Dear Prof. Treude,

We would like to thank you and the two anonymous referees for reviewing our paper and the insightful and constructive feedback. Please find our replies below where we explain how we addressed each comment and the changes that were made in our manuscript. Unless otherwise indicated, the line numbers in our replies refer to the revised manuscript. We include a marked-up version of our manuscript that highlights all the relevant changes that were made, and we now uploaded the high-resolution version of the figures at the end of our manuscript.

Kind regards,

Martijn Hermans

Prof. Treude’s comments:

Instead of 'bottom water' you could also use the term "supernatant" or 'bottom-near' water. I was in the past criticized by physical oceanographers for using the term 'bottom water' in my sediment studies, because it is a fixed term in oceanography and covers a large volume (an entire water mass).

Reply: We prefer to use “bottom water” since for chemical oceanographers this is a common term. We now made more explicit what bottom water refers to see line No. 32: “Depletion of oxygen (O₂) in bottom waters (i.e. water directly above the seafloor)”

- The term "suboxic" opposite to oxic and anoxic is not ideal, because suboxic is also anoxic. But I am aware that the term is frequently used together with cable bacteria biogeochemistry. I am OK with keeping the term, but please double check that your final manuscript has a proper definition of the terms at the beginning (oxic = with O₂, suboxic = anoxic but not sulfidic, anoxic = anoxic and sulfidic).

Reply: We explain this in more detail in the revised version of our manuscript, see line No. 51-52: “This spatial coupling of surficial O₂ reduction with H₂S oxidation at several centimetres depth creates a suboxic zone that is devoid of any O₂ and H₂S”

- In Fig. 2B a red and a green profile are shown in the same graphs. Please use a different color to allow green-red handicapped people to enjoy it.

Reply: We updated the figure and changed the colours to blue and orange, so it can be read by colour blind people as well.
Anonymous Referee #1

Received and published: 13 August 2020

The study documents the effect of cable bacteria on sediment geochemistry in repacked sediment cores on sediment collected from the Black Sea. The findings support current paradigms in the cable bacteria literature. This is a nice case study that has been well presented and written up. The dominance of cable bacteria in the oxygen budget is an interesting finding. There is also some additional data on element mapping that sheds further light on the effect of cable bacteria on sediment geochemistry at different sites. I only have a few relatively minor comments and suggestions for improvement.

Reply: We thank Anonymous Referee #1 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

Comment #1 The authors used the classic sediment repacking method to start a cable growth cycle. What effect might homogenisation have had on the final findings? For example, would siderite be likely to be so close to the surface under normal circumstances.

Reply: The homogenisation likely did not have a significant effect on our final findings. Homogenisation of the sediment is essential to obtain proper replicate cores and is known to induce growth of cable bacteria. While some of the FeS and siderite might have oxidised during sieving, homogenising and repacking, this does not affect the conclusions of our experiment. There is still ample FeS and siderite present in our sediment cores at the start of the experiment, and our time-series indicate that both FeS and siderite were dissolved by cable bacteria activity over time. Both FeS and siderite are frequently observed in marine surface sediments (e.g. Sulu-Gambari et al. 2016).

Comment #2 Line 163 – no need to say ‘bottom water’ just water samples?

Reply: Here, we specifically use the term bottom water samples, since it refers to the overlying water in the cores. Water samples could also refer to water column or pore water samples. Therefore, we think it is best to use the term bottom water samples, to avoid potential confusion.

Comment #3 Line 172 – please elaborate a little on exactly what you mean by salt corrections.

Reply: Freeze-drying removes the water from samples. However, the salt stays behind in the dried sediment. Hence, the weight of the freeze-dried sediment used for the sequential extractions needs to be corrected for this salt content if we wish to calculate elemental/mineral concentrations, such as Fe oxide and metal bound P contents per gram of sediment. Without the salt correction the absolute
elemental/mineral concentrations would be underestimated. Hence, we subtract the weight of the salt from the freeze-dried sediment to calculate the ‘real’ weight of the dry sediment. This is now explained in more detail in our methods, see line No. 175-178: “After freeze-drying, the salt from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried sediment.”

Comment #4 Line 205 – could you add a sentence on how the embedding was achieved?

Reply: The resin embedding process is described in more detail, see line No. 209-214: “On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013). Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled glovebox and subsequently embedded with Spurr’s epoxy resin as described in Jilbert et al. (2008). After curing, the epoxy-embedded core was split vertically using a rock saw.”

Comment #5 Line 210 onwards – consider adding this to methods.

Reply: The methods used to obtain these elemental maps from the epoxy resin embedded surface sediments from the Gulf of Finland and Lake Grevelingen are described in detail in the other studies cited (Sulu-Gambari et al. 2016; Sulu-Gambari et al. 2018; Hermans et al. Submitted). Therefore, we prefer not to describe the sampling process of those resin embedded cores in section 2.4 of our methodology.

Comment #6 also Line 261 – Only Ca and Si fluxes are presented, I couldn’t see them?

Reply: These Ca and Si fluxes are presented in Fig. S10 and Table S4 in the Supplementary Information, see line No. 480 in the manuscript.

Comment #7 Line 384 not clear what you mean here, please elaborate.

Reply: We have used Fick’s law for the calculation of the diffusive fluxes, which not does take the effect of the electric field generated by cable bacteria on the diffusion potential into account. The Nernst-Planck equation, however, extends Fick's law, because solutes can also be moved with respect to the fluid by electrostatic forces. By using Fick’s law, we underestimate the SO$_4^{2-}$ reduction rate by a most ~10-20%. We will describe this in more detail, see line No. 393-397: “Solutes can also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the relatively low strength of the electric field in the cores (<0.073V m$^{-1}$ at day 18; as estimated from Fig 2B), including the contribution of ionic drift to the sulphate flux would lead to SO$_4^{2-}$ reduction rates that are at most 10-20% higher.”
Comment #8 Line 415 – could it be that the nitrate is just denitrified? Also on this point, it seems that not flux measurements were made for nitrate. It seems likely that some nitrate is released to the water. It might be worth a brief discussion of a few scenarios here. All the nitrate is released to the water column, all the nitrate is denitrified by sediment bacteria and all the nitrate is denitrified by cable bacteria.

Reply: Unfortunately, we do not have data for NO$_3^-$.

The problem of the abovementioned scenarios is that, if we would include the role of other groups of bacteria, and the potential release of NO$_3^-$ to the water column, the mass balance for O$_2$ would have an even greater mismatch. When looking at the stoichiometry of NO$_3^-$ by cable bacteria and the conversion of N to N$_2$, we cannot close the budget fully that way. The text is now modified and explains this in more detail, see line No. 425-430: “These findings can be explained, however, if we assume that at least part of the NO$_3^-$ that is being formed near the sediment-water interface is also used for the metabolic activity of cable bacteria. It has been shown that cable bacteria can couple the oxidation of $\Sigma$H$_2$S to NO$_3^-$ in the absence of O$_2$ (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments where O$_2$ is present in concert with NO$_3^-$ near the sediment-water interface. However, we cannot exclude release of NO$_3^-$ to the water column or denitrification by other bacteria in the sediment.”

Comment #9 Line 485 – very interesting!

Reply: Thank you, this potential niche for vivianite formation is indeed an interesting finding.

Comment #10 Line 522 – not obvious to me from Fig 8A, it is interesting, can you make this clearer?

Reply: This is now more explicit in the text, see line No. 524-527: “While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P is strongly correlated with Ca.”

Comment #11 Line 546 I agree this is likely driven by cables, but how is this different from a straight reaction diffusion scenario (given ubiquity of cables, such a scenario does seem unlikely though). I think this idea needs a little more development and explanation as to how it might actually be applied.

Reply: Focusing of Fe and Mn oxides and associated P can indeed also occur in sediments overlain by oxic waters, where no cable bacteria are active. However, as demonstrated by our experiment (and various other studies (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016), in sediment populated by active cable bacteria, the upward fluxes of Fe$^{3+}$ and Mn$^{2+}$ are higher due to the dissolution of FeS, Fe- and Mn carbonates. This allows strong focusing of Fe and Mn oxides in a thin layer within a relatively short time
frame. We added the following section in our manuscript, see line No. 567-577: “Focussing of Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and can also occur in the absence of cable bacteria. However, the upward fluxes of Fe$^{2+}$ and Mn$^{2+}$ in sediments populated by cable bacteria are higher due to active dissolution of Fe and Mn minerals at depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame following an environmental perturbation (such as a transition to oxic bottom waters after a period of anoxia or mixing of the sediment), more Fe$^{2+}$ and Mn$^{2+}$ can oxidise upon contact with O$_2$ near the sediment-water interface and thus stronger enrichments of Fe and Mn minerals will be observed. Hence, focussing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in sediments populated by active cable bacteria compared to sediments where no cable bacteria are active under such conditions.”

Comment #12 Figure 3 not clear how this was generated. Based on the pictures?

Reply: The depth intervals of the oxic, suboxic and anoxic zone are based on the micro-electrode data. We made this more explicit in the caption of Fig. 3., see line No. 948-950: “Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two different cores.”
References


Anonymous Referee #2

Received and published: 31 August 2020

The manuscript represents a very comprehensive study of potential processes and effects of cable bacteria in sediments. Investigations on cable bacteria and their influence on biogeochemical processes are still in the beginning, but more and more studies show their importance for the element cycling; importance of cable bacteria activity on the oxygen demand in coastal sediments. In the present study, the authors used sediment cores from the coastal area of the Black Sea, which they homogenized and freed from macrofauna. This probably increased the availability of labile organic material and its distribution in deeper sediment layers. Furthermore, the sediment was anoxically stored until the experiment, during the experiment the overlying bottom water was saturated with oxygen so that a steady state must be established at the beginning of the incubations. This fact does not reduce the results of the experiment or the quality of the manuscript.

Reply: We thank Anonymous Referee #2 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

Comment #1 However, the authors should consider the study presented here as potential processes and not directly related to a coastal region (in this case the Black Sea). Therefore, I would strongly suggest to rewrite the manuscript and change the focus of the manuscript by concentrating on the "potential processes and biogeochemical impacts" rather than to directly relate it to coastal sediments of the Black Sea.

Reply: We are aware that the outcomes from our experiment are potential processes, which cannot be directly translated to the field site. This is why, in our title we specifically used the term “on coastal Black Sea sediment” instead of “in coastal Black Sea sediments”. Other examples of sentences in our manuscript (here these line numbers refer to the old version of our manuscript) that indicate that we are not directly relating our results to a coastal region are:

- Line No. 100: “In this study, we assess whether cable bacteria activity can establish in sediments that are relatively poor in FeS in an incubation experiment using siderite-bearing sediments from a coastal site in the Black Sea.”
- Line No. 552: “We can only speculate about the possible in-situ relevance of cable bacteria at the coastal site in the western Black Sea where the sediment for our incubation was collected.”
- Line No. 565-566: “Further field studies are required to assess the role of cable bacteria at our field site, preferably including an assessment of the burrow structures.”
We added the following additional text in the abstract, introduction and conclusion sections to further emphasise that we are referring to potential processes.

