



- 1 Extracellular enzyme production in the coastal upwelling system off
- 2 Peru during different upwelling scenarios: a mesocosm experiment
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## 18 Abstract

19	The Peruvian upwelling system is a highly productive ecosystem that could be altered by
20	ongoing global changes. We carried out a mesocosm experiment off Peru, with the addition of
21	water masses from the regional oxygen minimum zone (OMZ) collected at two different sites
22	simulating two different upwelling scenarios. Here we focus on pelagic remineralization of
23	organic matter by extracellular enzyme production of leucine aminopeptidase (LAP) and alkaline
24	phosphatase activity (APA). After addition of the OMZ water, dissolved inorganic nitrogen (N)
25	was depleted, but the standing stock of phytoplankton was relatively high even after nutrient
26	depletion (mostly >4 $\mu$ g chlorophyll <i>a</i> L <sup>-1</sup> ). During the initial phase of the experiment, APA was
27	0.6 nmol $L^{-1}$ h <sup>-1</sup> even though the PO <sub>4</sub> <sup>3-</sup> concentration was >0.5 µmol $L^{-1}$ . Initially, the dissolved
28	organic phosphorus (DOP) decreased, coinciding with an increase in PO4 <sup>3-</sup> concentration
29	probably linked to the APA. The LAP activity was very high with most of the measurements in
30	the range 200-800 nmol L <sup>-1</sup> h <sup>-1</sup> . This enzyme degrades amino acids, and these high values are
31	probably linked to the highly productive, but N-limited coastal ecosystem. Also, the experiment
32	took place during a rare coastal El Niño event with higher-than-normal surface temperatures,
33	which could have affected the enzyme production. Using a non-parametric multidimensional
34	scaling analysis (NMDS) with a generalized additive model (GAM), we found that
35	biogeochemical variables (e.g. nutrient and chlorophyll a concentrations), phytoplankton and
36	bacterial communities explained up to 64% of the variability in APA. The bacterial community
37	explained best the variability (34%) in LAP. The high hydrolysis rates for this enzyme suggests
38	that pelagic N remineralization supported the high standing stock of primary producers in the
39	mesocosms after N depletion.





- 41
- 42 Introduction

43	The Peruvian upwelling system is one of the most productive marine ecosystems in the world
44	(FAO, 2018). Its high productivity is driven by the upwelling of deep, nutrient rich water that
45	fuels primary production when reaching the sunlit surface ocean. The fate of the biomass
46	produced is of great importance for higher trophic levels and biogeochemical cycles. The
47	primary limiting nutrient is nitrogen (N), but iron (Fe) availability is also an important driver for
48	phytoplankton biomass production in addition to light (Chavez et al., 2008; Messié and Chavez,
49	2015). Part of the phytoplankton biomass passes to higher trophic levels through grazing and
50	predation. As the upwelled water parcel is transported further offshore by Ekman transport, part
51	of the biomass settles out of the euphotic zone and is decomposed in intermediate water layers
52	creating an extensive oxygen minimum zone (OMZ; Kalvelage et al., 2013).
53	The ongoing warming of surface waters is projected to have several consequences on marine
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54 55 56 57 58	ecosystems. For example, increasing temperatures lead to a reduction in gas solubility causing a decrease in oxygen concentrations; warming will also increase thermal stratification and reduce the ventilation of the deeper ocean (Keeling et al., 2010). Both of these effects will lead to expanding OMZs with potential consequences for biogeochemical cycling (Oschlies et al., 2018). Biogeochemical cycles of nitrogen (N) and phosphorus (P) are affected by O <sub>2</sub> depletions,
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54 55 56 57 58 59 60	ecosystems. For example, increasing temperatures lead to a reduction in gas solubility causing a decrease in oxygen concentrations; warming will also increase thermal stratification and reduce the ventilation of the deeper ocean (Keeling et al., 2010). Both of these effects will lead to expanding OMZs with potential consequences for biogeochemical cycling (Oschlies et al., 2018). Biogeochemical cycles of nitrogen (N) and phosphorus (P) are affected by O <sub>2</sub> depletions, e.g. through denitrification and sediment P release (Canfield et al., 2005). Hence, expanding OMZs may decrease the inorganic N : P ratio in the upwelled water potentially affecting the





63	After inorganic nutrients (primarily N) have been depleted, the productive surface layer is driven
64	by recycled production. In this process, dissolved organic matter (DOM) must first be broken
65	down into simpler forms before the DOM elements become biologically available. The
66	decomposition of DOM is not a uniform process as it is affected by both abiotic and biotic
67	variables. Extracellular enzymes hydrolyze complex dissolved organic molecules and is the first
68	step in remineralization of these DOM elements (Arnosti, 2011). Quantifying the rates of pelagic
69	remineralization is important for understanding recycled production and element fluxes in the
70	uppermost water masses. There are a range of different enzymes that are used for hydrolyzing
71	DOM, and two of the most studied ones are Leucine aminopeptidase (LAP) and Alkaline
72	phosphatase (AP).
73	LAP is a protein degrading enzyme that is used extracellularly in aquatic systems by bacteria,
74	some phytoplankton and fungi (Hoppe et al., 1988; Stoecker and Gustafson, 2003; Gutiérrez et
75	al., 2011). It hydrolyses a broad spectrum of substrates with a free amino group, but it has
76	preference for N-terminal leucine and related amino acids (Burley et al., 1990).

77 The AP enzyme is produced by a wide range of different organisms including aquatic bacteria 78 and phytoplankton. Its main function is related to the hydrolysis of phosphate monoesters that 79 separate orthophosphate (PO<sub>4</sub>) from an organic compound (Perry, 1972; Hoppe, 2003). AP exists 80 either as ectoenzyme (on the cell wall) or is excreted extracellularly and has for phytoplankton 81 commonly been related to P-limitation in aquatic environments (Rose and Axler, 1997; Nausch, 82 1998). Bacterial AP activity (APA) is more complex, as some, in particular particle attached 83 bacteria, take up and use C and N from the organic molecule after hydrolysis, and may for this 84 reason produce AP even under P replete conditions (Benitez-Nelson and Buesseler, 1999; 85 Hoppe, 2003; Labry et al., 2016).





In this study, we were interested in the dynamics of LAP and APA after an upwelling event in relation to biogeochemical variables and communities of plankton and bacterioplankton, and our main aim was to understand how much of the variability in enzyme activities could be explained by biogeochemical variables (e.g. nutrient concentrations) and microbial communities. This was done during a mesocosm experiment set up off the coast of Peru.

91

92 Materials and methods

A detailed description of the mesocosm set up and collection and addition of deep-water can be 93 94 found in Bach et al. (2020) within this special issue. In short, the mesocosm bags were 2 m in 95 diameter and extended from the surface down to 19 m depth, where the last 2 m was a conical sediment trap. Eight mesocosm bags were used and they were moored at 12.0555°S; 77.2348°W 96 97 just north of Isla San Lorenzo where the water depth is ~30 m. The mesocosms were closed by 98 attaching the sediment trap to the bottom and pulling the top above the surface on 25 Feb, 2017. 99 The bags were regularly cleaned from the inside and outside and sampled every second day with 100 integrated water samplers (0-10 m depth, IWS, Hydro-Bios). For a full detailed sampling and 101 cleaning timetable see Bach et al. (2020).

