Effect of land use legacy on the future carbon sink for the conterminous U.S.

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16 Abstract

17 Modeling the effects of the terrestrial carbon sink in the future depends upon not just current-day 18 19 land use and land cover (LULC), but also the legacy of past LULC change (LULCC), which is often not considered. The age distribution of trees in the forest depends upon the history of past 20 disturbances, while the nutrients in the soil depend upon past LULC. Thus establishing the 21 correct initial state of the vegetation and soil is crucial to model accurately the effect of 22 biogeochemical cycling with environmental change in the future. This study models the effects 23 of LULCC from 1750 to 2014 using the Land Use Harmonization dataset (LUH2) of land use 24 transitions with the Terrestrial Ecosystems Model (TEM) for the conterminous U.S. Modeled 25 LULC include plant functional types (PFTs) of potential vegetation, as well as managed 26 cropland, pastureland, and urban areas. LULCC is treated using a cohort approach, in which a 27 28 separate cohort occurs every year there is a land use transition, thereby ensuring proper age 29 structure of forests and regrowth with the correct soil nutrients. From 2000-2014 the modeled 30 Net Ecosystem Productivity (NEP) is 989 TgCyr⁻¹ for the conterminous U.S., but only -15 TgCyr⁻¹ if accounting for carbon lost from land use transitions and management. 31 32 33 The hypothesis is that the initial state of the vegetation and soils significantly affects the future

state of the terrestrial carbon sink. In this study, LULC remains constant in the future, with the 34 NCAR CCSM4 RCP8.5 climate used to force the TEM-Hydro model. The following 35 experiments are run from 2015 to 2100, including a) restarting from existing cohorts in 2014 36 (RESTART), b) reinitializing in 2015 based on condensing the cohorts for each PFT into a single 37 38 cohort (CONDENSED), and c) restarting from average cohort conditions for each PFT (AVERAGE). The NEP is too low when using condensed cohorts without reinitializing due to a 39 larger increase in heterotrophic respiration (R_h) resulting from the assumption of mature forests. 40 The carbon stocks are overestimated larger than using all the cohorts if condensed cohorts are 41 42 reinitialized due to the assumption of mature, equilibrated forests. Where nitrogen-limited, 43 forest regrowth is enhanced if regrowth starts from more nutrient-rich conditions. Water fluxes are dominated by environmental factors, but can be slightly dependent upon the underlying 44 carbon dynamics. It is therefore necessary to account for past disturbances when modeling 45

46 future changes in carbon dynamics.47

48 1 Introduction

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Globally, during the 21st century, land use and land cover change (LULCC) has accounted for 50 14% of the total anthropogenic carbon emissions (Friedlingstein et al., 2020). LULCC have 51 52 been responsible for the largest losses of carbon from the land in the conterminous U.S. since the 53 1700s, with growth enhancements from CO_2 fertilization and nitrogen deposition only partially countering this loss since the 1950s (Felzer and Jiang, 2018). Reforestation and afforestation 54 have been the primary drivers for this enhanced sink (Kondo et al., 2018), especially growing 55 back with rising CO₂ levels (Strassmann et al., 2008). This paper addresses the question of the 56 57 role of land legacy in the future carbon sink in conterminous U.S. How inappropriate is it to initialize a model with current-day land use and land cover (LULC) for a 21st century simulation, 58 which avoids the disturbance history and forest recovery from the 20th century and earlier? 59 60

61 Many modeling studies have been conducted to explore the role of LULCC relative to other environmental factors like CO2 fertilization, N deposition, and ozone both historically and into 62 the future. For example, studies have shown LULCC to be the most important cause of reduced 63 64 carbon inventory in the future due to loss of forest (Mahowald et al., 2017), while CO2 fertilization increases the sink (Tharammal et al., 2019). Reforestation, including regrowth from 65 timber harvest, and avoided deforestation, can increase the carbon sink in the future (Arneth et 66 al., 2017; Zhao et al., 2013). Remotely sensed data from 1973-2010 have shown that both 67 reduced forest area and older forest age have contributed to a reduced C sink in the conterminous 68 69 US (Sleeter et al., 2018). Legacy carbon fluxes from deforestation can be in the form of emissions from dead biomass, soils, and forest products, or uptake in regrowing secondary 70 forests (Houghton et al., 2012). 71 72 73 Only a few models (e.g. (Felzer and Jiang, 2018; Shevliakova et al., 2009)) have included forest 74 demography, to accurately track the effects of disturbance in regrowing forests. Krause et al.

