# Assessing MODIS Vegetation Continuous Fields tree cover product (collection 6): performance and applicability in tropical forests and savannas.

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Abstract. The Moderate Resolution Imaging Spectroradiometer vegetation continuous fields (MODIS VCF) Earth observation product is widely used to estimate forest cover changes, parameterise vegetation and Earth System models, and as a reference for validation or calibration where field data areis limited. However, while although limited independent validations of MODIS VCF have shown that MODIS VCF's accuracy decreases when estimating tree cover in sparsely-vegetated areas such as tropical savannas, no study has yet assessed the impact this may have on the VCF-based tree cover datadistributions used by many in their research. Using tropical forest and savanna inventory data collected by the TROpical Biomes in Transition (TROBIT) project, we produce a series of corrections that take into account (i) the spatial disparity between the in-situ plot size and the MODIS VCF pixel, and (ii) the trees' spatial distribution within in-situ plots. We then applied our corrections to areas identified as forest or savanna in the International Geosphere-Biosphere Programme (IGBP) land cover mapping product. All IGBP classes identified as 'savanna' show substantial increases in cover after correction, indicating that the most recent version of MODIS VCF consistently underestimates woody cover in tropical savannas. We estimate that MODIS VCF could be underestimating tropical tree cover by as much as 29between 9 to 15, %. Models that use MODIS VCF as their benchmark could therefore be underestimating the carbon uptake in forest-savanna areas and misrepresenting forestsavanna dynamics. Because of the limited While more detailed in-situ plot number, our results are designed to be used as an indicator of where the product is potentially more or less reliable. Until more in-situfield data are available is necessary, to produce more accurate and reliable corrections, we recommend caution when using <u>uncalibrated</u> MODIS VCF in tropical savannas.

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### 1 Introduction

Tree cover values derived from Earth observation (EO)E.O.) data form a fundamental part of ecological research. They are used to estimate forest cover change, biomass, and carbon stocks (Bastin et al., 2019; Giriraj et al., 2017; Saatchi et al., 2011; Song et al., 2014); help identify key areas for conservation efforts (Miles et al., 2006); and are used as a basis for climatic and vegetation modelling and model evaluation (Brovkin et al., 2013; Burton et al., 2019; Kelley et al., 2013). All this research, in turn, plays a vital role in informing local, regional, and global environmental policies (Harris et al., 2012). As such, an EOE.O. product's accuracy is important to consider, as any errors in the initial tree cover estimate can be further compounded in downstream work.

Only a handful of EOE.O, products provide global maps of percentage tree cover or forest and shrub cover distributions (Bartholomé and Belward, 2005; Bicheron et al., 2008), and fewer still provide information stretching over at least a decade (Friedl et al., 2002; Hansen et al., 2003). Of these, one of the products most widely used in ecological modelling is the Moderate Resolution Imaging Spectroradiometer Vegetation Continuous Fields (MODIS VCF) product (DiMiceli, 2017). MODIS VCF is a yearly product that provides percent tree cover globally at a spatial resolution of 250 m and is available for the years 2000 through to 2020. Its quantitative measure of woody cover is recorded annually and is described as a percentage of ground cover, making it particularly suited for use in evaluating dynamic global models (Lasslop et al., 2018; Rabin et al., 2017), as a proxy for in-situ data that areis harder to collect (Kelley et al., 2019), and to help define parameters for calculating global tree restoration potential (Bastin et al., 2019). Collection 6 is the most recent iteration of the product.

As the VCF product has progressed from Collection 1 to its current Collection 6, several validations using in-situ field data or higher-resolution remotely sensed data as a reference measurement have been carried out. These have been few and limited to sites within a biome (Montesano et al., 2009a), a region (Hansen et al., 2005; White et al., 2005), or within a country (Gao et al., 2014; Sexton et al., 2013). The MODIS VCF product evaluated was the most recent collection available at the time (i.e...)

Hansen et al., 2005 and White et al., 2005 for Collection 3; Montesano et al., 2009a for Collection 4; and Gao et al., 2015 and Sexton et al., 2013 for Collection 5). To our knowledge, no such independent validation experiment has yet been conducted on Collection 6, which produces tree cover estimates in the same manner as Collection 5 but with improvements made to the upstream inputs to enhance its accuracy (DiMiceli, 2017).

The validations found that MODIS VCF may be less suitable for estimating tree cover in sparsely\_vegetated areas. Huang & Siegert (2006) noted that MODIS VCF classified large areas of land as 'bare' where their land cover classification system identified it as sparsely-vegetated. Montesano et al. (2009) found that MODIS VCF data (Collection 4) overestimated cover in areas of low tree cover in taiga-tundra transition zones. Sexton et al. (2013) found that the Collection 5 product overestimated cover in areas of low cover (below 20 %) and underestimated in areas of higher tree cover, while Gao et al. (2015) found that MODIS VCF can only partially discriminate between tropical forest and nonforest, struggling in areas that have greater heterogeneity. What is clear from the history of these validation and comparison experiments is that MODIS VCF has accuracy issues in areas with low woody vegetation cover, which has implications when its tree cover estimates are treated as accurately representative of real-world conditions. Failure to accurately account for the product's difficulty in estimating low woody covers can, therefore, lead to miscalibrated models and estimations that do not reflect real-world conditions. This, in turn, has knock-on effects on environmental policy-making, conservation efforts, and future ecological research, especially in areas with vegetation cover types that are most prone to error.

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Tropical savannas have woody covers that fall within the range particularly affected by the reported MODIS VCF errors. A large proportion of these savannas can be found in tropical developing countries (Boval and Dixon, 2012), and are predicted to be home to half of the world's population by 2050 (State of the Tropics, 2020). Tropical savannas are therefore highly vulnerable to anthropogenic change. In the face of a growing population, land fragmentation, and changing climate, a savanna's ability to maintain robust ecosystem functions is directly linked to the amount of woody cover present (Sankaran et al., 2006). As a result, the ability to accurately monitor the state, dynamics, and woody cover trends of tropical savannas is a vital part of understanding how and why savannas are changing in the tropics (Harris et al., 2012; Miles et al., 2006), while also improving modelled climate projections and vegetation dynamics for this complex biome.

In this study, we validate the accuracy of MODIS VCF Collection 6 in tropical savannas and forests by comparing the tree cover percentage of the product to corresponding field data. We then characterise the observed bias in woody covers across both savanna and forest ecosystems and apply our corrections across the tropics to highlight the regions most likely to be affected by these inaccuracies in the MODIS VCF product.

### 275 **2 Methods**

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We used the MODIS VCF Collection 6 product (spatial resolution of 250 m, DiMiceli, 2017) with tree cover values averaged across the years 2006 through to 2009 to reflect the range of the field data collection period. The in-situ field data were was sourced from the 'TROpical Biomes In

Transition'InTransition', project (TROBIT)

(www.geog.leeds.ac.uk/TROBIT.www.geog.leeds.ac.uk/TROBIT., Torello-Raventos et al., 2013) and accessed via the Forestplots.net database (Lopez-Gonzalez et al., 2011; Lopez-Gonzalez et al., 2009). The data we used.which include the corner locations and the Canopy Area Index (CAI) values for 1724 forest and 3124 savanna sites distributed across Australia, Brazil, Bolivia, Cameroon, and Ghana (Fig. A1 and Table A1, Fig. 2 in Torello-Raventos et al., 2013). The TROBIT field campaigns were carried out over a 3-year period, from 2006 to 2009, and the field plotssites used in this study are 1 hectare 100 m x 100 m in size (except for BFI-01 (03 sites from Ghana that are 50 m x 100 m). All the sites fall within the tropics, that is, within 23,5 ha), BFI-02 (0.5 ha), BFI-03 (0.5 ha), CTC-01 (0.93 ha), degrees north and VCR-01 (0.6 ha). south of the equator.