Water (100 m<sup>3</sup>) from the oxygen minimum zone (OMZ) was collected from two locations. The first was collected on day 5 from 12.028323°S; 77.223603°W from 30 m depth, and the second one from 12.044333°S; 77.377583°W from 90 m depth. The original aim was to collect severe and moderate OMZ signature water (differing in e.g. nitrate concentrations) from the first and second site, respectively. This assumption was based on long-term monitoring data, however, the chemical properties (e.g. nitrate concentration) was more similar in these water masses than





108	anticipated reflecting low and very low OMZ signatures from site 1 and 2 respectively, but this
109	was discovered only after the collection. Deep-water was added to the mesocosms in two steps
110	on day 11 and 12 after the enclosure of the mesocosms. Approximately ~20 $\text{m}^3$ of the mesocosm
111	water was exchanged with OMZ water, and both deep-water stations were pumped into four
112	replicate mesocosms. The water removed was pumped out from 11-12 m depth whereas the
113	deep-water was pumped into carefully moving the input hose between 14-17 m depth. The water
114	collected at 30 m depth was pumped into mesocosms M1, M4, M5 and M8 having low OMZ
115	signature and deep-water from 90 m depth into mesocosms M2, M3, M6 and M7 having very
116	low OMZ signature.
117	At the site of the mesocosms, the OMZ is close to the surface (<10 m depth; Graco et al., 2017)
440	
118	and consequently the bottom part of the mesocosm was low in oxygen. In order to keep the
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125

126 Nutrient concentrations

127 Inorganic nutrients were determined from filtered (0.45 µm filter, Sterivex, Merck) samples

128 immediately after the water arrived in the laboratory. For the measurements, we used a

129 continuous flow analyzer (QuAAtro AutoAnalyzer, SEAL Analytical) connected to a





- 130 fluorescence detector (FP-2020, JASCO). Phosphate ( $PO_4^{3-}$ ), nitrate ( $NO_3^{-}$ ) and nitrite ( $NO_2^{-}$ )
- 131 were determined colorimetrically (Murphy and Riley, 1962; Morris and Riley, 1963) and
- 132 corrected with the refractive index method reported by Coverly et al. (2012). Ammonium (NH<sub>4</sub><sup>+</sup>)
- 133 concentrations were determined fluorometrically (Kérouel and Aminot, 1997). Dissolved
- inorganic nitrogen (DIN) was calculated by summing NO<sub>3</sub><sup>-</sup>, NO<sub>2</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup>. Further details on
- 135 measurement accuracy can be found in Bach et al. (2020).
- 136 To measure total dissolved nitrogen (TDN) and phosphorus (TDP), the samples were first
- 137 filtered through pre-combusted (5 h, 450°C) Whatman GF/F filters (pore size 0.7 μm). The
- 138 filtrate was collected in 50 mL acid-cleaned high-density polyethylene (HDPE) bottles and
- 139 placed directly into a freezer (-20°C). Later the filtrates were thawed at room temperature over a
- 140 period of 24 hours and divided in two. The first half was used to determine inorganic nutrient
- 141 concentrations as described above. From the other half we determined the TDN and TDP
- 142 concentrations. An oxidizing reagent (Oxisolv, Merck) was added, and the samples were
- 143 autoclaved for 30 minutes. TDN and TDP were measured spectrophotometrically (QuAAtro,
- 144 Seal Analytical). Dissolved organic nitrogen (DON) concentrations were calculated by
- 145 subtracting DIN from TDN. Dissolved organic phosphorus (DOP) was calculated as the
- 146 difference between TDP and  $PO_4^{3-}$ .

147

148 Fluorescent dissolved organic matter and PARAFAC analysis

- 149 Fluorescent dissolved organic matter (FDOM) was determined by measuring fluorescence in
- 150 water samples with a Cary Eclipse (Agilent Technologies) spectrofluorometer, using excitation





- and emission slit widths of 10 nm. Wavelength ranges were set to 230-456 nm for excitation,
- 152 with 2 nm increments, and the 290-600 nm for emission with 5 nm increments. The
- 153 measurements were collected into excitation-emission matrices (EEM). Blanks were measured
- 154 with the same settings using ultrapure water.
- 155 Raw measurements were processed using the DOMFluor toolbox (v. 1.7; Stedmon and Bro,
- 156 2008) for Matlab (R2017a). The processing consisted in 1) blank subtraction from seawater
- 157 EEMs, 2) EEMs normalization to the Raman area (RA), estimated applying the trapezoidal rule
- 158 of integration on the emission scan at the 350 nm excitation wavelength in the blank EEMs, and
- 159 3) cropping of the 1st and 2nd order Rayleigh scatter bands. Inner filter correction was not
- 160 performed as for the duration of the experiment the absorption coefficient at 250 nm (a250)
- displayed values (mean  $\pm$  sd = 1.56  $\pm$  0.91 m<sup>-1</sup>) well below 10 m<sup>-1</sup>, above which correction is
- 162 considered necessary (Stedmon and Bro, 2008).
- 163 The processed EEMs were analyzed applying a Parallel Factor Analysis (PARAFAC) using the
- 164 DOMFluor toolbox. The PARAFAC model was constructed based on 125 samples (outliers were
- removed) and validated using split-half validation and random initialization. The resulting model
- 166 consisted of 4 components (C1-C4; supplementary material Fig S1). For each of them, the
- 167 fluorescence maximum (Fmax) was recorded. The identified fluorophores were compared to
- 168 others found in the literature using the OpenFluor database (openfluor.lablicate.com; Murphy et
- 169 al., 2014).

170





172	Phytoplankton	community a	and chlorophyll a

- 173 Flow Cytometry subsamples were transferred from the IWS into 50 mL beakers and stored cool
- in the dark until analysis max. 8 hours after sampling. Each sample (650 µL) was analyzed with
- an Accuri C6 flow cytometer (BD Biosciences) set to a high flow rate (i.e. 66 µL/min).
- 176 Phytoplankton groups were differentiated based on the strength of the forward scatter (FSC-A),
- the side scatter (SSC-A), the red fluorescence (FL3-A) and orange fluorescence (FL2-A) signal (
- 178 "A" refers to the area of the signal integral). Furthermore, we used sequential filtrations with
- 179 different polycarbonate filters (Whatman, pore-sizes 0.2, 0.4, 0.8, 2, 3, 5, 8 µm) to distinguish
- 180 populations in the cytogram based on size. This procedure was helpful to approximate how FSC-
- 181 A values corresponded with size. We defined the following phytoplankton groups:
- 182 Synechococcus-like cells (Syn; 0.2-2µm), Cryptophyte-like cells (Crypto; ~90% between 2-5
- 183 μm), picoeukaryotes (Peuks; 0.2-2 μm), Nanoeukaryotes (Nano; 2-20 μm, mostly in the lower
- range), Microeukaryotes 1 (Mikro1; ~15-40 μm, occasionally overlapping with Nano),
- 185 Microeukaryotes 2 (Mikro2; ~>40 µm, cluster dominated by Akashiwo sanguineum from about
- 186 day 20 onward), elongated cells "chains" determined by the ratio of FSC-A to FSC-H where "H"
- 187 refers to the height of the forward scatter signal (details about this approach are provided in Paul
- 188 et al., this issue. The goal of this was to detect chain-forming diatoms which we expected to be
- an important component of the community).
- 190 Samples for chlorophyll *a* (chl-*a*) determination were filtered onto GF/F filters (Whatman) and
- 191 flash frozen in liquid nitrogen and stored at -80 °C (or dry ice for a brief period during air
- 192 transfer; ~2 days) until measurement. The chl-*a* concentration was measured using high-
- 193 performance liquid chromatography. The chl-*a* autofluorescence of the phytoplankton
- 194 community was measured with a handheld fluorometer (AquaPen, Photon Systems Instruments)





- 195 using 450 nm excitation light. The photochemical efficiency was calculated based on the
- 196 relationship between the variable to maximal fluorescence (Fv/Fm).
- 197
- 198 16S-rRNA gene based bacterial community determination
- 199 One liter of surface water obtained from the individual sampling sites was filtered through sterile
- 200 Millipore Express PLUS membrane filters (polyethersulfon) with a cut-off of 0.22 µm and a
- 201 diameter of 47 mm (Merck Millipore). After filtration, the filters were flash frozen in liquid
- 202 nitrogen and stored at -80°C until nucleic acid extraction. Nucleic acid extraction was performed
- 203 using the NucleoSpin TriPrep- Kit (Machery-Nagle) according to manufacturer's instruction
- with an additional step at the beginning of the extraction using a pestle to properly homogenize
- the sample.
- 206 Primers applied for the amplification of the bacterial 16S rRNA gene fragments were annealing
- to the variable region 1 and 2 and consisted of an initial standardized Illumina adapter (regular),
- 208 followed by an 8 nucleotide barcode (X's), a linker region (underlined) and a primer sequence
- 209 (bold). The sequences were for the forward primer Bac27 5'-

## 210 AATGATACGGCGACCACCGAGATCTACACXXXXXXX<u>TATGGTAATTGT</u>AGAGTTT

- 211 GATCCTGGCTCAG-3' and reverse Bac338 5'-
- 212 CAAGCAGAAGACGGCATACGAGATXXXXXXX<u>AGTCAGTCAGCCTGCTGCCTCCC</u>
- 213 GTAGGAGT-3'. The individual PCR reaction contained 100 ng of the extracted DNA. PCR
- conditions and purification of the amplification product were previously described (Fischer et al.
- 215 2019a). The final library pool for sequencing was combined from the eluates and contained 100
- 216 ng of DNA. Amplicon library sequencing was performed on a MiSeq instrument. Library





217 therefore was prepared according to the manufacturer's instructions and sequenced using the v3

218 chemistry with 2 x 300bp paired-end.

