As shown in previous work that we refer to in our manuscript (Seitaj et al. 2015; Sulu-Gambari et al. 2016), cable bacteria can induce formation of Fe and Mn oxides in seasonally hypoxic coastal systems during periods of oxygenated bottom waters in spring. As explained in line No. 41-44, the presence of these Fe and Mn oxides can delay the transition towards euxinia by removing hydrogen sulfide. Our experimental setting can be compared to the onset of bottom water re-oxygenation in spring in such seasonally hypoxic environments (i.e. the sediment was stored anoxically, and then exposed to oxygenated water). This allows study of the mineral dissolution and formation reactions in sediments populated by cable bacteria under such conditions, which is the goal of this work.

Comment #7 line2 121/122: with overlying waterWas this bottom water taken from the site or artificial water, as used for the aquarium?

Reply: The 20 cm long cores (filled with 15 cm sediment) were gently submerged in the two aquaria. Hence, the overlying water in the cores at the start of each incubation was the same as the artificial water used in the aquaria.

Comment #8 line 153: core was place outside the aquariumWhy was the core taken out ? was the incubation temperature maintained?

Reply: The entire experiment was carried out at in a temperature controlled laboratory at 20 °C. We made explicit that the flux incubations also took place at 20 °C, see line No. 153-154: “At each time point, one core was placed outside the aquarium at 20 °C...”

Comment #9 Was the overlying water during the 24-hour incubation for the solute flux measurements stirred to avoid stratification? This could have influenced the flux across the sediment-water interface

because stagnant waters lead to an increase of the Diffusive Boundary Layer, which controls the solute exchange.

Reply: The overlying water was mixed continuously by bubbling it continuously with air. The airflow was set in such a way that the water was effectively mixed, while the surface layer of the sediment was left undisturbed. We made this more explicit in the methods, see line **No. 153-157**: “At each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of overlying water in the core was continuously aerated. Potential stratification of the overlying water was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent evaporation.”

Comment #10 Pore water profiles (especially Fig 1a, Fig 2) are very small and it is difficult to recognize the different profiles (O₂, pH, H₂S) different; graphs should be enlarged.

Reply: We enlarged these graphs and we increased the font sizes to improve the readability of the figures.

References

- Bockris, J. O. M., and Reddy, A. K. (1998). Ion-ion interactions. *Modern Electrochemistry 1: Ionics*: 225-359.
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1 **Biogeochemical Impact of Cable Bacteria on Coastal Black Sea Sediment**

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12 **ABSTRACT**

13 Cable bacteria can strongly alter sediment biogeochemistry. Here, we used laboratory
14 incubations to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and
15 sulphur (S). Microsensor depth profiles of oxygen, sulphide and pH in combination with electric
16 potential profiling and FISH analyses showed a rapid development (<5 days) of cable bacteria,
17 followed by a long period of activity (>200 days). During most of the experiment, the current density
18 correlated linearly with the oxygen demand. Sediment oxygen uptake was attributed to activity of
19 cable bacteria and the oxidation of reduced products from anaerobic degradation of organic matter,
20 such as ammonium. Pore water sulphide was low (<5 μM) throughout the experiment. Sulphate
21 reduction acted as the main source of sulphide for cable bacteria. Pore water Fe^{2+} reached levels of up
22 to 1.7 mM during the incubations, due to the dissolution of FeS (30%) and siderite, an Fe carbonate
23 mineral (70%). Following upward diffusion of Fe^{2+} , a surface enrichment of Fe oxides formed. Hence,
24 besides FeS, siderite may act as a major source of Fe for Fe oxides in coastal surface sediments where
25 cable bacteria are active. Using μXRF , we show that the enrichments in Fe oxides induced by cable
26 bacteria are located in a thin subsurface layer of 0.3 mm. We show that similar subsurface layers

27 enriched in Fe and P are also observed at field sites where cable bacteria were recently active and
28 little bioturbation occurs. This suggests that such subsurface Fe oxide layers, which are not always
29 visible to the eye, could potentially be a marker for recent activity of cable bacteria.

30 Key words: cable bacteria, elemental cycling, solute fluxes, iron

31 **1. INTRODUCTION**

32 Depletion of oxygen (O₂) in bottom waters (i.e. water directly above the seafloor) of coastal areas
33 is increasing worldwide, as a consequence of eutrophication and climate change (Diaz and Rosenberg
34 2008; Schmidtke et al. 2017; Breitburg et al. 2018). Low O₂ can lead to the development of coastal
35 ‘dead zones’ characterised by recurrent mortality of marine life (Rabalais et al. 2002; Diaz and
36 Rosenberg 2008). Progressive eutrophication induces a characteristic response of coastal systems with
37 transient and seasonal hypoxia (O₂ < 63 μM) transitioning into permanent anoxia (O₂ = 0 μM). In this
38 later stage, free sulphide (H₂S) may escape from the sediment and accumulate in the bottom water, a
39 condition referred to as euxinia (Diaz and Rosenberg 2008; Kemp et al. 2009; Rabalais et al. 2014).
40 As H₂S is highly toxic to higher fauna, the development of euxinia may aggravate the ecological
41 consequences. However, the presence of iron (Fe) and manganese (Mn) oxides in surface sediments
42 may delay this transition towards euxinia by removing H₂S and thus preventing an efflux of H₂S to
43 the overlying water (Kristiansen et al. 2002; Kristensen et al. 2003; Diaz and Rosenberg 2008).

44 Cable bacteria are multicellular filamentous sulphur(S)-oxidising bacteria (Pfeffer et al. 2012)
45 that strongly enhance the formation of Fe and Mn oxides and efficiently remove H₂S from surface
46 sediments (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Sulu-Gambari et al. 2016a). Cable
47 bacteria belong to the *Desulfobulbaceae* family of the Deltaproteobacteria (Trojan et al. 2016;
48 Kjeldsen et al. 2019). Cable bacteria can spatially link the oxidation of H₂S in deeper sediments to the
49 reduction of O₂ near the sediment-water interface by transporting electrons over centimetre scale
50 distances (Pfeffer et al. 2012) through a conductive fibre network that is embedded in the cell
51 envelope (Meysman et al. 2019). This spatial coupling of surficial O₂ reduction with H₂S oxidation at
52 several centimetres depth creates a suboxic zone that is devoid of any O₂ and H₂S, and provides cable

53 bacteria a competitive advantage over other S-oxidising bacteria in aquatic environments (Meysman
54 2018). Cable bacteria have been documented in a range of fresh water (Risgaard-Petersen et al. 2015;
55 Müller et al. 2016) and marine environments (Malkin et al. 2014; Burdorf et al. 2017), however, they
56 appear to be particularly active in sediments overlain by seasonally hypoxic bottom waters (Seitaj et
57 al. 2015; Burdorf et al. 2018).

58 The metabolic activity of cable bacteria establishes an electrical circuit in the sediment, which
59 involves an electron current through the cable bacteria filaments (Bjerg et al. 2018), and an ionic
60 current through the pore water in the opposite direction (Naudet and Revil 2005; Revil et al. 2010;
61 Risgaard-Petersen et al. 2012). As a consequence, an electric potential (EP) is generated in the
62 sediment, which can be used as a reliable indicator for activity of cable bacteria (Risgaard-Petersen et
63 al. 2014).

64 Cable bacteria activity additionally generates a distinct biogeochemical signature, that can be
65 assessed by pH, O₂ and H₂S depth profiling (Nielsen et al. 2010). Their activity leads to the
66 development of a suboxic zone (i.e. a zone where O₂ and H₂S are both absent), and also induces a pH
67 profile that strongly changes with depth. Cathodic O₂ reduction ($O_2 + 4H^+ + 4e^- \rightarrow 2H_2O$) in the oxic
68 zone of the sediment results in a pH maximum (~9) due to proton consumption, whereas anodic
69 sulphide oxidation ($H_2S + 4 H_2O \rightarrow SO_4^{2-} + 10H^+ + 8e^-$) causes a pH minimum (<6.5) in the anoxic
70 zone (Fig. 1A; Nielsen et al. 2010; Meysman et al. 2015).

71 The presence of cable bacteria in sediments can strongly impact the elemental cycling of Fe,
72 Mn, Ca and S (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Rao et al. 2016; Sulu-Gambari et al.
73 2016a; van de Velde et al. 2016). Pore water acidification induced by cable bacteria activity can lead
74 to dissolution of calcium (Ca) carbonates, Fe carbonates (siderite), Mn carbonates and FeS in the zone
75 where the pH is low, thus generating high concentrations of Fe²⁺ and Mn²⁺ in the pore water
76 (Risgaard-Petersen et al. 2012; Rao et al. 2016). When these dissolved species diffuse upward this can
77 lead to strong enrichments of Fe and Mn oxides upon contact with O₂, or for dissolved Fe²⁺, also upon
78 contact with Mn oxides (Wang and Van Cappellen 1996; Seitaj et al. 2015; Sulu-Gambari et al.

79 2016a). These metal oxides are capable of efficiently buffering the benthic release of H₂S and
80 phosphate (HPO₄²⁻) during periods with low bottom water O₂. This so-called ‘firewall’ for H₂S and
81 alteration of the timing of HPO₄²⁻ release linked to this buffering can play a key role in regulating
82 water quality in seasonally hypoxic coastal systems (Seitaj et al. 2015; Sulu-Gambari et al. 2016b;
83 Hermans et al. 2019a).

84 In coastal sediments, O₂ typically penetrates to a depth of only several mm’s below the
85 sediment-water interface (Rasmussen and Jørgensen 1992; Rabouille et al. 2003; Glud 2008). This
86 also holds true for sediments inhabited by active cable bacteria (Nielsen et al. 2010; Pfeffer et al.
87 2012; Larsen et al. 2015). Hence, the oxidation of upward diffusing Fe²⁺ and Mn²⁺ is expected to take
88 place below and not at the sediment-water interface. We hypothesise that, as a consequence, in the
89 initial stages of cable bacteria activity and in the absence of bioturbation, most Fe and Mn oxide
90 enrichments will be restricted to a thin subsurface layer of the sediment. However, the sample
91 resolution and timing of the collection of solid phase data in field and laboratory studies published so
92 far do not allow an assessment of this hypothesis.

93 Cable bacteria are suggested to thrive in coastal sediments characterised by high rates of H₂S
94 production due to high rates of organic matter mineralisation (Malkin et al. 2014; Burdorf et al. 2017;
95 Hermans et al. 2019a). Laboratory and model studies have shown that the dissolution of FeS accounts
96 for 12 to 94% of the H₂S consumed by cable bacteria, while the other source is H₂S production from
97 the reduction of SO₄²⁻ (Risgaard-Petersen et al. 2012; Meysman et al. 2015; Burdorf et al. 2018). At
98 present, it is not known if cable bacteria activity can establish in sediments that are relatively low in
99 FeS and dissolved H₂S.

100 In this study, we assess whether cable bacteria activity can establish and thrive in sediments
101 that are relatively poor in FeS. Although, this will be done in a controlled incubation experiment with
102 siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural
103 environments populated by cable bacteria. The metabolic activity of cable bacteria is monitored using
104 microsensor profiles of pH, O₂, H₂S and EP. We also use sediment Fe and P speciation and μ XRF of

105 resin-embedded sediments to test whether we find evidence for subsurface enrichments in Fe oxides
106 and associated P. We find a rapid establishment of cable bacteria (<5 days) and the development of an
107 Fe oxide-rich subsurface layer, with the majority of the Fe ~70% supplied through dissolution of
108 siderite induced by cable bacteria activity. The depth of the Fe oxide layer was directly related to the
109 O₂ penetration depth and we propose that such subsurface enrichments in Fe, which also can contain P
110 and Mn, can be used as a marker for recent cable bacteria activity.

111 **2. METHODS AND MATERIALS**

112 **2.1. Study Area and Experimental Set-up**

113 In September 2015, 16 sediment cores (ø10 cm) were retrieved at a coastal site on the north-
114 western shelf of the Black Sea (27 m water depth; Fig. 1B; Table 1) using a multicorer (Oktopus
115 GmbH, Germany) as described in Lenstra et al. (2019). The overlying water was discarded, and the
116 upper 10 cm of the sediment was transferred into nitrogen purged aluminium bags that were sealed
117 and stored at 4 °C for several months. The anoxic storage is expected to have led to the death of all
118 macrofauna and most meiofauna (Coull and Chandler 2001; Riedel et al. 2012). Prior to incubation,
119 the sediment was passed through a 4 mm sieve to remove large debris and homogenised.
120 Subsequently, the sediment was transferred to 18 transparent polycarbonate cores (ø6 cm; 20 cm
121 length).

122 The bottom 15 cm of these cores was filled with sediment and the upper 5 cm with overlying
123 water. The cores were placed in two aquaria filled with artificial seawater (Instant Ocean Sea Salt +
124 Ultra High Quality (UHQ) water) with a salinity of 17.9, identical to the bottom water salinity at the
125 study site. The artificial sea water contained negligible concentrations of NH₄⁺, NO₃⁻, Fe, Mn and P as
126 described in Atkinson (1997) and Hovanec and Coshland (2004). The aquaria were kept in the dark at
127 a constant temperature (~20 °C), and the water was continuously aerated by two aquarium pumps.
128 Sixteen out of eighteen cores were exposed to oxygenated overlying water in the aquaria, whereas the
129 two remaining cores served as an anoxic control treatment. The control cores were tightly sealed with
130 rubber stoppers, to prevent the growth of cable bacteria by excluding O₂ (Nielsen et al. 2010).

131 Sampling for pore water and solid-phase analyses was performed at eight time points over a
132 total incubation period of 621 days. Each time point involved a three day procedure. On the first day,
133 microsensor depth profiles of EP, O₂, pH and H₂S were obtained in two randomly selected oxic cores
134 and the two anoxic control cores (O₂ profiling was not performed in the anoxic cores). On the second
135 day, solute fluxes were measured in the same oxic cores that were used for microsensor depth
136 profiling on the previous day. On the third day, the two cores were sectioned, of which only one core
137 was processed further for pore water and solid-phase analyses. Photographs were taken at four time
138 points (day 12; 33; 170 and 621) from one oxic core to follow the visual development of the surface
139 sediment during the experiment.

140 **2.2. High-resolution Microsensor Depth Profiling**

141 High-resolution depth profiles of pH, O₂ and H₂S were obtained (50- μ m depth resolution; 3
142 replicate profiles per oxic core; 2 replicate profiles per anoxic core) using commercial micro
143 electrodes (Unisense A.S., Denmark). The O₂ sensor was re-calibrated prior to each measurement,
144 using saturated bottom water (100% [O₂]) and the deeper sediment horizons (0% [O₂]) as calibration
145 points. Calibrations of the pH and H₂S electrodes were performed as described in Hermans et al.
146 (2019b). pH values are reported on the total scale. For depth profiling of EP (500- μ m resolution; 3
147 replicates per core), micro electrodes were used that were custom built at Aarhus University, as
148 described in Damgaard et al. (2014). A robust reference electrode (Ref-RM, Unisense, A.S.,
149 Denmark) was used during EP and pH measurements. To exclude turbulence-induced variations in the
150 potential of the reference electrode during EP profiling, a silicon tube filled with foam was mounted
151 on the tip of the reference electrode.

