1	Different responses of ecosystem CO_2 and N_2O emissions and CH4 uptake to
2	seasonally asymmetric warming in an alpine grassland of the Tianshan
3	Mountains
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10	Abstract:
11	An experiment was conducted to investigate the effect of seasonally asymmetric
12	warming on ecosystem respiration (Re), CH4 uptake, and N2O emissions in alpine
13	grassland of the Tianshan Mountains of Central Asia, from October 2016 to
14	September 2019. The annual mean of Re, CH4, and N2O fluxes in growing season
15	were 42.83 mg C m^-2 h^-1, -41.57 μg C m^-2 h^-1, and 4.98 μg N m^-2 h^-1, respectively.
16	Furthermore, warming during the non-growing season increased Re and CH4 uptake
17	by 7.9% and 10.6% in growing season, 10.5% and 9.2% in non-growing seasons,
18	respectively. However, the increase in N_2O emission in the growing season was
19	mainly caused by the warming during the growing season (by 29.7%). the warming
20	throughout the year and warming during the non-growing season increased N_2O
21	emissions by 101.9% and 192.3% in non-growing seasons, respectively. The Re, CH4
22	uptake, and N ₂ O emissions were positively correlated with soil temperature. Our

results suggested that Re, CH₄ uptake and N₂O emissions were regulated by soil 23 temperature, rather than soil moisture, in the case of seasonally asymmetric warming. 24 25 In addition, the response rate was defined by the changes in greenhouse gas fluxes driven by warming. In our field experiment, we observed the stimulatory effect of 26 27 warming during the non-growing season on Re and CH₄ uptake. In contrast, the response rates of Re and N₂O emissions were gradually attenuated by long-term 28 annual warming and the response rate of *Re* was also weakened by warming over the 29 growing season. These findings highlight the importance of warming in the 30 31 non-growing season in regulating greenhouse gas fluxes, a finding which is crucial for improving our understanding of C and N cycles under the scenarios of global 32 warming. 33

34 Keywords: Alpine steppe; Extreme climatic event; Greenhouse gas fluxes;
35 Warming of open-top chambers

36 **1. Introduction**

37 Since the industrial revolution, human activities have intensified global warming. 38 The global surface temperature increased by about 0.85°C from 1880 to 2012 (IPCC, 2013). Furthermore, the temperature is expected that the surface temperature will 39 increase by about 1.1–6.4 °C by the end of this century (IPCC, 2007, 2013). The rise 40 41 in atmospheric temperature over the year is not continuous on the temporal scale but there is asymmetrical warming across the seasons (Xia et al., 2014). The 3rd and 4rd 42 43 Assessment Report of the Inter-Governmental Panel on Climate Change (IPCC) proposed that, against the backdrop of global warming, the temperature change shows 44

45 that the warming amplitude in the winter is greater than that in the summer, with the 46 warming amplitude at high latitude being greater than that at low latitude, and 47 confirmed that the warming shows asymmetric trends on a seasonal scale (Easterling 48 et al., 1997; IPCC, 2001, 2007).

Carbon dioxide (CO_2) , methane (CH_4) , and nitrous oxide (N_2O) are three of the 49 major greenhouse gases (GHGs) in the atmosphere that directly cause global climate 50 warming, with their contributions to global warming being 60%, 20%, and 6%, 51 52 respectively (IPCC, 2007, 2013). Experimental warming is known to influence 53 ecosystem respiration (*Re*), CH₄ uptake, and N₂O emission (Pärn et al., 2018; Treat et al., 2018; Wang et al., 2019). Information on Re, CH₄ uptake, and N₂O emission will 54 enhance our understanding of ecosystem C and N cycling processes and improve our 55 56 predictions of the response of ecosystems to global climate change (Li et al., 2020; Wang et al., 2019). 57