219 Reads generated with amplicon sequencing were trimmed using the trimmomatic software 220 version 0.33 (Bolger et al., 2014) as described in Fischer et al. (2019b). Briefly, reads were 221 analyzed with a sliding window of 4 bp and regions were trimmed if the average Phred score 222 (Ewing and Green, 1998; Ewing et al., 1998) within the window was below 30. Trimmed reads 223 were kept within the dataset if the forward and reverse read both survived the quality trimming 224 and were longer than 36 bp. Afterwards, 20,000 reads per sample were kept in the dataset 225 (exceptions were sample M1 on day 10 (5817 reads) and M7 on day 24 (17660 reads) for further 226 analysis.

227 Quality trimmed sequences were analyzed using MOTHUR software, version 1.35.1 (Schloss et 228 al., 2009) as described in Fischer et al. (2019a). The quality filtered and subsampled reads were 229 concatenated to 1,040,321 contiguous sequences (contigs) using the command make.contig. 230 Contigs were filtered for ambiguous bases, homopolymers longer than 8 bases or sequences 231 longer than 552 bases using the command screen.seqs. The resulting 754,310 contigs were 232 checked for redundant sequences using the command unique.seqs and clustered to 199,746 233 unique sequences. The sequences were consecutively aligned to a modified version of the 234 SILVA database release version 132 (Pruesse et al., 2012) containing only the hypervariable 235 regions V1 and V2 by the command align.seqs. Sequences not aligning in the expected region 236 were removed from the dataset using the command screen.seqs. The alignment was further 237 optimized by removing gap-only columns with the command filter.seqs. The alignment 238 contained 717,217 sequences (148,760 unique). Rare and closely related sequences were 239 clustered using the commands unique seqs and precluster seqs. The latter was used to cluster





240	sequences with up to 3 positional differences compared to larger sequence clusters together.
241	Chimeric sequences were removed using the implemented software UCHIME (Edgar et al.,
242	2011) using the command chimera.uchime, followed by remove.seqs leaving 551,142 sequences
243	(29,519 unique) in the dataset. The classification of the sequences was performed against the
244	SILVA database and was done with a bootstrap threshold of 80 %. Operational taxonomic units
245	(OTUs) were formed using the average neighbor clustering method with the command
246	cluster.split. A sample-by-OUT table on the 97 % level, containing 10,258 OTUs, was generated
247	using the command make.shared. These OTUs were used for the subsequent analysis. After the
248	removal of mitochondria, chloroplast and singletons, 3225 OTUs were retained. These OTUs
249	were used for downstream analysis.

250

251 Extracellular enzymes

252 The leucine aminopeptidase (LAP) activity was determined using the method described by

253 Stoecker and Gustafson (2003) using L-leucine 7-amido-4-methyl-coumarin (Leu-AMC; Sigma

Aldrich) as a substrate. Leu-AMC was added to a final concentration of 500 µmol L<sup>-1</sup>, which was 254

255 determined in separate kinetics tests to saturate the enzyme activity. The samples  $(100-200 \,\mu l)$ 

256 were incubated in the dark at in situ surface temperature for a minimum of four hours. The

257 fluorescence was measured every 30-60 min with a Cary Eclipse (Agilent Technologies)

258 spectrofluorometer using 380 nm excitation and 440 nm emission wavelengths. The results were

259 compared with a standard curve determined using 7-amino-4-methyl-coumarin (AMC; Sigma

Aldrich) dissolved in DMSO, and the LAP activity calculated by linear regression. 260





Measurements of alkaline phosphatase activity (APA) were conducted with 20 ml subsamples of 261 initial/incubated seawater using 100 nmol L<sup>-1</sup> 4-methylumbelliferyl phosphate (MUF-P; Sigma-262 263 Aldrich) as the organic phosphate substrate (Ammerman, 1993). Fluorescence was measured on a BIOTEK Microplate Reader with a Cary Eclipse (Agilent Technologies) spectrofluorometer 264 using 355 nm excitation light and 460 nm emission detection. Following MUF-P addition, 265 fluorescence measurements were performed at 0, 1.5, and 3 h and APA (h<sup>-1</sup>) was calculated 266 from the linear increase in fluorescence and calibrated against 4-methylumbelliferone (MUF; 267 268 Sigma-Aldrich). The assays were performed and incubated in the dark. Ultrapure water (Milli-Q) 269 blanks and paraformaldehyde-killed controls generally yielded fluorescence values similar to t =270 0 readings.

271

## 272 Statistical analysis

273 Before comparisons between the two experimental treatments were conducted, we first 274 constructed a cumulative value where each measured value was summed up for each sampling 275 day. The linear regressions of the cumulative enzyme activity from the two treatments (n = 4)276 were compared with Student's t-test. In addition, the effect of biogeochemical, phytoplankton 277 and bacterioplankton community composition to APA and LAP was determined, using the 278 ordination scores of the first and second axis of a non-parametric multidimensional scaling 279 (NMDS) as explanatory variables in generalized additive models (GAMs) with APA or LAP as 280 dependent variable. The NMDS was applied separately to each group of variables: 281 biogeochemical, phytoplankton community and bacterioplankton community. The individual





282	explanatory power of each MDS score was estimated with a univariate GAM. The visualization
283	of the links was done for each explanatory variable through the prediction from the full model
284	object, setting all other explanatory variables at their mean value. In addition, links to the scores
285	of the biogeochemical variables and phytoplankton community NMDS were estimated with one
286	GAM model. It was not possible to include the bacterioplankton community into this model due
287	to the different sampling regime (lower number of samples) and this was treated with a second
288	model. NMDS was estimated with the metaMDS function in the Vegan package (Oksanen et al.,
289	2019), and GAMs were fitted using the gam function in the mgcv package (Wood, 2017). For
290	explaining the deviance, an adjusted coefficient of determination ( $\mathbb{R}^2$ ) was used. An adjusted $\mathbb{R}^2$
291	takes into account the model complexity and is more conservative than a non-adjusted R <sup>2</sup> .

292

293 RESULTS

294

295 Nutrients

296 The addition of deep-water increased the phosphate concentrations whereas the dissolved

inorganic nitrogen (DIN) was >2  $\mu$ mol L<sup>-1</sup> in the mesocosms until after the addition of deep-

water (days 11 and 12 of the experiment). After the addition of the deep-water, the DIN

concentration rapidly declined and was depleted at day 15 in most mesocosms except in M3

300 where DIN depletion occurred a week later (day 22; Fig 1). The  $PO_4^{3-}$  concentration increased

after closing the mesocosm and reached ~1.9  $\mu$ mol L<sup>-1</sup> in all mesocosms after the deep-water





addition. There was only a slight reduction to approximately  $1.5 \,\mu$ mol PO<sub>4</sub><sup>3-</sup> L<sup>-1</sup> over the course