Abstract; line No. 14-15: “to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and sulphur (S).”

Introduction; line No. 100-103: “In this study, we assess whether cable bacteria activity can establish and thrive in sediments that are relatively poor in FeS. Although, this will be done in a controlled incubation experiment with siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural environments populated by cable bacteria.”

Conclusion; line No. 599-605: “The results of our laboratory incubation (with a total duration of 621 days) show that cable bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The strong acidity of the pore water associated with the activity of cable bacteria, which was monitored using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a source of Fe that can form an Fe oxide-enriched surface layer.”

Comment #2 The difference between the natural distribution of cable bacteria and the experiment is also evident when looking at Fig. 1c. The authors can use their main results as shown here, but the focus should be on the conditions used in their experiment, which are rather artificial, but very nicely show the potential of cable bacteria in the biogeochemical cycling.

Reply: See our reply to comment #1 and the associated changes in the text. We are aware that the outcomes of our experiment regarding the impact of cable bacteria are amplified when compared to field conditions and that we cannot directly link this to the field site. This is because we optimized conditions to sustain metabolic activity of cable bacteria, and their growth:

1) The sediment was homogenised, which is known to induce growth of cable bacteria.
2) There was no bioturbation by meiofauna and macrofauna.
3) The bottom water was continuously oxygenated.

This optimisation was deliberately done to study the effects of cable bacteria on sediment biogeochemistry. We note that the approach used is common in studies of the biogeochemical impact of cable bacteria (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016).
Comment #3 In a second step the transfer to coastal sediments and their biogeochemical conditions can be done. Here the manuscript lacks the coherence (hypoxia and oxy-gen depletion as mentioned in the Introduction). In a final paragraph the transfer of the laboratory experiment to natural sediments and possible variations in biogeochemical processes as well as the influence of macrofauna (bioturbation and bioirrigation) can be discussed.

Reply: Sections 4.1 to 4.4 focus only on the experiment. In sections 4.5 and 4.6 we discuss the implications for the field. We added text to section 4.5 and 4.6 to clarify when we are referring to other laboratory experiments and results of field studies and the link with hypoxia.

Line No. 520-522: “This colour zonation is typical for sediments that have been geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016)”

Line No. 565-567: “may act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory experiments and at field sites, also in cases where visual observations are not conclusive.”

Line 577-578: “Macrofaunal activity within natural environments likely counteracts or prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer.”

Line 584-586: “At this site, which is in a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m⁻²) and polychaetes (up to ~1700 ind. m⁻²)…”

Comment #4 Is there any information about the organic carbon content of the sediment and how this changes over the incubation period? I would assume that this is the major driver for the development of biogeochemical zonation.

Reply: We now include the organic carbon content of the upper 0-0.5 cm of the sediment at the field site, as determined by Lenstra et al. (2019), in Table 1. We did not determine the change in organic carbon contents during the experiment because, at the typical rates of organic matter degradation expected here, only a small change in organic carbon content would be observed. Hence, we chose to focus on pore water NH₄⁺ profiles to obtain insight in rates of (anaerobic) organic matter degradation.

Comment #5 How does the development of the oxic zone, as shown in the experiment, relate to natural variations in coastal sediments?
Reply: The range in O2 penetration observed in the experiment is comparable to that observed in coastal systems with seasonal hypoxia (e.g. Seitaj et al. 2015). We included this in the manuscript: Line No. 531-534: “During the experiment, O2 penetration varied within a narrow range and was initially fixed between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth of 2 mm (Fig. 8A). Such a range in O2 penetration is in accordance with observations in coastal sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by…”

Comment #6 How does the experiment relate to the development of hypoxia and depletion of oxygen in coastal areas? The experiment shows the opposite reaction (from anoxic surface layer to an oxygenated layer).

Reply: As shown in previous work that we refer to in our manuscript (Seitaj et al. 2015; Sulu-Gambari et al. 2016), cable bacteria can induce formation of Fe and Mn oxides in seasonally hypoxic coastal systems during periods of oxygenated bottom waters in spring. As explained in line No. 41-44, the presence of these Fe and Mn oxides can delay the transition towards euxinia by removing hydrogen sulfide. Our experimental setting can be compared to the onset of bottom water re-oxygenation in spring in such seasonally hypoxic environments (i.e. the sediment was stored anoxically, and then exposed to oxygenated water). This allows study of the mineral dissolution and formation reactions in sediments populated by cable bacteria under such conditions, which is the goal of this work.

Comment #7 line2 121/122: .... with overlying water ....Was this bottom water taken from the site or artificial water, as used for the aquarium?

Reply: The 20 cm long cores (filled with 15 cm sediment) were gently submerged in the two aquaria. Hence, the overlying water in the cores at the start of each incubation was the same as the artificial water used in the aquaria.

Comment #8 line 153: ....... core was place outside the aquarium .....Why was the core taken out? was the incubation temperature maintained?

Reply: The entire experiment was carried out at in a temperature controlled laboratory at 20 °C. We made explicit that the flux incubations also took place at 20 °C, see line No. 153-154: “At each time point, one core was placed outside the aquarium at 20 °C...”

Comment #9 Was the overlying water during the 24-hour incubation for the solute flux measurements stirred to avoid stratification? This could have influenced the flux across the sediment-water interface
because stagnant waters lead to an increase of the Diffusive Boundary Layer, which controls the solute exchange.

**Reply:** The overlying water was mixed continuously by bubbling it continuously with air. The airflow was set in such a way that the water was effectively mixed, while the surface layer of the sediment was left undisturbed. We made this more explicit in the methods, see line **No. 153-157:** “At each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of overlying water in the core was continuously aerated. Potential stratification of the overlying water was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent evaporation.”

**Comment #10** Pore water profiles (especially Fig 1a, Fig 2) are very small and it is difficult to recognize the different profiles (O₂, pH, H₂S) different; graphs should be enlarged.

**Reply:** We enlarged these graphs and we increased the font sizes to improve the readability of the figures.
References


Biogeochemical Impact of Cable Bacteria on Coastal Black Sea Sediment

Martijn Hermans¹, Nils Risgaard-Petersen⁵, Filip J.R. Meysman⁴,⁵ and Caroline P. Slomp¹

¹Department of Earth Sciences, Faculty of Geosciences, Utrecht University, P.O Box 80021, 3508 TA Utrecht, the Netherlands
²Center for Geomicrobiology, Section for Microbiology, Department of Bioscience, Aarhus University, Aarhus, Denmark
³Center for Electromicrobiology, Section for Microbiology, Department of Bioscience, Aarhus University, Aarhus, Denmark
⁴Center of Excellence for Microbial Systems Technology, Department of Biology, University of Antwerp, Wilrijk, Belgium
⁵Department of Biotechnology, Delft University of Technology, Delft, the Netherlands

ABSTRACT

Cable bacteria can strongly alter sediment biogeochemistry. Here, we used laboratory incubations to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and sulphur (S). Microsensor depth profiles of oxygen, sulphide and pH in combination with electric potential profiling and FISH analyses showed a rapid development (<5 days) of cable bacteria, followed by a long period of activity (>200 days). During most of the experiment, the current density correlated linearly with the oxygen demand. Sediment oxygen uptake was attributed to activity of cable bacteria and the oxidation of reduced products from anaerobic degradation of organic matter, such as ammonium. Pore water sulphide was low (<5 μM) throughout the experiment. Sulphate reduction acted as the main source of sulphide for cable bacteria. Pore water Fe²⁺ reached levels of up to 1.7 mM during the incubations, due to the dissolution of FeS (30%) and siderite, an Fe carbonate mineral (70%). Following upward diffusion of Fe²⁺, a surface enrichment of Fe oxides formed. Hence, besides FeS, siderite may act as a major source of Fe for Fe oxides in coastal surface sediments where cable bacteria are active. Using µXRF, we show that the enrichments in Fe oxides induced by cable bacteria are located in a thin subsurface layer of 0.3 mm. We show that similar subsurface layers
enriched in Fe and P are also observed at field sites where cable bacteria were recently active and
little bioturbation occurs. This suggests that such subsurface Fe oxide layers, which are not always
visible to the eye, could potentially be a marker for recent activity of cable bacteria.

Key words: cable bacteria, elemental cycling, solute fluxes, iron

1. INTRODUCTION

Depletion of oxygen (O<sub>2</sub>) in bottom waters (i.e. water directly above the seafloor) of coastal areas
is increasing worldwide, as a consequence of eutrophication and climate change (Diaz and Rosenberg
2008; Schmidtko et al. 2017; Breitburg et al. 2018). Low O<sub>2</sub> can lead to the development of coastal
‘dead zones’ characterised by recurrent mortality of marine life (Rabalais et al. 2002; Diaz and
Rosenberg 2008). Progressive eutrophication induces a characteristic response of coastal systems with
transient and seasonal hypoxia (O<sub>2</sub> < 63 µM) transitioning into permanent anoxia (O<sub>2</sub> = 0 µM). In this
later stage, free sulphide (H<sub>2</sub>S) may escape from the sediment and accumulate in the bottom water, a
condition referred to as euxinia (Diaz and Rosenberg 2008; Kemp et al. 2009; Rabalais et al. 2014).
As H<sub>2</sub>S is highly toxic to higher fauna, the development of euxinia may aggravate the ecological
consequences. However, the presence of iron (Fe) and manganese (Mn) oxides in surface sediments
may delay this transition towards euxinia by removing H<sub>2</sub>S and thus preventing an efflux of H<sub>2</sub>S to
the overlying water (Kristiansen et al. 2002; Kristensen et al. 2003; Diaz and Rosenberg 2008).

Cable bacteria are multicellular filamentous sulphur(S)-oxidising bacteria (Pfeffer et al. 2012)
that strongly enhance the formation of Fe and Mn oxides and efficiently remove H<sub>2</sub>S from surface
sediments (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Sulu-Gambari et al. 2016a). Cable
bacteria belong to the Desulfobulbaceae family of the Deltaproteobacteria (Trojan et al. 2016;
Kjeldsen et al. 2019). Cable bacteria can spatially link the oxidation of H<sub>2</sub>S in deeper sediments to the
reduction of O<sub>2</sub> near the sediment-water interface by transporting electrons over centimetre scale
distances (Pfeffer et al. 2012) through a conductive fibre network that is embedded in the cell
envelope (Meysman et al. 2019). This spatial coupling of surficial O<sub>2</sub> reduction with H<sub>2</sub>S oxidation at
several centimetres depth creates a suboxic zone that is devoid of any O<sub>2</sub> and H<sub>2</sub>S, and provides cable
bacteria a competitive advantage over other S-oxidising bacteria in aquatic environments (Meysman 2018). Cable bacteria have been documented in a range of fresh water (Risgaard-Petersen et al. 2015; Müller et al. 2016) and marine environments (Malkin et al. 2014; Burdorf et al. 2017), however, they appear to be particularly active in sediments overlain by seasonally hypoxic bottom waters (Seitaj et al. 2015; Burdorf et al. 2018).

