

**21 October 2020**

Dear Prof. Treude,

We now included the University of Helsinki in the affiliations, since I currently work here as a postdoctoral researcher and I forgot to include this in the version I submitted on October 19th. We also added more details to the other affiliations. Because of this, the line numbers that we refer to in our replies have been updated again in this version.

Kind regards,

Martijn Hermans

**19 October 2020**

Dear Prof. Treude,

We would like to thank you and the two anonymous referees for reviewing our paper and the insightful and constructive feedback. Please find our replies below where we explain how we addressed each comment and the changes that were made in our manuscript. Unless otherwise indicated, the line numbers in our replies refer to the revised manuscript. We include a marked-up version of our manuscript that highlights all the relevant changes that were made, and we now uploaded the high-resolution version of the figures at the end of our manuscript.

Kind regards,

Martijn Hermans

Prof. Treude's comments:

Instead of 'bottom water' you could also use the term "supernatant" or 'bottom-near" water. I was in the past criticized by physical oceanographers for using the term 'bottom water' in my sediment studies, because it is a fixed term in oceanography and covers a large volume (an entire water mass).

**Reply:** We prefer to use “bottom water” since for chemical oceanographers this is a common term. We now made more explicit what bottom water refers to see line **No. 35**: “Depletion of oxygen ( $O_2$ ) in bottom waters (i.e. water directly above the seafloor)”

- The term "suboxic" opposite to oxic and anoxic is not ideal, because suboxic is also anoxic. But I am aware that the term is frequently used together with cable bacteria biogeochemistry. I am OK with keeping the term, but please double check that your final manuscript has a proper definition of the terms at the beginning (oxic = with  $O_2$ , suboxic = anoxic but not sulfidic, anoxic = anoxic and sulfidic).

**Reply:** We explain this in more detail in the revised version of our manuscript, see line **No. 54-55**: “This spatial coupling of surficial  $O_2$  reduction with  $H_2S$  oxidation at several centimetres depth creates a suboxic zone that is devoid of any  $O_2$  and  $H_2S$ ”

- In Fig. 2B a red and a green profile are shown in the same graphs. Please use a different color to allow green-red handicapped people to enjoy it.

**Reply:** We updated the figure and changed the colours to blue and orange, so it can be read by colour blind people as well.

## **Anonymous Referee #1**

Received and published: 13 August 2020

The study documents the effect of cable bacteria on sediment geochemistry in repacked sediment cores on sediment collected from the Black Sea. The findings support current paradigms in the cable bacteria literature. This is a nice case study that has been well presented and written up. The dominance of cable bacteria in the oxygen budget is an interesting finding. There is also some additional data on element mapping that sheds further light on the effect of cable bacteria on sediment geochemistry at different sites. I only have a few relatively minor comments and suggestions for improvement.

**Reply:** We thank Anonymous Referee #1 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

**Comment #1** The authors used the classic sediment repacking method to start a cable growth cycle. What effect might homogenisation have had on the final findings? For example, would siderite be likely to be so close to the surface under normal circumstances.

**Reply:** The homogenisation likely did not have a significant effect on our final findings. Homogenisation of the sediment is essential to obtain proper replicate cores and is known to induce growth of cable bacteria. While some of the FeS and siderite might have oxidised during sieving, homogenising and repacking, this does not affect the conclusions of our experiment. There is still ample FeS and siderite present in our sediment cores at the start of the experiment, and our time-series indicate that both FeS and siderite were dissolved by cable bacteria activity over time. Both FeS and siderite are frequently observed in marine surface sediments (e.g. Sulu-Gambari et al. 2016).

**Comment #2** Line 163 – no need to say ‘bottom water’ just water samples?

**Reply:** Here, we specifically use the term bottom water samples, since it refers to the overlying water in the cores. Water samples could also refer to water column or pore water samples. Therefore, we think it is best to use the term bottom water samples, to avoid potential confusion.

**Comment #3** Line 172 – please elaborate a little on exactly what you mean by salt corrections.

**Reply:** Freeze-drying removes the water from samples. However, the salt stays behind in the dried sediment. Hence, the weight of the freeze-dried sediment used for the sequential extractions needs to be corrected for this salt content if we wish to calculate elemental/mineral concentrations, such as Fe oxide and metal bound P contents per gram of sediment. Without the salt correction the absolute

elemental/mineral concentrations would be underestimated. Hence, we subtract the weight of the salt from the freeze-dried sediment to calculate the ‘real’ weight of the dry sediment. This is now explained in more detail in our methods, see line **No. 178-181**: “After freeze-drying, the salt from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried sediment.”

**Comment #4** Line 205 – could you add a sentence on how the embedding was achieved?

**Reply:** The resin embedding process is described in more detail, see line **No. 212-217**: “On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013). Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled glovebox and subsequently embedded with Spurr’s epoxy resin as described in Jilbert et al. (2008). After curing, the epoxy-embedded core was split vertically using a rock saw.”

**Comment #5** Line 210 onwards – consider adding this to methods.

**Reply:** The methods used to obtain these elemental maps from the epoxy resin embedded surface sediments from the Gulf of Finland and Lake Grevelingen are described in detail in the other studies cited (Sulu-Gambari et al. 2016; Sulu-Gambari et al. 2018; Hermans et al. Submitted). Therefore, we prefer not to describe the sampling process of those resin embedded cores in section 2.4 of our methodology.

**Comment #6** also Line 261 – Only Ca and Si fluxes are presented, I couldn’t see them?

**Reply:** These Ca and Si fluxes are presented in Fig. S10 and Table S4 in the Supplementary Information, see line **No. 488** in the manuscript.

**Comment #7** Line 384 not clear what you mean here, please elaborate.

**Reply:** We have used Fick’s law for the calculation of the diffusive fluxes, which not does take the effect of the electric field generated by cable bacteria on the diffusion potential into account. The Nernst-Planck equation, however, extends Fick’s law, because solutes can also be moved with respect to the fluid by electrostatic forces. By using Fick’s law, we underestimate the  $\text{SO}_4^{2-}$  reduction rate by a most ~10-20%. We will describe this in more detail, see line **No. 396-400**: “Solute can also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the relatively low strength of the electric field in the cores ( $<0.073\text{V m}^{-1}$  at day 18; as estimated from Fig 2B), including the contribution of ionic drift to the sulphate flux would lead to  $\text{SO}_4^{2-}$  reduction rates that are at most 10-20% higher.”

**Comment #8** Line 415 – could it be that the nitrate is just denitrified? Also on this point, it seems that not flux measurements were made for nitrate. It seems likely that some nitrate is released to the water. It might be worth a brief discussion of a few scenarios here. All the nitrate is released to the water column, all the nitrate is denitrified by sediment bacteria and all the nitrate is denitrified by cable bacteria.

**Reply:** Unfortunately, we do not have data for  $\text{NO}_3^-$ . The problem of the abovementioned scenarios is that, if we would include the role of other groups of bacteria, and the potential release of  $\text{NO}_3^-$  to the water column, the mass balance for  $\text{O}_2$  would have an even greater mismatch. When looking at the stoichiometry of  $\text{NO}_3^-$  by cable bacteria and the conversion of N to  $\text{N}_2$ , we cannot close the budget fully that way. The text is now modified and explains this in more detail, see line **No. 428-433**: “These findings can be explained, however, if we assume that at least part of the  $\text{NO}_3^-$  that is being formed near the sediment-water interface is also used for the metabolic activity of cable bacteria. It has been shown that cable bacteria can couple the oxidation of  $\sum\text{H}_2\text{S}$  to  $\text{NO}_3^-$  in the absence of  $\text{O}_2$  (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments where  $\text{O}_2$  is present in concert with  $\text{NO}_3^-$  near the sediment-water interface. However, we cannot exclude release of  $\text{NO}_3^-$  to the water column or denitrification by other bacteria in the sediment.”

**Comment #9** Line 485 – very interesting!

**Reply:** Thank you, this potential niche for vivianite formation is indeed an interesting finding.

**Comment #10** Line 522 – not obvious to me from Fig 8A, it is interesting, can you make this clearer?

**Reply:** This is now more explicit in the text, see line **No. 528-531**: “While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P is strongly correlated with Ca.”

**Comment #11** Line 546 I agree this is likely driven by cables, but how is this different from a straight reaction diffusion scenario (given ubiquity of cables, such a scenario does seem unlikely though). I think this idea needs a little more development and explanation as to how it might actually be applied.

**Reply:** Focusing of Fe and Mn oxides and associated P can indeed also occur in sediments overlain by oxic waters, where no cable bacteria are active. However, as demonstrated by our experiment (and various other studies (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016), in sediment populated by active cable bacteria, the upward fluxes of  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  are higher due to the dissolution of FeS, Fe- and Mn carbonates. This allows strong focusing of Fe and Mn oxides in a thin layer within a relatively short time

frame. We added the following section in our manuscript, see line **No. 571-581**: “Focussing of Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and can also occur in the absence of cable bacteria. However, the upward fluxes of  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  in sediments populated by cable bacteria are higher due to active dissolution of Fe and Mn minerals at depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame following an environmental perturbation (such as a transition to oxic bottom waters after a period of anoxia or mixing of the sediment), more  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  can oxidise upon contact with  $\text{O}_2$  near the sediment-water interface and thus stronger enrichments of Fe and Mn minerals will be observed. Hence, focussing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in sediments populated by active cable bacteria compared to sediments where no cable bacteria are active under such conditions.”

**Comment #12** Figure 3 not clear how this was generated. Based on the pictures?

**Reply:** The depth intervals of the oxic, suboxic and anoxic zone are based on the micro-electrode data. We made this more explicit in the caption of Fig. 3., see line **No. 924-926**: “Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two different cores.”

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- Sulu-Gambari, F., Seitaj, D., Behrends, T., Banerjee, D., Meysman, F. J., and Slomp, C. P. (2016). Impact of cable bacteria on sedimentary iron and manganese dynamics in a seasonally-hypoxic marine basin. *Geochimica et Cosmochimica Acta* **192**: 49-69.

## Anonymous Referee #2

Received and published: 31 August 2020

The manuscript represents a very comprehensive study of potential processes and effects of cable bacteria in sediments. Investigations on cable bacteria and their influence on biogeochemical processes are still in the beginning, but more and more studies show their importance for the element cycling; importance of cable bacteria activity on the oxygen demand in coastal sediments. In the present study, the authors used sediment cores from the coastal area of the Black Sea, which they homogenized and freed from macrofauna. This probably increased the availability of labile organic material and its distribution in deeper sediment layers. Furthermore, the sediment was anoxically stored until the experiment, during the experiment the overlying bottom water was saturated with oxygen so that a steady state must be established at the beginning of the incubations. This fact does not reduce the results of the experiment or the quality of the manuscript.

**Reply:** We thank Anonymous Referee #2 for reviewing our paper and the insightful and constructive feedback. Please find our replies to each comment below.

**Comment #1** However, the authors should consider the study presented here as potential processes and not directly related to a coastal region (in this case the Black Sea). Therefore, I would strongly suggest to rewrite the manuscript and change the focus of the manuscript by concentrating on the "potential processes and bio geochemical impacts" rather than to directly relate it to coastal sediments of the Black Sea.

**Reply:** We are aware that the outcomes from our experiment are potential processes, which cannot be directly translated to the field site. This is why, in our title we specifically used the term "on coastal Black Sea sediment" instead of "in coastal Black Sea sediments". Other examples of sentences in our manuscript (here these line numbers refer to the old version of our manuscript) that indicate that we are not directly relating our results to a coastal region are:

- **Line No. 100:** "In this study, we assess whether cable bacteria activity can establish in sediments that are relatively poor in FeS in an incubation experiment using siderite-bearing sediments from a coastal site in the Black Sea."
- **Line No. 552:** "We can only speculate about the possible *in-situ* relevance of cable bacteria at the coastal site in the western Black Sea where the sediment for our incubation was collected."
- **Line No. 565-566:** "Further field studies are required to assess the role of cable bacteria at our field site, preferably including an assessment of the burrow structures."



We added the following additional text in the abstract, introduction and conclusion sections to further emphasise that we are referring to potential processes.

Abstract; line No. 17-18: “to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and sulphur (S).”

Introduction; line No. 103-106: “In this study, we assess whether cable bacteria activity can establish and thrive in sediments that are relatively poor in FeS. Although, this will be done in a controlled incubation experiment with siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural environments populated by cable bacteria.”

Conclusion; line No. 603-609: “The results of our laboratory incubation (with a total duration of 621 days) show that cable bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The strong acidity of the pore water associated with the activity of cable bacteria, which was monitored using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a source of Fe that can form an Fe oxide-enriched surface layer.”

**Comment #2** The difference between the natural distribution of cable bacteria and the experiment is also evident when looking at Fig. 1c. The authors can use their main results as shown here, but the focus should be on the conditions used in their experiment, which are rather artificial, but very nicely show the potential of cable bacteria in the biogeochemical cycling.

**Reply:** See our reply to comment #1 and the associated changes in the text. We are aware that the outcomes of our experiment regarding the impact of cable bacteria are amplified when compared to field conditions and that we cannot directly link this to the field site. This is because we optimized conditions to sustain metabolic activity of cable bacteria, and their growth:

- 1) The sediment was homogenised, which is known to induce growth of cable bacteria.
- 2) There was no bioturbation by meiofauna and macrofauna.
- 3) The bottom water was continuously oxygenated.

This optimisation was deliberately done to study the effects of cable bacteria on sediment biogeochemistry. We note that the approach used is common in studies of the biogeochemical impact of cable bacteria (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016).

**Comment #3** In a second step the transfer to coastal sediments and their biogeochemical conditions can be done. Here the manuscript lacks the coherence (hypoxia and oxy-gen depletion as mentioned in the Introduction). In a final paragraph the transfer of the laboratory experiment to natural sediments and possible variations in biogeochemical processes as well as the influence of macrofauna (bioturbation and bioirrigation) can be discussed.

**Reply:** Sections 4.1 to 4.4 focus only on the experiment. In sections 4.5 and 4.6 we discuss the implications for the field. We added text to section 4.5 and 4.6 to clarify when we are referring to other laboratory experiments and results of field studies and the link with hypoxia.

**Line No. 524-526:** “This colour zonation is typical for sediments that have been geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016)”

**Line No. 569-571:** “may act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory experiments and at field sites, also in cases where visual observations are not conclusive.”

**Line 581-582:** “Macrofaunal activity within natural environments likely counteracts or prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer.”

**Line 588-590:** “At this site, which is in a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m<sup>-2</sup>) and polychaetes (up to ~1700 ind. m<sup>-2</sup>)...”

**Comment #4** Is there any information about the organic carbon content of the sediment and how this changes over the incubation period? I would assume that this is the major driver for the development of biogeochemical zonation.

**Reply:** We now include the organic carbon content of the upper 0-0.5 cm of the sediment at the field site, as determined by Lenstra et al. (2019), in Table 1. We did not determine the change in organic carbon contents during the experiment because, at the typical rates of organic matter degradation expected here, only a small change in organic carbon content would be observed. Hence, we chose to focus on pore water NH<sub>4</sub><sup>+</sup> profiles to obtain insight in rates of (anaerobic) organic matter degradation.

**Comment #5** How does the development of the oxic zone, as shown in the experiment, relate to natural variations in coastal sediments ?

**Reply:** The range in O<sub>2</sub> penetration observed in the experiment is comparable to that observed in coastal systems with seasonal hypoxia (e.g. Seitaj et al. 2015). We included this in the manuscript: Line No. 535-538: “During the experiment, O<sub>2</sub> penetration varied within a narrow range and was initially fixed between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth of 2 mm (Fig. 8A). Such a range in O<sub>2</sub> penetration is in accordance with observations in coastal sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by...”

**Comment #6** How does the experiment relate to the development of hypoxia and depletion of oxygen in coastal areas ? The experiment shows the opposite reaction (from anoxic surface layer to an oxygenated layer).

**Reply:** As shown in previous work that we refer to in our manuscript (Seitaj et al. 2015; Sulu-Gambari et al. 2016), cable bacteria can induce formation of Fe and Mn oxides in seasonally hypoxic coastal systems during periods of oxygenated bottom waters in spring. As explained in line No. 44-46, the presence of these Fe and Mn oxides can delay the transition towards euxinia by removing hydrogen sulfide. Our experimental setting can be compared to the onset of bottom water re-oxygenation in spring in such seasonally hypoxic environments (i.e. the sediment was stored anoxically, and then exposed to oxygenated water). This allows study of the mineral dissolution and formation reactions in sediments populated by cable bacteria under such conditions, which is the goal of this work.

**Comment #7** line2 121/122: .... with overlying water ....Was this bottom water taken from the site or artificial water, as used for the aquarium?

**Reply:** The 20 cm long cores (filled with 15 cm sediment) were gently submerged in the two aquaria. Hence, the overlying water in the cores at the start of each incubation was the same as the artificial water used in the aquaria.

**Comment #8** line 153: ..... core was place outside the aquarium .....Why was the core taken out ? was the incubation temperature maintained?

**Reply:** The entire experiment was carried out at in a temperature controlled laboratory at 20 °C. We made explicit that the flux incubations also took place at 20 °C, see line No. 156-157: “At each time point, one core was placed outside the aquarium at 20 °C...”

**Comment #9** Was the overlying water during the 24-hour incubation for the solute flux measurements stirred to avoid stratification? This could have influenced the flux across the sediment-water interface

because stagnant waters lead to an increase of the Diffusive Boundary Layer, which controls the solute exchange.

**Reply:** The overlying water was mixed continuously by bubbling it continuously with air. The airflow was set in such a way that the water was effectively mixed, while the surface layer of the sediment was left undisturbed. We made this more explicit in the methods, see line **No. 156-160**: “At each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of overlying water in the core was continuously aerated. Potential stratification of the overlying water was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent evaporation.”

**Comment #10** Pore water profiles (especially Fig 1a, Fig 2) are very small and it is difficult to recognize the different profiles (O<sub>2</sub>, pH, H<sub>2</sub>S) different; graphs should be enlarged.

**Reply:** We enlarged these graphs and we increased the font sizes to improve the readability of the figures.

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# 1 Biogeochemical Impact of Cable Bacteria on Coastal Black Sea Sediment

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## 15 ABSTRACT

16 Cable bacteria can strongly alter sediment biogeochemistry. Here, we used laboratory  
17 incubations to determine the potential impact of their activity on the cycling of Fe, phosphorus (P) and  
18 sulphur (S). Microsensor depth profiles of oxygen, sulphide and pH in combination with electric  
19 potential profiling and FISH analyses showed a rapid development (<5 days) of cable bacteria,  
20 followed by a long period of activity (>200 days). During most of the experiment, the current density  
21 correlated linearly with the oxygen demand. Sediment oxygen uptake was attributed to activity of  
22 cable bacteria and the oxidation of reduced products from anaerobic degradation of organic matter,  
23 such as ammonium. Pore water sulphide was low (<5  $\mu\text{M}$ ) throughout the experiment. Sulphate  
24 reduction acted as the main source of sulphide for cable bacteria. Pore water  $\text{Fe}^{2+}$  reached levels of up  
25 to 1.7 mM during the incubations, due to the dissolution of FeS (30%) and siderite, an Fe carbonate  
26 mineral (70%). Following upward diffusion of  $\text{Fe}^{2+}$ , a surface enrichment of Fe oxides formed. Hence,

27 besides FeS, siderite may act as a major source of Fe for Fe oxides in coastal surface sediments where  
28 cable bacteria are active. Using  $\mu$ XRF, we show that the enrichments in Fe oxides induced by cable  
29 bacteria are located in a thin subsurface layer of 0.3 mm. We show that similar subsurface layers  
30 enriched in Fe and P are also observed at field sites where cable bacteria were recently active and  
31 little bioturbation occurs. This suggests that such subsurface Fe oxide layers, which are not always  
32 visible to the eye, could potentially be a marker for recent activity of cable bacteria.

33 Key words: cable bacteria, elemental cycling, solute fluxes, iron

## 34 **1. INTRODUCTION**

35 Depletion of oxygen ( $O_2$ ) in bottom waters (i.e. water directly above the seafloor) of coastal areas  
36 is increasing worldwide, as a consequence of eutrophication and climate change (Diaz and Rosenberg  
37 2008; Schmidtko et al. 2017; Breitburg et al. 2018). Low  $O_2$  can lead to the development of coastal  
38 ‘dead zones’ characterised by recurrent mortality of marine life (Rabalais et al. 2002; Diaz and  
39 Rosenberg 2008). Progressive eutrophication induces a characteristic response of coastal systems with  
40 transient and seasonal hypoxia ( $O_2 < 63 \mu M$ ) transitioning into permanent anoxia ( $O_2 = 0 \mu M$ ). In this  
41 later stage, free sulphide ( $H_2S$ ) may escape from the sediment and accumulate in the bottom water, a  
42 condition referred to as euxinia (Diaz and Rosenberg 2008; Kemp et al. 2009; Rabalais et al. 2014).  
43 As  $H_2S$  is highly toxic to higher fauna, the development of euxinia may aggravate the ecological  
44 consequences. However, the presence of iron (Fe) and manganese (Mn) oxides in surface sediments  
45 may delay this transition towards euxinia by removing  $H_2S$  and thus preventing an efflux of  $H_2S$  to  
46 the overlying water (Kristiansen et al. 2002; Kristensen et al. 2003; Diaz and Rosenberg 2008).

