

# Spruce bark beetles (*Ips typographus*) cause up to 700 times higher bark BVOC emission rates compareds to healthy from Norway spruce (*Picea abies*)

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## 10 Abstract

Emissions of biogenic volatile organic compound (BVOC) from the bark of Norway spruce (*Picea abies*) trees can be affected by stress, such as infestation of spruce bark beetles (*Ips typographus*). ~~We studied the~~The difference in emission rates ~~between from~~ healthy ~~spruce bark~~ and infested ~~Norway~~ spruce bark ~~was studied, as well as~~ the influence of time since spruce bark beetle infestation started and the difference in emission rates from bark beetle drilled ~~entry holes and exit holes~~entry and exit holes. Bark chamber measurements on both ~~healthy trees and infested trees~~healthy and infested trees were performed during the summer of 2019 ~~at Hyltemossa and Norunda research station~~at two sites in Sweden. To consider the seasonal pattern of the spruce bark beetle, ~~we divided~~the emission rates from infested trees ~~were divided~~ into two seasons, an early season dominated by entry holes and a late season with mainly exit holes. ~~The~~Our findings results ~~showed~~ a significant difference in emission rates from 20 healthy and infested trees, independent of season. The seasonal average standardized BVOC emission rates ~~off from~~ healthy trees was  $321.89 \pm 521.67 \mu\text{g m}^{-2} \text{h}^{-1}$  (mean  $\pm$  standard deviation), while the average standardized BVOC emission rates ~~off from~~ infested trees were  $6700385 \mu\text{g m}^{-2} \text{h}^{-1}$  and  $2000402 \mu\text{g m}^{-2} \text{h}^{-1}$  during early and late season respectively. ~~A~~We also found an exponentially decreasing relationship ~~was found with~~ BVOC emission rates ~~decreased exponentially with time~~and time since infestation started ~~where the emission rates and indicated induced~~ 25 ~~emission rates for about one year after which the emission rates were similar to those from~~ reached the same level ~~as the constitutive BVOC emission rates from healthy~~ bark ~~after around one year~~. When comparing bark monoterpene BVOC emission rates with emission rates from needles, ~~we found that the constitutive~~constitutive needle emission rates ~~were found to be were~~ 11 times higher than the ~~healthy constitutive~~ bark emissions. However, the emission rates from infested Norway spruce tree bark were instead 6 to 20 times higher than the constitutive 30 needle emissions, causing substantial increases in the total tree BVOC emission rate ~~(550 % to 1900 % increase)~~. ~~This study adds evidence that spruce bark beetle induced bark BVOC emissions are higher than previously thought and highlights the need for further research with more samples more frequently throughout the season to fully understand the impact, which is required to quantify spruce bark beetle infestations impacts on the atmospheric chemistry and climate change.~~

## 35 1 Introduction

In Europe, forest damage caused by outbreaks of the European spruce bark beetle (*Ips typographus*) is the third largest disturbance after storm-felling and forest fires (Jönsson et al., 2012; Schelhaas et al., 2003). In Sweden, the drought in the summer of 2018 led to increased bark beetle outbreaks which, in 2020, were estimated to affect

about 8 million m<sup>3</sup> (standing timber volume) Norway spruce (*Picea abies*) forest (Wulff and Roberge, 2020). This is the largest stock of forest volume killed by spruce bark beetles recorded in a single year in Sweden; in the period of 1990-2010, around 150,000 m<sup>3</sup> forest in southern Sweden was damaged on average per year (Wulff and Roberge, 2020). Climate change amplifies the risk of bark beetle outbreaks as the elevated risk of storm felling and drought favor the bark beetles with easier access to weakened trees (Jönsson et al., 2012). Higher temperatures and a longer growing season can also lead to an additional generation of spruce bark beetles per year (Jakoby et al., 2019; Jönsson et al., 2012). A larger bark beetle population, triggered by weather extremes, is associated with an increased risk of attacks on healthy spruce trees, with outbreaks leading to extensive damage to the forests (Jakoby et al., 2019; Seidl et al., 2014).

Biogenic volatile organic compounds (BVOCs) emitted from trees function for example as a defense system against heat and oxidative stress (Loreto and Schnitzler, 2010). They are highly volatile and chemically reactive and can react directly with oxidizing species or act as membrane stabilizers (Brilli et al., 2009; Kleist et al., 2012; Sharkey et al., 2001). The efficiency of BVOCs to form oxidation products depends on the specific BVOC's molecular structure (Bonn and Moortgat, 2002; Roldin et al., 2019; Thomsen et al., 2021). As BVOCs are emitted, they can enhance chemical reactions which in turn can lead to increased tropospheric ozone concentration, or BVOCs can get oxidized and may foster the formation of secondary organic aerosol (SOA; Kulmala et al., 2003). Stress-induced BVOC emissions alter the oxidation capacity as some BVOC species are more efficient to act as secondary organic aerosol (SOA) SOA precursors and foster particle growth (Roldin et al., 2019; Thomsen et al., 2021). Boreal forests experiencing abiotic or biotic stress due to large-scale forest disturbances might thus increase the production of BVOC species highly efficient as precursors of SOA to the atmosphere. As BVOCs are emitted, they can enhance chemical reactions that in turn can lead to increased tropospheric ozone concentration or be oxidized and foster the formation of SOA (Kulmala et al., 2003). This results in high uncertainties regarding the contribution to either a negative feedback loop (cloud formation and radiation scattering (Paasonen et al., 2013)) or a positive feedback loop (increased tropospheric ozone) for the climate (Arneth et al., 2010; Jia et al., 2019). The impact of aerosol formation including the BVOC-SOA feedback still remains the largest uncertainty in our understanding of their radiative forcing (IPCC, 2014; Jia et al., 2019). Emission of BVOCs due to plant stress further increase the uncertainties of their impact in a changing climate.

Increased BVOC concentrations in plant tissue can also fight off predators (Laothawornkitkul et al., 2009; Li et al., 2019; Rieksta et al., 2020), and conifer trees use BVOCs as a defense mechanism against spruce bark beetles (Celedon and Bohlmann, 2019; Krokene, 2015; Raffa and Berryman, 1982). The parental bark beetles attack spruce Norway spruce trees by drilling entry holes into the bark and form eggs galleries in the phloem. After about eight weeks, the new generation starts to leave the tree by boring/drilling exit holes (Öhrn et al., 2014). To prevent a successful attack, the spruce Norway spruce increase the resin flow which submerges the parental bark beetles and egg galleries, potentially killing beetles or pushing them out of the entry hole (Raffa, 1991). The resin serves as a storage pool containing BVOCs that volatilize when the resin is flowing out of the tree, making the resin harden and close the wound. BVOCs are emitted constitutively from the trunk of the spruce Norway spruce, but when the emissions are induced as a stress-response, they have been shown to be toxic to spruce bark beetles (Celedon and Bohlmann, 2019; Krokene, 2015), especially certain compounds like myrcene and α-terpinene (Everaerts et al., 1988). Studies on conifers attacked by bark beetles found evidence of increased monoterpene (MT) content at the attacked location (Amin et al., 2013; Ghimire et al., 2016; Zhao et al., 2011b). Qand occurrence of the oxygenated MT eucalyptol has been found to indicate induced defense and higher survival rates from

Norway spruce attacked by spruce bark beetles (Schiebe et al., 2012). ~~When c~~Comparing BVOC emission sources from different parts on conifer trees, trunk emissions are suggested to potentially contribute a lot more to the whole tree emissions than previously thought, even when not attacked by bark beetles (Greenberg et al., 2012).

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~~The defense mechanism of Norway spruce is poorly understood~~There is still a lot to learn about the defense mechanism of Norway spruce, ~~only aas~~ few studies have analyzed the induced BVOC emission from the trunk following an attack of the European spruce bark beetle (Ghimire et al., 2016; Zhao et al., 2011b). The aim of this study was to ~~(i) investigate~~study the BVOC emissions from Norway spruce trunks and the impact of spruce bark beetles and to ~~(ii) study~~investigate the ~~relation connections between~~of BVOC emission rates, number of bark beetle drilled holes and time. The aim was also to ~~(iii) put our study in a broader perspective by compare and connect~~connect ing the spruce bark beetle induced BVOC emission rates to needle emissions and ~~other stresses like~~ heat stress. Based on previous findings, ~~we formulated~~three hypotheses were formulated: (H1)-~~Our first hypothesis was that~~ infested trees have higher bark emission rates than healthy trees, and ~~that~~ infestation changes the emission blend, (H2)-~~Our second hypothesis was that~~ BVOC emission rates ~~are would be~~ highest at the start of infestation, and decrease over time in response to declining tree vitality and eventual death of the tree ~~and~~ (H3)-~~We also wanted to investigate the connections of emission rates and number and type of bark beetle holes, where our third hypothesis was that~~ the type of the bark beetle drilled holes ~~matter for~~influence the BVOC emission rates where there is a relationship between the number of entry holes and emission rates rather than ~~a high~~the total amount ~~number~~ of holes. ~~The~~We reasoning behind H3 ~~wased~~ that after a successful infestation, the number of holes increase after some weeks when the new generation leave the tree through exit holes, but the emission rates would decrease as the tree is dying due to bark damage and ~~blue stain fungi~~ preventing transport of water and nutrients. In a sub-study, ~~two tree~~tree individuals with different initial health status were selected and followed throughout the growing season by repeating ~~the~~measurements during ~~at~~the successful attack and infestation of bark beetles. ~~From~~The aim (iv) of this sub-study ~~we were hoping~~was to see if the difference in initial health status of the trees would result in different emission rates and emission blends, and ~~to~~also analyze how the individual emission blend changed over time after a successful infestation.

## 2 Methods

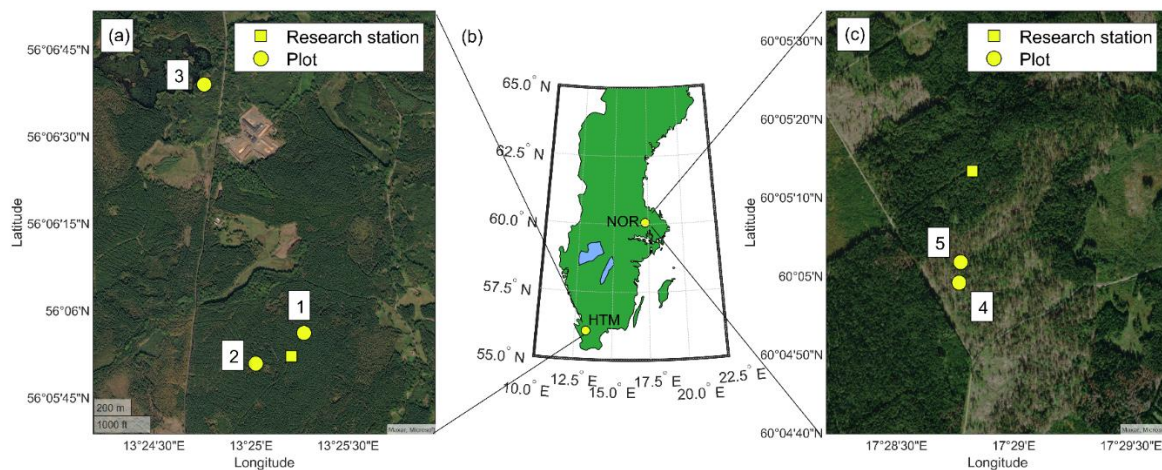
### 2.1 Site description

110 ~~Six~~Five measurement campaigns were carried out from May to August 2019 (Table 1), where five campaigns were located at the ICOS (Integrated Carbon Observation System, ICOS-Sweden.se) research station ~~in~~ Hyltemossa (HTM, 56°06'N, 13°25'E; Fig. 1b) and one ~~additional~~ campaign at the ICOS research station in Norunda (NOR, 60°05'N, 17°29'E; Fig. 1b). The forest in HTM is dominated (>97% of the species composition) by Norway spruce (*Picea abies*) with a small fraction (<3%) of Scots pine (*Pinus sylvestris*) and deciduous trees. The understory vegetation is sparse, containing mostly mosses (Heliasz et al., n.d.). The forest in NOR is dominated by Norway spruce (54%) and Scots pine (37%) with a small fraction (9%) of deciduous trees and an understory vegetation with shrubs of mostly blueberries, cranberries, mosses and flowers (Mölder et al., n.d.). Both facilities are located inside managed forests, but the age and height of the trees differ. In HTM the trees are around 40 years old with an average height of 19 m ~~in 2019~~ and NOR has a forest stand of mixed ages around 60-80 years and up to 110 years with a height of around 25 m for the dominating trees in 2019 (Heliasz et al., 2021; Mölder et al., 2021).

**Table 1. Time table of the six campaigns conducted during 2019 at the sites Hyltemossa (HTM) and Norunda (NOR). Indicated is also the number of collected samples during each campaign at each plot and the status of the Norway spruce when measurements were collected (healthy or infested).**

Month	Date	Site	Number of collected samples										
			Plot 1		Plot 2		Plot 3		Plot 4		Plot 5		Total
			Healthy	Infested	Healthy	Infested	Healthy	Infested	Healthy	Infested	Healthy	Infested	
May	4th-6th	HTM	12	3	12	-	12	-					39
June	4th-6th & 13th	HTM	9	1	12	-	6	6					34
July	2nd-4th	HTM	12	-	12	-	6	6					36
July-August	30th-1st	HTM	12	-	12	-	6	6					36
August	21st-22nd	NOR								9		9	18
August	26th-27th	HTM	12	-	12	-							24

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130 **Figure 1: The location of the study sites in Sweden (b), with Hyltemossa (HTM; a) displayed to the left, the location in Sweden in the middle (b) and Norunda (NOR; c) displayed to the right.** Measurement plots (yellow circle) at HTM (1-3) and NOR (4-5) are shown in the site-specific maps and their location relative to the ICOS station (yellow square). Healthy Norway spruce trees were measured in plot 1-3 and infested spruce trees were measured in plot 1, 3, 4 & 5. The sub-study was conducted in plot 3. The figure is created in MATLAB and Mapping Toolbox release 2021a  
 135 (The MathWorks, Inc., Natick, MA, USA).

Three plots in HTM and two plots in NOR were selected for the study (Fig. 1a,c), and two plots in NOR (Fig. 1e). Two plots in HTM were located inside the Norway spruce plantation used by ICOS, while the third plot was located around 1.6 km north of the ICOS station in an older (about 100 years) forest stand. In NOR the locations were chosen based on availability of bark beetle-affected trees inside the forest plantation. Four Norway spruce trees were selected at each plot in HTM, and three trees at each plot in NOR. A total of 18 trees were measured, whereof 12 were measured repeatedly during the growing season in HTM. For all plots in HTM healthy trees were selected where the tree health was determined by visual examination in close contact with the forest manager employed by the forest owner, at Gustafsborgs Säteri AB, in May 2019. Trees potentially that could have been stressed by forestry machinery or pests were not selected for the study. The infested trees were selected based on the signs of spruce bark beetle infestation. In addition, two trees seemingly stressed from signs of late bark beetle infestation from the previous year, in 2018, were found on two trees, were chosen, one at plot 1 only measured in the early season and one at plot 3 which later got infested again and used for the comparison of healthy

~~and stressed trees when infested. These trees were selected for the study to analyze long-term infestation effects.~~

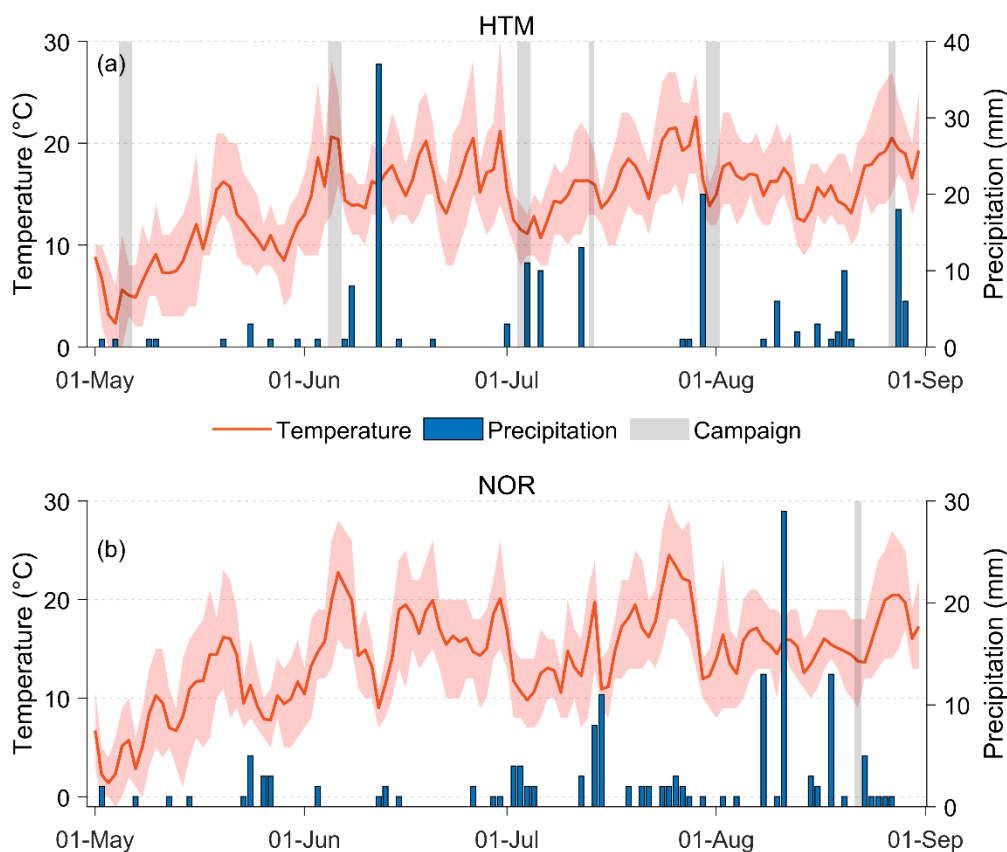
150 ~~Only infested trees were selected in NOR. As an active spruce bark beetle outbreaks was~~ occurring in NOR during 2019 ~~and only infested trees were selected at that site.~~

~~The sub-study was conducted at plot 3 in HTM where two trees were selected, one already stressed from previous infestation in the previous year and one healthy (Fig. 1a).~~

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~~To ensure bark beetle infestation. To enable the measurements of healthy trees which later got infested, a The healthy tree at plot 3 in HTM (Fig. 1a) was baited using a bark beetle slit trap with pheromones to facilitate bark beetle infestation. The trap was installed between the 1<sup>st</sup> and 2<sup>nd</sup> campaigns in May and June (Table 1). One bag of biological attractant was used containing tThe pheromone\_s used in the trap were a combination of 2,3,2~~  
160 ~~Methylbutenol, cis-Verbenol and Ipsdienol (Typosan P306, Plantskydd AB, Ljungbyhed, Sweden IPS), and one standard bag was inserted in the slit trap. Two trees were successfully infested and measured repeatedly to track the emission pattern after infestation. Both of the trees were successfully infested and measurements were repeated throughout the season to see the effects of infestation over time.~~

165 The weather during the measurement periods varied from cold and humid to warm and dry conditions. The average temperature during the growing season (May-August) was 14.6 °C (± 4.6 °C) in HTM and 14.2 °C (± 4.2 °C) in NOR with the sum of the precipitation over the growing season being 168 mm in HTM and 151 mm in NOR (Fig. 2). The daily average temperatures during the measurement periods ranged from 5 °C to 22 °C for both sites, with a daily total rainfall up to 3 mm (Fig. 2). In HTM, it was ~~coldest (1 °C to 11 °C) during the measurement~~  
170 ~~campaign in May and~~ warmest (12 °C to -28 °C) during the measurement campaign 2 in June ~~and coldest (-11 °C) during the 1<sup>st</sup> campaign in May.~~



175 **Figure 2: The ~~total daily precipitation (blue bars) and daily average temperature (red line) with daily minimum and maximum temperature (red shade) and total daily precipitation (blue bars) with daily minimum and maximum temperature (red shade)~~ for the study sites in (a) Hyltemossa and (b) Norunda. The times for the campaigns are marked in grey. ~~The weather data was acquired from the Data: ICOS research stations at the study sites~~ (Heliasz, 2020; Mölder, 2021).**

## 2.2 Experimental design

180 The bark emissions from the trees were measured using a tree trunk chamber connected with PTFE tubing (Teflon, Swagelok, Solon, OH, USA) to a pump box system consisting of a diaphragm pump (1420 VPD, Gardner Denver Thomas GmbH, Memmingen, Germany) and a flow meter (GFM mass flow meter, Aalborg Instruments & Controls Inc., USA). The pump box system was used to provide purge air with a flowrate of 0.7 lpm (liters per minute) to the trunk chamber with a volume of 0.6-0.9 Litres (Fig. 3). Adsorbent tubes were used to take air

185 samples from the chamber. A hydrocarbon-BVOC filtertrap (Hydrocarbon trap, Alltech, Associates Inc., USA) containing activated carbon and MnO<sub>2</sub>-coated copper nets was mounted between the pump box and chamber to scrub the purge air of remove VOCs, BVOCs and O<sub>3</sub> to ensure that only clean air entered the chamber. The chamber consisted of a metal frame and a flexible polyethylene foam base ~~to fit the tree trunk~~ and was fastened with straps around the tree trunk. The inside of the chamber had been carefully wrapped with pre-conditioned (oven-cleaned

190 40 °C, 3 hours) polyamid bags (Toppits, Cofresco Frischhalteprodukte GmbH, Germany) to avoid contamination with BVOC from the chamber foam base. During measurements, the chamber was closed with a metal lid with in- and outgoing PTFE tubing lines (ø 6.35 mm 1/4") for purge air and sample collection was used to close the chamber during the measurement. for purge and sample flow. Temperature of the air Air temperature within the chamber was measured with a temperature probe (HI 145, Hanna Instruments, RI, USA) during sampling BVOC

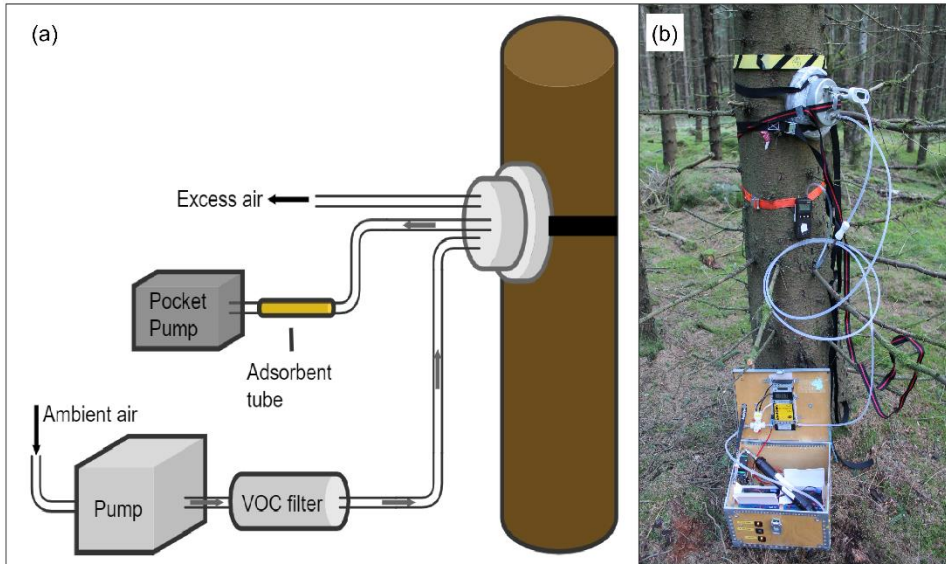
195 collection, and the bark surface temperature was measured with an infrared thermometer (IRT260, Biltema,

Sweden) ~~was used to take bark surface temperature before and after each sample from bark~~ inside the chamber before and after each BVOC collection.