The metabolic activity of cable bacteria establishes an electrical circuit in the sediment, which involves an electron current through the cable bacteria filaments (Bjerg et al. 2018), and an ionic current through the pore water in the opposite direction (Naudet and Revil 2005; Revil et al. 2010; Risgaard-Petersen et al. 2012). As a consequence, an electric potential (EP) is generated in the sediment, which can be used as a reliable indicator for activity of cable bacteria (Risgaard-Petersen et al. 2014).

Cable bacteria activity additionally generates a distinct biogeochemical signature, that can be assessed by pH, O₂ and H₂S depth profiling (Nielsen et al. 2010). Their activity leads to the development of a suboxic zone (i.e. a zone where O₂ and H₂S are both absent), and also induces a pH profile that strongly changes with depth. Cathodic O₂ reduction (O₂ + 4H⁺ + 4e⁻ → 2H₂O) in the oxic zone of the sediment results in a pH maximum (~9) due to proton consumption, whereas anodic sulphide oxidation (H₂S + 4 H₂O → SO₄²⁻ + 10H⁺ + 8e⁻) causes a pH minimum (<6.5) in the anoxic zone (Fig. 1A; Nielsen et al. 2010; Meysman et al. 2015).

The presence of cable bacteria in sediments can strongly impact the elemental cycling of Fe, Mn, Ca and S (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Rao et al. 2016; Sulu-Gambari et al. 2016a; van de Velde et al. 2016). Pore water acidification induced by cable bacteria activity can lead to dissolution of calcium (Ca) carbonates, Fe carbonates (siderite), Mn carbonates and FeS in the zone where the pH is low, thus generating high concentrations of Fe²⁺ and Mn²⁺ in the pore water (Risgaard-Petersen et al. 2012; Rao et al. 2016). When these dissolved species diffuse upward this can lead to strong enrichments of Fe and Mn oxides upon contact with O₂ or for dissolved Fe²⁺, also upon contact with Mn oxides (Wang and Van Cappellen 1996; Seitaj et al. 2015; Sulu-Gambari et al. 2016).
These metal oxides are capable of efficiently buffering the benthic release of \( \text{H}_2\text{S} \) and phosphate (\( \text{HPO}_4^{2-} \)) during periods with low bottom water \( \text{O}_2 \). This so-called ‘firewall’ for \( \text{H}_2\text{S} \) and alteration of the timing of \( \text{HPO}_4^{2-} \) release linked to this buffering can play a key role in regulating water quality in seasonally hypoxic coastal systems (Seitaj et al. 2015; Sulu-Gambari et al. 2016b; Hermans et al. 2019a).

In coastal sediments, \( \text{O}_2 \) typically penetrates to a depth of only several mm’s below the sediment-water interface (Rasmussen and Jørgensen 1992; Rabouille et al. 2003; Glud 2008). This also holds true for sediments inhabited by active cable bacteria (Nielsen et al. 2010; Pfeffer et al. 2012; Larsen et al. 2015). Hence, the oxidation of upward diffusing \( \text{Fe}^{2+} \) and \( \text{Mn}^{2+} \) is expected to take place below and not at the sediment-water interface. We hypothesise that, as a consequence, in the initial stages of cable bacteria activity and in the absence of bioturbation, most \( \text{Fe} \) and \( \text{Mn} \) oxide enrichments will be restricted to a thin subsurface layer of the sediment. However, the sample resolution and timing of the collection of solid phase data in field and laboratory studies published so far do not allow an assessment of this hypothesis.

Cable bacteria are suggested to thrive in coastal sediments characterised by high rates of \( \text{H}_2\text{S} \) production due to high rates of organic matter mineralisation (Malkin et al. 2014; Burdorf et al. 2017; Hermans et al. 2019a). Laboratory and model studies have shown that the dissolution of \( \text{FeS} \) accounts for 12 to 94\% of the \( \text{H}_2\text{S} \) consumed by cable bacteria, while the other source is \( \text{H}_2\text{S} \) production from the reduction of \( \text{SO}_4^{2-} \) (Risgaard-Petersen et al. 2012; Meysman et al. 2015; Burdorf et al. 2018). At present, it is not known if cable bacteria activity can establish in sediments that are relatively low in \( \text{FeS} \) and dissolved \( \text{H}_2\text{S} \).

In this study, we assess whether cable bacteria activity can establish and thrive in sediments that are relatively poor in \( \text{FeS} \). Although, this will be done in a controlled incubation experiment with siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural environments populated by cable bacteria. The metabolic activity of cable bacteria is monitored using microsensor profiles of \( \text{pH} \), \( \text{O}_2 \), \( \text{H}_2\text{S} \) and EP. We also use sediment \( \text{Fe} \) and P speciation and \( \mu \text{XRF} \) of
resin-embedded sediments to test whether we find evidence for subsurface enrichments in Fe oxides and associated P. We find a rapid establishment of cable bacteria (<5 days) and the development of an Fe oxide-rich subsurface layer, with the majority of the Fe ~70% supplied through dissolution of siderite induced by cable bacteria activity. The depth of the Fe oxide layer was directly related to the O₂ penetration depth and we propose that such subsurface enrichments in Fe, which also can contain P and Mn, can be used as a marker for recent cable bacteria activity.

2. METHODS AND MATERIALS

2.1. Study Area and Experimental Set-up

In September 2015, 16 sediment cores (⌀10 cm) were retrieved at a coastal site on the north-western shelf of the Black Sea (27 m water depth; Fig. 1B; Table 1) using a multicorer (Oktopus GmbH, Germany) as described in Lenstra et al. (2019). The overlying water was discarded, and the upper 10 cm of the sediment was transferred into nitrogen purged aluminium bags that were sealed and stored at 4 °C for several months. The anoxic storage is expected to have led to the death of all macrofauna and most meiofauna (Coull and Chandler 2001; Riedel et al. 2012). Prior to incubation, the sediment was passed through a 4 mm sieve to remove large debris and homogenised. Subsequently, the sediment was transferred to 18 transparent polycarbonate cores (⌀6 cm; 20 cm length).

The bottom 15 cm of these cores was filled with sediment and the upper 5 cm with overlying water. The cores were placed in two aquaria filled with artificial seawater (Instant Ocean Sea Salt + Ultra High Quality (UHQ) water) with a salinity of 17.9, identical to the bottom water salinity at the study site. The artificial sea water contained negligible concentrations of NH₄⁺, NO₃⁻, Fe, Mn and P as described in Atkinson (1997) and Hovanec and Coshland (2004). The aquaria were kept in the dark at a constant temperature (~20 °C), and the water was continuously aerated by two aquarium pumps. Sixteen out of eighteen cores were exposed to oxygenated overlying water in the aquaria, whereas the two remaining cores served as an anoxic control treatment. The control cores were tightly sealed with rubber stoppers, to prevent the growth of cable bacteria by excluding O₂ (Nielsen et al. 2010).
Sampling for pore water and solid-phase analyses was performed at eight time points over a total incubation period of 621 days. Each time point involved a three day procedure. On the first day, microsensor depth profiles of EP, O₂, pH and H₂S were obtained in two randomly selected oxic cores and the two anoxic control cores (O₂ profiling was not performed in the anoxic cores). On the second day, solute fluxes were measured in the same oxic cores that were used for microsensor depth profiling on the previous day. On the third day, the two cores were sectioned, of which only one core was processed further for pore water and solid-phase analyses. Photographs were taken at four time points (day 12; 33; 170 and 621) from one oxic core to follow the visual development of the surface sediment during the experiment.

2.2. High-resolution Microsensor Depth Profiling

High-resolution depth profiles of pH, O₂ and H₂S were obtained (50-μm depth resolution; 3 replicate profiles per oxic core; 2 replicate profiles per anoxic core) using commercial micro electrodes (Unisense A.S., Denmark). The O₂ sensor was re-calibrated prior to each measurement, using saturated bottom water (100% [O₂]) and the deeper sediment horizons (0% [O₂]) as calibration points. Calibrations of the pH and H₂S electrodes were performed as described in Hermans et al. (2019b). pH values are reported on the total scale. For depth profiling of EP (500-μm resolution; 3 replicates per core), micro electrodes were used that were custom built at Aarhus University, as described in Damgaard et al. (2014). A robust reference electrode (Ref-RM, Unisense, A.S., Denmark) was used during EP and pH measurements. To exclude turbulence-induced variations in the potential of the reference electrode during EP profiling, a silicon tube filled with foam was mounted on the tip of the reference electrode.

2.3. Solute Flux Measurements

Solute flux incubations were performed for NH₄⁺, Fe²⁺, Mn²⁺, Ca²⁺, HPO₄²⁻ and H₄SiO₄. At each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of overlying water in the core was continuously aerated. Potential stratification of the overlying water was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent evaporation. Water samples of 3 mL were retrieved at 7 time points over 24 hours. The same volume
of fresh artificial seawater was added to the cores directly after taking each sample. The samples were filtered (0.45 μm), and subsamples were taken for ammonium (1 mL) and for metals (1 mL; acidified with 10 μL Suprapur® HCl (35%) per mL sample), which were stored at -20°C and 4°C respectively until further analysis.

2.4. Pore Water and Sediment Collection

At each time point, two cores were sectioned at 0.5-1 cm resolution with an UWITEC push-up pole in a nitrogen-purged glovebag, but only samples for one core were used for sediment and pore water collection and analyses. Bottom water samples were retrieved from the overlying water in the cores. Slices for each depth interval were centrifuged at 3500 rpm for 20 minutes for pore water retrieval. Samples (1 mL) for NH₄⁺ were taken and stored at -20°C until analysis. Samples (1 mL) for pore water S, Fe, Mn, Ca, P and Si were also collected and acidified with 10 μL Suprapur® HCl (35%) per mL sample, which were stored at 4°C until analysis. Centrifuged sediment samples were freeze-dried and ground to a fine powder in a nitrogen-purged glovebox under a strictly anoxic environment to prevent oxidation (Kraal et al. 2009; Kraal and Slomp 2014). Only the top 5 cm of the solid-phase samples were analysed in further detail. The porosity (Supporting Information 1.1; Table S1) was calculated from the weight loss upon freeze-drying, using a sediment density of 2.65 g cm⁻³ (Burdige 2006). Salt corrections were performed on the solid-phase data using the gravimetric water content and salinity to determine the amount of salt after freeze-drying. After freeze-drying, the salt from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried sediment.