75 (2020) showed that including land legacy effects increases future carbon storage as ecosystems 76 regrow and adapt to higher levels of CO2 and N deposition. Since ecosystems are not in 77 equilibration with current-day land use, there will be continued carbon uptake even if climate 78 change and land use are held constant, due to regrowth from abandoned agriculture and CO₂ fertilization (Krause et al., 2019). Pugh et al. (2019) surmiseds that there will be a large carbon 79 sink from regrowth in the future regardless of environmental change as long as current 80 disturbance rates continue at historical levels. Lu et al. (2015) found that using corrected Forest 81 82 Inventory Analysis (FIA) data (Pan et al., 2011) applied to a dataset of annual land use 83 transitions (Hurtt et al., 2011) nearly doubled the carbon sink due to younger forests in the corrected data. Thom et al. (2018) points out that it is important to develop initial conditions to 84 account for past disturbance in order to capture the observed state. This idea is tested in the 85 current study by determining the difference in future carbon sink between initial conditions that 86 87 do capture disturbance since 1750 and reinitialized initial conditions.

88 89 Two factors that determine the carbon sink strength of regrowing forests are the stand age 90 distribution of the trees in the forest and the nutrient levels of the soil. The age distribution depends upon the timing and magnitude of past disturbances. Soil nutrient conditions depend 91 upon the prior history of land use and management. Several studies show that forest regrowing 92 93 from nutrient-rich fertilized agricultural land exhibit less resilience for climate change but higher growth rates. European beech trees on former agricultural land had lower C:N and higher P, 94 which resulted in less carbon allocation to roots, reducing resilience to drought (Mausolf et al., 95 2018). Similarly, Von Ohemib (2014) found these same changes led to higher tree ring width 96 97 due to more litter decomposition and higher N mineralization rates, as well as reduced resiliency. In terms of Net Ecosystem Productivity (NEP), reforestation sites exhibited reduced NEP due to 98 loss of carbon from the forest floor or soils during early recovery (Pan et al., 2011) but enhanced 99 NEP in afforestation sites due to replacement of depleted pools (Post and Kwon, 2000). 100 101

This study explores the question of land legacy on the future carbon sink by comparing model simulations with full forest demography with those based on reinitializing initial conditions to the present. The analysis looks at both carbon fluxes and stocks to determine how these vary regionally and integrated over the entire conterminous U.S. It explores the role of forest stand

age and soil nutrients in determining forest regrowth and tests the hypothesis that it is crucial to capture the effects of historical land legacy in order to accurately model the future carbon sink.

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110 2 Methods

This study uses the Terrestrial Ecosystems Model- Hydro Version 2 (TEM-HYDRO2) to explore
the role of historical land use legacy (from 1750 to 2014) on future (2014-2099) carbon storage.
The recent LUH2 version of land use transitions (Hurtt et al., 2020) is used to reconstruct the full
cohort of LULCC since 1750, while LULC is kept constant for the 21st century. Three sets of
experiments explore the role of fully accounting for past land legacy, reinitializing initial
conditions and not accounting for land legacy at all, and initial conditions based on averaging the
final state of the full cohorts in 2014 to determine if corrected initial conditions are sufficient.

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120 2.1 Model Description

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122 The Terrestrial Ecosystems Model version Hydro (TEM-Hydro – (Felzer, 2012; Felzer et al.,

2009; Felzer et al., 2011) is a fully prognostic biogeochemical model of carbon, nitrogen and

124 water dynamics between vegetation and soils. A complete description of the model can be found

in Felzer et al. (2009) (2011) and Felzer (2012). The model structure is illustrated in summary figures (Fig. S1a) along with how human disturbance is treated, which is relevant to this paper

(Fig. S1b). A cohort approach is developed to convert a dataset of land use transitions (Hurtt et

al. (2011; 2020) to annual cohorts of land use and land cover change (Hayes et al., 2011; Lu et

al., 2015), whose purpose is to retain the soil characteristics of the cohort from which disturbance

130 occurred and maintain appropriate growth and stand age of newly developed cohorts (Fig. S2a).

