

1 **Different responses of ecosystem respiration, CH₄ uptake, and N₂O emissions to**
2 **seasonally asymmetric warming in an alpine grassland of the Tianshan**
3 **Mountains**

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10 **Abstract:**

11 An experiment was conducted to investigate the effect of seasonally asymmetric
12 warming on ecosystem respiration (*Re*), CH₄ uptake, and N₂O emissions in alpine
13 grassland of the Tianshan Mountains of Central Asia, from October 2016 to
14 September 2019. Our results indicated that the annual mean of *Re*, CH₄, and N₂O
15 fluxes in growing season were 42.83 mg C m⁻² h⁻¹, -41.57 μg C m⁻² h⁻¹, and 4.98 μg N
16 m⁻² h⁻¹, respectively. Furthermore, warming during the non-growing season increased
17 *Re* and CH₄ uptake in both the growing season and non-growing seasons. However,
18 the increase in N₂O emission in the growing season was mainly caused by the
19 warming during the growing season. The *Re*, CH₄ uptake, and N₂O emissions were
20 positively correlated with soil temperature. Our results suggested that *Re*, CH₄ uptake
21 and N₂O emissions were regulated by soil temperature, rather than soil moisture, in
22 the case of seasonally asymmetric warming. In addition, the response rate was defined

23 by the changes in greenhouse gas fluxes driven by warming. In our field experiment,
24 we observed the stimulatory effect of warming during the non-growing season on *Re*
25 and CH₄ uptake. In contrast, the response rates of *Re* and N₂O emissions were
26 gradually attenuated by long-term annual warming, and the response rate of *Re* was
27 also weakened by warming over the growing season. These findings highlight the
28 importance of warming in the non-growing season in regulating greenhouse gas fluxes,
29 a finding which is crucial for improving our understanding of C and N cycles under
30 the scenarios of global warming.

31 **Keywords:** Greenhouse gas flux; Extreme climatic event; Temperature
32 sensitivity; Warming of open-top chambers

33 **1. Introduction**

34 Since the industrial revolution, human activities have intensified global warming.
35 The global surface temperature increased by about 0.85°C from 1880 to 2012, and it
36 is expected that the surface temperature will increase by about 1.1–6.4°C by the end
37 of this century (IPCC, 2007, 2013). The rise in atmospheric temperature over the year
38 is not continuous on the temporal scale, but there is asymmetrical warming across the
39 seasons (Xia et al., 2014). The 3rd and 4rd Assessment Report of the
40 Inter-Governmental Panel on Climate Change (IPCC) proposed that, against the
41 backdrop of global warming, the temperature change shows that the warming
42 amplitude in the winter is greater than that in the summer, with the warming
43 amplitude at high latitude being greater than that at low latitude, and confirmed that
44 the warming shows asymmetric trends on a seasonal scale (Easterling et al., 1997;

45 IPCC, 2001, 2007).

46 Carbon dioxide (CO₂), methane (CH₄), and nitrous oxide (N₂O) are three of the
47 major greenhouse gases (GHGs) in the atmosphere that directly cause global climate
48 warming, with their contributions to global warming being 60 %, 20 %, and 6 %,
49 respectively (IPCC, 2007, 2013). Experimental warming is known to influence
50 ecosystem respiration (*Re*), CH₄ uptake, and N₂O emission (Pärn et al., 2018; Treat et
51 al., 2018; Wang et al., 2019). Information on *Re*, CH₄ uptake, and N₂O emission and
52 their sensitivity to warming, will enhance our understanding of ecosystem C and N
53 cycling processes and improve our predictions of the response of ecosystems to global
54 climate change (Li et al., 2020; Wang et al., 2019).

55 At present, most studies focus on the influence of warming on GHG flux in
56 terrestrial ecosystems during the summer months (Keenan et al., 2014; Li et al., 2011;
57 Yang et al., 2014). Nevertheless, data on the influence of asymmetric warming on the
58 GHG flux on a seasonal scale are scarce. A study of the Alaskan tundra found that
59 summer warming (using open-top chambers to increase air temperatures in the
60 growing season) significantly increased *Re* in the growing season by about 20 %
61 (Natali et al., 2011). Compared with the slight effect of winter warming on the CO₂
62 fluxes in the growing season, warming increased CO₂ fluxes during the snow-covered
63 non-growing season by more than 50% (Natali et al., 2011). Studies have shown that
64 the response of soil CH₄ uptake rates to temperature increases in alpine meadows of
65 the Qinghai-Tibet Plateau are not consistent seasonally, with CH₄ uptake in the
66 non-growing season being more sensitive to temperature (increasing by 162 %) than