152 **2.3. Solute Flux Measurements**

153 Solute flux incubations were performed for NH₄⁺, Fe²⁺, Mn²⁺, Ca²⁺, HPO₄²⁻ and H₄SiO₄. At
154 each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of
155 overlying water in the core was continuously aerated. Potential stratification of the overlying water
156 was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent
157 evaporation. Water samples of 3 mL were retrieved at 7 time points over 24 hours. The same volume

158 of fresh artificial seawater was added to the cores directly after taking each sample. The samples were
159 filtered (0.45 μm), and subsamples were taken for ammonium (1 mL) and for metals (1 mL; acidified
160 with 10 μL Suprapur® HCl (35%) per mL sample), which were stored at -20°C and 4°C respectively
161 until further analysis.

162 **2.4. Pore Water and Sediment Collection**

163 At each time point, two cores were sectioned at 0.5-1 cm resolution with an UWITEC push-
164 up pole in a nitrogen-purged glovebag, but only samples for one core were used for sediment and pore
165 water collection and analyses. Bottom water samples were retrieved from the overlying water in the
166 cores. Slices for each depth interval were centrifuged at 3500 rpm for 20 minutes for pore water
167 retrieval. Samples (1 mL) for NH_4^+ were taken and stored at -20°C until analysis. Samples (1 mL) for
168 pore water S, Fe, Mn, Ca, P and Si were also collected and acidified with 10 μL Suprapur® HCl
169 (35%) per mL sample, which were stored at 4°C until analysis. Centrifuged sediment samples were
170 freeze-dried and ground to a fine powder in a nitrogen-purged glovebox under a strictly anoxic
171 environment to prevent oxidation (Kraal et al. 2009; Kraal and Slomp 2014). Only the top 5 cm of the
172 solid-phase samples were analysed in further detail. The porosity (Supporting Information 1.1; Table
173 S1) was calculated from the weight loss upon freeze-drying, using a sediment density of 2.65 g cm^{-3}
174 (Burdige 2006). Salt corrections were performed on the solid-phase data using the gravimetric water
175 content and salinity to determine the amount of salt after freeze-drying. *After freeze-drying, the salt
176 from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry
177 sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried
178 sediment.*

179 **2.5. Chemical Analysis of the Water and Sediment**

180 Concentrations of NH_4^+ in the pore water and solute flux samples were determined using the
181 phenol hypochlorite method (Koroleff 1969). The total Fe, Mn, Ca, P and Si concentrations (which
182 are assumed to represent Fe^{2+} , Mn^{2+} , Ca^{2+} , HPO_4^{2-} and H_4SiO_4) in the pore water and solute flux
183 samples were determined using Inductively Coupled Plasma-Optical Emission Spectroscopy (ICP-
184 OES, Spectro Arcos). Dissolved Fe and Mn are assumed to be present in the form of Fe^{2+} and Mn^{2+} ,

185 however some Mn^{3+} (Madison et al. 2013) or colloidal and nanoparticulate Fe and Mn might also be
186 available (Boyd and Ellwood 2010; Raiswell and Canfield 2012). Concentrations of P and S are
187 assumed to represent HPO_4^{2-} and SO_4^{2-} respectively. The colourimetric detection limit for NH_4^+ was
188 0.5 μM . The practical detection limit on the ICP-OES for Fe, Mn and P was 0.73, 0.11 and 7.30 μM ,
189 respectively.

190 Solid-phase Fe was fractionated into [1] labile ferric Fe (hydr)oxides and ferrous Fe ($FeS +$
191 $FeCO_3$), [2] crystalline Fe minerals, [3] magnetite and [4] pyrite (Supporting Information 1.2; Table
192 S2), using a combination of two operational extraction methods (Poulton and Canfield 2005; Claff et
193 al. 2010) as described by Kraal et al. (2017). Concentrations of Fe in all extracts were determined
194 using the colourimetric phenanthroline method (APHA 2005). Solid-phase S was separated into [1]
195 acid volatile sulphur (AVS; representing FeS) and [2] chromium reducible sulphur (CRS; representing
196 FeS_2 ; Table S1) using the method after Burton et al. (2006; 2008) as modified by Kraal et al. (2013).
197 Sulphide released during the S extraction was trapped as ZnS in alkaline Zn acetate traps.
198 Concentrations of S were determined by iodometric titration (APHA 2005). Solid-phase siderite
199 ($FeCO_3$) was determined by subtracting AVS from the labile ferrous concentrations retrieved from the
200 first step of the Fe extraction. Solid-phase P was fractionated into [1] exchangeable P, [2] citrate-
201 dithionite-bicarbonate (CDB)-P, [3] authigenic P, [4] detrital P and [5] organic P (Table S2) after
202 Ruttenberg (1992) as modified by Slomp et al. (1996). The sum of exchangeable P and CDB-P
203 represents metal bound P, as described in Hermans et al. (2019b). Concentrations of P in all extracts,
204 except CDB, were measured with the molybdenum blue colourimetric method (Murphy and Riley
205 1958). The P, Mn (assuming to represent Mn oxides; Hermans et al. 2019b) and Si (assuming to
206 represent metal oxide bound Si; Kostka and Luther III 1994; Rao et al. 2016) in CDB extracts was
207 determined using ICP-OES.

208 **2.6. Elemental Mapping of Fe, Mn, P and Ca**

209 On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin
210 embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013).
211 Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm

212 diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled
213 glovebox and subsequently embedded with Spurr's epoxy resin as described in Jilbert et al. (2008).
214 After curing, the epoxy-embedded core was split vertically using a rock saw. The surface was
215 smoothed by applying a 0.3 μm alumina powder layer. Elemental maps of Fe, Mn, P and Ca (30 μm
216 resolution) were retrieved using a Desktop EDAX Orbis μXRF analyser (Rh tube set at 30 kV,
217 500 μA , 300 ms dwell-time, equipped with a poly-capillary lens). Similar μXRF maps for Fe, Mn and
218 P in epoxy embedded surface sediment were obtained for two field sites: (1) the Gulf of Finland, for
219 sediments collected in June 2016 as described by Hermans et al. (Submitted), and Lake Grevelingen,
220 for sediments collected in January and May 2012 as described in Sulu-Gambari et al. (2016a; 2018).

221 **2.7. Fluorescence *In-situ* Hybridisation**

222 Fluorescence *in-situ* hybridisation (FISH; Pernthaler et al. 2001) was used to microscopically
223 quantify the abundance of cable bacteria filaments, as described in Seitaj et al. (2015). FISH analysis
224 was performed on one intact sediment core retrieved at our sampling site, and the sediment cores from
225 our incubation experiment used for pore water collection at three time points (days 5, 26 and 207).
226 These cores were sectioned at 0.5 cm depth resolution for the first 2.5 cm. Each sediment slice was
227 homogenised and fixed with 0.5 mL ethanol ($\geq 99.8\%$ purity), and stored in a 2 mL Eppendorf tube at
228 $-20\text{ }^{\circ}\text{C}$. For FISH analysis, a volume of 100 μL was retrieved from the Eppendorf tubes and mixed
229 with a 1:1 solution of PBS/ethanol (500 μL). Then 10 μL of this mixture was filtered through a
230 polycarbonate membrane (type GTTP; pore size 0.2 μm , Millipore, USA). Cable bacteria were
231 classified with a *Desulfobulbaceae*-specific oligonucleotide probe (DSB706; 5-ACC CGT ATT CCT
232 CCC GAT-3') after counter staining with DAPI (1 $\mu\text{g}/\text{mL}$) under an epifluorescence microscope
233 (Zeiss Axioplan, Germany) at 100x magnification. The abundance of cable bacteria was quantified by
234 determining the length and diameter of all observed filaments in a field ($105 \times 141\text{ }\mu\text{m}$) on the filter at
235 100x magnification (200 fields per sample). Cable bacterial abundances are expressed as filament
236 length per volumetric unit (m cm^{-3}) or depth integrated per unit area of sediment surface (m cm^{-2}),
237 consistent with previous studies (Schauer et al. 2014; Malkin et al. 2017).

238 **2.8. Scanning Electron Microscopy**

239 Cable bacteria filaments were taken from surface sediments from the oxic zone (upper 2 mm)
240 after 40 days using a microscope and were transferred to a 15 mL centrifuge tube. The tube was filled
241 to a volume of ~10 mL using ultra clean water, and was subsequently centrifuged at 2100 rpm for 2
242 min, after which the water was discarded. This washing step was repeated three times. The washed
243 samples were then transferred to a sample stub, where the sediment was air-dried over-night prior to
244 gold coating. The filaments were subsequently subjected to scanning electron microscopy (SEM)
245 imaging on a Phenom ProX Desktop SEM (Phenom-World B.V., the Netherlands) to obtain high-
246 resolution images, as described in Geerlings et al. (2019). SEM images were generated under 0.1-0.3
247 mbar vacuum, and a high accelerating voltage (10 or 15 kV).

248 **2.9. Data Analysis and Calculations**

249 The diffusive uptake of O₂ was calculated from the high-resolution O₂ depth profiles using the
250 PROFILE software package (Berg et al. 1998). Total H₂S ($\Sigma\text{H}_2\text{S} = \text{H}_2\text{S} + \text{HS}^- + \text{S}^{2-}$) was calculated as
251 a function of the recorded H₂S and pH values, accounting for temperature and salinity (Millero et al.
252 1988; Jeroschewski et al. 1996).

253 The EP depth profiles were normalised by subtracting the background EP signal in the
254 overlying water from the EP depth profiles, to calculate the EP value relative to that in the overlying
255 water (Damgaard et al. 2014). The electric field in the sediment was calculated from the linear slope
256 of the EP depth profiles (average of triplicates) in the surface sediments (Risgaard-Petersen et al.
257 2014). The magnitude of the current density was subsequently calculated from the gradient in the EP,
258 the so-called electric field, using Ohm's law:

$$J = \sigma_{sed} \cdot E \quad (1)$$

259 where J represents the magnitude of the current density (mA m⁻²), σ_{pw} is the conductivity of
260 the sediment matrix (S m⁻¹) and E (mV m⁻¹) represents the electric field. The conductivity of the pore
261 water was corrected for tortuosity and calculated as a function of the temperature and salinity using
262 the equations provided by Fofonoff and Millard Jr (1983).

263 The solute fluxes were calculated as described in Glud (2008) and Rao et al. (2016):

$$J = \frac{\Delta C_{ow}}{\Delta t} \cdot \frac{V_{ow}}{A} \quad (2)$$

264 where J represents the diffusive flux ($\text{mmol m}^{-2} \text{d}^{-1}$), ΔC_{ow} represents the concentration change in the
265 overlying water (mmol m^{-3}), Δt is the incubation time (d), V_{ow} is the volume in the overlying water
266 (m^3) and A the surface area of sediment in the core (m^2). In our experimental setup, only those fluxes
267 were measurable for NH_4^+ , Fe^{2+} , Mn^{2+} and HPO_4^{2-} , that were >0.08 , >0.06 , >0.01 and $>0.55 \text{ mmol m}^{-2}$
268 d^{-1} , respectively. However, for these four solutes, fluxes were always too low to be detected. Hence,
269 only Ca^{2+} and H_4SiO_4 fluxes are presented.

270 Diffusive downward fluxes of SO_4^{2-} and diffusive upward fluxes of NH_4^+ , Fe^{2+} , Mn^{2+} and
271 Ca^{2+} were calculated from linearized pore water gradients using Fick's first law (Berner 1980):

$$J = -\phi D_s \cdot \frac{dC}{dz} \quad (3)$$

272

273 The molecular diffusion coefficient was calculated as a function of pressure, salinity and
274 temperature using the R package *marelac* (Soetaert et al. 2010) and corrected for the ambient
275 tortuosity using the relations listed in Boudreau (1997).

276 3. RESULTS

277 3.1. Abundance of Cable Bacteria

278 Examination of the top 2.5 cm of the surface sediments using FISH showed the presence of
279 filamentous cable bacteria (Fig. 1C; Fig. S1). The *in-situ* cable bacterial abundance in the sediment at
280 our field site was low (14 m cm^{-2}). However, after 5 days of incubation in the laboratory, the
281 abundance increased strongly (724 m cm^{-2}). At day 26 the abundance of cable bacteria was even
282 higher (1035 m cm^{-2}). After 207 days, the cable bacterial abundance in the surface sediment was low
283 again (131 m cm^{-2}). SEM imaging confirmed that filaments were indeed cable bacteria (Fig. 1D), as
284 the external surface of the filament was characterised by a parallel pattern of ridges and grooves along

285 its latitudinal axis, which is a typical feature of cable bacteria (Cornelissen et al. 2018; Geerlings et al.
286 2019).

287 **3.2. High-resolution Depth Profiles of pH, O₂ ΣH₂S and EP**

288 High-resolution depth profiles of pH showed the development of a distinct peak near the
289 sediment-water interface at day 5, and acidification of the pore water in the deeper sediment (Fig.
290 2A). The width of this pore water acidification zone increased with time and reached its maximum at
291 day 26, followed by a decrease in the acidification. The distinct pH peak near the sediment-water
292 interface gradually disappeared after 33 days. The depth of O₂ penetration in the sediment remained
293 constant within the first 40 days of incubation (~1.1 mm) and subsequently moved downwards with
294 time to 9.6 mm (Fig. 2A; Fig. 3; Fig. S2). The dissolved ΣH₂S concentrations remained low (<5 μM)
295 throughout the experiment (Fig. 2A). The ΣH₂S appearance depth was initially equivalent to the O₂
296 penetration depth, and shifted downwards within 5 days, creating a suboxic zone where O₂ and ΣH₂S
297 remained below detection (Fig. 2A; Fig. 3). The width of the suboxic zone remained relatively
298 constant with time (~25 mm; Fig. 3), and only slight decreased after 207 days.

299 The EP depth profiles indicate a rapid establishment of an electric current after 5 days (0.4
300 mV; Fig. 2B). The time-series of depth profiles show that the EP increased and also accumulated over
301 a thicker depth horizon. At day 26 the EP reached its maximum value (1.2 mV), followed by a
302 decrease with time. Long-distance electron transport was not active in the anoxic control core (Fig.
303 S3).

304 **3.3. Diffusive Uptake of O₂ and Current Density**

305 The diffusive O₂ uptake of the sediment was highest after 5 days and gradually decreased
306 with time from ~30 to ~3.6 mmol m⁻² d⁻¹ (Fig. 4A). The current density rapidly increased from day 0
307 to day 18, from 6 to 128 e⁻ mmol m⁻² d⁻¹, and then gradually decreased with time (Fig. 4B). The
308 duplicate measurements show the same trend for the diffusive O₂ uptake and the current density,
309 which indicates that the results are reproducible.

310 **3.4. Pore Water Profiles**

311 Concentrations of NH_4^+ were low near the sediment-water interface and increased with
312 sediment depth reaching maximum levels of up to 1.7 mM at depth in the sediment (Fig. 5). The time-
313 series suggest a gradual decrease in production of dissolved NH_4^+ in the sediment leading to
314 decreasing concentrations with time. The pore water depth profiles of dissolved SO_4^{2-} show a decline
315 with sediment depth at all time points. However, SO_4^{2-} concentrations remained relatively constant
316 within the top 2 cm of surface sediment between day 12 and 33. Dissolved Fe^{2+} , Mn^{2+} and Ca^{2+} all
317 show the development of distinct peaks in the pore water with time, and after 40 days those peaks
318 disappear again. Pore water concentrations of HPO_4^{2-} generally increased with sediment depth for all
319 time points, and concentrations within the top 2 cm were below the detection limit indicating removal.
320 Dissolved H_4SiO_4 increased with sediment depth reaching concentration of up to 1 mM.

