

1 **Effect of the presence of plateau pikas on the ecosystem services of alpine meadows**

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12

13 **Abstract**

14 The activity of small mammalian herbivores influences grassland ecosystem services in arid
15 and semi-arid regions. Plateau pika (*Ochotona curzoniae*) was considered as a focal organism
16 to investigate the effect of small mammalian herbivores on meadow ecosystem services in
17 alpine regions. In this study, a home-range scale was used to measure the forage available to
18 livestock, water conservation, carbon sequestration, and soil nutrient maintenance (total
19 nitrogen, phosphorus, and potassium) in the topsoil layer; and a quadrat scale was used to
20 assess the biodiversity conservation of alpine meadows. This study showed that the forage
21 available to livestock and water conservation were 19.74% and 15.86% lower in the presence
22 of plateau pikas than in their absence, and biodiversity conservation, carbon sequestration,
23 soil nitrogen, and phosphorus maintenance were 14.58%, 29.15%, 9.97% and 8.89% higher in
24 the presence of plateau pikas than in their absence. In contrast, it had no impact on soil
25 potassium maintenance of meadow ecosystems in alpine regions. With the increase of
26 disturbance intensity of plateau pikas, the forage available to livestock, biodiversity
27 conservation, and soil nutrient maintenance of meadow ecosystems in alpine regions first
28 increased and then decreased; the water conservation tended to decrease linearly with the
29 increasing disturbance intensity of plateau pikas. These results present a pattern of plateau
30 pikas influencing the ecosystem services of meadow ecosystems in alpine regions, richening
31 the small mammalian herbivores in relation to grassland ecosystem services.

32 **1 Introduction**

33 Grasslands provide multiple ecosystem services, mainly including provisioning services
34 of food and water, regulating services of carbon sequestration and water conservation,
35 supporting services of soil nutrient maintenance and biodiversity conservation, and cultural

36 services of landscapes and recreation tourism (Millennium Ecosystem Assessment, 2005).
37 These ecosystem services sustain animal production, flora and fauna, and other human
38 welfare (Costanza et al., 1997; Zhang et al., 2018; Dong et al., 2020); however, they are
39 affected by multiple biotic factors, such as soil microbial communities (Van Eekeren et al.,
40 2010), grazing by large herbivores (Lu et al., 2017), and the presence of small herbivores
41 (Delibes-Mateos et al., 2011; Martínez-Estévez et al., 2013).

42 Small mammalian herbivores are common biotic factors (Davidson et al., 2012). These
43 herbivores usually create extensive disturbances on grassland vegetation and soil (Pang et al.,
44 2020a, 2020b) by developing burrow systems (Delibes-Mateos et al., 2008; Sun et al., 2015),
45 excreting feces and urine (Zhang et al., 2016), consuming plants (Eldridge and Myers, 2001;
46 Liu et al., 2017), clipping tall plants (Zhang et al., 2020), and producing bare soil patches
47 (Guo et al., 2012a, 2012b; Yu et al., 2017a, 2017b) or mounds (Yang et al., 2021). Previous
48 studies have shown that the presence of prairie dogs (*Cynomys ludovicianus*) can increase the
49 forage available to livestock, water conservation, carbon sequestration, and biodiversity
50 conservation of grassland ecosystems in arid regions (Ceballos et al., 1999, Martínez-Estévez
51 et al., 2013), whereas the presence of European rabbit (*Oryctolagus cuniculus*) can decrease
52 the forage available to livestock (Delibes-Mateos et al., 2008; Eldridge and Myers, 2001), and
53 increase the biodiversity conservation (Delibes-Mateos et al., 2008) and nitrogen maintenance
54 (Willott et al., 2000) of grassland ecosystems in semi-arid regions. In addition to grasslands in
55 arid and semi-arid regions, vast alpine meadows exist in high latitude and altitude regions
56 throughout the world (Zhang et al., 2018; Dong et al., 2020). However, how small
57 mammalian herbivores influence the ecosystem services in alpine meadows as much as they

58 do in arid and semi-arid regions has not been well documented.

59 The plateau pika (*Ochotona curzoniae*) is a common, small mammalian herbivore that
60 mainly lives in alpine meadows of the Qinghai-Tibetan Plateau (Smith and Foggin, 1999).
61 This small mammalian herbivore with an average weight of 150 g are diurnally active and
62 non-hibernating (Smith and Wang, 1991; Fan et al., 1999), and preferentially consume
63 dicotyledons (Zhao et al., 2013; Pang and Guo, 2017). Plateau pikas, a sexual monomorphism
64 (Dobson et al., 1998) , often construct a family warren with numerous burrow entrances and
65 develop a complex burrow system with an average length and depth of 13 m and 30 cm (Fan
66 et al., 1999). This mammalian herbivore is social and philopatric (Dobson et al., 1998) and its
67 young offspring stay with its family during its birth year (Wang et al., 2020). Plateau pikas are
68 generally considered a pest in China (Harris, 2010; Pang and Guo, 2017) as they often
69 exacerbate the degradation of alpine meadows (Liu et al., 2013; Zhang et al., 2016). However,
70 some studies have argued that plateau pika is a key species in alpine meadow ecosystems
71 (Smith and Foggin, 1999; Delibes-Mateos et al., 2011). This disagreement has encouraged
72 professionals to re-evaluate the role of plateau pikas in alpine meadow ecosystems. Thus, the
73 effects of plateau pikas' presence on ecosystem services of alpine meadows allow insight into
74 the role of plateau pikas in alpine meadow ecosystems. Previous studies have demonstrated
75 that the presence of plateau pikas decreases (Liu et al., 2013) or has no significant effect on
76 plant biomass (Pang and Guo, 2017), increases (Liu et al., 2017; Pang and Guo, 2017) or
77 decreases (Sun et al., 2015) plant-species richness, and increases (Yu et al., 2017a; Pang et al.,
78 2020a, 2020b) or decreases (Sun et al., 2015) soil carbon and nutrients. In addition, previous
79 studies have shown that the disturbance intensity of plateau pikas affects plant-species

80 richness, and soil nutrient stocks of alpine meadows (Yu et al., 2017a; Pang and Guo, 2018).
81 These findings imply that plateau pikas may have an impact on the ecosystem services of
82 alpine meadows. Thus, further studies are needed to test whether the presence of plateau pikas
83 and its disturbance intensity influence the ecosystem services of alpine meadows, which can
84 richen the presence of small mammalian herbivores in relation to grassland ecosystem
85 services.