For each campaign, trees from one plot were sampled per day with BVOC collection, typically starting around 200 08:00 (LT) and ending around 19:00 (LT) alternating sampling between the trees. The chamber bases were secured in place in the North or East orientation of the trunk onto the tree trunks every morning. The chambers were and were left open during the day to avoid built up concentrations of BVOCs inside the chambers. Prior to the sampling the bark temperature was measured at four different points inside the base prior to the sampling after which the lid was fastened and the chamber was flushed for 15 minutes before sampling started. Air temperature inside 205 the chamber was measured at the start and at the end of the BVOC collection to note potential temperature differences during the sample period. After sampling, the lid was removed and the bark temperature was measured again.

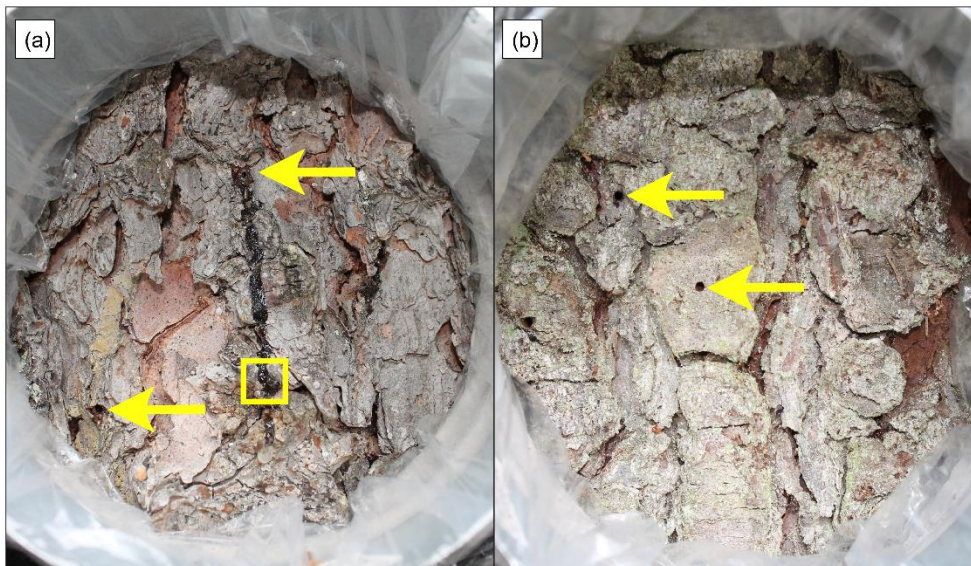
The start of the infestation was determined by the beetles' swarming time in relation to when the tree infestation 210 was detected. For plot 3 in HTM, the start of the infestation was seen during the campaign in June (Table 1) witnessed on site, For the other plots the swarming time was retrieved using data from Skogsstyrelsen Statistical Database (Skogsstyrelsen, n.d.) taken at the plots closest to the measurement sites (Supplementary material, Table S1, Table S2) which showed a A late swarm in HTM was detected around week 25-28 for 2018 (Fig. S1), and a main swarm in NOR around week 20-21 for 2019- (Fig. S2)(Skogsstyrelsen, n.d.), S1 in the 215 Supplement).

The bark inside the chambers was controlled visually before each measurement to count the number of bark beetle holes and to assess potential lichen and algal cover as seen in Table 1. By looking at bark photographs, the holes were later separated into entry or exit holes for the infested trees (Fig. 4). The separation depended on the 220 characteristics of the hole where entry holes were determined to have more resin bleed compared to exit holes, which also had a rounder shape as seen in Fig. 4. Table 1 The number of holes counted inside the chamber area is listed in Table 2 along with an extrapolation of the counted holes from the chamber area to square meter bark area. Entry holes were found for measurements taken up to 100 days after infestation and exit holes were predominantly found for measurements taken after 100 days of infestation, with the latest measurement occurring 350 days after 225 estimated infestation start. As the two hole types were consistently occurring before or after 100 days since infestation start, the measurements with mainly entry holes is referred to as the early season and the measurements with mainly exit holes as the late season. ~~describes the number of visual holes inside the chamber and the extrapolated number per square meter of bark area.~~



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Figure 3: The experimental setuphematics (a) and a field photo (b) contained of the tree trunk chamber mounted on a Norway spruce trunk. The chamber is tree, attached-connected to a pump box used to provide BVOC- and O<sub>3</sub>-free purge air. The BVOC samples were collected with adsorbent tubes connected to the chamber by extracting the air from the chamber using a pocket pump. A photograph of how the setup looked in the field is displayed to the right (b).



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Figure 4: Examples of infested Norway spruce trees with (a) entry holes and (b) exit holes, where the arrows points towards indicate examples of bark beetle drilled holes and the square box frames shows a bark beetle. There are more holes in the pictures than indicated by the arrows. More holes can be found in the picture than pointed to.

240

Table 1Table 2. The infested Norway spruce trees indicated by their tree ID, and their location by their site, and plot and their respective and the number of holes counted inside the respective chamber at the given date for each tree. The counted holes were upscaled extrapolated to holes per square meter of bark surface and the majority of the hole type was determined to either mostly entry exit holes or mostly exit entry holes. The Norway spruce tree with the ID S1S1 was infested during the in-late season 2018 and had thus a already a majority of exit holes early in 2019.

Tree ID	Site	Plot	Date	Number of holes inside chamber	Upscaled to holes per m <sup>2</sup>	Hole type majority
S1S1	HTM	1	2019-05-04	12	1062	exit
S1S1	HTM	1	2019-06-05	8	708	exit
S3S2	HTM	3	2019-06-04	4	354	entry



S3S3	<u>HTM</u>	<u>3</u>	2019-06-04	5	442	entry
S3S2	<u>HTM</u>	<u>3</u>	2019-07-03	3	286	entry
S3S3	<u>HTM</u>	<u>3</u>	2019-07-03	5	465	entry
S3S2	<u>HTM</u>	<u>3</u>	2019-08-01	6	531	entry
S3S3	<u>HTM</u>	<u>3</u>	2019-08-01	15	1273	entry
S4S1	<u>NOR</u>	<u>4</u>	2019-08-21	4	354	exit
S4S2	<u>NOR</u>	<u>4</u>	2019-08-21	5	442	exit
S4S3	<u>NOR</u>	<u>4</u>	2019-08-21	5	442	exit
S5S1	<u>NOR</u>	<u>5</u>	2019-08-22	5	442	exit
S5S2	<u>NOR</u>	<u>5</u>	2019-08-22	4	340	exit
S5S3	<u>NOR</u>	<u>5</u>	2019-08-22	7	619	exit

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### 2.3 BVOC sampling and analysis

250 A total of 187 samples were taken, where 147 samples were from healthy Norway spruce and 40 were from infested Norway spruce (Table 1). Stainless steel cartridges (Markes International Limited, Llantrisant, UK) packed with adsorbents Tenax TA (a porous organic polymer) and Carbograph 1TD (graphitized carbon black) were used to sample BVOCs. The BVOCs were sampled from the chambers using flow-controlled pocket pumps (Pocket Pump, SKC Ltd., Dorset, UK) ~~which extracted by extracting~~ the air through the steel cartridges at a flow rate of 200 ml min<sup>-1</sup> and a sampling time of ca. 30 minutes. ~~The~~ collected volume for each sample was between 5 to 6 Liter. Blank samples were collected ~~from air entering the chamber inlet air twice per day, once once before the first sample and once once after the last sample of the day to capture from air entering the chamber twice a day to capture~~ possible background contamination ~~of the filtered purge air. T~~air temperature ~~was measured inside the chamber was measured at the start of the BVOC sampling and at the end of the BVOC collection to note potential temperature differences during the sample period. After sampling was finished~~ ~~When sampling was done, the chamber lid was removed and, the bark temperature was measured again following the same procedure as mentioned above.~~ ~~The same~~ method was repeated throughout the day until all trees of that plot were measured 260 three times.

After collecting the BVOC samples, the adsorbent cartridges were capped and stored in a refrigerator (at ~3 °C) before being analyzed using a two-stage automated thermal desorption apparatus coupled to a gas-chromatograph mass-spectrometer. Desorption was done on a Turbomatrix ATD 650 (PerkinElmer, Waltham, MA, USA) ~~by~~ ~~Cartridges were initially primary heating the cartridge~~ ~~sted~~ to 280 °C in a flow of purified helium (He, ALPHAGAZ 1, Air Liquide Gas AB, Sweden) for 10 minutes, in order for the ~~VOCs- BVOCs~~ to volatilize. After the primary desorption, ~~VOCs- BVOCs~~ were cryo-focused downstream on a Tenax TA cold trap maintained at -30 °C. The cold trap was flash-heated (40 °C sec<sup>-1</sup>) to 300 °C for 6 minutes to perform a second desorption. The volatilized ~~VOCs- BVOCs~~ were passed via a heated transfer line using He as carrier gas, to a gas chromatograph-mass 270 spectrometry system (GC-MS, Shimadzu QP2010 Plus, Shimadzu Corporation, Japan). The BVOCs were separated using a BPX5 capillary column (50 m, I.D. 0.32 mm, film thickness 1.0 µm, Trajan Scientific, Australia) and the oven temperature was initially held at 40 °C for 1 minute, raised to 210 °C at a rate of 5 °C min<sup>-1</sup> and further increased to 250 °C at a rate of 20 °C min<sup>-1</sup> and lastly held for 2 minutes. Pure standard solutions of isoprene, ~~α~~-pinene, ~~β~~-pinene, p-cymene, eucalyptol, limonene, 3-carene, linalool, ~~α~~-humulene, ~~β~~-caryophyllene, 275 longifolene and myrcene were pre-prepared in methanol (Merck KGaA, Darmstadt, Germany) and injected onto adsorbent cartridges in a stream of Helium and analyzed ~~with the same conditions~~ as samples. When quantifying

BVOCs for which no standards were available,  $\alpha$ -pinene was used for MTs, and  $\alpha$ -humulene for sesquiterpenes (SQT). ~~For other BVOCs which could not be quantified during the study did not match any standard, the amount of the compound present on the sample was calculated as a percentage of the total amount on the sample using the chromatogram peak area.~~ The peaks of longifolene and  $\beta\beta$ -caryophyllene were coeluted in the chromatography and are therefore presented together as a sum of two compounds in this study. The chromatogram peaks were identified based on comparison with retention times and mass spectra of standards and the mass spectra in the NIST08F08 library. LabSolutions GCMS post run analysis program was used for data processing (Version 4.30, Shimadzu Corporation, [Japan](#)). Detection limit was set to 0.4 ng in the analysis software based on the analysis of blank samples.

~~Two outliers in the BVOC samples were found from two Norway spruce trees located at plot 1 in HTM (Fig. 1a). The bark was examined with bark photos and it was revealed it was detected that the chamber in both cases had been placed upon a small emerging branch with some spots of resin as well as with one single needle stuck on the bark. This was believed to have caused the outliers and these samples were considered unusable and excluded from further analysis. All samples from one Norway spruce at plot 2 were also excluded from the analysis after discovering placements on top of a bark hole likely not originating from spruce bark beetles, and thus not suitable in this study.~~

#### 2.4 Emission rate calculation and standardization

The BVOC concentrations obtained from the sample ~~air~~ analysis were converted to emission rate (ER) ( $\mu\text{g m}^{-2} \text{h}^{-1}$ ) according to Eq. (1), following Ortega & Helmig (2008):

$$ER = \frac{[C_{out} - C_{in}]Q}{A}, \quad (1)$$

where  $C_{out}$  ( $\mu\text{g l}^{-1}$ ) is the concentration of each compound within the chamber, and  $C_{in}$  ( $\mu\text{g l}^{-1}$ ) is the concentration of the compound in the filtered inlet air,  $Q$  is the flow rate through the chamber ( $\text{l min}^{-1}$ ) and  $A$  is the bark surface area ( $\text{m}^{-2}$ ) ~~covered by inside~~ the chamber.

The ER per hole was calculated as the average by dividing the ER derived from Eq. (1) for the respective sample with the number of counted holes (~~the number of holes can be found in Table 1 Table 2~~) according to Eq. (2):

$$ER \text{ per hole} = \frac{ER}{\# \text{ holes}}, \quad (2)$$

Finally, the ER for one square meter of bark surface,  $ER_{sqm}$ , was extrapolated based on the number of holes within the chamber and the chamber's bark area according to Eq. (3):

$$ER_{sqm} = ER \cdot \# \text{ holes} \cdot \frac{1}{A}, \quad (3)$$

~~The emission rates of the infested Norway spruce trees were scaled with an average of the holes per square meter found for the measured trees in HTM and NOR (Table 2). By doing this, any variation in emission rate caused by a difference in amount of holes was removed which enabled more accurate comparison between the infested trees. To remove the influence from the variations of the emission rates due to the difference in amount of bark beetle holes, the emission rates were scaled to represent the same number of holes using an average number of bark beetle holes per square meter calculated using the counts from this study in Table 1 Table 2. The average holes per~~

315 ~~square meter was applied to the emission rates for the infested trees, enabling comparison of emission rates from all infested trees.~~

~~The hole type majority was determined from bark photos and used to separate entry and exit holes in the data set. A majority of entry holes were found for, measurements taken up to 100 days after infestation was determined to have and a majority of exit holes for and measurements later than 100 days after since infestation, with the latest measurement occurring 350 days after infestation start, had a majority of exit holes. The timeframe of 100 days was determined by confirming the hole type majority from bark photographs. As the two hole types were consistently occurring before or after 100 days since infestation start, the measurements with mainly entry holes is referred to as the early season and the measurements with mainly exit holes as the late season.~~

325

As the bark surface temperature varied over the season and between the days, the emission rates were standardized using the algorithm for stored, temperature dependent BVOCs (G93) by Guenther et al., (1993; G93) according to Eq. (4):

$$330 \quad M = M_s \cdot e^{(\beta(T-T_s))}, \quad (4)$$

where  $M$  is the emission rate ( $\mu\text{g m}^{-2} \text{h}^{-1}$ ) at a given bark temperature,  $T$ , and  $\beta$  ( $0.09 \text{ K}^{-1}$ ) is an empirical coefficient establishing the temperature dependency (Guenther et al., 1993).  $M_s$  is the emission rate at standard temperature  $T_s$  of  $30 \text{ }^\circ\text{C}$ .

335

The temperature sensitivity of compound emission rates was calculated using a  $Q_{10}$  relationship (Lloyd and Taylor, 1994) following Seco et al. (2020) where the  $Q_{10}$  coefficient represents the factor by which the compound emission rate increases for every  $10 \text{ }^\circ\text{C}$  temperature increase from a reference emission rate,  $F_0$ . Only compounds appearing in more than three individual samples were selected for further analysis using this method. Log transformed  
340 emission rates were binned into  $1 \text{ }^\circ\text{C}$  bins and the mean emission rate per bin was calculated except for bins with only one value. An orthogonal distance regression was applied to the binned mean emission rates weighed by their standard deviation to determine  $Q_{10}$  and  $F_0$  using Eq. (5):

$$345 \quad F = F_0 \cdot Q_{10}^{(T-T_0)/10}, \quad (5)$$

where  $F_0$  is the reference emission rate at temperature  $T_0$  ( $=30 \text{ }^\circ\text{C}$ ),  $F$  is the flux rate at bark surface temperature  $T$  ( $^\circ\text{C}$ ), and  $Q_{10}$  is the temperature coefficient.

Based on the Guenther algorithm (G93, Guenther et al., 1993; Eq. (4)) and the  $Q_{10}$  temperature dependency calculation ( $Q_{10}$ , Lloyd and Taylor, 1994; Eq. (5)), an estimation of the ~~total BVOC constitutive total bark VOC~~  
350 emission rate from healthy Norway spruce bark and the needleleaf VOC- BVOC emission rate throughout the season was calculated. Both algorithms were used to calculate bark BVOC emissions, while only G93 was used to calculate the needleleaf emission, as well as bark SQT emission rate. The ~~calculated modelled~~ emissions for bark was based on the measured tree trunk temperature from the ICOS ecosystem data in HTM (Heliasz, 2020), taken  
355 at 3 meter height. An average of the trunk temperature measurements was taken in the North and East orientation of the trunk was used for this, because the trunk BVOC sampling was taken in these orientations.

The needleleaf emissions for MT and SQT were calculated according to Eq. (4) using the standardized seasonal average emission rate ( $M_s$ ),  $1.25 \mu\text{g g(dw)}^{-1} \text{h}^{-1}$  for MT and  $0.34$  for SQT  $\mu\text{g g(dw)}^{-1} \text{h}^{-1}$ , taken from van Meeningen et al. (2017) measured in HTM during 2016. The temperature input was the canopy-level air temperature measured at 24 meter agl. taken from the HTM ICOS station (Heliasz, 2020). The output of Eq. (4) was scaled from  $\text{g(dw)}$  to  $\text{m}^2$  by using a specific leaf area (SLA) of  $38.4 \text{ cm}^2 \text{ g}^{-1}$  calculated from Wang et al. (2017).

~~Two outliers in the BVOC samples were found and after examination with bark photography, it was discovered that the chamber in both cases had been placed upon a small emerging branch with some spots of resin as well as with one single needle stuck on the bark. These samples were considered unusable and excluded from further analysis. All samples from one spruce at plot 2 were also excluded from the analysis after discovering placements on top of a bark hole likely not originating from spruce bark beetles, and thus not suitable in this study.~~

## 2.5 Statistical analysis

All samples were tested for normality by creating normal probability curves (normplot, MATLAB R2021a, The MathWorks, Inc., MA, USA) which indicated ~~ing~~ no normal distribution in the data. Statistical analysis of all measurements were thus performed using a Kruskal-Wallis test (MATLAB R2021a, The MathWorks, Inc., MA, USA) with a level of significance set to  $P < 0.05$ . To assure ~~there werethat~~ no deviation between the plots of the healthy Norway spruce trees in HTM occurred, ~~the we tested the following scenario~~ was tested: 1) the difference in emission rates from the healthy trees at plot 1, 2 and 3 in HTM. To test the study ~~our~~ aim and hypotheses, ~~t-we tested the following scenarios~~ were tested: 2) the difference in emission rates from healthy and infested trees from all plots and sites (H1), 3) the difference in emission rates from one initially healthy spruce and one initially stressed spruce (aim (iv)) and 4) the difference between the calculated  $Q_{10}$  coefficient and  $F_0$  for the healthy and infested trees (aim (iii)).