All the sites fall within the tropics, that is, within 23.5 degrees north and south of the equator, and were selected in regions where savannas and forests were in close proximity and exist within ecotones or 'zones of tension.' As such, the sites sampled show a large variation in physiognomy and edaphic and climatic conditions (Table S1, Veenendaal et al., 2015).

The classification of the TROBIT plotssites as either 'forest' or 'savanna' is based on the parameters described in Torello-Raventos et al. (2013) and Veenendaal et al. (2015). A 'savanna' is a natural land cover that is not a forest, bare ground, or a body of water. 'Forest' is, with 'forest' being defined as woody vegetation with an average tree a mean height of or exceeding 6 m for trees with a diameter at breast height (dbh) exceeding 10 cm, and a canopy area index (CAI) valuetree cover of at least 0.3 for 'open forests' and 0.7 for 'forests.' In addition, floristic differences (i.e., dominance of 'savanna' species) are used to differentiate forests from taller-growing savannas that have similar CAIs and tree heights (see Fig. 9, Torello-Raventos et al., 2013).

or exceeding ~60 %...There is some ambiguity in how 'savannas' and 'grasslands' are defined. Some, with some modelling-based research treattreating the two biomes as different (Whitley et al., 2017). while studies based on plant functional traits group them together (Solofondranohatra et al., 2018; White et al., 2000). As there is some concern that MODIS VCF will struggle to pick up woody cover in

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CAI is defined as the sum of the projected areas of individual tree <u>crownseanopies</u> divided by the ground area. In the TROBIT project <u>(Torello-Raventos et al. (2013) and Veenendal et al. (2015))</u>, plotwide CAI is made up of the sum of, the upper-stratum, mid-stratum, and subordinate-stratum crown areas, are factored into the CAI values and were recorded using methods outlined in Torello Raventos et al. (2013), Membership to a stratumthe strata is determined by the tree's dbh (upper-stratum: dbh > 10 cm, mid-stratum: 2.5 cm < dbh < 10 cm, and subordinate-stratum: dbh < 2.5 cm, height > 1.5 m). About 50 trees per stratum per plot were measured to derive plot-specific allometric relations between stem diameter and crown area (supplement B of Torello Raventos et al. (2013)). These were then applied to the whole plot to establish plot-level CAI. For the allometric relationships, tree crowns were treated as circles, and the individual tree projected crown area was determined using the average of crown radii measured along the four cardinal points (i.e., from the centre of the stem to the distance furthest from the stem).

between individual tree canopies. We account for this by converting each CAI value into a probability distribution function incorporating the following two extreme scenarios: "enforced overlap." where the location probability of individual canopies increases linearly from 0 to 1 across a site; and "unenforced overlap." where individual canopies follow a uniform random distribution pattern and canopy overlap is not purposefully introduced (Fig. A2). We repeated this 1000 times per CAI measurement to determine the probability distribution of expected CAI for each field plotsite.

Unlike CAI, which is the fraction of ground covered by tree crowns, the percent tree cover value from MODIS VCF is defined as "the portion of the skylight orthogonal to the surface which is intercepted by trees" (Hansen et al. 2002). To make the MODIS VCF tree cover value comparable to the tree cover derived from the CAI of a TROBIT plot CAIssite, we divided the MODIS VCF values by 0.8 as suggested by Hansen et al. (2002). This is the standard approach in most modelling studies that use MODIS VCF (e.g., Lasslop et al., 2020; Kelley et al., 2013; Burton et al., 2019). The 0.8 value can be thought of as a gap correction factor (GCF) that accounts for within-canopy gaps. Although the GCF has been shown to vary with vegetation type (Lloyd et al., 2008; 0.34 - 0.60) and crown cover (Tang et al., 2019: 0.96 - 0.7), we opted to use 0.8 as we found that it yielded more conservative results compared to a variable GCF. It also avoided introducing additional parameters into our analysis.

Next, to account for the difference in size between the MODIS VCF pixel (250 m x 250 m) and the smaller field plot size (100 m x 100 m), we calculated the possible percent tree cover an area the size of a TROBIT field plotsite could have, given the MODIS VCF percent tree cover for a MODIS-sized pixel. This was done for two extreme scenarios: "enforced clumping," where all the tree cover for the given MODIS VCF value is forcibly 'clumped' on one side of the pixel, or "unenforced clumping," where 'clumping' is not enforced, and tree cover is distributed randomly within the pixel (Fig. A3). The Basically, the clumping scenarios introduce possible variations in percent cover due to the area and location mismatch between a TROBIT field plotsite and a MODIS pixel. A probability distribution was generated for each MODIS VCF pixel by calculating percent tree cover values for 1000 samples (100 m x 100 m) randomly placed within the 250 m x 250 m MODIS VCF pixel.

We thereby compared MODIS VCF and TROBIT under four different scenarios: 1) unenforced overlap and clumping; 2) enforce overlap and unenforced clumping; 3) unenforced overlap and enforced clumping; 4) enforced overlap and clumping. Comparisons were conducted by fitting the following logit function:

 $logit(VCF) = C_{0A} + \Delta \times log(C^{\tau_1}/(1 - C^{\tau_2}))$  (Equation 1)

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Where  $\mathcal{L}_{0}$ ,  $\mathcal{L}_{n}\tau_{1n}\tau_{2n}$  are optimised parameters and VCF and C are the MODIS VCF pixel and TROBIT site probability distributions, respectively. This is similar to a standard linear regression of logit transformed data, accounting for maximum and minimum bounds of 0 - 100 % tree cover, with  $\tau_{10}\tau_{20}$ allowing for a non-symmetric transformation of tree cover. To account for the probability density of each point, we inferred the parameters in Equation 1 using a Total Least Squares Bayesian Inference technique using a Metropolis-Hastings Markov Chain Monte Carlo step. Priors were uninformed but physically bounded (i.e.,  $\tau_{14}\tau_{24} > 0$ ) to assume an increasing relationship between MODIS VCF and C. Our logit transformed data in Equation 1, along with  $\tau_{\perp}$ ,  $\tau_{\perp}$  correcting for symmetry, allowed us to assume normally distributed model errors, thus describingand so our conditional probability of observations for a given parameter combination was described by a normal distribution (Gelman et al., 2013). Each combination was run over 10 chains, with 1000 warm-up iterations and 10,000 sampling iterations. Optimisation was performed using the rstan2.19.2 (Stan Development Team, 2019) package in R3.5.2 (R Core Team, 2018). Our optimization accounts for potential errors in TROBIT cover, which includes those caused by the allometric construction of the CAI, provided that the errors are unbiased and remain roughly consistent across sites (Gelman et al., 2013). As the TROBIT plots have relatively small total errors associated with the allometric relationships (Table B1, Torello-Raventos et al., 2013), systematic errors are unlikely to affect our results.

We evaluated the impact of the MODIS VCF biases inferred from this correction across the tropics by inverting our calculation of MODIS VCF bias (Fig. A4) as follows: first, the inverse (i.e., solving for C) of Equation 1 was applied to MODIS VCF values after conversion to a 100 m x 100 m pixel size grid (matching the field site area); then this corrected value was translated back to the original 250 m x 250 m VCF pixel size. As the inverse of Equation 1 has no analytical solution, we found the rounded percent value of C that minimises the absolute difference between the left- and right-hand side of the equation. For computational feasibility, we constructed maps of the tropics with corrected MODIS VCF values (Fig. 2) by sampling 5 iterations that were randomly sampled from each of our 10 optimisation chains (50 in total) and masking out pixels with cover types not considered as 'forest' or 'savanna',

We used the 500 m MODIS Land Cover Type (MCD12Q1 - collection 6) product to identify the areas of 'forest' and 'savanna' across the tropics in the MODIS VCF product. MCD12Q1 is widely used by the global land surface modelling community (e.g., Sellar et al., 2019; Wiltshire et al., 2020) and describes land cover in terms of 17 global land cover classes as per the International Geosphere-Biosphere Programme (IGBP, Table 3 in Sulla-Menashe and Friedl, 2018). The product is based on the same spectroradiometer (MODIS) and temporal resolution as the VCF product. Referring to the definition of 'savanna' of Veenendaal et al. (2015), the following land cover classes were chosen to represent 'savanna': Closed Shrubland, Open Shrubland, Woody Savanna, Savanna, and Grassland, while 'forest' encompasses: Evergreen Needleleaf Forests, Evergreen Broadleaf Forests, Deciduous Needleleaf Forests, Deciduous Broadleaf Forests, and Mixed Forests. We subset MCD12Q1 to the tropical zone between +/- 30° North and took the median class for the 2006 to 2009 period, matching the field data collection period.