47 Cable bacteria are multicellular filamentous sulphur(S)-oxidising bacteria (Pfeffer et al. 2012)  
48 that strongly enhance the formation of Fe and Mn oxides and efficiently remove  $H_2S$  from surface  
49 sediments (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Sulu-Gambari et al. 2016a). Cable  
50 bacteria belong to the *Desulfobulbaceae* family of the Deltaproteobacteria (Trojan et al. 2016;  
51 Kjeldsen et al. 2019). Cable bacteria can spatially link the oxidation of  $H_2S$  in deeper sediments to the  
52 reduction of  $O_2$  near the sediment-water interface by transporting electrons over centimetre scale

53 distances (Pfeffer et al. 2012) through a conductive fibre network that is embedded in the cell  
54 envelope (Meysman et al. 2019). This spatial coupling of surficial O<sub>2</sub> reduction with H<sub>2</sub>S oxidation at  
55 several centimetres depth creates a suboxic zone that is devoid of any O<sub>2</sub> and H<sub>2</sub>S, and provides cable  
56 bacteria a competitive advantage over other S-oxidising bacteria in aquatic environments (Meysman  
57 2018). Cable bacteria have been documented in a range of fresh water (Risgaard-Petersen et al. 2015;  
58 Müller et al. 2016) and marine environments (Malkin et al. 2014; Burdorf et al. 2017), however, they  
59 appear to be particularly active in sediments overlain by seasonally hypoxic bottom waters (Seitaj et  
60 al. 2015; Burdorf et al. 2018).

61 The metabolic activity of cable bacteria establishes an electrical circuit in the sediment, which  
62 involves an electron current through the cable bacteria filaments (Bjerg et al. 2018), and an ionic  
63 current through the pore water in the opposite direction (Naudet and Revil 2005; Revil et al. 2010;  
64 Risgaard-Petersen et al. 2012). As a consequence, an electric potential (EP) is generated in the  
65 sediment, which can be used as a reliable indicator for activity of cable bacteria (Risgaard-Petersen et  
66 al. 2014).

67 Cable bacteria activity additionally generates a distinct biogeochemical signature, that can be  
68 assessed by pH, O<sub>2</sub> and H<sub>2</sub>S depth profiling (Nielsen et al. 2010). Their activity leads to the  
69 development of a suboxic zone (i.e. a zone where O<sub>2</sub> and H<sub>2</sub>S are both absent), and also induces a pH  
70 profile that strongly changes with depth. Cathodic O<sub>2</sub> reduction ( $O_2 + 4H^+ + 4e^- \rightarrow 2H_2O$ ) in the oxic  
71 zone of the sediment results in a pH maximum (~9) due to proton consumption, whereas anodic  
72 sulphide oxidation ( $H_2S + 4H_2O \rightarrow SO_4^{2-} + 10H^+ + 8e^-$ ) causes a pH minimum (<6.5) in the anoxic  
73 zone (Fig. 1A; Nielsen et al. 2010; Meysman et al. 2015).

74 The presence of cable bacteria in sediments can strongly impact the elemental cycling of Fe,  
75 Mn, Ca and S (Risgaard-Petersen et al. 2012; Seitaj et al. 2015; Rao et al. 2016; Sulu-Gambari et al.  
76 2016a; van de Velde et al. 2016). Pore water acidification induced by cable bacteria activity can lead  
77 to dissolution of calcium (Ca) carbonates, Fe carbonates (siderite), Mn carbonates and FeS in the zone  
78 where the pH is low, thus generating high concentrations of Fe<sup>2+</sup> and Mn<sup>2+</sup> in the pore water



79 (Risgaard-Petersen et al. 2012; Rao et al. 2016). When these dissolved species diffuse upward this can  
80 lead to strong enrichments of Fe and Mn oxides upon contact with O<sub>2</sub>, or for dissolved Fe<sup>2+</sup>, also upon  
81 contact with Mn oxides (Wang and Van Cappellen 1996; Seitaj et al. 2015; Sulu-Gambari et al.  
82 2016a). These metal oxides are capable of efficiently buffering the benthic release of H<sub>2</sub>S and  
83 phosphate (HPO<sub>4</sub><sup>2-</sup>) during periods with low bottom water O<sub>2</sub>. This so-called ‘firewall’ for H<sub>2</sub>S and  
84 alteration of the timing of HPO<sub>4</sub><sup>2-</sup> release linked to this buffering can play a key role in regulating  
85 water quality in seasonally hypoxic coastal systems (Seitaj et al. 2015; Sulu-Gambari et al. 2016b;  
86 Hermans et al. 2019a).

87 In coastal sediments, O<sub>2</sub> typically penetrates to a depth of only several mm’s below the  
88 sediment-water interface (Rasmussen and Jørgensen 1992; Rabouille et al. 2003; Glud 2008). This  
89 also holds true for sediments inhabited by active cable bacteria (Nielsen et al. 2010; Pfeffer et al.  
90 2012; Larsen et al. 2015). Hence, the oxidation of upward diffusing Fe<sup>2+</sup> and Mn<sup>2+</sup> is expected to take  
91 place below and not at the sediment-water interface. We hypothesise that, as a consequence, in the  
92 initial stages of cable bacteria activity and in the absence of bioturbation, most Fe and Mn oxide  
93 enrichments will be restricted to a thin subsurface layer of the sediment. However, the sample  
94 resolution and timing of the collection of solid phase data in field and laboratory studies published so  
95 far do not allow an assessment of this hypothesis.

96 Cable bacteria are suggested to thrive in coastal sediments characterised by high rates of H<sub>2</sub>S  
97 production due to high rates of organic matter mineralisation (Malkin et al. 2014; Burdorf et al. 2017;  
98 Hermans et al. 2019a). Laboratory and model studies have shown that the dissolution of FeS accounts  
99 for 12 to 94% of the H<sub>2</sub>S consumed by cable bacteria, while the other source is H<sub>2</sub>S production from  
100 the reduction of SO<sub>4</sub><sup>2-</sup> (Risgaard-Petersen et al. 2012; Meysman et al. 2015; Burdorf et al. 2018). At  
101 present, it is not known if cable bacteria activity can establish in sediments that are relatively low in  
102 FeS and dissolved H<sub>2</sub>S.

103 In this study, we assess whether cable bacteria activity can establish and thrive in sediments  
104 that are relatively poor in FeS. Although, this will be done in a controlled incubation experiment with

105 siderite-bearing sediments from a coastal site in the Black Sea, our findings are relevant for natural  
106 environments populated by cable bacteria. The metabolic activity of cable bacteria is monitored using  
107 microsensor profiles of pH, O<sub>2</sub>, H<sub>2</sub>S and EP. We also use sediment Fe and P speciation and  $\mu$ XRF of  
108 resin-embedded sediments to test whether we find evidence for subsurface enrichments in Fe oxides  
109 and associated P. We find a rapid establishment of cable bacteria (<5 days) and the development of an  
110 Fe oxide-rich subsurface layer, with the majority of the Fe ~70% supplied through dissolution of  
111 siderite induced by cable bacteria activity. The depth of the Fe oxide layer was directly related to the  
112 O<sub>2</sub> penetration depth and we propose that such subsurface enrichments in Fe, which also can contain P  
113 and Mn, can be used as a marker for recent cable bacteria activity.

## 114 **2. METHODS AND MATERIALS**

### 115 **2.1. Study Area and Experimental Set-up**

116 In September 2015, 16 sediment cores ( $\varnothing$ 10 cm) were retrieved at a coastal site on the north-  
117 western shelf of the Black Sea (27 m water depth; Fig. 1B; Table 1) using a multicorer (Oktopus  
118 GmbH, Germany) as described in Lenstra et al. (2019). The overlying water was discarded, and the  
119 upper 10 cm of the sediment was transferred into nitrogen purged aluminium bags that were sealed  
120 and stored at 4 °C for several months. The anoxic storage is expected to have led to the death of all  
121 macrofauna and most meiofauna (Coull and Chandler 2001; Riedel et al. 2012). Prior to incubation,  
122 the sediment was passed through a 4 mm sieve to remove large debris and homogenised.  
123 Subsequently, the sediment was transferred to 18 transparent polycarbonate cores ( $\varnothing$ 6 cm; 20 cm  
124 length).

125 The bottom 15 cm of these cores was filled with sediment and the upper 5 cm with overlying  
126 water. The cores were placed in two aquaria filled with artificial seawater (Instant Ocean Sea Salt +  
127 Ultra High Quality (UHQ) water) with a salinity of 17.9, identical to the bottom water salinity at the  
128 study site. The artificial sea water contained negligible concentrations of NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, Fe, Mn and P as  
129 described in Atkinson (1997) and Hovanec and Coshland (2004). The aquaria were kept in the dark at  
130 a constant temperature (~20 °C), and the water was continuously aerated by two aquarium pumps.

131 Sixteen out of eighteen cores were exposed to oxygenated overlying water in the aquaria, whereas the  
132 two remaining cores served as an anoxic control treatment. The control cores were tightly sealed with  
133 rubber stoppers, to prevent the growth of cable bacteria by excluding O<sub>2</sub> (Nielsen et al. 2010).

134 Sampling for pore water and solid-phase analyses was performed at eight time points over a  
135 total incubation period of 621 days. Each time point involved a three day procedure. On the first day,  
136 microsensor depth profiles of EP, O<sub>2</sub>, pH and H<sub>2</sub>S were obtained in two randomly selected oxic cores  
137 and the two anoxic control cores (O<sub>2</sub> profiling was not performed in the anoxic cores). On the second  
138 day, solute fluxes were measured in the same oxic cores that were used for microsensor depth  
139 profiling on the previous day. On the third day, the two cores were sectioned, of which only one core  
140 was processed further for pore water and solid-phase analyses. Photographs were taken at four time  
141 points (day 12; 33; 170 and 621) from one oxic core to follow the visual development of the surface  
142 sediment during the experiment.

## 143 **2.2. High-resolution Microsensor Depth Profiling**

144 High-resolution depth profiles of pH, O<sub>2</sub> and H<sub>2</sub>S were obtained (50- $\mu$ m depth resolution; 3  
145 replicate profiles per oxic core; 2 replicate profiles per anoxic core) using commercial micro  
146 electrodes (Unisense A.S., Denmark). The O<sub>2</sub> sensor was re-calibrated prior to each measurement,  
147 using saturated bottom water (100% [O<sub>2</sub>]) and the deeper sediment horizons (0% [O<sub>2</sub>]) as calibration  
148 points. Calibrations of the pH and H<sub>2</sub>S electrodes were performed as described in Hermans et al.  
149 (2019b). pH values are reported on the total scale. For depth profiling of EP (500- $\mu$ m resolution; 3  
150 replicates per core), micro electrodes were used that were custom built at Aarhus University, as  
151 described in Damgaard et al. (2014). A robust reference electrode (Ref-RM, Unisense, A.S.,  
152 Denmark) was used during EP and pH measurements. To exclude turbulence-induced variations in the  
153 potential of the reference electrode during EP profiling, a silicon tube filled with foam was mounted  
154 on the tip of the reference electrode.

### 155 **2.3. Solute Flux Measurements**

156 Solute flux incubations were performed for  $\text{NH}_4^+$ ,  $\text{Fe}^{2+}$ ,  $\text{Mn}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{HPO}_4^{2-}$  and  $\text{H}_4\text{SiO}_4$ . At  
157 each time point, one core was placed outside the aquarium at 20 °C, and the isolated volume of  
158 overlying water in the core was continuously aerated. Potential stratification of the overlying water  
159 was prevented by actively bubbling it. Parafilm was wrapped on top of the cores to prevent  
160 evaporation. Water samples of 3 mL were retrieved at 7 time points over 24 hours. The same volume  
161 of fresh artificial seawater was added to the cores directly after taking each sample. The samples were  
162 filtered (0.45  $\mu\text{m}$ ), and subsamples were taken for ammonium (1 mL) and for metals (1 mL; acidified  
163 with 10  $\mu\text{L}$  Suprapur® HCl (35%) per mL sample), which were stored at -20°C and 4°C respectively  
164 until further analysis.

### 165 **2.4. Pore Water and Sediment Collection**

166 At each time point, two cores were sectioned at 0.5-1 cm resolution with an UWITEC push-  
167 up pole in a nitrogen-purged glovebag, but only samples for one core were used for sediment and pore  
168 water collection and analyses. Bottom water samples were retrieved from the overlying water in the  
169 cores. Slices for each depth interval were centrifuged at 3500 rpm for 20 minutes for pore water  
170 retrieval. Samples (1 mL) for  $\text{NH}_4^+$  were taken and stored at -20°C until analysis. Samples (1 mL) for  
171 pore water S, Fe, Mn, Ca, P and Si were also collected and acidified with 10  $\mu\text{L}$  Suprapur® HCl  
172 (35%) per mL sample, which were stored at 4°C until analysis. Centrifuged sediment samples were  
173 freeze-dried and ground to a fine powder in a nitrogen-purged glovebox under a strictly anoxic  
174 environment to prevent oxidation (Kraal et al. 2009; Kraal and Slomp 2014). Only the top 5 cm of the  
175 solid-phase samples were analysed in further detail. The porosity (Supporting Information 1.1; Table  
176 S1) was calculated from the weight loss upon freeze-drying, using a sediment density of 2.65  $\text{g cm}^{-3}$   
177 (Burdige 2006). Salt corrections were performed on the solid-phase data using the gravimetric water  
178 content and salinity to determine the amount of salt after freeze-drying. After freeze-drying, the salt  
179 from seawater stays behind in the solid-phase fraction. To determine the actual weight of the dry  
180 sediment, it is necessary to subtract the weight of the salt from the total weight of freeze-dried  
181 sediment.

## 182 **2.5. Chemical Analysis of the Water and Sediment**

183 Concentrations of  $\text{NH}_4^+$  in the pore water and solute flux samples were determined using the  
184 phenol hypochlorite method (Koroleff 1969). The total Fe, Mn, Ca, P and Si concentrations (which  
185 are assumed to represent  $\text{Fe}^{2+}$ ,  $\text{Mn}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{HPO}_4^{2-}$  and  $\text{H}_4\text{SiO}_4$ ) in the pore water and solute flux  
186 samples were determined using Inductively Coupled Plasma-Optical Emission Spectroscopy (ICP-  
187 OES, Spectro Arcos). Dissolved Fe and Mn are assumed to be present in the form of  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$ ,  
188 however some  $\text{Mn}^{3+}$  (Madison et al. 2013) or colloidal and nanoparticulate Fe and Mn might also be  
189 available (Boyd and Ellwood 2010; Raiswell and Canfield 2012). Concentrations of P and S are  
190 assumed to represent  $\text{HPO}_4^{2-}$  and  $\text{SO}_4^{2-}$  respectively. The colourimetric detection limit for  $\text{NH}_4^+$  was  
191  $0.5 \mu\text{M}$ . The practical detection limit on the ICP-OES for Fe, Mn and P was 0.73, 0.11 and  $7.30 \mu\text{M}$ ,  
192 respectively.

193 Solid-phase Fe was fractionated into [1] labile ferric Fe (hydr)oxides and ferrous Fe ( $\text{FeS} +$   
194  $\text{FeCO}_3$ ), [2] crystalline Fe minerals, [3] magnetite and [4] pyrite (Supporting Information 1.2; Table  
195 S2), using a combination of two operational extraction methods (Poulton and Canfield 2005; Claff et  
196 al. 2010) as described by Kraal et al. (2017). Concentrations of Fe in all extracts were determined  
197 using the colourimetric phenanthroline method (APHA 2005). Solid-phase S was separated into [1]  
198 acid volatile sulphur (AVS; representing  $\text{FeS}$ ) and [2] chromium reducible sulphur (CRS; representing  
199  $\text{FeS}_2$ ; Table S2) using the method after Burton et al. (2006; 2008) as modified by Kraal et al. (2013).  
200 Sulphide released during the S extraction was trapped as  $\text{ZnS}$  in alkaline Zn acetate traps.  
201 Concentrations of S were determined by iodometric titration (APHA 2005). Solid-phase siderite  
202 ( $\text{FeCO}_3$ ) was determined by subtracting AVS from the labile ferrous concentrations retrieved from the  
203 first step of the Fe extraction. Solid-phase P was fractionated into [1] exchangeable P, [2] citrate-  
204 dithionite-bicarbonate (CDB)-P, [3] authigenic P, [4] detrital P and [5] organic P (Table S2) after  
205 Ruttenberg (1992) as modified by Slomp et al. (1996). The sum of exchangeable P and CDB-P  
206 represents metal bound P, as described in Hermans et al. (2019b). Concentrations of P in all extracts,  
207 except CDB, were measured with the molybdenum blue colourimetric method (Murphy and Riley  
208 1958). The P, Mn (assuming to represent Mn oxides; Hermans et al. 2019b) and Si (assuming to

209 represent metal oxide bound Si; Kostka and Luther III 1994; Rao et al. 2016) in CDB extracts was  
210 determined using ICP-OES.

## 211 **2.6. Elemental Mapping of Fe, Mn, P and Ca**

212 On day 47, an undisturbed core (first 5 cm of surface sediment) was sampled for epoxy resin  
213 embedding for high-resolution elemental mapping (Jilbert et al. 2008; Jilbert and Slomp 2013).  
214 Sediment was carefully pushed upwards from the experimental core into a shorter (7 cm length; 1 cm  
215 diameter) mini core. This mini sub-core was then transferred to an acetone bath in a argon-filled  
216 glovebox and subsequently embedded with Spurr's epoxy resin as described in Jilbert et al. (2008).  
217 After curing, the epoxy-embedded core was split vertically using a rock saw. The surface was  
218 smoothed by applying a 0.3  $\mu\text{m}$  alumina powder layer. Elemental maps of Fe, Mn, P and Ca (30  $\mu\text{m}$   
219 resolution) were retrieved using a Desktop EDAX Orbis  $\mu\text{XRF}$  analyser (Rh tube set at 30 kV,  
220 500  $\mu\text{A}$ , 300 ms dwell-time, equipped with a poly-capillary lens). Similar  $\mu\text{XRF}$  maps for Fe, Mn and  
221 P in epoxy embedded surface sediment were obtained for two field sites: (1) the Gulf of Finland, for  
222 sediments collected in June 2016 as described by Hermans et al. (Submitted), and Lake Grevelingen,  
223 for sediments collected in January and May 2012 as described in Sulu-Gambari et al. (2016a; 2018).

## 224 **2.7. Fluorescence *In-situ* Hybridisation**

225 Fluorescence *in-situ* hybridisation (FISH; Pernthaler et al. 2001) was used to microscopically  
226 quantify the abundance of cable bacteria filaments, as described in Seitaj et al. (2015). FISH analysis  
227 was performed on one intact sediment core retrieved at our sampling site, and the sediment cores from  
228 our incubation experiment used for pore water collection at three time points (days 5, 26 and 207).  
229 These cores were sectioned at 0.5 cm depth resolution for the first 2.5 cm. Each sediment slice was  
230 homogenised and fixed with 0.5 mL ethanol ( $\geq 99.8\%$  purity), and stored in a 2 mL Eppendorf tube at  
231  $-20\text{ }^{\circ}\text{C}$ . For FISH analysis, a volume of 100  $\mu\text{L}$  was retrieved from the Eppendorf tubes and mixed  
232 with a 1:1 solution of PBS/ethanol (500  $\mu\text{L}$ ). Then 10  $\mu\text{L}$  of this mixture was filtered through a  
233 polycarbonate membrane (type GTTP; pore size 0.2  $\mu\text{m}$ , Millipore, USA). Cable bacteria were  
234 classified with a *Desulfobulbaceae*-specific oligonucleotide probe (DSB706; 5-ACC CGT ATT CCT  
235 CCC GAT-3') after counter staining with DAPI (1  $\mu\text{g}/\text{mL}$ ) under an epifluorescence microscope

236 (Zeiss Axioplan, Germany) at 100x magnification. The abundance of cable bacteria was quantified by  
237 determining the length and diameter of all observed filaments in a field ( $105 \times 141 \mu\text{m}$ ) on the filter at  
238 100x magnification (200 fields per sample). Cable bacterial abundances are expressed as filament  
239 length per volumetric unit ( $\text{m cm}^{-3}$ ) or depth integrated per unit area of sediment surface ( $\text{m cm}^{-2}$ ),  
240 consistent with previous studies (Schauer et al. 2014; Malkin et al. 2017).