At present, most studies focus on the influence of warming on GHG flux in 58 59 terrestrial ecosystems during the summer months (Keenan et al., 2014; Li et al., 2011; 60 Yang et al., 2014). Nevertheless, data on the influence of asymmetric warming on the GHG flux on a seasonal scale are scarce. A study of the Alaskan tundra found that 61 summer warming (using open-top chambers to increase air temperatures in the 62 growing season) significantly increased Re in the growing season by about 20% 63 (Natali et al., 2011). Compared with the slight effect of winter warming on the 64 ecosystem respiration in the growing season, warming increased ecosystem 65 respiration during the snow-covered non-growing season by more than 50% (Natali et 66

al., 2011). Lin et al. (2015) reported that the response of soil CH₄ uptake rates to 67 temperature increases in alpine meadows of the Qinghai-Tibet Plateau were not 68 consistent seasonally, with CH₄ uptake in the non-growing season being more 69 sensitive to temperature (increasing by 162%) than the corresponding value in the 70 71 growing season. A study by Cantarel et al. (2012) in an alpine grassland ecosystem showed that the response of N₂O emission to warming showed clear seasonal 72 differences, with the N₂O emission in the growing season showing significant 73 differences between the warming treatments, whereas the response of N₂O emission 74 75 to the warming treatments in November was not obvious. A recent study showed that seasonal variations in carbon flux were more closely related to air temperature in the 76 meadow steppe (Zhao et al., 2019). Another study found that experimental warming 77 78 enhanced CH₄ uptake in the relatively arid alpine steppe, but had no significant effects on CH₄ emission in the moist swamp meadow (Li et al., 2020). Furthermore, soil CH₄ 79 uptake was not significantly affected by warming in the alpine meadow of the Tibetan 80 81 Plateau (Wu et al., 2020). In contrast, a global meta-analysis showed that experimental warming stimulates ecosystem respiration in grassland ecosystems, and 82 the response of ecosystem respiration to warming strongly varies across the different 83 grassland types, with greater warming responses in cold than in temperate and semi-84 arid grasslands (Wang et al., 2019). Across the data set, Li et al. (2020) demonstrated 85 that N₂O emissions were significantly enhanced by whole-year warming treatments. 86 In contrast, no significant effects on soil N2O emissions were observed by in 87 short-season warming. 88

In summary, the GHG fluxes in terrestrial ecosystems shows significant 89 interannual, and seasonal variations, and its response to warming also varies over 90 91 different temporal scales. After long-term uniform warming, the biotic and abiotic factors have adapted to the temperature increase, and the GHG fluxes response to 92 93 increasing temperature is smaller than that in the early stages of warming. For example, over longer time periods of warming, accelerated carbon decomposition and 94 increased plant N uptake may decrease soil organic C and N pools (Wu et al., 2012), 95 and the microbial community with variable C use efficiency may reduce the 96 97 temperature sensitivity of heterotrophic respiration (Zhou et al., 2012). Moreover, climate warming is often unstable, with most of it occurring as extreme events 98 (Jentsch et al., 2007). The heterogeneity of warming may change the adaptability of 99 100 GHG fluxes to warming, and thus affect the carbon and nitrogen cycles in terrestrial ecosystems. In this study, we hypothesize the stimulatory effect of warming during the 101 non-growing season on Re, CH₄ uptake and N₂O emissions. However, the response 102 103 rates of Re, CH₄ uptake and N₂O emissions were gradually attenuated by long-term 104 annual warming and warming over the growing season, respectively.

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2. Materials and methods

The experiment was conducted from October 2016 to September 2019 at the
Bayinbuluk Grassland Ecosystem Research Station, Chinese Academy of Sciences
(42°52.76′ ~ 42°53.17′ N, 83°41.90′ ~ 83°43.12′ E, 2460 m above sea level), which is
located in the southern Tianshan mountains of Central Asia, Xinjiang Uyghur
Autonomous Region of China. Permafrost is present in the Bayinbuluk alpine

2018) being more than 250 cm. The mean annual temperature was -4.8 °C per decade, 112 with the lowest monthly temperature in January (-27.4 °C) and the highest in July 113 (11.2 °C), and the mean annual precipitation amounted to 265.7 mm, with 78.1% 114 occurring during the growing season, from June to September (Geng et al., 2019). 115 Variations in soil temperature, soil moisture, air temperature and precipitation are 116 shown in Figure S1, S2, S3 and S4, respectively. The site was fenced since 2005, all 117 the plots were dominated by Stipa purpurea, Festuca ovina, Oxytropis glabra, and 118 Potentilla multifida. The soil was sub-alpine steppe soil, the parent material of the soil 119 was Loess, and the average annual soil moisture was 5.9% (2017-2019). 120

grassland, with the average maximum frozen depth (from 2000 to 2011, Zhang et al.,