321 **3.5. Diffusive Fluxes**

322 Calculated diffusive fluxes of NH_4^+ into the oxic zone decreased during the incubation
323 experiment from 4.7 to 1.8 $\text{mmol m}^{-2} \text{d}^{-1}$ (Fig. 6A; Fig. S4; Table S3). Rates of SO_4^{2-} reduction
324 estimated from the linear gradient of the decrease in pore water SO_4^{2-} in the surface sediment with
325 depth generally also showed a decrease with time (Fig. 6B; Fig. S5; Table S3). The upward diffusive
326 flux of Fe^{2+} greatly increased from day 5 to day 12 and then gradually decreased with time (Fig 6C;
327 Fig. S6; Table S3). The upward diffusive flux of Mn showed an increase in the initial stage of the
328 experiment and reached its maximum at day 18, followed by a decrease with time (Fig. 6D; Fig. S7;
329 Table 3). The upward diffusive flux of Ca^{2+} showed no clear trend with time, however after 207 days
330 the flux became extremely low (Fig. 6E; Fig. S8). The upward diffusive flux of H_4SiO_4 also showed
331 an increase in the initial stage of the experiment, and reached its maximum at day 12, followed by a
332 decrease with time (Fig. 6F; Fig. S9).

333 **3.6. Solid-phase Profiles**

334 The surface sediment in the oxic cores became more enriched in Fe oxides with time, with
335 concentrations increasing from 53 to 485 $\mu\text{mol g}^{-1}$ (Figure 7). The deeper sediment in the oxic cores
336 and the entire anoxic control core had low or no Fe oxides. At day 5, FeS was strongly depleted

337 within the top 1 cm of the surface sediment and was gradually lost further with time. At day 621, most
338 of the FeS within the top 2.5 cm of the surface sediment had been dissolved. The anoxic core did not
339 show such a depletion of FeS in the surface sediment and even showed a slight increase in FeS. Solid-
340 phase siderite remained rather constant with depth from day 5 to 33, but afterwards was gradually lost
341 from the surface sediment. At day 621 a large proportion of the siderite was dissolved within the top 2
342 cm. Solid-phase siderite concentrations remained constant with depth in the anoxic control core.
343 Solid-phase depth profiles of Mn oxides, metal bound P and metal oxide bound Si all showed a
344 gradual increase in the surface sediment with time.

345 **3.7. High-resolution Elemental Mapping**

346 High-resolution desktop μ XRF mapping of Fe, Mn, P and Ca of our core after 47 days of
347 incubations revealed a subsurface sediment layer highly enriched in Fe and P (Fig. 8A). Subsurface
348 enrichments in Fe, P and Mn in relatively thin layers were also observed in sediments populated by
349 cable bacteria in the Gulf of Finland and Lake Grevelingen (Fig. 8B and C). In the latter system, the
350 layers enriched in Fe, P, Mn broadened upon recolonization by macrofauna (Fig. 8D).

351 **3. DISCUSSION**

352 **4.1. Metabolic Activity of Cable Bacteria**

353 Cable bacteria in our incubation experiment demonstrated a rapid growth, since their
354 abundance greatly increased after 5 days, and reached its peak at day 26 (Fig 1C). Such high
355 abundances are similar to those observed in previous experiments, in which FeS-rich marine
356 sediments from Aarhus Bay and Lake Grevelingen were incubated (Schauer et al. 2014; Burdorf et al.
357 2018). The activity of cable bacteria exerted a strong impact on the pore water depth profiles of pH,
358 O_2 , and $\sum H_2S$, as evident from the development of a pH maximum near the sediment-water interface,
359 the strong pore water acidification in the deeper sediment and the development of a suboxic zone (Fig.
360 2A). These pore water depth profiles resemble the distinct biogeochemical fingerprint typical for
361 active cable bacteria, as observed in previous laboratory incubation experiments (Risgaard-Petersen et
362 al. 2012; Malkin et al. 2014; Schauer et al. 2014; Vasquez-Cardenas et al. 2015; Rao et al. 2016;

363 Burdorf et al. 2018). The widening of the suboxic zone with time (Fig 3) is a consequence of the
364 downward expansion of the cable bacteria filament network (Schauer et al. 2014; Vasquez-Cardenas
365 et al. 2015).

366 The EP depth profiles demonstrated that long-distance electron transport by cable bacteria
367 was already active 5 days after the start of the experiment, as indicated by the increase of EP at depth
368 to 0.4 mV). With time, the EP signal increased to higher values and also accumulated over a thicker
369 depth horizon (Fig. 2B), indicating that cable bacteria activity both increased and extended to deeper
370 sediment depth, which is also a consequence of the downward expansion of cable bacteria filaments.
371 The EP reached a maximum after 18 days (1.3 mV; Fig. 2B) in concert with the highest current
372 density of $\sim 130 \text{ mmol e}^- \text{ m}^{-2} \text{ d}^{-1}$ (Fig 4B). This maximum EP value and current density are similar in
373 magnitude to those found in sediment incubations with seawater with a similar salinity (Damgaard et
374 al. 2014). From day 18 onwards the EP and current density flux gradually decreased with time to 13
375 $\text{mmol e}^- \text{ m}^{-2} \text{ d}^{-1}$ after 207 days (Fig. 4B), which implies a decrease in the metabolic activity of cable
376 bacteria. The suboxic zone persisted long after the current density had decreased (Fig. 3).

377 To summarise, the metabolic activity of cable bacteria was likely highest between day 18 and
378 day 26 based on the cable bacterial abundances, the extent of acidification of the pore water and the
379 current density (Fig 1C; Fig 2A and Fig 4B).

380 **4.2. Organic Matter Degradation**

381 Ammonium fluxes are assumed to reflect rates of anaerobic degradation of organic matter (Fig.
382 5), and the observed decline during the experiment coincides with the decrease in activity of cable
383 bacteria based on the EP profiles and current density (Fig. 2B; Fig 4B). This suggests that the
384 availability of easily degradable organic matter plays a role in sustaining the metabolic activity of
385 cable bacteria, most likely by controlling the rate of SO_4^{2-} reduction (Nielsen and Risgaard-Petersen
386 2015).

387 Rates of SO_4^{2-} reduction estimated from the linear gradient of the decrease in pore water SO_4^{2-} in
388 the surface sediment with depth indeed also showed a decline during the experiment. We note,

389 however, that a direct measurement of SO_4^{2-} reduction rates (Fossing and Jørgensen 1989; Kallmeyer
390 et al. 2004) would provide a better indicator, because SO_4^{2-} estimated from pore water profiles are in
391 general lower than rates estimated from tracer experiments (Hermans et al. 2019a; Sandfeld et al.
392 2020). Another cause for a slight underestimation of our SO_4^{2-} reduction rates, is due to the effect of
393 the electric field imposed by cable bacteria, which is not taken into account in Fick's law. **Solutes can**
394 **also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the**
395 **relatively low strength of the electric field in the cores ($<0.073\text{V m}^{-1}$ at day 18; as estimated from Fig.**
396 **2B), including the contribution of ionic drift to the sulphate flux would lead to SO_4^{2-} reduction rates**
397 **that are at most 10-20% higher.**

398 The metabolic activity of cable bacteria can lead to the production of SO_4^{2-} in the suboxic zone
399 via anodic sulphide oxidation (Risgaard-Petersen et al. 2012; Rao et al. 2016). We suspect that this
400 also explains the lack of change in pore water SO_4^{2-} with depth in the upper 2 cm of the sediment in
401 our experiment between 12 and 40 days (Fig. 5). Despite relatively high SO_4^{2-} reduction rates ranging
402 from 5.4 to 17.6 $\text{mmol m}^{-2} \text{d}^{-1}$ (Fig. 6B; Table S3), pore water concentrations of $\Sigma\text{H}_2\text{S}$ remained very
403 low throughout the experiment (Fig. 2A). This is likely due to the direct consumption of $\Sigma\text{H}_2\text{S}$
404 through the activity of cable bacteria, preventing $\Sigma\text{H}_2\text{S}$ from accumulating in the pore water, or
405 alternatively, precipitation of FeS by dissolved Fe^{2+} released from the dissolution of siderite.

406 Laboratory experiments have shown that S-oxidation by cable bacteria can play a dominant role
407 in the O_2 uptake of coastal sediments (Nielsen et al. 2010; Schauer et al. 2014; Nielsen and Risgaard-
408 Petersen 2015), and model analysis predicts up to 93% of the total O_2 uptake (Meysman et al. 2015).
409 When we plot diffusive uptake of O_2 against the current density (i.e. upward flux of electrons towards
410 the oxic zone), a linear relationship – with some scatter - emerges for days 12 to 621 (Fig. 9).
411 However, the data points for day 0 and 5 during the initial stages of our experiment do not follow this
412 linear relationship. We explain these findings as follows: At day 0, the cable bacteria were not active
413 yet and other processes, such as aerobic respiration and oxidation of NH_4^+ and other solutes (Table 2)
414 and solids (FeS) dominated the consumption of O_2 . At day 5 and 12, the activity of cable bacteria and
415 the oxidation of reduced products from anaerobic degradation of organic matter both contributed to

416 consumption of O₂. From day 12 onwards, both the O₂ consumption and electron flux follow a
417 downward decrease with time (Fig. 9). If cable bacteria would account for all of the O₂ consumption,
418 a ratio between the diffusive uptake of O₂ and the current density of 1:4 is expected (Fig. 1A; Nielsen
419 et al. 2010). We find that from day 12 onwards, most data points plot rather close to the line for this
420 1:4 relationship (Fig. 9), suggesting that cathodic O₂ reduction by cable bacteria is responsible for
421 nearly all O₂ consumption in the sediment (in line with the model results of Meysman et al. 2015).
422 This however poses a problem for the nitrogen budget, because our data indicate complete removal of
423 the NH₄⁺ that diffuses upward into the oxic zone (Fig. 6A), and based on the solute fluxes, no escape
424 to the overlying water (see section 2.4). This implies substantial O₂ consumption due to nitrification
425 (Table 2). These findings can be explained, however, if we assume that at least part of the NO₃⁻ that is
426 being formed near the sediment-water interface is also used for the metabolic activity of cable
427 bacteria. It has been shown that cable bacteria can couple the oxidation of ΣH₂S to NO₃⁻ in the
428 absence of O₂ (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments
429 where O₂ is present in concert with NO₃⁻ near the sediment-water interface. However, we cannot
430 exclude release of NO₃⁻ to the water column or denitrification by other bacteria in the sediment.
431 Another explanation is that cable bacteria might consume O₂ directly above the sediment-water
432 interface, as recently has been proposed by Burdorf et al. (2018). Lastly, the current density might be
433 slightly overestimated, since it ignores other sources that can create an electric potential, such as the
434 diffusion potential (Revil et al. 2012; Nielsen and Risgaard-Petersen 2015).

435 **4.3. Impact of Cable Bacteria on Fe, Mn and S Cycling**

436 The activity of cable bacteria had a strong impact on the biogeochemistry of the surface sediment
437 in our experiment (Fig. 7). Cable bacteria activity induced an intense acidification of the pore water in
438 the suboxic zone (Fig 2A), which led to the dissolution of Fe and Mn minerals in deeper sediment
439 layers, as can be inferred from the sharp maxima in dissolved Fe²⁺ and Mn²⁺ in the pore water
440 reaching concentrations of up to ~1700 and ~80 μM, respectively (Fig. 5). The twenty-fold higher
441 dissolved Fe²⁺ concentrations with respect to pore water Mn²⁺ can be attributed to the relatively higher
442 availability of FeS and siderite compared to the availability of Mn carbonates in the sediment that was

443 used for incubation (Lenstra et al. 2020). The peaks in dissolved Fe^{2+} and Mn^{2+} in the pore water
444 broadened over time spanning a depth of $>5\text{cm}$ (Fig. 5; Fig. S6; Fig. S7).

445 The upward diffusive flux of dissolved Fe^{2+} and Mn^{2+} was highest after 12 days, reaching values
446 of up to 3.16 and $0.16 \text{ mmol m}^{-2} \text{ d}^{-1}$ respectively. Fluxes subsequently gradually decreased with time
447 (Fig. 6C and D). The continuous upward diffusion of dissolved Fe^{2+} and Mn^{2+} led to enrichments of
448 poorly crystalline Fe and Mn oxides in the surface sediment (Fig. 7). Despite high upward fluxes of
449 dissolved Fe^{2+} and Mn^{2+} towards the sediment-water interface, our solute flux incubations indicate
450 there was little escape of Fe^{2+} and Mn^{2+} to the overlying water (see section 2.4). This implies that all
451 Fe^{2+} and Mn^{2+} that diffused upward was precipitated as Fe and Mn oxides upon contact with O_2 or
452 NO_3^- (Buresh and Moraghan 1976; Kuz'minskii et al. 1994; Straub et al. 1996). Little or no escape of
453 dissolved Fe^{2+} from the sediment into the overlying water, was suggested previously for a field site
454 with active cable bacteria based on diffusive flux calculations (Lake Grevelingen; Sulu-Gambari et al.
455 2016a) and was determined in flux incubations of cores during a laboratory experiment with cable
456 bacteria (Rao et al. 2016).

457 At the start of the experiment, the sedimentary FeS content was ($\sim 25 \mu\text{mol g}^{-1}$), which is not
458 unusual for coastal sediments on the north-western Black Sea margin (Wijsman et al. 2001), but is
459 low compared to sediments in eutrophic coastal systems (e.g. Morgan et al. 2012; Kraal et al. 2013;
460 Hermans et al. 2019a). The solid-phase depth profiles reveal a gradual removal of the FeS in the
461 surface sediment in our experiment over time (Fig. 7). At the end of our experiment (621 days), there
462 was no longer any FeS within the top 1.5 cm of the sediment. While approximately 90 mmol m^{-2} of
463 FeS was removed from the surface sediment within the first 5 days, a total of $\sim 240 \text{ mmol m}^{-2}$ was
464 removed after 621 days (Fig. 10; Table 3). Likely, part of the FeS that was removed from the surface
465 sediment within the first 5 days was removed through oxidation upon contact with O_2 , rather than the
466 metabolic activity of cable bacteria itself. The pore water acidification associated with cable bacteria
467 activity led to a strong loss of siderite within the top 2 cm of the sediment, with a total removal of
468 $\sim 560 \text{ mmol m}^{-2}$ during the experiment (Fig. 7; Fig. 10; Table 3). The depletion of sedimentary FeS

469 and siderite was directly proportional to the formation of Fe oxides near the sediment-water interface
470 (Fig. 10), and accounted for 30% and 70% of the Fe oxides, respectively (Table 3).

471 With these data we cannot accurately determine the role of FeS versus SO_4^{2-} reduction in
472 supplying the $\Sigma\text{H}_2\text{S}$ sustaining the activity of cable bacteria throughout the experiment. This is
473 primarily related to the variability between cores, and for this type of calculation, the low temporal
474 resolution of sampling. However, we can make an estimation of the relative role of SO_4^{2-} reduction
475 and FeS dissolution in $\Sigma\text{H}_2\text{S}$ production, based on the pore water profiles of SO_4^{2-} and dissolved Fe^{2+} ,
476 and the solid-phase mass balance of FeS and siderite (Fig. 6B and C; Table 4). This estimation shows
477 that SO_4^{2-} was mainly responsible for $\Sigma\text{H}_2\text{S}$ production, accounting for 85-99% (Table 4), and thus
478 that the dissolution of FeS only played a minor role in providing $\Sigma\text{H}_2\text{S}$.

