Persistent impacts of the 2018 drought on forest disturbance regimes in Europe

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Abstract. Europe was affected by an extreme drought in 2018, compounding with an extensive heatwave in the same and subsequent years. Here we provide a first rapid assessment of the impacts this compounding event had on forest disturbance regimes in Europe. We find that the 2018 drought has caused unprecedented levels of forest disturbance across large parts of Europe, persisting up to two years post drought. The 2018 drought pushed forest disturbance regimes in Europe to the edge of their past range of variation, especially in Central and Eastern Europe. Increased levels of forest disturbance were statistically linked to low soil water availability in 2018, and were further modulated by high vapor pressure deficit from 2018 to 2020. We also document the emergence of novel spatiotemporal disturbance patterns following the 2018 drought (i.e., more and larger disturbances, occurring with higher spatiotemporal autocorrelation) that will have long-lasting impacts on forest structure, and raise concerns about a potential loss of forest resilience. We conclude that the 2018 drought had unprecedented impacts on forest disturbance regimes in Europe, highlighting the urgent need to adapt Europe's forests to a hotter and drier future with more disturbance.

Main text

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Europe was affected by a severe drought in 2018, characterized by extreme soil moisture deficits and high atmospheric aridity (Peters et al., 2020). The event was consistent with emerging climatic extremes under global change, characterized by prolonged precipitation-free periods coinciding with elevated water loss due to increased temperatures (Ault, 2020). Such compounding drought events are thought to be major drivers of forest disturbances though direct tree mortality, and through

25 facilitating insect outbreaks and wildfire (Allen et al., 2015; Brodribb et al., 2020). Increased forest disturbances from droughts can push ecosystems beyond their historic range of variation (Johnstone et al., 2016), leaving the 'safe operating space' these systems have functioned in for decades to centuries. As a consequence, emerging novel drought regimes pose a substantial thread to global forest resilience (Trumbore et al., 2015; Millar and Stephenson, 2015). In Europe, drought is considered a major driver of forest disturbances (Senf et al., 2020). Increased forest mortality and early leaf-shedding have also been

30 reported in response to the 2018 drought (Schuldt et al., 2020; Brun et al., 2020). Yet, evidence remains largely anecdotal and the large-scale effect of the 2018 drought on forest disturbance regimes in Europe remains yet not quantified. We here conducted a rapid assessment of the 2018 drought impacts on forest disturbance regimes in Europe by providing an update of a satellite-based pan-European forest disturbance map (Senf and Seidl, 2021) until 2020, and by analyzing changes in disturbance regimes following the 2018 drought. We hypothesized that the low soil moisture availability following the 2018

35 drought and the high atmospheric water demand in 2018-2020 would lead to persistent increases in forest disturbances, which have pushed forest disturbance regimes to the edge of their past range of variation.

We found a substantial increase (up to +500 % compared to the average of 1986 – 2015; Fig. 1) in forest disturbances in large parts of Europe in 2018, which was significantly related to soil moisture and vapor pressure deficit anomalies in the summer of 2018 (Fig. 2). The increase in disturbance was persistent beyond 2018, with disturbance rates remaining above

40 average until 2020 (Fig. 1). The high disturbance levels of 2019 and 2020 could also be explained by soil moisture anomalies in 2018 (Fig. 2), suggesting that the 2018 drought had persistent impacts on forest disturbances for at least two years post drought. Overall, 2018 summer soil moisture and annual vapor pressure deficit anomalies explained 11.5 % of the variance in the forest disturbance anomaly across Europe for 2018 – 2020. We further found strong evidence for an increasingly important interaction between soil moisture and vapor pressure deficit anomalies. Specifically, we found highest forest disturbance