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The open-top chambers (OTCs) were made of 5 mm thick tempered glass. To 121 122 reduce the impact of precipitation and snow, the OTC was constructed with a hexagonal round table which was 100 cm high, and the diagonals of the bottom and 123 top were 100 cm and 60 cm, respectively. Four treatments were simulated using OTCs: 124 125 warming throughout the year (AW), warming in the non-growing season (October 1 to 126 May 31 of the next year) only (NGW), warming in the growing season (June 1 to September 30) only (GW) and no warming (NW). After the warming in the NGW or 127 GW, the tempered glass was removed and the frame was retained. Three replicate 128 plots were established for each treatment, each plot measuring $1 \text{ m} \times 1 \text{ m}$, with a 3-m 129 wide buffer zone between adjacent plots, making a total of 12 plots. Soil temperature 130 and soil moisture were measured at a frequency of every half an hour by an outdoor 131 temperature and humidity data recorder (at 10 cm depth; HOBO U23-001; Onset 132

Computer Corporation, Bourne, USA). The air temperature inside the OTCs is also recorded at a frequency of every half an hour using HOBO Pro RH/TEMP Data LOGGERS (hanged in the center of the OTCs, 50cm above the surface; Onset Computer Corporation, Bourne, USA). Soil temperature and air temperature were increased about 2.3 °C and 4 °C by the warming treatment, respectively (Figure S1 and S3). Soil moisture was reduced about 5% by the warming treatment (Figure S2).

Re, CH₄ and N₂O fluxes were measured using static chambers, made of PVC 139 tubing with diameter 0.25 m and height 0.17 m, with one chamber in each of the 12 140 plots. Gas samples were taken 0, 10, 20 and 30 minutes after the lid of the static 141 chamber was sealed in between 12:00 and 14:00 (GMT + 8), collecting once or twice 142 a week. The rates of ecosystem respiration, CH₄ and N₂O fluxes were calculated 143 144 based on the change in concentration of CO₂, N₂O and CH₄ in each chamber over time by a linear or non-linear equation (P < 0.05, $r^2 > 0.95$) (the positive flux values 145 represent emission, and the negative flux values represent uptake; Liu et al. 2012; 146 147 Wang et al. 2013). Concentrations of individual gases in samples were measured using a gas chromatograph (GC) (Agilent 7890A; Agilent Technologies, Santa Clara, CA, 148 USA). 149

Effects of seasonally asymmetric warming on Re, CH₄ uptake, and N₂O emissions were analyzed by two-way repeated-measures analysis of variance (ANOVA). One-way ANOVA was used to compare soil temperature, soil moisture and air temperature differences, respectively. Nonlinear regression analyses (Exponential Growth, Single, 3 Parameter) was used to identify the relationship

between ecosystem respiration (Re) and soil temperature (at 10-cm depth) from 155 October 2016 to September 2019. General linear analyses were used to identify 156 significant linear correlations and regressions between soil temperature, or soil 157 moisture variation and the responses of CH₄ uptake, or N₂O emissions. 158 Variation-partitioning analysis was used to disentangled the influence of soil 159 temperature and soil moisture on Re, CH₄ uptake and N₂O emission under the four 160 treatments in the growing season and the non-growing season, respectively. The 161 natural logarithm of the response ratio (RR) was used to reflect the effects of 162 seasonally asymmetric warming on alpine grassland GHG fluxes (Hedges et al., 1999). 163 The RR is the ratio of the mean value of the chosen variable in the warming group 164 (\overline{W}_t) to that in the control group (NW; \overline{W}_c), and is an index of the effect of 165 166 seasonally asymmetric warming on the corresponding variable (Eq. 1). All statistical analyses were conducted using SPSS (version 20.0) (IBM, Armonk, NY, USA) with 167 the statistically significant difference threshold set at P < 0.05. 168

$$RR = \ln\left(\frac{\overline{W}_{t}}{\overline{W}_{c}}\right) = \ln\left(\overline{W}_{t}\right) - \ln\left(\overline{W}_{c}\right)$$
(1)