86 Since soil carbon and nutrients differ between vegetated and bare soil patches in the
87 home range (Yu et al., 2017b), Pang et al. (2020a; 2020b) proposed that the home-range scale
88 is a better proxy than the quadrat scale to estimate the complete effects of the presence of
89 plateau pikas on soil carbon and nutrient stocks. Although the provisioning, regulation,
90 support, and cultural services of alpine meadows can be estimated by multiple indicators
91 (Egoh et al., 2012; Brown et al., 2014), one or two can be used to verify whether the presence
92 and intensity of plateau pikas influence each ecosystem service. In previous studies, palatable
93 plant biomass for livestock has been used to evaluate the provisioning services
94 (Martínez-Estévez et al., 2013; Wen et al., 2013); soil-water storage and soil organic carbon
95 stock have been used to evaluate the regulating services (Wen et al., 2013; Li and Xie, 2015;
96 Tang et al., 2019;); and plant-species richness and soil total nutrient stocks can be used to
97 evaluate the supporting services (Wen et al., 2013). Notably, cultural services are particularly
98 related to the spatial scale, as many are perceived visually over distant views (Norton et al.,
99 2012). The plateau pika is territorial and its habitat use is patchy within a given area (Pang et
100 al., 2020a), which leads to mismatches between the spatial scale and cultural services (De
101 Groot et al., 2010). Therefore, the present study used ecological services of forage available to

102 livestock, water conservation, carbon sequestration and soil nutrient maintenance, and
103 biodiversity conservation to test how the presence of plateau pikas influences the ecosystem
104 services of alpine meadows across five sites. In this study, we hypothesized that (1) the
105 presence of plateau pikas leads to lower forage available to livestock because of lower
106 palatable plant biomass in the presence of small mammalian herbivores; (2) the presence of
107 plateau pikas leads to higher water conservation and carbon sequestration because small
108 mammalian herbivores can increase soil-water storage and carbon stocks; and (3) the
109 presence of plateau pikas leads to higher biodiversity conservation and soil nutrient
110 maintenance because small mammalian herbivores can increase plant-species richness and
111 soil nutrient stocks.

112 **2 Materials and methods**

113 **2.1 Study site descriptions**

114 Plateau pikas can live in various habitats with different soil types, topographies, and
115 microclimates on the Qinghai-Tibetan Plateau. To determine how the presence of plateau
116 pikas generally influences the ecosystem services of alpine meadows, five survey sites were
117 selected in Luqu (102°22'12"E, 34°15'51"N), Gangcha (100°26'26"E, 37°36'12"N), Haiyan
118 (100°54'33"E, 36°57'50"N), Qilian (100°34'48"E, 37°43'26"N), and Gonghe (99°47'11"E,
119 36°43'48"N) counties on the Qinghai-Tibetan Plateau. These five survey sites have a similar
120 typical plateau continental climate, with elevations ranging from 3194 m at the Gonghe
121 survey site to 3550 m at the Luqu survey site. Based on 5-year weather data, the mean annual
122 temperatures are 3.1, 0.9, 1.9, 2.2, 3.3°C at Luqu, Gangcha, Haiyan, Qilian, and Gonghe,
123 respectively, of which the average temperature in warm season are 9.3, 8.3, 9.6, 10.3, 9.9°C

124 and in cold season are -3.1, -6.5, -5.8, -5.9, -3.4°C. The mean annual precipitation is 439.5,
125 258.9, 257.4, 257.0, and 239.8 mm, of which the warm season accounts for 83.4%, 92.8%,
126 89.3%, 91.5%, 91.4% at Luqu, Gangcha, Haiyan, Qilian, and Gonghe, respectively.
127 According to the Chinese soil classification system (Gong, 2001), the soil type at each site is
128 alpine meadow soil, similar to Cambisol in the WRB soil classification system.

129 Animal husbandry is the dominant use of alpine meadows on the Qinghai-Tibetan
130 Plateau, and herders traditionally graze their livestock seasonally on cold and warm
131 grasslands. The survey sites in this study were all situated in cold grasslands, in which alpine
132 meadows were fenced from mid-April to September, and fences were opened to grazing yaks
133 from mid-October to early April (Zhang et al., 2020). All field data were collected in August
134 when the annual population of plateau pikas was the highest and reproduction had largely
135 ceased (Qu et al., 2013; Pang et al., 2020a, 2020b). In addition, the growing season for plants
136 is short on the Qinghai-Tibetan Plateau, and some plants **don't** turn green until July. Therefore,
137 sampling in August is optimal because August is good time to identify all plants and ensure
138 an accurate survey of the plant species. Notably, the small burrowing herbivore at each survey
139 site was only plateau pikas.

140 **2.2 Field survey design**

141 The adult dispersal of plateau pikas is a gradual process (Pang et al., 2020a), it is easy to
142 identify reference sites without plateau pikas, even though these sites might be potential as
143 suitable habitats. In this study, a home-range scale was used to calculate the forage available
144 to livestock, water conservation, carbon sequestration, and soil nutrient maintenance, and a
145 quadrat scale was used to calculate the biodiversity conservation.