## 241 **2.8. Scanning Electron Microscopy**

242 Cable bacteria filaments were taken from surface sediments from the oxic zone (upper 2 mm)  
243 after 40 days using a microscope and were transferred to a 15 mL centrifuge tube. The tube was filled  
244 to a volume of ~10 mL using ultra clean water, and was subsequently centrifuged at 2100 rpm for 2  
245 min, after which the water was discarded. This washing step was repeated three times. The washed  
246 samples were then transferred to a sample stub, where the sediment was air-dried over-night prior to  
247 gold coating. The filaments were subsequently subjected to scanning electron microscopy (SEM)  
248 imaging on a Phenom ProX Desktop SEM (Phenom-World B.V., the Netherlands) to obtain high-  
249 resolution images, as described in Geerlings et al. (2019). SEM images were generated under 0.1-0.3  
250 mbar vacuum, and a high accelerating voltage (10 or 15 kV).

## 251 **2.9. Data Analysis and Calculations**

252 The diffusive uptake of  $\text{O}_2$  was calculated from the high-resolution  $\text{O}_2$  depth profiles using the  
253 PROFILE software package (Berg et al. 1998). Total  $\text{H}_2\text{S}$  ( $\Sigma\text{H}_2\text{S} = \text{H}_2\text{S} + \text{HS}^- + \text{S}^{2-}$ ) was calculated as  
254 a function of the recorded  $\text{H}_2\text{S}$  and pH values, accounting for temperature and salinity (Millero et al.  
255 1988; Jeroschewski et al. 1996).

256 The EP depth profiles were normalised by subtracting the background EP signal in the  
257 overlying water from the EP depth profiles, to calculate the EP value relative to that in the overlying  
258 water (Damgaard et al. 2014). The electric field in the sediment was calculated from the linear slope  
259 of the EP depth profiles (average of triplicates) in the surface sediments (Risgaard-Petersen et al.  
260 2014). The magnitude of the current density was subsequently calculated from the gradient in the EP,  
261 the so-called electric field, using Ohm's law:

$$J = \sigma_{sed} \cdot E \quad (1)$$

262 where  $J$  represents the magnitude of the current density ( $\text{mA m}^{-2}$ ),  $\sigma_{pw}$  is the conductivity of  
 263 the sediment matrix ( $\text{S m}^{-1}$ ) and  $E$  ( $\text{mV m}^{-1}$ ) represents the electric field. The conductivity of the pore  
 264 water was corrected for tortuosity and calculated as a function of the temperature and salinity using  
 265 the equations provided by Fofonoff and Millard Jr (1983).

266 The solute fluxes were calculated as described in Glud (2008) and Rao et al. (2016):

$$J = \frac{\Delta C_{ow}}{\Delta t} \cdot \frac{V_{ow}}{A} \quad (2)$$

267 where  $J$  represents the diffusive flux ( $\text{mmol m}^{-2} \text{d}^{-1}$ ),  $\Delta C_{ow}$  represents the concentration change in the  
 268 overlying water ( $\text{mmol m}^{-3}$ ),  $\Delta t$  is the incubation time (d),  $V_{ow}$  is the volume in the overlying water  
 269 ( $\text{m}^3$ ) and  $A$  the surface area of sediment in the core ( $\text{m}^2$ ). In our experimental setup, only those fluxes  
 270 were measurable for  $\text{NH}_4^+$ ,  $\text{Fe}^{2+}$ ,  $\text{Mn}^{2+}$  and  $\text{HPO}_4^{2-}$ , that were  $>0.08$ ,  $>0.06$ ,  $>0.01$  and  $>0.55$   $\text{mmol m}^{-2}$   
 271  $\text{d}^{-1}$ , respectively. However, for these four solutes, fluxes were always too low to be detected. Hence,  
 272 only  $\text{Ca}^{2+}$  and  $\text{H}_4\text{SiO}_4$  fluxes are presented.

273 Diffusive downward fluxes of  $\text{SO}_4^{2-}$  and diffusive upward fluxes of  $\text{NH}_4^+$ ,  $\text{Fe}^{2+}$ ,  $\text{Mn}^{2+}$  and  
 274  $\text{Ca}^{2+}$  were calculated from linearized pore water gradients using Fick's first law (Berner 1980):

$$J = -\phi D_s \cdot \frac{dC}{dz} \quad (3)$$

275

276 The molecular diffusion coefficient was calculated as a function of pressure, salinity and  
 277 temperature using the R package *marelac* (Soetaert et al. 2010) and corrected for the ambient  
 278 tortuosity using the relations listed in Boudreau (1997).

### 279 3. RESULTS

#### 280 3.1. Abundance of Cable Bacteria

281 Examination of the top 2.5 cm of the surface sediments using FISH showed the presence of  
 282 filamentous cable bacteria (Fig. 1C; Fig. S1). The *in-situ* cable bacterial abundance in the sediment at



283 our field site was low ( $14 \text{ m cm}^{-2}$ ). However, after 5 days of incubation in the laboratory, the  
284 abundance increased strongly ( $724 \text{ m cm}^{-2}$ ). At day 26 the abundance of cable bacteria was even  
285 higher ( $1035 \text{ m cm}^{-2}$ ). After 207 days, the cable bacterial abundance in the surface sediment was low  
286 again ( $131 \text{ m cm}^{-2}$ ). SEM imaging confirmed that filaments were indeed cable bacteria (Fig. 1D), as  
287 the external surface of the filament was characterised by a parallel pattern of ridges and grooves along  
288 its latitudinal axis, which is a typical feature of cable bacteria (Cornelissen et al. 2018; Geerlings et al.  
289 2019).

### 290 **3.2. High-resolution Depth Profiles of pH, $\text{O}_2$ , $\Sigma\text{H}_2\text{S}$ and EP**

291 High-resolution depth profiles of pH showed the development of a distinct peak near the  
292 sediment-water interface at day 5, and acidification of the pore water in the deeper sediment (Fig.  
293 2A). The width of this pore water acidification zone increased with time and reached its maximum at  
294 day 26, followed by a decrease in the acidification. The distinct pH peak near the sediment-water  
295 interface gradually disappeared after 33 days. The depth of  $\text{O}_2$  penetration in the sediment remained  
296 constant within the first 40 days of incubation ( $\sim 1.1 \text{ mm}$ ) and subsequently moved downwards with  
297 time to  $9.6 \text{ mm}$  (Fig. 2A; Fig. 3; Fig. S2). The dissolved  $\Sigma\text{H}_2\text{S}$  concentrations remained low ( $<5 \mu\text{M}$ )  
298 throughout the experiment (Fig. 2A). The  $\Sigma\text{H}_2\text{S}$  appearance depth was initially equivalent to the  $\text{O}_2$   
299 penetration depth, and shifted downwards within 5 days, creating a suboxic zone where  $\text{O}_2$  and  $\Sigma\text{H}_2\text{S}$   
300 remained below detection (Fig. 2A; Fig. 3). The width of the suboxic zone remained relatively  
301 constant with time ( $\sim 25 \text{ mm}$ ; Fig. 3), and only slight decreased after 207 days.

302 The EP depth profiles indicate a rapid establishment of an electric current after 5 days ( $0.4$   
303  $\text{mV}$ ; Fig. 2B). The time-series of depth profiles show that the EP increased and also accumulated over  
304 a thicker depth horizon. At day 26 the EP reached its maximum value ( $1.2 \text{ mV}$ ), followed by a  
305 decrease with time. Long-distance electron transport was not active in the anoxic control core (Fig.  
306 S3).

### 307 **3.3. Diffusive Uptake of O<sub>2</sub> and Current Density**

308 The diffusive O<sub>2</sub> uptake of the sediment was highest after 5 days and gradually decreased  
309 with time from ~30 to ~3.6 mmol m<sup>-2</sup> d<sup>-1</sup> (Fig. 4A). The current density rapidly increased from day 0  
310 to day 18, from 6 to 128 e<sup>-</sup> mmol m<sup>-2</sup> d<sup>-1</sup>, and then gradually decreased with time (Fig. 4B). The  
311 duplicate measurements show the same trend for the diffusive O<sub>2</sub> uptake and the current density,  
312 which indicates that the results are reproducible.

### 313 **3.4. Pore Water Profiles**

314 Concentrations of NH<sub>4</sub><sup>+</sup> were low near the sediment-water interface and increased with  
315 sediment depth reaching maximum levels of up to 1.7 mM at depth in the sediment (Fig. 5). The time-  
316 series suggest a gradual decrease in production of dissolved NH<sub>4</sub><sup>+</sup> in the sediment leading to  
317 decreasing concentrations with time. The pore water depth profiles of dissolved SO<sub>4</sub><sup>2-</sup> show a decline  
318 with sediment depth at all time points. However, SO<sub>4</sub><sup>2-</sup> concentrations remained relatively constant  
319 within the top 2 cm of surface sediment between day 12 and 33. Dissolved Fe<sup>2+</sup>, Mn<sup>2+</sup> and Ca<sup>2+</sup> all  
320 show the development of distinct peaks in the pore water with time, and after 40 days those peaks  
321 disappear again. Pore water concentrations of HPO<sub>4</sub><sup>2-</sup> generally increased with sediment depth for all  
322 time points, and concentrations within the top 2 cm were below the detection limit indicating removal.  
323 Dissolved H<sub>4</sub>SiO<sub>4</sub> increased with sediment depth reaching concentration of up to 1 mM.

### 324 **3.5. Diffusive Fluxes**

325 Calculated diffusive fluxes of NH<sub>4</sub><sup>+</sup> into the oxic zone decreased during the incubation  
326 experiment from 4.7 to 1.8 mmol m<sup>-2</sup> d<sup>-1</sup> (Fig. 6A; Fig. S4; Table S3). Rates of SO<sub>4</sub><sup>2-</sup> reduction  
327 estimated from the linear gradient of the decrease in pore water SO<sub>4</sub><sup>2-</sup> in the surface sediment with  
328 depth generally also showed a decrease with time (Fig. 6B; Fig. S5; Table S3). The upward diffusive  
329 flux of Fe<sup>2+</sup> greatly increased from day 5 to day 12 and then gradually decreased with time (Fig. 6C;  
330 Fig. S6; Table S3). The upward diffusive flux of Mn showed an increase in the initial stage of the  
331 experiment and reached its maximum at day 18, followed by a decrease with time (Fig. 6D; Fig. S7;  
332 Table S3). The upward diffusive flux of Ca<sup>2+</sup> showed no clear trend with time, however after 207 days  
333 the flux became extremely low (Fig. 6E; Fig. S8). The upward diffusive flux of H<sub>4</sub>SiO<sub>4</sub> also showed

334 an increase in the initial stage of the experiment, and reached its maximum at day 12, followed by a  
335 decrease with time (Fig. 6F; Fig. S9).

### 336 **3.6. Solid-phase Profiles**

337 The surface sediment in the oxic cores became more enriched in Fe oxides with time, with  
338 concentrations increasing from 53 to 485  $\mu\text{mol g}^{-1}$  (Fig. 7). The deeper sediment in the oxic cores and  
339 the entire anoxic control core had low or no Fe oxides. At day 5, FeS was strongly depleted within the  
340 top 1 cm of the surface sediment and was gradually lost further with time. At day 621, most of the  
341 FeS within the top 2.5 cm of the surface sediment had been dissolved. The anoxic core did not show  
342 such a depletion of FeS in the surface sediment and even showed a slight increase in FeS. Solid-phase  
343 siderite remained rather constant with depth from day 5 to 33, but afterwards was gradually lost from  
344 the surface sediment. At day 621 a large proportion of the siderite was dissolved within the top 2 cm.  
345 Solid-phase siderite concentrations remained constant with depth in the anoxic control core. Solid-  
346 phase depth profiles of Mn oxides, metal bound P and metal oxide bound Si all showed a gradual  
347 increase in the surface sediment with time.

### 348 **3.7. High-resolution Elemental Mapping**

349 High-resolution desktop  $\mu\text{XRF}$  mapping of Fe, Mn, P and Ca of our core after 47 days of  
350 incubations revealed a subsurface sediment layer highly enriched in Fe and P (Fig. 8A). Subsurface  
351 enrichments in Fe, P and Mn in relatively thin layers were also observed in sediments populated by  
352 cable bacteria in the Gulf of Finland and Lake Grevelingen (Fig. 8B and C). In the latter system, the  
353 layers enriched in Fe, P, Mn broadened upon recolonization by macrofauna (Fig. 8D).

## 354 **3. DISCUSSION**

### 355 **4.1. Metabolic Activity of Cable Bacteria**

356 Cable bacteria in our incubation experiment demonstrated a rapid growth, since their  
357 abundance greatly increased after 5 days, and reached its peak at day 26 (Fig. 1C). Such high  
358 abundances are similar to those observed in previous experiments, in which FeS-rich marine  
359 sediments from Aarhus Bay and Lake Grevelingen were incubated (Schauer et al. 2014; Burdorf et al.

360 2018). The activity of cable bacteria exerted a strong impact on the pore water depth profiles of pH,  
361 O<sub>2</sub>, and  $\Sigma\text{H}_2\text{S}$ , as evident from the development of a pH maximum near the sediment-water interface,  
362 the strong pore water acidification in the deeper sediment and the development of a suboxic zone (Fig.  
363 2A). These pore water depth profiles resemble the distinct biogeochemical fingerprint typical for  
364 active cable bacteria, as observed in previous laboratory incubation experiments (Risgaard-Petersen et  
365 al. 2012; Malkin et al. 2014; Schauer et al. 2014; Vasquez-Cardenas et al. 2015; Rao et al. 2016;  
366 Burdorf et al. 2018). The widening of the suboxic zone with time (Fig. 3) is a consequence of the  
367 downward expansion of the cable bacteria filament network (Schauer et al. 2014; Vasquez-Cardenas  
368 et al. 2015).

369 The EP depth profiles demonstrated that long-distance electron transport by cable bacteria  
370 was already active 5 days after the start of the experiment, as indicated by the increase of EP at depth  
371 to 0.4 mV). With time, the EP signal increased to higher values and also accumulated over a thicker  
372 depth horizon (Fig. 2B), indicating that cable bacteria activity both increased and extended to deeper  
373 sediment depth, which is also a consequence of the downward expansion of cable bacteria filaments.  
374 The EP reached a maximum after 18 days (1.3 mV; Fig. 2B) in concert with the highest current  
375 density of  $\sim 130 \text{ mmol e}^- \text{ m}^{-2} \text{ d}^{-1}$  (Fig. 4B). This maximum EP value and current density are similar in  
376 magnitude to those found in sediment incubations with seawater with a similar salinity (Damgaard et  
377 al. 2014). From day 18 onwards the EP and current density flux gradually decreased with time to 13  
378  $\text{mmol e}^- \text{ m}^{-2} \text{ d}^{-1}$  after 207 days (Fig. 4B), which implies a decrease in the metabolic activity of cable  
379 bacteria. The suboxic zone persisted long after the current density had decreased (Fig. 3).

380 To summarise, the metabolic activity of cable bacteria was likely highest between day 18 and  
381 day 26 based on the cable bacterial abundances, the extent of acidification of the pore water and the  
382 current density (Fig. 1C; Fig. 2A and Fig. 4B).

#### 383 **4.2. Organic Matter Degradation**

384 Ammonium fluxes are assumed to reflect rates of anaerobic degradation of organic matter ( Fig  
385 6A), and the observed decline during the experiment coincides with the decrease in activity of cable

386 bacteria based on the EP profiles and current density (Fig. 2B; Fig. 4B). This suggests that the  
387 availability of easily degradable organic matter plays a role in sustaining the metabolic activity of  
388 cable bacteria, most likely by controlling the rate of  $\text{SO}_4^{2-}$  reduction (Nielsen and Risgaard-Petersen  
389 2015).

390 Rates of  $\text{SO}_4^{2-}$  reduction estimated from the linear gradient of the decrease in pore water  $\text{SO}_4^{2-}$  in  
391 the surface sediment with depth indeed also showed a decline during the experiment. We note,  
392 however, that a direct measurement of  $\text{SO}_4^{2-}$  reduction rates (Fossing and Jørgensen 1989; Kallmeyer  
393 et al. 2004) would provide a better indicator, because  $\text{SO}_4^{2-}$  estimated from pore water profiles are in  
394 general lower than rates estimated from tracer experiments (Hermans et al. 2019a; Sandfeld et al.  
395 2020). Another cause for a slight underestimation of our  $\text{SO}_4^{2-}$  reduction rates, is due to the effect of  
396 the electric field imposed by cable bacteria, which is not taken into account in Fick's law. **Solutes can  
397 also move with respect to the fluid by electrostatic forces (Bockris and Reddy 1998). Given the  
398 relatively low strength of the electric field in the cores ( $<0.073\text{V m}^{-1}$  at day 18; as estimated from Fig.  
399 2B), including the contribution of ionic drift to the sulphate flux would lead to  $\text{SO}_4^{2-}$  reduction rates  
400 that are at most 10-20% higher.**

401 The metabolic activity of cable bacteria can lead to the production of  $\text{SO}_4^{2-}$  in the suboxic zone  
402 via anodic sulphide oxidation (Risgaard-Petersen et al. 2012; Rao et al. 2016). We suspect that this  
403 also explains the lack of change in pore water  $\text{SO}_4^{2-}$  with depth in the upper 2 cm of the sediment in  
404 our experiment between 12 and 40 days (Fig. 5). Despite relatively high  $\text{SO}_4^{2-}$  reduction rates ranging  
405 from 5.4 to 17.6  $\text{mmol m}^{-2} \text{d}^{-1}$  (Fig. 6B; Table S3), pore water concentrations of  $\Sigma\text{H}_2\text{S}$  remained very  
406 low throughout the experiment (Fig. 2A). This is likely due to the direct consumption of  $\Sigma\text{H}_2\text{S}$   
407 through the activity of cable bacteria, preventing  $\Sigma\text{H}_2\text{S}$  from accumulating in the pore water, or  
408 alternatively, precipitation of FeS by dissolved  $\text{Fe}^{2+}$  released from the dissolution of siderite.

409 Laboratory experiments have shown that S-oxidation by cable bacteria can play a dominant role  
410 in the  $\text{O}_2$  uptake of coastal sediments (Nielsen et al. 2010; Schauer et al. 2014; Nielsen and Risgaard-  
411 Petersen 2015), and model analysis predicts up to 93% of the total  $\text{O}_2$  uptake (Meysman et al. 2015).

412 When we plot diffusive uptake of O<sub>2</sub> against the current density (i.e. upward flux of electrons towards  
413 the oxic zone), a linear relationship – with some scatter - emerges for days 12 to 621 (Fig. 9).  
414 However, the data points for day 0 and 5 during the initial stages of our experiment do not follow this  
415 linear relationship. We explain these findings as follows: At day 0, the cable bacteria were not active  
416 yet and other processes, such as aerobic respiration and oxidation of NH<sub>4</sub><sup>+</sup> and other solutes (Table 2)  
417 and solids (FeS) dominated the consumption of O<sub>2</sub>. At day 5 and 12, the activity of cable bacteria and  
418 the oxidation of reduced products from anaerobic degradation of organic matter both contributed to  
419 consumption of O<sub>2</sub>. From day 12 onwards, both the O<sub>2</sub> consumption and electron flux follow a  
420 downward decrease with time (Fig. 9). If cable bacteria would account for all of the O<sub>2</sub> consumption,  
421 a ratio between the diffusive uptake of O<sub>2</sub> and the current density of 1:4 is expected (Fig. 1A; Nielsen  
422 et al. 2010). We find that from day 12 onwards, most data points plot rather close to the line for this  
423 1:4 relationship (Fig. 9), suggesting that cathodic O<sub>2</sub> reduction by cable bacteria is responsible for  
424 nearly all O<sub>2</sub> consumption in the sediment (in line with the model results of Meysman et al. 2015).  
425 This however poses a problem for the nitrogen budget, because our data indicate complete removal of  
426 the NH<sub>4</sub><sup>+</sup> that diffuses upward into the oxic zone (Fig. 6A), and based on the solute fluxes, no escape  
427 to the overlying water (see section 2.4). This implies substantial O<sub>2</sub> consumption due to nitrification  
428 (Table 2). These findings can be explained, however, if we assume that at least part of the NO<sub>3</sub><sup>-</sup> that is  
429 being formed near the sediment-water interface is also used for the metabolic activity of cable  
430 bacteria. It has been shown that cable bacteria can couple the oxidation of ΣH<sub>2</sub>S to NO<sub>3</sub><sup>-</sup> in the  
431 absence of O<sub>2</sub> (Marzocchi et al. 2014). Our data suggest that this process may also occur in sediments  
432 where O<sub>2</sub> is present in concert with NO<sub>3</sub><sup>-</sup> near the sediment-water interface. However, we cannot  
433 exclude release of NO<sub>3</sub><sup>-</sup> to the water column or denitrification by other bacteria in the sediment.  
434 Another explanation is that cable bacteria might consume O<sub>2</sub> directly above the sediment-water  
435 interface, as recently has been proposed by Burdorf et al. (2018). Lastly, the current density might be  
436 slightly overestimated, since it ignores other sources that can create an electric potential, such as the  
437 diffusion potential (Revil et al. 2012; Nielsen and Risgaard-Petersen 2015).