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170 **3. Results**

Our study showed that the Bayinbuluk alpine grassland exhibited a low Re, was a net CH₄ sink, and a negligible N₂O source. The annual mean values of Re, CH₄ uptake, and N₂O emissions in the growing season were 42.83 mg C m⁻² h⁻¹, 41.57 µg C m⁻² h⁻¹, and 4.98 µg N m⁻² h⁻¹, respectively, from October 2016 to September 2019. One-way ANOVA results of Re, CH₄ uptake and N₂O emissions among the four warming treatments were not significant, with the exception that the soil CH₄ uptake in the growing season 2019 under GW treatment was significantly higher than that of the AW and NGW treatments (P < 0.05). Compared with the control group (NW), the *Re* was decreased by 7.5% and 4.0% in the growing season and non-growing season, respectively, under AW and decreased by 2.4% and 8.5% under GW in the growing season and non-growing season, respectively. However, compared with the control group, the *Re* under NGW increased by 7.9% and 10.5% in the growing season and non-growing season, respectively, averaged over the three years (Figure 2 a).

The AW temperature change induced a 6.4% increase in CH₄ uptake in the 184 growing season and a 3.8% decrease in the non-growing season. The GW treatment 185 resulted in 7.1% and 10.2% increases in CH₄ uptake in the growing season and 186 non-growing season, respectively. On the contrary, the NGW generated a 10.6% and 187 188 9.2% decrease in CH₄ uptake in the growing season and non-growing season, respectively (Figure 2 b). The AW and NGW treatments resulted in 5.8% and 2.2% 189 decreases, respectively, in N₂O emission in the growing season, and 101.9% and 190 192.3% increases, respectively, in N₂O emission in the non-growing season. 191 Compared with the control, NW group, the N₂O emission increased by 29.7% and 192 decreased by 24.4% under GW in the growing season and non-growing season, 193 respectively (Figure 2 c). 194

195 The results of two-way repeated measures ANOVA showed significant 196 interannual differences of Re in the growing season (P < 0.05, Figure 1 a), whereas 197 the CH₄ uptake under the warming treatment exhibited significant differences in the 198 growing season (P < 0.01; Figure 1 b), and the interannual N₂O emission showed

199	significant differences in both the growing season and non-growing season ($P < 0.05$,
200	Figure 1 c). Therefore, interannual variation was larger than the impact of the
201	warming treatment (for Re and N2O emissions, Figure 1), whereas the warming
202	treatment had a significant impact on CH4 uptake. Under the four warming treatments,
203	Re exhibited exponential growth, respectively ($P < 0.05$; Figure S5 a). we observed
204	increasing CH ₄ uptake with increasing soil temperature ($P < 0.05$; Figure S5 b). On
205	the other hand, the N_2O emission showed a significantly positive linear correlation
206	with soil temperature, but only under NGW ($P < 0.05$; Figure S5 c).
207	The soil moisture was reduced by warming in the alpine grassland (Figure S2).
208	However, Re, CH ₄ uptake and N ₂ O emission were not significant linearly correlated
209	with soil moisture, respectively ($P \ge 0.05$; Figure S6). We disentangled the influence
210	of soil temperature and soil moisture on Re, CH4 uptake, and N2O emission by
211	variation-partitioning analysis under the four treatments in the growing season and the
212	non-growing season (Figure 4). Under the NGW treatment, Re, CH ₄ uptake, and N ₂ O
213	emission in the non-growing season were more influenced by soil temperature than by
214	soil moisture. Under the GW treatment, there was the single effect of soil temperature
215	on CH4 uptake and N2O emission in the non-growing season. In contrast, there were
216	the joint effects of soil temperature and moisture on Re in the non-growing season
217	under the GW treatment. Re in the growing season was influenced more by soil
218	moisture than soil temperature under the GW treatment. Annual Re under the AW
219	treatment was influenced by the joint effects of soil temperature and moisture.

4. Discussion

221	Our study found that the response rate of Re to temperature significantly
222	decreased with the increase in soil temperature (Δ ST_{AW} and $~\Delta$ ST_{GW}) under AW and
223	GW treatments, respectively (Figure 3 a, c; $P < 0.05$). This finding indicated that the
224	response of Re to soil temperature became less and less sensitive to soil temperature
225	with warming throughout the year (or the growing season) in the alpine grasslands.
226	On the contrary, NGW significantly increased the response rate of Re to temperature
227	change (Δ ST_{NGW}), indicating that warming in the non-growing season amplified the
228	sensitivity of Re to temperature change (Figure 3 b, $P < 0.05$). In addition, Zou et al.
229	(2018) showed that the soil fluxes of CO_2 increased exponentially with increasing
230	temperature, but warming decreased the temperature sensitivity by 23% in the
231	grassland. Furthermore, Natali et al. (2011) also confirmed that, compared with the
232	CO_2 flux in the growing season, the CO_2 flux in the nongrowing season was more
233	sensitive to the temperature increase.