479 **4.4. Impact of Cable Bacteria on Ca, P and Si Cycling**

480 Cable bacteria activity is known to lead to dissolution of Ca carbonates, because of the strong
481 acidification of the pore water (Risgaard-Petersen et al. 2012; Rao et al. 2016). We indeed find similar
482 maxima in pore water Ca^{2+} during the experiment (Fig. 5) and a high upward flux of Ca^{2+} (up to ~18
483 $\text{mmol m}^{-2} \text{d}^{-1}$; Fig. 6E; Fig. S8) of which a substantial fraction (up to ~55%) escapes to the overlying
484 water (Fig. S10; Table S4), which is consistent with a previous incubation experiment Rao et al.
485 (2016).

486 Pore water depth profiles of HPO_4^{2-} reveal a production at depth and removal of all upward
487 diffusing HPO_4^{2-} within the first 1-3 cm of the surface sediment (Fig. 5). A major proportion of this
488 HPO_4^{2-} is bound to Fe oxides (Fig. 7). Given that a large proportion of the Fe oxides in our sediment
489 cores derive from the dissolution of siderite, this suggests that the buffer mechanism that delays the
490 benthic release of HPO_4^{2-} through retention of P associated with newly formed Fe oxides (Sulu-
491 Gambari et al. 2016b), might also be active in systems that are relatively poor in sedimentary FeS.

492 The shape of the pore water HPO_4^{2-} profiles suggests that some of the HPO_4^{2-} is removed below
493 the zone where Fe and Mn oxides are present (Fig. 5; Fig. 7). A possible explanation could be the
494 formation of vivianite, an Fe(II) phosphate mineral. Vivianite formation in sediments typically occurs

495 when pore water levels of Fe^{2+} and HPO_4^{2-} are high and concentrations of $\Sigma\text{H}_2\text{S}$ are low (Nriagu
496 1972), as observed in our study. In our experiment, free $\Sigma\text{H}_2\text{S}$ does not accumulate in the pore water,
497 which we attribute to removal through the activity of cable bacteria and FeS formation at depth (Fig.
498 2A; Fig. 7). Hence, cable bacteria may create a geochemical niche that allows the formation of
499 vivianite in the suboxic zone. Further work with sediments with higher P concentrations would be
500 needed to assess this with direct measurement techniques, such as X-ray spectroscopy (Egger et al.
501 2015; Kraal et al. 2017; Sulu-Gambari et al. 2018). Other sediment P pools, i.e. organic, authigenic
502 and detrital P remained constant over time, indicating that the P contents determined for discrete
503 sediment slices using sequential extractions were not affected by pore water acidification as a result of
504 cable bacteria activity (Table S5).

505 Pore water H_4SiO_4 profiles show a typical increase with depth as observed upon dissolution of
506 biogenic silica in marine sediments (Aller 2014). Fluxes of H_4SiO_4 towards the sediment-water
507 interface range up to $\sim 2.8 \text{ mmol m}^{-2} \text{ d}^{-1}$ and gradually decreased with time (Fig. 6F; Fig. S9). The
508 results of the solute flux incubations indicate that most of this H_4SiO_4 escaped to the overlying water
509 (ranging from 28 to 92%; Table S4; Fig. S10). The decline in the benthic release flux of H_4SiO_4
510 contrasts with results of a previous incubation experiment by Rao et al. (2016) with similar pore water
511 concentrations of H_4SiO_4 reaching values up to $\sim 1 \text{ mM}$. In their study, the flux remained constant
512 over time, possibly because of differences in the amount of biogenic Si in the sediment. The solid-
513 phase metal oxide bound Si pool in the surface sediment increased directly proportional to the
514 formation of Fe oxides throughout the experiment (Fig. 7). Silica is known to adsorb to Fe oxides
515 (Sigg and Stumm 1981; Davis et al. 2002). Hence, the results suggest that the Fe oxides formed
516 through the activity of cable bacteria captured some of the upward diffusing H_4SiO_4 .

517 **4.5. Sediment Marker for Cable Bacteria Activity**

518 Visual observations of core photographs reveal the gradual development of an orange layer (oxic
519 zone) up to 9 mm thick, overlying a grey layer (suboxic zone) and a black layer (sulphidic zone)
520 during the experiment (Fig. S11). This colour zonation is typical for sediments that have been
521 geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and

522 Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016a). High-resolution
523 elemental maps of our sediments reveal the development of a ~0.3 mm thin subsurface layer highly
524 enriched in Fe oxides and associated P, 47 days after the start of the incubation (Fig. 8A). While the
525 Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the
526 sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P
527 is strongly correlated with Ca. Below, we describe the formation of this layer in more detail and
528 explain why such subsurface enrichments, detected with μ XRF, may act as an additional sediment
529 marker for present or recent cable bacteria activity, also in cases where visual observations are not
530 conclusive.

531 During the experiment, O_2 penetration varied within a narrow range and was initially fixed
532 between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth
533 of 2 mm (Fig. 8A). Such a range in O_2 penetration is in accordance with observations in coastal
534 sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by rapid
535 oxidation of upward diffusing Fe^{2+} upon contact with O_2 (and possibly NO_3^- ; Fig. 6C). Directly, above
536 the Fe oxide layer a broader ~0.8 mm thick Mn oxide layer was observed (Fig. 8A). This contrast in
537 zonation between Fe and Mn is likely due to the slower oxidation kinetics of Mn^{2+} compared to Fe^{2+}
538 (Burdige 1993; Luther 2010; Learman et al. 2011).

539 While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P
540 close to the sediment-water interface (Fig. 8A). In this layer, P is strongly correlated with Ca. This
541 layer likely consists of carbonate fluorapatite (CFA), a Ca-P mineral, which is typically formed in
542 marine sediments (Van Cappellen and Berner 1988; Ruttenberg and Berner 1993). Possibly, the high
543 pore water pH near the sediment-water interface (resulting from cathodic O_2 reduction by cable
544 bacteria; Fig. 2A), promotes apatite formation (Bellier et al. 2006), and the elevated Ca^{2+}
545 concentrations (Fig. 5) created a biogeochemical niche for the formation of CFA.

546 Such focusing of Fe, Mn, P and associated elements within a thin subsurface layer, as a
547 consequence of cable bacteria activity, also occurs in the field. This was demonstrated by Hermans et

548 al. (Submitted) in a study of a coastal site in the Gulf of Finland where cable bacteria were recently
549 active. Here, μ XRF mapping of resin embedded sediments revealed strong focusing of Fe oxides,
550 Mn(II) phosphates and Fe bound P within a 3 mm thick layer near the sediment-water interface (Fig.
551 8B). A re-assessment of the μ XRF data of Sulu-Gambari et al. (2016a; 2018) of surface sediments
552 with active cable bacteria from seasonally hypoxic marine Lake Grevelingen in January also revealed
553 similar subsurface enrichments in Fe, Mn and P (Fig. 8C). Importantly, no visual signals for cable
554 bacteria based on the colour pattern of the sediment were observed at the time.

555 Macrofaunal activity likely counteracts or prevents strong focusing of Fe oxides and associated P
556 within such a thin subsurface layer at field sites. Bioturbation, i.e. mixing of the sediment, typically
557 leads to oxidation from the sediment surface downwards (Norkko et al. 2012). Bioirrigation can
558 efficiently pump O_2 into the pore water and thereby enhance the oxidation of dissolved Fe^{2+}
559 (Kristensen et al. 2012; Norkko et al. 2012), but is not expected to lead to such a sharp oxidation front
560 (Norkko et al. 2012; Hermans et al. 2019a). This is also evident from high-resolution elemental maps
561 of the surface sediment from Lake Grevelingen in May, which shows the disappearance of the thin
562 layer highly enriched in Fe and P formed by cable bacteria in January as a consequence of
563 macrofaunal activity in May (Fig. 8D; Seitaj et al. 2015; Sulu-Gambari et al. 2016b).

564 We conclude that the focusing of Fe, Mn and associated P within a thin layer below the sediment-
565 water interface is likely a consistent feature in sediments populated by active cable bacteria and may
566 act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory
567 experiments and at field sites, also in cases where visual observations are not conclusive. Focussing of
568 Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and
569 can also occur in the absence of cable bacteria. However, the upward fluxes of Fe^{2+} and Mn^{2+} in
570 sediments populated by cable bacteria, are higher due to active dissolution of Fe and Mn minerals at
571 depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame
572 following an environmental perturbation (such as a transition to oxic bottom waters after a period of
573 anoxia or mixing of the sediment), more Fe^{2+} and Mn^{2+} can oxidise upon contact with O_2 near the
574 sediment-water interface and stronger enrichments of Fe and Mn minerals will be observed. Hence,

575 focusing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in
576 sediments populated by active cable bacteria compared to sediments where no cable bacteria are
577 active under such conditions. Macrofaunal activity within natural environments likely counteracts or
578 prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer. When
579 using standard techniques for sediment sampling (i.e. core slicing and chemical analysis of these
580 slices), these layers may be missed due to the relatively coarse depth resolution. Hence, μ XRF
581 mapping of epoxy embedded sediment is recommended.

582 **4.6. Cable Bacteria Activity at the Field Site**

583 We can only speculate about the possible *in-situ* relevance of cable bacteria at the coastal site
584 in the western Black Sea where the sediment for our incubation was collected. **At this site, which is in**
585 **a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m⁻²)**
586 **and polychaetes (up to ~1700 ind. m⁻²)** were observed at the time of sampling (Lenstra et al. 2019).
587 Macrofauna can inhibit the activity of cable bacteria through bioturbation by physically cutting and
588 damaging the filaments, rendering them unable to transport electrons (Malkin et al. 2014). Recent
589 work has shown, however, that in some cases, cable bacterial communities can also thrive in
590 sediments with macrofauna (Burdorf et al. 2017; Malkin et al. 2017; Aller et al. 2019). In a study of
591 bivalve reefs, cable bacteria were found to efficiently remove highly toxic $\Sigma\text{H}_2\text{S}$, which is beneficial
592 for bivalves (Malkin et al. 2017). Cable bacteria can also be abundant in bioturbated deposits, when
593 associated with stable subdomains of the bioturbated zone, such as worm tubes (Aller et al. 2019). In
594 such settings, a more complex precipitation pattern, e.g. along tube linings is observed (Aller et al.
595 2019), than described here for laboratory experiments with defaunated sediments and field sediments
596 with an impoverished macrofaunal population (Fig. 8A). Further field studies are required to assess
597 the role of cable bacteria at our field site, preferably including an assessment of the burrow structures.

598 **Conclusions**

599 The results of our laboratory incubation (with a total duration of 621 days) show that cable
600 bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The
601 strong acidity of the pore water associated with the activity of cable bacteria, which was monitored

602 using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and
603 formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental
604 results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a
605 source of Fe that can form an Fe oxide-enriched surface layer. Both FeS and SO_4^{2-} reduction provided
606 the $\sum\text{H}_2\text{S}$ required by cable bacteria to sustain their activity. Pore water $\sum\text{H}_2\text{S}$ was always low (<5
607 μM). Using μXRF mapping of epoxy embedded sediment, we show that the activity of cable bacteria
608 led to the development of a thin subsurface sediment layer (0.3 mm) that was highly enriched in Fe
609 and P. The position of this layer in the sediment was directly proportional to the O_2 penetration depth
610 during the experiment. We show that a similar layer highly enriched in Fe and P was also formed in
611 sediments of field locations populated by cable bacteria (i.e. marine Lake Grevelingen and the
612 brackish Gulf of Finland). We suggest that such layers, which are not necessarily visible by eye, may
613 be used as a marker of cable bacteria activity in sediments with low macrofaunal activity.

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626 **Code availability.** All data, if not directly available from the tables and supplementary information,
627 will be made available in the PANGAEA database. In the meantime data are available upon request to
628 the main author.

629 **Author contributions.** MH and CPS designed the experiment. MH carried out the experiment and
630 analysis. All authors interpreted the data. MH and CPS wrote the paper with comments provided by
631 NRP and FJRM.

632 **Competing interests.** The authors declare that they have no conflict of interest.

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909

910 **TABLES AND FIGURES**

911 **Table 1.** Key site characteristics: latitude, longitude, water depth, bottom water O₂ concentration, *in-situ* O₂ uptake, *in-situ*
 912 O₂ penetration depth in the sediment, porosity and salinity. These data were retrieved from Lenstra et al. (2019). Our study
 913 site is station 9 in Lenstra et al. (2019).

Black Sea (Station 9)		Unit
Latitude	44°34.9'	N
Longitude	29°11.4'	E
Water depth	27	m
Bottom water O ₂	92	μM
O ₂ uptake	25.8 ± 1.77	mmol m ⁻² d ⁻¹
O ₂ penetration depth	2.25	mm
Porosity	0.86	-
Salinity	17.881	-
Avg. organic carbon content (0-0.5 cm)	1.8%	

914
 915 **Table 2.** Mass balance of O₂ consumption. The diffusive uptake of O₂ as calculated from the O₂ depth profiles (column 1)
 916 was compared to the potential O₂ demand from the oxidation of NH₄⁺, Fe²⁺ and Mn²⁺ (column 2-4). The O₂ consumption of
 917 the oxidation of NH₄⁺, Fe²⁺ and Mn²⁺ was determined based on the stoichiometry of NH₄⁺, Fe²⁺ and Mn²⁺ oxidation with O₂
 918 as described in Reed et al. (2011). The oxidation of dissolved Fe²⁺ and Mn²⁺ only played a minor role in the total O₂
 919 consumption during the experiment, contributing only 0.9 to 3.8% and 0.1 to 0.4%, respectively.

	O₂ [mmol m⁻² d⁻¹]	Potential O₂ Demand			e⁻ [mmol m⁻² d⁻¹]
		NH₄⁺ [mmol m⁻² d⁻¹]	Fe²⁺ [mmol m⁻² d⁻¹]	Mn²⁺ [mmol m⁻² d⁻¹]	
Day 5	-23.35	9.42	0.21	0.05	82.68
Day 12	-23.24	8.46	0.89	0.09	111.94
Day 18	-21.10	8.04	0.70	0.08	127.97
Day 26	-23.00	7.58	0.63	0.07	97.55
Day 33	-22.80	5.06	0.62	0.05	84.16
Day 40	-19.60	4.88	0.60	0.06	76.31
Day 207	-6.90	3.52	0.08	0.01	13.10
Day 621	-3.25	N/A	0.03	0.01	9.47

920

921 **Table 3.** Mass balance of Fe. Time-series of the depth integrated (0-5 cm) increase in Fe oxides and the depth integrated (0-5
 922 cm) depletion of FeS and FeCO₃ (siderite) in mmol m⁻². All values are reported in mmol Fe m⁻². Negative values represent a
 923 decrease, whereas positive values indicate an increase in the mineral pools.

	$\Delta\text{Fe oxides}$ [mmol m ⁻²]	ΔFeS [mmol m ⁻²]	ΔFeCO_3 [mmol m ⁻²]
Day 5	120	-90	-42
Day 12	170	-90	-126
Day 18	189	-105	-92
Day 26	276	-174	-99
Day 33	315	-176	-109
Day 40	412	-223	-200
Day 207	523	-236	-341
Day 621	874	-242	-566

924

925 **Table 4.** Sources of $\Sigma\text{H}_2\text{S}$ calculated from the reduction of SO₄²⁻ and the dissolution of FeS. The numbers are presented
 926 either as mmol m⁻² d⁻¹ or as the relative percentage of the $\Sigma\text{H}_2\text{S}$ production. The amount of S from the dissolution of FeS
 927 was estimated from the upward diffusive flux of Fe²⁺ (Fig. 6C) and the relative fraction of FeS (FeS/FeS+siderite) based on
 928 the mass balance calculations (Table 3).