### 438 **4.3. Impact of Cable Bacteria on Fe, Mn and S Cycling**

439 The activity of cable bacteria had a strong impact on the biogeochemistry of the surface sediment  
440 in our experiment (Fig. 7). Cable bacteria activity induced an intense acidification of the pore water in  
441 the suboxic zone (Fig. 2A), which led to the dissolution of Fe and Mn minerals in deeper sediment  
442 layers, as can be inferred from the sharp maxima in dissolved  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  in the pore water  
443 reaching concentrations of up to  $\sim 1700$  and  $\sim 80$   $\mu\text{M}$ , respectively (Fig. 5). The twenty-fold higher  
444 dissolved  $\text{Fe}^{2+}$  concentrations with respect to pore water  $\text{Mn}^{2+}$  can be attributed to the relatively higher  
445 availability of FeS and siderite compared to the availability of Mn carbonates in the sediment that was  
446 used for incubation (Lenstra et al. 2020). The peaks in dissolved  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  in the pore water  
447 broadened over time spanning a depth of  $>5\text{cm}$  (Fig. 5; Fig. S6; Fig. S7).

448 The upward diffusive flux of dissolved  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  was highest after 12 days, reaching values  
449 of up to  $3.16$  and  $0.16$   $\text{mmol m}^{-2} \text{d}^{-1}$  respectively. Fluxes subsequently gradually decreased with time  
450 (Fig. 6C and D). The continuous upward diffusion of dissolved  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  led to enrichments of  
451 poorly crystalline Fe and Mn oxides in the surface sediment (Fig. 7). Despite high upward fluxes of  
452 dissolved  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  towards the sediment-water interface, our solute flux incubations indicate  
453 there was little escape of  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  to the overlying water (see section 2.4). This implies that all  
454  $\text{Fe}^{2+}$  and  $\text{Mn}^{2+}$  that diffused upward was precipitated as Fe and Mn oxides upon contact with  $\text{O}_2$  or  
455  $\text{NO}_3^-$  (Buresh and Moraghan 1976; Kuz'minskii et al. 1994; Straub et al. 1996). Little or no escape of  
456 dissolved  $\text{Fe}^{2+}$  from the sediment into the overlying water, was suggested previously for a field site  
457 with active cable bacteria based on diffusive flux calculations (Lake Grevelingen; Sulu-Gambari et al.  
458 2016a) and was determined in flux incubations of cores during a laboratory experiment with cable  
459 bacteria (Rao et al. 2016).

460 At the start of the experiment, the sedimentary FeS content was ( $\sim 25$   $\mu\text{mol g}^{-1}$ ), which is not  
461 unusual for coastal sediments on the north-western Black Sea margin (Wijsman et al. 2001), but is  
462 low compared to sediments in eutrophic coastal systems (e.g. Morgan et al. 2012; Kraal et al. 2013;  
463 Hermans et al. 2019a). The solid-phase depth profiles reveal a gradual removal of the FeS in the  
464 surface sediment in our experiment over time (Fig. 7). At the end of our experiment (621 days), there

465 was no longer any FeS within the top 1.5 cm of the sediment. While approximately 90 mmol m<sup>-2</sup> of  
466 FeS was removed from the surface sediment within the first 5 days, a total of ~240 mmol m<sup>-2</sup> was  
467 removed after 621 days (Fig. 10; Table 3). Likely, part of the FeS that was removed from the surface  
468 sediment within the first 5 days was removed through oxidation upon contact with O<sub>2</sub>, rather than the  
469 metabolic activity of cable bacteria itself. The pore water acidification associated with cable bacteria  
470 activity led to a strong loss of siderite within the top 2 cm of the sediment, with a total removal of  
471 ~560 mmol m<sup>-2</sup> during the experiment (Fig. 7; Fig. 10; Table 3; Table S5). The depletion of  
472 sedimentary FeS and siderite was directly proportional to the formation of Fe oxides near the  
473 sediment-water interface (Fig. 10), and accounted for 30% and 70% of the Fe oxides, respectively  
474 (Table 3).

475 With these data we cannot accurately determine the role of FeS versus SO<sub>4</sub><sup>2-</sup> reduction in  
476 supplying the ∑H<sub>2</sub>S sustaining the activity of cable bacteria throughout the experiment. This is  
477 primarily related to the variability between cores, and for this type of calculation, the low temporal  
478 resolution of sampling. However, we can make an estimation of the relative role of SO<sub>4</sub><sup>2-</sup> reduction  
479 and FeS dissolution in ∑H<sub>2</sub>S production, based on the pore water profiles of SO<sub>4</sub><sup>2-</sup> and dissolved Fe<sup>2+</sup>,  
480 and the solid-phase mass balance of FeS and siderite (Fig. 6B and C; Table 4). This estimation shows  
481 that SO<sub>4</sub><sup>2-</sup> was mainly responsible for ∑H<sub>2</sub>S production, accounting for 85-99% (Table 4), and thus  
482 that the dissolution of FeS only played a minor role in providing ∑H<sub>2</sub>S.

#### 483 **4.4. Impact of Cable Bacteria on Ca, P and Si Cycling**

484 Cable bacteria activity is known to lead to dissolution of Ca carbonates, because of the strong  
485 acidification of the pore water (Risgaard-Petersen et al. 2012; Rao et al. 2016). We indeed find similar  
486 maxima in pore water Ca<sup>2+</sup> during the experiment (Fig. 5) and a high upward flux of Ca<sup>2+</sup> (up to ~18  
487 mmol m<sup>-2</sup> d<sup>-1</sup>; Fig. 6E; Fig. S8) of which a substantial fraction (up to ~55%) escapes to the overlying  
488 water (Fig. S10; Table S4), which is consistent with a previous incubation experiment Rao et al.  
489 (2016).



490 Pore water depth profiles of  $\text{HPO}_4^{2-}$  reveal a production at depth and removal of all upward  
491 diffusing  $\text{HPO}_4^{2-}$  within the first 1-3 cm of the surface sediment (Fig. 5). A major proportion of this  
492  $\text{HPO}_4^{2-}$  is bound to Fe oxides (Fig. 7). Given that a large proportion of the Fe oxides in our sediment  
493 cores derive from the dissolution of siderite, this suggests that the buffer mechanism that delays the  
494 benthic release of  $\text{HPO}_4^{2-}$  through retention of P associated with newly formed Fe oxides (Sulu-  
495 Gambari et al. 2016b), might also be active in systems that are relatively poor in sedimentary FeS.

496 The shape of the pore water  $\text{HPO}_4^{2-}$  profiles suggests that some of the  $\text{HPO}_4^{2-}$  is removed below  
497 the zone where Fe and Mn oxides are present (Fig. 5; Fig. 7). A possible explanation could be the  
498 formation of vivianite, an Fe(II) phosphate mineral. Vivianite formation in sediments typically occurs  
499 when pore water levels of  $\text{Fe}^{2+}$  and  $\text{HPO}_4^{2-}$  are high and concentrations of  $\Sigma\text{H}_2\text{S}$  are low (Nriagu  
500 1972), as observed in our study. In our experiment, free  $\Sigma\text{H}_2\text{S}$  does not accumulate in the pore water,  
501 which we attribute to removal through the activity of cable bacteria and FeS formation at depth (Fig.  
502 2A; Fig. 7). Hence, cable bacteria may create a geochemical niche that allows the formation of  
503 vivianite in the suboxic zone. Further work with sediments with higher P concentrations would be  
504 needed to assess this with direct measurement techniques, such as X-ray spectroscopy (Egger et al.  
505 2015; Kraal et al. 2017; Sulu-Gambari et al. 2018). Other sediment P pools, i.e. organic, authigenic  
506 and detrital P remained constant over time, indicating that the P contents determined for discrete  
507 sediment slices using sequential extractions were not affected by pore water acidification as a result of  
508 cable bacteria activity (Table S6).

509 Pore water  $\text{H}_4\text{SiO}_4$  profiles show a typical increase with depth as observed upon dissolution of  
510 biogenic silica in marine sediments (Aller 2014). Fluxes of  $\text{H}_4\text{SiO}_4$  towards the sediment-water  
511 interface range up to  $\sim 2.8 \text{ mmol m}^{-2} \text{ d}^{-1}$  and gradually decreased with time (Fig. 6F; Fig. S9). The  
512 results of the solute flux incubations indicate that most of this  $\text{H}_4\text{SiO}_4$  escaped to the overlying water  
513 (ranging from 28 to 92%; Table S4; Fig. S10). The decline in the benthic release flux of  $\text{H}_4\text{SiO}_4$   
514 contrasts with results of a previous incubation experiment by Rao et al. (2016) with similar pore water  
515 concentrations of  $\text{H}_4\text{SiO}_4$  reaching values up to  $\sim 1 \text{ mM}$ . In their study, the flux remained constant  
516 over time, possibly because of differences in the amount of biogenic Si in the sediment. The solid-

517 phase metal oxide bound Si pool in the surface sediment increased directly proportional to the  
518 formation of Fe oxides throughout the experiment (Fig. 7). Silica is known to absorb to Fe oxides  
519 (Sigg and Stumm 1981; Davis et al. 2002). Hence, the results suggest that the Fe oxides formed  
520 through the activity of cable bacteria captured some of the upward diffusing  $H_4SiO_4$ .

#### 521 **4.5. Sediment Marker for Cable Bacteria Activity**

522 Visual observations of core photographs reveal the gradual development of an orange layer (oxic  
523 zone) up to 9 mm thick, overlying a grey layer (suboxic zone) and a black layer (sulphidic zone)  
524 during the experiment (Fig. S11). This colour zonation is typical for sediments that have been  
525 geochemically affected by cable bacteria activity, as seen both in laboratory experiments (Nielsen and  
526 Risgaard-Petersen 2015) and at coastal field sites (Sulu-Gambari et al. 2016a). High-resolution  
527 elemental maps of our sediments reveal the development of a ~0.3 mm thin subsurface layer highly  
528 enriched in Fe oxides and associated P, 47 days after the start of the incubation (Fig. 8A). While the  
529 Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P very close to the  
530 sediment-water interface (Fig. 8A). This layer is located above the Fe oxide layer, and in this layer P  
531 is strongly correlated with Ca. Below, we describe the formation of this layer in more detail and  
532 explain why such subsurface enrichments, detected with  $\mu$ XRF, may act as an additional sediment  
533 marker for present or recent cable bacteria activity, also in cases where visual observations are not  
534 conclusive.

535 During the experiment,  $O_2$  penetration varied within a narrow range and was initially fixed  
536 between 1 and 2 mm depth (Fig. 3A), with the layer highly enriched in Fe forming mostly at a depth  
537 of 2 mm (Fig. 8A). Such a range in  $O_2$  penetration is in accordance with observations in coastal  
538 sediments (e.g. Seitaj et al. 2015). The formation of the Fe-enriched layer can be explained by rapid  
539 oxidation of upward diffusing  $Fe^{2+}$  upon contact with  $O_2$  (and possibly  $NO_3^-$ ; Fig. 6C). Directly, above  
540 the Fe oxide layer a broader ~0.8 mm thick Mn oxide layer was observed (Fig. 8A). This contrast in  
541 zonation between Fe and Mn is likely due to the slower oxidation kinetics of  $Mn^{2+}$  compared to  $Fe^{2+}$   
542 (Burdige 1993; Luther 2010; Learman et al. 2011).

543 While the Fe oxide layer is clearly enriched in P, we also observed a second layer enriched in P  
544 close to the sediment-water interface (Fig. 8A). In this layer, P is strongly correlated with Ca. This  
545 layer likely consists of carbonate fluorapatite (CFA), a Ca-P mineral, which is typically formed in  
546 marine sediments (Van Cappellen and Berner 1988; Ruttenberg and Berner 1993). Possibly, the high  
547 pore water pH near the sediment-water interface (resulting from cathodic O<sub>2</sub> reduction by cable  
548 bacteria; Fig. 2A), promotes apatite formation (Bellier et al. 2006), and the elevated Ca<sup>2+</sup>  
549 concentrations (Fig. 5) created a biogeochemical niche for the formation of CFA.

550 Such focusing of Fe, Mn, P and associated elements within a thin subsurface layer, as a  
551 consequence of cable bacteria activity, also occurs in the field. This was demonstrated by Hermans et  
552 al. (Submitted) in a study of a coastal site in the Gulf of Finland where cable bacteria were recently  
553 active. Here,  $\mu$ XRF mapping of resin embedded sediments revealed strong focusing of Fe oxides,  
554 Mn(II) phosphates and Fe bound P within a 3 mm thick layer near the sediment-water interface (Fig.  
555 8B). A re-assessment of the  $\mu$ XRF data of Sulu-Gambari et al. (2016a; 2018) of surface sediments  
556 with active cable bacteria from seasonally hypoxic marine Lake Grevelingen in January also revealed  
557 similar subsurface enrichments in Fe, Mn and P (Fig. 8C). Importantly, no visual signals for cable  
558 bacteria based on the colour pattern of the sediment were observed at the time.

559 Macrofaunal activity likely counteracts or prevents strong focusing of Fe oxides and associated P  
560 within such a thin subsurface layer at field sites. Bioturbation, i.e. mixing of the sediment, typically  
561 leads to oxidation from the sediment surface downwards (Norkko et al. 2012). Bioirrigation can  
562 efficiently pump O<sub>2</sub> into the pore water and thereby enhance the oxidation of dissolved Fe<sup>2+</sup>  
563 (Kristensen et al. 2012; Norkko et al. 2012), but is not expected to lead to such a sharp oxidation front  
564 (Norkko et al. 2012; Hermans et al. 2019a). This is also evident from high-resolution elemental maps  
565 of the surface sediment from Lake Grevelingen in May, which shows the disappearance of the thin  
566 layer highly enriched in Fe and P formed by cable bacteria in January as a consequence of  
567 macrofaunal activity in May (Fig. 8D; Seitaj et al. 2015; Sulu-Gambari et al. 2016b).

568 We conclude that the focusing of Fe, Mn and associated P within a thin layer below the sediment-  
569 water interface is likely a consistent feature in sediments populated by active cable bacteria and may  
570 act as an additional sediment marker for present or recent cable bacteria activity, both in laboratory  
571 experiments and at field sites, also in cases where visual observations are not conclusive. Focussing of  
572 Fe and Mn oxides in the surface sediment is not exclusively tied to the activity of cable bacteria, and  
573 can also occur in the absence of cable bacteria. However, the upward fluxes of Fe<sup>2+</sup> and Mn<sup>2+</sup> in  
574 sediments populated by cable bacteria, are higher due to active dissolution of Fe and Mn minerals at  
575 depth (e.g. Risgaard-Petersen et al. 2012; Rao et al. 2016). Hence, within the same time frame  
576 following an environmental perturbation (such as a transition to oxic bottom waters after a period of  
577 anoxia or mixing of the sediment), more Fe<sup>2+</sup> and Mn<sup>2+</sup> can oxidise upon contact with O<sub>2</sub> near the  
578 sediment-water interface and stronger enrichments of Fe and Mn minerals will be observed. Hence,  
579 focusing of Fe and Mn oxides in subsurface sediments is likely more prominent and stronger in  
580 sediments populated by active cable bacteria compared to sediments where no cable bacteria are  
581 active under such conditions. Macrofaunal activity within natural environments likely counteracts or  
582 prevents strong focusing of Fe oxides and associated P within such a thin subsurface layer. When  
583 using standard techniques for sediment sampling (i.e. core slicing and chemical analysis of these  
584 slices), these layers may be missed due to the relatively coarse depth resolution. Hence,  $\mu$ XRF  
585 mapping of epoxy embedded sediment is recommended.

#### 586 **4.6. Cable Bacteria Activity at the Field Site**

587 We can only speculate about the possible *in-situ* relevance of cable bacteria at the coastal site  
588 in the western Black Sea where the sediment for our incubation was collected. At this site, which is in  
589 a region that is subject to seasonal hypoxia (Capet et al. 2013), both bivalves (up to ~7200 ind. m<sup>-2</sup>)  
590 and polychaetes (up to ~1700 ind. m<sup>-2</sup>) were observed at the time of sampling (Lenstra et al. 2019).  
591 Macrofauna can inhibit the activity of cable bacteria through bioturbation by physically cutting and  
592 damaging the filaments, rendering them unable to transport electrons (Malkin et al. 2014). Recent  
593 work has shown, however, that in some cases, cable bacterial communities can also thrive in  
594 sediments with macrofauna (Burdorf et al. 2017; Malkin et al. 2017; Aller et al. 2019). In a study of

595 bivalve reefs, cable bacteria were found to efficiently remove highly toxic  $\Sigma\text{H}_2\text{S}$ , which is beneficial  
596 for bivalves (Malkin et al. 2017). Cable bacteria can also be abundant in bioturbated deposits, when  
597 associated with stable subdomains of the bioturbated zone, such as worm tubes (Aller et al. 2019). In  
598 such settings, a more complex precipitation pattern, e.g. along tube linings is observed (Aller et al.  
599 2019), than described here for laboratory experiments with defaunated sediments and field sediments  
600 with an impoverished macrofaunal population (Fig. 8A). Further field studies are required to assess  
601 the role of cable bacteria at our field site, preferably including an assessment of the burrow structures.

## 602 **Conclusions**

603         The results of our laboratory incubation (with a total duration of 621 days) show that cable  
604 bacteria can potentially strongly impact the Fe, Mn, P and S dynamics in coastal sediments. The  
605 strong acidity of the pore water associated with the activity of cable bacteria, which was monitored  
606 using microsensor profiling of the EP during the experiment, led to dissolution of FeS and siderite and  
607 formation of Fe and Mn oxides and Ca-P in mineral form near the sediment surface. Our experimental  
608 results provide conclusive evidence for siderite dissolution driven by cable bacteria activity as a  
609 source of Fe that can form an Fe oxide-enriched surface layer. Both FeS and  $\text{SO}_4^{2-}$  reduction provided  
610 the  $\Sigma\text{H}_2\text{S}$  required by cable bacteria to sustain their activity. Pore water  $\Sigma\text{H}_2\text{S}$  was always low (<5  
611  $\mu\text{M}$ ). Using  $\mu\text{XRF}$  mapping of epoxy embedded sediment, we show that the activity of cable bacteria  
612 led to the development of a thin subsurface sediment layer (0.3 mm) that was highly enriched in Fe  
613 and P. The position of this layer in the sediment was directly proportional to the  $\text{O}_2$  penetration depth  
614 during the experiment. We show that a similar layer highly enriched in Fe and P was also formed in  
615 sediments of field locations populated by cable bacteria (i.e. marine Lake Grevelingen and the  
616 brackish Gulf of Finland). We suggest that such layers, which are not necessarily visible by eye, may  
617 be used as a marker of cable bacteria activity in sediments with low macrofaunal activity.

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630 **Code availability.** All data, if not directly available from the tables and supplementary information,  
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632 the main author.

633 **Author contributions.** MH and CPS designed the experiment. MH carried out the experiment and  
634 analysis. All authors interpreted the data. MH and CPS wrote the paper with comments provided by  
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886 **TABLES AND FIGURES**

887 **Table 1.** Key site characteristics: latitude, longitude, water depth, bottom water O<sub>2</sub> concentration, *in-situ* O<sub>2</sub> uptake, *in-situ*  
 888 O<sub>2</sub> penetration depth in the sediment, porosity and salinity. These data were retrieved from Lenstra et al. (2019). Our study  
 889 site is station 9 in Lenstra et al. (2019).

<b>Black Sea (Station 9)</b>		<b>Unit</b>
Latitude	44°34.9'	N
Longitude	29°11.4'	E
Water depth	27	m
Bottom water O <sub>2</sub>	92	μM
O <sub>2</sub> uptake	25.8 ± 1.77	mmol m <sup>-2</sup> d <sup>-1</sup>
O <sub>2</sub> penetration depth	2.25	mm
Porosity	0.86	-
Salinity	17.881	-
Avg. organic carbon content (0-0.5 cm)	1.8%	

890  
 891 **Table 2.** Mass balance of O<sub>2</sub> consumption. The diffusive uptake of O<sub>2</sub> as calculated from the O<sub>2</sub> depth profiles (column 1)  
 892 was compared to the potential O<sub>2</sub> demand from the oxidation of NH<sub>4</sub><sup>+</sup>, Fe<sup>2+</sup> and Mn<sup>2+</sup> (column 2-4). The O<sub>2</sub> consumption of  
 893 the oxidation of NH<sub>4</sub><sup>+</sup>, Fe<sup>2+</sup> and Mn<sup>2+</sup> was determined based on the stoichiometry of NH<sub>4</sub><sup>+</sup>, Fe<sup>2+</sup> and Mn<sup>2+</sup> oxidation with O<sub>2</sub>  
 894 as described in Reed et al. (2011). The oxidation of dissolved Fe<sup>2+</sup> and Mn<sup>2+</sup> only played a minor role in the total O<sub>2</sub>  
 895 consumption during the experiment, contributing only 0.9 to 3.8% and 0.1 to 0.4%, respectively.