146 A stratified random and paired design was used to select plots. The home range of the
147 plateau pika was approximately 1262.5 m² (Fan et al., 1999), and the plot size was 35 × 35 m,
148 which was similar to the average area of the plateau pika's home range. At each of the five
149 sites, this study first selected 10 plots where plateau pikas were present, or where active
150 burrow entrances were observed. The second plot was identified along the road when the first
151 plot with plateau pikas was selected. The distance between the two plots with plateau pikas
152 was more than 3 km, which ensured that plateau pikas of the same family would not appear in
153 two plots at the same time. Second, a paired adjacent plot without plateau pikas and active
154 burrow entrances was selected for each plot with plateau pikas. The plots without plateau
155 pikas were in any direction of plots with plateau pikas. The distance between each plot with
156 plateau pikas and its paired plot without plateau pikas ranged from 500 to 1000 m. If the
157 distance between each paired plot was too close, the plateau pikas could move between plots
158 with and without plateau pikas. To ensure that each plot with plateau pikas was paired with a
159 plot without plateau pikas, each paired plot shared the alpine meadow with same dominant
160 plant, with no obvious differences in soil type, topography, or microclimate. In total, there
161 were 10 paired plots at each site and 100 plots across five sites, including 50 with and 50
162 without plateau pikas. Each paired plot shared the same grazing intensity during the cold
163 season; however, 50 paired plots consisted of different yak grazing intensity, and this can
164 permit the general pattern relating to the effect of plateau pika disturbance on alpine meadow
165 ecosystem services.

166 **2.3 Field sampling**

167 Field surveys and sampling were conducted in early August 2020. First, the active

168 burrow entrance at each plot with plateau pikas was estimated by the “plugging tunnels
169 method,” in which the burrow entrances were plugged with hay for 3 days (Sun et al., 2015),
170 and the number of plugs cleared by the plateau pikas to allow access to the meadow surface
171 was recorded (Guo et al., 2012a). The average number of burrow entrances with cleared plugs
172 after 3 days was taken as the density of active burrow entrances. For plots with plateau pikas,
173 the density of active burrow entrances was used as a proxy for the intensity of the disturbance
174 (Guo et al., 2012a; Sun et al., 2015). Second, this study was restricted to plateau pikas in
175 relation to the ecosystem services of alpine meadows. However, bare soil patches caused by
176 other factors (no plateau pikas) is simultaneously existed on the vegetated surface in the
177 presence/absence of plateau pikas. To actual quantify the effect of plateau pikas on ecosystem
178 services of alpine meadows, this study only measured the area of bare soil patches caused by
179 plateau pikas, although there exist multiple types of bare soil patches in alpine meadows. **The**
180 **soil bare patches caused by plateau pikas is easily to identify because one soil bare patch**
181 **caused by plateau pikas is paired with a visible burrow entrance (Pang et al., 2021a). Other**
182 **soil bare patches are not paired with visible burrow entrance.** The area of each bare soil patch
183 (created by plateau pikas) in the plot with plateau pikas was measured. Each bare soil patch
184 was identified as regular shape or irregular shape. If one bare soil patch was identified as
185 regular shape, such as rectangle, circle, trapezoid, etc; a ruler was used to measure its length,
186 width, height, diameter, upper and lower bottom, and these data was used to calculate the area
187 of that bare soil patch. If one bare soil patch was identified as irregular shape, this bare soil
188 patch was divided into several regular shapes; the areas of these regular shapes were
189 calculated, respectively; the area sum of these regular shapes form irregular bare soil patch

190 was considered as the area of that irregular bare soil patch (Han et al., 2011). Then, the sum of
191 all bare soil patch areas in each plot with plateau pikas was calculated and defined as the bare
192 soil area for that plot. Third, five vegetated quadrats (1×1 m) were placed on the vegetated
193 surface approximately 8 m apart along the shape of a W pattern in all plots (with or without
194 plateau pikas). In plot with plateau pikas, if quadrat was justly covered with the bare patches
195 caused by plateau pikas, the quadrat was slightly moved to avoid it; if quadrat was justly
196 covered with the bare patches caused by other factors, the quadrat was not moved. Fourth,
197 alpine meadows in plot with plateau pikas consisted of bare and vegetated surface, and a bare
198 soil patch was selected as a paired bare soil quadrat for each vegetated quadrat in the plot with
199 plateau pikas, and the distance between each paired bare soil quadrat and vegetated quadrat
200 was as short as possible (less than 1 m). Bare soil patch quadrat and each vegetated quadrat
201 were beneficial to accurately measure the soil nutrient, carbon concentrations and plant
202 biomass, which reflected the effect of plateau pikas' presence on ecosystem services in alpine
203 meadows by comparing the parameters between plots with and without plateau pikas at home
204 range scale. Thus, there were 5 paired quadrats, consisting of 5 vegetated and 5 bare soil
205 quadrats in each plot with plateau pikas. Additionally, there were 5 vegetated quadrats in each
206 plot without plateau pikas, since this study focused on bare soil patches induced by plateau
207 pikas.

208 In each vegetated quadrat of the plot with or without plateau pikas, all vascular plant
209 species were identified, and the number of plant species were recorded as plant-species
210 richness. Then, all plants rooted in a quadrat were harvested into palatable and unpalatable
211 plants, in which palatable plants or unpalatable plants were for yak and Tibetan sheep (Pang

212 and Guo, 2017). Finally, all palatable plant samples were placed in envelopes and transported
213 to the laboratory.

214 Generally, most burrows derived from the activities of plateau pikas are less than 20 cm
215 in depth (Yu et al., 2017b), although a few burrows extend to depths of 60 cm (Fan et al.,
216 1999). In addition, the majority of plant roots in alpine meadows of the Qinghai-Tibetan
217 Plateau are in the top 20 cm of the soil. The soil samples were collected at a depth of 20 cm.
218 Soil samples were collected from vegetated and bare soil quadrats for each plot with plateau
219 pikas and vegetated quadrats for each plot without plateau pikas. Before collecting the soil
220 samples, plants and litter were removed from the soil surface. First, a 5 cm diameter soil
221 auger was used to collect soil samples, which were used to measure soil organic carbon and
222 soil nutrient concentrations (total nitrogen, phosphorus, and potassium). Second, soil profiles
223 in each quadrat were obtained using a spade, and a stainless-steel cutting ring (with a volume
224 of 100 cm³) was used to collect soil cores to determine soil bulk density and soil water
225 content. Soil samples used to determine soil bulk density were packed into aluminum boxes
226 with recorded weights, and each aluminum box was numbered. The aluminum boxes
227 containing fresh soil were immediately weighed, recorded, stored at 4°C, and then transported
228 to the laboratory. Thus, in this study, 10 soil samples were collected to analyze the soil carbon,
229 nitrogen, phosphorus, and potassium concentrations, and 10 soil samples were obtained to
230 determine the soil bulk density in each plot with plateau pikas. Because this study is
231 constricted with bare soil patch caused by plateau pikas, bare soil quadrats was not considered
232 in plot without plateau pikas; therefore, 5 soil samples were used to determine the soil carbon,
233 nitrogen, phosphorus, and potassium concentrations, and 5 samples were obtained for the

