

## Author's response

Dear editor,

Please find hereafter our point by point reply to the reviewers' comments, including the changes brought to the manuscript mentioned by their line numbers in the manuscript pdf file.

### Response to reviewer #1

First of all, we would like to thank the anonymous reviewer for their kind attention to our article and the time that they allocated to read it and provided us with comments and suggestions for enhancing our paper. Hereafter are our answers to the 4 comments raised by the reviewer.

#### Comment #1:

It is not clear whether the sensitivity results will hold if there will be source-sink terms for organic matter.

#### Response #1:

Each of the three sensitivity analyses are conducted with varying total organic matter (TOC) sources between 1 to 10 mgC/L (Fig. 7-8). Therefore, the source term is indeed considered and adding a new source will not change the results.

As for organic matter sink, since we had designed the study under a situation where the system shall not be depleted of organic matter and dissolved oxygen at any moment, as it is our goal to see how the parameters behave in their presence, this study cannot stand under a sink term that would lead to depletion, especially of organic matter. Indeed, in a depleted environment, the influence of the organic parameters cannot be studied.

To summarize, the fact that we conducted many numerical experiments under various organic matter conditions already answers the reviewer's concern. The answer to the effect of a poor or rich environment in organic matter is already largely discussed throughout the paper.

[This limitation and other upcoming restrictions of this study are addressed in a separate subsection in the revised manuscript under Discussion part, section 4.2, lines 470-474.](#)

#### Comment #2:

No attention is paid to the radiation effects of the bacteria population that is most pronounced at low flows

#### Response #2:

We would like to thank you for bringing up this point. Indeed, our model currently lacks this process and it is something which needs to be incorporated in the model source code. However, its implementation and reanalysis of the work will take far more time than the review period of this article. Moreover experimental data on the subject needs to be found in order to conduct a sensitivity analysis. This specific question may be the subject of further research. [We discussed this limitation of our approach in lines 475-480 of the manuscript.](#)

#### Comment #3:

The role of the hyporheic exchange in bacteria population dynamics at low flows is ignored and can be substantial.

Response #3:

The contribution of groundwater to downstream rivers is well known to be negligible with respect to the discharge of those rivers (Strahler order > 6). Our case study mimics such rivers, and especially the Seine river crossing the Paris urban area. For such a system, Pryet et al. (2015) provide estimates of aquifer contribution to the Seine River. More specifically, the hyporheic exchange rate is very limited in this area, with a maximum value of  $0.005 \text{ m}^3 \cdot \text{s}^{-1} \cdot \text{km}^{-1}$ , which corresponds to a maximum of  $1 \text{ m}^3 \cdot \text{s}^{-1}$  over a 200 km stretch of downstream river. With respect to the actual discharge of the Seine river of  $80 \text{ m}^3 \cdot \text{s}^{-1}$ , this process is not relevant to account for in our study. However the reviewer is correct that if studying headwater streams functioning, the hyporheic exchanges would be of primary importance. [The negligence of hyporheic exchanges is mentioned in the initial conditions section lines 250-252.](#)

Comment #4:

The introduction of constant parameters to simulate the repartitioning is a gross simplification. Monitoring data show that the ratios vary.

Response #4:

Without any doubt, the organic matter partitioning parameters vary with time. However, from a sensitivity analysis point of view, we need to keep them constant during the simulation period so that the influence of their increment could be studied on the model total variance and then ranked using the Sobol criteria. That is why we have 360,000 ( $N(2D+2)$ ) parameter combinations where each parameter value is changed within its variation range given in Table 1 but unchanged during the simulation period to study their impact on model output. Fortunately, the concept of time varying parameters has been addressed in our data assimilation study (<https://doi.org/10.1016/j.envsoft.2022.105382>) which is the second step after the sensitivity analysis as it has been explained in discussion subsection “Consequences of the results on data assimilation (DA) strategy”. In that work, we had estimated the time evolution of parameters thanks to observation data and a new work is under way to test an automatic detection of boundary conditions organic matter content using data assimilation. To our knowledge this would be a major step forward for the modeling of the water quality of downstream river systems.