For a more detailed land-cover-specific evaluation, we resampled the corrected 250 m MODIS VCF pixels to a 500 m grid and combined it with the MCD12Q1 product to construct land-cover-specific MODIS VCF tree cover frequency distributions (Fig. A5). Our tree cover correction by cover type (Fig. 3) for the four clumping/overlap regression combinations was then calculated by multiplying each cover type MODIS VCF frequency distribution (Fig. A5) with curves representing the median, 5 %, and 95 % confidence lines of the correction equation ensembles.

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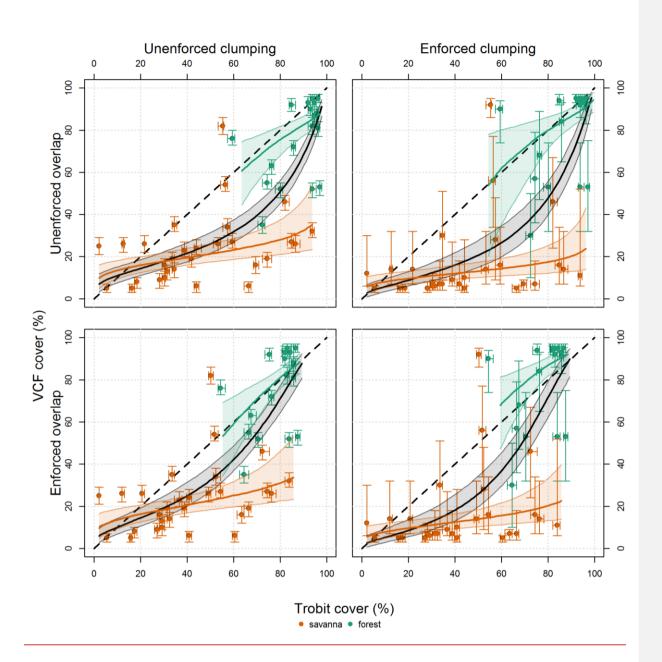
3.1 Comparing MODIS VCF to tree cover from TROBIT field sites

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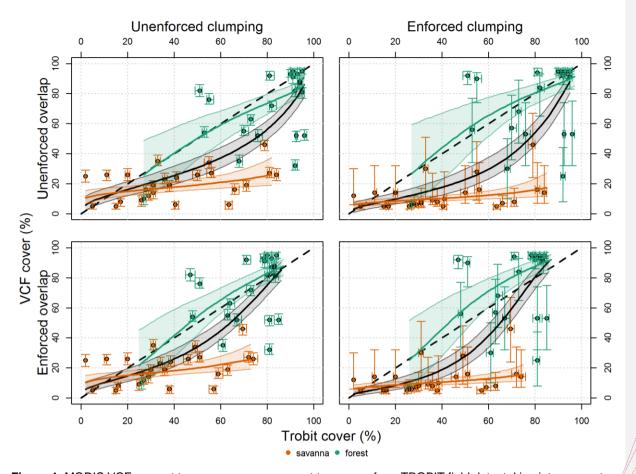


Figure 1, MODIS VCF percent tree cover versus percent tree cover from TROBIT field data, taking into account uncertainties associated with tree cover spatial distributions within a MODIS pixel and field plotsite. The 4 combinations are: (1) no overlap and no clumping, where tree canopies are randomly distributed within both pixel and site; (2) no overlap and maximum clumping, where tree canopies are clustered in one area of the pixel, and randomly distributed throughout the field site; (3) with overlap and no clumping, where tree canopies are randomly distributed within the pixel, but overlap substantially within the field site; and (4) with overlap and maximum clumping, where tree canopies are clustered to one side within a pixel, and overlap substantially within the site. The bolded dashed line in black shows the 1:1 relationship. The solid lines represent the median of the respective regressions (green for forest; orange for savanna; black for forest and savanna combined). The), and the thin lines represent the 5 and 95 % confidence interval of their respective regression lines. The vertical error bars represent uncertainty introduced by clumping; the horizontal error bars represent the uncertainty introduced by overlap.

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MODIS VCF underestimates tree cover within the 1923, % to 8172, % range across all four combinations of enforced-unenforced overlap and clumping (black line, Fig. 1). Below 12 %, MODIS VCF tree cover values do not significantly disagree with TROBIT field data, and may instead be overestimating tree cover (50 % confidence, dashed line, Fig. 1). A similar pattern is seen when tree cover exceeds 84-80, %: MODIS VCF does not differ significantly from TROBIT when there is no enforced overlap (i.e., when tree canopies are spaced randomly within the site - Fig. A2 left), but may overestimate tree cover when overlap is enforced (i.e., trees are clustering towards one side increasing the degree of canopy overlap - Fig. A2 right).

There is a clear difference in how accurately MODIS VCF estimates tree cover in forested areas (in green, Figure 1) as opposed to areas identified as savannas (in orange, Fig. 1). In savanna sites, MODIS VCF significantly and consistently underestimates tree cover regardless of the amount of overlap and clumping. Significant underestimation (at 95 % confidence) occurs when *in-situ* tree cover exceeds 18 – 19 % (without enforced clumping) or 9 – 10 % (with enforced clumping). In forest sites, MODIS VCF does not show the same pattern of systematic underestimation generally estimates tree cover well. Divergence does occur at high covers, depending on the enforcement of overlap or clumping. MODIS VCF overestimates underestimates, tree cover where tree cover exceeds 7879 % (at the 95 % confidence interval) when neither overlap nor clumping is enforced, and overestimates

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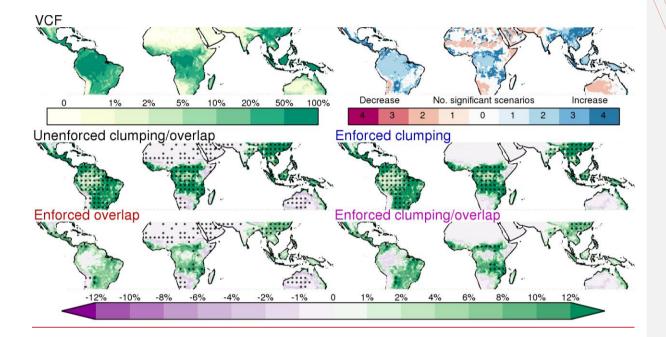
where tree cover exceeds 9058 % (at 5 % confidence interval) when both overlap and clumping are enforced.

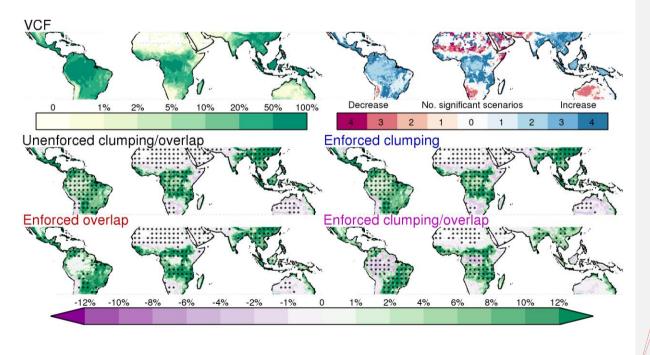
### 3.2 Global estimates of change in tropical tree cover

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We assessed the impact of MODIS VCF's VCFs underestimation of intermediate tree covers across the tropics restricted to the (IGBP forest and savanna land cover classes we identified as being either 'forest' or 'savanna,'), using a 'correction'eorrection based on the combined forest and savanna sites (black curve, Fig. 1). We did not use1) instead of using the savanna-only sites for a savanna-specific correction (orange curve, Fig. 1)). This is because there were few TROBIT sites representing savanna with MODIS VCF tree cover values exceeding 40 %, and global land cover maps disagree on the distribution of savannas within the forest-savanna ecotone (Herold et al., 2008).