234 analysis of soil bulk density in each plot without plateau pikas. 5 soil samples in each plot
235 were individually measured. The average value of five soil samples in one plot was
236 considered as the representative data of that plot.

237 **2.4 Analysis of samples**

238 In the laboratory, palatable plant samples were dried in an oven at 80°C for 48 h and
239 weighed. The soil samples used to measure soil bulk density and soil-water content were
240 dried to a constant weight at 105±2°C, and the aluminum boxes with dry soil were weighed
241 and the values were recorded. The soil samples used to measure soil organic carbon, total
242 nitrogen, phosphorus, and potassium concentrations were air-dried, gravel and roots were
243 manually picked out, and the proportion of gravel larger than 2.0 mm in the soil sample was
244 determined by passing through a 2.0 mm sieve. Finally, soil samples were sieved at 0.15 mm
245 to analyze soil organic carbon, nitrogen, phosphorus, and potassium concentrations in the
246 laboratory. Soil organic carbon was measured using the dichromate heating-oxidation
247 (Naelson and Sommers, 1982). Soil total nitrogen concentration was measured using the
248 Kjeldahl procedure. Soil total phosphorus concentration was measured using the
249 Molybdenum blue colorimetric method. Soil total potassium concentration was measured
250 using flame photometry.

251 Soil bulk density, soil organic carbon, and nutrient concentrations (total nitrogen,
252 phosphorus, and potassium) were used to calculate the soil organic carbon, total nitrogen,
253 phosphorus, and potassium stocks. Soil bulk density and soil-water content were used to
254 calculate the soil-water storage (Jia et al., 2020).

255 **2.5 Calculations**

256 The bare soil area consisted of all bare soil patches, and the vegetated surface area was
257 estimated from the plot areas minus the bare soil areas. This study only measured the area of
258 bare soil patches caused by plateau pikas, although there were other kinds of bare soil patches
259 in alpine meadows. Therefore, bare soil areas caused by plateau pikas were considered to be
260 zero in each plot without plateau pikas, and the vegetated surface area was considered to be
261 100%.

262 The palatable plant biomass was calculated using the following equation:

$$263 \quad GB = B_q \times \delta_{va} \quad (1)$$

264 where GB , B_q , and δ_{va} are the palatable plant biomass of the plot, palatable plant biomass
265 on the quadrat scale (g m^{-2}), and vegetated surface area, respectively.

266 The plant-species richness in a quadrat ($1 \times 1 \text{ m}$) was measured using the species number
267 of each quadrat.

268 Soil-water storage was determined using the method recommended by Jia et al. (2020),
269 and it was calculated by the following equation:

$$270 \quad SWS_{pika} = (SWC_{BA} \times BD_{BA} \times T \times (1 - \theta_{BA}) \times 0.01 \times BA) + (SWC_{VA} \times BD_{VA} \times T \times (1 - \theta_{VA}) \times 0.01 \times VA) \quad (2)$$

271 Where SWS_{pika} , SWC_{BA} , BD_{BA} , and θ_{BA} were soil-water storage in a plot with plateau
272 pikas, water content (g kg^{-1}), soil bulk density (g cm^{-3}) and soil fraction of gravel larger than 2
273 mm in bare soil areas of plots with plateau pikas, respectively; BA was the percentage of
274 bare soil areas in plots with plateau pikas; SWC_{VA} , BD_{VA} , and θ_{VA} were water content (g kg^{-1}),
275 soil bulk density (g cm^{-3}) and soil fraction of gravel larger than 2 mm in vegetated areas of a
276 plot with plateau pikas, respectively; and T was soil thickness (20 cm); VA was the percentage
277 of vegetated surface area in plots with plateau pikas; SWC_{BA} and SWC_{VA} was measured by

278 oven drying method.

$$279 \quad SWS_{no\ pika} = SWC_{no\ pika} \times BD_{no\ pika} \times T \times (1 - \theta_{no\ pika}) \times 0.01 \times 100\% \quad (3)$$

280 Where $SWS_{no\ pika}$, $SWC_{no\ pika}$, $BD_{no\ pika}$ and $\theta_{no\ pika}$ were soil-water storage in a plot without
281 plateau pikas, soil water content (g kg^{-1}), soil bulk density (g cm^{-3}) and soil fraction of gravel
282 larger than 2 mm in plots without plateau pikas, respectively; and T is soil thickness (20 cm).

283 The soil organic carbon stock per plot was estimated using the method described by Pang
284 et al. (2020b), and it was calculated by following equation:

$$285 \quad SOCS_{pika} = (SOC_{BA} \times BD_{BA} \times T \times (1 - \theta_{BA}) \times 0.01 \times BA) + (SOC_{VA} \times BD_{VA} \times T \times (1 - \theta_{VA}) \times 0.01 \times VA) \quad (4)$$

286 Where $SOCS_{pika}$ was soil organic carbon stock in a plot with plateau pikas (kg m^{-2});
287 SOC_{BA} , BD_{BA} , and θ_{BA} were soil organic carbon concentration (g kg^{-1}), soil bulk density (g
288 cm^{-3}) and soil fraction of gravel larger than 2 mm in bare soil areas of plots with plateau pikas,
289 respectively; BA was the percentage of bare soil areas in plots with plateau pikas; SOC_{VA} ,
290 BD_{VA} , and θ_{VA} were organic carbon concentration (g kg^{-1}), soil bulk density (g cm^{-3}) and soil
291 fraction of gravel larger than 2 mm in vegetated areas of a plot with plateau pikas,
292 respectively; and T was soil thickness (20 cm); VA was the percentage of vegetated surface
293 area in plots with plateau pikas.