## **Response to reviewer #2**

First of all, I would like to thank the reviewer for their positive comments on our paper and valuable propositions for improvement. Here are our subsequent responses, clarifications and the modifications that we added in the manuscript:

*In response to their comment:*

*“What is the novelty this manuscript brings to the field?”*

**[R1]** Our article provides the following novelties compared to the previous works and [are summarized in the introduction part of the manuscript in lines 65-69.](#):

1. In the previous works, Wang et al., (2018) have focused on both high & low flow under different conditions of bloom, non-bloom and post-bloom, whereas here we have explored further the summer post-bloom low flow condition where significant discrepancies are observed between water quality model results and dissolved oxygen (DO) observations (Wang et al., 2022). We already knew that bacterial physiological parameters control DO evolution during low flow and that was already quantified in Wang et al., (2018). However, what we didn't know is the extent to which the characteristics of the organic matter (OM), particularly its dissolved biodegradable fraction, control the oxygen dynamics at low water levels, and whether these characteristics are important with respect to the physiological properties of heterotrophic bacteria;
2. To explore this question through a sensitivity analysis, we had to explicitly add the parameters of OM model inside the software CRIVE itself, and especially in the boundary conditions of the model, where CRIVE used to read the six pools of OM (DOM1, DOM2, DOM3, POM1, POM2, POM3) as time varying state variables defined by the user. We therefore added an OM repartition model inside CRIVE so that the DOM<sub>i</sub> and POM<sub>i</sub> state variables are now calculated by the model based on only one value provided by the user, i.e. the total organic carbon (TOC), and 5 new model parameters ( $t$ ,  $b_1$ ,  $b_2$ ,  $s_1$ ,  $s_2$ ). This repartition model not only distributes TOC among the six CRIVE pools using the 5 parameters whose variation ranges were found using a bibliography review (Table 1) , but it also gives us the possibility to do a sensitivity analysis for evaluating their role and influence on DO variation. The main difference is that instead of reading the 6 pools (not varying) directly, now it reads TOC and uses the 5 parameters to convert it into the 6 pools (now varying because the 5 parameters have a range).
3. Compared to the previous work (Wang et al., 2018) that has studied only the direct impact of each parameter, we went further and looked into intra-parameter interactions (higher order Sobol indices) and we found that certain parameters hide the influence of other parameters due to their interactions. Thanks to that, we designed the 2nd and 3rd Sobol' sensitivity analysis which allowed us to better quantify how the share of OM influences DO in river systems with regards to the physiological parameters of heterotrophic bacteria. We believe that this methodology may also be of interest for future sensitivity analysis where parameter interactions may hide the effect of other parameters.
4. In the previous study, the sensitivity analysis was conducted under a constant OM load of TOC =3.2 mgC/L whereas in this work, we evaluated the evolution of sensitivity indices for various TOC loadings ranging from 1 up to 10 mgC/L which represents the OM load from river, treatment plants and combined sewage overflows (Fig. 7b and Fig 8).
5. Conducting long-term sensitivity analysis: In the previous work, the influence of model parameters are usually studied over a short period of time, for instance a 4 day period in Wang et al. (2018). Here we looked deeper inside the system dynamics by extending the period up to 45 days. Such strategy led to a better understanding of the mid-term effect of slowly biodegradable OM. Even though those effects appear rather negligible, this result is important for improving and simplifying water quality models.

The reviewer advises to make the findings clearer in the paper. We agree with this comment and propose to add a first section to the discussion to wrap up those findings as stated here after.

*In response to their comment:*

*“The importance of heterotrophic bacteria activity and properties of the dissolved organic matter pool are pinpointed as important parameters to explain uncertainties of water quality models in the introduction (lines 49 to 60). Then, what is this manuscript offering new (or different) from previous studies? “*

**[R2]** The reviewer was misled by a flawed formulation of this section of the introduction. We rephrased it following those clarifications in lines 50-54. Formerly, Wang et al., (2022) simply assumed from their study that the OM degradation and OM repartition are playing a role in the model discrepancies during low flow, without explicitly quantifying their relative influences. In our paper we tested those hypotheses and therefore extended the parameters of interest to include 3 parameters representing OM kinetics and 5 representing OM repartition to quantify the sensitivity of DO variation with respect to those with a Sobol sensitivity analysis. We found that  $b_1$ , the share of BDOM, has a significant influence on DO variations in certain circumstances, such as the presence of fast growing heterotrophic bacteria. In that case, a low  $b_1$  value may lead to a depletion of BDOM by heterotrophic bacteria, while high  $b_1$  allows the micro-organism to grow without limit, leading to significant oxygen depletions.

*In response to their comment:*

*“Please, specify better what are the research questions or objectives of this work?”*

**[R3]** We added the following main research questions in the introduction of the article in lines 70-76:

- What are the influential parameters controlling DO during a post-bloom summer low flow period where discrepancies are observed in different water quality models? Is a model that includes bacteria physiological parameters only sufficient to describe DO variation ?
- To what extent is the knowledge of the quantity of OM share, especially that of BDOM influential for water quality modeling?
- What is the hierarchy among the influential parameters ?

*In response to their proposal on the discussion part:*

**[R4]** We totally overhauled and restructured the discussion section by first answering the research questions with the current section 3.4. This new section is called “Hierarchy of the most influential parameters during low flow period”.

Then, we discussed the limitations and assumptions of this study as the second subsection as indicated in the response to the comments of the first reviewer. We also provided recommendations for future studies in order to incorporate these limitations lines.