	S from SO ₄ ²⁻ reduction [mmol m ⁻² d ⁻¹]	S from FeS dissolution [mmol m ⁻² d ⁻¹]	S from SO ₄ ²⁻ reduction [%]	S from FeS dissolution [%]
Day 5	10.49	0.56	95%	5%
Day 12	17.60	1.48	92%	8%
Day 18	8.87	1.50	86%	14%
Day 26	11.15	1.61	87%	13%
Day 33	8.54	1.52	85%	15%
Day 40	7.57	1.25	86%	14%
Day 207	10.38	0.13	99%	1%
Day 621	5.36	0.03	99%	1%

929

930 **Fig. 1. (A)** Geochemical pore water fingerprint typical for cable bacteria activity. This fingerprint is defined by a distinct pH
931 profile (light grey line) and a sub-oxic zone that is devoid of O₂ (red line) and H₂S (blue line). The cable bacteria filaments
932 are depicted in yellow. On the background, the sediment core photograph, taken 278 days after the start of the experiment,
933 shows a distinct colour zonation where (1) the oxic zone displays an orange colour (2) the suboxic zone has a grey colour
934 and (3) the sulphidic zone has a black colour. The scale bar denotes a distance of 6 cm, with 0.5 cm intervals. **(B)**
935 Bathymetric map of the Black Sea. The purple star indicates the location of our study site (44°34.93'N, 29°11.38'E), which
936 was sampled with R/V *Pelagia* in September 2015. Further details are provided in Lenstra et al. (2019). **(C)** Volumetric
937 density of cable bacteria [m cm⁻³] in the top 2.5 cm of the sediment, for *in-situ* as well as for three time points during the
938 incubation experiment **(D)** SEM image of a cable bacteria filament that was extracted from the surface sediment after 40
939 days.

940

941 **Fig. 2. (A)** Time-series of the pore water pH (black), O₂ (red) and ΣH₂S (blue) signatures of the incubated sediment. **(B)**
942 Development of the EP depth profile in the incubated sediment over time. The dashed-line at 0 mm depth represents the
943 sediment-water interface. The blue boxes indicate the overlying water, whereas the underlying light grey boxes represent the
944 sediment. The EP depth profiles represent an average of 3 replicate measurements. The error bars indicate the minimum and
945 maximum EP values that were observed. The orange depth profiles represent duplicate measurement performed on a
946 different core.

947

948 **Fig. 3.** Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone
949 (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two
950 different cores.

951

952 **Fig. 4.** Time-series of the **(A)** diffusive O₂ uptake in mmol O₂ m⁻² d⁻¹ and **(B)** current density as a consequence of long-
953 distance electron transport (e⁻) in mmol e⁻ m⁻² d⁻¹ in the sediment incubation.

954

955 **Fig. 5.** Time-series of pore water depth profiles of NH₄⁺ (orange), SO₄²⁻ (purple), Fe²⁺ (red), Mn²⁺ (green), Ca²⁺ (grey),
956 HPO₄²⁻ (blue) and H₄SiO₄ (yellow). The control core was sampled at day 621.

957

958 **Fig. 6.** Time-series of diffusive fluxes calculated from the linear gradient of the pore water profiles of **(A)** NH₄⁺, **(B)** SO₄²⁻,
959 **(C)** Fe²⁺, **(D)** Mn²⁺, **(E)** Ca²⁺ and **(F)** H₄SiO₄ in mmol m⁻² d⁻¹ towards the oxic zone of the sediment, based on the linear pore
960 water gradients (Section 1.6; Fig. S3-S4). Here, a positive value indicates an upward flux, whereas a negative value
961 represents a downward flux. N/A = not available. The control core was sampled at day 621.

962

963 **Fig. 7.** Time-series of solid-phase depth profiles of Fe oxides (red), FeS (black), siderite (grey), Mn oxides (green), metal
964 bound P (blue) and metal oxide bound Si (yellow).

965

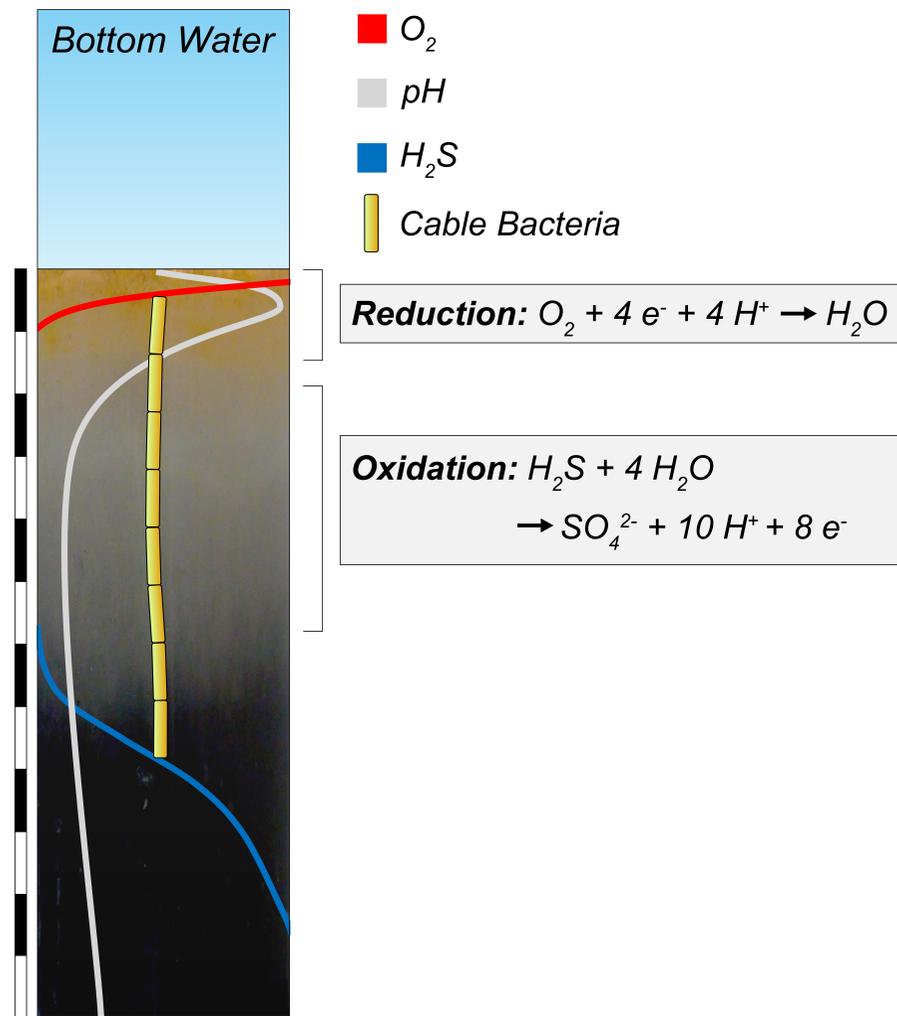
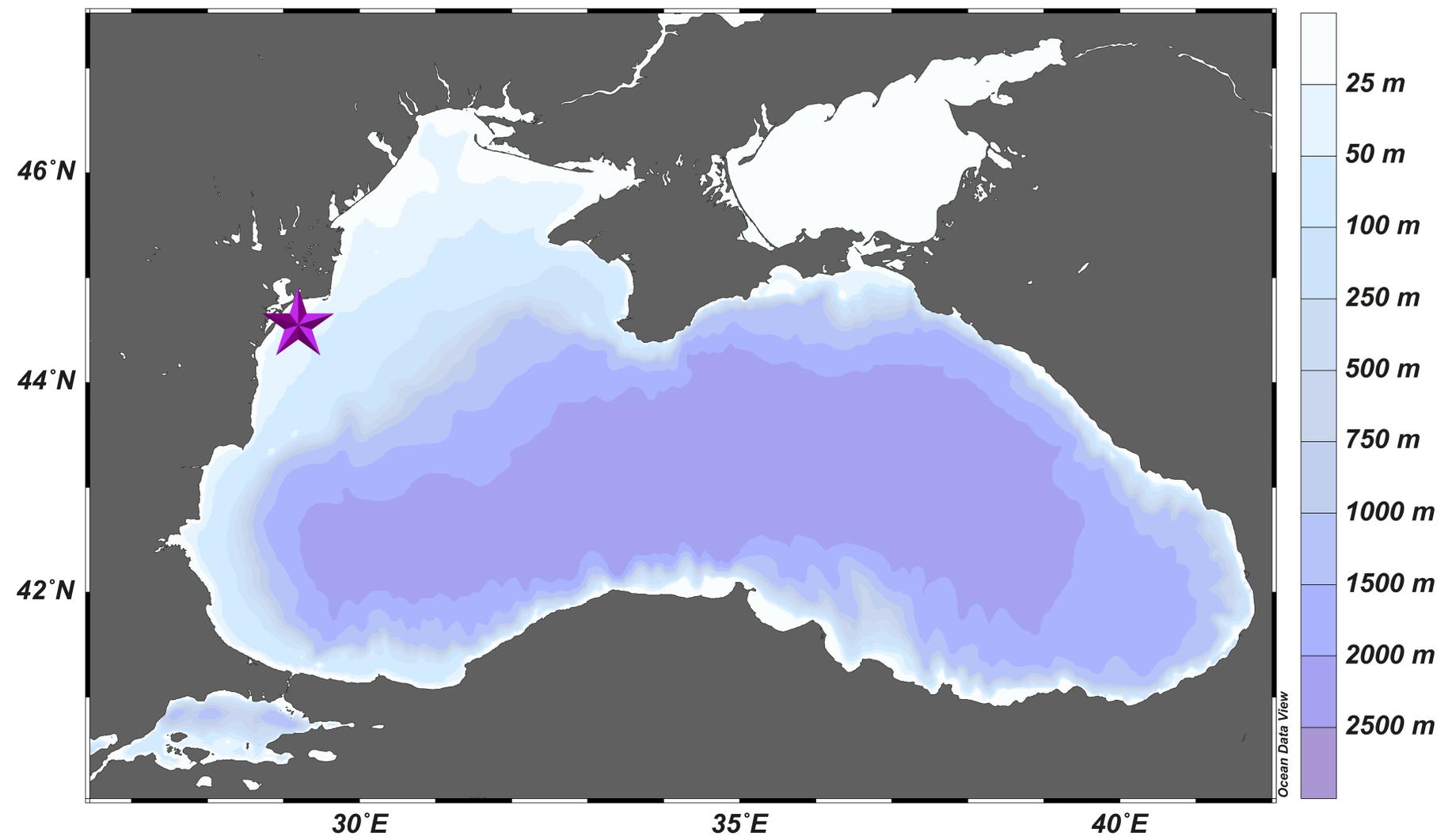
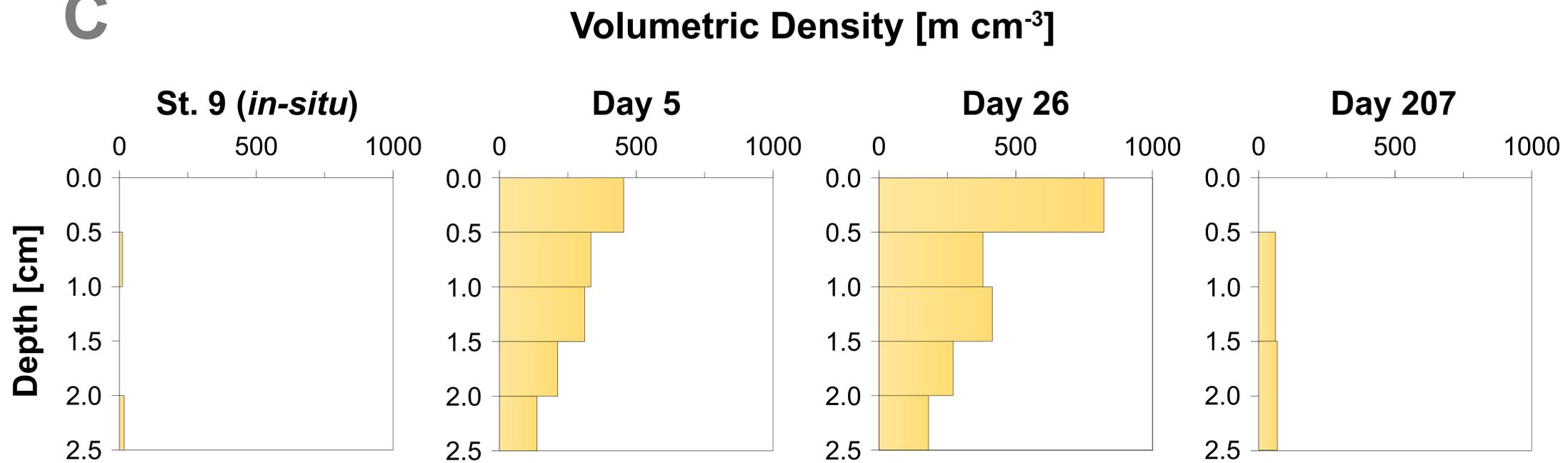
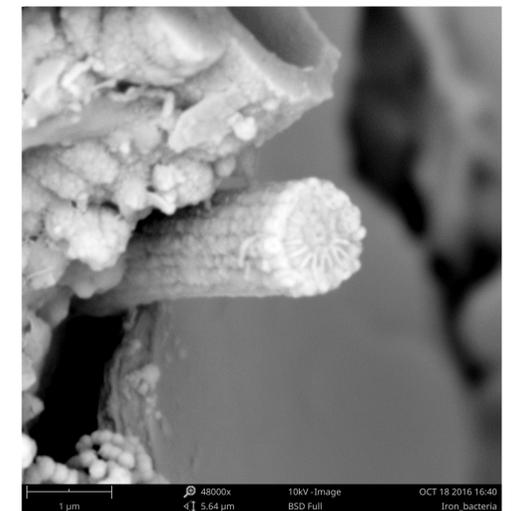
966 **Fig. 8.** High-resolution elemental maps of Fe (red), Mn (green), P (blue) and Ca (white) of surface sediments. These maps
967 are shown in true vertical orientation and the colours accentuate the relative count intensities adjusted for brightness and
968 contrast to highlight the features in the sediment. The tick marks represent 1 mm intervals. μ XRF maps of the surface
969 sediment (A) from the incubation experiment, (B) from the Gulf of Finland at site GOF5 in June (Hermans et al. Submitted),
970 (C) from Lake Grevelingen in January (when cable bacteria become active) and (D) from Lake Grevelingen in May
971 (showing the effects of bioturbation as described in Seitaj et al. (2015)).