67 the corresponding value in the growing season (Lin et al., 2015). A study by Cantarel
68 et al. (2012) in an alpine grassland ecosystem showed that the response of N₂O
69 emission to warming showed clear seasonal differences, with the N₂O emission in the
70 growing season showing significant differences between the warming treatments,
71 whereas the response of N₂O emission to the warming treatments in November was
72 not obvious. A recent study showed that seasonal variations in carbon flux were more
73 closely related to air temperature in the meadow steppe (Zhao et al., 2019). Another
74 study found that experimental warming enhanced CH₄ uptake in the relatively arid
75 alpine steppe, but had no significant effects on CH₄ emission in the moist swamp
76 meadow (Li et al., 2020). Furthermore, soil CH₄ uptake was not significantly affected
77 by warming in the alpine meadow of the Tibetan Plateau (Wu et al., 2020). In contrast,
78 a global meta-analysis showed that experimental warming stimulates C fluxes in
79 grassland ecosystems, and the response of C fluxes to warming strongly varies across
80 the different grassland types, with greater warming responses in cold than in
81 temperate and semi-arid grasslands (Wang et al., 2019). Across the data set, Li et al.
82 (2020) demonstrated that N₂O emissions were significantly enhanced by whole-year
83 warming treatments. In contrast, no significant effects on soil N₂O emissions were
84 observed by in short-season warming.

85 In summary, the GHG flux in terrestrial ecosystems shows significant interannual,
86 and seasonal variations, and its response to warming also varies over different
87 temporal scales. After long-term uniform warming, the biotic and abiotic factors have
88 adapted to the temperature increase, and the GHG flux response to increasing

89 temperature is smaller than that in the early stages of warming. For example, over
90 longer time periods of warming, accelerated carbon decomposition and increased
91 plant N uptake may decrease soil organic C and N pools (Wu et al., 2012), and the
92 microbial community with variable C use efficiency may reduce the temperature
93 sensitivity of heterotrophic respiration (Zhou et al., 2012). Moreover, climate
94 warming is often unstable, with most of it occurring as extreme events (Jentsch et al.,
95 2007). The heterogeneity of warming may change the adaptability of GHG fluxes to
96 warming, and thus affect the carbon and nitrogen cycles in terrestrial ecosystems.
97 Therefore, we hypothesize that warming in the non-growing season will stimulate
98 GHG flux (especially during the non-growing season) in the alpine steppe. However,
99 continuous warming throughout the year and during the growing season will reduce
100 the sensitivity of GHG flux to warming. This current short communication will help
101 to assess this variation with respect to GHG flux response to increasing temperatures
102 against the backdrop of global climate change, by carrying out seasonally
103 asymmetrical warming studies in alpine grasslands.

104 **2. Materials and methods**

105 The experiment was conducted from October 2016 to September 2019 at the
106 Bayinbuluk Grassland Ecosystem Research Station, Chinese Academy of Sciences
107 (42°52.76' ~ 42°53.17' N, 83°41.90' ~ 83°43.12' E, 2460 m above sea level), which is
108 located in the southern Tianshan mountains of Central Asia, Xinjiang Uyghur
109 Autonomous Region of China. Permafrost is present in the Bayinbuluk alpine
110 grassland, with the average maximum frozen depth (from 2000 to 2011, Zhang et al.,

111 2018) being more than 250 cm. The mean annual temperature was -4.8°C per decade,
112 with the lowest monthly temperature in January (-27.4°C) and the highest in July
113 (11.2°C), and the mean annual precipitation amounted to 265.7 mm, with 78.1%
114 occurring during the growing season, from June to September (Geng et al., 2019).
115 Variations in soil temperature, soil moisture, air temperature and precipitation are
116 shown in Figure S1, S2, S3 and S4, respectively. Ungrazed since 2005, all the plots
117 were dominated by *Stipa purpurea*, *Festuca ovina*, *Oxytropis glabra*, and *Potentilla*
118 *multifida*. The soil was sub-alpine steppe soil, the parent material of the soil was
119 Loess, and the average annual soil moisture was 5.9 % (2017-2019).