- 303 of the experiment (Fig 1).
- 304 The dissolved organic nitrogen (DON) and phosphorus (DOP) concentrations were initially 9 –
- $12 \mu mol L^{-1}$  and  $0.6 1.0 \mu mol L^{-1}$ , respectively. There was no drastic change in DON with the
- 306 deep-water addition and there was an overall decrease in DON to 6.0 7.9  $\mu$ mol L<sup>-1</sup> on day 30
- 307 after which it increased somewhat again. The DOP concentrations decreased rapidly the first 8
- $\label{eq:constraint} 308 \qquad \mbox{days to } 0.19 \mbox{-} 0.32 \ \mbox{\mu mol } L^{\mbox{-}1} \ \mbox{but increased after the deep-water addition and remained within } 0.2$
- $-0.7 \mu mol L^{-1}$  interval for the rest of the experiment.
- 310 The PARAFAC modelling of the EEMs yielded four FDOM components (C1-C4; Fig 2, Fig S1).
- 311 Using the OpenFluor database we identified multiple fluorophores with strong similarity
- 312 (TCCex $\cdot$ em > 0.95) to our components (Table S1). Components 1 and 3 had characteristics
- 313 resembling amino acid/protein compounds whereas components 2 and 4 were more humic-like
- 314 (Table S1). All FDOM components increased sharply at day 18. This did not take place in
- 315 Pacific seawater sampled outside the mesocosm where the FDOM was relatively stable
- throughout the experiment. After the increase at day 18, humic-like components (C2 and C4)
- 317 were relatively stable but decreased slightly after day 28-30. The amino acid-like components
- 318 (C1 and C3) exhibited higher variability among mesocosms, and C3 had overall higher
- 319 variability throughout the experiment. Both humic-like and amino acid-like components
- 320 maintained fluorescence values above the initial ones until the end of the experiment, but there
- 321 were no clear differences between the treatments. However, towards the end of the experiment
- 322 M1 and M2 had highest concentrations of C1, M1 also had highest concentration of C2 and C3
- 323 whereas M3 had the highest concentration of C4 at the end of the experiment.





324

325	Chlorophyll, photochemical efficiency and phytoplankton community
326	After deep-water addition, the chl- <i>a</i> concentration increased from 2-4 $\mu$ g L <sup>-1</sup> to 4-8 $\mu$ g L <sup>-1</sup> except
327	for mesocosms M3 and M4 where the increase was not as pronounced (Fig 3). The chl- $a$
328	concentration in M3 increased after day 22 to ~4 $\mu$ g Chl- <i>a</i> L <sup>-1</sup> , whereas in M4 the chl- <i>a</i>
329	concentration remained low (<2 $\mu$ g L <sup>-1</sup> ) throughout most of the experiment (Fig 3). The
330	photochemical efficiency (Fv/Fm) was approximately 0.7 throughout the whole experiment
331	without major difference between mesocosms, except for M4 where it was consistently lower
332	(<0.6) during the last week of experiment (Fig 3).
333	The initial community was dominated by diatoms in terms of biomass but this group gradually
334	reduced in numbers after the enclosure of the mesocosms and instead the mixotrophic
335	dinoflagellate Akashiwo sanguineum appeared (Fig 4). The cell counts done with the flow
336	cytometer were checked with a microscope and this was the primary species in terms of biomass
337	in the Microeukaryote 2 group (Fig 4). The exceptions were mesocosms M3 and M4 where this
338	dinoflagellate was not abundant (M4) or bloomed later (M3) and where there were more
339	Chrysophytes. In M4 there was in addition a bloom of picoeukaryotes starting after day 20 (Fig
340	4).
341	

342 Bacterial community

The bacterial community was dominated by the class Alphaproteobacteria throughout the wholeexperiment and in all the mesocosms units, reaching values between 60 to 88% of the total





- 345 sequences at day16 (Fig 5). Within Alphaproteobacteria, the *Roseobacter* lineage (genera
- 346 HIMB11, Ascidiaceihabitans, Amylibacter and Planktomarina in M1) of the order
- 347 Rhodobacterales contributed most to the bacterial community in all the mesocosms (10-55 %) in
- 348 particular on day 16, except in M8 where the SAR11 Ia clade dominated the community (55% of
- the total sequence at day 16). The order Parvibaculales had high relative abundances (12-20% of
- the total sequences) in M4, M5, M6 and M7 before the deep-water addition (day 10) decreasing
- in the following week. The relative abundance of order Rickettsiales peaked at day 16 in all the
- 352 mesocosms except in M8, decreasing after one week. The class Gammaproteobacteria comprised
- between 20 to 45% of the total relative abundance. Within Gammaproteobacteria, the order
- 354 Thiomicospirales had high relative abundance (8-17% total sequences) at day 10 in most of the
- 355 mesocosms, whereas the order Cellvibrionales and order Oceanospirillales (genus
- 356 *Pseudohongiella*) increased from day 24 and by the end of the experiment, respectively. In M8,
- 357 the abundances of orders Thiomicospirales and Pseudomonadales (14% of total sequences)
- increased at day 24. Other groups that increased in abundance in the second half of the
- experiment were the deltaproteobacterial orders Desulfobacteriales (7-20% in M2, M3, M4 and
- 360 M5) and Bdellovibrionales (5-8% in M2, M3 and M4). The order Flavobacteriales dominated
- 361 within Bacteroidetes and the relative abundance ranged from 1 to 25% throughout the
- 362 experiment, being generally high (10-20%) at day 10. The flavobacterial genus Aurantivirga
- 363 contributed > 7% in M1, M2 and M3.

364

365 Enzyme activity





366	The initial LAP activity before the deep-water addition was relatively low (average 359 nmol L <sup>-1</sup>
367	$h^{-1} \pm 81 \text{ nmol } L^{-1} h^{-1} \text{ SD}$ ) but increased after the addition of deep-water in some of the
368	mesocosms (Fig 6). In M3 the LAP activity was high, reaching 1600 nmol $L^{-1}$ h <sup>-1</sup> directly after
369	the deep-water addition, but decreased after that. The highest cumulative LAP activity at the end
370	of the experiment was in M7 where the LAP activity was 716 nmol $L^{-1} h^{-1}$ after deep-water
371	addition and the average after day 16 was 657 nmol $L^{-1} h^{-1} \pm 142 \text{ nmol } L^{-1} h^{-1}$ (SD). There was a
372	difference between the treatments in the cumulative LAP after the addition of the deep-water
373	until day 16, with the very low OMZ signature (lowest NO3 concentration) water producing the
374	highest LAP activity (Students' t-test, $p = 0.047$ ), but this difference disappeared after day 16 (p
375	= 0.44).
376	The alkaline phosphatase activity (APA) was 0.5-0.6 nmol $L^{-1} h^{-1}$ at the beginning of the
377	experiment but decreased to undetectable levels after day 30 (Fig 7). There was a noticeable drop
378	in APA after the addition of the deep-water, and the decrease continued gradually until day 28
379	after which the APA was very low (<0.1 nmol $L^{-1} d^{-1}$ ). The APA was similar in all the
380	mesocosms and there was no treatment effect ( $p = 0.81$ ). The exception to this was M3 where the
381	APA was lower, compared to all other mesocosms for most of the experiment (Fig 7).

382 The variability in APA was better explained by the measured variables than LAP (Fig 8). The

383 biogeochemical variables and bacterioplankton community separately explained 62% of the

384 variability in APA, whereas the phytoplankton community alone explained 57% of the

variability. Combining both the biogeochemical variables and the phytoplankton community

386 increased the explanatory power to 74% (bacterioplankton community not included as the

387 number of sample points were less). The variability in LAP was best explained by the

388 bacterioplankton community (38%) followed by biogeochemical variables (20%) and





- 389 phytoplankton community (18%). The combined biochemical variables and phytoplankton
- 390 community explained 28% of the LAP variability.
- 391
- 392
- 393 DISCUSSION