2.5. Chemical Analysis of the Water and Sediment

Concentrations of NH₄⁺ in the pore water and solute flux samples were determined using the phenol hypochlorite method (Koroleff 1969). The total Fe, Mn, Ca, P and Si concentrations (which are assumed to represent Fe²⁺, Mn²⁺, Ca²⁺, HPO₄²⁻ and H₄SiO₄) in the pore water and solute flux samples were determined using Inductively Coupled Plasma-Optical Emission Spectroscopy (ICP-OES, Spectro Arcos). Dissolved Fe and Mn are assumed to be present in the form of Fe²⁺ and Mn²⁺.
however some Mn\(^{3+}\) (Madison et al. 2013) or colloidal and nanoparticulate Fe and Mn might also be available (Boyd and Ellwood 2010; Raiswell and Canfield 2012). Concentrations of P and S are assumed to represent HPO\(_4^{2-}\) and SO\(_4^{2-}\) respectively. The colourimetric detection limit for NH\(_4^+\) was 0.5 μM. The practical detection limit on the ICP-OES for Fe, Mn and P was 0.73, 0.11 and 7.30 μM, respectively.

Solid-phase Fe was fractionated into [1] labile ferric Fe (hydr)oxides and ferrous Fe (FeS + FeCO\(_3\)), [2] crystalline Fe minerals, [3] magnetite and [4] pyrite (Supporting Information 1.2; Table S2), using a combination of two operational extraction methods (Poulton and Canfield 2005; Claff et al. 2010) as described by Kraal et al. (2017). Concentrations of Fe in all extracts were determined using the colourimetric phenanthroline method (APHA 2005). Solid-phase S was separated into [1] acid volatile sulphur (AVS; representing FeS) and [2] chromium reducible sulphur (CRS; representing FeS\(_2\); Table S1) using the method after Burton et al. (2006; 2008) as modified by Kraal et al. (2013). Sulphide released during the S extraction was trapped as ZnS in alkaline Zn acetate traps. Concentrations of S were determined by iodometric titration (APHA 2005). Solid-phase siderite (FeCO\(_3\)) was determined by subtracting AVS from the labile ferrous concentrations retrieved from the first step of the Fe extraction. Solid-phase P was fractionated into [1] exchangeable P, [2] citrate-dithionite-bicarbonate (CDB)-P, [3] authigenic P, [4] detrital P and [5] organic P (Table S2) after Ruttenberg (1992) as modified by Slomp et al. (1996). The sum of exchangeable P and CDB-P represents metal bound P, as described in Hermans et al. (2019b). Concentrations of P in all extracts, except CDB, were measured with the molybdenum blue colourimetric method (Murphy and Riley 1958). The P, Mn (assuming to represent Mn oxides; Hermans et al. 2019b) and Si (assuming to represent metal oxide bound Si; Kostka and Luther III 1994; Rao et al. 2016) in CDB extracts was determined using ICP-OES.

### 2.6. Elemental Mapping of Fe, Mn, P and Ca

On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013). Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm
diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled glovebox and subsequently embedded with Spurr’s epoxy resin as described in Jilbert et al. (2008). After curing, the epoxy-embedded core was split vertically using a rock saw. The surface was smoothed by applying a 0.3 μm alumina powder layer. Elemental maps of Fe, Mn, P and Ca (30 μm resolution) were retrieved using a Desktop EDAX Orb is μXRF analyser (Rh tube set at 30 kV, 500 μA, 300 ms dwell-time, equipped with a poly-capillary lens). Similar μXRF maps for Fe, Mn and P in epoxy embedded surface sediment were obtained for two field sites: (1) the Gulf of Finland, for sediments collected in June 2016 as described by Hermans et al. (Submitted), and Lake Grevelingen, for sediments collected in January and May 2012 as described in Sulu-Gambari et al. (2016a; 2018).

2.7. Fluorescence In-situ Hybridisation

Fluorescence in-situ hybridisation (FISH; Pernthaler et al. 2001) was used to microscopically quantify the abundance of cable bacteria filaments, as described in Seitaj et al. (2015). FISH analysis was performed on one intact sediment core retrieved at our sampling site, and the sediment cores from our incubation experiment used for pore water collection at three time points (days 5, 26 and 207). These cores were sectioned at 0.5 cm depth resolution for the first 2.5 cm. Each sediment slice was homogenised and fixed with 0.5 mL ethanol (≥99.8% purity), and stored in a 2 mL Eppendorf tube at -20 °C. For FISH analysis, a volume of 100 μL was retrieved from the Eppendorf tubes and mixed with a 1:1 solution of PBS/ethanol (500 μL). Then 10 μL of this mixture was filtered through a polycarbonate membrane (type GTTP; pore size 0.2 μm, Millipore, USA). Cable bacteria were classified with a Desulfobulbaceae-specific oligonucleotide probe (DSB706; 5-ACC CGT ATT CCT 3′) after counter staining with DAPI (1 μg/mL) under an epifluorescence microscope (Zeiss Axioplan, Germany) at 100x magnification. The abundance of cable bacteria was quantified by determining the length and diameter of all observed filaments in a field (105 × 141 μm) on the filter at 100x magnification (200 fields per sample). Cable bacterial abundances are expressed as filament length per volumetric unit (m cm⁻³) or depth integrated per unit area of sediment surface (m cm⁻²), consistent with previous studies (Schauer et al. 2014; Malkin et al. 2017).
2.8. Scanning Electron Microscopy

Cable bacteria filaments were taken from surface sediments from the oxic zone (upper 2 mm) after 40 days using a microscope and were transferred to a 15 mL centrifuge tube. The tube was filled to a volume of ~10 mL using ultra clean water, and was subsequently centrifuged at 2100 rpm for 2 min, after which the water was discarded. This washing step was repeated three times. The washed samples were then transferred to a sample stub, where the sediment was air-dried over-night prior to gold coating. The filaments were subsequently subjected to scanning electron microscopy (SEM) imaging on a Phenom ProX Desktop SEM (Phenom-World B.V., the Netherlands) to obtain high-resolution images, as described in Geerlings et al. (2019). SEM images were generated under 0.1-0.3 mbar vacuum, and a high accelerating voltage (10 or 15 kV).

2.9. Data Analysis and Calculations

The diffusive uptake of O$_2$ was calculated from the high-resolution O$_2$ depth profiles using the PROFILE software package (Berg et al. 1998). Total H$_2$S ($\Sigma$H$_2$S = H$_2$S + HS$^-$ + S$^{2-}$) was calculated as a function of the recorded H$_2$S and pH values, accounting for temperature and salinity (Millero et al. 1988; Jeroschewski et al. 1996).

The EP depth profiles were normalised by subtracting the background EP signal in the overlying water from the EP depth profiles, to calculate the EP value relative to that in the overlying water (Damgaard et al. 2014). The electric field in the sediment was calculated from the linear slope of the EP depth profiles (average of triplicates) in the surface sediments (Risgaard-Petersen et al. 2014). The magnitude of the current density was subsequently calculated from the gradient in the EP, the so-called electric field, using Ohm’s law:

$$ J = \sigma_{sed} \cdot E $$

where $J$ represents the magnitude of the current density (mA m$^{-2}$), $\sigma_{pw}$ is the conductivity of the sediment matrix (S m$^{-1}$) and $E$ (mV m$^{-1}$) represents the electric field. The conductivity of the pore water was corrected for tortuosity and calculated as a function of the temperature and salinity using the equations provided by Fofonoff and Millard Jr (1983).
The solute fluxes were calculated as described in Glud (2008) and Rao et al. (2016):

$$J = \frac{\Delta C_{ow}}{\Delta t} \cdot \frac{V_{ow}}{A}$$ (2)

where $J$ represents the diffusive flux (mmol m$^{-2}$ d$^{-1}$), $\Delta C_{ow}$ represents the concentration change in the overlying water (mmol m$^{-3}$), $\Delta t$ is the incubation time (d), $V_{ow}$ is the volume in the overlying water (m$^3$) and $A$ the surface area of sediment in the core (m$^2$). In our experimental setup, only those fluxes were measurable for $\text{NH}_4^+$, $\text{Fe}^{2+}$, $\text{Mn}^{2+}$ and $\text{HPO}_4^{2-}$, that were $>0.08$, $>0.06$, $>0.01$ and $>0.55$ mmol m$^{-2}$ d$^{-1}$, respectively. However, for these four solutes, fluxes were always too low to be detected. Hence, only $\text{Ca}^{2+}$ and $\text{H}_4\text{SiO}_4$ fluxes are presented.

Diffusive downward fluxes of $\text{SO}_4^{2-}$ and diffusive upward fluxes of $\text{NH}_4^+$, $\text{Fe}^{2+}$, $\text{Mn}^{2+}$ and $\text{Ca}^{2+}$ were calculated from linearized pore water gradients using Fick’s first law (Berner 1980):

$$J = -\phi D_s \cdot \frac{dC}{dz}$$ (3)

The molecular diffusion coefficient was calculated as a function of pressure, salinity and temperature using the R package marelac (Soetaert et al. 2010) and corrected for the ambient tortuosity using the relations listed in Boudreau (1997).

3. RESULTS

3.1. Abundance of Cable Bacteria

Examination of the top 2.5 cm of the surface sediments using FISH showed the presence of filamentous cable bacteria (Fig. 1C; Fig. S1). The in-situ cable bacterial abundance in the sediment at our field site was low (14 m cm$^{-2}$). However, after 5 days of incubation in the laboratory, the abundance increased strongly (724 m cm$^{-2}$). At day 26 the abundance of cable bacteria was even higher (1035 m cm$^{-2}$). After 207 days, the cable bacterial abundance in the surface sediment was low again (131 m cm$^{-2}$). SEM imaging confirmed that filaments were indeed cable bacteria (Fig. 1D), as the external surface of the filament was characterised by a parallel pattern of ridges and grooves along...
its latitudinal axis, which is a typical feature of cable bacteria (Cornelissen et al. 2018; Geerlings et al. 2019).

3.2. High-resolution Depth Profiles of pH, O₂, \( \sum H_2S \) and EP

High-resolution depth profiles of pH showed the development of a distinct peak near the sediment-water interface at day 5, and acidification of the pore water in the deeper sediment (Fig. 2A). The width of this pore water acidification zone increased with time and reached its maximum at day 26, followed by a decrease in the acidification. The distinct pH peak near the sediment-water interface gradually disappeared after 33 days. The depth of O₂ penetration in the sediment remained constant within the first 40 days of incubation (~1.1 mm) and subsequently moved downwards with time to 9.6 mm (Fig. 2A; Fig. 3; Fig. S2). The dissolved \( \sum H_2S \) concentrations remained low (<5 μM) throughout the experiment (Fig. 2A). The \( \sum H_2S \) appearance depth was initially equivalent to the O₂ penetration depth, and shifted downwards within 5 days, creating a suboxic zone where O₂ and \( \sum H_2S \) remained below detection (Fig. 2A; Fig. 3). The width of the suboxic zone remained relatively constant with time (~25 mm; Fig. 3), and only slight decreased after 207 days.