	<b>O<sub>2</sub></b> <b>[mmol m<sup>-2</sup> d<sup>-1</sup>]</b>	<b>Potential O<sub>2</sub> Demand</b>			<b>e<sup>-</sup></b> <b>[mmol m<sup>-2</sup> d<sup>-1</sup>]</b>
		<b>NH<sub>4</sub><sup>+</sup></b> <b>[mmol m<sup>-2</sup> d<sup>-1</sup>]</b>	<b>Fe<sup>2+</sup></b> <b>[mmol m<sup>-2</sup> d<sup>-1</sup>]</b>	<b>Mn<sup>2+</sup></b> <b>[mmol m<sup>-2</sup> d<sup>-1</sup>]</b>	
Day 5	-23.35	9.42	0.21	0.05	82.68
Day 12	-23.24	8.46	0.89	0.09	111.94
Day 18	-21.10	8.04	0.70	0.08	127.97
Day 26	-23.00	7.58	0.63	0.07	97.55
Day 33	-22.80	5.06	0.62	0.05	84.16
Day 40	-19.60	4.88	0.60	0.06	76.31
Day 207	-6.90	3.52	0.08	0.01	13.10
Day 621	-3.25	N/A	0.03	0.01	9.47

896

897 **Table 3.** Mass balance of Fe. Time-series of the depth integrated (0-5 cm) increase in Fe oxides and the depth integrated (0-5  
898 cm) depletion of FeS and FeCO<sub>3</sub> (siderite) in mmol m<sup>-2</sup>. All values are reported in mmol Fe m<sup>-2</sup>. Negative values represent a  
899 decrease, whereas positive values indicate an increase in the mineral pools.

	$\Delta\text{Fe oxides}$ [mmol m <sup>-2</sup> ]	$\Delta\text{FeS}$ [mmol m <sup>-2</sup> ]	$\Delta\text{FeCO}_3$ [mmol m <sup>-2</sup> ]
Day 5	120	-90	-42
Day 12	170	-90	-126
Day 18	189	-105	-92
Day 26	276	-174	-99
Day 33	315	-176	-109
Day 40	412	-223	-200
Day 207	523	-236	-341
Day 621	874	-242	-566

900

901 **Table 4.** Sources of  $\Sigma\text{H}_2\text{S}$  calculated from the reduction of SO<sub>4</sub><sup>2-</sup> and the dissolution of FeS. The numbers are presented  
902 either as mmol m<sup>-2</sup> d<sup>-1</sup> or as the relative percentage of the  $\Sigma\text{H}_2\text{S}$  production. The amount of S from the dissolution of FeS  
903 was estimated from the upward diffusive flux of Fe<sup>2+</sup> (Fig. 6C) and the relative fraction of FeS (FeS/FeS+siderite) based on  
904 the mass balance calculations (Table 3).

	<b>S from SO<sub>4</sub><sup>2-</sup> reduction</b> [mmol m <sup>-2</sup> d <sup>-1</sup> ]	<b>S from FeS dissolution</b> [mmol m <sup>-2</sup> d <sup>-1</sup> ]	<b>S from SO<sub>4</sub><sup>2-</sup> reduction</b> [%]	<b>S from FeS dissolution</b> [%]
Day 5	10.49	0.56	95%	5%
Day 12	17.60	1.48	92%	8%
Day 18	8.87	1.50	86%	14%
Day 26	11.15	1.61	87%	13%
Day 33	8.54	1.52	85%	15%
Day 40	7.57	1.25	86%	14%
Day 207	10.38	0.13	99%	1%
Day 621	5.36	0.03	99%	1%

905

906 **Fig. 1. (A)** Geochemical pore water fingerprint typical for cable bacteria activity. This fingerprint is defined by a distinct pH  
907 profile (light grey line) and a sub-oxic zone that is devoid of O<sub>2</sub> (red line) and H<sub>2</sub>S (blue line). The cable bacteria filaments  
908 are depicted in yellow. On the background, the sediment core photograph, taken 278 days after the start of the experiment,  
909 shows a distinct colour zonation where (1) the oxic zone displays an orange colour (2) the suboxic zone has a grey colour  
910 and (3) the sulphidic zone has a black colour. The scale bar denotes a distance of 6 cm, with 0.5 cm intervals. **(B)**  
911 Bathymetric map of the Black Sea. The purple star indicates the location of our study site (44°34.93'N, 29°11.38'E), which  
912 was sampled with R/V *Pelagia* in September 2015. Further details are provided in Lenstra et al. (2019). **(C)** Volumetric  
913 density of cable bacteria [m cm<sup>-3</sup>] in the top 2.5 cm of the sediment, for *in-situ* as well as for three time points during the  
914 incubation experiment **(D)** SEM image of a cable bacteria filament that was extracted from the surface sediment after 40  
915 days.

916

917 **Fig. 2. (A)** Time-series of the pore water pH (black), O<sub>2</sub> (red) and ΣH<sub>2</sub>S (blue) signatures of the incubated sediment. **(B)**  
918 Development of the EP depth profile in the incubated sediment over time. The dashed-line at 0 mm depth represents the  
919 sediment-water interface. The blue boxes indicate the overlying water, whereas the underlying light grey boxes represent the  
920 sediment. The EP depth profiles represent an average of 3 replicate measurements. The error bars indicate the minimum and  
921 maximum EP values that were observed. The orange depth profiles represent duplicate measurement performed on a  
922 different core.

923

924 **Fig. 3.** Time-series of the development of the oxic zone (orange), suboxic zone (light grey) and the anoxic/sulphidic zone  
925 (dark grey) in the sediment. These zones were calculated from 3 replicate microelectrode depth profiles retrieved from two  
926 different cores.

927

928 **Fig. 4.** Time-series of the **(A)** diffusive O<sub>2</sub> uptake in mmol O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> and **(B)** current density as a consequence of long-  
929 distance electron transport (e<sup>-</sup>) in mmol e<sup>-</sup> m<sup>-2</sup> d<sup>-1</sup> in the sediment incubation.

930

931 **Fig. 5.** Time-series of pore water depth profiles of NH<sub>4</sub><sup>+</sup> (orange), SO<sub>4</sub><sup>2-</sup> (purple), Fe<sup>2+</sup> (red), Mn<sup>2+</sup> (green), Ca<sup>2+</sup> (grey),  
932 HPO<sub>4</sub><sup>2-</sup> (blue) and H<sub>4</sub>SiO<sub>4</sub> (yellow). The control core was sampled at day 621.

933

934 **Fig. 6.** Time-series of diffusive fluxes calculated from the linear gradient of the pore water profiles of **(A)** NH<sub>4</sub><sup>+</sup>, **(B)** SO<sub>4</sub><sup>2-</sup>,  
935 **(C)** Fe<sup>2+</sup>, **(D)** Mn<sup>2+</sup>, **(E)** Ca<sup>2+</sup> and **(F)** H<sub>4</sub>SiO<sub>4</sub> in mmol m<sup>-2</sup> d<sup>-1</sup> towards the oxic zone of the sediment, based on the linear pore  
936 water gradients (Section 1.6; Fig. S4-S9). Here, a positive value indicates an upward flux, whereas a negative value  
937 represents a downward flux. N/A = not available. The control core was sampled at day 621.

938

939 **Fig. 7.** Time-series of solid-phase depth profiles of Fe oxides (red), FeS (black), siderite (grey), Mn oxides (green), metal  
940 bound P (blue) and metal oxide bound Si (yellow).

941

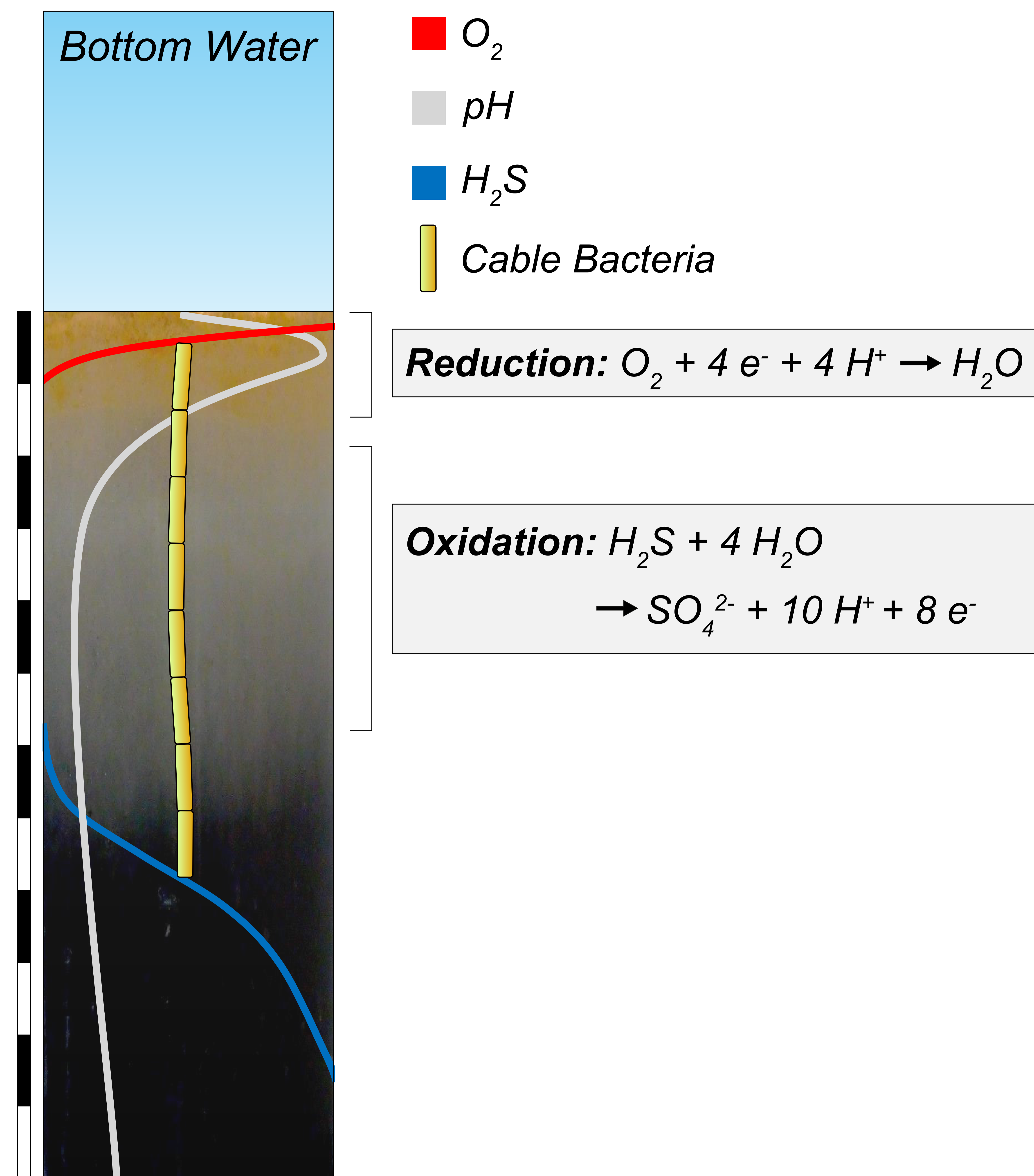
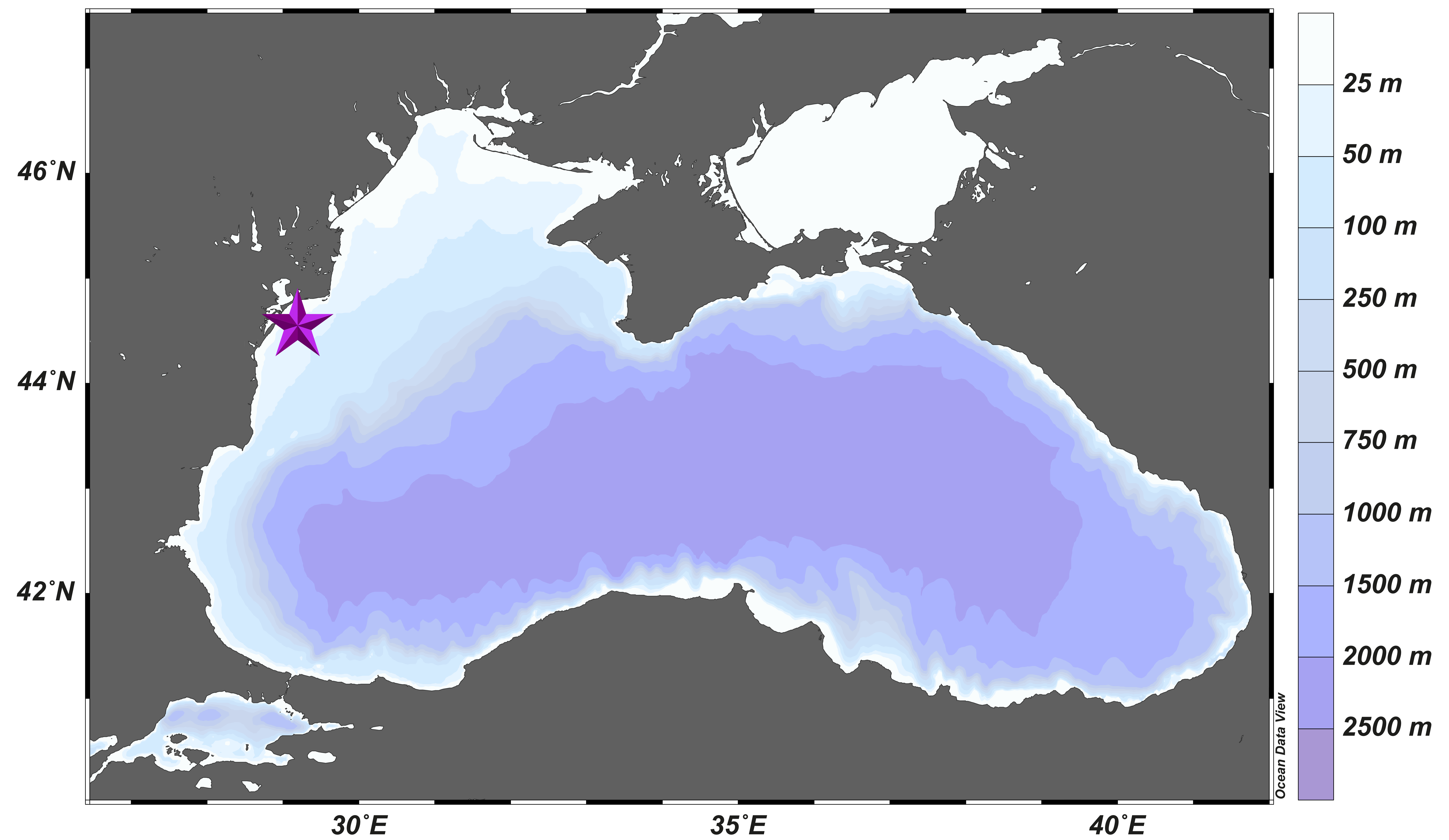
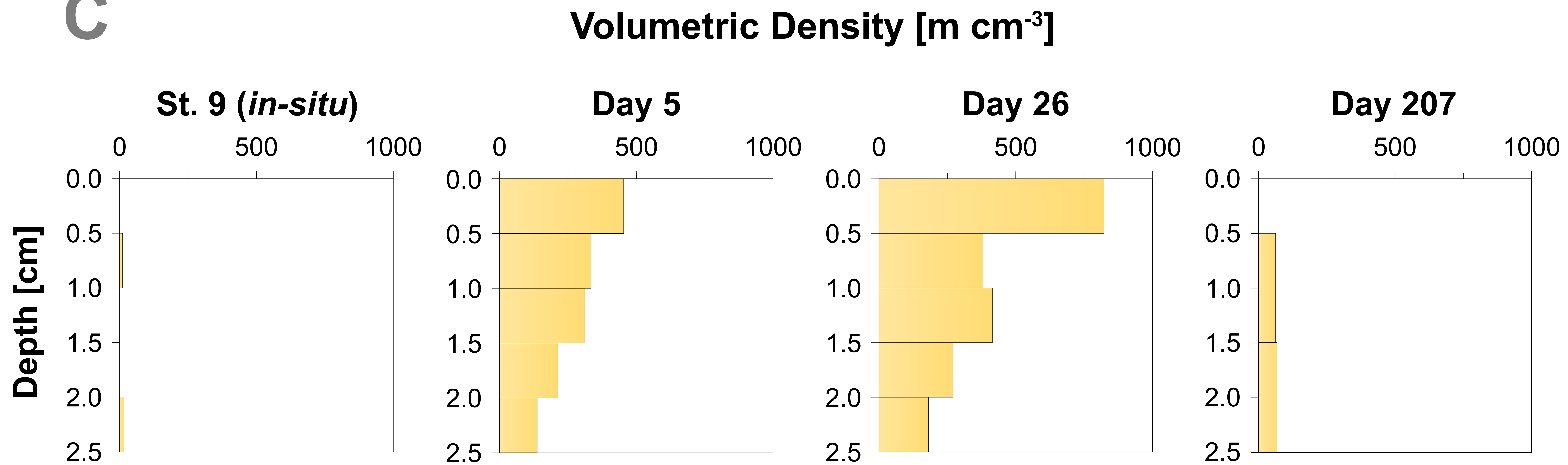
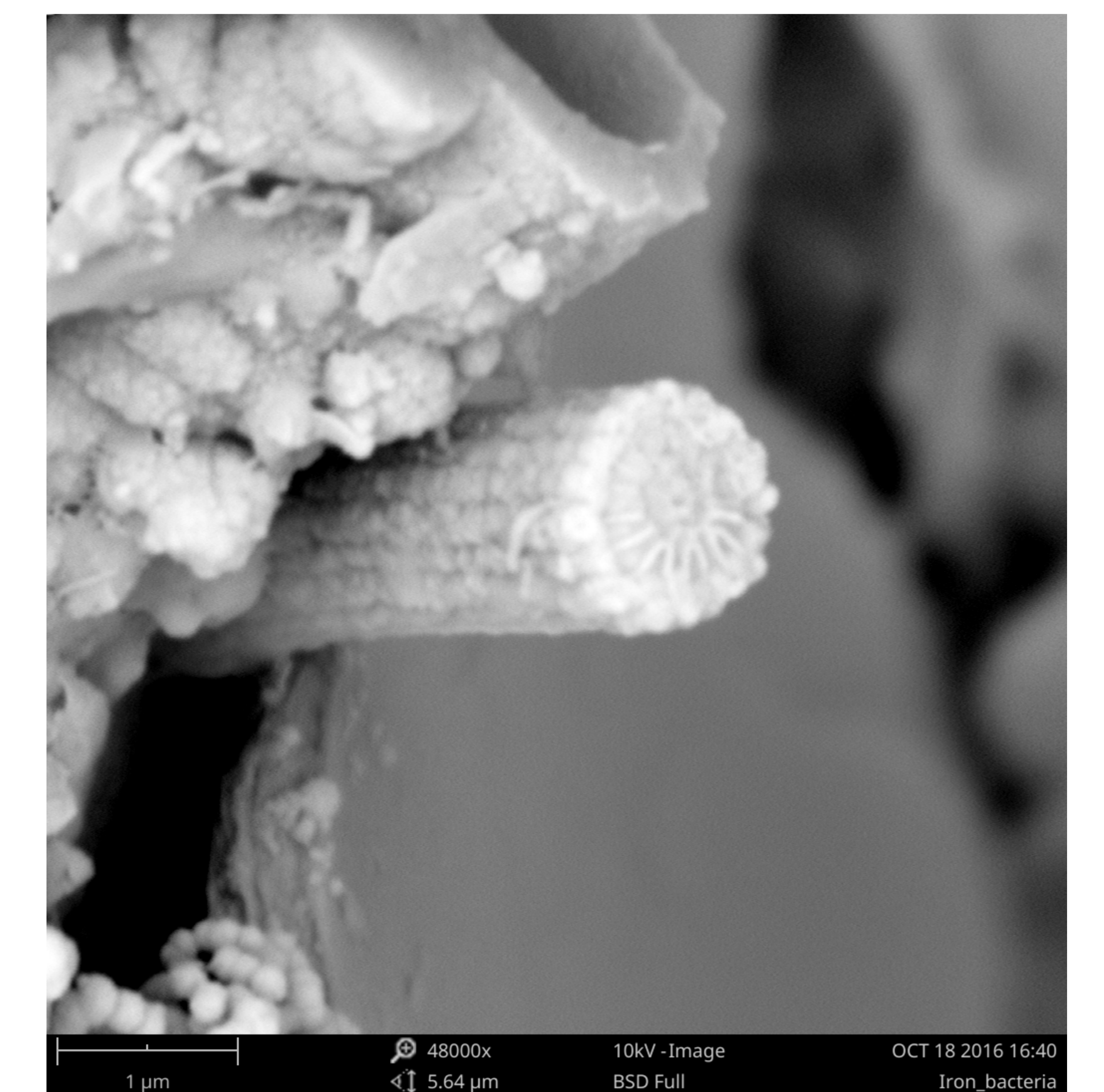
942 **Fig. 8.** High-resolution elemental maps of Fe (red), Mn (green), P (blue) and Ca (white) of surface sediments. These maps  
943 are shown in true vertical orientation and the colours accentuate the relative count intensities adjusted for brightness and  
944 contrast to highlight the features in the sediment. The tick marks represent 1 mm intervals.  $\mu$ XRF maps of the surface  
945 sediment (**A**) from the incubation experiment, (**B**) from the Gulf of Finland at site GOF5 in June (Hermans et al. Submitted),  
946 (**C**) from Lake Grevelingen in January (when cable bacteria become active) and (**D**) from Lake Grevelingen in May  
947 (showing the effects of bioturbation as described in Seitaj et al. (2015)).