$$294 \quad SOCS_{no\ pika} = SOC_{no\ pika} \times BD_{no\ pika} \times T \times (1 - \theta_{no\ pika}) \times 0.01 \times 100\% \quad (5)$$

295 Where $SOCS_{no\ pika}$ was soil organic carbon stock in the plot without plateau pikas (kg
296 m^{-2}); and $SOC_{no\ pika}$, $BD_{no\ pika}$ and $\theta_{no\ pika}$ were soil organic carbon concentration (g kg^{-1}), soil
297 bulk density (g cm^{-3}) and soil fraction of gravel larger than 2 mm in plots without plateau
298 pikas, respectively; and T was soil thickness (20 cm).

299 The soil total nitrogen, phosphorus, and potassium stocks per plot were quantified using
300 the method described by Pang et al. (2020a), and it was calculated by the following

301 equation:

$$302 \quad SNSi_{pika} = (SNi_{BA} \times BD_{BA} \times T \times (1 - \theta_{BA}) \times 0.01 \times BA) + (SNi_{VA} \times BD_{VA} \times T \times (1 - \theta_{VA}) \times 0.01 \times VA) \quad (6)$$

303 Where $SNSi_{pika}$ was soil total nitrogen, phosphorus, potassium stock in plot with plateau
304 pikas (kg m^{-2}), and SNi_{BA} , BD_{BA} , and θ_{BA} were soil nutrient concentration (g kg^{-1}), soil bulk
305 density (g cm^{-3}) and soil fraction of gravel larger than 2 mm in bare soil area of plots with
306 plateau pikas, respectively; BA was the percentage of bare soil areas in plots with plateau
307 pikas; SNi_{VA} , BD_{VA} , and θ_{VA} were soil nutrient concentration (g kg^{-1}), soil bulk density (g cm^{-3})
308 and soil fraction of gravel larger than 2 mm in vegetated areas of a plot with plateau pikas,
309 respectively; and T was soil thickness (20 cm); VA was the percentage of vegetated surface
310 area in plots with plateau pikas.

$$311 \quad SNSi_{no\ pika} = SNi_{no\ pika} \times BD_{no\ pika} \times T \times (1 - \theta_{no\ pika}) \times 0.01 \times 100\% \quad (7)$$

312 Where $SNSi_{no\ pika}$ was soil total nitrogen, phosphorus, potassium stock in the plot without
313 plateau pikas (kg m^{-2}), $SNi_{no\ pika}$, $BD_{no\ pika}$ and $\theta_{no\ pika}$ were soil nutrient concentration (g kg^{-1}),
314 soil bulk density (g cm^{-3}) and soil fraction of gravel larger than 2 mm in plots without plateau
315 pikas, respectively; and T was soil thickness (20 cm).

316 2.6 Data analysis

317 Data from 50 disturbed plots and 50 undisturbed plots were used to examine the
318 difference in ecosystem services of alpine meadows between the presence of plateau pikas
319 and the absence of plateau pikas; and then data from 50 disturbed plots were used to examine
320 the responses of each ecosystem service of alpine meadows to the disturbance intensity of
321 plateau pikas.

322 All data variables (palatable plant biomass, plant-species richness, soil-water storage,
323 soil organic carbon stock, soil total nitrogen stock, soil total phosphorus stock, and soil total

324 potassium stock) were assessed for the normality and homogeneity by using the Shapiro-Wilk
325 test. If necessary, the data were base-10 log-transformed to fit the assumption of normality
326 and homogeneity for further variance analysis.

327 A Linear Mixed Model (LMM) with the function “lmer” from the lme4 package was
328 used to examine differences in palatable plant biomass, plant-species richness, soil-water
329 storage, soil organic carbon stock, soil total nitrogen stock, soil total phosphorus stock, and
330 soil total potassium stock between the presence and absence of plateau pikas across the five
331 sites. In linear mixed models, the abovementioned parameters acted as response variables, the
332 absence/presence were introduced as fixed factor, and the paired plots nested within each site
333 as a random factor.

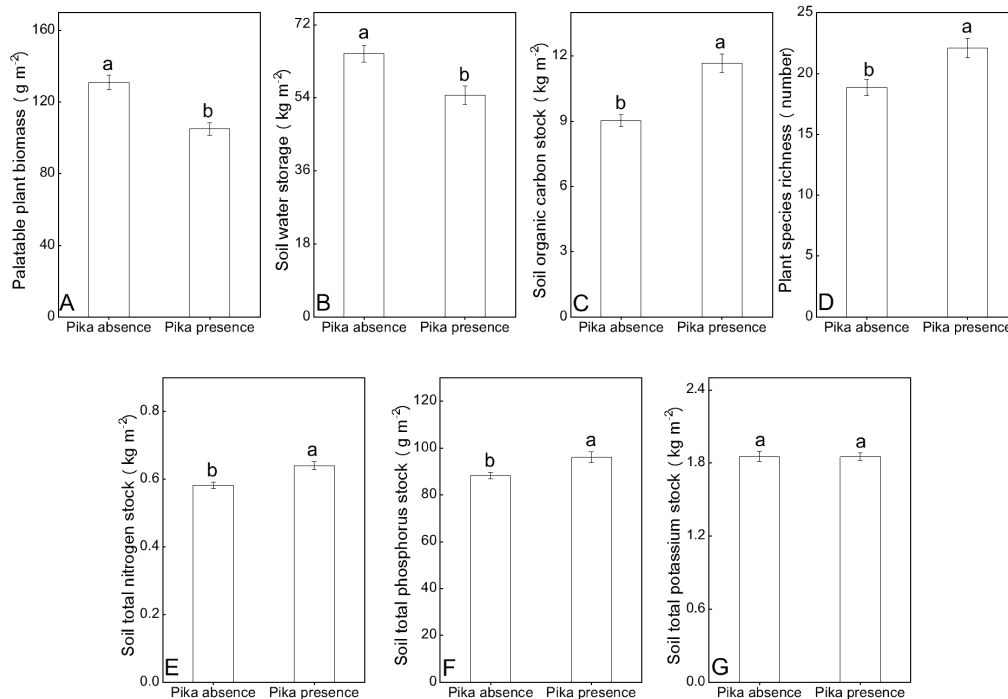
334 To clarify the responses of palatable plant biomass, plant-species richness, soil-water
335 storage, soil organic carbon stock, soil total nitrogen stock, soil total phosphorus stock, and
336 soil total potassium stock to the disturbance caused by plateau pikas, a linear model (LM) was
337 used to examine the relationships between these variables and active burrow entrance
338 densities in all plots with plateau pikas. The densities of active burrow entrances by plateau
339 pikas were considered to be the fixed factor, and were used to construct the regression
340 analysis between palatable plant biomass, plant-species richness, soil-water storage, soil
341 organic carbon stock, soil total nitrogen stock, soil total phosphorus stock, and active burrow
342 entrances densities. To select the final regression models, likelihood ratio tests were used to
343 compare simple linear regression and polynomial regression models. After likelihood ratio
344 tests, the models with $p < 0.05$ and the smaller Akaike Information Criterion (AIC) were used
345 as the final regression models.