120 The open-top chambers (OTCs) were made of 5 mm thick tempered glass. To
121 reduce the impact of precipitation and snow, the OTC was constructed with a
122 hexagonal round table which was 100 cm high, and the diagonals of the bottom and
123 top were 100 cm and 60 cm, respectively. Four treatments were simulated using OTCs:
124 warming throughout the year (AW), warming in the non-growing season (October to
125 May) only (NGW), warming in the growing season (June to September) only (GW)
126 and no warming (NW). Three replicate plots were established for each treatment, each
127 plot measuring $1\text{ m} \times 1\text{ m}$, with a 3-m wide buffer zone between adjacent plots,
128 making a total of 12 plots. Soil temperature and soil moisture were measured at 10 cm
129 depth by an outdoor temperature and humidity data recorder (HOBO U23-001; Onset
130 Computer Corporation, Bourne, USA).

131 *Re*, CH_4 and N_2O fluxes were measured using static chambers, made of PVC
132 tubing with diameter 0.25 m and height 0.17 m, with one chamber in each of the 12

133 plots. Gas samples were taken 0, 10, 20 and 30 minutes after the lid of the static
134 chamber was sealed in between 12:00 and 14:00 (GMT + 8) every day. The rates of
135 ecosystem respiration, CH₄ and N₂O fluxes were calculated based on the change in
136 concentration of CO₂, N₂O and CH₄ in each chamber over time by a linear or
137 non-linear equation ($P < 0.05$, $r^2 > 0.95$) (the positive flux values represent emission,
138 and the negative flux values represent uptake; Liu et al. 2012; Wang et al. 2013). A
139 total of 232 samples were taken, collecting once or twice a week. Concentrations of
140 individual gases in samples were measured using a gas chromatograph (GC) (Agilent
141 7890A; Agilent Technologies, Santa Clara, CA, USA).

142 Effects of seasonally asymmetric warming on R_e , CH₄ uptake, and N₂O
143 emissions were analyzed by two-way repeated-measures analysis of variance
144 (ANOVA). One-way ANOVA was used to compare soil temperature differences.
145 General linear analyses were used to identify significant linear correlations and
146 regressions between soil temperature variation and the responses of R_e , CH₄ uptake,
147 or N₂O emissions. The natural logarithm of the response ratio (RR) was used to reflect
148 the effects of seasonally asymmetric warming on alpine grassland GHG fluxes
149 (Hedges et al., 1999). The RR is the ratio of the mean value of the chosen variable in
150 the warming group (\bar{W}_t) to that in the control group (NW; \bar{W}_c), and is an index of
151 the effect of seasonally asymmetric warming on the corresponding variable (Eq. 1).
152 All statistical analyses were conducted using SPSS (version 20.0) (IBM, Armonk, NY,
153 USA) with the statistically significant difference threshold set at $P < 0.05$.

$$RR = \ln\left(\frac{\overline{W}_t}{\overline{W}_c}\right) = \ln(\overline{W}_t) - \ln(\overline{W}_c) \quad (1)$$

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3. Results

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Our study showed that the Bayinbuluk alpine grassland exhibited a low Re , was a net CH_4 sink, and a negligible N_2O source. The annual mean values of Re , CH_4 uptake, and N_2O emissions in the growing season were $42.83 \text{ mg C m}^{-2} \text{ h}^{-1}$, $41.57 \text{ } \mu\text{g C m}^{-2} \text{ h}^{-1}$, and $4.98 \text{ } \mu\text{g N m}^{-2} \text{ h}^{-1}$, respectively, from October 2016 to September 2019 (Figure 1). Compared with the control group (NW), the Re was decreased by 7.5% and 4.0% in the growing season and non-growing season, respectively, under AW and decreased by 2.4% and 8.5% under GW in the growing season and non-growing season, respectively. However, compared with the control group, the Re under NGW increased by 7.9% and 10.5% in the growing season and non-growing season, respectively, averaged over the three years (Figure 2 a). The AW temperature change induced a 6.4% increase in CH_4 uptake in the growing season and a 3.8% decrease in the non-growing season. The GW treatment resulted in 7.1% and 10.2% increases in CH_4 uptake in the growing season and non-growing season, respectively. On the contrary, the NGW generated a 10.6% and 9.2 % decrease in CH_4 uptake in the growing season and non-growing season, respectively (Figure 2 b). The AW and NGW treatments resulted in 5.8% and 2.2 % decreases, respectively, in N_2O emission in the growing season, and 101.9% and 192.3% increases, respectively, in N_2O emission in the non-growing season. Compared with the control, NW group, the N_2O emission increased by 29.7% and decreased by 24.4% under GW in the growing season and non-growing season, respectively (Figure 2 c). One-way ANOVA results

176 of Re , CH_4 uptake and N_2O emissions among the four warming treatments were not
177 significant, with the exception that the soil CH_4 uptake in the growing season 2019
178 under GW treatment was significantly higher than that of the AW and NGW
179 treatments ($P < 0.05$).