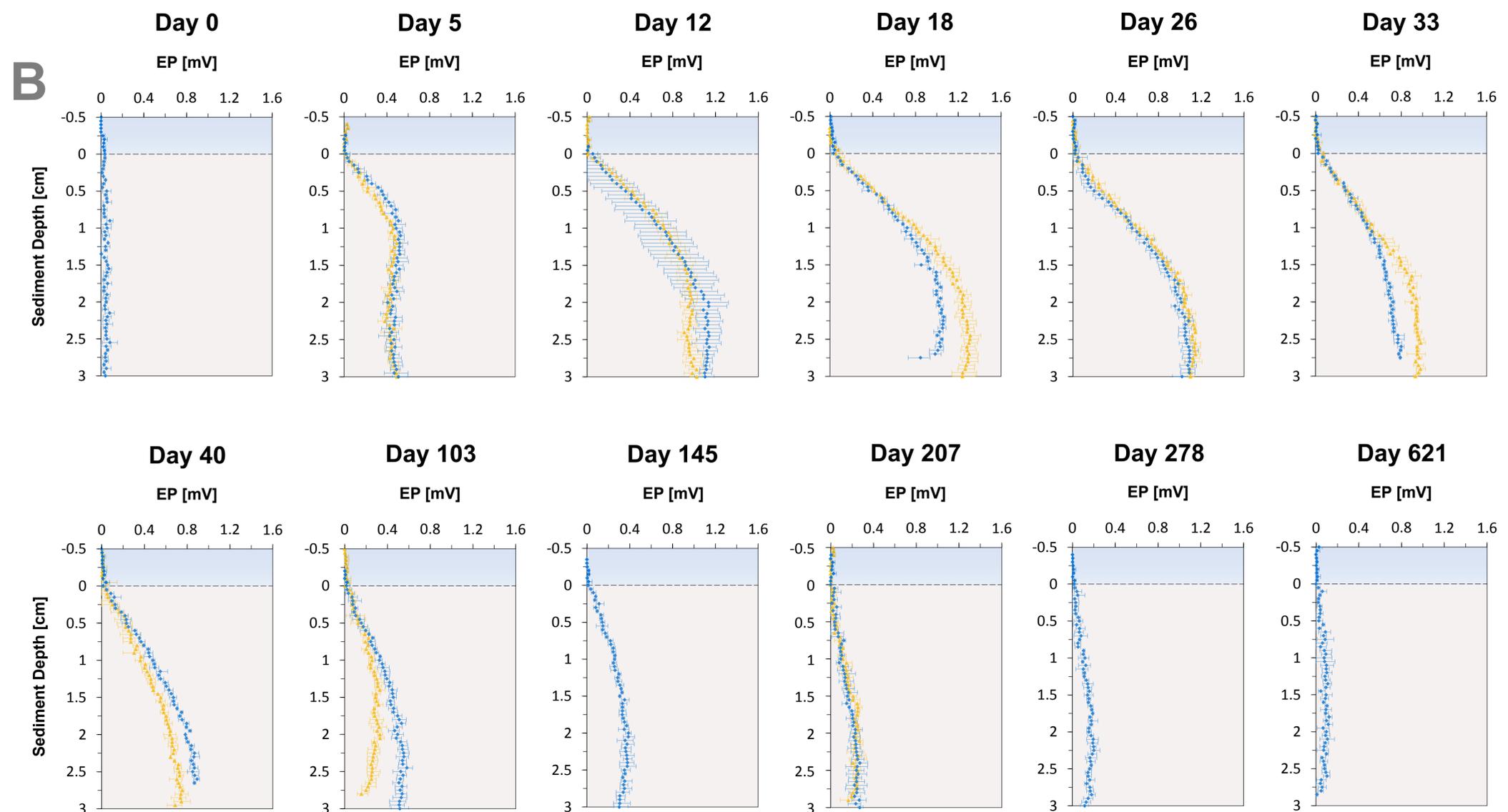
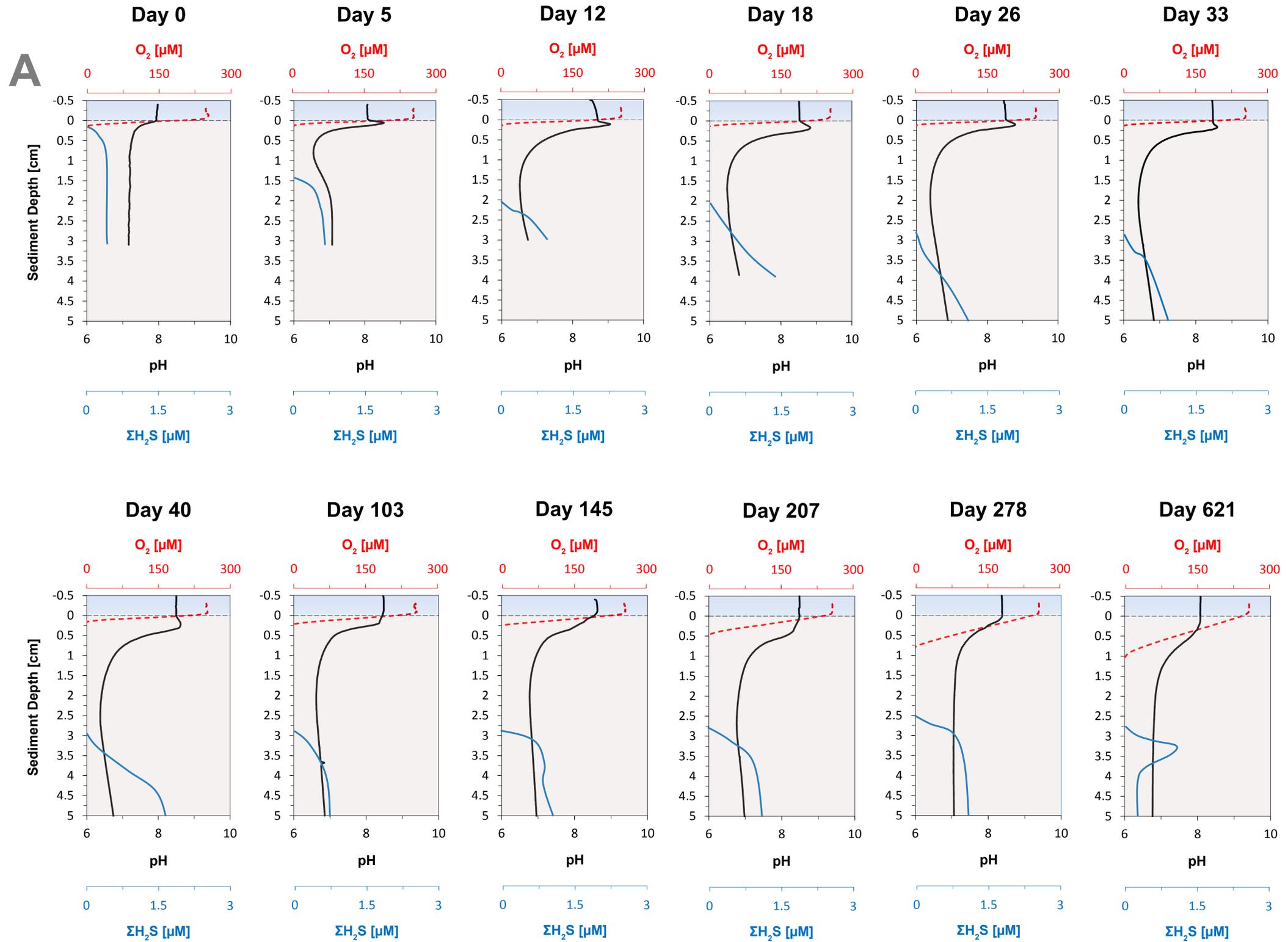
972

973 **Fig. 9.** The relationship between the diffusive uptake of O_2 ($mmol O_2 m^{-2} d^{-1}$) and the current density of long-distance
974 electron transport ($mmol e^{-} m^{-2} d^{-1}$). Red triangles are data for days 0 and 5. Green diamonds are data for all other time
975 points. The blue line represents the expected correlation between the cathodic O_2 consumption rate and the current density
976 assuming a 1:4 ratio (Nielsen et al. 2010). Here, a positive value indicates an upward flux, whereas a negative value
977 represents a downward flux.

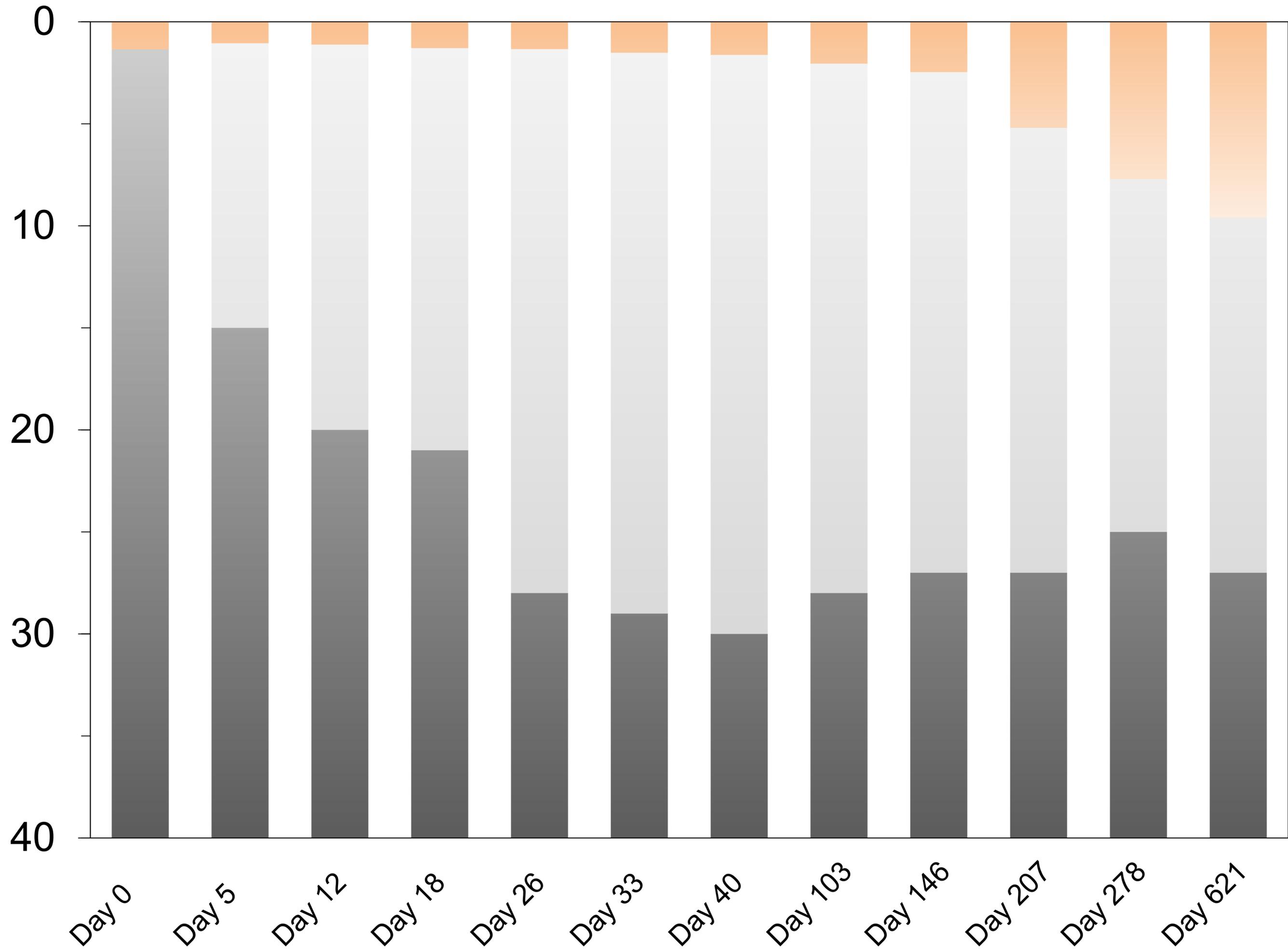
978

979 **Fig. 10.** Time-series of the depth integrated (0-5 cm) increase in Fe oxides (red) and the depletion of FeS (black) and siderite
980 (grey) in $mmol m^{-2}$. Negative values represent a decrease, whereas positive values indicate an increase in the mineral pools.

A**B****C****D**

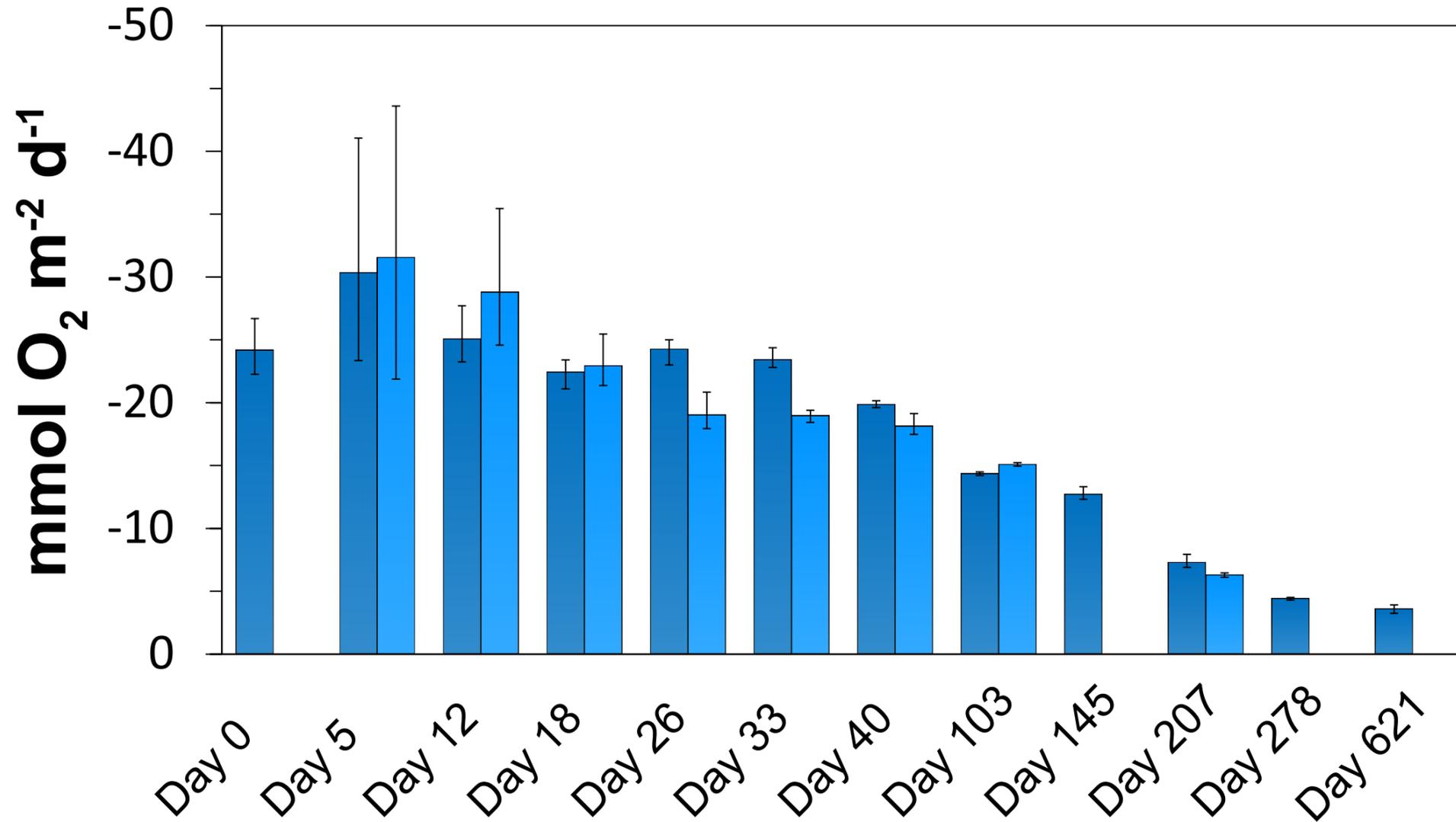


Depth Zonation [mm]