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The distribution of tree cover change after calibrating against field data are post correction is similar across the four scenarios (Fig. 2), and the regions where all four scenarios agree on the direction of change (positive and negative) are substantial. However, there are some differences caused by the uncertainty introduced by different extents of overlap and clumping. While we see a significant increase in tree cover across all clumping-overlap combinations in many regions of tropical savannas and grasslands (Pennington et al., 2018), such as in the forest-savanna mosaics that surround Congolian rainforests, we do not see the same pattern in the Cerrado of Brazil. This is likely because the African forest-savanna regions fall within the range of MODIS VCF values that consistently undergo a positive correction (~30 - 50 %, see Fig. A4), while the Cerrado of Brazil does not.

We also see a significant tree cover decrease in the Sahel post-correction in most or all of the scenarios, which runs counter to the results of Brandt et al. (2020) that found that tree cover was underestimated in the region. This disparity may be explained by our lack of field sites in more arid regions. As these corrections were based on a limited number of sites in a limited number of regions, it is important to note that the maps shown in Figure 2 are not definitive. Instead, it should be used to identify areas where MODIS VCF estimates may be more or less reliable.

3.3 Change in tree cover within different vegetation classes in tropical ecosystems

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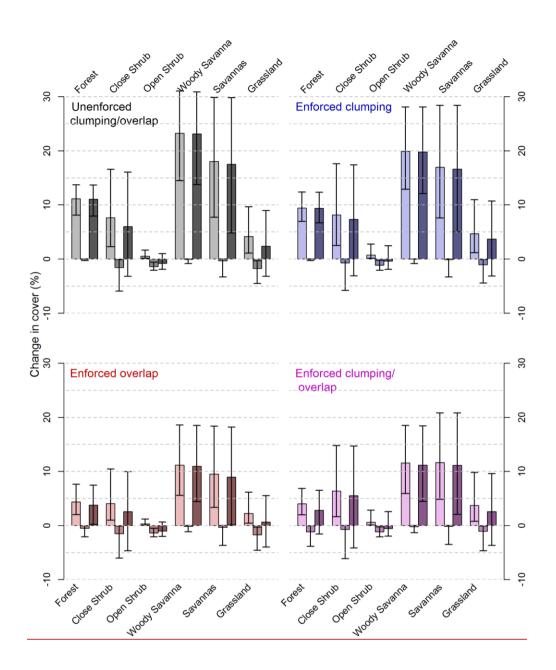
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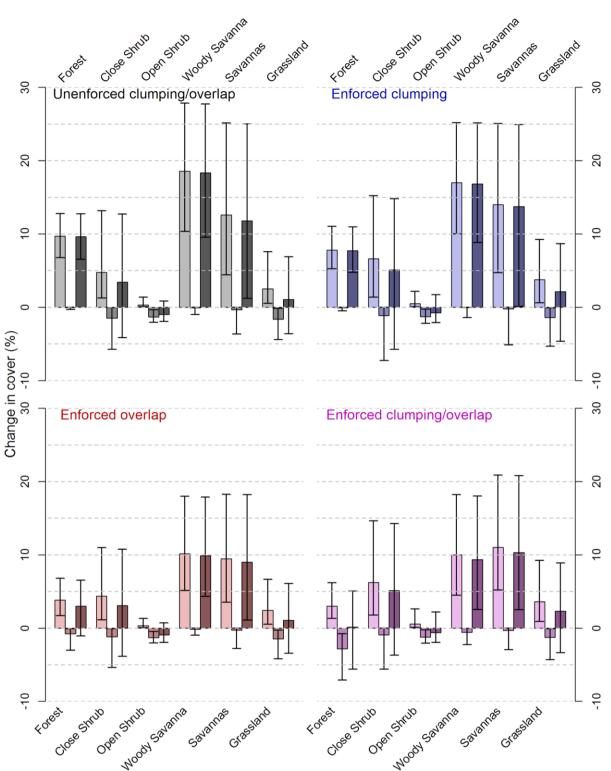


Figure 3. Percent change in tree cover after the application of the appropriate relevant correction (clockwise: no enforced clumping or overlap (black); enforced clumping and no enforced overlap (blue); no enforced clumping and enforced overlap (red); enforced clumping and overlap (pink) in the 'forest' supercategory and the 5 savanna classes. Palest tone indicates positive change, mid-tone indicates negative change, and the darkest tone indicates net change. Error bars denote the 5-95 % confidence interval; if the error bar extends past the x-axis, the post-correction change is not considered significant.

When looking at our correction in more detail, we see that MODIS VCF significantly underestimates tree cover in all the IGBP land cover classes that we considered, regardless of overlap or clumping (95 % confidence interval) (%Fig. 3). The most substantial and significant underestimation is in the classes 'woody savannas' and 'savannas.'savannas'. The underestimation is the largest in woody savannas, except when clumping and overlap are enforced (in purple, Fig. 3). This is because the

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13 <del>13</del> peak in the tree cover <u>frequency</u> distribution for savannas aligns with where the correction for maximum overlap and clumping is the largest (i.e., at about 20 % tree cover, see Fig. A5), while the peak in cover distribution for woody savannas (<u>26 - 6730-70</u> %, Fig. A5) aligns with the cover range that undergoes the greatest correction (Fig. 4, Fig. A5) in the other clumping and overlap scenarios.

'Open shrublands' only showshows a small underestimation of tree cover, despite its woodyhaving a tree cover definition (10 - 60 %) matchingrange that closely resembles the range where MODIS VCF most underestimates tree cover (26 - 67i.e. 20-60 % cover). The discrepancy may be because the majority of the 'open shrublands' IGBP class commission error is definitions may not entirely correspond with the 'grasslands' class (see Table S6 in Sulla-Menashe et al., 2019). The tree cover percentage identified by MODIS VCF, or there is some issue with how apt the IGBP description of shrublands is in terms of tree cover. For example, IGBP distinguishes 'shrublands' from 'savannas' based on a woody perennial height threshold of < 2 m height, while MODIS VCF tree cover only observes canopies of trees exceeding 5 m in areas classified as 'open shrublands' height. MODIS VCF tree cover in open shrubland is therefore likely to be lower than the IGBP definition would suggest tree cover (see Fig. A5), resulting in corrections that are more conservative.

We found significant increases in tree cover for 'forests' in every correction scenario, though net change is only significant (95 % confidence) when overlap is unenforced. This can be explained by the presence of both negative and positive corrections in the higher ranges of tree cover when overlap is enforced. Similarly, the net change is insignificant across all clumping and overlap scenarios for the IGBP classes matching the lower ranges of tree cover (grassland, close shrubland and open shrubland).

### 4 Discussion

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While MODIS VCF is a powerful and accessible tool to map tree cover, our field data-based 765 corrections indicate that the latest MODIS VCF collection 6 is missing a lot of woody cover even when uncertainty introduced by site canopy overlap and clumping within the MODIS VCF pixel are accounted for. Our maps (Fig. 2) highlight that this potential underestimation of woody cover is mainly occurring in tropical savannas. Moreover, the highest underestimation in the savanna classes occurs when there is no enforced overlap (i.e., when there is a uniform random distribution of trees) which is the most likely scenario for the TROBIT savanna plots as evidenced by work done by Veenendaal et 770 al. (2015), where) to test the plots were tested for complete spatial randomness and only. Only, minor indications of overlap were found. Woody savannas, as an example, may have their tree cover underestimated by up to 3227, % (95 % confidence) when neither clumping nor overlap is enforced (in black, Fig. 3). If our results are representative of the tropics, then overall, MODIS VCF may be 775 underestimating tropical tree cover by between 7 - 295-22, % for unenforced clumping and overlap or 0 - 21-22, % for when either clumping or overlap are enforced (5 - -95 % confidence).