180 The results of two-way repeated measures ANOVA showed significant
181 interannual differences of Re in the growing season ($P < 0.05$, Figure 1 a), whereas
182 the CH_4 uptake under the warming treatment exhibited significant differences in the
183 growing season ($P < 0.01$; Figure 1 b), and the interannual N_2O emission showed
184 significant differences in both the growing season and non-growing season ($P < 0.05$,
185 Figure 1 c). Therefore, interannual variation was larger than the impact of the
186 warming treatment (for Re and N_2O emissions, Figure 1), whereas the warming
187 treatment had a significant impact on CH_4 uptake. Under the four warming treatments,
188 Re was significantly positively linearly correlated with soil temperature ($P < 0.05$;
189 Figure S5 a). we observed increasing CH_4 uptake with increasing soil temperature (P
190 < 0.05 ; Figure S5 b). On the other hand, the N_2O emission showed a significantly
191 positive linear correlation with soil temperature, but only under NGW ($P < 0.05$;
192 Figure S5 c).

193 **4. Discussion**

194 Our study found that the response rate of Re to temperature significantly
195 decreased with the increase in soil temperature (ΔST_{AW} and ΔST_{GW}) under AW and
196 GW treatments, respectively (Figure 3 a, c; $P < 0.05$). This finding indicated that the
197 response of Re to soil temperature became less and less sensitive to soil temperature

198 with warming throughout the year (or the growing season) in the alpine grasslands.
199 On the contrary, NGW significantly increased the response rate of R_e to temperature
200 change (ΔST_{NGW}), indicating that warming in the non-growing season amplified the
201 sensitivity of R_e to temperature change (Figure 3 b, $P < 0.05$). In addition, Zou et al.
202 (2018) showed that the soil fluxes of CO_2 increased exponentially with increasing
203 temperature, but warming decreased the temperature sensitivity by 23% in the
204 grassland. Furthermore, Natali et al. (2011) also confirmed that, compared with the
205 CO_2 flux in the growing season, the CO_2 flux in the nongrowing season was more
206 sensitive to the temperature increase.

207 Ecosystem CH_4 flux is the net result of CH_4 production and consumption,
208 occurring simultaneously under the action of methanogenic archaea and
209 methane-oxidizing bacteria (e.g., Mer and Roger, 2001). In addition, our results
210 demonstrated that warming increased CH_4 uptake in the growing season, but
211 decreased CH_4 uptake in the non-growing season in the alpine grassland, findings
212 similar to those from other grassland ecosystems (Lin et al., 2015; Wu et al., 2020;
213 Zhu et al., 2015). Our results also demonstrated that seasonally asymmetric warming
214 did not significantly affect the response rate of CH_4 uptake (Figure 3 d-f, $P > 0.05$).
215 CH_4 flux depended on temperature, pH, and the availability of substrate (e.g., Treat et
216 al., 2015). The CH_4 uptake observed during the three growing season and
217 non-growing season implied that the alpine grassland soil could act as an atmospheric
218 CH_4 sink, a finding which agrees with the results of many previous studies in similar
219 regions (Wei et al., 2015; Zhao et al., 2017). Hu et al. (2016) suggested that
220 asymmetrical responses of CH_4 fluxes to warming and cooling should be taken into
221 account when evaluating the effects of climate change on CH_4 uptake in the alpine

222 meadow on the Tibetan plateau. Unlike CH₄ flux in alpine grasslands, Treat et al.
223 (2018) confirmed that wetland was a small CH₄ source in the non-growing season,
224 whereas uplands varied from CH₄ sinks to CH₄ sources. The latest research confirmed
225 that warming in the Arctic had become more apparent in the non-growing season than
226 in the typical growing season (Bao et al., 2020). Hereby, Bao et al. (2020) found that
227 the CH₄ emissions during the spring thaw and the autumn freeze contributed
228 approximately one-quarter of the annual total CH₄ emissions. That experimental
229 warming is stimulating soil CH₄ uptake in the growing season implies that the
230 grasslands of the Bayinbuluk may have the potential to remove more CH₄ from the
231 atmosphere under future global warming conditions.