346 The Bonferroni's test used to adjust P values and made to correct for experiment-wise
 347 error rates. All statistical analyses were performed with R 4.0.2 (R Foundation for Statistical
 348 Computing, Vienna, Austria).

349 3 Results

350 3.1 Effects of the presence of plateau pikas on the ecosystem services of alpine meadows

351 The palatable plant biomass (Fig. 1A) and soil-water storage (Fig. 1B) were 19.74% and
 352 15.86% lower in the plots with plateau pikas than in the plots without plateau pikas, whereas
 353 soil organic carbon stock (Fig. 1C), plant-species richness (Fig. 1D), soil total nitrogen (Fig.
 354 1E) and total phosphorus stocks (Fig. 1F) in the plots with plateau pikas were 29.15%,
 355 14.58%, 9.97% and 8.89% higher than those in the plots without plateau pikas. In addition,
 356 there was no difference in the soil total potassium stock between the plots with and without
 357 plateau pikas (Fig. 1G).

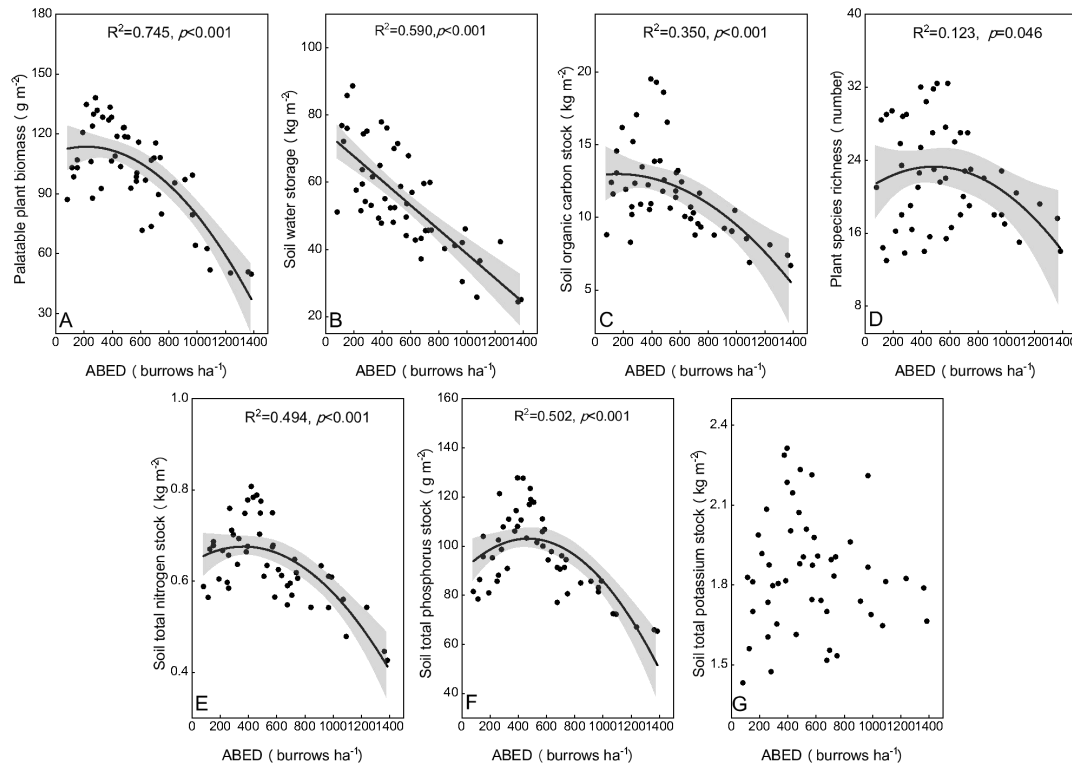


358
 359 **Figure 1.** Palatable plant biomass (A, $F = 46.254$, $p < 0.001$), soil-water storage (B, $F =$

360 35.189, $p < 0.001$), soil organic carbon stock (C, $F = 87.628$, $p < 0.001$), plant-species
361 richness (D, $F = 63.569$, $p = 0.003$), soil total nitrogen stock (E, $F = 22.477$, $p < 0.001$), soil
362 total phosphorus stock (F, $F = 11.724$, $p = 0.004$), and soil total potassium stock (G, $F = 0.026$,
363 $p = 0.88$) of plots with and without plateau pikas (mean \pm standard error). Lower case
364 represents a significant difference between the absence and presence of pika based on an
365 LMM.