To test H2 and H3, an exponential function (Curve Fitting Toolbox, MATLAB R2021a, The MathWorks, Inc., MA, USA) was fitted ~~applied~~ to the data using (Eq. 6):

$$f(x) = a \cdot e^{b \cdot x} \quad (6)$$

~~( $f(x) = a \cdot e^{b \cdot x}$ , where  $x$  is the emission rate in  $\mu\text{g m}^{-2} \text{h}^{-1}$ . The following scenarios were compared:)~~

~~1) emission rates from infested Norway spruce and the evolution over time (H2) and 2) emission rates from infested Norway spruce and the number and type of bark beetle holes (H3).~~

The significance of the difference between the healthy tree sites, the emission rates from control trees and infested trees, the difference in emission rates from the two infested trees S3S2 and S3S3, and the difference between  $Q_{10}$  and  $F_0$  for healthy and infested trees was analyzed using a Kruskal Wallis test. The level of significance was  $p < 0.05$ .

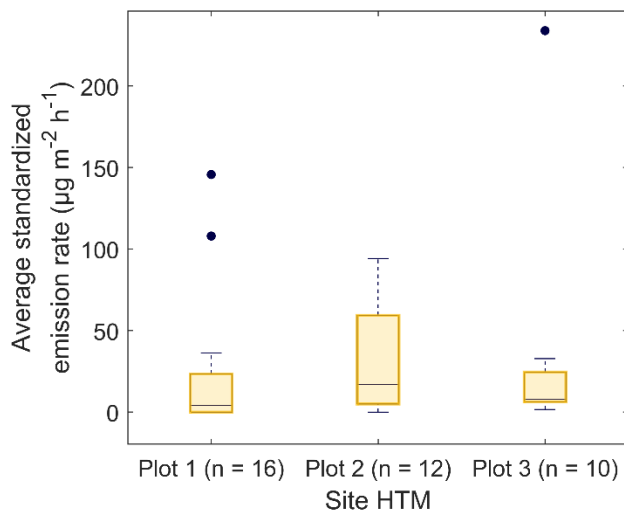
## 3 Results

### 3.1 ~~Constitutive and induced~~ Bark BVOC emissions from healthy and infested Norway spruce ~~bark~~

For the healthy ~~spruce~~ Norway spruce trees in HTM, the average total ~~temperature standardized~~ temperature standardized bark BVOC emission rate from all samples ( $n=113$ ) was  ~~$324.89 \pm 51.672$~~   $\mu\text{g m}^{-2} \text{h}^{-1}$  (mean  $\pm$  standard

deviation; [Table 3](#)). The most dominant BVOC group was MTs ( $29.37 \pm 51.01 \mu\text{g m}^{-2} \text{h}^{-1}$ ; [Table 3](#)) followed by SQTs ( $2.12 \pm 3.217 \mu\text{g m}^{-2} \text{h}^{-1}$ ; [Table 3](#)). Isoprene emissions were detected in 58% of total samples from the healthy spruce tree bark with an average emission rate of  $0.40 \pm 0.985 \mu\text{g m}^{-2} \text{h}^{-1}$  ([Table 3](#)).

The variability of the emission rates differed little between the plots in HTM sites, where plot 1 had a daily average total temperature standardized bark emission rate of  $23.26 \pm 42.53 \mu\text{g m}^{-2} \text{h}^{-1}$ , plot 2 of  $30.38 \pm 32.35 \mu\text{g m}^{-2} \text{h}^{-1}$  and plot 3 of  $47.54 \pm 79.22 \mu\text{g m}^{-2} \text{h}^{-1}$ . The standardized emission rates were ranging from 0-145.7  $\mu\text{g m}^{-2} \text{h}^{-1}$ , for plot 1, 0-94.2  $\mu\text{g m}^{-2} \text{h}^{-1}$  for plot 2 and 1.5-235.4  $\mu\text{g m}^{-2} \text{h}^{-1}$  for plot 3 where the median emission rates were 4.1, 17 and 87.9  $\mu\text{g m}^{-2} \text{h}^{-1}$  respectively (Fig. 5). No statistically significant difference ( $p\text{-value} > 0.3$ ) was found for their standardized emission rates between the sites-plots and no clear pattern of diurnal variation was found in the samples.

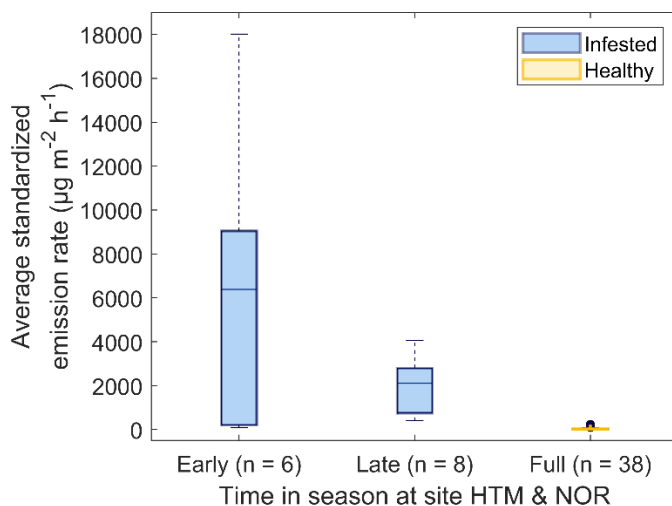


**Figure 5: Boxplots of the temperature standardized emission rates of the control healthy trees for the plot 1-3 in Hyltemossa, where plot 3 is located furthest away from the station in an older forest stand. The number of samples taken at each plot is indicated by n, the black dots indicates outliers, the whiskers indicate the minimum and maximum value and the line indicates the median. The difference in emission rates was tested using a Kruskal-Wallis test (MATLAB R2021a, The MathWorks, Inc., MA, USA) and indicated as not statistically significant difference for the daily average of the total temperature standardized emission rates ( $p > 0.3$ ).**

For the bark beetle infested trees located in both sites (HTM and NOR), the calculations of seasonal average emission rate was separated into early season and late season as the bark beetle typically drill entry holes earlier, and exit holes later. The seasons were separated based on infestation start, where the early season was less than 100 days since infestation start and the late season after more than 100 days of infestation. The average total temperature standardized bark emission rate for Norway spruce infested in from the early season bark beetle infested spruce Norway spruce trees (n = 6) was  $6.70690 \pm 6.90860 \mu\text{g m}^{-2} \text{h}^{-1}$  (mean  $\pm$  standard deviation; [Table 3](#)). T, while the average for trees infested in the late season (n = 8) was  $2.004970 \pm 1.3400 \mu\text{g m}^{-2} \text{h}^{-1}$  ([Table 3](#)). MTs was the most dominant BVOC group throughout the season with an average of  $6.6030 \pm 6.7040 \mu\text{g m}^{-2} \text{h}^{-1}$  for the early season and  $1.9050 \pm 1.3050 \mu\text{g m}^{-2} \text{h}^{-1}$  for the late season, followed by SQTs (early:  $53 \pm 74 \mu\text{g m}^{-2} \text{h}^{-1}$ , late:  $18 \pm 24 \mu\text{g m}^{-2} \text{h}^{-1}$ ; [Table 3](#)). Throughout the season, isoprene was also found in 42% of the samples with an average emission rate of  $3.436 \pm 6.769 \mu\text{g m}^{-2} \text{h}^{-1}$  during the early season and  $0.14 \pm 0.20 \mu\text{g m}^{-2} \text{h}^{-1}$  for the late season ([Table 3](#)).

For all measured Norway spruce trees at both sites, a total of 74 individual VOCs-BVOCs were found throughout the measurement period for all samples (n = 151) whereof 32 were MTs, 5 were SQTs and 37 were classified as other BVOCs including isoprene. For the healthy spruceNorway spruce tree samples in HTM (n = 113), 44  
 435 individual compounds were found in total, where 12 were MTs, 2 were SQTs and 30 other BVOCs including isoprene. The Despite the lower sample count for the infested spruceNorway spruce trees measured at both sites had less samples (n = 38) compared to the healthy tree samples, but a higher number of individual compounds was found with 52 compounds in total where the majority of the compounds were MTs (n = 30) which was more than  
 440 the double compared to the healthy trees. There were also more SQTs (n = 5) found in the infested tree samples, but less other BVOCs including isoprene; (n = 17) compared to the healthy tree samples. For the infested trees, there was also a difference in how many compounds were found early in the season compared to later, in total 40  
 and 33 individual compounds were found for the early and 33 compounds for the late season respectively (Table  
3). For MTs and SQTs, more individual compounds were found in the early season (27 MTs & 5 SQTs) compared to the late season (17 MTs & 2 SQTs), but for the other BVOCs more were found in the later season which had 14  
 445 individual compounds identified compared to 8 in the early season.

A significant difference was found for the daily average of the total standardized bark BVOC emission rate when comparing healthy trees and infested trees healthy and infested trees from all plots and sites, for both early and late season ( $p < 0.0001$ ; Fig. 6). During the early season, the infested trees had a median emission rate of 6,400  
 450  $\mu\text{g m}^{-2} \text{h}^{-1}$  and  $2,100 \mu\text{g m}^{-2} \text{h}^{-1}$  during the late season (Fig. 6). The emission rates for infested trees during the early and late season were around This is around 740- and 240-fold times higher for the infested trees during early and  
late season compared to the median of the healthy trees (The median of the daily average emission rates is 8.6  $\mu\text{g m}^{-2} \text{h}^{-1}$ ;  
Fig. 6), for the healthy control trees, and 6385  $\mu\text{g m}^{-2} \text{h}^{-1}$  and  $2102 \mu\text{g m}^{-2} \text{h}^{-1}$  for the infested trees during  
 early and late season respectively, i.e. around 740 and 240 times higher for the infested individuals early and late  
 455 in the season, respectively.



**Figure 6:** Boxplots of the temperature standardized emission rates of the infested Norway spruce trees (blue) during the early and late season and the healthy Norway spruce trees (yellow) during the full season. The healthy comparison between the daily average total standardized emission rate per measured tree and time for healthy trees were measured in Hyltemossa and infested trees were measured both in Hyltemossa and Norunda. To consider the seasonal pattern of the spruce bark beetle, the infested trees were divided into early and late season, where the early season is dominated by entry holes and the late season by exit holes. The number of samples taken at each plot is indicated by n and the black dots indicates outliers, the whiskers indicate the minimum and maximum value and the line indicates the median. The difference in emission rates for the infested and healthy trees was tested using a Kruskal-Wallis test (MATLAB R2021a, The MathWorks, Inc., MA, USA) which indicates significantly higher emission rates from infested trees during both seasons compared to the healthy trees (P < .001 for both). The emission

rate from the control trees are clearly lower than of the infested trees for both the early and late season, with a significance of  $p < 0.00$  for both. The emission rates are standardized to temperature (30°C).

470 A difference between the ~~healthy trees and infested trees~~ healthy and infested trees from both sites was also  
apparent in the occurrence of the compounds throughout the samples (Table 2 Table 3). The most common MT  
compounds among the healthy ~~spruce~~ Norway spruce trees were  $\alpha$ -pinene (76 %, relative occurrence in all  
samples),  $\beta$ -pinene (55 %), 3-carene (48 %) and limonene (44 %). For infested ~~spruce~~ Norway spruce trees in  
both seasons, the mentioned MTs were also the most occurring compounds ~~but they also, however, they~~ occurred  
475 in more samples (88-100 %). The late season also had 100 % occurrence of (1S)-camphene while that compound  
~~only~~ occurred in 75 % of the samples for the early season. For SQTs,  $\alpha$ -humulene occurred most among the  
healthy trees (47 %) followed by longifolene+ $\beta$ -caryophyllene (17 %). For the infested trees in the ~~early~~ late  
season ~~the pattern was reversed~~, the SQTs occurring most were longifolene+ $\beta$ -caryophyllene (56 %) followed  
by  $\alpha$ -humulene (31 %). ~~For the early season, The late season showed a similar pattern -where the occurrence~~  
480 ~~was similar for~~ longifolene+ $\beta$ -caryophyllene occurred most (56 %) followed by ~~but~~  $\alpha$ -humulene ~~occurred in~~  
~~more samples (18%) which occurred in fewer samples compared to the early season compared to the late season.~~  
The SQTs germaacrene D, isodene and  $\beta$ -cubebene were found to be emitted from one ~~of the~~ infested trees  
~~during~~ in the early season, but was not discovered ~~at~~ any other time. Isoprene was found to be mostly occurring  
among the other BVOCs for both the healthy (58 %) and infested ~~spruce~~ Norway spruce during the early and late  
485 season (63 % and 27 % respectively). After isoprene, decanal (45 %), benzene (45 %), nonanal (38 %) and toluene  
(21 %) were occurring most for the healthy ~~spruce~~ Norway spruce. For the early season infested ~~spruce~~ Norway  
spruce trees, 2-methyl-1-phenylpropene (38 %) and 2-methyl-3-buten-2-ol (19 %) were occurring most after  
isoprene, however, this was not the case for the infested trees in the late season where 2-methyl-3-buten-2-ol was  
not emitted and 2-methyl-1-phenylpropene occurred in 9 % of the samples.

490

Comparing ~~to the~~ constitutive emission rates from ~~healthy infested~~ trees ~~to, the emission rates of healthy infested~~  
trees, the emission rates were ~~shown~~ showed ~~to~~ increases for all individual compounds with increases ranging  
from a 200 % 3-fold increase to a 250,000 % 2580-fold increase for both early and late season (Table 2 Table 3).  
The group of MTs had the highest (230-fold) increase of 22,467,800 % during the early season compared to while  
495 SQTs (increased by 2,400 % 25-fold increase) and isoprene (8-fold increase) e increased with 740 %. The emission  
rates during the late season ~~also contributed~~ showed to a 65-fold increase of 6,600 % for the MTs and 70049 %  
8-fold for the SQTs, however, isoprene was found to have a 0.4-fold decrease ~~from~~ from the infested tree emission  
rates in the late season ~~compared to healthy tree emission rates. (-65 %)~~. The compound (+)-sabinene had the  
highest increase of all individual compounds (2580-fold (2579007863 %)) during the early season when comparing  
500 healthy and infested tree emission rates (Table A1). The compounds: tricyclene, eucalyptol, 4-carene,  $\zeta$ -  
fenchene,  $\alpha$ -phellandrene, ~~trans,trans(4E,6E)-~~alloocimene, norbornane,  $\gamma$ -terpinene,  $\alpha$ -fenchene, 2-  
carene,  $\alpha$ -thujene,  $\alpha$ -terpinene ~~were~~ only found ~~only~~ occurred emitted from ~~in~~ infested trees and indicate a  
change in the chemical composition of the emitted BVOCs ~~when a tree is infested~~ (full table is found in Appendix  
Table A1).

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**Table 2. The compounds occurring most from all samples throughout the season for healthy trees and infested trees. Presented is the temperature standardized seasonal average emission rate ( $\mu\text{g m}^{-2} \text{h}^{-1} \pm$  one standard deviation) for each compound and the groups of MT, SQT and other BVOCs and the occurrence (%) of each compound in all samples. The increase (%) is presented for the infested trees as an increase from healthy to infested. The compounds that were**

510 identified but unable to quantify is presented as n.q. (no quantification). A full list of all identified compounds is found in Table A1.

-	Healthy		Infested early season			Infested late season		
Compound name	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	increase (%)	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	increase (%)	occurrence (%)
<b>Monoterpenes</b>	29.37 $\pm$ 51.01	-	6630 $\pm$ 6740	22678	-	1950 $\pm$ 1350	6608	-
alpha-Pinene	11.49	76.11	911.14	7830	100	824.64	7077	100
beta-Pinene	8.22	55.75	954.17	11508	100	225.28	2641	100
3-Carene	2.48	48.67	285.16	11398	100	33.07	1233	95
Limonene	1.89	44.25	320.82	16875	88	85.43	4420	100
p-Cymene	0.49	39.82	241.45	49176	63	53	10716	77
beta-Myrcene	0.32	17.70	159.76	49825	79	6.3	1869	86
beta-Phellandrene	2.7	10.62	673.13	24831	44	189.47	6917	68
(1S)-Camphene	1.7	6.19	1516	89076	75	388.82	22772	100
(+)-Sabinene	0.08	0.88	206.37	257863	44	2.93	3563	5
<b>Sesquiterpenes</b>	2.12 $\pm$ 3.17	-	53.0 $\pm$ 74	2400	-	18 $\pm$ 24	749	-
Longifolene+beta-Caryophyllene	0.7	17.70	37.65	5279	56	13.6	1843	55
alpha-Humulene	1.42	47.79	4.86	242	31	4.52	218	18
Germaerene-D	-	-	3.86	-	19	-	-	-
Isoledene	-	-	3.25	-	19	-	-	-
beta-Cubebene	-	-	3.2	-	19	-	-	-
<b>Other BVOCs</b>	0.40 $\pm$ 0.85	-	3.36 $\pm$ 6.69	-	-	0.14 $\pm$ 0.20	-	-
Isoprene	0.4	58.41	3.36	740	63	0.14	-65	27
Decanal	n.q.	45.13	n.q.	-	13	n.q.	-	14
Benzene	n.q.	45.13	-	-	-	n.q.	-	14
Nonanal	n.q.	38.94	n.q.	-	6	n.q.	-	14
Toluene	n.q.	21.24	n.q.	-	6	n.q.	-	9
2-Methyl-1-phenylpropene	-	-	n.q.	-	38	n.q.	-	9
2-methyl-3-buten-2-ol	-	-	n.q.	-	19	-	-	-

Table 3. Seasonal average temperature standardized emission rate ( $\mu\text{g m}^{-2} \text{h}^{-1} \pm$  one standard deviation) from all Norway spruce trees located in Hyltemossa and Norunda. Presented areis the frequently occurring unique compounds, compound groups (Monoterpenes, sesquiterpenes and other BVOCs) and total BVOCsemission emitted from healthy and infested (early and late season) Norway spruce bark. The increase or decrease (%) is presented for the infested trees as a change in emission rate from healthy to infested. The occurrence (%)- is presented to indicates how often each compound appeared in the samples throughout the growing season. The compounds that were identified but unable to quantify areis presented as n.q. (no quantification). A full list of all identified compounds is found in the Appendix (Table A1).