232 Furthermore, with the increased variation in soil temperature, the response rate of
233 N₂O emission gradually decreased under AW treatment (Figure 3 g, $P < 0.05$). Our
234 results suggested that the response of N₂O emission to temperature increase was
235 limited by the warming that occurred throughout the year. However, our results
236 displayed N₂O emission peaks during the freeze–thaw periods (e.g., May 2017, June
237 2018 and April 2019). Warming increased N₂O emissions in the thawing period due to
238 disruption of the gas diffusion barrier and greater C and N availability for microbial
239 activity (Nyborg et al., 1997). Wagner-Riddle et al. (2017) also demonstrated that the
240 magnitude of the freeze/thaw-induced N₂O emissions was associated with the number
241 of days with soil temperatures below 0°C. Pärn et al. (2018) found that N₂O emission
242 from organic soils increases with rising soil NO₃⁻, follows a bell-shaped distribution
243 with soil moisture. Another study has shown that a whole - year warming treatment
244 significantly increased N₂O emissions, but daytime, night-time or short - season

245 warming did not have significant effects (Li et al., 2020). In addition, Cantarel et al.
246 (2010) suggested that the N₂O flux from cool and upland grasslands may be driven
247 primarily by response to elevated temperature under projected future climate
248 conditions.

249 The soil moisture was reduced by warming in the alpine grassland (Figure S2).
250 Therefore, we disentangled the influence of soil temperature and soil moisture on *Re*,
251 CH₄ uptake, and N₂O emission by variation-partitioning analysis under the four
252 treatments in the growing season and the non-growing season (Figure 4). Our results
253 showed that, under the NGW treatment, *Re*, CH₄ uptake, and N₂O emission in the
254 non-growing season were more influenced by soil temperature than by soil moisture.
255 Under the GW treatment, there was the single effect of soil temperature on CH₄
256 uptake and N₂O emission in the non-growing season. In contrast, there were the joint
257 effects of soil temperature and moisture on *Re* in the non-growing season under the
258 GW treatment. *Re* in the growing season was influenced more by soil moisture than
259 soil temperature under the GW treatment. Annual *Re* under the AW treatment was
260 influenced by the joint effects of soil temperature and moisture.

261 **5. Conclusions**

262 In summary, the effect of seasonally asymmetrical warming on *Re* and N₂O
263 emission was obvious, unlike the situation with CH₄ uptake. The *Re* and N₂O
264 emission were able to adapt to continuous warming, resulting in a reduced response
265 rates of the *Re* and N₂O emission to temperature increase. Warming in the
266 non-growing season increased the temperature dependence of the *Re*. Thus, we

267 believe that the study of climate change should pay greater attention to warming in the
268 non-growing season, to avoid underestimating the greenhouse effect on *Re* in alpine
269 grasslands.

270 **Data availability**

271 The measured CO₂, CH₄ and N₂O fluxes and soil temperature and soil water
272 content data are available in Zenodo (<http://doi.org/10.5281/zenodo.4244207>).

273 **Author contributions**

274 GYM, LYY and MA conceive the research question, designed the study approach,
275 led the field survey, ensured data curation and conducted formal analysis. YP and
276 LKH assisted with data collection and analysis. GYM wrote the first draft of the paper,
277 and all co-authors contributed to writing review and editing.