Then, we restructured the sub section “Consequences of the results on water quality monitoring in urban areas” by reformulating how important are our results in the context of water quality

monitoring and what information or experimental data is required to be supplied to the water quality models in order to provide better estimates of the river water quality. Indeed, we will first show how important it is to have better identification of bacterial parameters in any water quality monitoring network and second what we can do to get more information on b1 or BDOM. As recommended by the reviewer, we removed the parts related to the design of monitoring stations.

The last subsection of the discussion is dedicated to data assimilation where more clarifications are made for technical terms such as data assimilation and particle filter and also suggestions on how to conduct the data assimilation based on our results

*In response to their second question regarding the incorporation of new parameters:*

**[R5]** We restructured the material and method section 2.2.2 lines 195-220 to display how we have incorporated the organic matter repartition model consisting of 5 new parameters inside CRIVE instead of using the 6 forced user inputs that are not model parameters.

*In response to their specific questions on the addition of new parameter like:*

*What do authors mean with new parameters? New regarding what exactly? C-RIVE?*

**[R6]** Here, we have two types of new parameters. First, OM degradation kinetic parameters that already exist in CRIVE but whose influence was not studied in any other research. Secondly, OM repartitioning parameters (section 2.2.2) that as I have explained in the point #2 of R1. This is a novelty that did not exist in CRIVE before. Indeed, CRIVE used to read the share of each one of the 6 OM pools directly as an input (that was not variable, they were created in a form of database by multiplying TOC with certain assumed values), however, what we did as a novelty was that we gave CRIVE the possibility to read directly TOC (which comes from experimental data) and convert it into the above 6 OM pools using 5 parameters for which we did an extensive bibliography to find their variation range (t,b1,b2,s1,s2). Thereby, we created these 5 new parameters whose influence on DO could be studied and thanks to which we can now have varying 6 OM pools. This is something which was not possible before. [This is clarified in lines 216-222.](#)

*In response to the question: Another change is that authors pooled DOM1 and DOM2 fractions to create a new fraction called BDOM. Am I missing something? What is really new/different in this approach regarding to previous work in C-RIVE?*

**[R7]** If we look at equations 12 & 13, the variation range of BDOM is found using the variation range of b1, therefore it is not pooled by addition of DOM1 and DOM2. On the contrary, DOM1 and DOM2 are now derived from BDOM using the parameter s1. But why did we do this? In the first sensitivity analysis, we found b1 as an influential parameter, however for the second and third Sobol and in order to decrease the computation cost, we used BDOM as a parameter to get rid of the 5 initial parameters (as shown in Table 4, we went from 17 parameters to 12 parameters). BDOM is the equivalent of b1 ( $b1 = BDOM/DOM$ ). DOM is constant ( $DOM = t \times TOC$ ) because t was found to be non-influential in the first experiment and fixed here, therefore, having b1 or BDOM does not make any difference technically.

**[R8]** *Finally, regarding the use of repetitive and introductory paragraphs*, we think that it is a good habit to brief the reader regarding what they are going to expect in different sections of an article. This will give them the chance to fast access to their intended sections or subsections. Therefore, we believe that the short briefing of the article at the end of introduction or a fast description of the method section in its first paragraph is a good habit in modeling articles which helps the reader better understand how the different elements of the method section are interlinked. However, we will find the annoying sentences and will remove or rephrase them.

As required by the editor, clear definitions and clarifications were added for technical terms such as data assimilation and particle filter in order to facilitate a wider audience in lines 506-509 and 510-511, respectively. Finally, necessary modifications were made to the figures in order to make them compatible with the color blindness requirements.

#### References:

Pryet, A., Labarthe, B., Saleh, F., Akopian, M., and Flipo, N. (2015) Reporting of stream-aquifer flow distribution at the regional scale with a distributed process-based model, *Water Resour. Manage.*, 29, 139-159. doi: 10.1007/s11269-014-0832-7

Wang, S., Flipo, N., and Romary, T.: Time-Dependent Global Sensitivity Analysis of the C-RIVE Biogeochemical Model in Contrasted Hydrological and Trophic Contexts, *Water Research*, 144, 341–355, <https://doi.org/10.1016/j.watres.2018.07.033>, 2018.

Wang, S., Flipo, N., and Romary, T.: Oxygen Data Assimilation for Estimating Micro-Organism Communities' Parameters in River Systems, *Water Research*, 165, 115–121, <https://doi.org/10.1016/j.watres.2019.115021>, 2019.

Wang, S., Flipo, N., Romary, T., and Hasanyar, M.: Particle Filter for High Frequency Oxygen Data Assimilation in River Systems, *Environmental Modelling & Software*, 151, 105382, 2022 <https://doi.org/10.1016/j.envsoft.2022.105382>