515

-	Healthy		Infested early season			Infested late season		
Compound name	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	increase (%)	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	increase (%)	occurrence (%)
<b>Monoterpenes</b>	29 $\pm$ 51	-	6,600 $\pm$ 6,700	22,400	-	1,900 $\pm$ 1,300	6,500	-
$\alpha$ -Pinene	12 $\pm$ 20	76	910 $\pm$ 1030	7,800	100	820 $\pm$ 890	7,100	100
$\beta$ -Pinene	8 $\pm$ 19	56	950 $\pm$ 960	11,500	100	230 $\pm$ 170	2,600	100
3-Carene	3 $\pm$ 5	49	290 $\pm$ 420	11,400	100	30 $\pm$ 40	1,200	95
Limonene	2 $\pm$ 3	44	320 $\pm$ 320	16,900	88	90 $\pm$ 80	4,400	100
p-Cymene	1 $\pm$ 1	40	240 $\pm$ 320	49,200	63	50 $\pm$ 40	10,700	77



<u><math>\beta</math>-Myrcene</u>	<u>0.3 ± 0.8</u>	<u>18</u>	<u>160 ± 170</u>	<u>49,800</u>	<u>79</u>	<u>10 ± 6</u>	<u>1,900</u>	<u>86</u>
<u><math>\beta</math>-Phellandrene</u>	<u>3 ± 8</u>	<u>11</u>	<u>670 ± 660</u>	<u>24,800</u>	<u>44</u>	<u>190 ± 160</u>	<u>6,900</u>	<u>68</u>
<u>(1S)-Camphene</u>	<u>2 ± 6</u>	<u>6</u>	<u>1,520 ± 1,970</u>	<u>89,100</u>	<u>75</u>	<u>390 ± 230</u>	<u>22,800</u>	<u>100</u>
<u>(+)-Sabinene</u>	<u>0.1 ± 0</u>	<u>1</u>	<u>210 ± 200</u>	<u>257,900</u>	<u>44</u>	<u>3 ± 0</u>	<u>3,600</u>	<u>5</u>
<b><u>Sesquiterpenes</u></b>	<b><u>2.1 ± 3.2</u></b>	<b>-</b>	<b><u>53 ± 74</u></b>	<b><u>2,400</u></b>	<b>-</b>	<b><u>18 ± 24</u></b>	<b><u>700</u></b>	<b>-</b>
<u>Longifolene+<math>\beta</math>-Caryophyllene</u>	<u>0.7 ± 1.4</u>	<u>18</u>	<u>38 ± 70</u>	<u>5,300</u>	<u>56</u>	<u>14 ± 24</u>	<u>1,800</u>	<u>55</u>
<u><math>\alpha</math>-Humulene</u>	<u>1.4 ± 3</u>	<u>48</u>	<u>5 ± 9</u>	<u>200</u>	<u>31</u>	<u>4 ± 13</u>	<u>200</u>	<u>18</u>
<u>Germacrene D</u>	<u>-</u>	<u>-</u>	<u>4</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Isolatedene</u>	<u>-</u>	<u>-</u>	<u>3</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u><math>\beta</math>-Cubebene</u>	<u>-</u>	<u>-</u>	<u>3</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<b><u>Other BVOCs</u></b>	<b><u>0.4 ± 0.9</u></b>	<b>-</b>	<b><u>3.4 ± 6.7</u></b>	<b>-</b>	<b>-</b>	<b><u>0.1 ± 0.2</u></b>	<b>-</b>	<b>-</b>
<u>Isoprene</u>	<u>0.4 ± 0.9</u>	<u>58</u>	<u>3 ± 7</u>	<u>700</u>	<u>63</u>	<u>0.1 ± 0.2</u>	<u>-65</u>	<u>27</u>
<u>Decanal</u>	<u>n.q</u>	<u>45</u>	<u>n.q</u>	<u>-</u>	<u>13</u>	<u>n.q</u>	<u>-</u>	<u>14</u>
<u>Benzene</u>	<u>n.q</u>	<u>45</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>n.q</u>	<u>-</u>	<u>14</u>
<u>Nonanal</u>	<u>n.q</u>	<u>39</u>	<u>n.q</u>	<u>-</u>	<u>6</u>	<u>n.q</u>	<u>-</u>	<u>14</u>
<u>Toluene</u>	<u>n.q</u>	<u>21</u>	<u>n.q</u>	<u>-</u>	<u>6</u>	<u>n.q</u>	<u>-</u>	<u>9</u>
<u>2-Methyl-1-phenylpropene</u>	<u>-</u>	<u>-</u>	<u>n.q</u>	<u>-</u>	<u>38</u>	<u>n.q</u>	<u>-</u>	<u>9</u>
<u>2-Methyl-3-buten-2-ol</u>	<u>-</u>	<u>-</u>	<u>n.q</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<b><u>Total</u></b>	<b><u>32 ± 52</u></b>	<b>-</b>	<b><u>6,700 ± 6,900</u></b>	<b><u>20,900</u></b>	<b>-</b>	<b><u>2,000 ± 1,300</u></b>	<b><u>6,000</u></b>	<b>-</b>

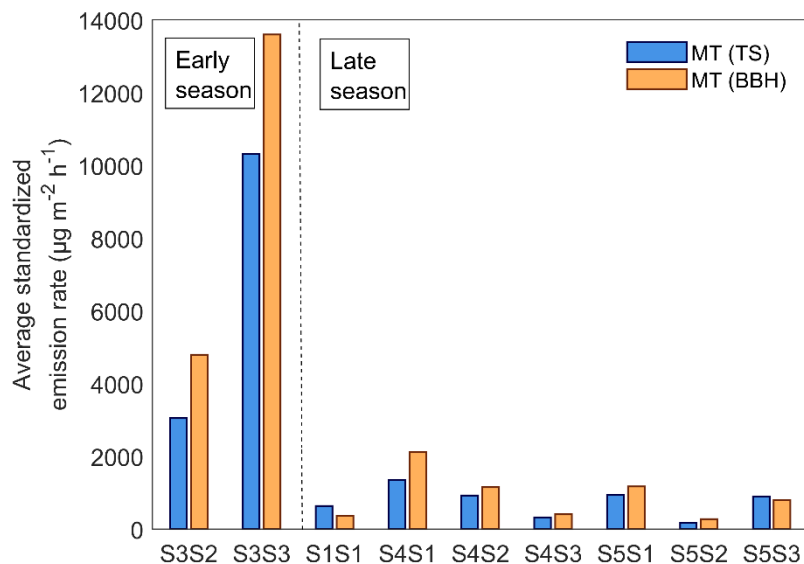
520

<u>-</u>	<b><u>Healthy</u></b>		<b><u>Infested early season</u></b>			<b><u>Infested late season</u></b>		
<b><u>Compound name</u></b>	<b><u>average ± std</u></b> <b><u>(<math>\mu\text{g m}^{-2} \text{h}^{-1}</math>)</u></b>	<b><u>occurrence</u></b> <b><u>(%)</u></b>	<b><u>average ± std</u></b> <b><u>(<math>\mu\text{g m}^{-2} \text{h}^{-1}</math>)</u></b>	<b><u>increase</u></b> <b><u>(%)</u></b>	<b><u>occurrence</u></b> <b><u>(%)</u></b>	<b><u>average ± std</u></b> <b><u>(<math>\mu\text{g m}^{-2} \text{h}^{-1}</math>)</u></b>	<b><u>increase</u></b> <b><u>(%)</u></b>	<b><u>occurrence</u></b> <b><u>(%)</u></b>
<b><u>Monoterpenes</u></b>	<b><u>29 ± 51</u></b>	<b>-</b>	<b><u>6600 ± 6700</u></b>	<b><u>22400</u></b>	<b>-</b>	<b><u>1900 ± 1300</u></b>	<b><u>6500</u></b>	<b>-</b>
<u><math>\alpha</math>-Pinene</u>	<u>11.5</u>	<u>76</u>	<u>910</u>	<u>7800</u>	<u>100</u>	<u>820</u>	<u>7100</u>	<u>100</u>
<u><math>\beta</math>-Pinene</u>	<u>8.2</u>	<u>56</u>	<u>950</u>	<u>11500</u>	<u>100</u>	<u>230</u>	<u>2600</u>	<u>100</u>
<u><math>\beta</math>-Carene</u>	<u>2.5</u>	<u>49</u>	<u>290</u>	<u>11400</u>	<u>100</u>	<u>30</u>	<u>1200</u>	<u>95</u>
<u>Limonene</u>	<u>1.9</u>	<u>44</u>	<u>320</u>	<u>16900</u>	<u>88</u>	<u>90</u>	<u>4400</u>	<u>100</u>
<u>p-Cymene</u>	<u>0.5</u>	<u>40</u>	<u>240</u>	<u>49200</u>	<u>63</u>	<u>50</u>	<u>10700</u>	<u>77</u>
<u><math>\beta</math>-Myrcene</u>	<u>0.3</u>	<u>18</u>	<u>160</u>	<u>49800</u>	<u>79</u>	<u>10</u>	<u>1900</u>	<u>86</u>
<u><math>\beta</math>-Phellandrene</u>	<u>2.7</u>	<u>11</u>	<u>670</u>	<u>24800</u>	<u>44</u>	<u>190</u>	<u>6900</u>	<u>68</u>
<u>(1S)-Camphene</u>	<u>1.7</u>	<u>6</u>	<u>1520</u>	<u>89100</u>	<u>75</u>	<u>390</u>	<u>22800</u>	<u>100</u>
<u>(+)-Sabinene</u>	<u>0.1</u>	<u>1</u>	<u>210</u>	<u>257900</u>	<u>44</u>	<u>0</u>	<u>3600</u>	<u>5</u>
<b><u>Sesquiterpenes</u></b>	<b><u>2.1 ± 3.2</u></b>	<b>-</b>	<b><u>53 ± 74</u></b>	<b><u>2400</u></b>	<b>-</b>	<b><u>18 ± 24</u></b>	<b><u>700</u></b>	<b>-</b>
<u>Longifolene+<math>\beta</math>-Caryophyllene</u>	<u>0.7</u>	<u>18</u>	<u>40</u>	<u>5300</u>	<u>56</u>	<u>10</u>	<u>1800</u>	<u>55</u>
<u><math>\alpha</math>-Humulene</u>	<u>1.4</u>	<u>48</u>	<u>4.9</u>	<u>200</u>	<u>31</u>	<u>4.5</u>	<u>200</u>	<u>18</u>
<u>Germacrene D</u>	<u>-</u>	<u>-</u>	<u>3.9</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Isolatedene</u>	<u>-</u>	<u>-</u>	<u>3.3</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u><math>\beta</math>-Cubebene</u>	<u>-</u>	<u>-</u>	<u>3.2</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<b><u>Other BVOCs</u></b>	<b><u>0.4 ± 0.9</u></b>	<b>-</b>	<b><u>3.4 ± 6.7</u></b>	<b>-</b>	<b>-</b>	<b><u>0.1 ± 0.2</u></b>	<b>-</b>	<b>-</b>
<u>Isoprene</u>	<u>0.4</u>	<u>58</u>	<u>3.4</u>	<u>700</u>	<u>63</u>	<u>0.1</u>	<u>-65</u>	<u>27</u>
<u>Decanal</u>	<u>-</u>	<u>45</u>	<u>-</u>	<u>-</u>	<u>13</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Benzene</u>	<u>-</u>	<u>45</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Nonanal</u>	<u>-</u>	<u>39</u>	<u>-</u>	<u>-</u>	<u>6</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Toluene</u>	<u>-</u>	<u>21</u>	<u>-</u>	<u>-</u>	<u>6</u>	<u>-</u>	<u>-</u>	<u>9</u>

<del>2-Methyl-1-phenylpropene</del>	=	=	=	=	<del>38</del>	=	=	<del>9</del>
<del>2-Methyl-3-buten-2-ol</del>	=	=	=	=	<del>19</del>	=	=	=
<b>Total</b>	<del>32 ± 52</del>	=	<del>6700 ± 6900</del>	<del>20900</del>	=	<del>2000 ± 1300</del>	<del>6000</del>	=

### 525 3.1.1 Scaling the infested tree bark emission with number of bark beetle holes

530 The temperature standardized emission rates (TS) for the total BVOCs from the bark beetle infested Norway spruce trees from both sites were ranging had seasonal averages ranging from around about 500 to 13,000  $\mu\text{g m}^{-2} \text{h}^{-1}$  for all trees throughout the measurement period (Fig. 7). The daily average TS emission rate per bark beetle hole for infested trees during the early season was ~~22.04 ± 298.69~~  $\mu\text{g hole}^{-1} \text{h}^{-1}$  which is emissions from mainly entrance holes and. The bark beetle holes from infested trees during the late season and with mainly exit holes had a daily average temperature standardized emission rate of ~~4.41 ± 3.51~~  $\mu\text{g hole}^{-1} \text{h}^{-1}$  for the late season. The average number of holes per square meter bark area found in this study was 554 based on the values in Table 1. When applying scaling the TS emission rates with the average number of bark beetle holes to emission rates with bark beetle holes (BBH), a comparison with only the TS emission rates showed that the total average of BBH emission rates were higher (Fig. 7). During the early season for the trees measured in the early season the BBH emission rates were about 2,500  $\mu\text{g m}^{-2} \text{h}^{-1}$  higher (~~9,200 ± 6,2030~~  $\mu\text{g m}^{-2} \text{h}^{-1}$ ) compared to TS emission rates (~~6,70690 ± 5,140000~~  $\mu\text{g m}^{-2} \text{h}^{-1}$ ) and for. For the trees measured in the late season the BBH emission rates were 150  $\mu\text{g m}^{-2} \text{h}^{-1}$  higher (~~900 ± 650~~  $\mu\text{g m}^{-2} \text{h}^{-1}$ ) compared to TS emission rates (~~750 ± 400~~  $\mu\text{g m}^{-2} \text{h}^{-1}$ ). For the individual trees and for the group of MT, the BBH emission rates of MT increased compared to TS emission rates for all trees but two (tree S1S1 and tree S5S3) where the TS emission rates were about 300  $\mu\text{g m}^{-2} \text{h}^{-1}$  higher (Fig. 7). The inconsistent variation in emission rates scaled with BBH or TS can be explained by the difference in number of bark beetle holes found per tree (Table 1 Table 2). The TS emission rates only consider the bark beetle holes inside the bark chamber while the BBH emission rates are calculated based on holes extrapolated to average holes per square meter. By applying and an average of the bark beetle holes found in this study the same average of bark beetle holes were used for all to all trees, any variations in emission rates caused by a different due to amount of holes can be disregarded can be disregarded. The results from the infested trees are were thus from here on presented as BBH emission unless stated otherwise.



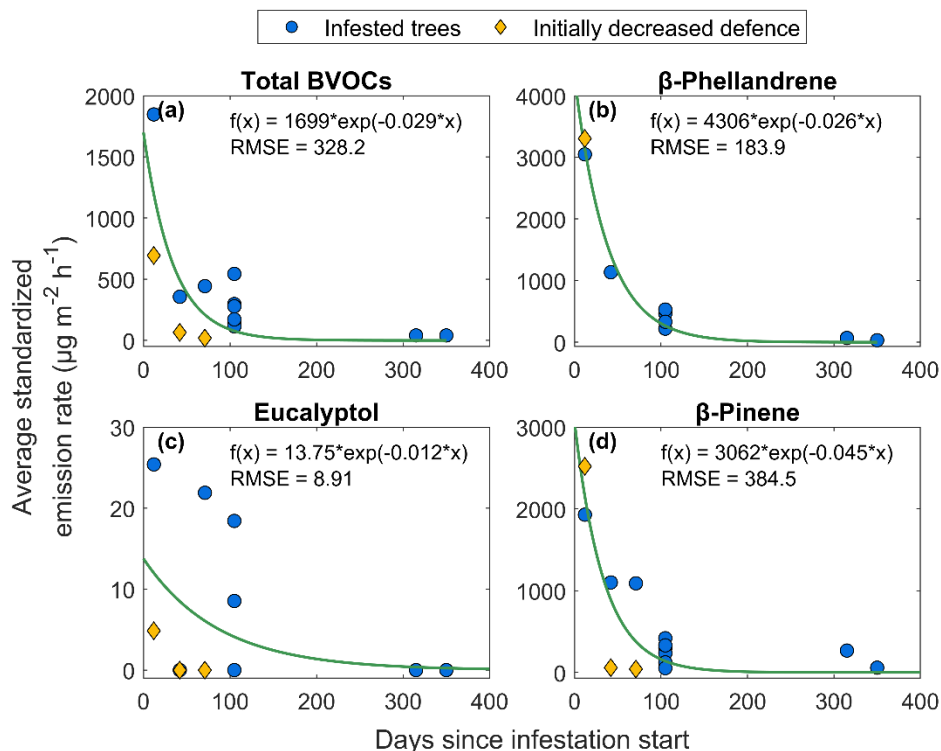
550 **Figure 7: The seasonal average temperature standardized emission rate for the group of monoterpenes (MT) from  
 all infested Norway spruce trees located in Hyltemossa and Norunda, for the group of monoterpenes (MT). The tree ID  
 is presented at the x-axis and is separated into early season (<100 days since infestation start) and late season (less or  
 555 more than 100 days since infestation start). The temperature standardized emission  
 rates (TS) are presented in blue, while the emission rates also scaled to the same by the average number of number of  
bark beetle holes (BBH) is presented in orange. The emission rates are a lot higher in the early season where the  
emissions occur mainly from entry holes compared to the late season with mainly exit holes. This is evident even if the  
emission rates were re-calculated to represent equal number of holes.**

### 3.2 The influence of time since infestation on emission rate from infested trees

To study the influence of time on the emission rates after a Norway spruce trees were infested by spruce bark  
 beetles, measurements were ~~done as different~~ taken at different occasions in relation ~~times to~~ after infestation start  
 560 in both HTM and NOR ~~for both sites~~. The earliest ~~After a tree was infested by bark beetles, its emission rates were~~  
~~found to decrease with time passed since the start of the infestation. When the measurements occurred~~ started after  
 12 days ~~since the start of after infestation and showed an,~~ the average emission rate ~~for for the total all VOCs~~  
 BVOCs of ~~where~~ around ~~1,90850~~  $\mu\text{g m}^{-2} \text{h}^{-1}$ , when excluding ~~at the~~ tree with lowered defence (as presented in Sect.  
 3.4.3.2.3; Fig. 8a; excluded tree is marked in yellow). An exponential function ( ~~$f(x) = a \times e^{b \times x}$ , where  $x$  is the~~  
 565 ~~emission rate in  $\mu\text{g m}^{-2} \text{h}^{-1}$~~ ) was fitted to all data points according to Eq. 6. Three compounds were selected for  
 further analysis,  $\beta\beta$ -phellandrene, eucalyptol and  $\beta\beta$ -pinene, using the same exponential function. The emission  
 rates after 12 days were different for the individual compounds compared to the total average,  $\beta\beta$ -phellandrene  
 and  $\beta\beta$ -pinene have emission rates of around  $3,000$ - $3,500 \mu\text{g m}^{-2} \text{h}^{-1}$  and  $2,000$ - $2,500 \mu\text{g m}^{-2} \text{h}^{-1}$  respectively (Fig.  
 8b, ~~and~~ d). Eucalyptol was emitted ~~at slightly~~ lower rates of around 4 and  $25 \mu\text{g m}^{-2} \text{h}^{-1}$ , ~~depending on if the tree~~  
 570 ~~had lowered defence or not~~, where the low emission rates came from the lowered defence tree (Fig. 8c). Some  
 compounds were not emitted from all infested trees: eucalyptol was only observed from 4 individual trees, and  
 $\beta\beta$ -phellandrene from 7 trees, while  $\beta\beta$ -pinene was emitted from all infested trees ( $n = 9$ ).

After about 100 days since start of infestation, the trees were showing signs of browning and loss of needles ~~and~~  
 575 ~~The emission rates for the total BVOCs had decreased with about 80 %~~ showed a 5-fold decrease on average,  
~~emitting around from the start to around~~  $300 \mu\text{g m}^{-2} \text{h}^{-1}$ . ~~The emission rates were~~ but still at levels higher than the  
 seasonal ~~constitutive~~ emissions from healthy Norway spruce trees in HTM (around  $30 \mu\text{g m}^{-2} \text{h}^{-1}$ ; Fig. 8).  
 Compared to the emission rate of the total BVOCs after 100 days since infestation, the emission rates from the  
 compounds  $\beta\beta$ -phellandrene and  $\beta\beta$ -pinene were at about the same level, however, the decrease on average since

580 the start of infestation was higher (88 % and 89 %, respectively around 9-fold). Eucalyptol did not have an as distinct decrease but had only decreased with, after 100 days the emission rates had only decreased with 10 % 1-fold on average on average. When the Norway spruce with ID S1S1 located in plot 1 in HTM was tree measured after more than 300 days since the start of infestation start it had lost almost had lost almost all of its needles and some bark. At that time, the, and the emission rates of the total BVOC emission rates from that trees were down to around 40  $\mu\text{g m}^{-2}\text{h}^{-1}$ , which was at the same level as the constitutive emissions from healthy trees at the same time (on average 38  $\mu\text{g m}^{-2}\text{h}^{-1}$ ; Fig. 8). No emissions of eucalyptol were found after more than 300 days, but the emission rates of  $\beta\beta$ -phellandrene and  $\beta\beta$ -pinene after 315 days post infestation were around 70 and 270  $\mu\text{g m}^{-2}\text{h}^{-1}$  respectively, however, after 350 days the emission rates went down to around 32 and 58  $\mu\text{g m}^{-2}\text{h}^{-1}$  respectively, also comparable with the constitutive emissions from healthy trees at that time (around 45  $\mu\text{g m}^{-2}\text{h}^{-1}$ ).



595 **Figure 8:** The relationship between average temperature standardized emission rate from all infested Norway spruce trees (blue circle) in Hyltemossa and Norunda and the number of days since after passed since start of infestation start for (a) total all BVOCs and the compounds, (b) beta-phellandrene, (c) eucalyptol and (d) beta-pinene. An exponential curve was fitted to the data according to Eq. 6 (green line). All trees are included in the exponential fitted curve, however one tree had initially lowered defence and from a late bark beetle attack previous season and is in the figure marked specifically in the figure (yellow diamond) in yellow and diamond shape for visualization.

### 3.3 The difference in BVOC emission rates from bark beetle entry holes and exit holes

600 No clear relationship was found between the total number of holes and emission rates, likely due to a mixed signal from the type (entry or exit) and time since infestation. The total BVOC temperature standardized emission rates were generally lower from exit holes compared to entry holes when the Norway spruce number of holes were trees had similar amounts of holes (Fig. 9a). The individual compounds emitted from both the entry and exit holes were dominated by  $\beta\beta$ -phellandrene,  $\beta\beta$ -pinene,  $\alpha\alpha$ -pinene and (1S)-camphene (Table 3). The compounds found from entry holes but not from exit holes were: 2-carene, 4-carene,  $\alpha\alpha$ -fenchene,  $\alpha\alpha$ -phellandrene,  $\alpha\alpha$ -terpinene,  $\alpha\alpha$ -thujene,  $\beta\beta$ -cubebene,  $\gamma$ -terpinene, germacrene D, isodene and (4E,6E)-trans-~~alloocimene~~ (Table 3). Generally lower emission rates from exit holes were also seen for the compounds  $\beta\beta$ -phellandrene and  $\beta\beta$ -pinene (Fig. 9b,d).

However, for the compound eucalyptol (Fig. 9c) emissions were only found from four individuals, which had similar emission rates regardless of entry or exit holes. The oxygenated ~~compounds-monoterpenes~~ myrtenal and bornyl acetate were only found in entry holes but could not be quantified (Table 3).