An overestimation at the lower end of the cover (< 20 %) (Hansen et al., 2002; Sexton et al., 2013) and underestimation in the lower to middle range of cover (20 % - 60 %), have been identified in validations of previous MODIS VCF collections (Gross et al., 2018; Yang and Crews, 2019). According to its definition, MODIS VCF only maps trees that are 5 m or taller (Hansen et al. 2003), while the TROBIT CAI includes all trees with a minimum dbh of 2.5 cm, as well as trees with a height exceeding 1.5 m when dbh < 2.5 cm. This could explain the observed underestimation in the lower tree cover ranges. However, because of how our field reference CAI is derived, we were not able to conclusively link the 5 m threshold to our observed underestimation. This could explain the observed underestimation in the lower tree cover ranges for the savanna sites. However, the discrepancies we found between the percent tree cover ranges of the IGBP class descriptions (i.e. MCD12Q1 product) and the corresponding MODIS VCF suggests that the 5 m height threshold may not always apply. Unlike in MODIS VCF, an IGBP 'tree' has a minimum height of 2 m, and 'shrubs' are shorter. Therefore, MODIS VCF recording tree cover in the open and closed shrubland classes of MCD12Q1 (Fig. A5) indicates one of three things: these shrublands contain trees tall enough to be measured by MODIS VCF (> 5 m), that MODIS VCF is able to spot shorter trees (< 5 m), or some combination of both scenarios. More importantly, for the MCD12Q1 savanna class, MODIS VCF yields a percent tree cover range that matches closely with the savanna class definition (between 10 % and 30 %), even though the MODIS VCF and IGBP tree thresholds are different. This would suggest that in areas classed as 'savanna', trees grow taller than 5 m, or, similar to the shrubland classes, MODIS VCF is picking up trees with heights below 5 m. Because of how our field reference CAI is derived, we were not able to apply a 5 m threshold to the reference data. As a result, we cannot conclusively link the 5 m

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threshold to our observed underestimation. We expect that the TROBIT CAI percent tree cover value will include tree cover that is not identifiable by MODIS VCF, and that our observed underestimation is not as extensive as shown. More work needs to be done to evaluate how effective MODIS VCF is at distinguishing trees above the 5 m threshold.

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On the other hand, when looking at the relationship between TROBIT's upper stratum canopy height and the difference between TROBIT and MODIS VCF we would have expected an increasing underestimation in the lower height ranges. Instead, we found a low R² and a mixture of under and overestimations in heights between 0 and 10 m (Fig. A6). This suggests that the inclusion of trees below 5 m height in the TROBIT inventory does not fully explain the observed underestimation. However, as the relationship between upper canopy heights and the subordinate strata composition (and canopy cover thereof) varies widely depending on factors including ecosystem type and altitude (Rutten et al., 2015), more research needs to be done with more in-situ height data.

We also found discrepancies between the tree cover values derived from MODIS VCF and the corresponding class definition of the MCD12Q1 product (Fig. A5), which again suggests that the 5 m height threshold may not always apply in MODIS VCF. For example, MODIS VCF recorded tree cover in the 'open shrublands' and 'closed shrublands' classes of the MCD12Q1 product (Fig. A5), even though the height range for these classes is 1 - 2 m. For the 'savannas' class, MODIS VCF yields a percent tree cover range that matches closely with the 'savannas' class definition (between 10 % and 30 %), despite the differing tree thresholds for MODIS VCF and IGBP (5 m minimum for MODIS VCF, and 2 m minimum for IGBP). These discrepancies suggest one of the following three things: 'open/closed shrublands' and 'savannas' contain trees taller than 5 m; MODIS VCF is distinguishing trees below the 5 m threshold; or some combination of both.

Another explanation for the discrepancy between the IGBP class definitions and those estimated through MODIS VCF could be the between-class omission and commission errors (Fig. 4, and Table S6 in Sulla-Menashe et al., 2019). For example, the accuracy for 'closed shrublands' is particularly low. It is mainly confused with 'open shrublands,' 'woody savannas,' and 'savannas.' The majority of the 'open shrublands' class commission error is with the 'grasslands' class, and there is confusion to a lesser extent between 'open shrubland,' 'woody savannas' and 'savannas.' Also, the 'cropland/natural vegetation mosaics' class is often mapped as 'closed shrubland,' 'woody savannas,' 'savannas' or 'grasslands.'

More work needs to be done to evaluate how effective both MODIS VCF and MCD12Q1 are at implementing the height thresholds in their respective 'tree' definitions, as this may have implications when MODIS VCF and MCD12Q1 are used for global model calibration or validation.

Overall, our Que, results suggest that the biases found in the previous collections may have persisted in collection 6, despite reported improvement in accuracy (DiMiceli et al., 2017). This indicates that the biases introduced by binning the training data (Gerard et al. 2017) and using a CART (Classification and Regression Tree) model (Hanan et al., 2013) are inherent and still present within this version of MODIS VCF. Models calibrated using MODIS VCF (Brandt et al., 2017; Lasslop et al., 2020; Burton et al., 2019; Kelley et al., 2019, 2020) risk inheriting these biases and should therefore be validated using other sources of data. We suggest that while MODIS VCF gives a good overview of tree cover on a global scale, it should be re-calibrated before it is used as a reference or training data. Special care should be taken in savannas, a biome that has long been noted as being challenging for EOE.O. products to characterise, as solitary trees in the landscape tend to be missed by global tree cover products (Jung et al., 2006, Brandt et al., 2020). The poor performance of MODIS VCF in savannas in particular (Gaughan et al., 2013; Gross et al., 2018; Kumar et al., 2019) emphasises the importance of continuous independent validation and re-calibration of the product. The ecosystem functions of savannas can vary drastically with just a slight difference in tree cover (Gaughan et al., 2013) and even slight errors may create issues in how we interpret the state and dynamics of the biome, which in turn affects how the land is managed.

Work on forest restoration potential would also be impacted. Bastin et al. (2019), for example, used MODIS VCF to estimate tree cover in agricultural land. As this tree cover is likely to have been underestimated substantially, the derived available land space for replanting may be less than projected, with the restoration potential overestimated. However, our results also indicate an underestimated tree cover in woodier savannas and forests. Accounting for this, the restoration

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ATO ensure the appropriate use of the product, we suggest that where field data are available, the MODIS VCF product should be calibrated for use in the target region. However, calibrating MODIS VCF on a large scale using field data as a reference do present several challenges. Firstly, different in-situ measurement techniques tend to measure different types of tree cover (e.g., Fiala et al., 2006; Korhonen et al., 2006; Rautiainen et al., 2005) and each will require a conversion to enable direct comparison with MODIS VCF. In our case, to account for gaps between tree crowns, we applied the 0.8 'gap correction factor' to the CAI. However, the GCF and resulting tree cover could vary widely on a plot-by-plot basis (Lloyd et al., 2008). With further in-situ data that describe tropical vegetation type-specific GCF variation, we may be able to incorporate site-specific GCFs into our analysis.

There is also the uncertainty associated with the field data collection. In our case, the site-specific CAI standard errors (supplement B in Torello-Raventos et al., 2013) are small and show no systematic bias and are therefore not expected to significantly change our results. Using field plots over a limited geographic extent creates additional uncertainty that may still be unaccounted for in our analysis when calibration is applied across the highly variable tropical forest-savanna ecotone. The bottom map in Fig. A7 combines our uncertainty maps (Fig. 2) with a map plotting the distance of a point from the sampled TROBIT plots, and highlights Southeast Asia, Central America, and Mexico as areas where additional in-situ observations would greatly help improve confidence. Field data from the north-western region of South America, the southeast of the African continent, and Madagascar would also help.