The EP depth profiles indicate a rapid establishment of an electric current after 5 days (0.4 mV; Fig. 2B). The time-series of depth profiles show that the EP increased and also accumulated over a thicker depth horizon. At day 26 the EP reached its maximum value (1.2 mV), followed by a decrease with time. Long-distance electron transport was not active in the anoxic control core (Fig. S3).

3.3. Diffusive Uptake of O₂ and Current Density

The diffusive O₂ uptake of the sediment was highest after 5 days and gradually decreased with time from ~30 to ~3.6 mmol m⁻² d⁻¹ (Fig. 4A). The current density rapidly increased from day 0 to day 18, from 6 to 128 e⁻ mmol m⁻² d⁻¹, and then gradually decreased with time (Fig. 4B). The duplicate measurements show the same trend for the diffusive O₂ uptake and the current density, which indicates that the results are reproducible.
3.4. Pore Water Profiles

Concentrations of NH$_4^+$ were low near the sediment-water interface and increased with sediment depth reaching maximum levels of up to 1.7 mM at depth in the sediment (Fig. 5). The time-series suggest a gradual decrease in production of dissolved NH$_4^+$ in the sediment leading to decreasing concentrations with time. The pore water depth profiles of dissolved SO$_4^{2-}$ show a decline with sediment depth at all time points. However, SO$_4^{2-}$ concentrations remained relatively constant within the top 2 cm of surface sediment between day 12 and 33. Dissolved Fe$^{2+}$, Mn$^{2+}$ and Ca$^{2+}$ all show the development of distinct peaks in the pore water with time, and after 40 days those peaks disappear again. Pore water concentrations of HPO$_4^{2-}$ generally increased with sediment depth for all time points, and concentrations within the top 2 cm were below the detection limit indicating removal. Dissolved H$_4$SiO$_4$ increased with sediment depth reaching concentration of up to 1 mM.

3.5. Diffusive Fluxes

Calculated diffusive fluxes of NH$_4^+$ into the oxic zone decreased during the incubation experiment from 4.7 to 1.8 mmol m$^{-2}$ d$^{-1}$ (Fig. 6A; Fig. S4; Table S3). Rates of SO$_4^{2-}$ reduction estimated from the linear gradient of the decrease in pore water SO$_4^{2-}$ in the surface sediment with depth generally also showed a decrease with time (Fig. 6B; Fig. S5; Table S3). The upward diffusive flux of Fe$^{2+}$ greatly increased from day 5 to day 12 and then gradually decreased with time (Fig. 6C; Fig. S6; Table S3). The upward diffusive flux of Mn showed an increase in the initial stage of the experiment and reached its maximum at day 18, followed by a decrease with time (Fig. 6D; Fig. S7; Table 3). The upward diffusive flux of Ca$^{2+}$ showed no clear trend with time, however after 207 days the flux became extremely low (Fig. 6E; Fig. S8). The upward diffusive flux of H$_4$SiO$_4$ also showed an increase in the initial stage of the experiment, and reached its maximum at day 12, followed by a decrease with time (Fig. 6F; Fig. S9).

3.6. Solid-phase Profiles

The surface sediment in the oxic cores became more enriched in Fe oxides with time, with concentrations increasing from 53 to 485 µmol g$^{-1}$ (Figure 7). The deeper sediment in the oxic cores and the entire anoxic control core had low or no Fe oxides. At day 5, FeS was strongly depleted.
within the top 1 cm of the surface sediment and was gradually lost further with time. At day 621, most of the FeS within the top 2.5 cm of the surface sediment had been dissolved. The anoxic core did not show such a depletion of FeS in the surface sediment and even showed a slight increase in FeS. Solid-phase siderite remained rather constant with depth from day 5 to 33, but afterwards was gradually lost from the surface sediment. At day 621 a large proportion of the siderite was dissolved within the top 2 cm. Solid-phase siderite concentrations remained constant with depth in the anoxic control core. Solid-phase depth profiles of Mn oxides, metal bound P and metal oxide bound Si all showed a gradual increase in the surface sediment with time.

3.7. High-resolution Elemental Mapping

High-resolution desktop µXRF mapping of Fe, Mn, P and Ca of our core after 47 days of incubations revealed a subsurface sediment layer highly enriched in Fe and P (Fig. 8A). Subsurface enrichments in Fe, P and Mn in relatively thin layers were also observed in sediments populated by cable bacteria in the Gulf of Finland and Lake Grevelingen (Fig. 8B and C). In the latter system, the layers enriched in Fe, P, Mn broadened upon recolonization by macrofauna (Fig. 8D).

3. DISCUSSION

4.1. Metabolic Activity of Cable Bacteria

Cable bacteria in our incubation experiment demonstrated a rapid growth, since their abundance greatly increased after 5 days, and reached its peak at day 26 (Fig 1C). Such high abundances are similar to those observed in previous experiments, in which FeS-rich marine sediments from Aarhus Bay and Lake Grevelingen were incubated (Schauer et al. 2014; Burdorf et al. 2018). The activity of cable bacteria exerted a strong impact on the pore water depth profiles of pH, O₂, and ∑H₂S, as evident from the development of a pH maximum near the sediment-water interface, the strong pore water acidification in the deeper sediment and the development of a suboxic zone (Fig. 2A). These pore water depth profiles resemble the distinct biogeochemical fingerprint typical for active cable bacteria, as observed in previous laboratory incubation experiments (Risgaard-Petersen et al. 2012; Malkin et al. 2014; Schauer et al. 2014; Vasquez-Cardenas et al. 2015; Rao et al. 2016;
Burdorf et al. 2018). The widening of the suboxic zone with time (Fig 3) is a consequence of the downward expansion of the cable bacteria filament network (Schauer et al. 2014; Vasquez-Cardenas et al. 2015).

The EP depth profiles demonstrated that long-distance electron transport by cable bacteria was already active 5 days after the start of the experiment, as indicated by the increase of EP at depth to 0.4 mV. With time, the EP signal increased to higher values and also accumulated over a thicker depth horizon (Fig. 2B), indicating that cable bacteria activity both increased and extended to deeper sediment depth, which is also a consequence of the downward expansion of cable bacteria filaments. The EP reached a maximum after 18 days (1.3 mV; Fig. 2B) in concert with the highest current density of ~130 mmol e\(^{-}\) m\(^{-2}\) d\(^{-1}\) (Fig 4B). This maximum EP value and current density are similar in magnitude to those found in sediment incubations with seawater with a similar salinity (Damgaard et al. 2014). From day 18 onwards the EP and current density flux gradually decreased with time to 13 mmol e\(^{-}\) m\(^{-2}\) d\(^{-1}\) after 207 days (Fig. 4B), which implies a decrease in the metabolic activity of cable bacteria. The suboxic zone persisted long after the current density had decreased (Fig. 3).

To summarise, the metabolic activity of cable bacteria was likely highest between day 18 and day 26 based on the cable bacterial abundances, the extent of acidification of the pore water and the current density (Fig 1C; Fig 2A and Fig 4B).

4.2. Organic Matter Degradation

Ammonium fluxes are assumed to reflect rates of anaerobic degradation of organic matter (Fig. 5), and the observed decline during the experiment coincides with the decrease in activity of cable bacteria based on the EP profiles and current density (Fig. 2B; Fig 4B). This suggests that the availability of easily degradable organic matter plays a role in sustaining the metabolic activity of cable bacteria, most likely by controlling the rate of SO\(_4^{2-}\) reduction (Nielsen and Risgaard-Petersen 2015).

Rates of SO\(_4^{2-}\) reduction estimated from the linear gradient of the decrease in pore water SO\(_4^{2-}\) in the surface sediment with depth indeed also showed a decline during the experiment. We note,
however, that a direct measurement of $SO_4^{2-}$ reduction rates (Fossing and Jørgensen 1989; Kallmeyer et al. 2004) would provide a better indicator, because $SO_4^{2-}$ estimated from pore water profiles are in general lower than rates estimated from tracer experiments (Hermans et al. 2019a; Sandfeld et al. 2020). Another cause for a slight underestimation of our $SO_4^{2-}$ reduction rates, is due to the effect of the electric field imposed by cable bacteria, which is not taken into account in Fick’s law. Solutes can also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the relatively low strength of the electric field in the cores (<0.073 V m$^{-1}$ at day 18; as estimated from Fig. 2B), including the contribution of ionic drift to the sulphate flux would lead to $SO_4^{2-}$ reduction rates that are at most 10-20% higher.

The metabolic activity of cable bacteria can lead to the production of $SO_4^{2-}$ in the suboxic zone via anodic sulphide oxidation (Risgaard-Petersen et al. 2012; Rao et al. 2016). We suspect that this also explains the lack of change in pore water $SO_4^{2-}$ with depth in the upper 2 cm of the sediment in our experiment between 12 and 40 days (Fig. 5). Despite relatively high $SO_4^{2-}$ reduction rates ranging from 5.4 to 17.6 mmol m$^{-2}$ d$^{-1}$ (Fig. 6B; Table S3), pore water concentrations of $\Sigma H_2S$ remained very low throughout the experiment (Fig. 2A). This is likely due to the direct consumption of $\Sigma H_2S$ through the activity of cable bacteria, preventing $\Sigma H_2S$ from accumulating in the pore water, or alternatively, precipitation of FeS by dissolved Fe$^{2+}$ released from the dissolution of siderite.