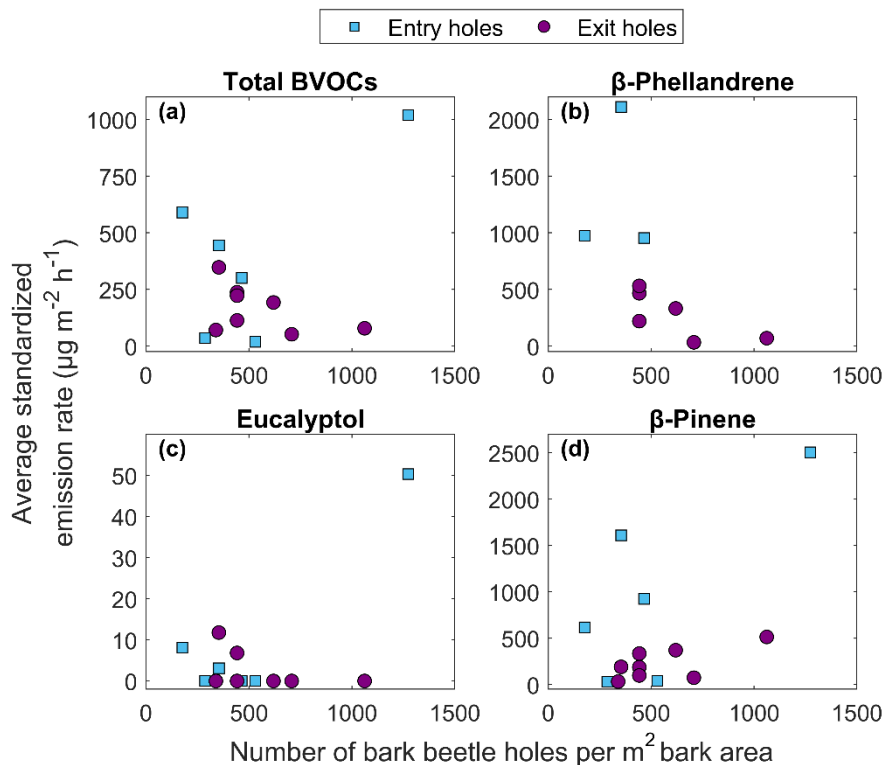


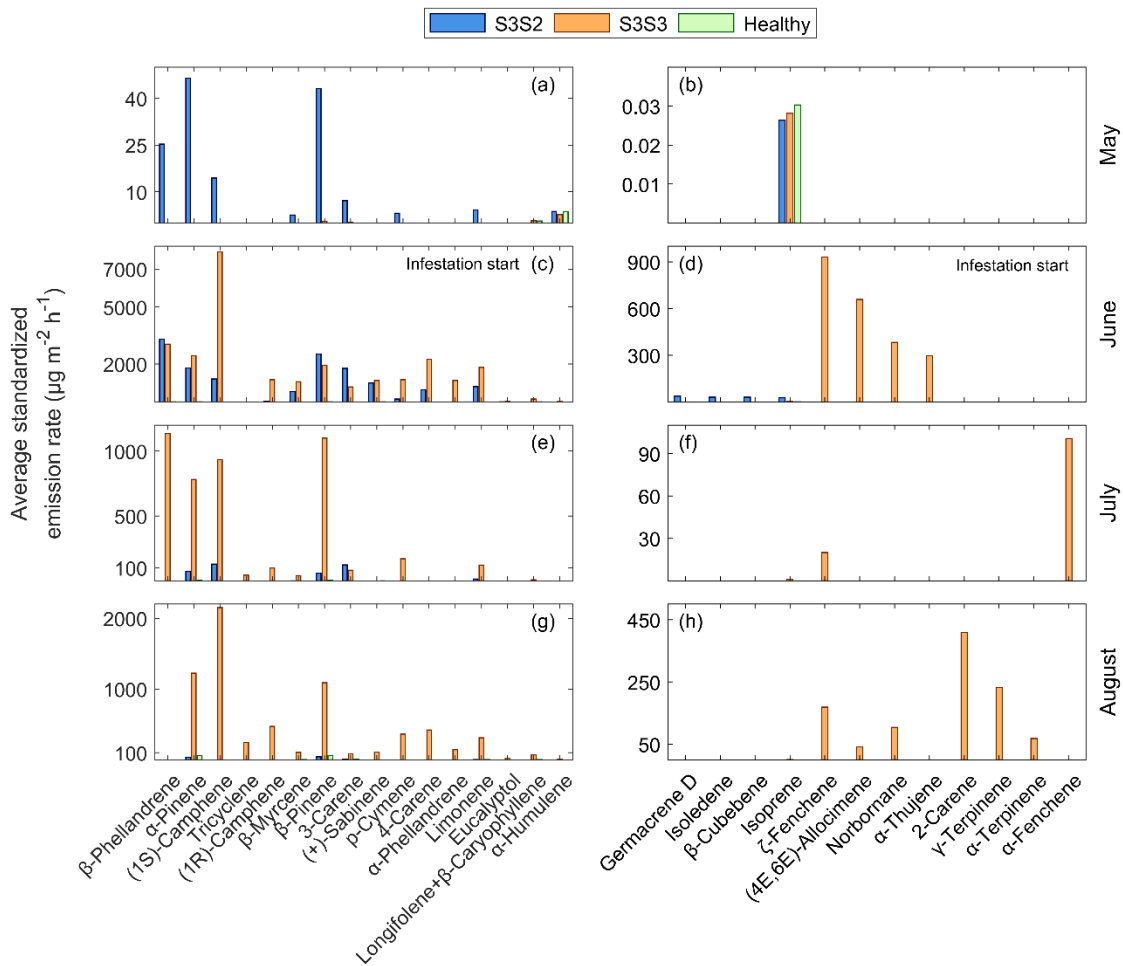
Figure 9: The relationship between average ~~temperature~~ standardized emission rate from infested Norway spruce trees in Hyltemossa and Norunda and the number of bark beetle holes per m<sup>2</sup> bark area for (a) ~~total~~ BVOCs, ~~and the compounds:~~ (b) ~~beta~~ phellandrene, (c) eucalyptol and (d) ~~beta-pinene~~ pinene. ~~There is a distinction between the entry~~ The bark beetle holes are separated into entry holes (bluegreen squares) and the exit holes (purple circles). ~~The exit holes were appearing later during the season (> 100 days since infestation start) which could explain the lower emission rates as the vitality of the trees decrease with higher number of holes. High emission rates from exit holes could indicate signals from blue stain fungi.~~

### 3.4 Bark beetle infestation impact over time from two trees with different initial health status

As a part of the sub-study, ~~at the~~ At plot 3, where the bark beetle trap was installed at plot 3 in HTM (Fig. 1). This resulted in, successful bark beetle infestation of two Norway spruce trees with different initial health status, one healthy (tree ID S3S3) and one stressed (tree ID S3S2) were infested by bark beetles. One of the trees had no visual stress signs and was assumed to be healthy (S3S3). The other tree (S3S2) had visual signs of stress already before the beetle attack with old resin flow located high on the stem, most likely due to a late summer attack the previous season. The different status of the trees can be identified in Fig. 10 (a-b) where S3S2 had significantly higher ( $p > 0.02$ ) total emission rate of bark BVOCs in May, before the infestation, compared to the healthy ~~spruce~~ Norway spruce (S3S3). ~~The remaining two Norway spruce trees on plot 3 were not infested and a~~ An average of the ~~emission rates for two control healthy trees at plot 3 were taken from these trees were taken to compare with the infested trees is also presented in Figure 8, however, only~~ Only four compounds ~~for the healthy trees~~ were found in May, longifolene + ~~ββ~~ caryophyllene, ~~αα~~ humulene and isoprene (Fig. 10a,b), which had similar emission rates as the other trees and ~~t~~. The ~~control healthy~~ trees remained at low emission rates for the remaining months (Fig. 10a-h) and are thus not included in further comparison of the trees (Fig. 10). The total emission rates for ~~S3S3 and S3S2 both trees~~ were induced in June when the bark beetle infestation started (Fig. 10c-d). ~~There was~~

~~no-but there was no~~ significant difference ( $p > 0.2$ ) in emission rates between the trees, but the compound blend differed between the S3S3 and S3S2. ~~The tree~~ However, a difference was seen in the compound blend. S3S3 had a higher emission rate from the compound ~~1S-camphene(1S)-Camphene~~ (about  $800 \mu\text{g m}^{-2}\text{h}^{-1}$ ) and was also emitting ~~zeta-fenchene, trans,trans(4E,6E)-alloocimene, norbornane and alpha-thujene, compounds~~ which were  
640 not emitted from S3S2. The samples from S3S2 in June (Fig.10c,d) did however contain the bark beetle pheromones germacrene D, isodene and  $\beta\beta$ -cubebene which were not found in the compound blend from S3S3. In July, the emission rate from S3S2 was significantly lower ( $p < 0.0325$ ) ~~and close to zero~~ compared to S3S3, which still had high emission rates of  $\beta\beta$ -phellandrene,  $\alpha\alpha$ -pinene,  $\beta\beta$ -pinene and ~~1S-camphene(1S)-Camphene~~ (Fig. 10e,f). A similar difference between the trees was apparent in August as well (Fig. 10e-f) where the emission  
645 rates from S3S2 ~~were~~ still significantly lower ( $p < 0.003$ ) compared to S3S3 (Fig. 10g,h). The ~~spruce~~Norway spruce S3S3 was also found to emit the compound verbenone in August, a compound which could not be quantified in the study (Table 3), ~~which was not found in S3S2.~~

The individual compound blend was also found to change over time for the healthy tree, S3S3 ~~as it was~~, ~~when it got~~ infested and ~~when~~as the infestation continued (Fig. A1). ~~In May, when the tree was healthy, Before infestation,~~ ~~in total~~ ~~in total~~ 10 compounds ~~in total~~ were identified, ~~with dominant BVOC being dominated by other BVOCs:~~ decanal (28 %), nonanal (20 %), toluene (16 %) and 1,3,5-trifluorobenzene (15 %) where the percentage represents the amount of the compound found in the sample relative to the total amount (Fig. A1) ~~by mass~~. After bark beetle infestation ~~started~~ in June, the number of detected compounds ~~increased~~~~raised~~ to 27 ~~and was now~~, dominated by  
655 MTs (~~1S-camphene(1S)-Camphene~~ (18%), ~~5-vinyl-m-xylene~~ (15%) and  $\beta\beta$ -phellandrene (9%); Fig. A1) ~~but also other BVOCs (5-vinyl-m-xylene (15%)).~~ The emissions during the campaign in July ~~also~~ consisted mainly of MTs with largest contributions from  $\beta\beta$ -phellandrene (23 %) and  $\beta\beta$ -pinene (23 % and 22 % ~~respectively~~) followed by ~~1S-camphene(1S)-Camphene~~ (19 %) and  $\alpha\alpha$ -pinene (19 % and 16 %; Fig A1 ~~respectively~~). The compound composition in August was similar as in June, with the majority of the blend consisting of MTs  
660 dominated by ~~1S-camphene(1S)-Camphene~~,  $\alpha\alpha$ -pinene and  $\beta\beta$ -pinene (22 %, 13 % & 11 % respectively) and other BVOCs (2-methyl-1-phenylpropene (6%); Fig. A1).



665 **Figure 10: The average temperature standardized BVOC emission rates for all compounds from Norway spruce at plot 3 in Hyltemossa: healthy trees (green), infested spruce with ID S3S2 (blue), and infested spruce with ID S3S3 (orange) and healthy trees (green). Measurements were taken in 2019 during (a,b) May, (c,d) June, (e,f) July and (g,h) August. The graphs are horizontally separated for visibility due to large differences in scale. The healthy trees are included in all graphs but the emission rates are not visible on the same scale as the infested trees in (c,d) June or (e,f) July. The bark beetle infestation had not started in (a,b) May, however, the spruce S3S2 was already subjected to stress from late bark beetle attacks during the previous season before the bark beetle infestation started again in (c,d) June, leading to higher emission rates already in May.**

670 **The average standardized emission rates for all compounds for the two spruces S3S2 (blue), S3S3 (orange) and control trees at site 3 (green) for May (a-b), June (c-d), July (e-f) and August (g-h). The graphs are horizontally separated for visibility due to large differences in scale. The control trees are included in all graphs but the emission rates are not visible on the same scale as the infested trees in June (c-d) or July (e-f). The bark beetle infestation had not started in May (a-b), however, the spruce S3S2 was already subjected to stress from late bark beetle attacks previous season before the bark beetle infestation started again in June (c-d), leading to higher emission rates in May.**

675

### 3.5 Reference emission rate at 30°C and calculated $Q_{10}$ coefficient

680 From Eq. (5), the reference emission rate at 30°C ( $F_0$ ) and the increase in emission rate with every 10°C ( $Q_{10}$  coefficient) were calculated for both the healthy and infested spruce Norway spruce trees as well as for the infested trees from both sites. Only compounds found in at least three samples were included in the calculation (Table A2). The result for emitted compounds from both healthy trees and infested trees shows a  $Q_{10}$  coefficient ranging spread from 0.1 to 576 for healthy trees and 1.3 to 9804 for infested trees where the  $Q_{10}$  coefficient was higher when the trees were infested increased for infested trees for all compounds but one, p-cymene (Table A2).

685 The  $F_0$  value, which indicates the emission rate that compound would have at 30°C for each specific compound, and for this value there was also showed a difference between the healthy and infested trees. The spread of  $F_0$  for healthy trees was ranging from 0.0105 to 932  $\mu\text{g m}^{-2} \text{h}^{-1}$  and from compared to 0.5 to 34,900  $\mu\text{g m}^{-2} \text{h}^{-1}$  for the

infested trees, ~~however, compared to the  $Q_{10}$  coefficient, the but in this case the  $F_0$  value~~ was higher for all  
690 compounds from the infested trees compared to the healthy (Table A2). The average  $Q_{10}$  coefficient for all  
compounds for healthy trees was 13 while it was 96 for infested trees, ~~indicating a 7-fold increase leading to an  
increase of the  $Q_{10}$  coefficient by 60034 %.~~ The same average for  $F_0$  was ~~210.8~~  $\mu\text{g m}^{-2} \text{h}^{-1}$  for healthy trees and  
~~2,65048~~  $\mu\text{g m}^{-2} \text{h}^{-1}$  for infested trees, a ~~127-fold~~ increase, ~~an increase of 12,600 %~~, which is in line with the  
increased emission rates when standardized according to G93 (Table A1). The highest increase in both  $Q_{10}$  and  $F_0$   
695 was seen in the compounds  $\beta\beta$ -pinene and longifolene +  $\beta\beta$ -caryophyllene (~~two SQTs quantified together~~) with  $Q_{10}$   
increasing ~~125-fold with 12,500 %~~ and ~~22,400 %~~ ~~225-fold~~ respectively, and  $F_0$  ~~increasing with 3,160-fold~~ ~~316,000  
%~~ and ~~209,800 %~~ ~~2,100-fold~~ respectively. The lowest change for  $Q_{10}$  was seen in  $\alpha\alpha$ -pinene and p-cymene, where  
 $\alpha\alpha$ -pinene ~~had a 1-fold increase~~ ~~was increasing with 0.7 %~~ from healthy to infested and the  $Q_{10}$  coefficient for p-  
cymene ~~was actually decreasing with 34 %~~ ~~had a 0.6-fold decrease~~ for infested trees compared to healthy. Despite  
700 the lower  $Q_{10}$  coefficient for p-cymene in infested trees, its  $F_0$  value was still higher for the infested trees, however,  
~~it had the lowest increase with the increase of 45049 %~~ ~~5-fold~~ ~~was low~~ compared to the other MT compounds.  
Isoprene was seen to have the overall lowest increase in  $F_0$ , increasing ~~with 89 %~~ ~~with 0.1-fold~~ from healthy to  
infested. A significant difference was found for the  $Q_{10}$  coefficients for healthy and infested trees ( $p < P_0 < 0.032$ )  
as well as for  $F_0$  ( $p < P_0 < 0.0106$ ).

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There were four compounds for which the requirements for the calculations were only fulfilled for infested trees.  
Those were eucalyptol, tricyclene, (1R)-camphene and (+)-sabinene, for which an increase or comparison  
between healthy and infested trees cannot be made, but this might indicate that these compounds could be limited  
to emissions from infested trees only.

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### 3.6 Calculated ~~seasonal constitutive~~ BVOC emissions from healthy Norway spruce bark over the season and comparison with needles/leaf emissions

~~The calculated Leafneedle emissions for over the growing season of 2019 in Hyltemossa varied between~~  
~~found to vary an average with an average of around 60 to 170  $\mu\text{g m}^{-2} \text{h}^{-1}$  for leafneedle MT in July and August and~~  
715 an average of around 25 to 100 in May, and 50 to 120  $\mu\text{g m}^{-2} \text{h}^{-1}$  in May and September, ~~respectively~~ (Fig. 11b).  
Bark ~~emissions have been~~ emissions were based on measured tree trunk temperature at 3m agl, averaged from 2  
directions (North and East) and the average standardized emissions ( $M_s$ ), 29  $\mu\text{g m}^{-2} \text{h}^{-1}$  for MT and 2  $\mu\text{g m}^{-2} \text{h}^{-1}$   
for SQT, and the  $Q_{10}$  approach for healthy trees. The calculated emission rates from bark reached a maximum  
around 16  $\mu\text{g m}^{-2} \text{h}^{-1}$  in July, which is ten times lower than the calculated leafneedle emissions at the same time  
720 (Fig. 11c). The bark emission rates remained below 10  $\mu\text{g m}^{-2} \text{h}^{-1}$  for most of the growing season. The estimated  
bark emission rates from healthy trees using the  $Q_{10}$  approach were generally about 5  $\mu\text{g m}^{-2} \text{h}^{-1}$  lower than the  
calculated emission rates using the G93 approach, but steeply increased during the warmest days to match the  
G93-emissions (Fig. 11c).

725

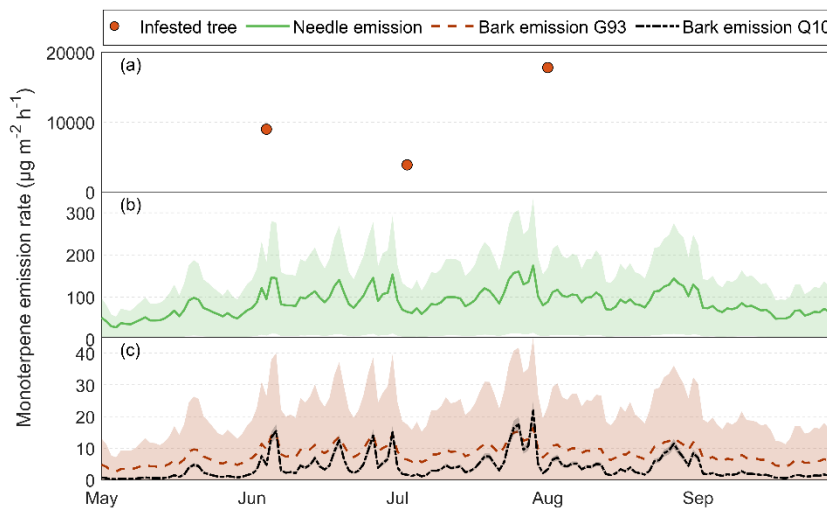
For SQTs, the leafneedle emissions peaked at 30  $\mu\text{g m}^{-2} \text{h}^{-1}$  in late July at the same time as when MT emissions  
were high, and also showed emissions up to 20  $\mu\text{g m}^{-2} \text{h}^{-1}$  earlier in June (Fig. A2b). For most of May and  
September, SQT emissions from needles/leaves were calculated to be below 5  $\mu\text{g m}^{-2} \text{h}^{-1}$ . Bark emissions of SQT  
for healthy trees estimated with G93 were well below 1  $\mu\text{g m}^{-2} \text{h}^{-1}$  throughout the season (maximum 0.75  $\mu\text{g m}^{-2}$   
 $\text{h}^{-1}$  in late July), and below 0.3  $\mu\text{g m}^{-2} \text{h}^{-1}$  most of the time (Fig. A2c).



However, when comparing estimated bark emission from healthy trees with actual measurements of infested trees, bark emissions from infested trees were much higher (Fig. 11a). The measured bark MT emission rate from the infested tree reached up to around 18,000  $\mu\text{g m}^{-2} \text{h}^{-1}$  as a daily average for one day, making the total MT emission rate (including leafneedle emissions) increase by almost a 100-fold when the tree was infested. The lowest measured infested tree emission rate (around 3,900  $\mu\text{g m}^{-2} \text{h}^{-1}$ ) for MT was found during the July campaign, however, despite it being the lowest, it was still considerably higher than the constitutive MT emission rate from healthy trees of that day (70  $\mu\text{g m}^{-2} \text{h}^{-1}$  including both needle and bark; Fig. 11), which was calculated to be around 70  $\mu\text{g m}^{-2} \text{h}^{-1}$  including both leafneedle and bark VOC BVOC emission. Bark MT emission rates from the bark of the infested trees were at least around 55 times higher than the total MT emission rate from both needlesleaves and bark of a healthy tree.

~~Emission rates from the bark of the infested trees were about 55 times higher than the total constitutive MT emission rate from both leaves and bark, even at the lowest measured emission rates.~~

For the SQT emission rates, the difference was not as distinct. The SQT emission rate reached maximum emission rate of around 0.75  $\mu\text{g m}^{-2} \text{h}^{-1}$ , for bark emissions (Fig. A2b) and around 30  $\mu\text{g m}^{-2} \text{h}^{-1}$  for leafneedle emission (Fig. A2a) while the measured bark emission rate from the infested tree peaked around 40  $\mu\text{g m}^{-2} \text{h}^{-1}$  as a daily average for one day, indicating a 1.3-fold increase when a tree is infested. The lowest measured infested tree emission rate was also in July for the SQTs, at around 1.4  $\mu\text{g m}^{-2} \text{h}^{-1}$ , which was still higher than the calculated healthy constitutive bark emission rate at around 0.2  $\mu\text{g m}^{-2} \text{h}^{-1}$ , but lower than the constitutive leafneedle emission rate of about 5  $\mu\text{g m}^{-2} \text{h}^{-1}$ .