45 anomalies in areas that were affected by both low soil moisture and high vapor pressure deficit (Fig. 2), highlighting the importance of both extreme soil moisture deficits and high atmospheric aridity during the drought of 2018.
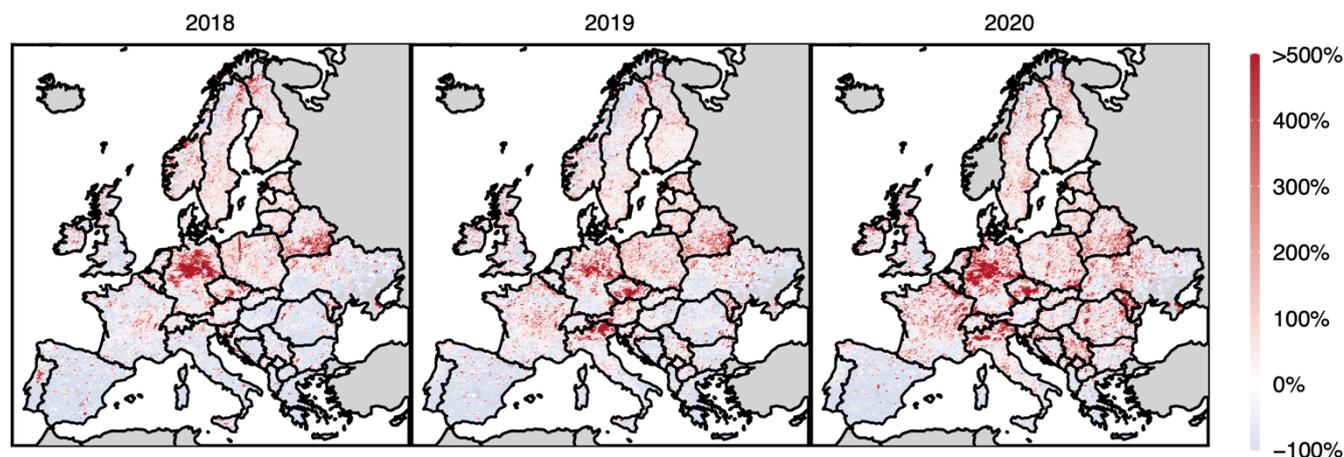
Figure 1: Forest disturbance anomalies in the years 2018-2020 in reference to 1986-2015, estimated from satellite-based disturbance maps across Europe. Anomalies are expressed in percent, that is +100% indicates a doubling of disturbed area in reference to the average disturbed area 1986-2015. Anomalies were calculated at a grid of ~9 km. Background maps are from https://gadm.org.
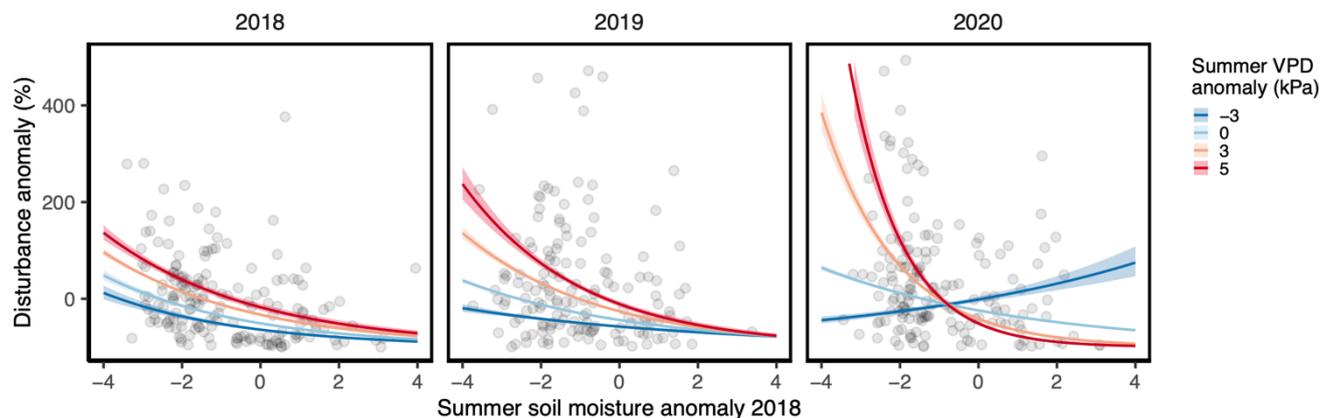
Figure 2: Relationship between forest disturbance anomaly in 2018, 2019 and 2020 (see Fig. 1) with local summer (June, July and August) soil moisture and summer vapor pressure deficit (VPD) anomalies (both in reference to 1986 – 2015). The black dots show a sample (~1 %) of the raw data. Ribbons around solid lines indicate the 95 % confidence interval. Note that disturbance anomalies were capped at +500 to improve visibility.