394	After the closure and addition of deep-water there was rapid phytoplankton growth in the upper 5
395	m of the mesocosms, with low light conditions limiting primary production deeper down (Bach
396	et al., 2020). The DIN concentrations were depleted around day 18 coinciding with an increase in
397	several of the FDOM components (both amino acid-like and humic-like components), also
398	matching the end of the phytoplankton bloom. There was, however, relatively constant and low
399	export of carbon out from the mesocosms (Bach et al., 2020) and at the same time relatively high
400	Chl- <i>a</i> concentration (mostly >4 $\mu$ g chl- <i>a</i> L <sup>-1</sup> ) under conditions with depleted DIN (Fig 3). In
401	addition, the photochemical efficiency was overall relatively high (>0.5) throughout the
402	experiment suggesting regenerated primary production driven by recycling of nutrients. The
403	measured hydrolysis rates, particularly LAP, indicated that extracellular enzyme activity plays an
404	important role for this recycled production.
405	The main aim of this study was to relate the biogeochemical and microbial community to the
406	extracellular enzyme activity and a more detailed description of the temporal development and
407	biomass comparison of microbial groups will be presented elsewhere in this special issue (e.g.
408	Bach et al., 2020). Among phytoplankton, diatoms are typically dominating following upwelling
409	events (Anabalón et al., 2016), whereas dinoflagellates tend to become more dominant after





410	establishment of stratification (Margalef et al., 1979). This was also seen in our mesocosm as the
411	dinoflagellate Akashiwo sanguinea, a mixotrophic species that may form red tides (Jeong et al.,
412	2005; Badylak et al., 2014), that quickly appeared in most mesocosm after OMZ water was
413	added with some exceptions. In M3 it appeared a little later and in M4 it did not bloom at all.
414	Interestingly these two mesocosms had a higher concentration of cryophytes and M4 had
415	additionally a bloom event of picoeukaryotes. Being mixotrophic, A. sanguinea is known to prey
416	on smaller species (Jeong et al., 2005) and lower grazing pressure could be the reason for the
417	bloom of picoeukaryotes in M4.
418	The bacterial community composition changed during the experiment but without clear treatment
419	effects. The dominant bacterial groups were the class Alphaproteobacteria, (Parvibaculales,
420	SAR11 subclade Ia, Roseobacter clade and Rickettsiales), class Gammaproteobacteria (SAR116
421	clade, Cellvibrionales, Oceanospirillales and SUP05 clade) and to lesser extent the class
422	Deltaproteobacteria (Desulfobacterales) and class Bacteroidea (order Flavobacteriales). SAR11
423	subclade Ia, Roseobacter clade, SAR116 clade, SUP05 clade and Desulfobacterales are known to
424	utilize inorganic and organic sulfur components such as hydrogen sulfide (H <sub>2</sub> S), sulfate (SO <sub>4</sub> )
425	and dimethylsulfoniopropionate (DMSP) metabolites for their metabolic requirements (Nowinski
426	et al., 2019) and are coupled with the nitrogen cycle (Schunck et al., 2013). Specifically, the
427	sulfur-oxidizing SPU05 oxidizes $H_2S$ coupled with the nitrate reduction and potentially produces
428	nitrite (Shah et al., 2017), whereas Desulfobacterales play an important role in $N_2$ fixation (Gier
429	et al., 2016). These bacterial taxa associated with the sulfur cycle are typically found in the OMZ
430	regions (Pajares et al., 2020). We observed a temporal shift in the bacterial community through
431	the experiment changing between sulfur-oxidizing (SUP05) and sulfate-reducing
432	(Desulfobacterales) bacteria, probably liked to the nitrate availability, i.e. more DIN at the





433	enclose of the mesocosms and thus more relative abundance of SUP05. We also observed a shift
434	within phytoplankton-associated bacteria (Roseobacter lineage, Gammaproteobacteria, and
435	Flavobacteriales) that likely responded to the availability of DOM supply during the experiment
436	(Buchan et al 2014, Chafee et al 2017). The high relative abundance of Flavobacteriales and
437	genera from the Roseobacter lineage on days 10 and 16, respectively, coincided with the increase
438	in chl-a and high LAP activity until day 16. Positive correlations have been observed between
439	chl-a, Bacteroides and Deltaproteobacteria and LAP during phytoplankton blooms (Shi et al
440	2019). However, we do not have gene expression data and cannot make any firm conclusion
441	about the connection between these groups and production of LAP.
442	The temporal shift in the bacterial community indicates niche partitioning between bacterial taxa
443	that assimilate different organic substrates or inorganic sulfur components, produced during
444	phytoplankton bloom events or from sulfidic events (Schunck et al., 2013; Callbeck et al., 2018;
445	Nowinski et al., 2019). Our results support previous studies that have demonstrated the important
446	role of the sulfur cycle in shaping the bacterial community composition in poorly oxygenated
447	waters (Schunck et al., 2013; Aldunate et al., 2018). It is worth to note that the conditions in the
448	bottom of the mesocosms were sub-oxic and there might have been a clear depth gradient in the
449	bacterial community that was not picked up by our integrated 0-10 m sampling.
450	Overall, there was a treatment effect of the different OMZ waters on the LAP activity, with
451	higher LAP in the very low OMZ signature addition, but this effect was only observed right after
452	the addition of the deep-water. There were also slightly higher NO3 concentrations in this water
453	(Bach et al., 2020). However, this difference in both DIN and LAP was relatively small and

- disappeared a week after the OMZ water addition, most likely because the collected deep-water
- 455 were more similar between the two locations than anticipated, with relatively similar





- 456 concentrations of DIN. Although there were differences between individual mesocosms in terms
- 457 of the plankton community structure, there were no clear differences between treatments, and we
- 458 can conclude that the availability of nutrients by itself can shift the LAP production.
- The LAP activity in our study was very high (~10-times higher compared with most literature
- data). In a comparable study but further offshore in Peru, the LAP activity was  $20 65 \text{ nmol } \text{L}^{-1}$
- 461  $h^{-1}$  in surface waters (Maßmig et al., 2020). Further to the south, in Chile (30° 30.80' S), values
- 462 up to 230 nmol L<sup>-1</sup> h<sup>-1</sup> have been recorded, with a clear seasonal cycle linked to upwelling events
- 463 (Gutiérrez et al., 2011). With most of our data ranging between 200 800 nmol L<sup>-1</sup> h<sup>-1</sup> it is clear
- that these LAP activities are linked to the upwelling, which is more intense near the coast and
- also more constant at the study site compared with further south. The enzyme activity in
- sediments can be up to three orders of magnitude higher than what we found (Hoppe et al.,
- 467 2002), and an order of magnitude higher values have been observed in a eutrophic, salt-water
- lake (Song et al., 2019). The high LAP activities are likely a reflection of the high microbial
- 469 activity in the Peruvian upwelling system. The experiment was also taking place during a rare
- 470 coastal El Niño event (Garreaud, 2018), with anomalous higher surface temperatures, which
- 471 could be a reason for the high values we recorded as LAP is known to increase with temperature472 (Christian and Karl, 1995).

There was also some loss of N due to denitrification, estimated to 0.2-4.2 nmol N<sub>2</sub> L<sup>-1</sup> h<sup>-1</sup> during the experiment (Schulz et al 2021). For comparison, the LAP activity suggests an average of 417 nmol L<sup>-1</sup> h<sup>-1</sup> hydrolyzation of N-containing compounds, but this should be seen as the maximal potential rather than the actual rate. The use of fluorescently labelled substrates for measuring extracellular activity is a proxy method that has some drawbacks. The primary one is that the molecular structure of the substrate used is never equivalent to the high molecular weight DOM





479	in the water. This means that the measured hydrolysis rates could be an overestimation of the
480	actual hydrolysis rates of DOM (e.g. Arnosti, 2011). The primary benefit of the method is that it
481	is straightforward and has been in widespread use for decades, which means that comparisons
482	with other ecosystems is possible, and for our purpose, we can use it for better understanding
483	how much of the variability can be explained by the other measured variables.
484	

485	Considering the APA, the most interesting aspect was that it was measurable in the beginning of
486	the experiment at high $PO_4^{3-}$ concentration. This high APA activity at high $PO_4^{3-}$ concentration
487	has been observed in deep oceans (Hoppe and Ullrich, 1999; Baltar et al 2016). Baltar and
488	collaborators (2016) also observed an increase in APA in experiments amended with organic
489	matter suggesting the activity of APA by organic matter supply, independently of the PO <sub>4</sub> <sup>3-</sup>
490	concentration. This could be due to bacterial APA, which is more complex than for
491	phytoplankton, in that it can be linked to the hydrolysis and acquisition of C (Hoppe, 2003). This
492	is supported by the initial decrease in DOP and increase in $PO_4^{3-}$ , which indicates that the AP
493	hydrolysis of DOP added to the $PO_4^{3-}$ pool. It is known that APA stays suspended and active for
494	a long time in marine environments, and cell-free APA was reduced by only 25% over 16 days in
495	the experiment by Thomson et al. (2019). If this enzyme is viable for this long, it suggests that
496	there was no new production of AP after the closure of the mesocosms, which is supported by
497	the dilution effect of adding the deep-water. In that case, the disappearance of the initial AP took
498	30 days.