**Figure 11. The measured and calculated BVOC emission rates for the group monoterpenes from Norway spruce in Hyltemossa: (a) the actual measured emission rates from infested Norway spruce bark (red dot), (b) calculated needle emission rates (green line) constitutive and (c) calculated healthy bark VOC emission. The needle emissions were calculated based on the Guenther algorithm (Guenther et al., 1993) and the measured emission rates were taken from van Meeningen et al., (2017), and specific needle area (SLA) was taken from Wang et al., (2017) and, using the air temperature at 24 m was taken from the HTM ICOS station (Heliasz, 2020). The healthy bark emission is calculated emission rates from the group monoterpenes based on the tree temperature taken at 3 meters height in the northNorth and eastEast orientation, data taken from the HTM ICOS station (Heliasz, 2020). The bark emissions are calculated using (black) is calculated based on the Guenther algorithm ((Guenther et al., 1993); orange) and the Q<sub>10</sub> temperature**

dependency (black) (orange) based on measured emission rates in this study (black). The shaded areas (green, shade and orange and black shade) represent the standard deviation from the mean for the respective calculation method. The (b) leaf emission rates (green) are calculated based on the Guenther algorithm and the measured emission rates are taken from van Meeningen et al., (2017) and specific leaf area (SLA) was taken from Wang et al., (2017), using the air temperature at 24 meters taken from the HTM ICOS station (Heliasz, 2020). For comparison, (a) the actual measured bark VOC emission rates from one infested tree over time, from this study, is included in the Figure (red dot).

## 4 Discussion

### 4.1 Constitutive and induced bark BVOC emissions from healthy or infested Norway spruce

Both emission rates and composition blend of bark BVOCs from Norway spruce trees were found to change when infested by spruce bark beetles, which is in line with previous studies on bark beetle infestation of conifer trees (Amin et al., 2013; Birgersson and Bergström, 1989; Ghimire et al., 2016; Heijari et al., 2011). In this study, we identified 29 compounds unique to infested trees were identified from which where the majority were MTs (n = 19; Table A1). Several of the identified compounds were only emitted from infested trees trees, for example, eucalyptol, isodene, (+) camphor, tricyclene,  $\alpha$ -phellandrene, which is were consistent with the findings of Ghimire et al. (2016), for example, eucalyptol, isodene, (+)-camphor, tricyclene,  $\alpha$ -phellandrene (Table A1). The findings in this study also show isoprene emission was also found to be emitted from both healthy trees and infested trees healthy and infested tree barks which initially, which was believed to originate from potential lichen cover as a study by (Zhang-Turpeinen et al., (2021), however, found a positive correlation between isoprene and lichen cover. However, when visually evaluating bark photos for algae and lichen coverage there is was no clear indication that higher coverage coincided with isoprene emissions ere seemed not to be a relationship when assessing the lichen cover from bark photographs, making the origin of the it uncertain if the isoprene emission did originate from the bark or not. emission uncertain. This is also consistent with the study by Ghimire et al. (2016) also assessed the lichen coverage and isoprene emission and their results are consistent with this study, in which they did not find any statistically significant relationship with isoprene emission from bark and lichen or algal cover.

With regard to the quantity of BVOC emission however, (The results of this study indicate a much greater difference increase in BVOC between emission rates from healthy trees and infested trees healthy and insect infested conifer trees compared to than previous findings (Amin et al., 2013; Ghimire et al., 2016; Heijari et al., 2011). In this study, We found the total Norway spruce constitutive bark BVOC emission rates was found to be from healthy trees to have a seasonal average of  $31.89 \pm 51.67 \mu\text{g m}^{-2}\text{h}^{-1}$  (mean  $\pm$  standard deviation), while the infested trees had an average of  $6630 \pm 6740 \mu\text{g m}^{-2}\text{h}^{-1}$  (mean  $\pm$  standard deviation) for the early season and  $1950 \pm 1350 \mu\text{g m}^{-2}\text{h}^{-1}$  for the late season. This implies that bark from infested trees emits 63 to 215-fold times higher when a Norway spruce was infested more BVOC to the atmosphere than compared to healthy trees, depending on how long the infestation had been ongoing (Fig. 6) where the emissions are higher earlier and decrease with time. Previous findings reported have found increases in emission rates up to 3-fold when Engelmann spruce (*Picea engelmannii*) was infested by spruce beetles (*Dendroctonus rufipennis*) (Amin et al., 2013), up to 10-fold when Scots pine (*Pinus sylvestris*) was infested by weevils (*Hylobius abietis*) (Heijari et al., 2011) and The measured emission rates in this study are 3 to 9 times higher than the emission rates in the study by Ghimire et al. (2016) in which they found up to 15-fold increase of emission rates from all BVOCs when comparing healthy Norway

spruce trees with trees infested by the spruce bark beetle (Ghimire et al., 2016). ~~The measured emission rates in this study are still 3 to 9 times higher than the emission rates in the study by Ghimire et al. (2016) which had the highest increase comparing healthy and infested Norway spruce bark. As the emission rates in the study by Ghimire et al. (2016) are also standardized according to G93, temperature is not an-should impacting factor on the emission rates to create this large difference. -A possible reason might however be for the difference in measured emission rates might be the time since of the infestation start in relation to the measurements. In our study we found an~~ An exponential decay in emission rates over time after infestation were found in this study (Fig. 8), suggesting that if measurements were taken early during the infestation the emission rates would be higher. However, the ~~measured~~ emission rates ~~in this study~~ from exit holes (measured after 100 days) during the late season ~~in this study are~~ still higher than the emission rates found in June for Ghimire et al. (2016) originating from unspecified hole type. As they did not specify how long time the infestation had been ongoing it makes a comparison difficult as their June measurement could have originated from older infestations. ~~In this~~ study, ~~we conducted~~ measurements were taken throughout the growing season, from the same spruce, starting before the bark beetle infestations ~~and. This allowed us to capture from the emission rates from very~~ early infestation to later stages, something that, to our knowledge, has not been done before. Birgersson and Bergström (1989) did measure volatiles emitted from entry holes in bark beetle infested Norway spruce during the first week of infestation, but not longer. This finding makes it very important to consider the influence of time since infestation when modelling emission rates of BVOCs from infested Norway spruce. We could see a trend with exponential decrease in emission rates with time for both the total BVOCs as well as selected compounds. The spruce bark beetles typically have a first swarm in May, followed by a sister brood in June and the initiation of a second generation in July (Jönsson et al., 2009). The trend with decreased emission rates we found is only related to the start of infestation, regardless of the time in the season and from which swarming period the infestation started.

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The results from this study found eEmission rates from infested Norway spruce bark were found to decrease with time, similarly to what (Eller et al., (2013) found when piercing holes in needles from Ponderosa pine (*Pinus ponderosa*). They found MT emission rates to increase by four orders of magnitude when the needle was pierced compared to undamaged needles and as the exposed resin hardened, the emission rates decreased exponentially until they reached similar levels as undamaged needles already after 30 days. In this study, induced we found increased emission rates from infested Norway spruce bark were seen to last up to 300 days before reaching similar levels as healthy bark, indicating the death of the tree (Fig. 8). This would suggest that the increased emission rates and their exponential decay over time is not only due to exposure of resin occurring as the spruce bark beetle first drill a hole, but also due to the ecological impact developmental process of the beetle and indicated by the visibility of the different hole types created. This is supported by the fact that The timing of the infestation plays an important role in the increased emission rates. wWhen comparing the total number of holes to emission rates for the individual spruce Norway spruce trees, there was no distinct pattern was found (Fig. 9). The hypothesis was that emission rates would be higher at the start of infestation and decrease over time with declining vitality of the trees, which could be explained by a relationship between emission rates and the type of bark beetle hole rather than the total number of holes. However, when the holes were separated When separating the holes into entry and exit, it was apparent that entry holes generally have higher emission rates compared to exit holes (Fig. 9). -which This can also be supported by the result indicating higher emission rates at the start of an infestation the relation to time since infestation where the emission rates were highest in the beginning (Fig. 8) when there is a majority of entry holes (Table 2) and entry holes generally appear at the start of an infestation. For the total VOCs- BVOCs there is a large spread in emission rates, and the second highest emission rate came from an individual (tree ID S3S2) at

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plot 3 in HTM that ~~with had~~ less than 3200 holes per square meter (Table 2); ~~and which was determined to have~~ mainly entry holes (Fig. 9). Some of the lowest emission rates came from ~~an individual an individual (tree ID S1S1)~~ with more than 1000 holes per square meter with mainly exit holes (Table 2). The same is true when looking at the compounds  $\beta\beta$ -phellandrene, eucalyptol and  $\beta\beta$ -pinene (Fig. 9). ~~The importance of hole type could be further~~  
850 ~~supported when the emission rates were scaled with temperature only and with additional scaling from the~~  
~~averaged bark beetle holes from our study (Fig. 7). Comparing all trees, there is a larger difference in emission~~  
~~rates comparing the two scaling approaches from tree S3S2 and S3S3, which both had a majority of entry holes.~~  
~~Based on~~ This result supports, the speculation is that the high emission rates are linked to the time since infestation  
855 ~~start, rather explained by the hole type than the total amount of holes. This indicates that when only taking~~  
~~temperature into account, the emission rates were lower, but when all trees had been scaled to have the same~~  
~~number of bark beetle holes, the difference in emission rates is more distinct – especially from the two trees with~~  
~~a majority of entry holes. This supports the conclusion that high emissions are rather explained by the type of~~  
~~holes, where entrance holes are more relevant than exit holes, than the total amount of holes.~~

860 ~~The number of identified BVOC compounds found to be emitted from entry holes were higher compared to exit~~  
~~holes (Table 3), which further supports the assumption that the emission rates are due to ecological impacts of the~~  
~~spruce bark beetles and not only exposed resin. To further support the speculation that the increased emission rates~~  
~~are due to ecological impacts of the spruce bark beetles and not only exposed resin, the number of identified BVOC~~  
~~compounds found emitted from entry holes were higher compared to exit hole (Table 3)~~The number of identified  
865 ~~compounds was also higher in emissions from entry holes.~~ This is consistent with Birgersson and Bergström (1989)  
who looked at volatiles emitted from entry holes in bark beetle infested ~~spruce~~Norway spruce. They did however  
not look at exit holes, but their findings show that during the early stage of an attack, the MT emissions are high  
and the concentration of the collected MTs during the first day is consistent with what ~~wase~~ found in ~~this our~~ study  
12 days after infestation. Two oxygenated ~~MTs-compounds~~ were found only from the entry holes which is  
870 consistent with the findings of Birgersson and Bergström (1989) and indicates emissions from the phloem. The  
bark beetle pheromones: germacrene D, isodene,  $\beta\beta$ -cubebene and 2-methyl-3-buten-2-ol (~~MB, not quantified~~)  
were also only found emitted from entry holes ~~in this study~~ – however, only from one tree (~~tree ID S3S2~~) and ~~one~~  
~~day throughout the study only during the June campaign (Fig. 10d). The presence of MB~~This could be ~~indications~~  
~~indicative~~ of beetles ~~present~~ in the galleries during the measurements of that individual tree (Birgersson et al.,  
875 1988; Zhao et al., 2011a). ~~The high increase in emission rates couldan also be an impact-a result fromof bark beetle~~  
~~associated blue-stain fungus (*E. polonica*) as a study by Mageroy et al., (2020) found that inoculation withof the~~  
~~fungus in Norway spruce bark was shown to increase the concentration of total terpenes 91-fold after 35 days~~  
~~compared to the concentration in healthy bark. They did however not measure the emission rates from the bark as~~  
~~done in this study, but their finding of high terpene concentration coincides with the high emission rates from~~  
880 ~~infested Norway spruce bark found in this study.-~~

#### 4.1.1 Indications of differences in emissions from healthy or stressed trees during infestation

~~At plot 3 in HTM we selected trees~~Two Norway spruce trees with ~~various-different~~ health status ~~were selected at~~  
~~plot 3 in HTM~~ to study if the health status might have an impact on induced tree emissions from bark beetle  
infestations. ~~Prior to the infestation, we found that the emission rates were different; one tree (ID S3S3) had lower~~  
885 ~~emission rates compared to the other (ID S3S2) which had significantly higher emission rates (Fig. 10a,b). The~~  
~~higher emission rates from S23S23 can be a sign of stress (Loreto and Schnitzler, 2010) and was not only visible~~  
~~in the BVOC emissions, but also from resin flow on the bark, supporting a theory that the high emission rates were~~

caused by a late summer attack from spruce bark beetles during the previous season. In June, the two trees were subjected to spruce bark beetle infestation, S3S3 for the first time and S3S2 for the second time and. Prior to the infestation, we found that emission rates for two spruces were different; one emitted significantly more BVOCs compared to the other, a state which was indicating stress (Fig. 10). This was not only visible in the BVOC emissions, but also from resin flow on the bark, supporting the claim that the high emission rates were from stress, something which was caused by a late summer attack from spruce bark beetles during the previous season. Despite the old infestation, the stressed tree was attacked by spruce bark beetles again during our measurement period. When both trees had been infested by bark beetles, and were 12 days into the infestation they both trees showed induced emission rates (Fig. 10c,d). There was no significant difference in their respective emission rates; however, the trees were emitting slightly different compound blends, which might have been caused by from the initial status of the trees. Tree S3S3 he initially healthy tree showed induced emissions of mostly MTs mostly induced MTs where of mainly (1S)-camphene,  $\beta\beta$ -phellandrene,  $\alpha\alpha$ -pinene and 4-carene were dominating the emissions, but there were also emissions of  $\zeta$ -fenchene, trans,trans(4E,6E)-alloocimene, norbornane and  $\alpha\alpha$ -thujene which were not found to be emitted from the initially stressed tree tree S3S2. The initially stressed tree Tree S3S2 was emitting several of the same MTs as S3S3 with the majority being emissions of  $\beta\beta$ -phellandrene and  $\beta\beta$ -pinene. But in addition to this, the the previously mentioned bark beetle pheromones were found in the samples taken 12 days after infestation from this tree previously mentioned being emitted once came from this initially stressed tree, and was only found 12 days into the infestation. The pheromones could indicate that there was ongoing blue stain fungi infection caused from the infestation that happened in the previous season or that the tree already had successful bark beetle infestation again (Birgersson et al., 1988; Zhao et al., 2011a). Previous studies have found that priming Norway spruce bark with methyl jasmonate (MeJA) as well as inoculating with blue-stain fungi have increased the spruce defense towards spruce bark beetles (Mageroy et al., 2020; Zhao et al., 2011a). As tree S3S2 had survived a previous infestation, a speculation can be that the tree also had it also had fungi present during the new attack. However, the tree defense defense for this tree was not found to be increased but it was rather found that the spruce bark beetles easily overtook it evident by as there was visible entry holes entry holes present, something which was found to be lower for trees previous studies have indicated should be lower compared to healthy trees when they are primed with MeJA or inoculated with fungi compared to healthy trees or inoculated beetles (Mageroy et al., 2020; Zhao et al., 2011a). Additional other evidence for lowered defense of tree S3S2 is a comparison to the initially healthy tree's (S3S3) in regards to BVOC emission rates (Fig. 10): something that could be supported as the tree was cut down at the end of the measurement period. As the infestation continued, the emission rates from the already stressed spruce tree S3S2 were not significantly lower than from compared to the initially healthy spruce S3S3, an indication of decreased vitality of spruce S3S2 while tree S3S3 had induced emission rates until August indicating ongoing defense, which continued at high emission rates until August. The However, the last measurement in August did however revealed occurrence of verbenone from the initially healthy tree tree S3S3, which have been found to be emitted with successful fungal establishment and has been shown to repel bark beetles (Bakke, 2009; Cale et al., 2019). The findings of verbenone could indicate that the bark beetles had successfully overtaken the spruce Norway spruce S3S3 in August, however, this could not be confirmed as the forest owner had to take down the trees which made further measurements impossible. As bark beetle outbreaks have been seen to increase in number, there might be an increase in the number of healthy trees being attacked and killed in addition to the typical attacks on already stressed trees (Jakoby et al., 2019). The results from this study revealed different blends of compounds when a tree was already stressed from previous infestation and attacked again compared to when the tree was healthy before the attack. Another important note on this is that the previously attacked tree (S3S2) indicated In addition to this, we could also see that the one with

~~previous infestation indicated~~ induced emission rates until the start of the next season (Fig. 10a,b), during which the tree was infested again with further induced emissions. The healthy tree (S3S3) did however have ~~higher induced~~ emission rates for longer when it was infested ~~compared to the stressed ones~~ healthy compared to when tree S3S2 was infested again. These results indicate that the second generation of spruce bark beetles, ~~attacking late during the growing season,~~ might lead to induced emission rates continuing until the next ~~season/year.~~ ~~If a tree attacked late during the growing season survives, it can be attacked again the next season along with attacks on healthy trees. This might indicate that when new attacks can occur on the same tree, as well as on healthy trees. The increased attacks on healthy trees as well as~~ the initiation of a second generation of spruce bark beetles ~~and the attack on healthy trees~~ might have a larger impact on the total bark BVOC emission rates from Norway spruce ~~where they are induced for longer.~~

The high BVOC emission rates from infested trees ~~does~~ not only affect the trees themselves, but ~~the emissions ultimately they~~ also impacts atmospheric processes. Induced emission rates ~~from~~ BVOCs due to insect herbivory have been found to potentially increase SOA yields when modelling an increase in emission rates (Bergström et al., 2014). ~~This which~~ would support the ~~speculation claim~~ that bark beetle induced BVOC emission rates ~~could bear~~ is important to consider when modelling or measuring SOA formation. Taking not only the quantitative aspects of bark beetle induced emission rates ~~into account,~~ but also the qualitative effects ~~into account,~~ the SQTs  $\alpha$ -humulene, longifolene and  $\beta\beta$ -caryophyllene ~~and MT  $\alpha$ -pinene~~ have been found to have highest SOA yields ~~compared to 16 other BVOCs, where  $\alpha$ -pinene had the 9<sup>th</sup> highest SOA yield~~ (Lee et al., 2006). ~~This study found~~ ~~In our study we found~~ ~~increased emission rates from~~ of longifolene+ $\beta\beta$ -caryophyllene (quantified together) ~~of around 54- and 20-fold from emission rates at around 5300 % and 1850 % from infested trees depending on the time since infestation~~ in the early and late season ~~was found in this study~~ (Table 2/3). ~~This could lead to potentially increased SOA yields when forests are subjected to bark beetle outbreaks.~~ The MTs limonene and myrcene were slightly below the SQTs in ranking of SOA yield, and according to ~~the our~~ findings ~~in this study,~~ they were seen to increase with an average of ~~4400 to 16900 %~~ 50- to 170-fold and ~~1900 to 49900 %~~ 30- to 530-fold respectively, depending on ~~time since infestation~~ the season, where the highest increase was in the early season. ~~Myrcene had the third highest percentage increase for the seasonal average of all compounds, and appeared in around 80 % of all samples from infested trees, compared to about 20 % from the healthy tree samples.~~ This change in compound blend could potentially lead to large impacts on SOA yield from bark beetle infested trees overall ~~and highlights the importance of measuring and accounting for bark BVOC emissions.-~~

In the comparison of the initially stressed tree and the initially healthy tree, it was apparent that there ~~originally~~ initially was no significant difference in the total emission rate, but different compound blends were emitted. Linking this to SOA, the higher emissions of limonene and myrcene from the initially healthy tree early during the infestation indicates that higher SOA yields might come from healthy ~~spruce~~ Norway spruce trees when infested. The high emission rates were also seen to continue until the trees were taken down in August ~~which implies~~ implying that potential increase in SOA yield might continue for longer. Compounds unique to both infested trees were emitted as well, where the stressed tree emitted bark beetle related pheromones and the initially healthy tree emitted a broader blend of MTs, these individual compounds might play a role in the SOA yield as well. ~~An increase of~~ ~~As the bark beetle infested trees generally would impact SOA yields, more~~ attacks on healthy trees might further affect the atmospheric processes, specifically production of SOA.