Based on our rapid assessment, we estimate that approximately 1.56 million hectares of forest were disturbed in Europe in 2018, and that 4.74 million hectares were disturbed over the period 2018 – 2020. This is an annual surplus of ~360.000 hectares for 2018 – 2020 compared to the average canopy mortality in 1986 – 2015. The strongest increase in forest disturbances was observed in Central Europe (Fig 3; mostly Germany, Czechia and Austria; Supplementary Table 1) and Eastern Europe (Fig. 3; Belarus and Ukraine; Supplementary Table 1). Yet also in Northern Europe mortality rates were among

the highest observed over the past 35 years (Fig. 3). In contrast, canopy mortality rates in Western and Southern Europe, i.e., areas not as strongly affected by the extreme drought of 2018, remained within their recent range of variation (Fig. 3).
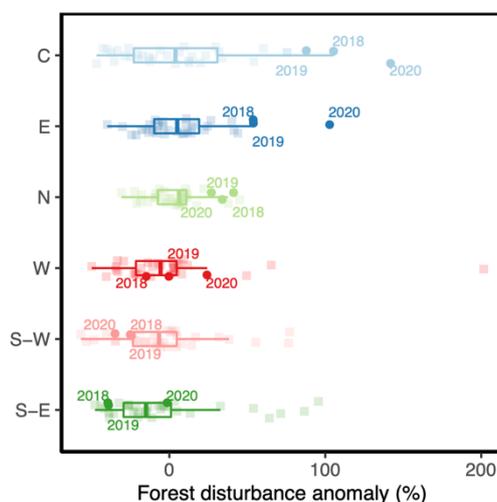
Figure 3: Forest disturbance anomalies at the regional level in reference to 1986-2015, with the years 2018-2020 highlighted. Anomalies are expressed in percent, that is +100% indicates a doubling of the annual area disturbed in reference to the average disturbance area 1986-2015. Abbreviations for the regions are: C = Central Europe; E = Eastern Europe; N = Northern Europe; S-E = South-Eastern Europe; S-W = South-Western Europe; W = Western Europe. See Supplementary Table 1 for details at the country level.

The persistent and widespread increase in forest disturbances after the 2018 drought suggests that – in addition to direct drought-related tree mortality (Choat et al., 2018) – indirect drought effects in the subsequent years were a major driver of increased disturbances. A particularly important indirect drought effect is the facilitation of insect disturbances (Allen et al., 2015; Seidl et al., 2017). In Central and Eastern Europe, large-scale outbreaks of bark beetles (mostly *Ips typographus* L.) led to a strong increase in infested conifers after 2018. According to national felling statistics, drought and insect activity nearly brought regular forestry to a halt in these regions, with at least ~ 50 % (Austria and Germany) and up to > 90 % (Czechia) of all harvests in 2019 being related to salvage logging after bark beetle (Knížek and Liška, 2020; Destatis, 2020; BMLRT, 2020). Widespread bark beetle mortality also explains the strong increase in forest disturbances in Belarus and Ukraine, where *Ips acuminatus* Gyll. caused widespread pine dieback (Food and Agricultural Organization of the United Nations, 2018). In addition to biotic disturbances, also fire activity increased in the areas affected by the 2018 drought. For example, Finland, Sweden and Norway experienced the highest fire activity on record in 2018, and sharp increases in area burned were also reported for many countries in Central Europe(San-Miguel-Ayanz et al., 2018, 2019). Yet, fire still only plays a minor role in both Central and Northern Europe, and was responsible for only ~ 3 % of the total disturbance area in 2018. Also, two major storm events occurred in 2018 affecting Poland and Northern Italy, constituting disturbances causally not related to the 2018 drought but emerging in our analysis (Fig 1). These two storms, while being the most extensive pulses of disturbances in the

affected regions for many decades, only explained ~80.000 ha of the 1.56 million ha of forest disturbances recorded for 2018 in our analysis (Forzieri et al., 2020).