The hydrolysis rates of AP were relatively low compared with most published data, probably linked to the clear surplus of  $PO_4^{3-}$ . It is worth to note, however, that we were most likely not





- 501 measuring the maximal potential hydrolysis rates as substrate addition was relatively low (100 nmol  $L^{-1}$ ) and would likely have been higher with more added substrate. This could be the reason 502 for the apparent discrepancy between the measured hydrolysis rates and the change in the PO<sub>4</sub><sup>3-</sup> 503 504 and DOP pools during the 10 first days of the experiment. During this time there was a decrease of approximately 0.5  $\mu$ mol DOP L<sup>-1</sup> and an increase of 0.6  $\mu$ mol PO<sub>4</sub><sup>3-</sup> L<sup>-1</sup>, suggesting an actual 505 hydrolysis rate of 2.0-2.5 nmol L<sup>-1</sup> h<sup>-1</sup> (assuming 500-600 nmol over 10 days). This is a factor 3-506 507 4 higher compared with the initially measured APA of  $\sim 0.6$  nmol L<sup>-1</sup> h<sup>-1</sup>. 508 The statistical model that we applied was better at explaining the APA variation than LAP. APA 509 gradually decreased during the initial phase of the experiment to undetectable levels after the 510 middle of the experiment. Any correlation does not mean causality and the higher coefficient of 511 determination is probably rather a reflection of the clear temporal development in APA. If the 512 AP was produced before the closure of the mesocosm and slowly degraded as discussed above, 513 any connection with the biogeochemical or plankton community must be due to unrelated
- temporal development. For example, the DIN also decreased over time but was likely not relatedto the APA.
- 516 For the LAP the overall explanatory power by the biogeochemical and plankton community

517 composition was less than for APA, but interestingly the bacterioplankton community

518 composition clearly explained the variability better (38%) than the combined biogeochemical

- and phytoplankton community (28%). Considering that the bacterial community was not
- sampled as frequently as the biogeochemical variables and flow cytometer counts, we suspect
- that the explanatory power would have increased with more frequent sampling. It is likely that
- 522 bacteria were producing the LAP and some taxa are more reliant on enzyme production for
- 523 nutrient acquisition than others (Ramin and Allison, 2019). Some dinoflagellates are also known





- to produce LAP and most of the mesocosms with high dinoflagellate biomass except M4.
- 525 However, the phytoplankton community only explained 18% of the variability in LAP, and these
- 526 dinoflagellates were likely not producing any substantial amount of this enzyme.

In conclusion, we found very high levels of LAP (mostly in the range 200 - 800 nmol L<sup>-1</sup> h<sup>-1</sup>), 527 528 which is an order of magnitude higher than most literature data. This is probably linked to the 529 upwelling supporting high levels of microbial activity in combination with the general DIN limitation in the coastal Peruvian upwelling. There was measurable APA at the start of the 530 531 experiment, but this gradually declined to undetectable levels in all of the mesocosms midway (~30 days) in the experiment. With high concentrations of  $PO_4^{3^-}$ , low APA is not surprising, and 532 533 AP is a relatively slowly degrading enzyme that could have been fully dissolved and produced 534 before the closure of the mesocosms. Our statistical mode explained more of the variability of 535 APA (74%) compared with LAP, probably due to its clear temporal development. The 536 bacterioplankton community composition explained best the variability of LAP (38%) compared 537 with the combined biochemical and phytoplankton community model (28%). With more than 538 50% of the variability unaccounted for, we are still missing important pieces of the puzzle 539 understanding the variability in this enzyme. The hydrolysis rates for LAP suggests that pelagic 540 N remineralization supported the relatively high standing stock of primary producers (mostly >4  $\mu$ g chl-*a* L<sup>-1</sup>) in the mesocosms after N depletion. 541

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544 Data availability





- 545 All data will be made available on the permanent repository www.pangaea.de after publication.
- 546 The DNA sequencing data will be submitted to NCBI SRA (in prep).

547

- 548 Author contribution
- 549 Samples were taken by KS, JP, JA, LB, EvdE, MF, NHH, JM and UR. In addition to the
- sampling crew, further data analysis was conducted by MTCG and MGL. UR developed the
- 551 experimental design and sampling strategy and coordinated the mesocosm campaign. All co-
- authors contributed to the data interpretation. KS wrote the manuscript with contributions fromall co-authors.

554

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#### 574 REFERENCES

- Aldunate, M., De la Iglesia, R., Bertagnolli, A. D., and Ulloa, O.: Oxygen modulates bacterial
- 576 community composition in the coastal upwelling waters off central Chile, Deep Sea Research
- 577 Part II: Topical Studies in Oceanography, 156, 68-79, 2018.
- 578 Ammerman, J.: Microbial cycling of inorganic and organic phosphorus in the water column,
- 579 Handbook of methods in aquatic microbial ecology, 1, 649-660, 1993.
- 580 Anabalón, V., Morales, C., González, H., Menschel, E., Schneider, W., Hormazabal, S.,
- 581 Valencia, L., and Escribano, R.: Micro-phytoplankton community structure in the coastal
- upwelling zone off Concepción (central Chile): Annual and inter-annual fluctuations in a highly
- dynamic environment, Prog Oceanogr, 149, 174-188, 2016.





- 584 Arnosti, C.: Microbial extracellular enzymes and the marine carbon cycle, Ann Rev Mar Sci, 3,
- 585 401-425, 2011.
- 586 Bach, L. T., Alvarez-Fernandez, S., Hornick, T., Stuhr, A., and Riebesell, U.: Simulated ocean
- 587 acidification reveals winners and losers in coastal phytoplankton, PloS one, 12, e0188198, 2017.
- 588 Bach, L. T., Paul, A. J., Boxhammer, T., Esch, E. v. d., Graco, M., Schulz, K. G., Achterberg, E.,
- 589 Aguayo, P., Aristegui, J., Ayon, P., Banos, I., Bernales, A., Boegeholz, A. S., Chavez, F., Chen,
- 590 S.-M., Doering, K., Filella, A., Fischer, M., Grasse, P., Haunost, M., Hennke, J., Hernandez-
- 591 Hernandez, N., Hopwood, M., Igarza, M., Kalter, V., Kittu, L., Kohnert, P., Ledesma, J.,
- 592 Lieberum, C., Lischka, S., Loescher, C., Ludwig, A., Mendoza, U., Meyer, J., Meyer, J.,
- 593 Minutolo, F., Cortes, J. O., Piiparinen, J., Sforna, C., Spilling, K., Sanchez, S., Spisla, C., Sswat,
- 594 M., Moreira, M. Z., and Riebesell, U.: Factors controlling plankton productivity, particulate
- 595 matter stoichiometry, and export flux in the coastal upwelling system off Peru, Biogeosciences
- 596 17: 4831-4852, 2020.
- 597 Badylak, S., Phlips, E. J., and Mathews, A. L.: Akashiwo sanguinea (Dinophyceae) blooms in a
- sub-tropical estuary: an alga for all seasons, Plankton and Benthos Research, 9, 147-155, 2014.
- 599 Benitez-Nelson, C. R., and Buesseler, K. O.: Variability of inorganic and organic phosphorus
- turnover rates in the coastal ocean, Nature, 398, 502-505, 1999.
- 601 Bolger, A. M., Lohse, M., and Usadel, B.: Trimmomatic: a flexible trimmer for Illumina
- 602 sequence data, Bioinformatics, 30, 2114-2120, 2014.