#### 4.2 Bark beetle induced BVOC emissions in relation to other stresses, ~~leaf~~needle emissions and modelling

Significant increases of the temperature standardized BVOC emissions of Norway spruce bark ~~of up to around 22000 % for the total group of MTs~~ were seen when trees were infested by bark beetles early in the season. Around a 230-fold increase in emission rate was seen for the group of MTs. This high increase in emission rate from insect stress for Norway spruce has not previously been observed according to the review by Yu et al. (2021), in which the highest recorded increase was around 2,000 %, including previous studies on spruce bark beetles, ~~Ips typographus~~. Heat stress ~~from higher temperatures~~ was also identified as a stressor but it ~~did not~~ increase BVOC emissions as much as stress from bark beetles (Yu et al., 2021). A study on Norway spruce with higher air temperatures of 40 °C found that BVOCs increased by 175 % compared to emission rates at air temperature of 30°C (Esposito et al., 2016). ~~This is not close, not comparable much lower than to~~ the increase found in emission rates from bark beetle infestation found in this study. However, the impact of combined stresses from temperature and insect attacks might further increase the BVOC emissions. This is illustrated by the increase of reference emission rates after infestation across all BVOCs ( $F_0$ , Table A2) which are standardized at a reference temperature and therefore the difference between  $F_0$  from healthy to infested trees could be interpreted as the change in stress induced BVOC emissions due to bark beetle attacks without the temperature effect. In our study we found the The temperature sensitivity as expressed by the  $Q_{10}$  coefficient was also found to increase for all compounds but one change when trees become infested. Our analysis of the  $Q_{10}$  coefficient showed that the coefficient increased for all compounds but one (of 15, namely p-cymene, Table A2) when trees became infested, indicating that the emission rates were accelerated by both a higher reference emission rate and an accelerated temperature response compared to healthy trees, as temperatures increased, indicating that the emission rates would increase if the temperature increased. The combination of bark beetle stress with increased temperature might thus lead to even higher increase in emission rates. This is however not the focus of this study, but as ~~we found~~ the BVOC compounds temperature sensitivity was found to increase when trees were infested, and as ~~we found~~ bark beetle infestations to increase bark BVOC emissions more than any other comparable stress, there might be high influences of BVOC emissions from combined stress, making it important to account for when modelling the emissions.

The increase of bark emission rates ~~we seen~~ found in this study is high enough to considerably add to the emission rates of a full tree when comparing with emission rates from needles – which is considered the part of the tree with the highest emission rates. When modelling the emission rates, two approaches were used for the bark MT emissions, the G93 algorithm and the  $Q_{10}$  approach. The results showed similarities in pattern but the  $Q_{10}$  approach had larger increases in emission rates with higher temperature increase, something that was expected. The G93 modelled emission rates were constantly higher than the  $Q_{10}$ , and had less variability, which might be explained by the empirical coefficients used in the G93 compared to ~~only taking temperature into account~~ fitting  $F_0$  and  $Q_{10}$  for each compound separately. For the needle emission rates, only G93 was used because of the light dependent nature of some BVOCs emitted from the needles that could not be explained by ~~the~~ temperature in the  $Q_{10}$  approach only. The seasonal average emission rates from ~~spruce~~ Norway spruce needles were measured during 2017 ~~6~~ in Hyltemossa in a study by van Meeningen et al. (2017). As the study was conducted at the same site, their results were applied to ~~this~~ our study as a comparison of bark BVOC emission to needle emission. It was clear that ~~constitutive~~ MT emissions from healthy bark does not compare to the needle emission (Fig. 11), where the seasonal average of the MT emissions were 11 times higher from needles than bark. However, when comparing seasonal average bark MT emission from infested trees with needle emissions, it was the other way around, ~~the~~ the bark emissions from infested trees were 6 to 20 times higher than the needle emissions depending on the time of

1015 season. The ~~bark constitutive~~-MT emission from ~~bark of~~ healthy ~~spruce~~Norway spruce trees accounted for 8 % of  
the total emission rates from bark and needles. However, if there ~~was~~were an ongoing infestation from bark beetles,  
the bark emission rates would account for 95 % of the total emission rates during the early season, and 85 % during  
the late season. ~~When comparing with the seasonal average of the emission rates from healthy trees, spruce bark  
beetle infestation could~~The infestation would lead to a 6- to 20-fold increase ~~in the~~ total emission rates from bark  
1020 and needles ~~by 550 % to 1,900 %, depending on the time in the season, when comparing with the seasonal average  
of the constitutive emission rates from healthy trees.~~

When a tree is infested, the emission rates increase significantly which can cause large local effects both for tree  
health ~~but also~~and SOA production. The BVOC emission increase ~~can~~ also cause more widespread effects, ~~if the~~  
1025 outbreaks are sustained at high levels, there would be large impacts regionally. During 2020 in Sweden, 8 million  
m<sup>3</sup> forest was affected by spruce bark beetles (Wulff and Roberge, 2020). This represents about 0.7 % of the total  
volume of Norway spruce trees with a diameter larger than 15 cm in Sweden (Skogsstyrelsen, n.d.). Using the  
seasonal average from the early ~~season~~and late season of the bark beetle infested emission rates of MT found in  
this study and the needle emission rates from van Meeningen et al. (2017), the infested trees during 2020 would  
1030 contribute to an increase of about 4 to 13 % of total MT emission rate from Norway spruce trees in ~~all of~~Sweden,  
including emissions from canopy and stem. The effects from insect herbivory and specifically spruce bark beetles  
might thus be underestimated both in emission and vegetation models (MEGAN, LPJ-GUESS; Guenther et al.,  
2006; Schurgers et al., 2009) and atmospheric chemistry models estimating BVOC impacts on oxidation capacity  
and SOA formation (ADCHEM; Roldin et al., 2011). ~~Evident by the difference in emission rates during the early  
or late season, it is also important to consider the influence of time after infestation when modelling emission rates  
of BVOCs from infested Norway spruce to get a correct estimation of the spruce bark beetle impact.~~

## 5 Conclusion

Norway spruce trees are emitting BVOCs from the bark as a stress response to spruce bark beetles, and as the  
number of spruce bark beetle outbreaks increase, it will impact the total emission of BVOCs. The aim of ~~the our~~  
1040 study was to examine how spruce bark beetles affect the BVOC emission rates from Norway spruce bark by  
looking at the difference between healthy and infested trees, the time ~~passed after since~~infestation ~~start~~ and the  
difference in emissions from different bark beetle ~~drilled~~ holes ~~types~~. ~~We also wanted~~One aim was also to provide  
an insight into how the BVOC emissions change from non-infested to infested, and following the infestation over  
time. ~~The Our results study~~ shows that there is a significant difference in BVOC emission rates from healthy ~~spruce~~  
1045 ~~bark~~and infested ~~spruce~~Norway spruce bark, but also a relationship between BVOC emissions from infested trees  
and the time ~~passed since~~after infestation start, which can be supported by ~~the our~~ results ~~indicating that indicated~~  
a difference in emissions from bark beetle drilled ~~entry holes and exit holes~~entry and exit holes. ~~We also saw that~~  
The initiation of a second generation of bark beetles, which can lead to late summer attacks, can potentially have  
prolonged impacts on the BVOC emissions as ~~we found~~ emission rates ~~were found~~ to be induced until the start of  
1050 the next season. When the tree was infested again, the emission rates ~~were~~as further induced to reach the same  
levels as the induced emissions of a tree that was healthy before infestation. As the infestation proceeded, ~~there~~  
~~was~~we saw a difference in the emission rate and compound blend when comparing the initially stressed tree with  
the initially healthy tree, where the emission rates were induced to high levels until August for the initially healthy  
tree, but not for the initially stressed tree. ~~Further studies are needed to support the findings and speculations of  
this study but also to analyze t~~The entire impact of spruce bark beetles on Norway spruce trees, ~~would require~~



1060 ~~further studies, T-and~~ the importance of ~~further~~~~such~~ studies is supported by ~~the~~~~our~~ findings that the bark beetle  
induced BVOC emission rates can be considerably higher than previously thought and could potentially ~~increase~~  
~~the~~~~lead to a 1.1-fold increase of~~ total MT emissions from Norway spruce in Sweden ~~with 4 to 13 %~~. Even further  
work would be needed in investigating the impact of coupled stress factors. ~~We found a~~~~A~~ potential link between  
1065 temperature stress and bark beetle stress ~~was identified in this study~~, where trees ~~seems~~ to become more sensitive  
to temperature ~~leading to~~~~with a potential to have even~~ ~~potentially~~ higher emission rates when temperatures  
increase in conjunction with bark beetle infestations. ~~We believe that~~~~Based on the findings of this study~~, bark  
beetle infestations ~~are believed to~~~~can~~ have higher impacts on the atmosphere and climate change than previously  
thought and samples from more trees and more frequently throughout the season ~~are~~~~is~~ needed in order to fully  
1065 understand the impact.

Appendix A

1070 **Table A1. Seasonal average temperature standardized emission rate ( $\mu\text{g m}^{-2} \text{h}^{-1} \pm$  one standard deviation) from all**  
**Norway spruce trees located in Hyltemossa and Norunda. Presented are all the identified compounds, compound**  
**groups (Monoterpenes, sesquiterpenes and other BVOCs) and total emission emitted from healthy and infested (early**  
**and late season) Norway spruce bark. The increase or decrease (%) is presented for the infested trees as a change in**  
**emission rate from healthy to infested. The occurrence (%) is presented to indicate how much indicates how often each**  
1075 **compound appeared in the samples. The compounds that were identified but unable to quantify is presented as n.q. (no**  
**quantification). All identified compounds from all samples throughout the season separated into healthy trees and**  
**infested trees. Presented is the seasonal average emission rate ( $\mu\text{g m}^{-2} \text{h}^{-1} \pm$  one standard deviation) for each compound**  
**and the groups of MT, SQT and other BVOCs and the occurrence (%) of each compounds for all samples. The increase**  
1080 **(%) is presented for the infested trees as an increase from healthy to infested. The compounds that were identified but**  
**unable to quantify is presented as n.q. (no quantification).**

-	Healthy		Infested early season			Infested late season		
Compound name	<u>average <math>\pm</math> std</u> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<u>occurrence</u> (%)	<u>average <math>\pm</math> std</u> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<u>increase</u> (%)	<u>occurrence</u> (%)	<u>average <math>\pm</math> std</u> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<u>increase</u> (%)	<u>occurrence</u> (%)
<b>Monoterpenes</b>	<u>29 <math>\pm</math> 51</u>	-	<u>6,600 <math>\pm</math> 6,700</u>	<u>22,400</u>	-	<u>1,900 <math>\pm</math> 1,300</u>	<u>6,500</u>	-
<u><math>\alpha</math>-pinene</u>	<u>12 <math>\pm</math> 20</u>	<u>76</u>	<u>910 <math>\pm</math> 1030</u>	<u>7,800</u>	<u>100</u>	<u>820 <math>\pm</math> 890</u>	<u>7,100</u>	<u>100</u>
<u><math>\beta</math>-pinene</u>	<u>8 <math>\pm</math> 19</u>	<u>56</u>	<u>950 <math>\pm</math> 960</u>	<u>11,500</u>	<u>100</u>	<u>230 <math>\pm</math> 170</u>	<u>2,600</u>	<u>100</u>
<u>3-Carene</u>	<u>3 <math>\pm</math> 5</u>	<u>49</u>	<u>290 <math>\pm</math> 420</u>	<u>11,400</u>	<u>100</u>	<u>30 <math>\pm</math> 40</u>	<u>1,200</u>	<u>95</u>
<u>Limonene</u>	<u>2 <math>\pm</math> 3</u>	<u>44</u>	<u>320 <math>\pm</math> 320</u>	<u>16,900</u>	<u>88</u>	<u>90 <math>\pm</math> 80</u>	<u>4,400</u>	<u>100</u>
<u>p-Cymene</u>	<u>1 <math>\pm</math> 1</u>	<u>40</u>	<u>240 <math>\pm</math> 320</u>	<u>49,200</u>	<u>63</u>	<u>50 <math>\pm</math> 40</u>	<u>10,700</u>	<u>77</u>
<u><math>\beta</math>-Myrcene</u>	<u>0.3 <math>\pm</math> 0.8</u>	<u>18</u>	<u>160 <math>\pm</math> 170</u>	<u>49,800</u>	<u>79</u>	<u>10 <math>\pm</math> 6</u>	<u>1,900</u>	<u>86</u>
<u><math>\beta</math>-Phellandrene</u>	<u>3 <math>\pm</math> 8</u>	<u>11</u>	<u>670 <math>\pm</math> 660</u>	<u>24,800</u>	<u>44</u>	<u>190 <math>\pm</math> 160</u>	<u>6,900</u>	<u>68</u>
<u>(1S)-Camphene</u>	<u>2 <math>\pm</math> 6</u>	<u>6</u>	<u>1,520 <math>\pm</math> 1,970</u>	<u>89,100</u>	<u>75</u>	<u>390 <math>\pm</math> 230</u>	<u>22,800</u>	<u>100</u>
<u>2-Cyclopentylcyclopentanone</u>	<u>n.q-</u>	<u>4</u>	-	-	-	<u>n.q</u>	-	<u>5</u>
<u><math>\alpha</math>-Terpineol</u>	<u>n.q-</u>	<u>3</u>	-	-	-	-	-	-
<u>5-Ethyl-m-xylene</u>	<u>n.q-</u>	<u>1</u>	-	-	-	-	-	-
<u>(+)-Sabinene</u>	<u>0.1 <math>\pm</math> 0</u>	<u>1</u>	<u>210 <math>\pm</math> 200</u>	<u>257,900</u>	<u>44</u>	<u>3 <math>\pm</math> 0</u>	<u>3,600</u>	<u>5</u>
<u>(1R)-Camphene</u>	-	-	<u>190 <math>\pm</math> 740</u>	-	<u>19</u>	<u>110 <math>\pm</math> 80</u>	-	<u>77</u>
<u>Tricyclene</u>	-	-	<u>170 <math>\pm</math> 250</u>	-	<u>50</u>	<u>20 <math>\pm</math> 20</u>	-	<u>36</u>
<u>Eucalyptol</u>	-	-	<u>10 <math>\pm</math> 20</u>	-	<u>44</u>	<u>2 <math>\pm</math> 5</u>	-	<u>27</u>
<u>(+)-Camphor</u>	-	-	-	-	-	<u>n.q-</u>	-	<u>46</u>
<u>Pinocarvone</u>	-	-	<u>n.q</u>	-	<u>44</u>	<u>n.q-</u>	-	<u>5</u>
<u>4-Carene</u>	-	-	<u>350 <math>\pm</math> 280</u>	-	<u>38</u>	-	-	-
<u><math>\zeta</math>-Fenchene</u>	-	-	<u>120 <math>\pm</math> 190</u>	-	<u>25</u>	<u>1 <math>\pm</math> 0</u>	-	<u>5</u>
<u><math>\alpha</math>-Phellandrene</u>	-	-	<u>110 <math>\pm</math> 200</u>	-	<u>31</u>	-	-	-
<u>(1R)-(-)-Myrtenal</u>	-	-	<u>n.q</u>	-	<u>31</u>	-	-	-
<u>(4E,6E)-Allocimene</u>	-	-	<u>50 <math>\pm</math> 80</u>	-	<u>25</u>	-	-	-
<u>5-Vinyl-m-xylene</u>	-	-	<u>n.q</u>	-	<u>25</u>	-	-	-
<u>3-Pinanone</u>	-	-	-	-	-	<u>n.q</u>	-	<u>14</u>
<u>Norbornane</u>	-	-	<u>60 <math>\pm</math> 80</u>	-	<u>19</u>	-	-	-
<u><math>\gamma</math>-Terpinene</u>	-	-	<u>90 <math>\pm</math> 0</u>	-	<u>13</u>	-	-	-
<u><math>\alpha</math>-Fenchene</u>	-	-	<u>10 <math>\pm</math> 0</u>	-	<u>13</u>	-	-	-
<u>2-Carene</u>	-	-	<u>160 <math>\pm</math> 0</u>	-	<u>13</u>	-	-	-

<u><math>\alpha</math>-Thujene</u>	=	=	<u>20 ± 0</u>	=	<u>13</u>	=	=	=
<u>Verbenone</u>	=	=	<u>n.q-</u>	=	<u>13</u>	=	=	=
<u>Myrtenal</u>	=	=	<u>n.q-</u>	=	<u>6</u>	=	=	=
<u><math>\alpha</math>-Terpinene</u>	=	=	<u>30 ± 0</u>	=	<u>6</u>	=	=	=
<b><u>Sesquiterpenes</u></b>	<u>2.1 ± 3.2</u>	-	<u>53 ± 74</u>	<u>2,400</u>	-	<u>18 ± 24</u>	<u>700</u>	-
<u>Longifolene+<math>\beta</math>-Caryophyllene</u>	<u>0.7 ± 1.4</u>	<u>18</u>	<u>38 ± 70</u>	<u>5,300</u>	<u>56</u>	<u>14 ± 24</u>	<u>1,800</u>	<u>55</u>
<u><math>\alpha</math>-Humulene</u>	<u>1.4 ± 3</u>	<u>48</u>	<u>5 ± 9</u>	<u>200</u>	<u>31</u>	<u>4 ± 13</u>	<u>200</u>	<u>18</u>
<u>Germacrene D</u>	=	=	<u>4 ± 0</u>	=	<u>19</u>	=	=	=
<u>Isoledene</u>	=	=	<u>3 ± 0</u>	=	<u>19</u>	=	=	=
<u><math>\beta</math>-Cubebene</u>	=	=	<u>3 ± 0</u>	=	<u>19</u>	=	=	=
<b><u>Other BVOCs</u></b>	<u>0.4 ± 0.9</u>	-	<u>3.4 ± 6.7</u>	-	-	<u>0.1 ± 0.2</u>	-	-
<u>Isoprene</u>	<u>0.4 ± 0.9</u>	<u>58</u>	<u>3 ± 7</u>	<u>700</u>	<u>63</u>	<u>0.1 ± 0.2</u>	<u>-65</u>	<u>27</u>
<u>Decanal</u>	<u>n.q-</u>	<u>45</u>	<u>n.q</u>	=	<u>13</u>	<u>n.q-</u>	=	<u>14</u>
<u>Benzene</u>	<u>n.q-</u>	<u>45</u>	=	=	=	<u>n.q-</u>	=	<u>14</u>
<u>Nonanal</u>	<u>n.q-</u>	<u>39</u>	<u>n.q-</u>	=	<u>6</u>	<u>n.q-</u>	=	<u>14</u>
<u>Toluene</u>	<u>n.q-</u>	<u>21</u>	<u>n.q-</u>	=	<u>6</u>	<u>n.q-</u>	=	<u>9</u>
<u>1,3,5-Trifluorobenzene</u>	<u>n.q-</u>	<u>14</u>	=	=	=	=	=	=
<u>Benzaldehyde</u>	<u>n.q-</u>	<u>12</u>	=	=	=	<u>n.q-</u>	=	<u>9</u>
<u>Butyl formate</u>	<u>n.q-</u>	<u>8</u>	=	=	=	<u>n.q-</u>	=	<u>5</u>
<u>Caprolactam</u>	<u>n.q-</u>	<u>7</u>	=	=	=	=	=	=
<u>Cyclopentanone</u>	<u>n.q-</u>	<u>5</u>	=	=	=	<u>n.q-</u>	=	<u>9</u>
<u>Methanesulfonic anhydride</u>	<u>n.q-</u>	<u>5</u>	=	=	=	<u>n.q-</u>	=	<u>5</u>
<u>Trimethylbenzol</u>	<u>n.q-</u>	<u>2</u>	=	=	=	=	=	=
<u>m-Xylene</u>	<u>n.q-</u>	<u>2</u>	=	=	=	=	=	=
<u>Ethylhexanol</u>	<u>n.q-</u>	<u>2</u>	=	=	=	=	=	=
<u>Acetic acid</u>	<u>n.q-</u>	<u>2</u>	=	=	=	=	=	=
<u>tert-Butylamine</u>	<u>n.q-</u>	<u>1</u>	=	=	=	<u>n.q</u>	=	<u>9</u>
<u>m-Ethyltoluene</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>o-Ethyltoluene</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Methyl 3-hydroxy-2,2-dimethylpropanoate</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>1-Pentene</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Butanal</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>1-Nonene</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Isobutenyl methyl ketone</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Diacetone alcohol</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Furfural</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>1,6-Anhydro-<math>\beta</math>-d-talopyranose</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>dl-3,4-Dehydroproline methyl ester</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>6,10,14-Trimethyl-2-pentadecanone</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Undecanal</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>Carbon disulfide</u>	<u>n.q-</u>	<u>1</u>	=	=	=	=	=	=
<u>2-Methyl-1-phenylpropene</u>	=	=	<u>n.q-</u>	=	<u>38</u>	<u>n.q</u>	=	<u>9</u>
<u>2-Methyl-3-buten-2-ol</u>	=	=	<u>n.q-</u>	=	<u>19</u>	=	=	=

Benzoic acid	=	=	=	=	=	n.q.	=	9
Acetophenone	=	=	=	=	=	n.q.	=	9
Methyl acetate	=	=	=	=	=	n.q.	=	9
(-)-Bornyl acetate	=	=	n.q.	=	13	=	=	=
Bornyl acetate	=	=	n.q.	=	13	=	=	=