366 **3.2 Effects of plateau pikas' disturbance intensity on the ecosystem services of alpine** 367 **meadows**

368 The palatable plant biomass (Fig. 2A), soil organic carbon stock (Fig. 2C), plant-species
369 richness (Fig. 2D), soil total nitrogen (Fig. 2E), and phosphorus (Fig. 2F) stocks significantly
370 increased at first and then decreased gradually as the disturbance intensity of plateau pikas
371 increased. While the soil-water storage of the topsoil layer (Fig. 2B) decreased linearly with
372 increasing disturbance intensity of plateau pikas. In addition, the disturbance intensity of
373 plateau pikas had no obvious relationship with soil total potassium (Fig. 2G).



374

375 **Figure 2.** The palatable plant biomass (A, $F = 68.534$), soil-water storage (B, $F = 69.102$),
 376 soil organic carbon stock (C, $F = 12.642$), plant-species richness (D, $F = 3.292$), soil total
 377 nitrogen stock (E, $F = 22.901$), soil total phosphorus stock (F, $F = 23.652$), soil total
 378 potassium stock (G) for different disturbance intensity of plateau pikas based on linear models
 379 (LMs). An adjusted local smoothed regression line (black) with its 95% confident interval
 380 (gray) was used to determine the relationship between the disturbance intensity of plateau
 381 pikas and the above indicators. ABED: active burrow entrance densities

382 4 Discussion

383 Prairie dogs and European rabbits have been shown to affect grassland ecosystem
 384 services in arid and semi-arid regions (Delibes-Mateos et al., 2011; Martínez-Estévez et al.,
 385 2013). This study combined the home-range scale and a quadrat scales to test how the
 386 presence of plateau pikas and its disturbance intensity influence the ecosystem services of
 387 alpine meadows, and found that the presence of plateau pikas and its disturbance intensity

388 indeed impacts the ecosystem services of alpine meadows, similar to prairie dogs and
389 European rabbits in grassland ecosystem services in arid and semi-arid regions.

390 Lower palatable plant biomass in the presence of plateau pikas indicates that the
391 presence of plateau pikas reduces the forage available to livestock, which is consistent with
392 the results of European rabbits in semi-arid regions (Eldridge and Myers, 2001;
393 Delibes-Mateos et al., 2008), and is not consistent with results from prairie dogs in arid
394 regions (Martínez-Estévez et al., 2013). Prairie dogs benefit perennial plants in arid
395 grasslands, in which blue gramma (*Bouteloua gracilis*) and vine mesquite (*Panicum obtusum*)
396 are palatable perennials for livestock (Sierra-Corona et al., 2015), whereas European rabbits
397 increase unpalatable plants (*Marrubium vulgare* and *Colchicum melitensis*) because they
398 prefer grasses (Leigh et al., 1989; Eldridge and Myers, 2001). Plateau pikas enable more
399 unpalatable broad-leaved plants to grow in alpine meadows (Pang and Guo, 2018) and can
400 bury many plants (Pang and Guo, 2017). However, their consumption patterns can benefit the
401 growth of palatable plants (Pang and Guo, 2017), because plateau pikas preferentially
402 consume unpalatable dicotyledons (Zhao et al., 2013; Pang and Guo, 2017). The tradeoff
403 between the decrease and increase in palatable plant biomass contributes to a negative effect
404 on palatable plant biomass on a home-range scale, resulting in a decrease in the forage
405 available to livestock. These results demonstrate that the presence of small mammalian
406 herbivores affects the forage available to livestock of grassland ecosystems may be related to
407 environmental conditions. Specific performance is that the presence of small mammalian
408 herbivores is disadvantageous to the forage available to livestock in semi-arid and alpine
409 regions, but it is beneficial to forage available to livestock in arid regions.

410 The presence of plateau pikas has different impacts on regulating services of alpine
411 meadows, when assessed by different indicators. The presence of plateau pikas leads to lower
412 soil-water storage, resulting in a decrease in the water conservation, whereas the presence of
413 plateau pikas can lead to higher soil organic carbon stock, implying that the presence of
414 plateau pikas can increase the carbon sequestration. Lower water conservation of alpine
415 meadows in relation to the presence of plateau pikas is consistent with the effect of European
416 rabbits' presence on the water conservation of grasslands in semi-arid regions (Eldridge et al.,
417 2010), whereas it is inconsistent with the presence of prairie dogs in relation to the water
418 conservation in arid regions (Martínez-Estévez et al., 2013). This difference is ascribed to
419 evaluation indicators for the water conservation. The water infiltration rate is considered as an
420 index to evaluate the effect of prairie dogs on the water conservation of grasslands in arid
421 regions (Martínez-Estévez et al., 2013). In contrast, the water storage of topsoil is used to
422 evaluate the effects of European rabbits and plateau pikas on the water conservation of
423 grasslands in semi-arid grassland and alpine meadow (Eldridge et al., 2010). The activities of
424 European rabbits and plateau pikas can reduce the crust cover of grasslands and increase
425 water infiltration from top soil to deep soil in semi-arid regions (Eldridge et al., 2010; Li et al.,
426 2015), contributing to a negative effect on the water conservation in the topsoil layer. This
427 study shows that the presence of plateau pikas leads to higher the carbon sequestration in
428 alpine meadows, similar to the effect of the presence of prairie dogs in arid regions
429 (Martínez-Estévez et al., 2013) and European rabbits in semi-arid regions (Delibes-Mateos et
430 al., 2011). Plateau pikas can input extra organic matter through the deposition of uneaten food
431 (Liu et al., 2009; Zhang et al., 2016; Yu et al., 2017a) and the excretion of urine and feces

432 (James et al., 2009; Yu et al., 2017b), which increases the soil organic carbon stock and
433 contributes to an increase in the carbon sequestration of alpine meadows. These results
434 indicate that the presence of small mammalian herbivores can increase the carbon
435 sequestration of grasslands.