- 603 Buchan, A., G. R. LeCleir, C. A. Gulvik, and J. M. González. 2014. Master recyclers: Features
- and functions of bacteria associated with phytoplankton blooms. Nat. Rev. Microbiol. 12: 686-
- 605 698. doi:10.1038/nrmicro3326.
- 606 Burley, S. K., David, P. R., Taylor, A., and Lipscomb, W. N.: Molecular structure of leucine
- aminopeptidase at 2.7-A resolution, Proc Natl Acad Sci, 87, 6878-6882, 1990.
- 608 Canfield D., Kristensen E., and Thamdrup B.: Aquatic geomicrobiology. Elsevier, 2005.
- 609 Callbeck, C. M., Lavik, G., Ferdelman, T. G., Fuchs, B., Gruber-Vodicka, H. R., Hach, P. F.,
- 610 Littmann, S., Schoffelen, N. J., Kalvelage, T., and Thomsen, S.: Oxygen minimum zone cryptic
- 611 sulfur cycling sustained by offshore transport of key sulfur oxidizing bacteria, Nature
- 612 communications, 9, 1-11, 2018.
- 613 Chafee M, Fernàndez-Guerra A, Buttigieg PL, Gerdts G, Eren AM, Teeling H, Amann RI (2017)
- 614 Recurrent patterns of microdiversity in a temperate coastal marine environment. The ISME J
- **615** 12:237
- 616 Chavez, F. P., Bertrand, A., Guevara-Carrasco, R., Soler, P., and Csirke, J.: The northern
- 617 Humboldt Current System: Brief history, present status and a view towards the future, Prog
- 618 Oceanogr, 79, 95-105, 2008.
- 619 Christian J.R., Karl D.M.: Bacterial ectoenzymes in marine waters: activity ratios and
- temperature responses in three oceanographic provinces. Limnol Oceanogr, 40:1042-1049, 1995.
- 621 Coverly, S., Kérouel, R., and Aminot, A.: A re-examination of matrix effects in the segmented-
- flow analysis of nutrients in sea and estuarine water, Analytica chimica acta, 712, 94-100, 2012.





- 623 Edgar, R. C., Haas, B. J., Clemente, J. C., Quince, C., and Knight, R.: UCHIME improves
- sensitivity and speed of chimera detection, Bioinformatics, 27, 2194-2200, 2011.
- 625 Ewing, B., and Green, P.: Base-calling of automated sequencer traces using phred. II. Error
- 626 probabilities, Genome research, 8, 186-194, 1998.
- 627 Ewing, B., Hillier, L., Wendl, M. C., and Green, P.: Base-calling of automated sequencer traces
- 628 usingPhred. I. Accuracy assessment, Genome research, 8, 175-185, 1998.
- 629 FAO: The state of world fisheries and aquaculture, Food and Agriculture Organization of the
- 630 United Nations, Rome, 223 pp., 2018.
- 631 Fischer, M. A., Güllert, S., Refai, S., Künzel, S., Deppenmeier, U., Streit, W. R., and Schmitz, R.
- A.: Long-term investigation of microbial community composition and transcription patterns in a
- biogas plant undergoing ammonia crisis, Microbial Biotechnology, 12, 305-323, 2019a.
- 634 Fischer, M. A., Ulbricht, A., Neulinger, S. C., Refai, S., Waßmann, K., Künzel, S., and Schmitz-
- 635 Streit, R. A.: Immediate effects of ammonia shock on transcription and composition of a biogas
- reactor microbiome, Front Microbiol, 10, 2064, 2019b.
- 637 Garreaud, R. D.: A plausible atmospheric trigger for the 2017 coastal El Niño, International
- 638 Journal of Climatology, 38, e1296-e1302, 2018.
- 639 Gier, J., Sommer, S., Löscher, C. R., Dale, A. W., Schmitz-Streit, R., and Treude, T.: Nitrogen
- 640 fixation in sediments along a depth transect through the Peruvian oxygen minimum zone,
- 641 Biogeosciences (BG), 13, 4065-4080, 2016.





- 642 Graco, M. I., Purca, S., Dewitte, B., Castro, C. G., Morón, O., Ledesma, J., Flores, G., and
- 643 Gutiérrez, D.: The OMZ and nutrient features as a signature of interannual and low-frequency
- 644 variability in the Peruvian upwelling system, 2017.
- 645 Gutiérrez, M., Pantoja, S., Tejos, E., and Quiñones, R.: The role of fungi in processing marine
- organic matter in the upwelling ecosystem off Chile, Mar Biol, 158, 205-219, 2011.
- 647 Hauss, H., Franz, J. M., and Sommer, U.: Changes in N: P stoichiometry influence taxonomic
- 648 composition and nutritional quality of phytoplankton in the Peruvian upwelling, J Sea Res, 73,
- 649 74-85, 2012.
- 650 Hoppe, H.-G., Kim, S.-J., and Gocke, K.: Microbial decomposition in aquatic environments:
- 651 combined process of extracellular enzyme activity and substrate uptake, Appl. Environ.
- 652 Microbiol., 54, 784-790, 1988.
- Hoppe, H.-G., Arnosti, C., and Herndl, G.: Ecological significance of bacterial enzymes in the
- marine environment, Enzymes in the Environment: Activity, Ecology, and Applications, 73-107,
- **655** 2002.
- 656 Hoppe, H.-G.: Phosphatase activity in the sea, Hydrobiol, 493, 2003.
- 657 Jeong, H. J., Du Yoo, Y., Park, J. Y., Song, J. Y., Kim, S. T., Lee, S. H., Kim, K. Y., and Yih,
- 658 W. H.: Feeding by phototrophic red-tide dinoflagellates: five species newly revealed and six
- species previously known to be mixotrophic, Aquat Microb Ecol, 40, 133-150, 2005.





- 660 Kalvelage, T., Lavik, G., Lam, P., Contreras, S., Arteaga, L., Löscher, C. R., Oschlies, A.,
- 661 Paulmier, A., Stramma, L., and Kuypers, M. M.: Nitrogen cycling driven by organic matter
- export in the South Pacific oxygen minimum zone, Nature Geosci, 6, 228-234, 2013.
- 663 Keeling, R. F., Körtzinger, A., and Gruber, N.: Ocean deoxygenation in a warming world, Ann
- 664 Rev Mar Sci, 2, 199-229, 2010.
- 665 Kérouel, R., and Aminot, A.: Fluorometric determination of ammonia in sea and estuarine waters
- by direct segmented flow analysis, Mar Chem, 57, 265-275, 1997.
- 667 Labry, C., Delmas, D., Youenou, A., Quere, J., Leynaert, A., Fraisse, S., Raimonet, M., and
- 668 Ragueneau, O.: High alkaline phosphatase activity in phosphate replete waters: The case of two
- 669 macrotidal estuaries, Limnol Oceanogr, 61, 1513-1529, 2016.
- 670 Margalef, R., Estrada, M., and Blasco, D.: Functional morphology of organisms involved in red
- tides, as adapted to decaying turbulence, in: Toxic dinoflagellate blooms, edited by: Taylor, D.
- 672 L., and Seliger, H. H., Elsevier-North Holland, Amsterdam, 89-94, 1979.
- 673 Maßmig, M., Lüdke, J., Krahmann, G., and Engel, A.: Bacterial degradation activity in the
- eastern tropical South Pacific oxygen minimum zone, Biogeosciences, 17, 215-230, 2020.
- 675 Messié, M., and Chavez, F. P.: Seasonal regulation of primary production in eastern boundary
- upwelling systems, Prog Oceanogr, 134, 1-18, 2015.
- 677 Morris, A., and Riley, J.: The determination of nitrate in sea water, Analytica Chimica Acta, 29,
- 678 272-279, 1963.