278 **Competing interests**

279 The authors declare that they have no conflicts of interest.

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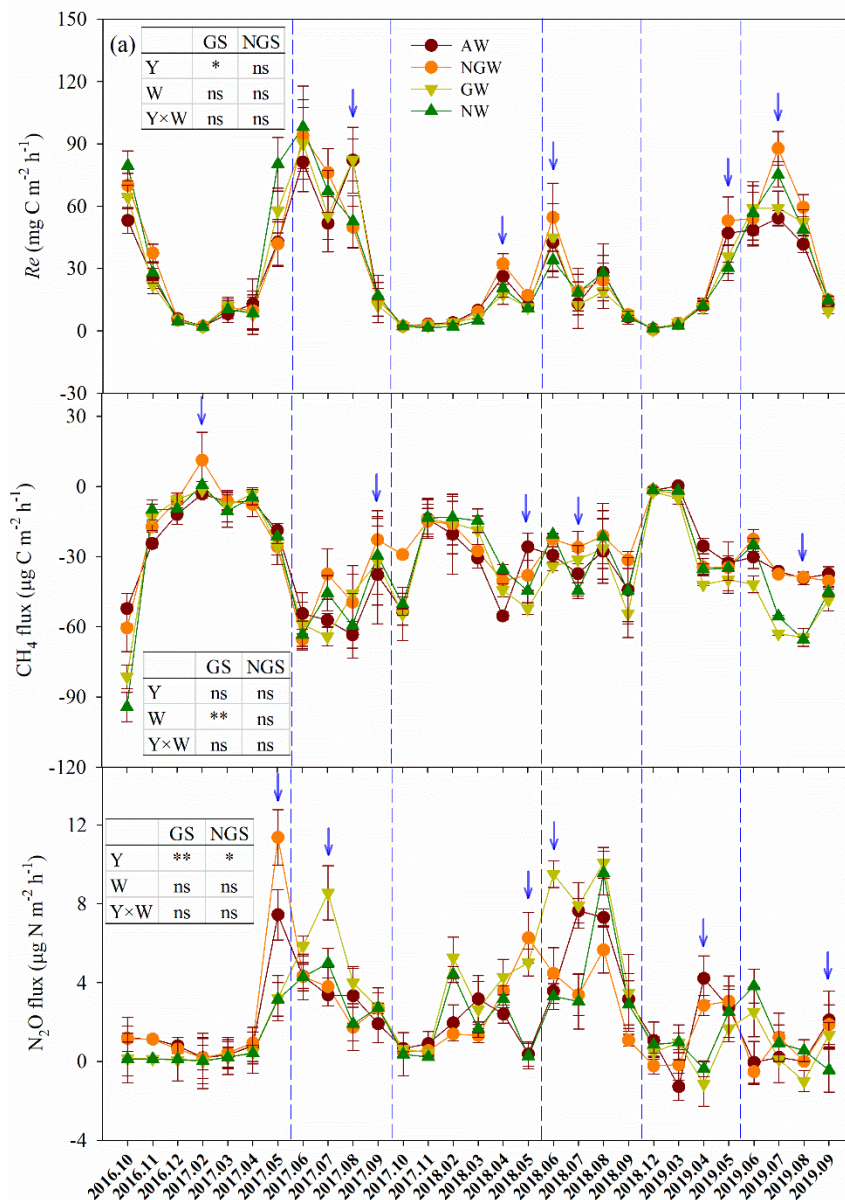
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430

431 Figure 1 Monthly variation of a). ecosystem respiration (Re), b). CH_4 uptake and c).

432 N_2O emissions under the four treatments from October 2016 to September 2019. AW,

433 warming throughout the year; NGW, warming in the non-growing season only; GW,