The unprecedented and persistent increase in forest disturbances reported herein will have long-lasting impacts on forest dynamics in Europe. In the past, wind was the most important natural disturbance agent on the continent (Schelhaas et al., 2003; Seidl et al., 2014). The single largest forest disturbance event reported in Europe since 1850 was the storm 'Lothar' in the winter of 1999/2000 (Gardiner et al., 2010). We here show that current forest disturbance levels exceeded this past extreme. In 2020, the level of forest disturbances was 1.42 times higher than in the year 2000 (i.e., the year in which we record the impact of storm 'Lothar'). This indicates that the drought of 2018 might be responsible for the biggest pulse of disturbances in Europe in at least 170 years (Schelhaas et al., 2003). This period of forest disturbance can have profound and long-lasting impacts on the structure of Europe's forests. Specifically, we found that not only the amount, but also the size, frequency and aggregation of forest disturbances increased beyond historic levels in 2018 – 2020 (Fig. 4). These attributes are of high relevance for forest dynamics as they shape forest development trajectories for decades to centuries, and are determinants of the resilience of forest ecosystems (Scheffer et al., 2015; Johnstone et al., 2016). While forests have returned swiftly to their historical attractor after past large-scale perturbations (such as storm 'Lothar'; Fig. 4), the 2018 drought has pushed forest disturbance regimes in Europe past their basin of attraction for at least three consecutive years, and it remains unclear if the disturbance regime will return within the next years. A continuation of Europe's forests along this new trajectory of increasing frequency, size, and aggregation of disturbances might result in the crossing of tipping-points, causing pervasive and irreversible shifts in forest ecosystem structure and functioning (McDowell et al., 2020; Anderegg et al., 2012).
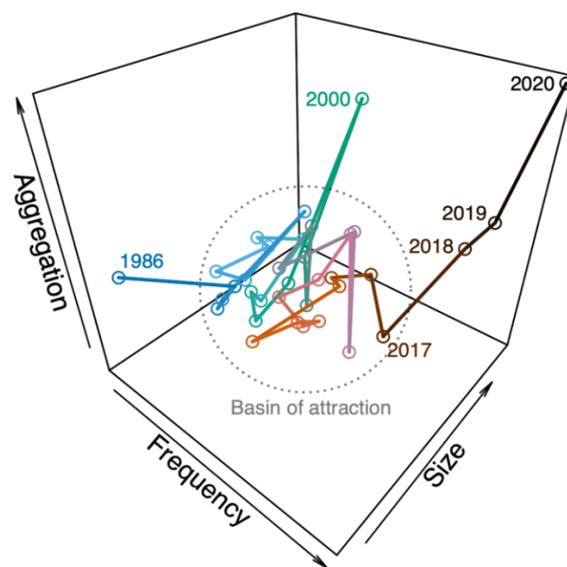
Figure 4: The development of disturbance regime characteristics in Europe's forests 1986 – 2020. Frequency denotes the average number of disturbances per unit forest area and year, size is the 95 % quantile of the patch size distribution of disturbances, and aggregation the average spatiotemporal autocorrelation of disturbance patches. The drought of 2018 has pushed Europe's forests disturbance regimes outside of their past basin of attraction.

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We here provide a first rapid assessment of the impacts of the 2018 drought on forest disturbance regimes in Europe. Our analyses show that forest disturbance regimes in Europe have changed profoundly following the drought of 2018, and suggest that this change is mainly driven by the drought of 2018. While we provide strong evidence based on satellite data, we note that satellite-based assessments only provide a coarse-scale view of ecosystem dynamics. Further research is thus needed

115   to improve our understanding of the impacts of the 2018 drought at the local and regional scale. Our assessment can help to guide these research efforts, and provide information needed to adapt forests to a hotter and drier future with more disturbance. Future projections indicate that drought events such as the one observed in 2018 will become the new normal in the near future (Samaniego et al., 2018; Toreti et al., 2019). Waves of forest disturbances such as the one observed in recent years are thus likely also in the coming decades. We conclude that the causes and consequences of changing forest disturbance regimes

120   should be a key priority for science and policy.

Materials and Methods

We updated an existing pan-European forest disturbance map originally covering the time period 1986-2016 (Senf and Seidl, 2021) until the year 2020. We applied the same workflow as used for creating the first version of the map in order to ensure consistency over time. The initial map product had an overall accuracy of 92.5 ± 2.1 % with a disturbance commission error

125   of 14.6 ± 1.8 % and a disturbance omission error of 32.8 ± 0.3 %, indicating that the map is conservative (i.e., higher omission of true disturbances than commission of false disturbances). We performed a visual quality screening of the map update and could not identify any inconstancies that might flag a rapid decrease in map accuracy for the recent years. Yet, due to a limited number of clear satellite observations in Norway for the year 2020, we identified some artefacts stemming from clouds in the final maps for Norway. To reduce bias in our analysis, we excluded data from Norway in 2020. The updated map products are

130   available at https://zenodo.org/record/4570157#.YIrB0C0evOQ (Version 1.1.0).