948

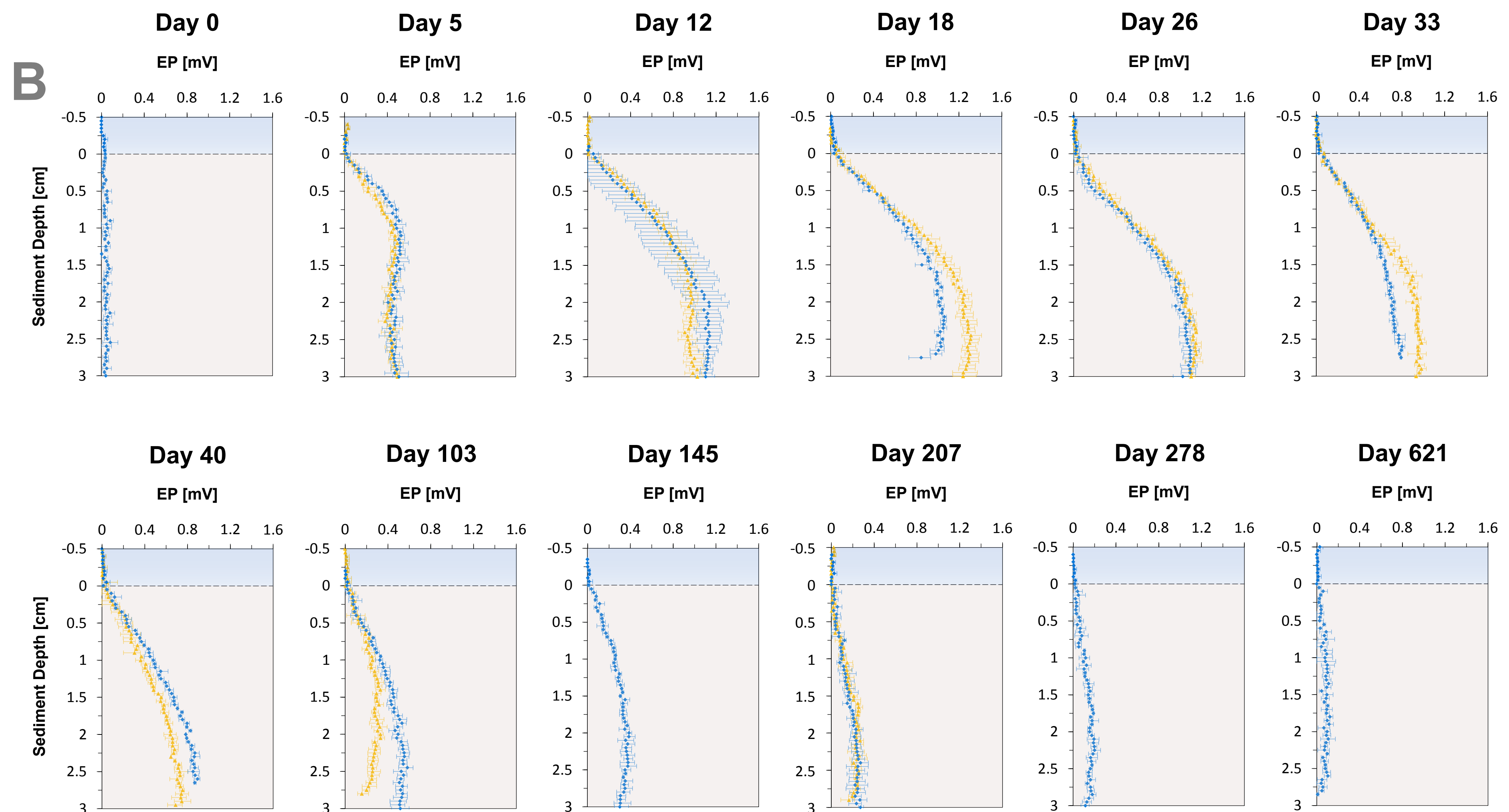
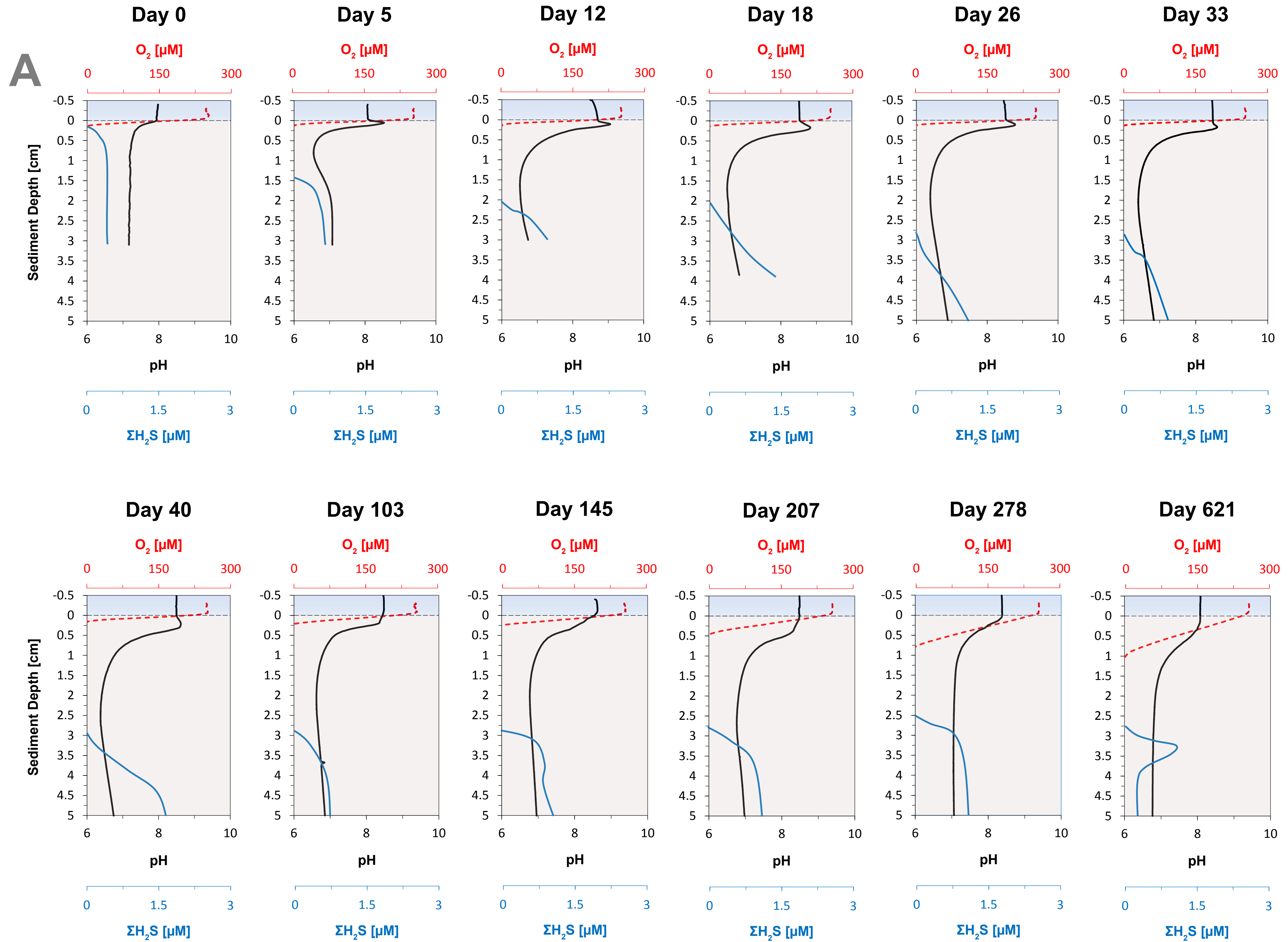
949 **Fig. 9.** The relationship between the diffusive uptake of  $O_2$  ( $mmol O_2 m^{-2} d^{-1}$ ) and the current density of long-distance  
950 electron transport ( $mmol e^{-} m^{-2} d^{-1}$ ). Red triangles are data for days 0 and 5. Green diamonds are data for all other time  
951 points. The blue line represents the expected correlation between the cathodic  $O_2$  consumption rate and the current density  
952 assuming a 1:4 ratio (Nielsen et al. 2010). Here, a positive value indicates an upward flux, whereas a negative value  
953 represents a downward flux.

954

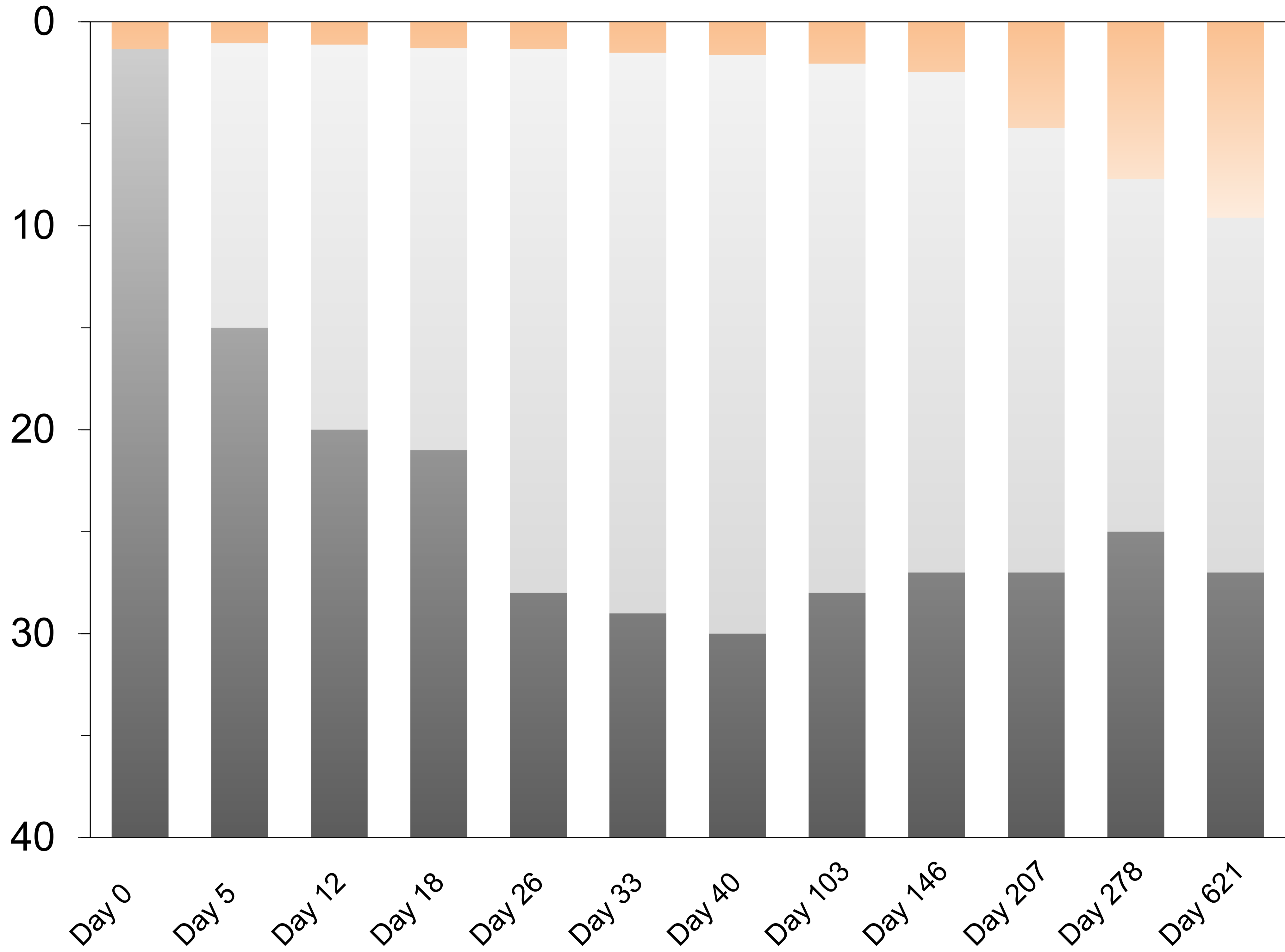
955 **Fig. 10.** Time-series of the depth integrated (0-5 cm) increase in Fe oxides (red) and the depletion of FeS (black) and siderite  
956 (grey) in  $mmol m^{-2}$ . Negative values represent a decrease, whereas positive values indicate an increase in the mineral pools.