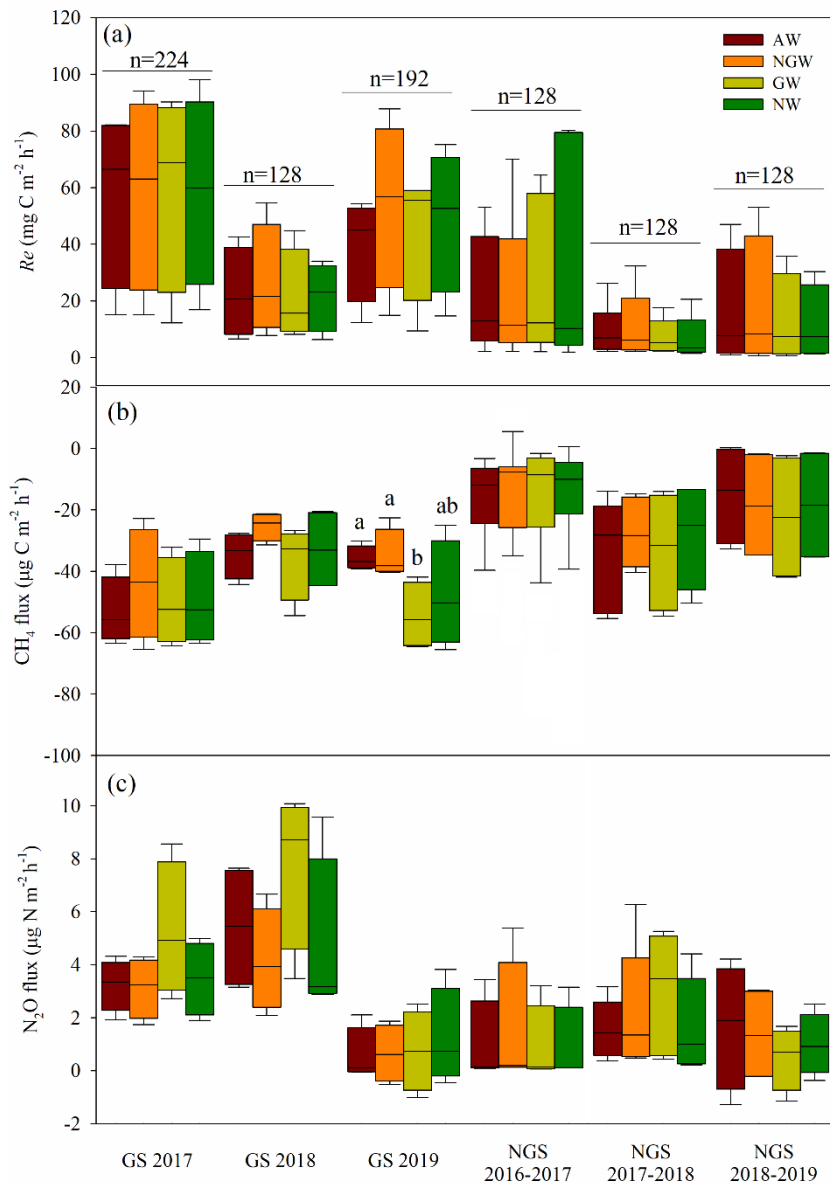
434 warming in the growing season only; NW, non-warming. The blue arrows indicate

435 warming effects. The data points represent mean \pm standard error, SE. The tables

436 illustrate the tests of significance for year (Y) and warming (W) on Re , CH_4 uptake

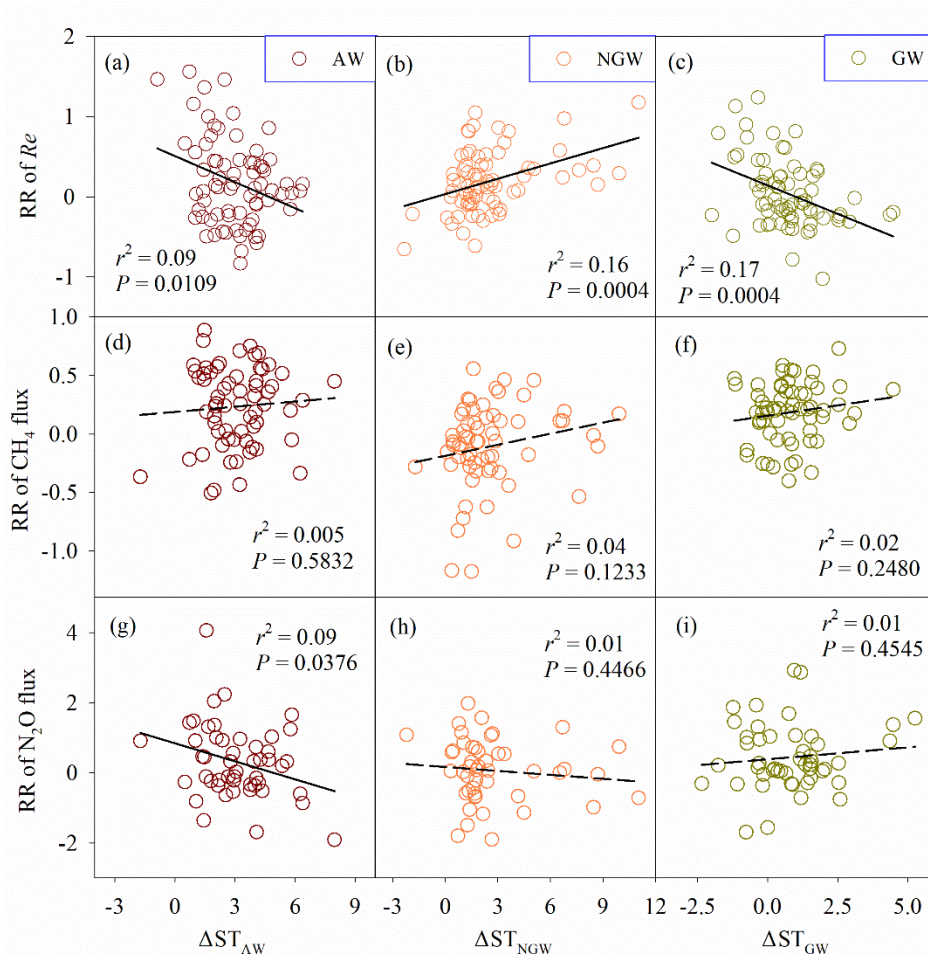
437 and N_2O emission by two-way repeated-measures analysis of variance (ANOVA) in