Finally, factors such as cloud cover, landscape heterogeneity, phenology, vegetation type, and soil type affect the accuracy of remotely-sensed products like MODIS VCF (Hansen et al., 2003; Huete et al., 1997; Smith et al., 2002). Data characterising these at the plot level would help identify potential confounding factors affecting MODIS VCF performance, and so help further constrain uncertainties.

is available, the height threshold of MODIS VCF should be evaluated, and the product itself should be calibrated for use in the target region. Alternatively, comparing MODIS VCF to other land cover maps or higher-resolution remotely sensed data areis recommended (Gross et al., 2018; Lary and Lait, 2006), though without a large-scale effort to re-calibrate MODIS VCF, the question of how appropriate MODIS VCF is for use in both forests and savannas in the tropics will remain. By highlighting the extent to which MODIS VCF struggles to estimate tree cover in tropical forests and savannas, we hope to inform the future use of this product to improve its useability.

# **5 Conclusion**

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We found that MODIS VCF significantly underestimates tree cover in tropical forests and savannas, even when within-field site and field site-pixel variation are accounted for during validation. As MODIS VCF is a product that is commonly used in a wide variety of ecological research including vegetation modelling, estimating restoration potential, and identifying forest-savanna bimodality, we stress that more independent work on validating and re-calibrating is required before its tree cover estimates can be relied upon in the tropics.

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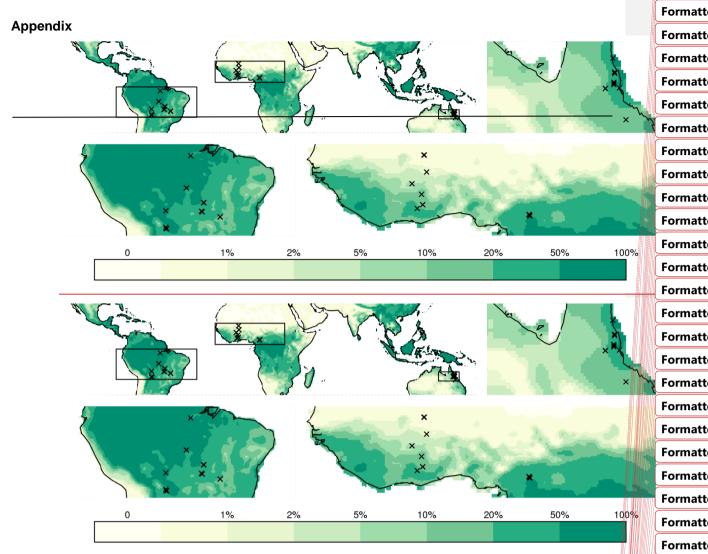
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1005 Figure A1. Location of sampling sites in Africa, Australia, and South America from the TROBIT Project (based on Fig. 2, Torello-Raventos et al., 2013) shown on MODIS VCF (DiMiceli, 2017). sites with available GPS coordinates were selected. (https://www.forestplots.net),

<del>No.</del>	Site Name	Country	Latitu de	Longitu de	MODIS VCF Percent Tree Cover	Cano py Area Index	Average Upper Stratum Height (m)	Cover Type	TROBIT Site ◆ Description ◆
1	ALC-01	Brazil	-2.53	-54.91	12. <u>5</u> 50	0.32	<u>6.56</u>	Savann a	Savanna woodland
2	ALF-01	Brazil	-9. <u>6</u> 60	-55.94	77 <mark>.00</mark>	2.31	37.02	Forest	Tall forest -
3	ALF-02	Brazil	-9.58	-55.92	76 <del>.00</del>	2.65	41.32	Forest	Tall forest ◆
4	ASU-01	Ghana	7.14	-2.45	41.33	2.54	<u>45.27</u>	Forest	<u>Tall forest</u> ◀
5	BBI-01	Burkina Faso	12.73	-1.17	1.33	0.52	12.53	Savann a	Savanna woodland