234 Ecosystem CH₄ flux is the net result of CH₄ production and consumption, occurring simultaneously under the action of methanogenic archaea and 235 236 methane-oxidizing bacteria (e.g., Mer and Roger, 2001). In this study, warming increased CH₄ uptake in the growing season, but decreased CH₄ uptake in the 237 non-growing season in the alpine grassland, findings similar to those from other 238 grassland ecosystems (Lin et al., 2015; Wu et al., 2020; Zhu et al., 2015). Our results 239 also demonstrated that seasonally asymmetric warming did not significantly affect the 240 response rate of CH₄ uptake (Figure 3 d-f, P > 0.05). CH₄ flux depended on 241 temperature, pH, and the availability of substrate (e.g., Treat et al., 2015). The CH₄ 242 uptake observed during the three growing season and non-growing season implied 243 244 that the alpine grassland soil could act as an atmospheric CH₄ sink, a finding which

agrees with the results of many previous studies in similar regions (Wei et al., 2015; 245 Zhao et al., 2017). Hu et al. (2016) suggested that asymmetrical responses of CH₄ 246 fluxes to warming and cooling should be taken into account when evaluating the 247 effects of climate change on CH₄ uptake in the alpine meadow on the Tibetan plateau. 248 Unlike CH₄ flux in alpine grasslands, Treat et al. (2018) confirmed that wetland was a 249 250 small CH₄ source in the non-growing season, whereas uplands varied from CH₄ sinks to CH₄ sources. The latest research confirmed that warming in the Arctic had become 251 more apparent in the non-growing season than in the typical growing season (Bao et 252 253 al., 2020). Hereby, Bao et al. (2020) found that the CH₄ emissions during the spring thaw and the autumn freeze contributed approximately one-quarter of the annual total 254 255 CH₄ emissions. That experimental warming is stimulating soil CH₄ uptake in the growing season implies that the grasslands of the Bayinbuluk may have the potential 256 to remove more CH₄ from the atmosphere under future global warming conditions. 257

258 Furthermore, with the increased variation in soil temperature, the response rate of N₂O emission gradually decreased under AW treatment (Figure 3 g, P < 0.05). The 259 response of N₂O emission to temperature increase was limited by the warming that 260 261 occurred throughout the year. However, N₂O emission peaks were displayed during the freeze-thaw periods (e.g., May 2017, June 2018 and April 2019). Warming 262 increased N₂O emissions in the thawing period due to disruption of the gas diffusion 263 barrier and greater C and N availability for microbial activity (Nyborg et al., 1997). 264 Wagner-Riddle et al. (2017) also demonstrated that the magnitude of the 265 freeze/thaw-induced N₂O emissions was associated with the number of days with soil 266 267 temperatures below 0°C. Pärn et al. (2018) found that N₂O emission from organic soils increases with rising soil NO₃⁻, follows a bell-shaped distribution with soil 268

269 moisture. Another study has shown that a whole - year warming treatment 270 significantly increased N_2O emissions, but daytime, night-time or short - season 271 warming did not have significant effects (Li et al., 2020). In addition, Cantarel et al. 272 (2010) suggested that the N_2O flux from cool and upland grasslands may be driven 273 primarily by response to elevated temperature under projected future climate 274 conditions.

275 **5.** Conclusions

In summary, the effect of seasonally asymmetrical warming on Re and N₂O 276 emission was obvious, unlike the situation with CH₄ uptake. The Re and N₂O 277 emission were able to adapt to continuous warming, resulting in a reduced response 278 rates of the Re and N₂O emission to temperature increase. Warming in the 279 280 non-growing season increased the temperature dependence of the Re. Thus, we believe that the study of climate change should pay greater attention to warming in the 281 non-growing season, to avoid underestimating the greenhouse effect on *Re* in alpine 282 283 grasslands.

284 Data availability

The measured CO_2 , CH_4 and N_2O fluxes and soil temperature and soil water content data are available in Zenodo (http://doi.org/10.5281/zenodo.4244207).