438 the growing season (GS) and the non-growing season (NGS), respectively; * $P < 0.05$;
 439 ** $P < 0.01$; ns, non-significant.



440
 441 Figure 2 Boxplot presentation of variations in ecosystem respiration (Re), CH_4
 442 uptake, and N_2O emission under four treatments in the growing season and
 443 non-growing season from October 2016 to September 2019. The median is
 444 represented by the line in the box. The box (the interquartile range) represents the
 445 middle 50% of the data, whereas the whiskers represent the ranges for the bottom 25%
 446 and the top 25% of the data values, excluding outliers. GS, growing season; NGS,

447 non-growing season; AW, warming throughout the year; NGW, warming in the
 448 non-growing season only; GW, warming in the growing season only; NW,
 449 non-warming. No significant differences among AW, NGW, GW, and NW were
 450 reported from ANOVA; data points are the mean \pm standard error. One-way ANOVA
 451 results of Re , CH_4 uptake and N_2O emissions among the four warming treatments
 452 were not significant, except that the CH_4 uptake in the GS 2019 under the GW
 453 treatment was significantly higher than that of AW and NGW treatment ($P < 0.05$).



454

455 Figure 3 Response (presented by linear regression) of variation in ecosystem
 456 respiration (Re), CH_4 uptake, and N_2O emission to changes in soil temperature under
 457 AW, NGW and GW conditions in the alpine grassland, from 2016 to 2019. RR, the

458 natural logarithm of the response ratio of the mean value of the chosen variable in the
 459 warming group to that in the control (NW) group. ΔST_{AW} , soil temperature of AW
 460 minus that of NW; ΔST_{CW} , soil temperature of NGW minus that of NW; ΔST_{WW} , soil
 461 temperature of GW minus that of NW; AW, warming throughout the year; NGW,
 462 warming in the non-growing season only; GW, warming in the growing season only;
 463 NW, non-warming.

	<i>Re</i>	CH ₄ flux	N ₂ O flux
NGW-NGS %	a 41.6 c 0.8 b -1.6	75.0 -4.1 0.8	43.8 -1.4 -1.9
NGW-GS %	6.4 6.3 9.0	-2.9 0.2 -2.7	1.3 4.0 -0.3
GW-NGS %	0.7 36.5 22.2	51.3 7.4 0.9	29.6 10.2 -2.0
GW-GS %	22.6 -12.4 23.4	-2.6 0.4 -2.4	3.8 0.9 <0.1
AW-AY %	9.5 22.3 10.1	15.3 6.2 -0.9	7.7 4.5 -1.9
NW-AY %	7.6 26.7 5.0	18.5 4.7 -0.9	21.5 -3.7 3.5

464

465 Figure 4 Influence of soil temperature and soil moisture on ecosystem respiration
 466 (*Re*), CH₄ uptake, and N₂O emission by variation-partitioning analysis under four
 467 treatments in the growing season and non-growing season. a, Single effect of soil
 468 temperature (%); b, single effect of soil moisture (%); c, joint effects of soil
 469 temperature and moisture (%); NGW-NGS, greenhouse gas fluxes in non-growing
 470 season under non-growing season warming treatment; NGW-GS, greenhouse gas
 471 fluxes in growing season under non-growing season warming treatment; GW-NGS,

472 greenhouse gas fluxes in non-growing season under growing season warming
473 treatment; GW-GS, greenhouse gas fluxes in growing season under growing season
474 warming treatment; AW-AY, annual greenhouse gas fluxes under annual warming
475 treatment; NW-AY, annual greenhouse gas fluxes without warming.