**A****B****C****D**



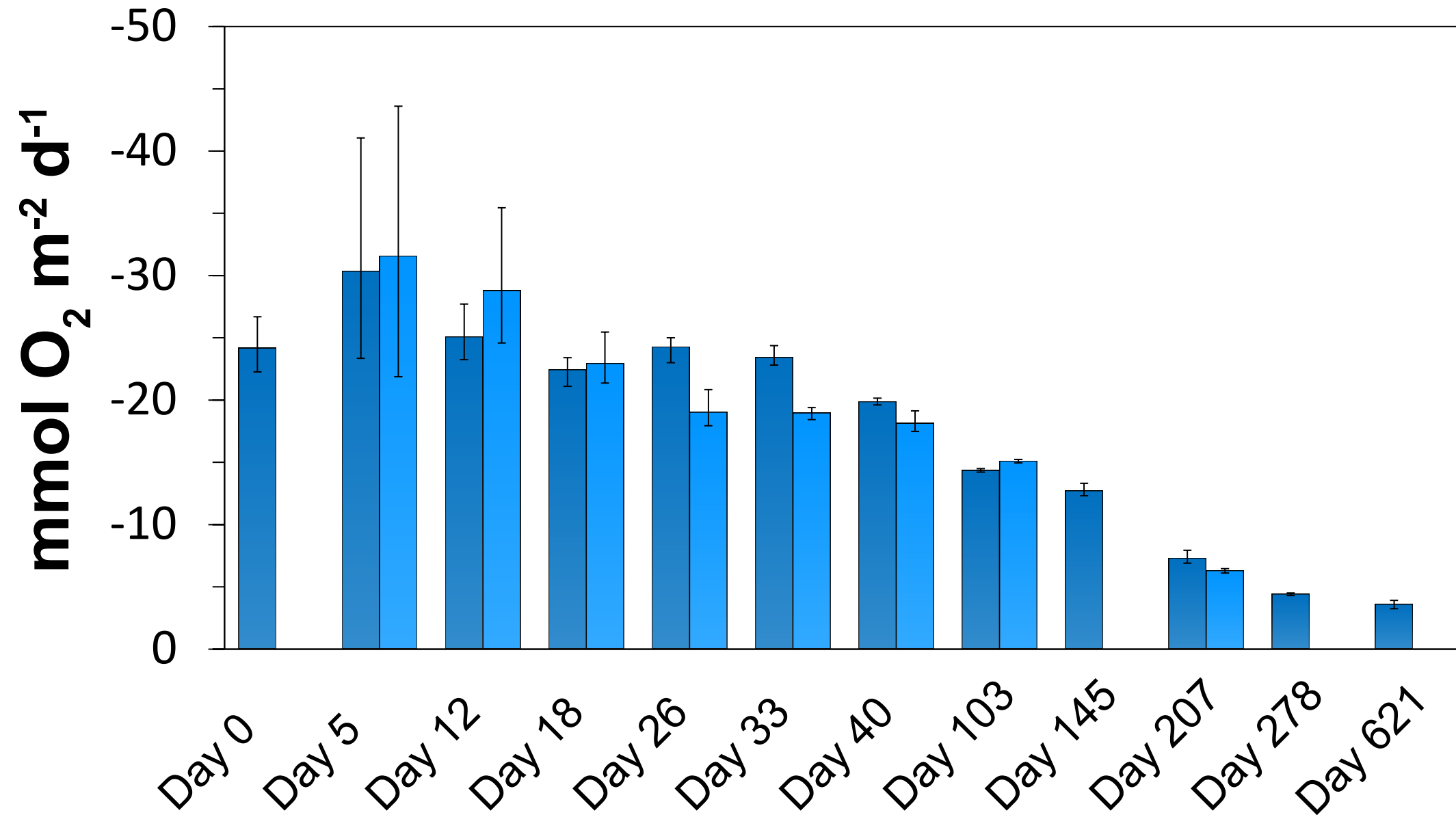


**Depth Zonation [mm]**

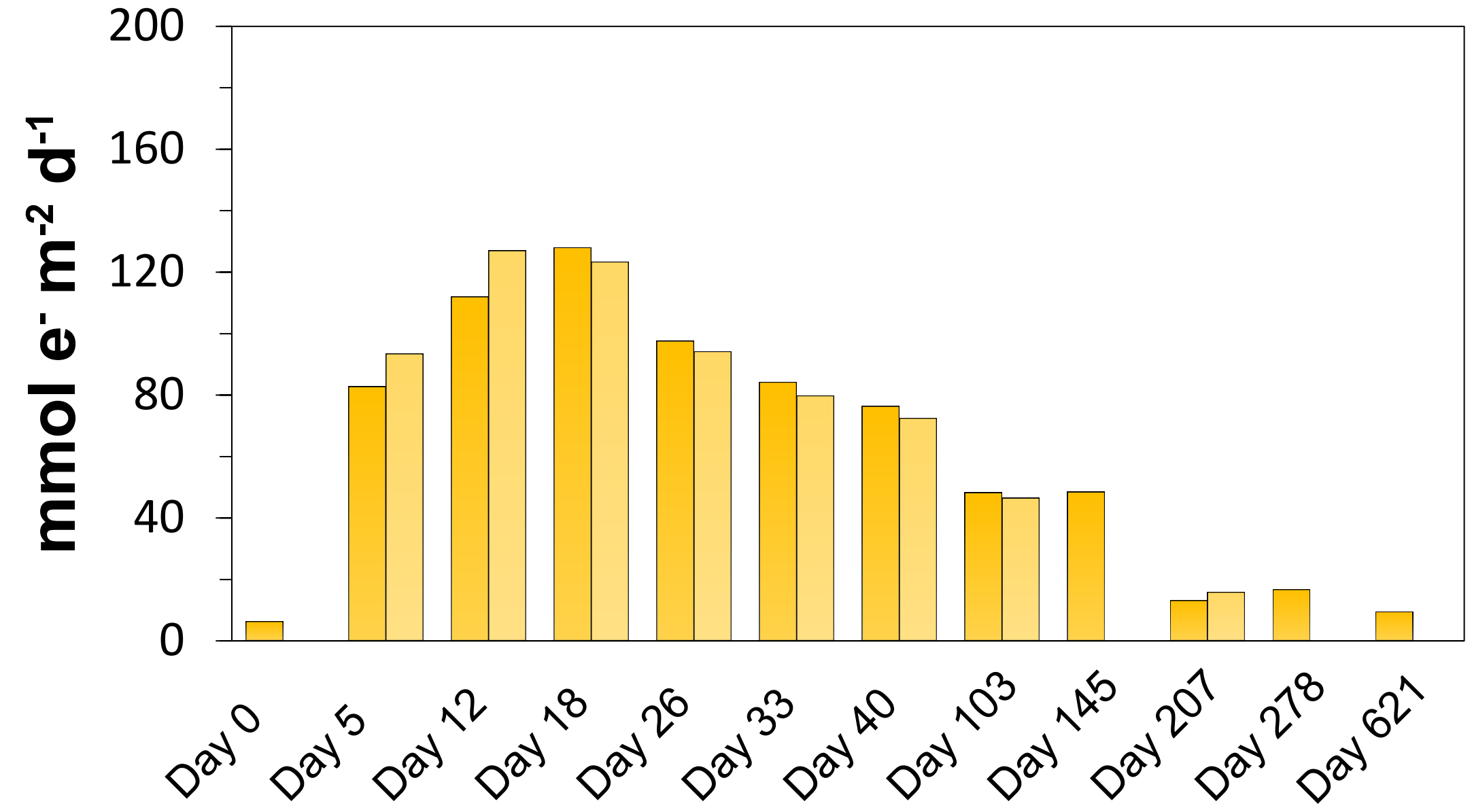


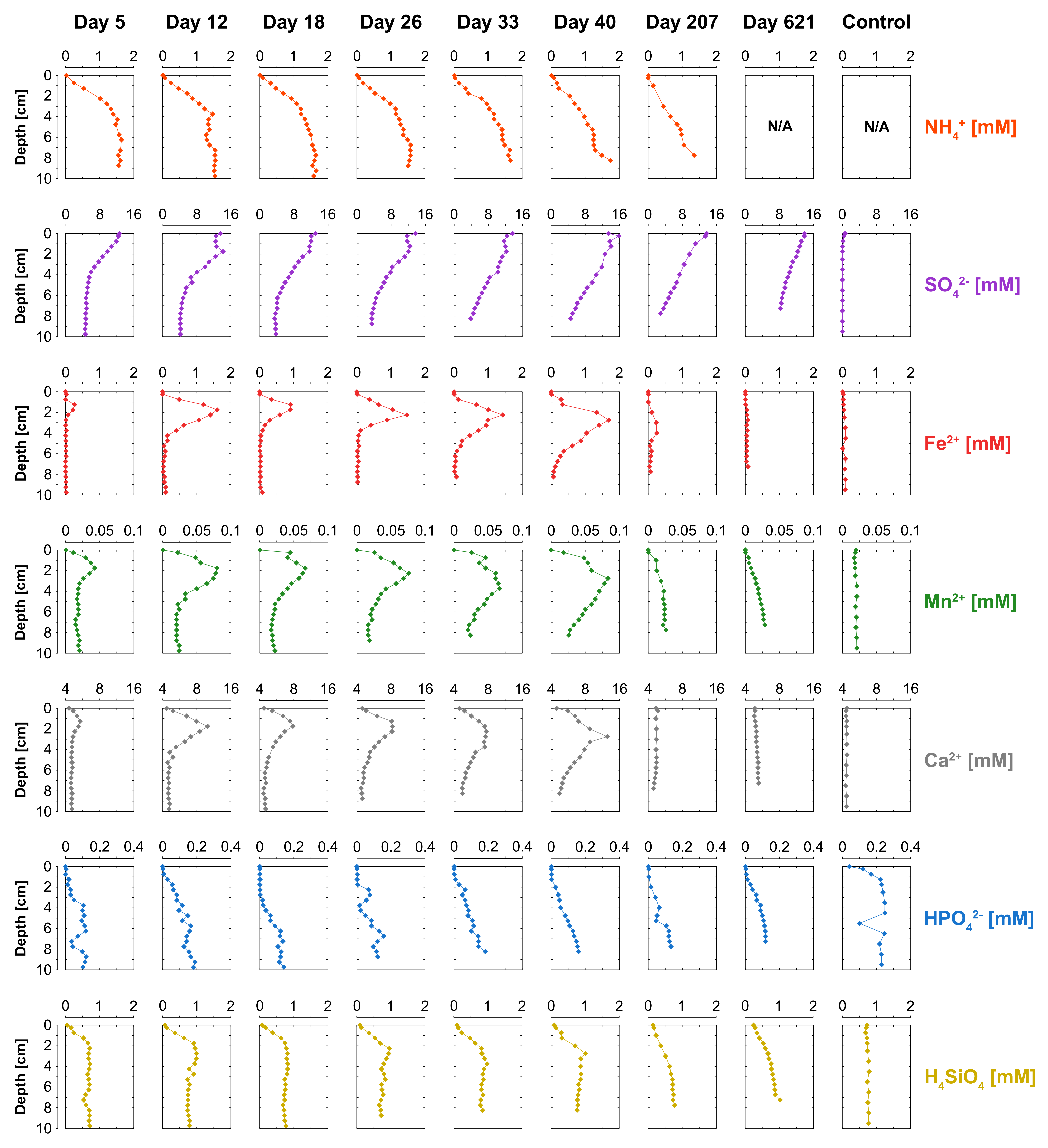
**A**

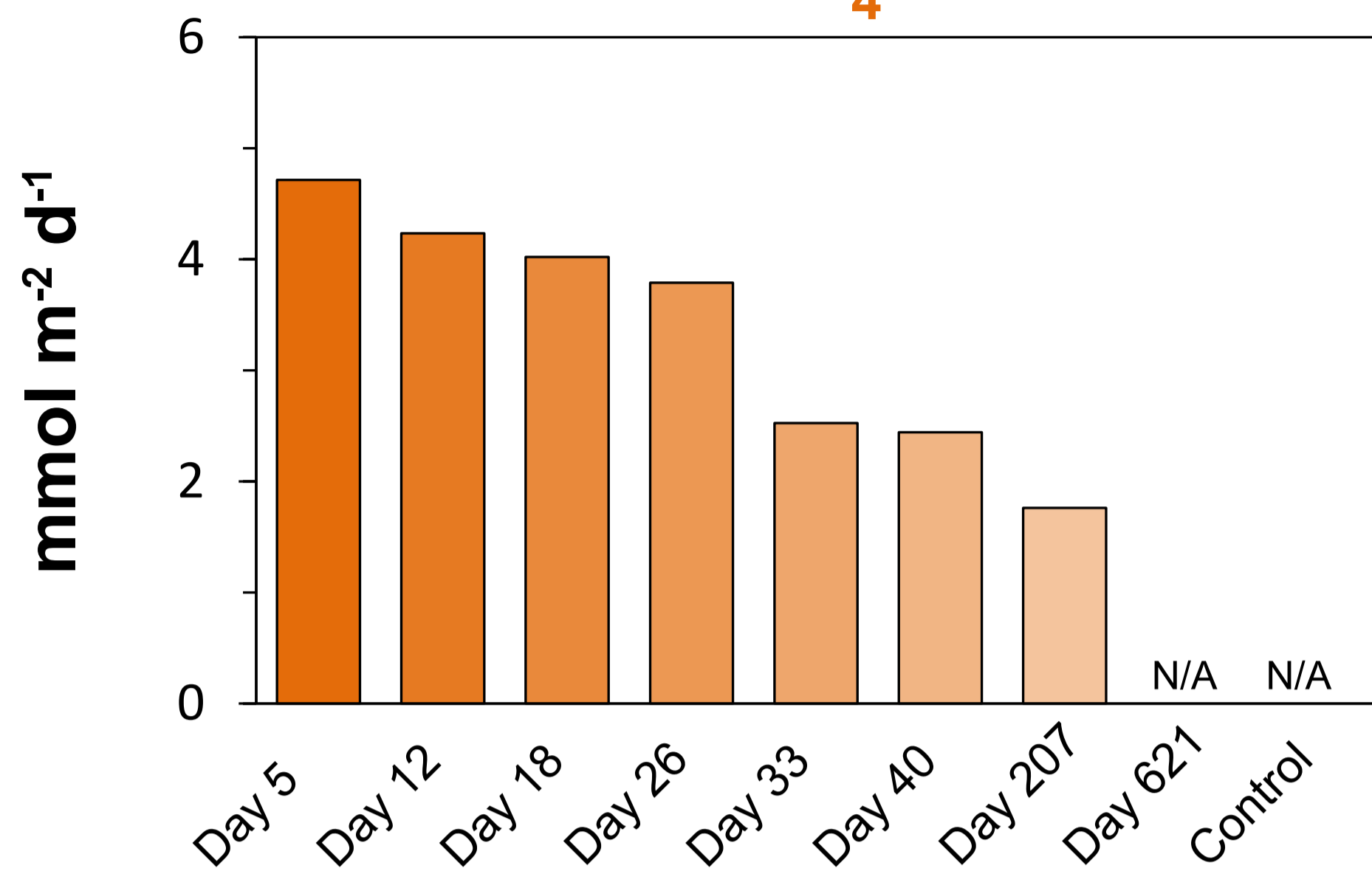
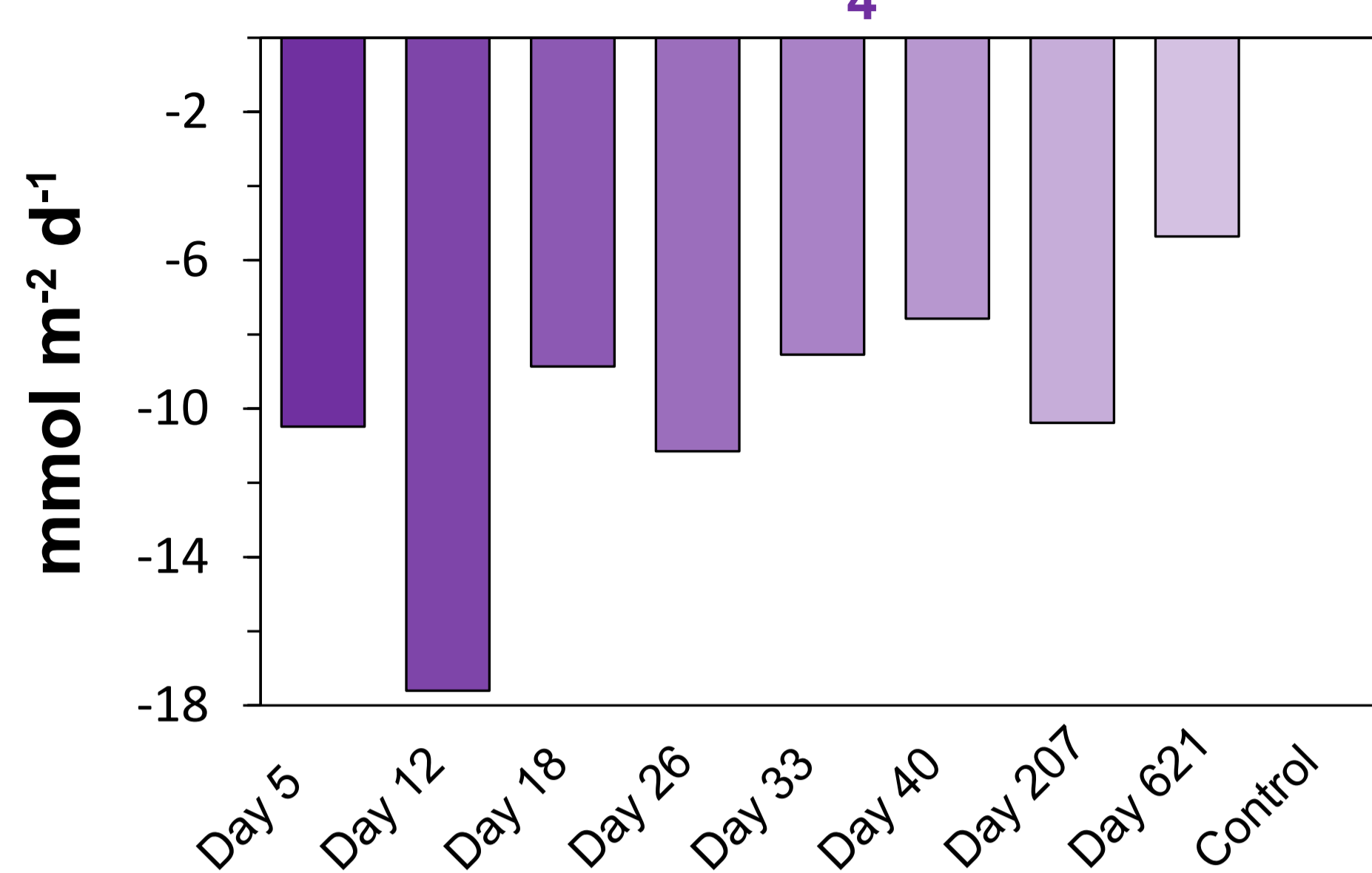
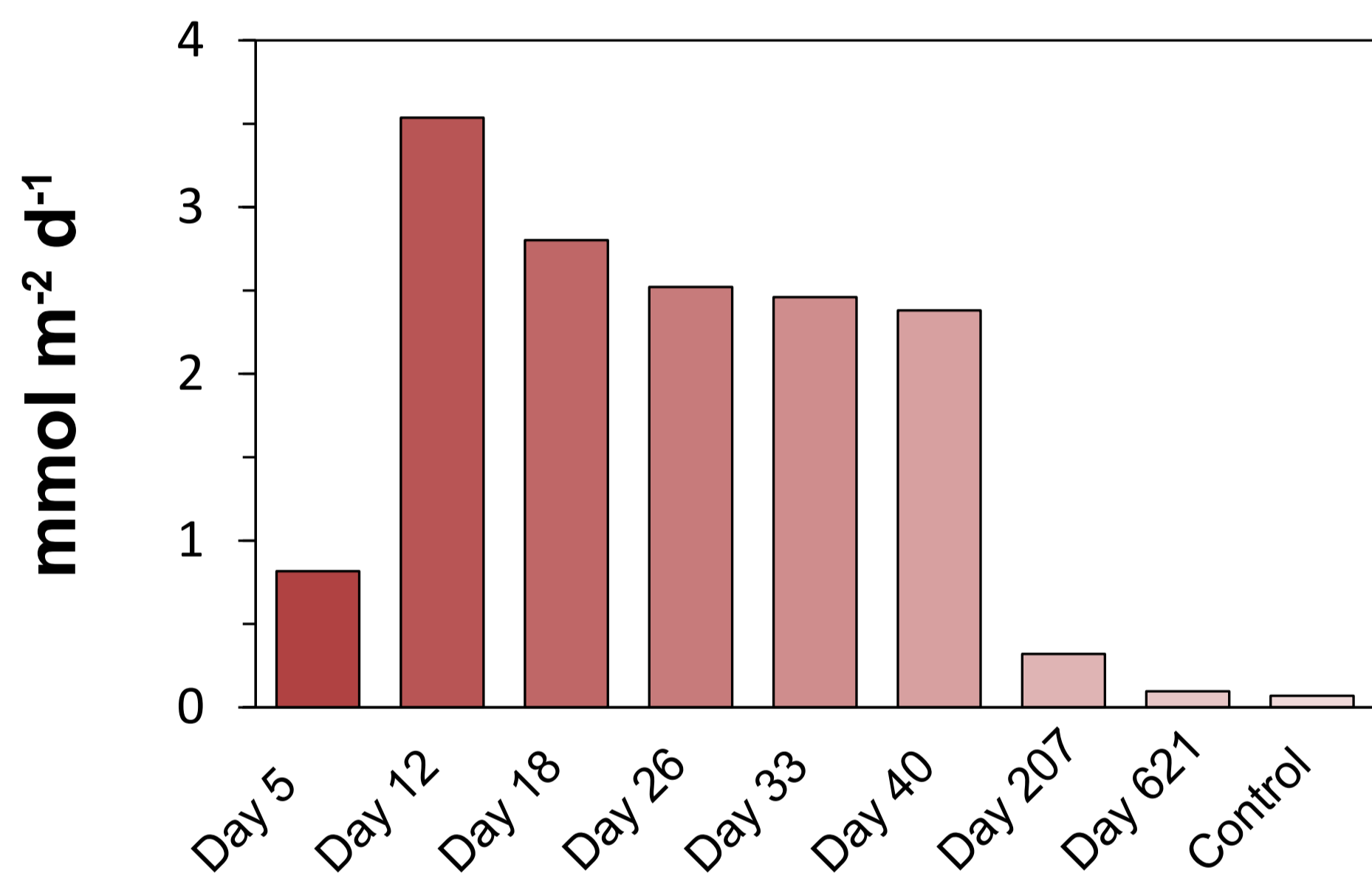
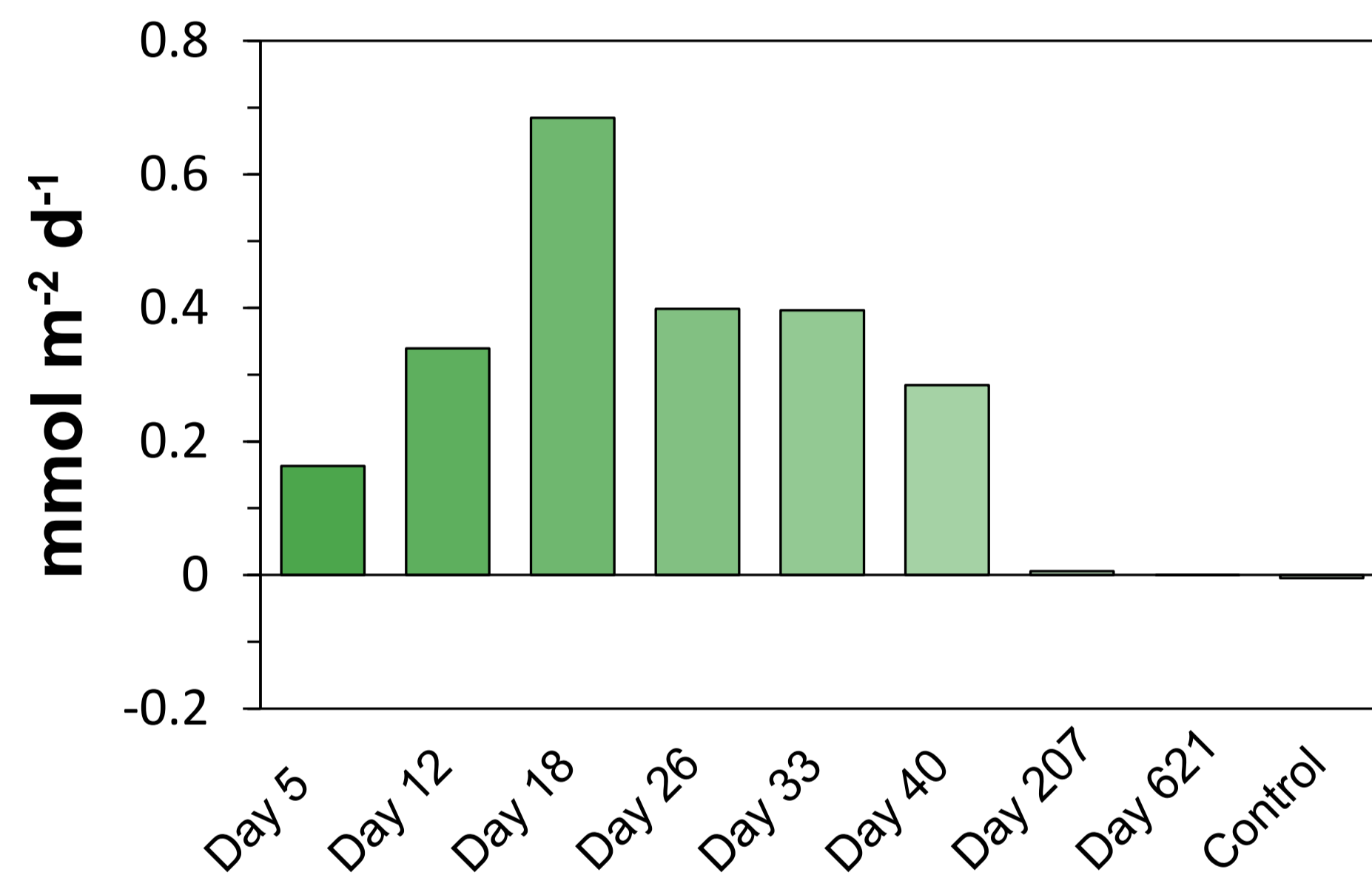
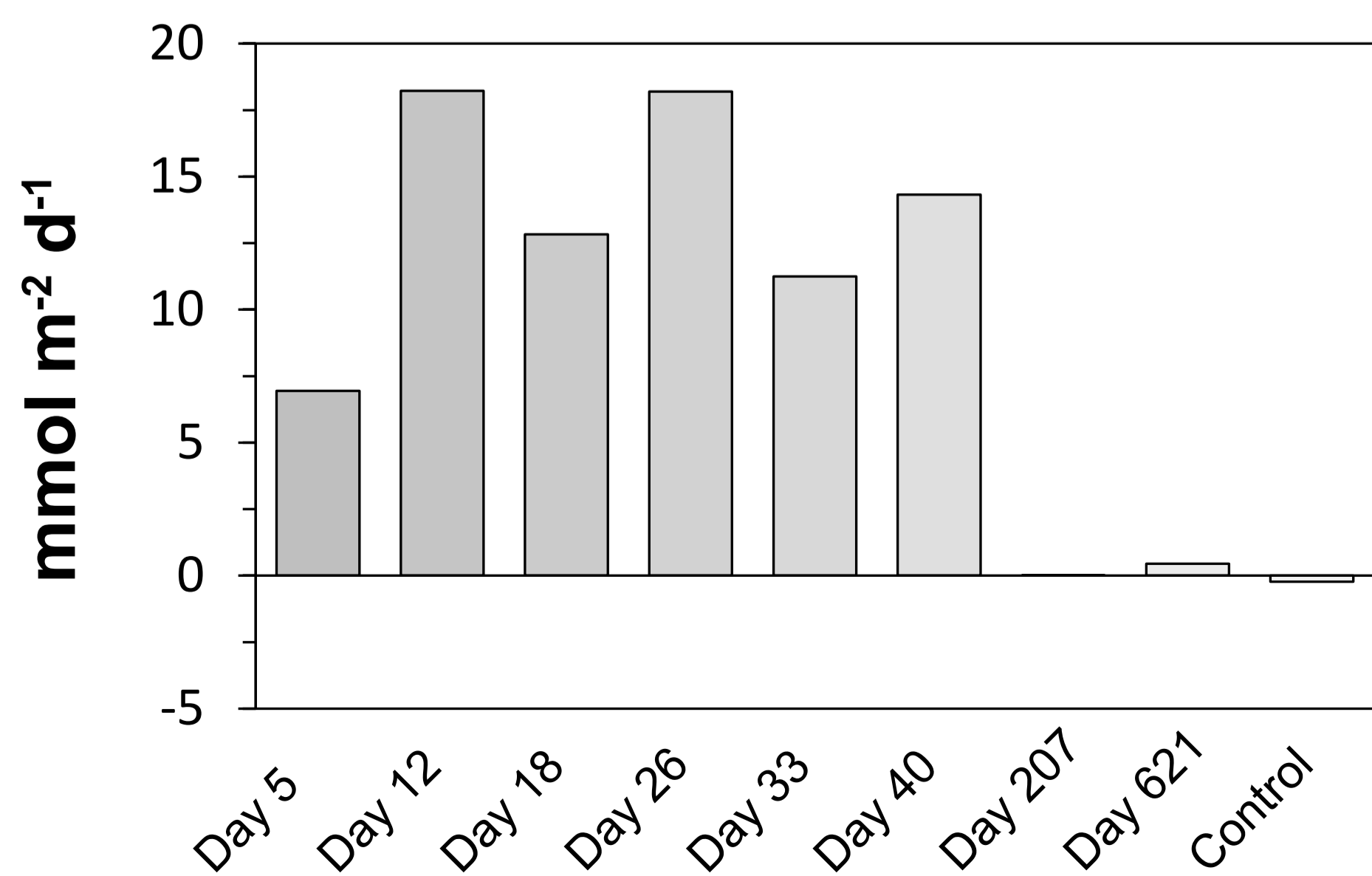
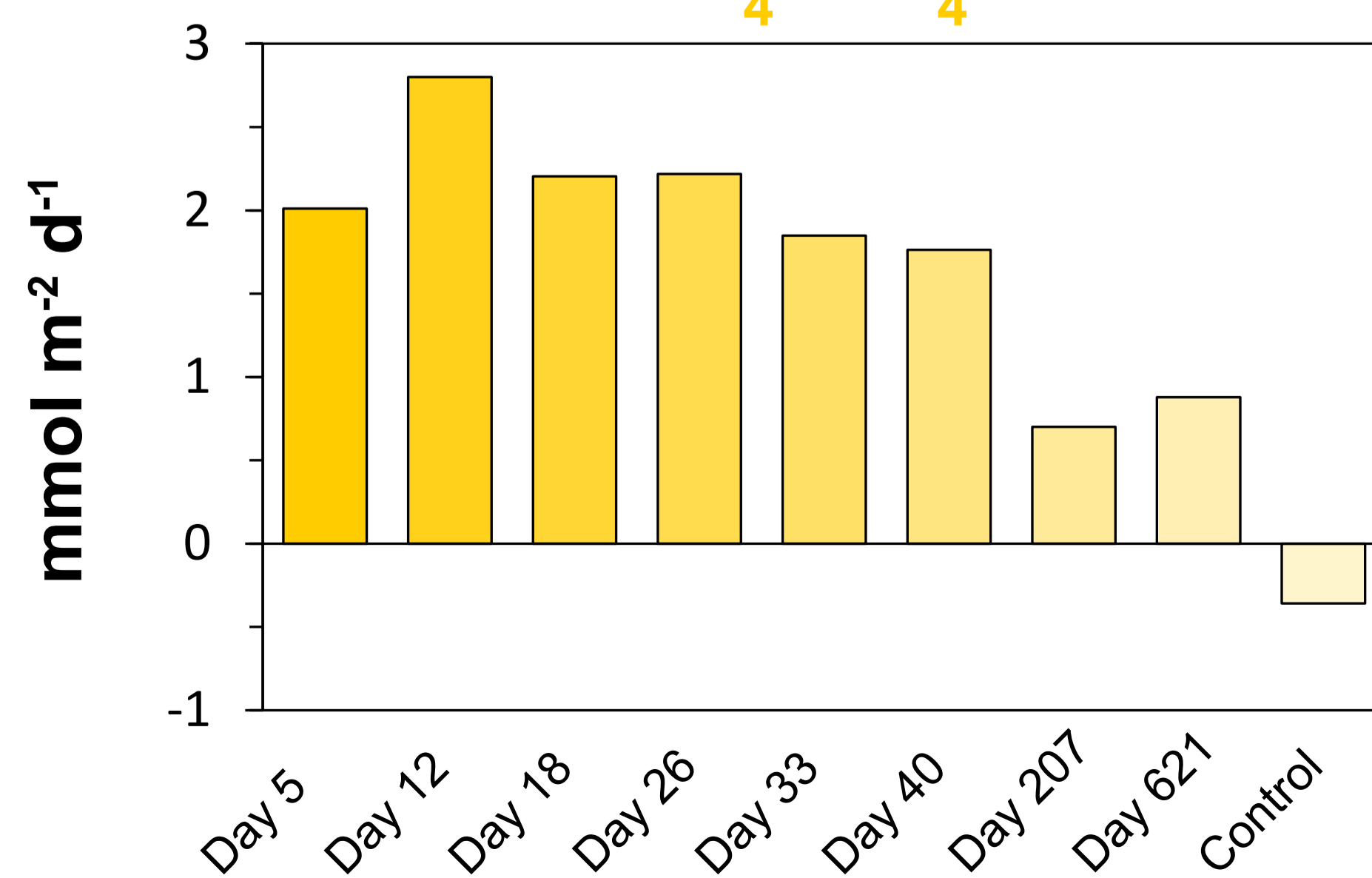
## Diffusive O<sub>2</sub> Uptake