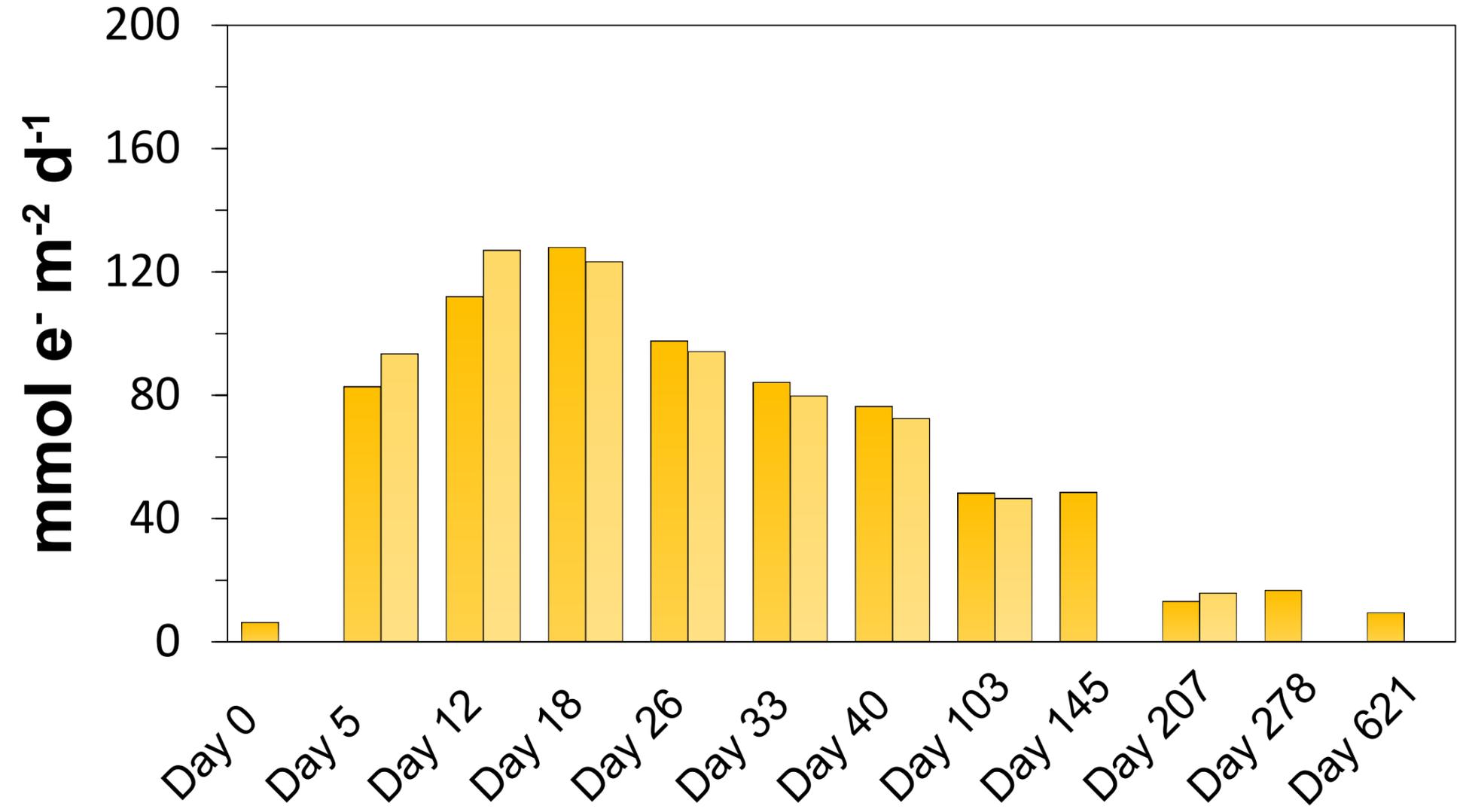


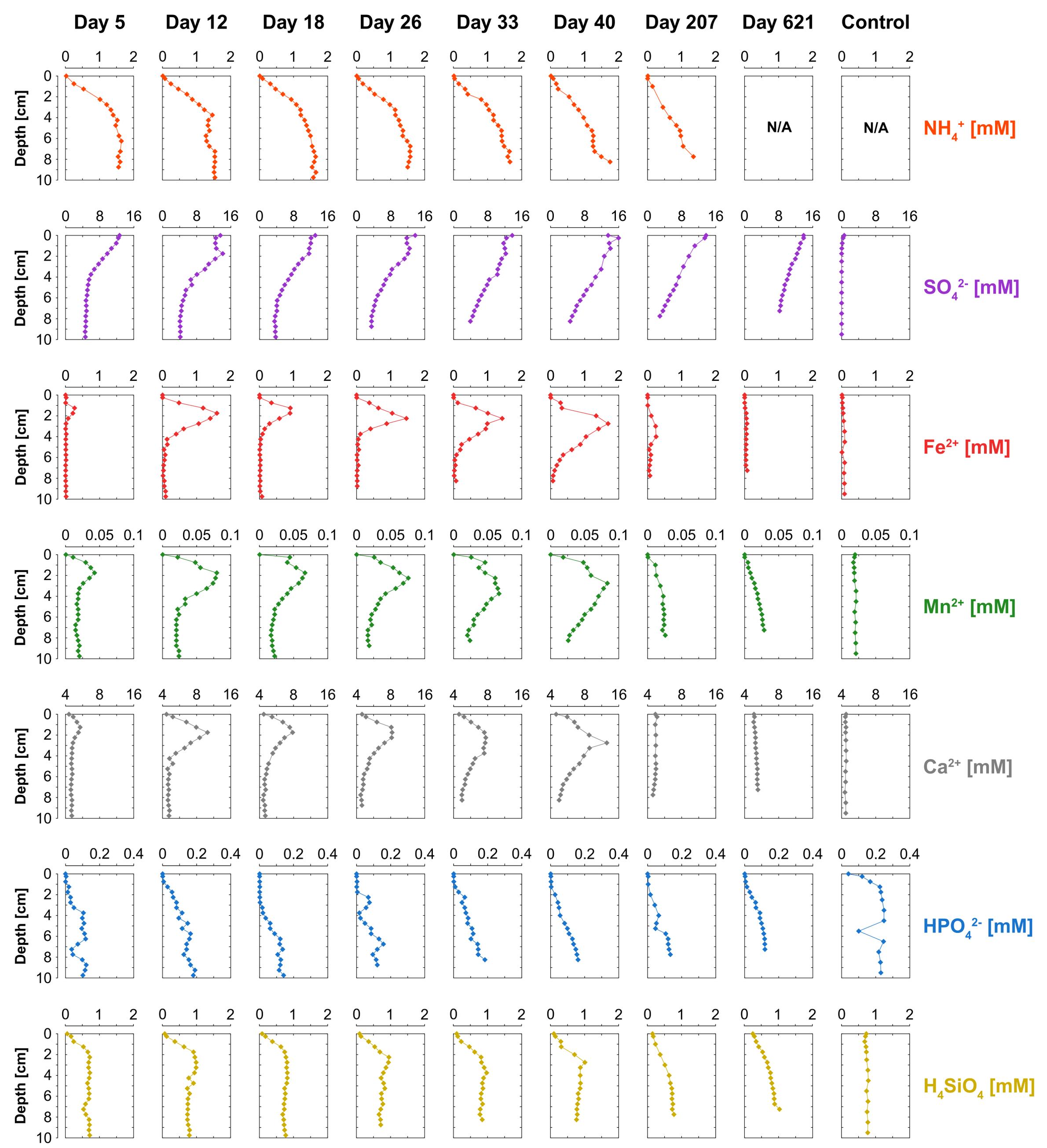
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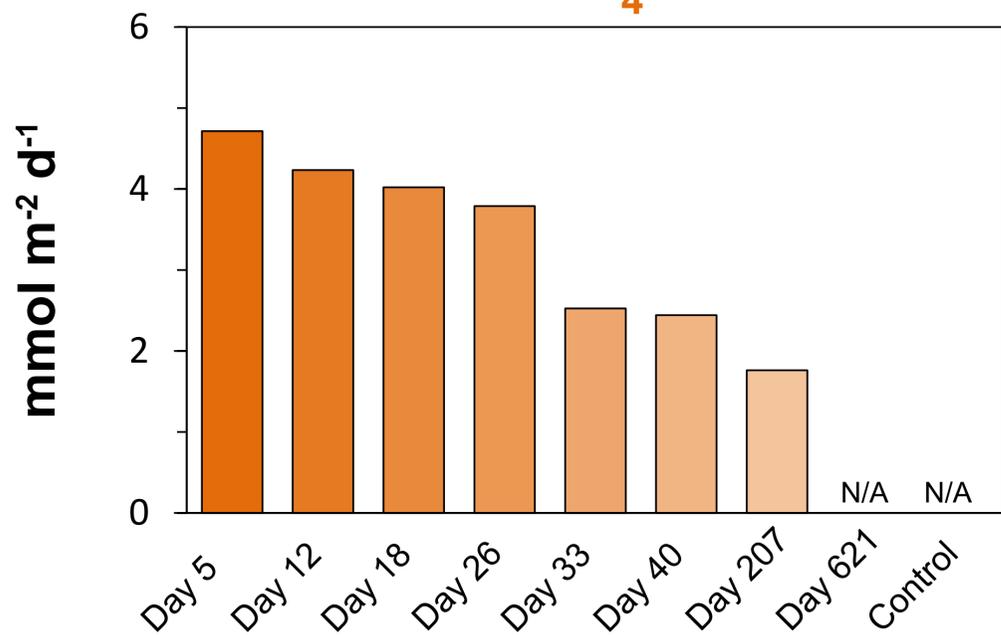
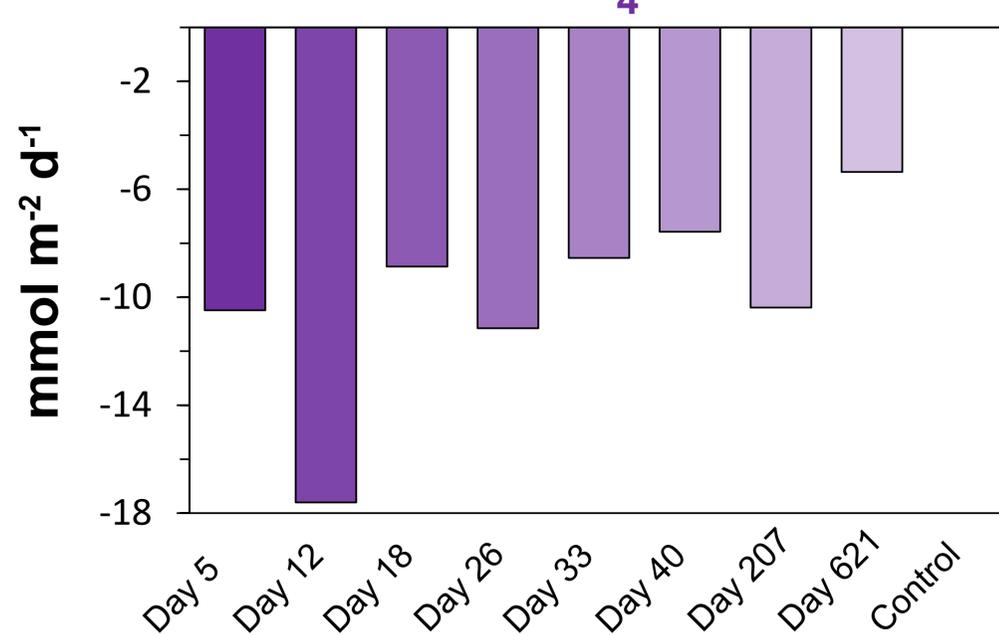
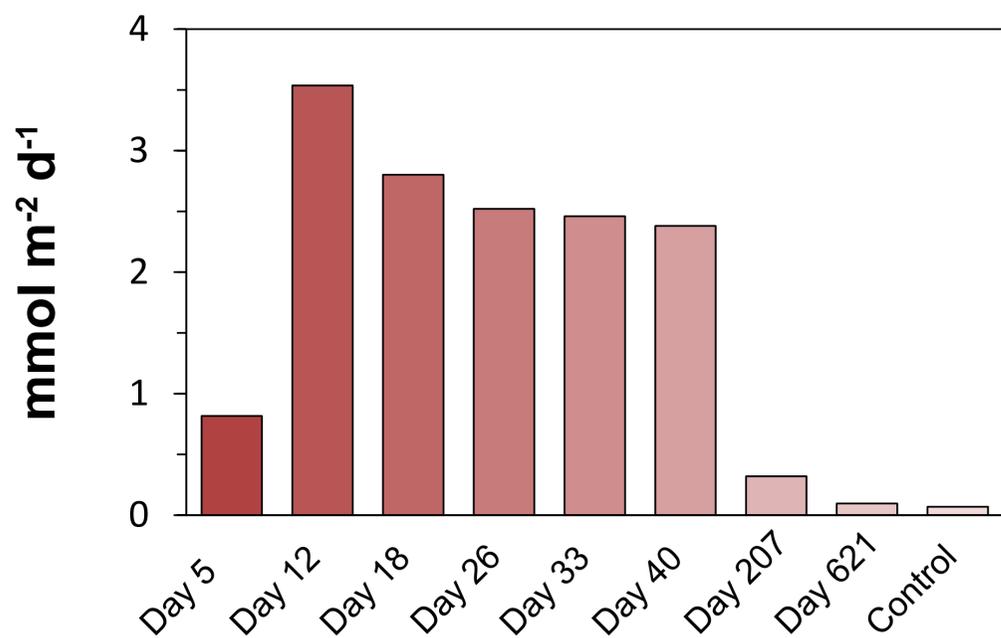
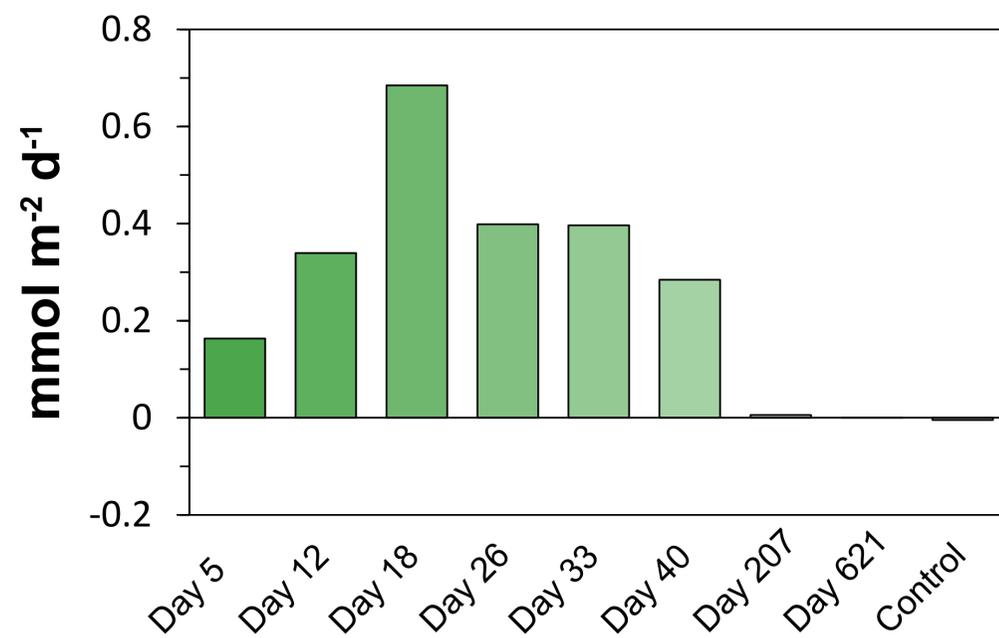
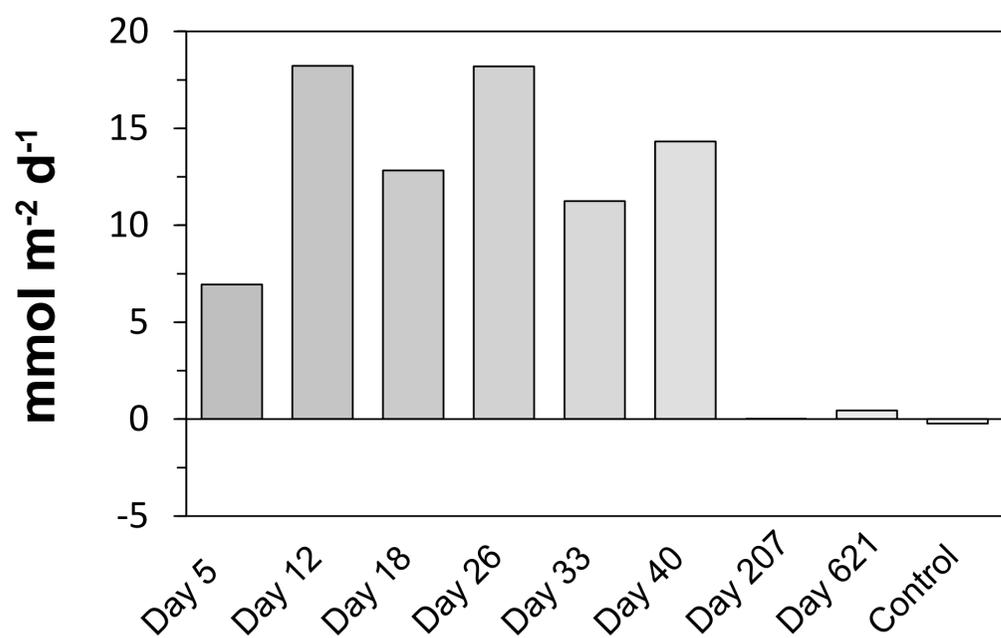
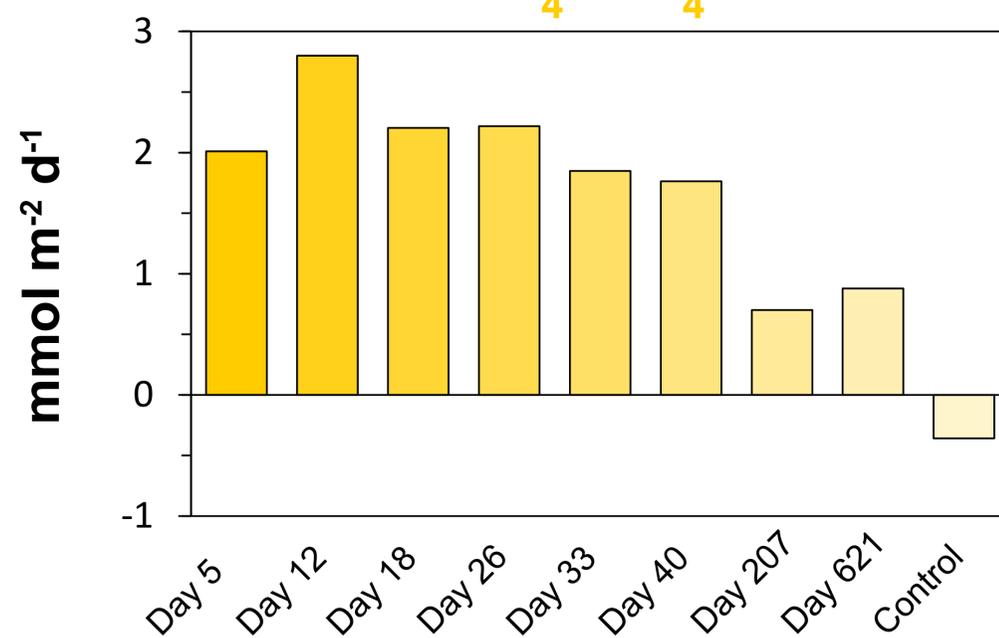
Diffusive O₂ Uptake