287 Autho

Author contributions

GYM, LYY and MA conceive the research question, designed the study approach,
led the field survey, ensured data curation and conducted formal analysis. YP and
LKH assisted with data collection and analysis. GYM wrote the first draft of the paper,

and all co-authors contributed to writing review and editing.

292 **Competing interests**

293 The authors declare that they have no conflicts of interest.

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Figure 1 Monthly variation of a). ecosystem respiration (Re), b). CH₄ uptake and c). N₂O emissions under the four treatments from October 2016 to September 2019. AW, warming throughout the year; NGW, warming in the non-growing season only; GW, warming in the growing season only; NW, non-warming. The blue arrows indicate warming effects. The data points represent mean \pm standard error, SE. The tables illustrate the tests of significance for year (Y) and warming (W) on Re, CH₄ uptake and N₂O emission by two-way repeated-measures analysis of variance (ANOVA) in



the growing season (GS) and the non-growing season (NGS), respectively; *P < 0.05;

453 **P < 0.01; ns, non-significant.



Figure 2 Boxplot presentation of variations in ecosystem respiration (Re), CH₄ uptake, and N₂O emission under four treatments in the growing season and non-growing season from October 2016 to September 2019. The median is represented by the black line in the box. The box (the interquartile range) represents

the middle 50% of the data, whereas the whiskers represent the ranges for the bottom 459 25% and the top 25% of the data values, excluding outliers. GS, growing season; NGS, 460 461 non-growing season; AW, warming throughout the year; NGW, warming in the non-growing season only; GW, warming in the growing season only; NW, 462 non-warming. No significant differences among AW, NGW, GW, and NW were 463 reported from ANOVA; data points are the mean \pm standard error. One-way ANOVA 464 results of Re, CH₄ uptake and N₂O emissions among the four warming treatments 465 were not significant, except that the CH₄ uptake in the GS 2019 under the GW 466 treatment was significantly higher than that of AW and NGW treatment (P < 0.05). 467





Figure 3 Response (presented by linear correlation) of variation in ecosystem

470 respiration (*Re*), CH₄ uptake, and N₂O emission to changes in soil temperature under

471	AW, NGW and GW conditions in the alpine grassland, from 2016 to 2019. RR, the
472	natural logarithm of the response ratio of the mean value of the chosen variable in the
473	warming group to that in the control (NW) group. $\triangle ST_{AW}$, soil temperature of AW
474	minus that of NW; \triangle ST _{CW} , soil temperature of NGW minus that of NW; \triangle ST _{WW} , soil
475	temperature of GW minus that of NW; AW, warming throughout the year; NGW,
476	warming in the non-growing season only; GW, warming in the growing season only;
477	NW, non-warming.

	Re	CH ₄ flux	N ₂ O flux		
NGW-NGS %	a c b 41.6 0.8 -1.6	75.0 -4.1 0.8	43.8 -1.4 -1.9		
NGW-GS %	6.4 6.3 9.0	-2.9 0.2 -2.7	1.3 4.0 -0.3		
GW-NGS %	0.7 36.5 22.2	51.3 7.4 0.9	29.6 10.2 -2.0		
GW-GS %	22.6 -12.4 23.4	-2.6 0.4 -2.4	3.8 0.9 <0.1		
AW-AY %	9.5 22.3 10.1	15.3 6.2 -0.9	7.7 4.5 -1.9		
NW-AY %	7.6 26.7 5.0	18.5 4.7 -0.9	21.5 -3.7 3.5		

478

Figure 4 Influence of soil temperature and soil moisture on ecosystem respiration 479 (Re), CH₄ uptake, and N₂O emission by variation-partitioning analysis under four 480 treatments in the growing season and non-growing season. a, Single effect of soil 481 temperature (%); b, single effect of soil moisture (%); c, joint effects of soil 482 temperature and moisture (%); NGW-NGS, greenhouse gas fluxes in non-growing 483 season under non-growing season warming treatment; NGW-GS, greenhouse gas 484

485	fluxes in growing season under non-growing season warming treatment; GW-NGS,
486	greenhouse gas fluxes in non-growing season under growing season warming
487	treatment; GW-GS, greenhouse gas fluxes in growing season under growing season
488	warming treatment; AW-AY, annual greenhouse gas fluxes under annual warming
489	treatment; NW-AY, annual greenhouse gas fluxes without warming.