- 679 Murphy, J., and Riley, J. P.: A modified single solution method for the determination of
- 680 phosphate in natural waters, Analytica chimica acta, 27, 31-36, 1962.
- 681 Murphy, K. R., Stedmon, C. A., Wenig, P., and Bro, R.: OpenFluor-an online spectral library of
- 682 auto-fluorescence by organic compounds in the environment, Analytical Methods, 6, 658-661,
- **683** 2014.
- 684 Nausch, M.: Alkaline phosphatase activities and the relationship to inorganic phosphate in the
- 685 Pomeranian Bight (southern Baltic Sea), Aquat Microb Ecol, 16, 87-94, 1998.
- 686 Nowinski, B., Motard-Côté, J., Landa, M., Preston, C. M., Scholin, C. A., Birch, J. M., Kiene, R.
- 687 P., and Moran, M. A.: Microdiversity and temporal dynamics of marine bacterial
- dimethylsulfoniopropionate genes, Env Microbiol, 21, 1687-1701, 2019.
- 689 Ordination methods, diversity analysis and other functions for community and vegetation
- 690 ecologists, 2019.
- 691 Oschlies, A., Brandt, P., Stramma, L., and Schmidtko, S.: Drivers and mechanisms of ocean
- 692 deoxygenation, Nature Geosci, 11, 467-473, 2018.
- 693 Pajares, S., Varona-Cordero, F., and Hernández-Becerril, D. U.: Spatial Distribution Patterns of
- Bacterioplankton in the Oxygen Minimum Zone of the Tropical Mexican Pacific, Microb Ecol,
- **695** 2020.
- 696 Perry, M.: Alkaline phosphatase activity in subtropical Central North Pacific waters using a
- 697 sensitive fluorometric method, Mar Biol, 15, 113-119, 1972.





- 698 Pruesse, E., Peplies, J., and Glöckner, F. O.: SINA: accurate high-throughput multiple sequence
- alignment of ribosomal RNA genes, Bioinformatics, 28, 1823-1829, 2012.
- 700 Rose, C., and Axler, R. P.: Uses of alkaline phosphatase activity in evaluating phytoplankton
- community phosphorus deficiency, Hydrobiol, 361, 145-156, 1997.
- 702 Ramin K.I., and Allison S.D.: Bacterial tradeoffs in growth rate and extracellular enzymes. Front
- 703 Microbiol, 10: 2956, 2019.
- 704 Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B.,
- 705 Lesniewski, R. A., Oakley, B. B., Parks, D. H., and Robinson, C. J.: Introducing mothur: open-
- source, platform-independent, community-supported software for describing and comparing
- microbial communities, Appl Env Microbiol, 75, 7537-7541, 2009.
- 708 Schulz, K. G., Achterberg, E. P., Arístegui, J., Bach, L. T., Baños, I., Boxhammer, T., Erler, D.,
- 709 Igarza, M., Kalter, V., Ludwig, A., Löscher, C., Meyer, J., Meyer, J., Minutolo, F., von der Esch,
- 710 E., Ward, B. B., and Riebesell, U.: Nitrogen loss processes in response to upwelling in a
- 711 Peruvian coastal setting dominated by denitrification a mesocosm approach, Biogeosciences,
- 712 18, 4305–4320, 2021.
- 713 Schunck, H., Lavik, G., Desai, D. K., Großkopf, T., Kalvelage, T., Löscher, C. R., Paulmier, A.,
- 714 Contreras, S., Siegel, H., and Holtappels, M.: Giant hydrogen sulfide plume in the oxygen
- 715 minimum zone off Peru supports chemolithoautotrophy, PloS one, 8, e68661, 2013.
- 716 Shah, V., Chang, B. X., and Morris, R. M.: Cultivation of a chemoautotroph from the SUP05
- 717 clade of marine bacteria that produces nitrite and consumes ammonium, The ISME journal, 11,
- 718 263-271, 2017.





- 719 Song, C., Cao, X., Zhou, Y., Azzaro, M., Monticelli, L. S., Maimone, G., Azzaro, F., La Ferla,
- 720 R. and Caruso, G.: Nutrient regeneration mediated by extracellular enzymes in water column and
- 721 interstitial water through a microcosm experiment. Sci Tot Env 670: 982-992, 2019.
- 722 Spilling, K., Camarena-Gómez, M.-T., Lipsewers, T., Martinez-Varela, A., Díaz-Rosas, F.,
- 723 Eronen-Rasimus, E., Silva, N., von Dassow, P., and Montecino, V.: Impacts of reduced inorganic
- N: P ratio on three distinct plankton communities in the Humboldt upwelling system, Mar Biol,
- 725 166, 114, 2019.
- 726 Stedmon, C. A., and Bro, R.: Characterizing dissolved organic matter fluorescence with parallel
- factor analysis: a tutorial, Limnology and Oceanography: Methods, 6, 572-579, 2008.
- 728 Stoecker, D. K., and Gustafson, D. E.: Cell-surface proteolytic activity of photosynthetic
- dinoflagellates, Aquat Microb Ecol, 30, 175-183, 2003.
- 730 Thomson, B., Wenley, J., Currie, K., Hepburn, C., Herndl, G. J., and Baltar, F.: Resolving the
- 731 paradox: continuous cell-free alkaline phosphatase activity despite high phosphate
- concentrations, Mar Chem, 214, 103671, 2019.
- 733 Wood, S. N.: Generalized additive models: an introduction with R, CRC press, 2017.

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738 Figure legends

739

740	Fig 1. The concentration of dissolved inorganic nitrogen (DIN), phosphate (PO <sub>4</sub> <sup>3-</sup> ), dissolved
741	organic nitrogen (DON) and phosphorus (DOP). The red and blue color are the mesocosm bags
742	with deep-water addition with low (closer to shore) and very low (further offshore) oxygen
743	minimum zone (OMZ) signature, respectively. The green dashed lines denote the time of OMZ
744	water addition.
745	
746	Fig 2. The fluorescence dissolved organic matter (FDOM) components (C1-C4) during the
747	experiment. The red and blue color are the mesocosm bags with deep-water addition with low

748 (closer to shore) and very low (further offshore) oxygen minimum zone (OMZ) signature,

respectively. The green dashed lines denote the time of OMZ water addition.

750

(lower graph). The red and blue color are the mesocosm bags with deep-water addition with low

(closer to shore) and very low (further offshore) oxygen minimum zone (OMZ) signature,

respectively. The green dashed lines denote the time of OMZ water addition.

755





757	Fig 4. Development of the main groups of phytoplankton enumerated by flow cytometry. The red
758	and blue color are the mesocosm bags with deep-water addition with low (closer to shore) and
759	very low (further offshore) oxygen minimum zone (OMZ) signature, respectively. The green
760	dashed lines denote the time of OMZ water addition.
761	

762

763	Fig 5. The bacterial community composition in the 8 mesocosms taken at different time points.
764	In the upper row are mesocosms with deep-water from low OMZ signature (30 m depth) and in
765	the second row with very low OMZ signature (90 m depth). The Y-axis indicates the relative
766	abundance of the bacterial taxa. Only the groups that contributed more than 0.5 % of the total
767	sequences are included and the rest are grouped as "Other Bacteria". The classification was
768	performed mainly in class, order and genus levels. The abbreviations indicate the main class
769	levels: Alphaproteobacteria (orange shades), Gammaproteobacteria (blue-pink shades),
770	Deltaproteobacteria (green shades), and Bacteroidia (yellow shades).
771	

- Fig 6. The leucine aminopeptidase (LAP) and cumulative LAP activity. The red and blue colorare the mesocosm bags with deep-water addition with low (closer to shore) and very low (further
- offshore) oxygen minimum zone (OMZ) signature, respectively. The green dashed lines denote
- the time of OMZ water addition.





# 777

Fig 7. The alkaline phosphatase activity (APA) and cumulative APA. The red and blue of
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- the mesocosm bags with deep-water addition with low (closer to shore) and very low (further
- offshore) oxygen minimum zone (OMZ) signature, respectively. The green dashed lines denote
- the time of OMZ water addition.

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783

- Fig 8. Non-parametric multidimensional scaling (NMDS) plots for biochemical, phytoplankton
- community and bacterioplankton community (upper row). From the NMDS scores, generalized
- additive models (GAMs) were made (lower two rows) where we used alkaline phosphatase
- 787 activity (APA) and leucine aminopeptidase (LAP) as dependent variables. The output scores
- (mds1 and mds2) of the NMDS are depicted in the lower two rows.

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800 801 FIG 4 802





803



805 FIG 5







807

808 Fig 6









812







813

814 FIG 8