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6	BBI-02	Burkina Faso	12.73	-1.16	1. <u>5</u> 50	0.99	13.6	Savann a	Savanna woodlane	1
7	BDA-01	Burkina Faso	10.94	-3.15	6.17	0. <u>3</u> 30	14.53	Savann a	Shrub-rich savanna woodland	<u></u>
8	BDA-02	Burkina Faso	10.94	-3.15	4. <u>5</u> 50	0.18	14.47	Savann a	Shrub-rich savanna woodland ◆	
9	BFI-01	Ghana	7.71	-1.69	15 <del>.00</del>	1.22	29.67	Savann a	Tall closed woodland	
<del>10</del>	BFI-02	Ghana	7.71	-1.69	12.83	1.08	28.2	Savann a	Tall savanna woodland	
<del>11</del>	BFI-03	Ghana	7.71	-1. <u>7</u> 70	25.83	2.54	Forest 45. 07.	Savann a	Tall savanna woodland	
<del>12</del>	CTC-01	Australia	- 16. <u>1</u> 10	145.45	72.67	2.35	40.37	Forest	Tall forest	
<del>13</del>	DCR-01	Australia	-17.02	145.58	21.67	1.67	<u>27.19</u>	Savann a	Tall savanna woodland	-
14	DCR-02	Australia	-17.03	145. <u>6</u>	65.67	0.71	Forest 22. 51.	Savann a	Tall savanna woodland	
<del>15</del>	EKP-01	Australia	-18.07	145.99	43. <u>5</u> 50	0.74	28.13For est	Savann a	Tall savanna woodland	
<del>16</del>	FLO-01	Brazil	-12.81	-51.85	65.67	2. <u>4</u> 40	28.21	Forest	<u>Forest</u>	
<del>17</del>	FMS-01	Australia	-18.09	144.84	7.67	0.32	Forest <u>20</u> .03	Savann a	Shrub-rich savanna woodland ◄	
<del>18</del>	FMS-02	Australia	-18.11	144.82	44.17	1.21	16.69	Forest	Stunted shrub-rich fores	t
<del>19</del>	HOM- 01	Mali	15.34	-1.47	0. <u>5</u> 50	0.05	3.87	Savann a	Savanna grassland	
<del>20</del>	HOM- 02	Mali	15.33	-1.55	0.83	0.16	<u>6.13</u>	Savann a	Savanna grassland	
21	JBG-01	Brazil	-15.95	-47.87	20.83	0.22	7.48	Savann a	Scrub savanna ◀	
<del>22</del>	JBG-02	Brazil	-15.95	-47.87	20 <del>.00</del>	0.02	6.29	Savann a	Scrub savanna ◀	
<del>23</del>	JBG-03	Brazil	-15.93	-47.87	20. <u>5</u> 50	0.12	8.01	Savann a	Scrub savanna ◀	
<del>24</del>	JBG-04	Brazil	-15.94	-47.86	27.17	0.77	12.65	Savann a	Savanna woodland	
<del>25</del>	KBL-01	Australia	-17.77	145.54	75 <del>.00</del>	1.69	39.5	Forest	Tall forest -	
<del>26</del>	KBL-02	Australia	-17.85	145.53	61.17	0.81	29.2Fore	Savann a	Tall savanna woodland	
<del>27</del>	KBL-03	Australia	-17.69	145.53	79. <u>5</u> 50	3.00	36.62	Forest	Tall forest ◄	
<del>28</del>	KCR-01	Australia	-17.11	145. <u>6</u>	78.83	2.44	42.37	Forest	Tall forest ◄	Ī
<del>29</del>	LFB-03	Bolivia	- 14. <u>6</u> 60	-60.85	28.17	0.39	9.93	Savann a	Shrub-rich savanna woodland	
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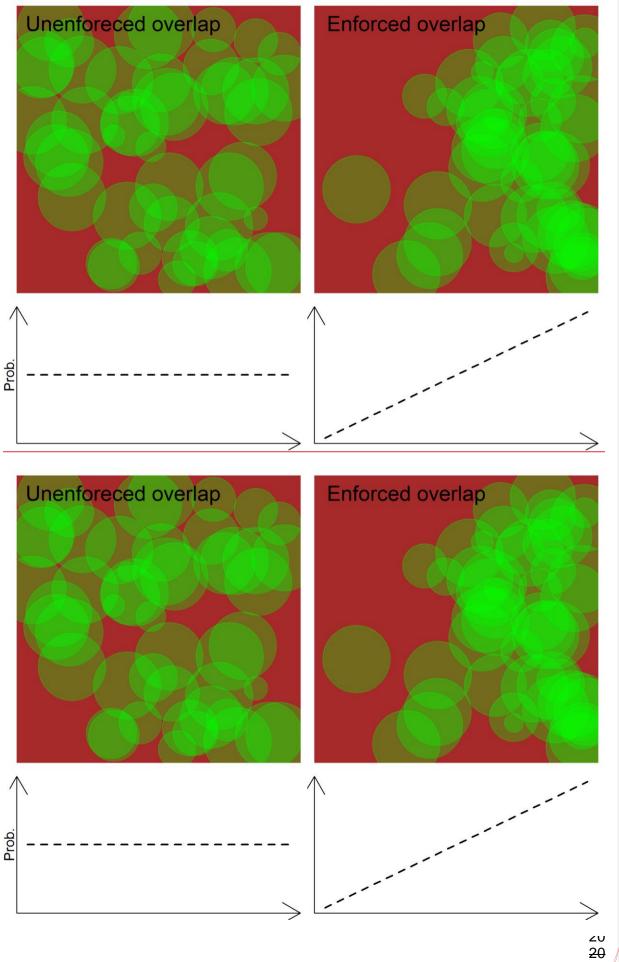
<del>30</del>	MDJ-01	Cameroon	6.17	12.83	42 <del>.00</del>	3.24	45	Forest	Tall forest
<del>31</del>	MDJ-02	Cameroon	6.16	12.82	18.67	0.44	<u>16.13</u>	Savann a	Long-grass savanna
<del>32</del>	MDJ-03	Cameroon	5.98	12.87	64.67	2.97	<u>36.53</u>	Forest	Stunted shrub-rich forest
33	MDJ-04	Cameroon	6 <del>.00</del>	12.87	15 <del>.00</del>	0.37	18.93	Savann a	Long-grass savanna
34	MDJ-05	Cameroon	5.98	12.87	70.33	2.85	21.27	Forest	Stunted shrub-rich forest
<del>35</del>	MDJ-06	Cameroon	6 <del>.00</del>	12.89	20. <u>5</u> 50	0.68	<u>15.27</u>	Savann a	Long-grass savanna
<del>36</del>	MDJ-07	Cameroon	6.01	12.89	57.33	1.75	42.67	Forest	<u>Tall forest</u> ◆
<del>37</del>	MDJ-08	Cameroon	6.21	12.75	15 <del>.00</del>	0.48	<u>18</u>	Savann a	Long-grass savanna
<del>38</del>	MLE-01	Ghana	9. <u>3</u> 30	-1.86	10 <del>.00</del>	0.34	14.67	Savann a	Savanna woodland
<del>39</del>	NXV-02	Brazil	- 14. <u>7</u> 70	-52.35	20.83	1.82	15.76For	Savann a	Tall closed woodland
40	RSC-01	Australia	-20.16	146.54	28 <del>.00</del>	1.15	13.14	Forest	Stunted forest -
41	SMT-01	Brazil	-12.82	-51.77	36.67	1.55	14.37	Savann a	Savanna woodlane
<del>42</del>	SMT-02	Brazil	-12.82	-51.77	41. <u>5</u> 50	1.44	14.64For	Savann a	Savanna woodlan <del>d</del>
43	SMT-03	Brazil	-12.83	-51.77	19.33	0.53	11.19	Savann a	Savanna woodland
44	TUC-01	Bolivia	-18.52	-60.81	50.33	1.29	14.9	Forest	Stunted forest •
4 <del>5</del>	TUC-02	Bolivia	-18.53	-60.63	21.67	0.81	12.05	Savann a	Shrub-rich woodlane
<del>46</del>	TUC-03	Bolivia	-18.19	-60.86	10.83	0.37	14.11	Savann a	Savanna woodland
47	VCR-01	Brazil	-14.83	-52.16	69. <u>5</u> 50	2.81	28.94	Forest	Tall forest
48	VCR-02	Brazil	-14.83	-52.17	69.67	2.74	30.93	Forest	<u>Forest</u> ◆

**Table A1.** Site names, locations, Canopy Area Index values, MODIS VCF percent tree cover values, and cover type, and TROBIT site descriptions of the 48 TROBIT Project plotssites, used in this study. TROBIT site descriptions are based on Table S1Details regarding the vegetation structure for each site can be found in Fig. 1 of Veenendaal Torello-Raventos et al., 20152013.

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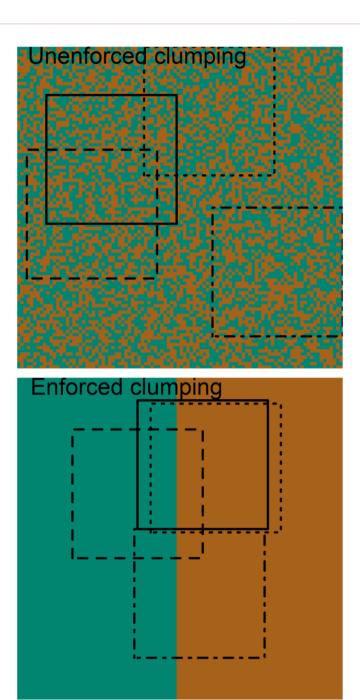
**Figure A2.** Visual representation of the effects of enforcing overlap within a (100 m x 100 m) TROBIT site with a given Canopy Area Index -(CAI). Left: Overlap is not enforced, and individual crowns follow a uniform random distribution. Right: Overlap is enforced by linearly increasing the probability of a canopy being located more on one side of the site (i.e., illustrated, here as the right side of the site) than the other. This results in tree canopies 'overlapping' to a greater extent, which affects how accurately CAI represents actual canopy cover.

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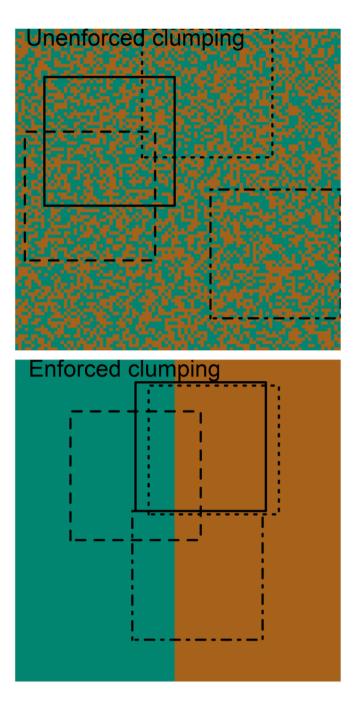
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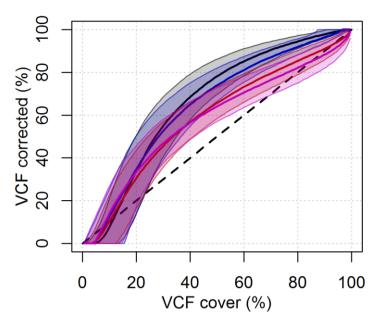
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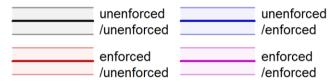
**Figure A3.** Visual representation of the effects of unenforced and enforced clumping in a 250 m x 250 m MODIS VCF pixel with 50 % tree cover. Clumping all the cover to one side of the pixel (bottom) affects the average canopy cover value of a 100 m x 100 m-sized average TROBIT site, illustrated here as the black boxes.