436 Higher plant-species richness in the presence of plateau pikas shows that the presence of
437 plateau pikas can lead to higher biodiversity conservation, similar to the effect of European
438 rabbits in semi-arid regions (Delibes-Mateos et al., 2008) and prairie dogs in arid regions
439 (Davidson et al., 2012). The mechanisms by which small mammalian herbivores lead to
440 higher plant-species richness have been discussed in many previous studies (Zhang et al.,
441 2020; Pang et al., 2021b). The presence of plateau pikas can lead to higher soil total nitrogen
442 and total phosphorus stocks, demonstrating that plateau pikas can increase the soil nitrogen
443 and phosphorus maintenance. In addition, there was no difference in the soil total potassium
444 stock between the areas with and without plateau pikas, indicating that the presence of plateau
445 pikas had no effect on the soil potassium maintenance. This effect was also observed with
446 prairie dogs and European rabbits in arid (Delibes-Mateos et al., 2011) and semi-arid regions
447 (Delibes-Mateos et al., 2008; Willott, 2001). Some of the following factors explain the higher
448 soil nitrogen and phosphorus stocks caused by plateau pikas. The presence of plateau pikas
449 can increase the input of soil organic material (Liu et al., 2013; Zhang et al., 2016; Pang et al.,
450 2020a). Secondly, the presence of plateau pikas can result in higher organic nitrogen and
451 phosphorus stocks (Yu et al., 2017b), which contributes to higher soil nitrogen and
452 phosphorus maintenance. These results suggest that a general pattern can be identified
453 regarding the effect of the presence of small mammalian herbivores on the supporting

454 services of biodiversity conservation, soil nitrogen, and phosphorus maintenance.

455 This study also shows that the disturbance intensity of plateau pikas also affects the
456 forage available to livestock, biodiversity conservation, water conservation, carbon
457 sequestration, and soil total nitrogen and phosphorus maintenance. As found in plant-species
458 richness and aboveground plant productivity (Dial and Roughgarden, 1998; Gao and Carmel,
459 2020), the response of plant-species richness and palatable plant biomass to the disturbance
460 intensity of plateau pikas follow the pattern for the intermediate disturbance hypothesis in this
461 study. In addition, the soil organic carbon stock, soil total nitrogen and phosphorus stocks at
462 home-range scale also support the intermediate disturbance hypothesis. However, the top soil
463 water storage does not conform the intermediate disturbance hypothesis.

464 At lower disturbance intensity, stronger competition of dominant sedges often restrains
465 the grass to grow well (Pang and Guo, 2018) and the rare plants to coexist (Wang et al., 2012),
466 which leads the forage available to livestock and biodiversity conservation of alpine meadows
467 to be maintained at a low level. Although the presence of plateau pikas can increase the input
468 of soil organic matter, this increase is low (Pang and Guo, 2017; Pang et al., 2020b), which
469 enables the soil organic carbon sequestration and soil nitrogen and phosphorus maintenance
470 of alpine meadows to maintain a relatively low level.

471 At intermediate disturbance intensity, the activities of plateau pikas improve the growth
472 potential of grass plants (Wang et al., 2012), and increase the input of organic matter, soil total
473 nitrogen (Li et al., 2014), organic carbon accumulation (Yu et al., 2017b), which contributes
474 to higher the biodiversity conservation, forage available to livestock, carbon sequestration,
475 soil total nitrogen and phosphorus maintenance services.

476 **At higher disturbance intensity of plateau pikas, frequent bioturbation can enable all**
477 **species to be at risk of going extinct (Dial and Roughgarden, 1998).** Low soil water content in
478 alpine meadows (Liu et al., 2013) only sustains the xerophytes and mesophytes, most of
479 which are unpalatable (Pang and Guo, 2018). This contributes to relatively lower forage
480 available to livestock and biodiversity conservation. Low vegetation biomass decreases the
481 input resources of soil organic matter (Sun et al., 2015; Pang and Guo, 2017), contributing to
482 a decrease in the soil organic carbon sequestration and soil nitrogen and phosphorus
483 maintenance of alpine meadows.

484 Additionally, the linearly negative relationship between the water conservation of alpine
485 meadow and disturbance intensity of plateau pikas is ascribed to evaporation and more water
486 infiltration on bare soil patches, as the amount of water evaporation and infiltration tends to
487 increase as the area of bare soil increases (Liu et al., 2013).

488 Together with previous studies (Delibes-Mateos et al., 2011; Martínez-Estévez et al.,
489 2013), this study demonstrates that the presence of small mammalian herbivores has similar
490 impacts on the biodiversity conservation, soil nutrient maintenance, and carbon sequestration
491 of grasslands throughout the arid, semi-arid, and alpine regions, whereas the effects of the
492 presence of small mammalian herbivores on the forage available to livestock and water
493 conservation are dependent on environmental conditions. This study further verifies that the
494 disturbance intensity of plateau pikas also has a significant impact on the ecosystem services
495 of alpine ecosystems. These results concur with the findings in research fields of small
496 mammalian herbivores in relation to grassland ecosystem services.

497 **5 Conclusions**

498 This study focused on plateau pikas to investigate the responses of forage available to
499 livestock, water conservation, carbon sequestration, soil nutrient maintenance, and
500 biodiversity conservation of meadow ecosystems to the presence of a small mammalian
501 herbivore and its disturbance intensity across five sites. This will provide insight into the
502 relationship between small mammalian herbivores and ecosystem services of grasslands. The
503 results of this study showed that the presence of plateau pikas led to higher biodiversity
504 conservation, soil nitrogen and phosphorus maintenance, and carbon sequestration of alpine
505 meadows, whereas it led to lower forage available to livestock and water conservation of
506 alpine meadows. Furthermore, this study found that the effect of plateau pikas disturbance
507 intensity on the forage available to livestock, biodiversity conservation, soil maintenance of
508 nitrogen and phosphorus, and carbon sequestration also conformed to the moderate
509 disturbance hypothesis. These results verified that plateau pikas could affect the ecosystem
510 services of meadow ecosystems in alpine regions and present a relatively complete pattern of
511 small mammalian herbivores influencing grassland ecosystem services.
512

513 *Author contributions.* YC and ZG conceived the ideas and designed the methodology; YC, XP,
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516

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