Laboratory experiments have shown that S-oxidation by cable bacteria can play a dominant role in the $O_2$ uptake of coastal sediments (Nielsen et al. 2010; Schauer et al. 2014; Nielsen and Risgaard-Petersen 2015), and model analysis predicts up to 93% of the total $O_2$ uptake (Meysman et al. 2015). When we plot diffusive uptake of $O_2$ against the current density (i.e. upward flux of electrons towards the oxic zone), a linear relationship – with some scatter - emerges for days 12 to 621 (Fig. 9). However, the data points for day 0 and 5 during the initial stages of our experiment do not follow this linear relationship. We explain these findings as follows: At day 0, the cable bacteria were not active yet and other processes, such as aerobic respiration and oxidation of NH$_4^+$ and other solutes (Table 2) and solids (FeS) dominated the consumption of $O_2$. At day 5 and 12, the activity of cable bacteria and the oxidation of reduced products from anaerobic degradation of organic matter both contributed to
consumption of O$_2$. From day 12 onwards, both the O$_2$ consumption and electron flux follow a downward decrease with time (Fig. 9). If cable bacteria would account for all of the O$_2$ consumption, a ratio between the diffusive uptake of O$_2$ and the current density of 1:4 is expected (Fig. 1A; Nielsen et al. 2010). We find that from day 12 onwards, most data points plot rather close to the line for this 1:4 relationship (Fig. 9), suggesting that cathodic O$_2$ reduction by cable bacteria is responsible for nearly all O$_2$ consumption in the sediment (in line with the model results of Meysman et al. 2015). This however poses a problem for the nitrogen budget, because our data indicate complete removal of the NH$_4^+$ that diffuses upward into the oxic zone (Fig. 6A), and based on the solute fluxes, no escape to the overlying water (see section 2.4). This implies substantial O$_2$ consumption due to nitrification (Table 2). These findings can be explained, however, if we assume that at least part of the NO$_3^-$ that is being formed near the sediment-water interface is also used for the metabolic activity of cable bacteria. It has been shown that cable bacteria can couple the oxidation of $\Sigma$H$_2$S to NO$_3^-$ in the absence of O$_2$ (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments where O$_2$ is present in concert with NO$_3^-$ near the sediment-water interface. However, we cannot exclude release of NO$_3^-$ to the water column or denitrification by other bacteria in the sediment. Another explanation is that cable bacteria might consume O$_2$ directly above the sediment-water interface, as recently has been proposed by Burdorf et al. (2018). Lastly, the current density might be slightly overestimated, since it ignores other sources that can create an electric potential, such as the diffusion potential (Revil et al. 2012; Nielsen and Risgaard-Petersen 2015).

4.3. Impact of Cable Bacteria on Fe, Mn and S Cycling

The activity of cable bacteria had a strong impact on the biogeochemistry of the surface sediment in our experiment (Fig. 7). Cable bacteria activity induced an intense acidification of the pore water in the suboxic zone (Fig 2A), which led to the dissolution of Fe and Mn minerals in deeper sediment layers, as can be inferred from the sharp maxima in dissolved Fe$^{2+}$ and Mn$^{2+}$ in the pore water reaching concentrations of up to $\sim$1700 and $\sim$80 $\mu$M, respectively (Fig. 5). The twenty-fold higher dissolved Fe$^{2+}$ concentrations with respect to pore water Mn$^{2+}$ can be attributed to the relatively higher availability of FeS and siderite compared to the availability of Mn carbonates in the sediment that was
used for incubation (Lenstra et al. 2020). The peaks in dissolved Fe$^{2+}$ and Mn$^{2+}$ in the pore water broadened over time spanning a depth of >5cm (Fig. 5; Fig. S6; Fig. S7).

The upward diffusive flux of dissolved Fe$^{2+}$ and Mn$^{2+}$ was highest after 12 days, reaching values of up to 3.16 and 0.16 mmol m$^{-2}$ d$^{-1}$ respectively. Fluxes subsequently gradually decreased with time (Fig. 6C and D). The continuous upward diffusion of dissolved Fe$^{2+}$ and Mn$^{2+}$ led to enrichments of poorly crystalline Fe and Mn oxides in the surface sediment (Fig. 7). Despite high upward fluxes of dissolved Fe$^{2+}$ and Mn$^{2+}$ towards the sediment-water interface, our solute flux incubations indicate there was little escape of Fe$^{2+}$ and Mn$^{2+}$ to the overlying water (see section 2.4). This implies that all Fe$^{2+}$ and Mn$^{2+}$ that diffused upward was precipitated as Fe and Mn oxides upon contact with O$_2$ or NO$_3^-$ (Buresh and Moraghan 1976; Kuz'minskii et al. 1994; Straub et al. 1996). Little or no escape of dissolved Fe$^{2+}$ from the sediment into the overlying water, was suggested previously for a field site with active cable bacteria based on diffusive flux calculations (Lake Grevelingen; Sulu-Gambari et al. 2016a) and was determined in flux incubations of cores during a laboratory experiment with cable bacteria (Rao et al. 2016).

At the start of the experiment, the sedimentary FeS content was (~25 µmol g$^{-1}$), which is not unusual for coastal sediments on the north-western Black Sea margin (Wijsman et al. 2001), but is low compared to sediments in eutrophic coastal systems (e.g. Morgan et al. 2012; Kraal et al. 2013; Hermans et al. 2019a). The solid-phase depth profiles reveal a gradual removal of the FeS in the surface sediment in our experiment over time (Fig. 7). At the end of our experiment (621 days), there was no longer any FeS within the top 1.5 cm of the sediment. While approximately 90 mmol m$^{-2}$ of FeS was removed from the surface sediment within the first 5 days, a total of ~240 mmol m$^{-2}$ was removed after 621 days (Fig. 10; Table 3). Likely, part of the FeS that was removed from the surface sediment within the first 5 days was removed through oxidation upon contact with O$_2$, rather than the metabolic activity of cable bacteria itself. The pore water acidification associated with cable bacteria activity led to a strong loss of siderite within the top 2 cm of the sediment, with a total removal of ~560 mmol m$^{-2}$ during the experiment (Fig. 7; Fig. 10; Table 3). The depletion of sedimentary FeS
and siderite was directly proportional to the formation of Fe oxides near the sediment-water interface (Fig. 10), and accounted for 30% and 70% of the Fe oxides, respectively (Table 3).

With these data we cannot accurately determine the role of FeS versus SO$_4^{2-}$ reduction in supplying the $\Sigma$H$_2$S sustaining the activity of cable bacteria throughout the experiment. This is primarily related to the variability between cores, and for this type of calculation, the low temporal resolution of sampling. However, we can make an estimation of the relative role of SO$_4^{2-}$ reduction and FeS dissolution in $\Sigma$H$_2$S production, based on the pore water profiles of SO$_4^{2-}$ and dissolved Fe$^{2+}$, and the solid-phase mass balance of FeS and siderite (Fig. 6B and C; Table 4). This estimation shows that SO$_4^{2-}$ was mainly responsible for $\Sigma$H$_2$S production, accounting for 85-99% (Table 4), and thus that the dissolution of FeS only played a minor role in providing $\Sigma$H$_2$S.

4.4. Impact of Cable Bacteria on Ca, P and Si Cycling

Cable bacteria activity is known to lead to dissolution of Ca carbonates, because of the strong acidification of the pore water (Risgaard-Petersen et al. 2012; Rao et al. 2016). We indeed find similar maxima in pore water Ca$^{2+}$ during the experiment (Fig. 5) and a high upward flux of Ca$^{2+}$ (up to $\sim$18 mmol m$^{-2}$ d$^{-1}$; Fig. 6E; Fig. S8) of which a substantial fraction (up to $\sim$55%) escapes to the overlying water (Fig. S10; Table S4), which is consistent with a previous incubation experiment Rao et al. (2016).

Pore water depth profiles of HPO$_4^{2-}$ reveal a production at depth and removal of all upward diffusing HPO$_4^{2-}$ within the first 1-3 cm of the surface sediment (Fig. 5). A major proportion of this HPO$_4^{2-}$ is bound to Fe oxides (Fig. 7). Given that a large proportion of the Fe oxides in our sediment cores derive from the dissolution of siderite, this suggests that the buffer mechanism that delays the benthic release of HPO$_4^{2-}$ through retention of P associated with newly formed Fe oxides (Sulu-Gambari et al. 2016b), might also be active in systems that are relatively poor in sedimentary FeS.

The shape of the pore water HPO$_4^{2-}$ profiles suggests that some of the HPO$_4^{2-}$ is removed below the zone where Fe and Mn oxides are present (Fig. 5; Fig. 7). A possible explanation could be the formation of vivianite, an Fe(II) phosphate mineral. Vivianite formation in sediments typically occurs
when pore water levels of Fe$^{2+}$ and HPO$_4^{2-}$ are high and concentrations of ΣH$_2$S are low (Nriagu 1972), as observed in our study. In our experiment, free ΣH$_2$S does not accumulate in the pore water, which we attribute to removal through the activity of cable bacteria and FeS formation at depth (Fig. 2A; Fig. 7). Hence, cable bacteria may create a geochemical niche that allows the formation of vivianite in the suboxic zone. Further work with sediments with higher P concentrations would be needed to assess this with direct measurement techniques, such as X-ray spectroscopy (Egger et al. 2015; Kraal et al. 2017; Sulu-Gambari et al. 2018). Other sediment P pools, i.e. organic, authigenic and detrital P remained constant over time, indicating that the P contents determined for discrete sediment slices using sequential extractions were not affected by pore water acidification as a result of cable bacteria activity (Table S5).

Pore water H$_4$SiO$_4$ profiles show a typical increase with depth as observed upon dissolution of biogenic silica in marine sediments (Aller 2014). Fluxes of H$_4$SiO$_4$ towards the sediment-water interface range up to ~2.8 mmol m$^{-2}$ d$^{-1}$ and gradually decreased with time (Fig. 6F; Fig. S9). The results of the solute flux incubations indicate that most of this H$_4$SiO$_4$ escaped to the overlying water (ranging from 28 to 92%; Table S4; Fig. S10). The decline in the benthic release flux of H$_4$SiO$_4$ contrasts with results of a previous incubation experiment by Rao et al. (2016) with similar pore water concentrations of H$_4$SiO$_4$ reaching values up to ~1 mM. In their study, the flux remained constant over time, possibly because of differences in the amount of biogenic Si in the sediment. The solid-phase metal oxide bound Si pool in the surface sediment increased directly proportional to the formation of Fe oxides throughout the experiment (Fig. 7). Silica is known to absorb to Fe oxides (Sigg and Stumm 1981; Davis et al. 2002). Hence, the results suggest that the Fe oxides formed through the activity of cable bacteria captured some of the upward diffusing H$_4$SiO$_4$.

**4.5. Sediment Marker for Cable Bacteria Activity**

Visual observations of core photographs reveal the gradual development of an orange layer (oxic zone) up to 9 mm thick, overlying a grey layer (suboxic zone) and a black layer (sulphidic zone) during the experiment (Fig. S11). This colour zonation is typical for sediments that have been geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and
Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016a). High-resolution elemental maps of our sediments reveal the development of a ~0.3 mm thin subsurface layer highly enriched in Fe oxides and associated P, 47 days after the start of the incubation (Fig. 8A). While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P is strongly correlated with Ca. Below, we describe the formation of this layer in more detail and explain why such subsurface enrichments, detected with µXRF, may act as an additional sediment marker for present or recent cable bacteria activity, also in cases where visual observations are not conclusive.

During the experiment, O₂ penetration varied within a narrow range and was initially fixed between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth of 2 mm (Fig. 8A). Such a range in O₂ penetration is in accordance with observations in coastal sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by rapid oxidation of upward diffusing Fe²⁺ upon contact with O₂ (and possibly NO₃⁻; Fig. 6C). Directly above the Fe oxide layer a broader ~0.8 mm thick Mn oxide layer was observed (Fig. 8A). This contrast in zonation between Fe and Mn is likely due to the slower oxidation kinetics of Mn²⁺ compared to Fe²⁺ (Burdige 1993; Luther 2010; Learman et al. 2011).