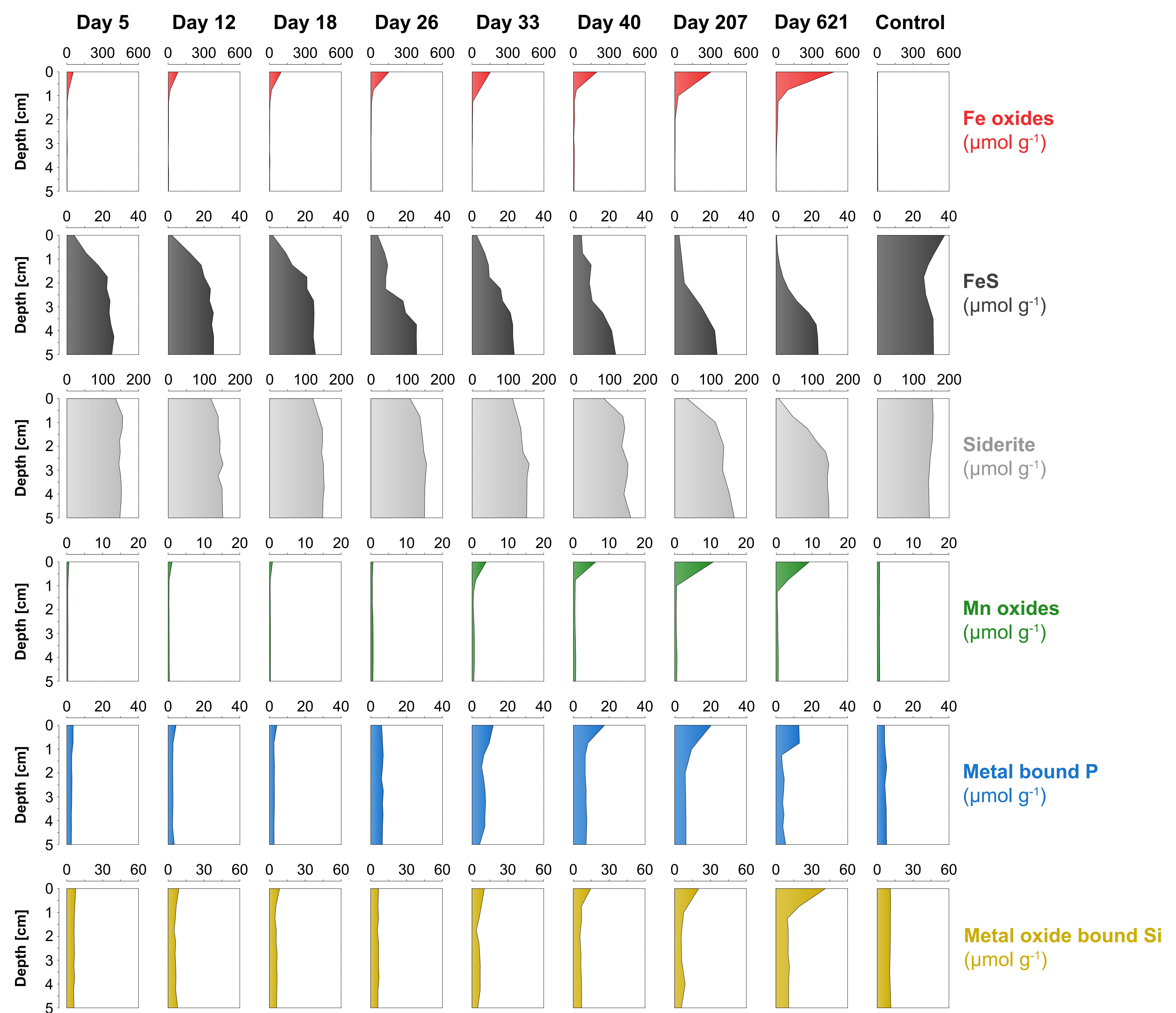
**B**

## Current Density



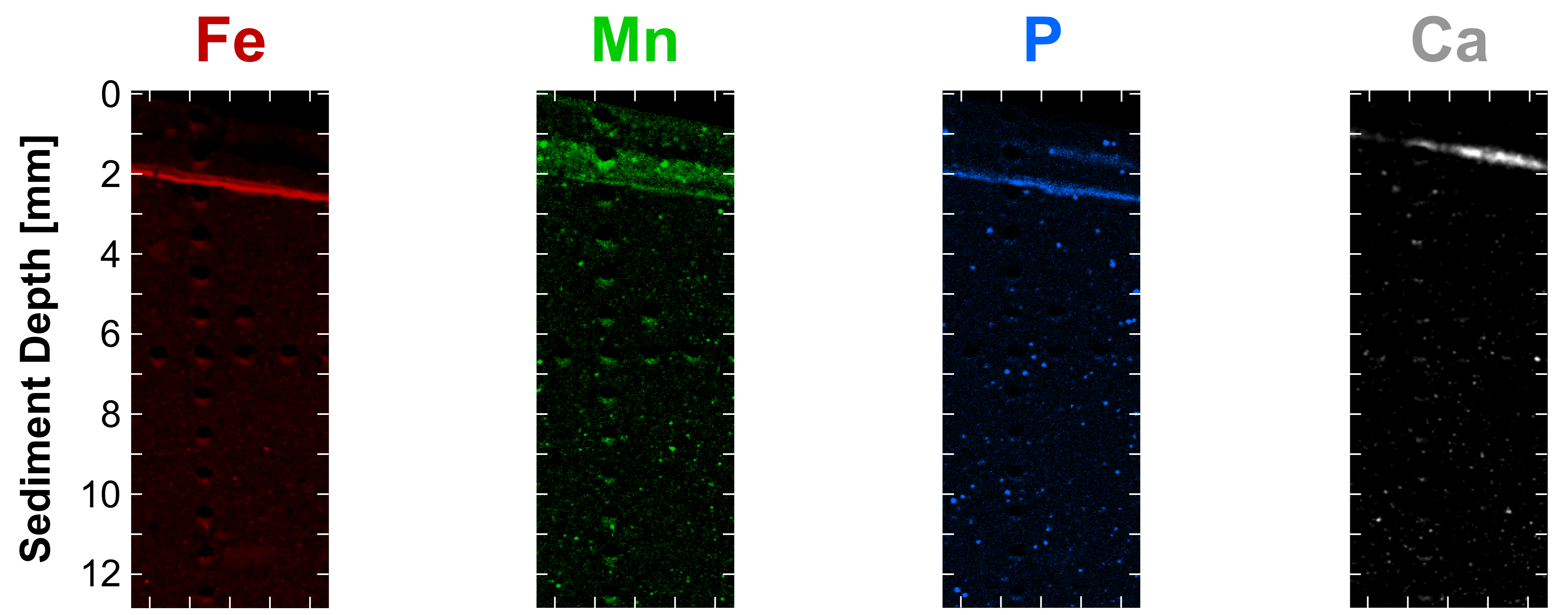


**A****B****C****D****E****F**

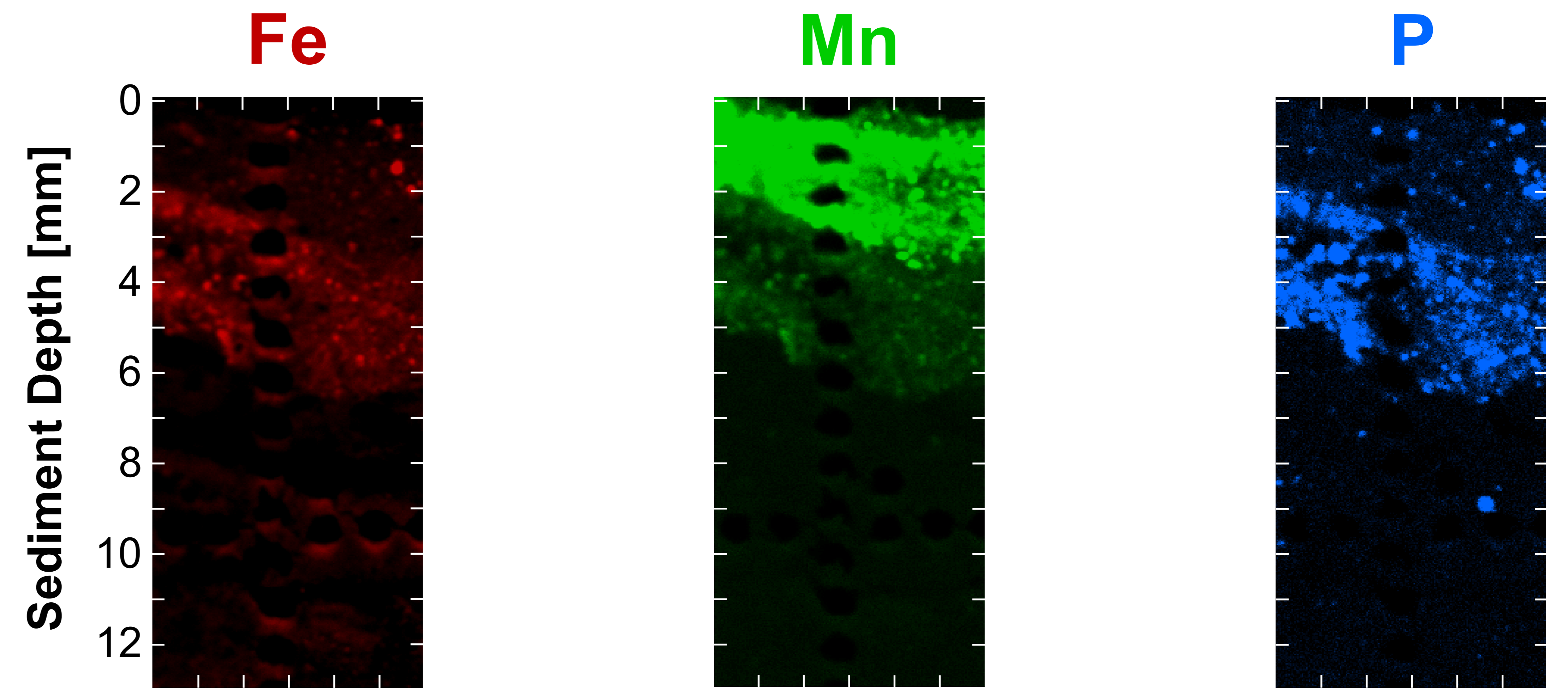


**A**

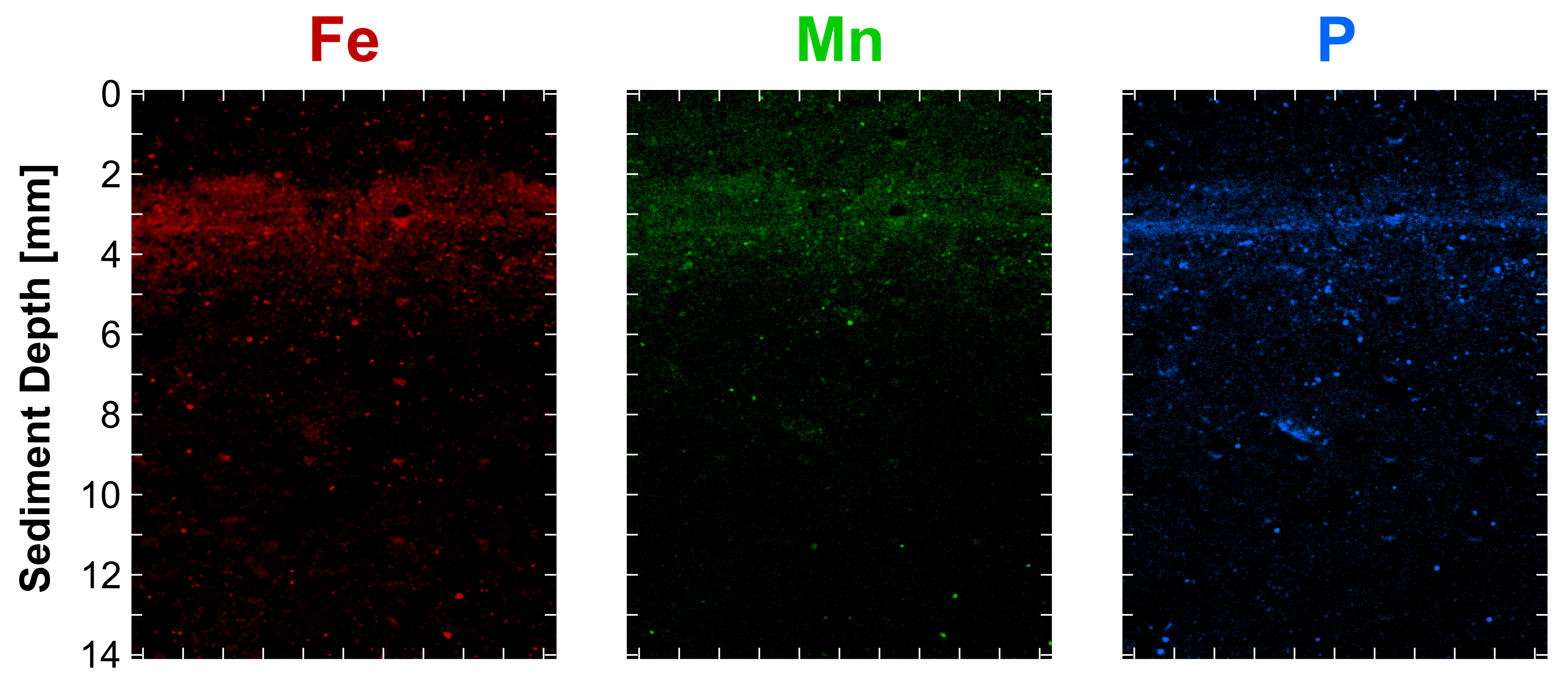
**Black Sea Experiment**  
Cable bacteria

**B**

**Gulf of Finland (GOF5)**  
Cable bacteria

**C**

**Lake Grevelingen in January**  
Cable bacteria

**D**

**Lake Grevelingen in May**  
Macrofauna