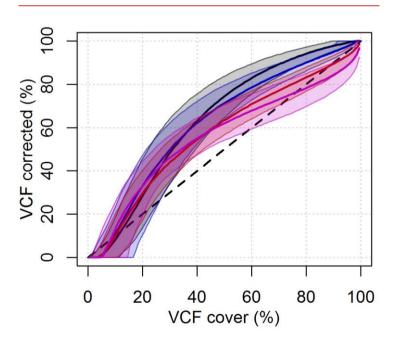
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# Clumping/Canopy overlap





# Clumping/Canopy overlap

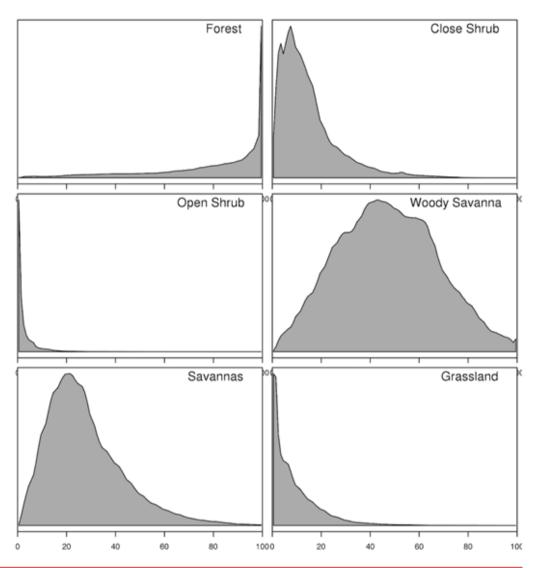


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**Figure A4.** The correction curves developed for MODIS VCF based on the 4 pixel-site mismatch scenarios (no clumping and no overlap; enforced clumping no overlap; no clumping enforced overlap; and enforced clumping and enforced overlap). The dashed line signifies the 'ideal' 1:1 relationship wherein corrected MODIS VCF is unchanged from the original MODIS VCF values. The shaded regions represent 5 to 95 % confidence intervals for the respective corrected MODIS VCF values.



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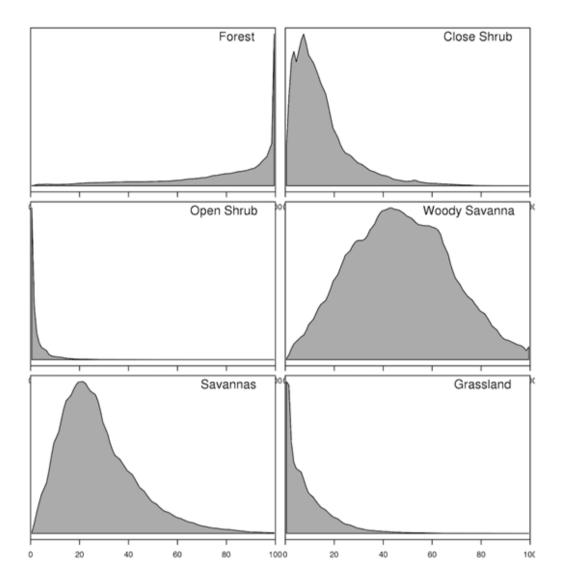
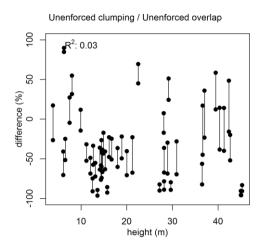


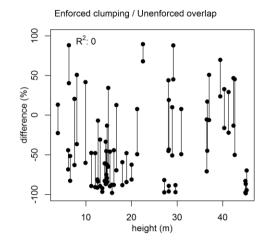
Figure A5. Frequency distributions of percent tree cover value as estimated by MODIS VCF across the 'forest' supercategory and the following IGBP classes that by our definition count as part of the 'savanna' domain: Closed Shrublands, Open Shrublands, Woody Savannas, Savannas, and Grasslands. Specific class definitions as per the User Guide for the MODIS Land Cover Product (Sulla-Menashe and Friedl, 2018).

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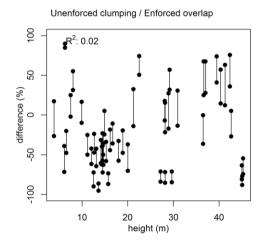
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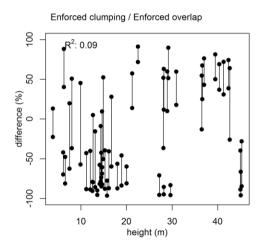
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**Figure A6**: TROBIT plot upper stratum height versus the difference between MODIS VCF and TROBIT percent tree cover in the four clumping-overlap scenarios. Upper and lower bars represent the uncertainty range's 10<sup>th</sup> and 90<sup>th</sup> percentile, respectively, based on the convolution of MODIS VCF and TROBIT cover uncertainties from Fig. 1.

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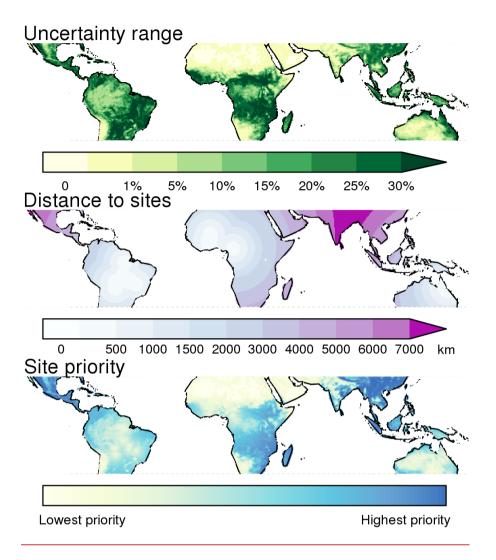


Figure A7. (Top) Uncertainty range of potential MODIS VCF mismatch, calculated as the 90<sup>th</sup> percentile (the highest value out of the four scenarios in Fig. 2) minus 10<sup>th</sup> percentile (the lowest value out of the four scenarios in Fig. 2). (Middle) Geographic distance to the closest TROBIT site sampled. (Bottom). Regions coloured to denote priority for field surveying to constrain map uncertainty, based on multiplying the (Top) and (Middle) maps.

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### Code/Data Availability.

The code and data used to support the findings of this study are archived at

1 155 <a href="https://github.com/douglask3/VCF">https://github.com/douglask3/VCF</a> vs sites revision number <a href="fdda3ff">fdda3ff</a>

#### **Author Contribution.**

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RA, DK, FGR.A., D.K., F.G., designed the TROBIT, MODIS VCF pixel-site comparison technique. RA, FGR.A., F.G., collated TROBIT and corresponding MODIS VCF values. DKD.K., and NDN.D., performed regression analysis and constructed global maps. RAR.A., wrote the first draft of the paper with input from DKD.K., and FG DKF.G. D.K., plotted the figures. MTR, EV, TRF, OLP, SL, BS, HT..., BSM, TD, LA, GD, M.T.R, E.V., T.R.F., O.L.P., S.L., B.S., H.T..., B.S.M., T.D., L.A., G.D., and JKJ.K., carried out the extensive TROBIT field campaigns. M.T.R. was the main person responsible for field data quality checking and digitising. RA, DK, FGR.A., D.K., F.G., and NDN.D., contributed to the final manuscript.

### Competing Interests.

1 165 The authors declare that they have no conflict of interest.

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Maps in Fig. 2 and A1 were constructed using raster2.6-7 (Hijmans, 2017) and mapproj1.2-5 (Brownrigg et al., 2017) in R 3.2.0 (R Core Team, 2015).(Hijmans, 2017) and mapproj1.2-5 (Brownrigg et al., 2017) in R 3.2.0 (R Core Team, 2015). Coastlines were obtained from mapsv3.1.0 (Becker et al., 2016).

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