While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P close to the sediment-water interface (Fig. 8A). In this layer, P is strongly correlated with Ca. This layer likely consists of carbonate fluorapatite (CFA), a Ca-P mineral, which is typically formed in marine sediments (Van Cappellen and Berner 1988; Ruttenberg and Berner 1993). Possibly, the high pore water pH near the sediment-water interface (resulting from cathodic O₂ reduction by cable bacteria; Fig. 2A), promotes apatite formation (Bellier et al. 2006), and the elevated Ca²⁺ concentrations (Fig. 5) created a biogeochemical niche for the formation of CFA.

Such focusing of Fe, Mn, P and associated elements within a thin subsurface layer, as a consequence of cable bacteria activity, also occurs in the field. This was demonstrated by Hermans et
al. (Submitted) in a study of a coastal site in the Gulf of Finland where cable bacteria were recently active. Here, µXRF mapping of resin embedded sediments revealed strong focusing of Fe oxides, Mn(II) phosphates and Fe bound P within a 3 mm thick layer near the sediment-water interface (Fig. 8B). A re-assessment of the µXRF data of Sulu-Gambari et al. (2016a; 2018) of surface sediments with active cable bacteria from seasonally hypoxic marine Lake Grevelingen in January also revealed similar subsurface enrichments in Fe, Mn and P (Fig. 8C). Importantly, no visual signals for cable bacteria based on the colour pattern of the sediment were observed at the time.

Macrofaunal activity likely counteracts or prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer at field sites. Bioturbation, i.e. mixing of the sediment, typically leads to oxidation from the sediment surface downwards (Norkko et al. 2012). Bioirrigation can efficiently pump O₂ into the pore water and thereby enhance the oxidation of dissolved Fe^{2+} (Kristensen et al. 2012; Norkko et al. 2012), but is not expected to lead to such a sharp oxidation front (Norkko et al. 2012; Hermans et al. 2019a). This is also evident from high-resolution elemental maps of the surface sediment from Lake Grevelingen in May, which shows the disappearance of the thin layer highly enriched in Fe and P formed by cable bacteria in January as a consequence of macrofaunal activity in May (Fig. 8D; Seitaj et al. 2015; Sulu-Gambari et al. 2016b).

We conclude that the focusing of Fe, Mn and associated P within a thin layer below the sediment-water interface is likely a consistent feature in sediments populated by active cable bacteria and may act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory experiments and at field sites, also in cases where visual observations are not conclusive. Focusing of Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and can also occur in the absence of cable bacteria. However, the upward fluxes of Fe^{2+} and Mn^{2+} in sediments populated by cable bacteria, are higher due to active dissolution of Fe and Mn minerals at depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame following an environmental perturbation (such as a transition to oxic bottom waters after a period of anoxia or mixing of the sediment), more Fe^{2+} and Mn^{2+} can oxidise upon contact with O₂ near the sediment-water interface and stronger enrichments of Fe and Mn minerals will be observed. Hence,
focusing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in sediments populated by active cable bacteria compared to sediments where no cable bacteria are active under such conditions. Macrofaunal activity within natural environments likely counteracts or prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer. When using standard techniques for sediment sampling (i.e. core slicing and chemical analysis of these slices), these layers may be missed due to the relatively coarse depth resolution. Hence, µXRF mapping of epoxy embedded sediment is recommended.

4.6. Cable Bacteria Activity at the Field Site

We can only speculate about the possible in-situ relevance of cable bacteria at the coastal site in the western Black Sea where the sediment for our incubation was collected. At this site, which is in a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m$^{-2}$) and polychaetes (up to ~1700 ind. m$^{-2}$) were observed at the time of sampling (Lenstra et al. 2019). Macrofauna can inhibit the activity of cable bacteria through bioturbation by physically cutting and damaging the filaments, rendering them unable to transport electrons (Malkin et al. 2014). Recent work has shown, however, that in some cases, cable bacterial communities can also thrive in sediments with macrofauna (Burdorf et al. 2017; Malkin et al. 2017; Aller et al. 2019). In a study of bivalve reefs, cable bacteria were found to efficiently remove highly toxic $\Sigma H_2S$, which is beneficial for bivalves (Malkin et al. 2017). Cable bacteria can also be abundant in bioturbated deposits, when associated with stable subdomains of the bioturbated zone, such as worm tubes (Aller et al. 2019). In such settings, a more complex precipitation pattern, e.g. along tube linings is observed (Aller et al. 2019), than described here for laboratory experiments with defaunated sediments and field sediments with an impoverished macrofaunal population (Fig. 8A). Further field studies are required to assess the role of cable bacteria at our field site, preferably including an assessment of the burrow structures.

Conclusions

The results of our laboratory incubation (with a total duration of 621 days) show that cable bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The strong acidity of the pore water associated with the activity of cable bacteria, which was monitored
using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a source of Fe that can form an Fe oxide-enriched surface layer. Both FeS and SO$_4^{2-}$ reduction provided the $\sum$H$_2$S required by cable bacteria to sustain their activity. Pore water $\sum$H$_2$S was always low (<5 µM). Using µXRF mapping of epoxy embedded sediment, we show that the activity of cable bacteria led to the development of a thin subsurface sediment layer (0.3 mm) that was highly enriched in Fe and P. The position of this layer in the sediment was directly proportional to the O$_2$ penetration depth during the experiment. We show that a similar layer highly enriched in Fe and P was also formed in sediments of field locations populated by cable bacteria (i.e. marine Lake Grevelingen and the brackish Gulf of Finland). We suggest that such layers, which are not necessarily visible by eye, may be used as a marker of cable bacteria activity in sediments with low macrofaunal activity.

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**Review statement.** This research was edited by Tina Treude and reviewed by two anonymous referees.
Code availability. All data, if not directly available from the tables and supplementary information, will be made available in the PANGAEA database. In the meantime data are available upon request to the main author.

Author contributions. MH and CPS designed the experiment. MH carried out the experiment and analysis. All authors interpreted the data. MH and CPS wrote the paper with comments provided by NRP and FJRM.

Competing interests. The authors declare that they have no conflict of interest.

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TABLES AND FIGURES

Table 1. Key site characteristics: latitude, longitude, water depth, bottom water O₂ concentration, in-situ O₂ uptake, in-situ O₂ penetration depth in the sediment, porosity and salinity. These data were retrieved from Lenstra et al. (2019). Our study site is station 9 in Lenstra et al. (2019).

<table>
<thead>
<tr>
<th>Black Sea (Station 9)</th>
<th>Unit</th>
</tr>
</thead>
<tbody>
<tr>
<td>Latitude</td>
<td>44°34.9' N</td>
</tr>
<tr>
<td>Longitude</td>
<td>29°11.4' E</td>
</tr>
<tr>
<td>Water depth</td>
<td>27 m</td>
</tr>
<tr>
<td>Bottom water O₂</td>
<td>92 µM</td>
</tr>
<tr>
<td>O₂ uptake</td>
<td>25.8 ± 1.77 mmol m⁻² d⁻¹</td>
</tr>
<tr>
<td>O₂ penetration depth</td>
<td>2.25 mm</td>
</tr>
<tr>
<td>Porosity</td>
<td>0.86 -</td>
</tr>
<tr>
<td>Salinity</td>
<td>17.881 -</td>
</tr>
<tr>
<td>Avg. organic carbon content (0-0.5 cm)</td>
<td>1.8%</td>
</tr>
</tbody>
</table>

Table 2. Mass balance of O₂ consumption. The diffusive uptake of O₂ as calculated from the O₂ depth profiles (column 1) was compared to the potential O₂ demand from the oxidation of NH₄⁺, Fe²⁺ and Mn²⁺ (column 2-4). The O₂ consumption of the oxidation of NH₄⁺, Fe²⁺ and Mn²⁺ was determined based on the stoichiometry of NH₄⁺, Fe²⁺ and Mn²⁺ oxidation with O₂ as described in Reed et al. (2011). The oxidation of dissolved Fe²⁺ and Mn²⁺ only played a minor role in the total O₂ consumption during the experiment, contributing only 0.9 to 3.8% and 0.1 to 0.4%, respectively.

<table>
<thead>
<tr>
<th>O₂ [mmol m⁻² d⁻¹]</th>
<th>Potential O₂ Demand</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>NH₄⁺ [mmol m⁻² d⁻¹]</td>
</tr>
<tr>
<td>Day 5</td>
<td>-23.35</td>
</tr>
<tr>
<td>Day 12</td>
<td>-23.24</td>
</tr>
<tr>
<td>Day 18</td>
<td>-21.10</td>
</tr>
<tr>
<td>Day 26</td>
<td>-23.00</td>
</tr>
<tr>
<td>Day 33</td>
<td>-22.80</td>
</tr>
<tr>
<td>Day 40</td>
<td>-19.60</td>
</tr>
<tr>
<td>Day 207</td>
<td>-6.90</td>
</tr>
<tr>
<td>Day 621</td>
<td>-3.25</td>
</tr>
</tbody>
</table>
Table 3. Mass balance of Fe. Time-series of the depth integrated (0-5 cm) increase in Fe oxides and the depth integrated (0-5 cm) depletion of FeS and FeCO$_3$ (siderite) in mmol m$^{-2}$. All values are reported in mmol Fe m$^{-2}$. Negative values represent a decrease, whereas positive values indicate an increase in the mineral pools.

<table>
<thead>
<tr>
<th></th>
<th>$\Delta$Fe oxides [mmol m$^{-2}$]</th>
<th>$\Delta$FeS [mmol m$^{-2}$]</th>
<th>$\Delta$FeCO$_3$ [mmol m$^{-2}$]</th>
</tr>
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<tbody>
<tr>
<td>Day 5</td>
<td>120</td>
<td>-90</td>
<td>-42</td>
</tr>
<tr>
<td>Day 12</td>
<td>170</td>
<td>-90</td>
<td>-126</td>
</tr>
<tr>
<td>Day 18</td>
<td>189</td>
<td>-105</td>
<td>-92</td>
</tr>
<tr>
<td>Day 26</td>
<td>276</td>
<td>-174</td>
<td>-99</td>
</tr>
<tr>
<td>Day 33</td>
<td>315</td>
<td>-176</td>
<td>-109</td>
</tr>
<tr>
<td>Day 40</td>
<td>412</td>
<td>-223</td>
<td>-200</td>
</tr>
<tr>
<td>Day 207</td>
<td>523</td>
<td>-236</td>
<td>-341</td>
</tr>
<tr>
<td>Day 621</td>
<td>874</td>
<td>-242</td>
<td>-566</td>
</tr>
</tbody>
</table>

Table 4. Sources of $\sum$H$_2$S calculated from the reduction of SO$_4^{2-}$ and the dissolution of FeS. The numbers are presented either as mmol m$^{-2}$ d$^{-1}$ or as the relative percentage of the $\sum$H$_2$S production. The amount of S from the dissolution of FeS was estimated from the upward diffusive flux of Fe$^{2+}$ (Fig. 6C) and the relative fraction of FeS (FeS/FeS+siderite) based on the mass balance calculations (Table 3).