=	Healthy		Infested early season			Infested late season		
	average $\pm$ std ( $\mu\text{g m}^{-2}\text{h}^{-1}$ )	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2}\text{h}^{-1}$ )	increase (%)	occurrence (%)	average $\pm$ std ( $\mu\text{g m}^{-2}\text{h}^{-1}$ )	increase (%)	occurrence (%)
<b>Monoterpenes</b>	<b>29 <math>\pm</math> 51</b>	=	<b>6.600 <math>\pm</math> 6.700</b>	<b>22.400</b>	=	<b>1.900 <math>\pm</math> 1.300</b>	<b>6.500</b>	=
$\alpha$ -pinene	12	76	910	7.800	100	820	7.100	100
$\beta$ -pinene	8	56	950	11.500	100	230	2.600	100
$\beta$ -Carene	3	49	290	11.400	100	30	1.200	95
Limonene	2	44	320	16.900	88	90	4.400	100
p-Cymene	1	40	240	49.200	63	50	10.700	77
$\beta$ -Myrcene	0	18	160	49.800	79	10	1.900	86
$\beta$ -Phellandrene	3	11	670	24.800	44	190	6.900	68
(1S)-Camphene	2	6	1.520	89.100	75	390	22.800	100
$\alpha$ -Cyclopentylcyclopentanone	=	4	=	=	=	=	=	5
$\alpha$ -Terpineol	=	3	=	=	=	=	=	=
5-Ethyl-m-xylene	=	1	=	=	=	=	=	=
(+)-Sabinene	0	1	210	257.900	44	3	3.600	5
(1R)-Camphene	=	=	190	=	19	110	=	77
Tricyclene	=	=	170	=	50	20	=	36
Eucalyptol	=	=	10	=	44	2	=	27
(+)-Camphor	=	=	=	=	=	=	=	46
Pinocarvone	=	=	=	=	44	=	=	5
4-Carene	=	=	350	=	38	=	=	=
$\zeta$ -Fenchene	=	=	120	=	25	1	=	5
$\alpha$ -Phellandrene	=	=	110	=	31	=	=	=
(1R)-(-)-Myrtenal	=	=	=	=	31	=	=	=
(4E,6E)-Allocimene	=	=	50	=	25	=	=	=
5-Vinyl-m-xylene	=	=	=	=	25	=	=	=
$\beta$ -Pinarone	=	=	=	=	=	=	=	14
Norbornane	=	=	60	=	19	=	=	=
$\gamma$ -Terpinene	=	=	90	=	13	=	=	=
$\alpha$ -Fenchene	=	=	10	=	13	=	=	=
$\beta$ -Carene	=	=	160	=	13	=	=	=
$\alpha$ -Thujene	=	=	20	=	13	=	=	=
Verbenone	=	=	=	=	13	=	=	=
Myrtenal	=	=	=	=	6	=	=	=
$\alpha$ -Terpinene	=	=	30	=	6	=	=	=
<b>Sesquiterpenes</b>	<b>2.1 <math>\pm</math> 3.2</b>	=	<b>53 <math>\pm</math> 74</b>	<b>2.400</b>	=	<b>18 <math>\pm</math> 24</b>	<b>700</b>	=
Longifolene+ $\beta$ -Caryophyllene	1	18	38	5.300	56	14	1.800	55
$\alpha$ -Humulene	1	48	5	200	31	5	200	18
Germaacrene-D	=	=	4	=	19	=	=	=
Isoledene	=	=	3	=	19	=	=	=
$\beta$ -Cubebene	=	=	3	=	19	=	=	=

<u>Other BVOCs</u>	<u>0.4 ± 0.9</u>	<u>-</u>	<u>3.4 ± 6.7</u>	<u>-</u>	<u>0.1 ± 0.2</u>	<u>-</u>	<u>-</u>	
<u>Isoprene</u>	<u>0</u>	<u>58</u>	<u>3</u>	<u>700</u>	<u>63</u>	<u>0</u>	<u>-65</u>	<u>27</u>
<u>Decanal</u>	<u>-</u>	<u>45</u>	<u>-</u>	<u>-</u>	<u>13</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Benzene</u>	<u>-</u>	<u>45</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Nonanal</u>	<u>-</u>	<u>39</u>	<u>-</u>	<u>-</u>	<u>6</u>	<u>-</u>	<u>-</u>	<u>14</u>
<u>Toluene</u>	<u>-</u>	<u>21</u>	<u>-</u>	<u>-</u>	<u>6</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>1,3,5-Trifluorobenzene</u>	<u>-</u>	<u>14</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Benzaldehyde</u>	<u>-</u>	<u>12</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>Butyl formate</u>	<u>-</u>	<u>8</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>5</u>
<u>Caprolactam</u>	<u>-</u>	<u>7</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Cyclopentanone</u>	<u>-</u>	<u>5</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>Methanesulfonic anhydride</u>	<u>-</u>	<u>5</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>5</u>
<u>Trimethylbenzol</u>	<u>-</u>	<u>2</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>m-Xylene</u>	<u>-</u>	<u>2</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Ethylhexanol</u>	<u>-</u>	<u>2</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Acetic acid</u>	<u>-</u>	<u>2</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>tert-Butylamine</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>m-Ethyltoluene</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>o-Ethyltoluene</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Methyl 3-hydroxy 2,2-dimethylpropanoate</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>1-Pentene</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Butanal</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>1-Nonene</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Isobutenyl-methyl ketone</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Diacetone alcohol</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Furfural</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>1,6-Anhydro-β-D-talopyranose</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>dl-3,4-Dehydroproline methyl ester</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>6,10,14-Trimethyl-2-pentadecanone</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Undecanal</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Carbon disulfide</u>	<u>-</u>	<u>1</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>2-Methyl-1-phenylpropene</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>38</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>2-Methyl-3-buten-2-ol</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>19</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Benzoic acid</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>Acetophenone</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>Methyl acetate</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>9</u>
<u>(-)-Bornyl acetate</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>13</u>	<u>-</u>	<u>-</u>	<u>-</u>
<u>Bornyl acetate</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>-</u>	<u>13</u>	<u>-</u>	<u>-</u>	<u>-</u>
<b>Total</b>	<b>32 ± 52</b>	<b>-</b>	<b>6.700 ± 6.900</b>	<b>20.900</b>	<b>-</b>	<b>2.000 ± 1.300</b>	<b>6.000</b>	<b>-</b>

-	<b>Healthy</b>		<b>Infested early season</b>			<b>Infested late season</b>		
<b>Compound name</b>	<i>average ± std</i> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<i>occurrence</i> (%)	<i>average ± std</i> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<i>increase</i> <i>se</i> (%)	<i>occurrence</i> (%)	<i>average ± std</i> ( $\mu\text{g m}^{-2} \text{h}^{-1}$ )	<i>increase</i> <i>se</i> (%)	<i>occurrence</i> (%)
<b>Monoterpenes</b>	29.37 ± 51.01	-	6690 ± 6860	22678.35	-	1970 ± 1310	6607.52	-
alpha-Pinene	11.49	76.11	911.14	7829.85	100.00	824.64	7077.02	100.00

beta-Pinene	8.22	55.75	954.17	11507.91	100.00	225.28	2640.63	100.00
3-Carene	2.48	48.67	285.16	11398.39	100.00	33.07	1233.47	95.45
Limonene	1.89	44.25	320.82	16874.60	87.70	85.43	4420.11	100.00
p-Cymene	0.49	39.82	241.45	49175.51	62.50	53	10716.33	77.27
beta-Myrcene	0.32	17.70	159.76	49825.00	78.95	6.3	1868.75	86.36
beta-Phellandrene	2.7	10.62	673.13	24830.74	43.75	189.47	6917.41	68.18
(1S)-Camphene	1.7	6.19	1516	89076.47	75.00	388.82	22771.76	100.00
2-Cyclopentylcyclopentane	n.q.	4.42	-	-	-	n.q.	-	4.54
alpha-Terpineol	n.q.	2.65	-	-	-	-	-	-
5-Ethyl-m-xylene	n.q.	0.88	-	-	-	-	-	-
(+)-Sabinene	0.08	0.88	206.37	25786.250	43.75	2.93	3562.50	4.54
(1R)-Camphene	-	-	186.16	-	18.75	112.32	-	77.27
Tricyclene	-	-	173.42	-	50.00	24.52	-	36.36
Eucalyptol	-	-	10.25	-	43.75	2.32	-	27.27
(+)-Camphor	-	-	-	-	-	n.q.	-	45.45
Pinocarvone	-	-	n.q.	-	43.75	n.q.	-	4.54
4-Carene	-	-	348.15	-	37.50	-	-	-
zeta-Fenchene	-	-	116.96	-	25.00	0.92	-	4.54
alpha-Phellandrene	-	-	114.43	-	31.25	-	-	-
(1R)-(-)-Myrtenal	-	-	n.q.	-	31.25	-	-	-
trans,trans-Allocimene	-	-	50.9	-	25.00	-	-	-
5-Vinyl-m-xylene	-	-	n.q.	-	25.00	-	-	-
3-Pinanone	-	-	-	-	-	n.q.	-	13.63
Norbornane	-	-	60.27	-	18.75	-	-	-
gamma-Terpinene	-	-	88.84	-	12.50	-	-	-
alpha-Fenchene	-	-	14.07	-	12.50	-	-	-
2-Carene	-	-	156.65	-	12.50	-	-	-
alpha-Thujene	-	-	15.83	-	12.50	-	-	-
Verbenone	-	-	n.q.	-	12.50	-	-	-
Myrtenal	-	-	n.q.	-	6.25	-	-	-
alpha-Terpinene	-	-	26.06	-	6.25	-	-	-
<b>Sesquiterpenes</b>	<b>2.12 ± 3.17</b>	<b>-</b>	<b>53.0 ± 74</b>	<b>2400.00</b>	<b>-</b>	<b>18 ± 24</b>	<b>749.06</b>	<b>-</b>
Longifolene+beta-Caryophyllene	0.7	17.70	37.65	5278.57	56.25	13.6	1842.86	54.54
alpha-Humulene	1.42	47.79	4.86	242.25	31.25	4.52	218.31	18.18
Germacrene-D	-	-	3.86	-	18.75	-	-	-
Isodene	-	-	3.25	-	18.75	-	-	-
beta-Cubebene	-	-	3.2	-	18.75	-	-	-
<b>Other BVOCs</b>	<b>0.40 ± 0.85</b>	<b>-</b>	<b>3.36 ± 6.69</b>	<b>740.00</b>	<b>-</b>	<b>0.14 ± 0.20</b>	<b>-65.00</b>	<b>-</b>
Isoprene	0.4	58.41	3.36	740.00	62.50	0.14	-65.00	27.27
Decanal	n.q.	45.13	n.q.	-	12.50	n.q.	-	13.63
Benzene	n.q.	45.13	-	-	-	n.q.	-	13.63
Nonanal	n.q.	38.94	n.q.	-	6.25	n.q.	-	13.63
Toluene	n.q.	21.24	n.q.	-	6.25	n.q.	-	9.09
1,3,5-Trifluorobenzene	n.q.	14.16	-	-	-	-	-	-

Benzaldehyde	n.g.	11.50	-	-	-	n.g.	-	9.09
Butyl formate	n.g.	7.96	-	-	-	n.g.	-	4.54
Caprolactam	n.g.	7.08	-	-	-	-	-	-
Cyclopentanone	n.g.	5.31	-	-	-	n.g.	-	9.09
Methanesulfonic anhydride	n.g.	5.31	-	-	-	n.g.	-	4.54
Trimethylbenzol	n.g.	1.77	-	-	-	-	-	-
m-Xylene	n.g.	1.77	-	-	-	-	-	-
Ethylhexanol	n.g.	1.77	-	-	-	-	-	-
Acetic acid	n.g.	1.77	-	-	-	-	-	-
tert-Butylamine	n.g.	0.88	-	-	-	n.g.	-	9.09
m-Ethyltoluene	n.g.	0.88	-	-	-	-	-	-
o-Ethyltoluene	n.g.	0.88	-	-	-	-	-	-
Methyl 3-hydroxy-2,2-dimethylpropanoate	n.g.	0.88	-	-	-	-	-	-
1-Pentene	n.g.	0.88	-	-	-	-	-	-
Butanal	n.g.	0.88	-	-	-	-	-	-
1-Nonene	n.g.	0.88	-	-	-	-	-	-
Isobutenyl methyl ketone	n.g.	0.88	-	-	-	-	-	-
Diacetone alcohol	n.g.	0.88	-	-	-	-	-	-
Furfural	n.g.	0.88	-	-	-	-	-	-
1,6-Anhydro-beta-D-talopyranose	n.g.	0.88	-	-	-	-	-	-
dl-3,4-Dehydroproline methyl ester	n.g.	0.88	-	-	-	-	-	-
6,10,14-Trimethyl-2-pentadecanone	n.g.	0.88	-	-	-	-	-	-
Undecanal	n.g.	0.88	-	-	-	-	-	-
Carbon disulfide	n.g.	0.88	-	-	-	-	-	-
2-Methyl-1-phenylpropene	-	-	n.g.	-	37.50	n.g.	-	9.09
alpha,alpha-Dimethylallyl alcohol	-	-	n.g.	-	18.75	-	-	-
Benzoic acid	-	-	-	-	-	n.g.	-	9.09
Acetophenone	-	-	-	-	-	n.g.	-	9.09
Methyl acetate	-	-	-	-	-	n.g.	-	9.09
(-) Bornyl acetate	-	-	n.g.	-	12.50	-	-	-
Bornyl acetate	-	-	n.g.	-	12.50	-	-	-

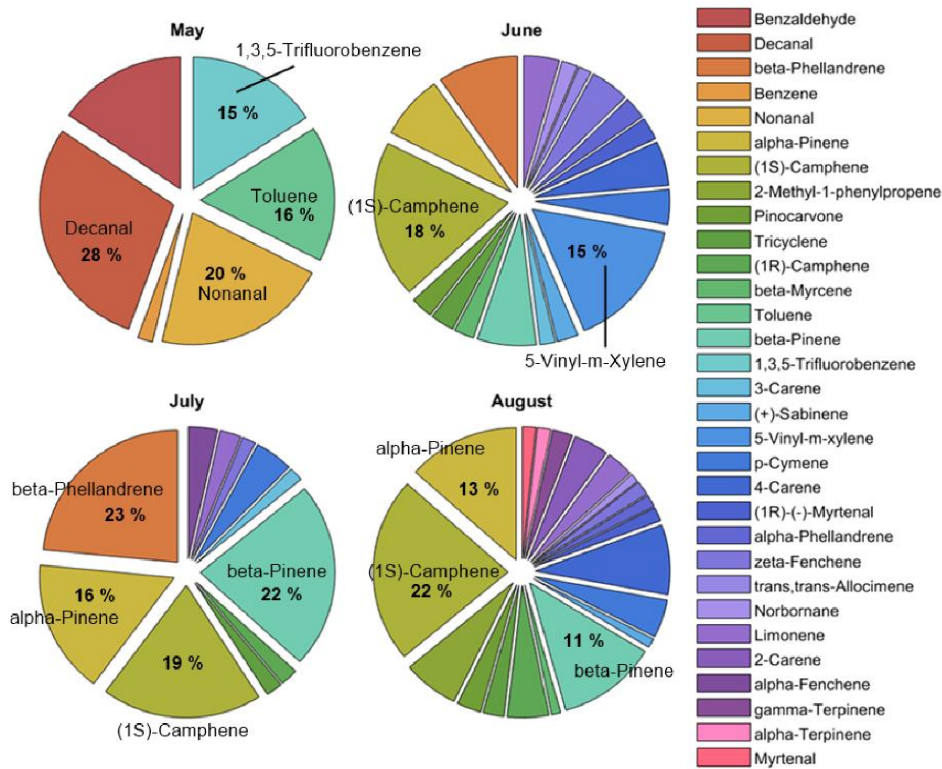


Figure A1: The daily average blend from the spruce Norway spruce S3S3 and percentage contribution in mass throughout the summer (May, June, July and August), showing only compounds with a mass contribution of at least 1 %.

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Table A2. The difference between healthy and infested trees when applying the calculations for the Q<sub>10</sub> temperature dependency. The reference emission rate at 30°C (F<sub>0</sub> is the reference emission rate standardized) entails higher emission rates at 30 °C with a higher number. The Q<sub>10</sub> coefficient indicates the emission rate change for every 10 °C temperature difference and is therefore a measure of temperature sensitivity, to temperature, leading to higher emission rates at higher temperatures when the coefficient is larger.

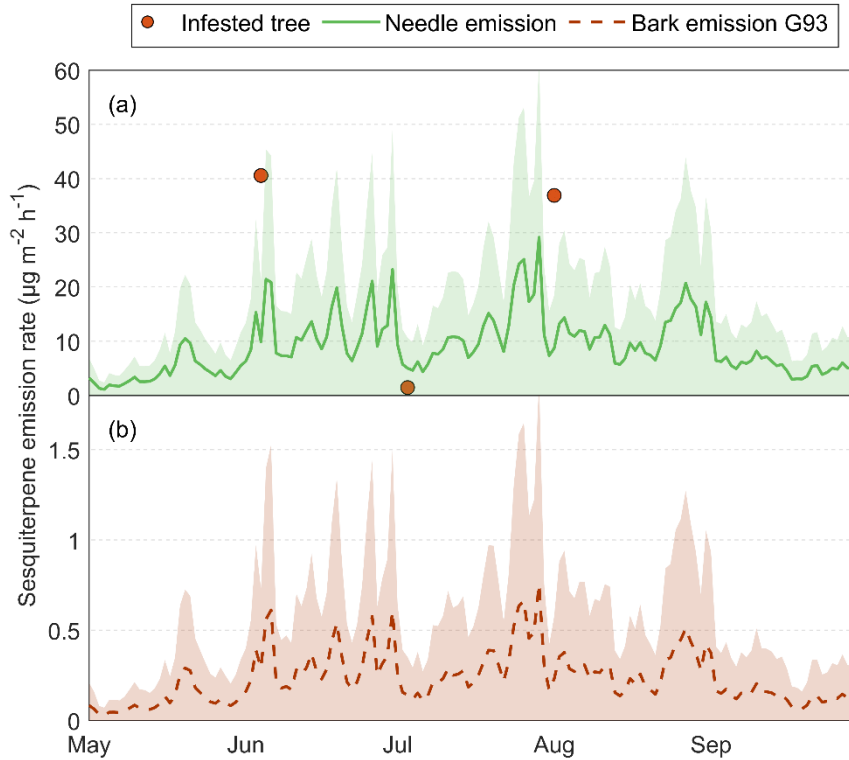
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Compound name	F <sub>0</sub> (µg m <sup>-2</sup> h <sup>-1</sup> )		Q <sub>10</sub>	
	Healthy	Infested	Healthy	Infested
<b>Monoterpenes</b>				
beta-Pinene	1111.0	34.90034901.4	87.8	980981.6
(1R)-Camphene	--	1.5001503.0	--	8080.2
beta-Phellandrene	2120.6	1.2401240.0	32.8	3837.8
alpha-Pinene	5555.0	880879.5	1918.7	1918.8
(1S)-Camphene	9392.6	470470.1	66.3	1414.3
beta-Myrcene	1615.8	250248.2	5756.7	170167.6
Limonene	1312.7	120116.4	1211.6	1716.9
3-Carene	54.6	110111.4	76.6	2625.8
p-Cymene	1615.7	9086.3	2222.4	1514.6
Tricyclene	--	8076.3	--	1413.6
(+)-Sabinene	--	7066.3	--	87.5
Eucalyptol	--	32.5	--	33.5
<b>Sesquiterpenes</b>				
Longifolene	0.010.01	2423.6	0.10.1	3333.1
alpha-Humulene	0.010.01	0.50.5	0.70.7	11.3



Other BVOCs

Isoprene	<u>11.3</u>	<u>22.4</u>	<u>10+0.2</u>	<u>24+24.4</u>
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**Figure A2. The measured and calculated BVOC emission rates for the group sesquiterpenes from Norway spruce in Hyltemossa: (a) the actual measured emission rates from infested Norway spruce bark (red dot) and calculated needle emission rates (green line) and (b) calculated healthy bark emission. The needle emissions are calculated based on the Guenther algorithm (Guenther et al., 1993) and the measured emission rates are taken from van Meeningen et al., (2017) and specific needle area (SLA) was taken from Wang et al., (2017), using the air temperature at 24 m taken from the HTM ICOS station (Heliasz, 2020). The healthy bark emission is calculated with the Guenther algorithm (orange) based on the tree temperature taken at 3 m height in the North and East orientation using data taken from the HTM ICOS station (Heliasz, 2020). The shaded areas (green, orange and black shade) represent the standard deviation from the mean for the respective calculation method.**

~~The modelled calculated emission rates for the group sesquiterpenes from (b) constitutive healthy bark VOC BVOC emission rates from the group sesquiterpenes based on the tree temperature taken at 3 meters height in the northNorth and eastEast orientation, data taken from the HTM ICOS station (Heliasz, 2020). The bark emission (black) is calculated based on the Guenther algorithm (Guenther et al., 1993) based on measured emission rates in this study. The (a) leafneedle emission rates (green) are calculated based on the Guenther algorithm and the measured emission rates are taken from van Meeningen et al., (2017) and specific leafneedle area (SLA) was taken from Wang et al., (2017) using the air temperature at 24 meters taken from the HTM ICOS station (Heliasz, 2020). For comparison, the actual measured bark VOC BVOC emission rates from one infested tree over time, from this study, is included in the Figure (red dot).~~

## 1120 **Author contribution**

EJ and TH designed and planned the campaigns. EJ performed the measurements. EJ performed the data analysis with contributions from KL, AG and AMJ. Funding was acquired by TH. EJ prepared the manuscript draft with contributions from all co-authors.

## **Competing interests**

1125 The authors declare that they have no conflict of interest.

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