We aggregated the disturbance map from its native 30 m resolution to a regular grid of 0.1° (~ 9 km) by calculating the absolute annual (t) area of forest disturbed (D_{ti}) per grid cell i. From the absolute annual area disturbed we subsequently calculated the long-term average annual canopy disturbance area for the period $1986 – 2015$ as reference ($D_{i,ref}$), in order to estimate the annual fractional anomaly A_{it} as $A_{it} = D_{it}/D_{i,ref}$. In the following, we refer to A_{it} as annual forest disturbance

135   anomaly per grid cell. The forest disturbance anomaly is the percent deviation of annual forest area disturbed relative to the long-term ($1986 – 2015$) mean. As anomalies can become unreliable when the reference level $D_{i,ref}$ is very low (i.e., a very small absolute increase can lead to a very large anomaly in such cases), we excluded all grid cells with < 1 ha of disturbance per year on average from the analysis (excluding n = 481,770 cells, representing 18 % of all cells). Besides calculating anomalies for each grid cell, we also calculated them for six European regions, first aggregating annual area disturbed to

140   regional level and subsequently calculating the anomalies. The regions considered were East (Belarus, Moldova, Ukraine), Central (Austria, Czechia, Germany, Hungary, Poland, Slovakia, Slovenia, Switzerland), West (Belgium, France, Ireland,

Netherlands, United Kingdom), North (Denmark, Estonia, Finland, Latvia, Lithuania, Norway, Sweden), South-East (Albania, Bosnia and Herzegovina, Bulgaria, Croatia, Greece, Montenegro, Romania, Serbia), and South-West (Italy, Portugal, Spain).

In addition to the forest disturbance anomaly, we also characterized changes in forest disturbance regimes in response
145 to the 2018 drought. Specifically, we calculated the patch size for each individual disturbance patch in Europe (n > 35 million patches), as well as the disturbance frequency (expressed as number of patches per hectare forest area per year). We further characterized the spatiotemporal aggregation of disturbance patches by calculating the proportion of all disturbance patches that occurred in the same year in a five-kilometer radial kernel around each individual disturbance patch. A value of one indicates that all disturbances in close proximity happened in the same year as the focal patch (high spatiotemporal
150 autocorrelation), whereas a value of zero indicates no other disturbances occurred in the same year and in proximity to the focal patch. This measure broadly quantifies the press-pulse-dichotomy of human versus natural disturbance regimes (Sebald et al., 2019). We finally aggregated all three measures to annual values by calculating the 95th quantile for patch sizes and the average of frequency and spatiotemporal aggregation. We used the 95th percentile for patch sizes instead of the average, as patch-size distributions are highly left-skewed with very heavy right tales, which can obscure the calculation of average patch
155 sizes. The 95th percentile gives a better indication of the width of the patch size distribution than the average.

To assess the impacts of the 2018 drought on disturbances, we used the most recent European Center for Medium-Range Weather Forecast (ECMWF) ERA5-land reanalysis data, which has a spatial resolution of 0.1° (~ 9 km) and is available from 1979 to present. We extracted the monthly averaged volumetric soil water content from 0 to 289 cm over June to August. We scaled the data to anomalies via z-transformation, using the mean and average of the reference period 1986 – 2015. We
160 further acquired mean temperate and mean dew point temperate for June to August to derive the mean summer vapor pressure deficit following formulas described in (Seager et al., 2015). We finally modelled the spatial variability in forest disturbance anomaly among grid cells (A_{it}) for the years 2018, 2019 and 2020 using soil moisture anomalies from 2018 (sm_{2018}), characterizing the overall drought effect, and vapor pressure deficit anomalies (vpd_{ti}) from 2018 to 2020, characterizing the confounding role of subsequent atmospheric drought (resulting from high temperatures) on the overall lack of soil moisture
165 caused by the 2018 drought. Anomalies were calculated in reference to the period 1986 – 2015. We expected that the soil moisture anomaly of 2018 can explain disturbance anomalies in 2018, 2019 and 2020 due to legacy effects of the 2018 drought on subsequent years. We furthermore expected that the strength of association would be significantly modulated by annual vapor pressure deficit anomalies. We included year as dummy variable to account for differences among years in both the average disturbance anomaly, as well as strength of association between predictors and response. We used a log-linear model
170 with Gaussian error distribution fitted in the statistical software R (R Core Team, 2020). We compared all potential alternative model formulations using an exhaustive model search implemented in the *MuMIn* package (Bartoń, 2009) and the model with the smallest AIC was selected as final model.

Data and code availability

All data and code are available under https://github.com/corneliussenf/Drought2018. A permanent version of this repository
175 will be made available upon publication of the manuscript.

Author contribution

CS and RS designed the research. CS conducted all analysis. CS wrote the manuscript, with revision from RS.

Competing interest

The authors declare that they have no conflict of interest.

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