<table>
<thead>
<tr>
<th></th>
<th>S from SO$_4^{2-}$ reduction [mmol m$^{-2}$ d$^{-1}$]</th>
<th>S from FeS dissolution [mmol m$^{-2}$ d$^{-1}$]</th>
<th>S from SO$_4^{2-}$ reduction [%]</th>
<th>S from FeS dissolution [%]</th>
</tr>
</thead>
<tbody>
<tr>
<td>Day 5</td>
<td>10.49</td>
<td>0.56</td>
<td>95%</td>
<td>5%</td>
</tr>
<tr>
<td>Day 12</td>
<td>17.60</td>
<td>1.48</td>
<td>92%</td>
<td>8%</td>
</tr>
<tr>
<td>Day 18</td>
<td>8.87</td>
<td>1.50</td>
<td>86%</td>
<td>14%</td>
</tr>
<tr>
<td>Day 26</td>
<td>11.15</td>
<td>1.61</td>
<td>87%</td>
<td>13%</td>
</tr>
<tr>
<td>Day 33</td>
<td>8.54</td>
<td>1.52</td>
<td>85%</td>
<td>15%</td>
</tr>
<tr>
<td>Day 40</td>
<td>7.57</td>
<td>1.25</td>
<td>86%</td>
<td>14%</td>
</tr>
<tr>
<td>Day 207</td>
<td>10.38</td>
<td>0.13</td>
<td>99%</td>
<td>1%</td>
</tr>
<tr>
<td>Day 621</td>
<td>5.36</td>
<td>0.03</td>
<td>99%</td>
<td>1%</td>
</tr>
</tbody>
</table>
**Fig. 1. (A)** Geochemical pore water fingerprint typical for cable bacteria activity. This fingerprint is defined by a distinct pH profile (light grey line) and a sub-oxic zone that is devoid of O$_2$ (red line) and H$_2$S (blue line). The cable bacteria filaments are depicted in yellow. On the background, the sediment core photograph, taken 278 days after the start of the experiment, shows a distinct colour zonation where (1) the oxic zone displays an orange colour (2) the suboxic zone has a grey colour and (3) the sulphidic zone has a black colour. The scale bar denotes a distance of 6 cm, with 0.5 cm intervals. **(B)** Bathymetric map of the Black Sea. The purple star indicates the location of our study site (44°34.93'N, 29°11.38’E), which was sampled with R/V Pelagia in September 2015. Further details are provided in Lenstra et al. (2019). **(C)** Volumetric density of cable bacteria [m cm$^{-3}$] in the top 2.5 cm of the sediment, for in-situ as well as for three time points during the incubation experiment **(D)** SEM image of a cable bacteria filament that was extracted from the surface sediment after 40 days.

**Fig. 2. (A)** Time-series of the pore water pH (black), O$_2$ (red) and $\Sigma$H$_2$S (blue) signatures of the incubated sediment. **(B)** Development of the EP depth profile in the incubated sediment over time. The dashed-line at 0 mm depth represents the sediment-water interface. The blue boxes indicate the overlying water, whereas the underlying light grey boxes represent the sediment. The EP depth profiles represent an average of 3 replicate measurements. The error bars indicate the minimum and maximum EP values that were observed. The orange depth profiles represent duplicate measurement performed on a different core.

**Fig. 3.** Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two different cores.

**Fig. 4.** Time-series of the **(A)** diffusive O$_2$ uptake in mmol O$_2$ m$^{-2}$ d$^{-1}$ and **(B)** current density as a consequence of long-distance electron transport (e$^-$) in mmol e$^-$ m$^{-2}$ d$^{-1}$ in the sediment incubation.

**Fig. 5.** Time-series of pore water depth profiles of NH$_4^+$ (orange), SO$_4^{2-}$ (purple), Fe$^{2+}$ (red), Mn$^{2+}$ (green), Ca$^{2+}$ (grey), HPO$_4^{2-}$ (blue) and H$_4$SiO$_4$ (yellow). The control core was sampled at day 621.

**Fig. 6.** Time-series of diffusive fluxes calculated from the linear gradient of the pore water profiles of **(A)** NH$_4^+$, **(B)** SO$_4^{2-}$, **(C)** Fe$^{2+}$, **(D)** Mn$^{2+}$, **(E)** Ca$^{2+}$ and **(F)** H$_4$SiO$_4$ in mmol m$^{-2}$ d$^{-1}$ towards the oxic zone of the sediment, based on the linear pore water gradients (Section 1.6; Fig. S3-S4). Here, a positive value indicates an upward flux, whereas a negative value represents a downward flux. N/A = not available. The control core was sampled at day 621.
Fig. 7. Time-series of solid-phase depth profiles of Fe oxides (red), FeS (black), siderite (grey), Mn oxides (green), metal bound P (blue) and metal oxide bound Si (yellow).

Fig. 8. High-resolution elemental maps of Fe (red), Mn (green), P (blue) and Ca (white) of surface sediments. These maps are shown in true vertical orientation and the colours accentuate the relative count intensities adjusted for brightness and contrast to highlight the features in the sediment. The tick marks represent 1 mm intervals. µXRF maps of the surface sediment (A) from the incubation experiment, (B) from the Gulf of Finland at site GOF5 in June (Hermans et al. Submitted), (C) from Lake Grevelingen in January (when cable bacteria become active) and (D) from Lake Grevelingen in May (showing the effects of bioturbation as described in Seitaj et al. (2015)).

Fig. 9. The relationship between the diffusive uptake of O\(_2\) (mmol O\(_2\) m\(^{-2}\) d\(^{-1}\)) and the current density of long-distance electron transport (mmol e\(^{-}\) m\(^{-2}\) d\(^{-1}\)). Red triangles are data for days 0 and 5. Green diamonds are data for all other time points. The blue line represents the expected correlation between the cathodic O\(_2\) consumption rate and the current density assuming a 1:4 ratio (Nielsen et al. 2010). Here, a positive value indicates an upward flux, whereas a negative value represents a downward flux.

Fig. 10. Time-series of the depth integrated (0-5 cm) increase in Fe oxides (red) and the depletion of FeS (black) and siderite (grey) in mmol m\(^{-2}\). Negative values represent a decrease, whereas positive values indicate an increase in the mineral pools.
**A**

![Bottom Water](image)

- **O₂**
- **pH**
- **H₂S**
- **Cable Bacteria**

**Reduction**: \( O₂ + 4 \text{e}^- + 4 H^+ \rightarrow H₂O \)

**Oxidation**: \( H₂S + 4 H₂O \rightarrow SO₄^{2-} + 10 H^+ + 8 \text{e}^- \)

**B**

Map showing the distribution of different parameters at various depths:

- **42°N**
- **44°N**
- **46°N**

**C**

**Volumetric Density [m cm⁻³]**

<table>
<thead>
<tr>
<th>St. 9 (in-situ)</th>
<th>Day 5</th>
<th>Day 26</th>
<th>Day 207</th>
</tr>
</thead>
<tbody>
<tr>
<td>0.0 0.5 1.0 1.5 2.0 2.5</td>
<td>0.0 0.5 1.0 1.5 2.0 2.5</td>
<td>0.0 0.5 1.0 1.5 2.0 2.5</td>
<td>0.0 0.5 1.0 1.5 2.0 2.5</td>
</tr>
</tbody>
</table>

**D**

Microscopic image of cable bacteria.
**Day 0**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 5**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 12**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 18**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 26**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 33**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 40**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 103**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 145**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 207**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 278**

- **EP [mV]**
- **Sediment Depth [cm]**

**Day 621**

- **EP [mV]**
- **Sediment Depth [cm]**
A. **Diffusive O₂ Uptake**

- **Day 0**: mmol O₂ m⁻² d⁻¹
- **Day 5**: mmol O₂ m⁻² d⁻¹
- **Day 12**: mmol O₂ m⁻² d⁻¹
- **Day 18**: mmol O₂ m⁻² d⁻¹
- **Day 26**: mmol O₂ m⁻² d⁻¹
- **Day 33**: mmol O₂ m⁻² d⁻¹
- **Day 40**: mmol O₂ m⁻² d⁻¹
- **Day 103**: mmol O₂ m⁻² d⁻¹
- **Day 145**: mmol O₂ m⁻² d⁻¹
- **Day 207**: mmol O₂ m⁻² d⁻¹
- **Day 278**: mmol O₂ m⁻² d⁻¹
- **Day 621**: mmol O₂ m⁻² d⁻¹

B. **Current Density**

- **Day 0**: mmol e⁻ m⁻² d⁻¹
- **Day 5**: mmol e⁻ m⁻² d⁻¹
- **Day 12**: mmol e⁻ m⁻² d⁻¹
- **Day 18**: mmol e⁻ m⁻² d⁻¹
- **Day 26**: mmol e⁻ m⁻² d⁻¹
- **Day 33**: mmol e⁻ m⁻² d⁻¹
- **Day 40**: mmol e⁻ m⁻² d⁻¹
- **Day 103**: mmol e⁻ m⁻² d⁻¹
- **Day 145**: mmol e⁻ m⁻² d⁻¹
- **Day 207**: mmol e⁻ m⁻² d⁻¹
- **Day 278**: mmol e⁻ m⁻² d⁻¹
- **Day 621**: mmol e⁻ m⁻² d⁻¹
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<th>Day 207</th>
<th>Day 621</th>
<th>Control</th>
</tr>
</thead>
<tbody>
<tr>
<td><img src="image" alt="NH₄⁺ concentration" /></td>
<td><img src="image" alt="SO₄²⁻ concentration" /></td>
<td><img src="image" alt="Fe²⁺ concentration" /></td>
<td><img src="image" alt="Mn²⁺ concentration" /></td>
<td><img src="image" alt="Ca²⁺ concentration" /></td>
<td><img src="image" alt="HPO₄²⁻ concentration" /></td>
<td><img src="image" alt="H₄SiO₄ concentration" /></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

- **NH₄⁺ [mM]**
- **SO₄²⁻ [mM]**
- **Fe²⁺ [mM]**
- **Mn²⁺ [mM]**
- **Ca²⁺ [mM]**
- **HPO₄²⁻ [mM]**
- **H₄SiO₄ [mM]**

**Depth [cm]**

0 8 16

N/A  N/A
Lake Grevelingen in January
Cable bacteria

Lake Grevelingen in May
Macrofauna

Black Sea Experiment
Cable bacteria

Gulf of Finland (GOF5)
Cable bacteria
Day 5  Day 12  Day 18  Day 26  Day 33  Day 40  Day 207  Day 621

mmol m⁻²

Fe oxides
FeS
Siderite