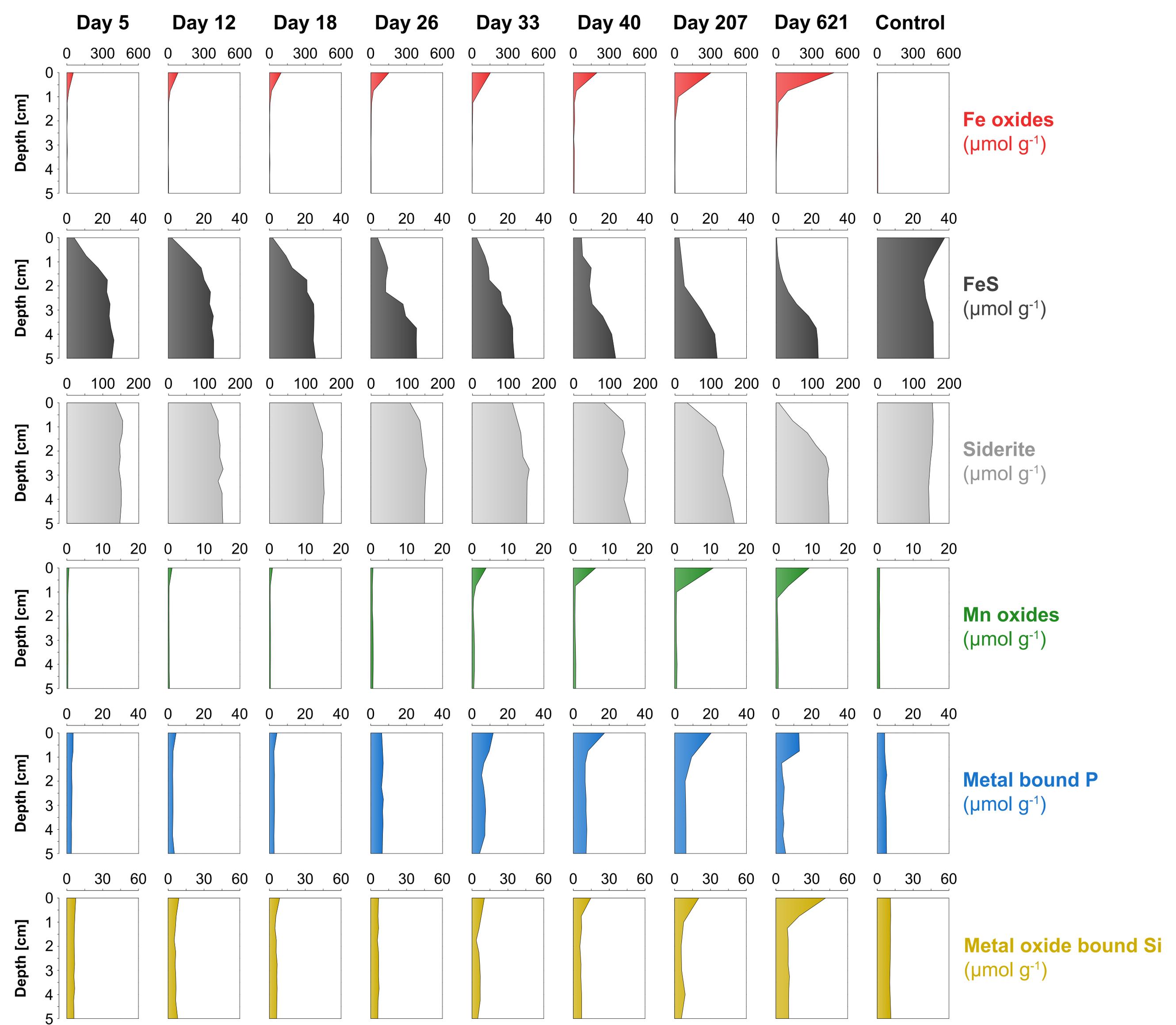
**B**

Current Density



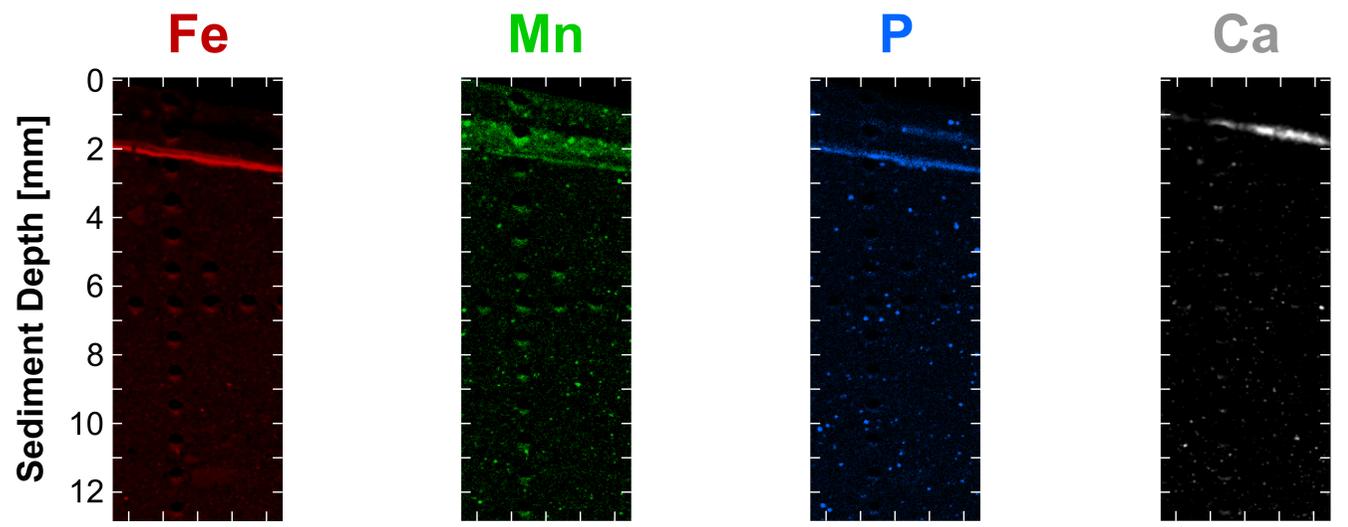


A**B****C****D****E****F**

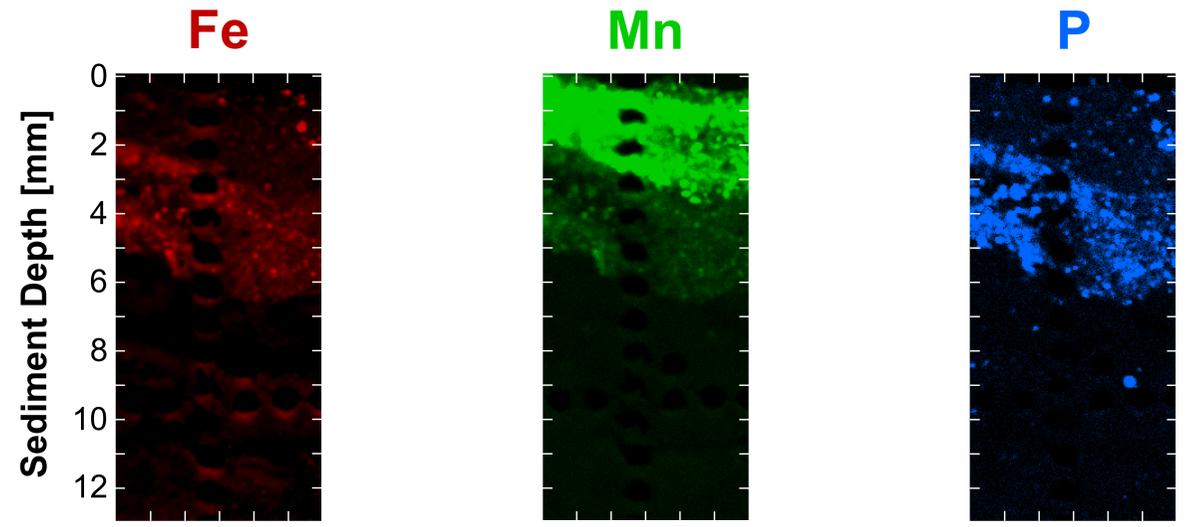


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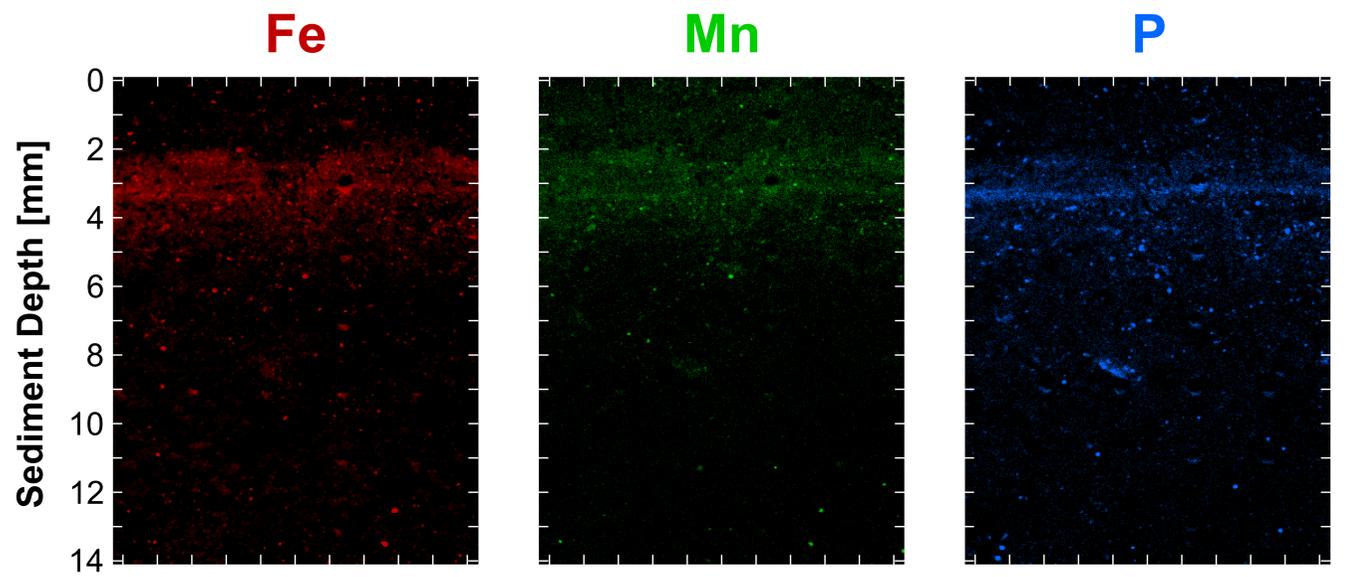
Black Sea Experiment
Cable bacteria

**B**

Gulf of Finland (GOF5)
Cable bacteria

**C**

Lake Grevelingen in January
Cable bacteria

**D**

Lake Grevelingen in May
Macrofauna