131 <u>This approach involves first using the LUH2 dataset to establish the fractional land cover type at</u>

the starting year of 1750. The primary and secondary vegetation are replaced with their potential

vegetation values (as described in (Raich et al., 1991)), while other managed lands include
 croplands, pasturelands, and urban, with the multiple types of crops and pastures combined into

single values for each, respectively. Disturbances (including timber harvest) involve the creation

of new cohorts, with the corresponding area adjusted from the original cohort. Therefore, soil

nutrients and forest stand age are tracked separately for each disturbance. The output are then

area-weighted for each of the cohorts. Since this approach tracks each cohort separately, it is

139 possible to end up with thousands of cohorts for a single grid cell by 2014. -A complete

description of this approach can be found in Felzer and Jiang (2018). New to this study is that the initial vegetation is started in 1750 (consistent with Allan et al. (2021) baseline period) and

subsequent transitions were determined until 2014 (Fig. S2b, c, d, e) to align with the temporal

range of climate datasets. The result for a single grid cell is usually hundreds of cohorts by the year 2014, accounting for all transitions between primary and secondary vegetation, cropland,

145 pastureland, and urban areas, as well as timber harvest.

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147 The partitioning of disturbance products and fluxes for agriculture and timber harvest and

148 management practices and calibration are described in Felzer and Jiang (2018). In this study

both croplands and turflawn (urban) are fertilized (using the approach taken in Felzer et al.

150 (2018), while no additional fertilization (beyond that provided by livestock) is applied to pasture.

151 A few additional modifications were made for this study. Irrigation was added to arid croplands,

because inorganic nitrogen was accumulating due to lack of leaching. The same scheme as used
in Felzer (2012) for turflawn was applied to croplands receiving less than 200 mm of water per
month during the growing season. The other change applies to abandoned cropland. Cropland
abandoned before there was major chemical fertilization in the 1960s were too nutrient depleted
in the model, and the forest regrowth occurred with reduced biomass, so 15 gN/m2/month during
the year of disturbance was added following crop abandonment to ensure at least limited forest
regrowth.

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160 2.2 Experimental Design

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162 SixFour simulations (Table 1) were designed to determine the effect of land legacy. The HISTORICAL run applies the full cohorts from 1750 to 2014, allowing for the Hurtt et al. (2020) 163 164 record of LULCC as described in the Methods. The HISTCONST run is the HISTORICAL run 165 but with LULC held constant at 2014 value, so includes other effects related to climate, CO2, N 166 deposition and ozone. Therefore the difference between HISTORICAL ad HISTCONST is the 167 effect of LULCC. The HISTCOND run is a transient run like HISTORICAL except that the 168 Plant Functional Types (PFTs) are condensed to a single cohort for each PFT in a give year. It is 169 essentially the fractional land cover per year, so the difference between HISTORICAL and HISTCONST illustrates the effect of forest demography. Because multiple land-use transitions 170 171 are incorporated into single cohorts, it is not possible to accurately incorporate the true 172 disturbance, so conversion and product fluxes results from land-use change are not included in 173 this run. The RESTART run uses restart files from the full suite of cohorts in 2014 to run from 174 2015 to 2099, keeping LULC constant with the 2014 cohorts. This run is essentially just a 175 continuation of the HISTORICAL run. The CONDENSED run reinitializes (i.e. reequilibrates) a 176 condensed version of the cohorts in 2014 to provide initial conditions for the 2015 to 2099 177 period. In this run, the 2014 cohorts are condensed to a single cohort for each pft (with primary 178 and secondary of the original PFT tracked separately), with the fractional areas determined based 179 on the 2014 cohorts. These condensed cohorts are then each reequilibrated at the start. The 180 TEMRESTART run uses a restart file for 2014 that is based on the average of the restart conditions for each of the cohorts, and then uses the condensed cohorts for the 2015 to 2099 181 period. Thus the TEMRESTART run uses the same number of cohorts as the CONDENSED 182 run, but does not reequilibrate at the start. So both the CONDENSED and TEMRESTART runs 183 184 used the simplified, condensed cohorts, but start with different initial conditions. The motivation 185 for these two condensed-PFT runs is to reduce computational time by eliminating the need to run 186 potentially thousands of land-use legacy cohorts for each grid when starting from present-day 187 conditions. The difference between the RESTART and CONDENSED runs shows the effect of 188 including land legacy on future carbon dynamics. Note that the RESTART run will also incorporate effects of changing climate, CO2, ozone, N deposition and fertilization, which cannot 189 be captured in the CONDENSED run. The TEMRESTART run shows if it is possible to 190 condense the initial conditions from a full suite of cohorts to produce the same results as the 191 **RESTART** run. 192

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194 Table 1: Model Experiments195

Experiment	Number Cohorts	Initialization	Time Period
HISTORICAL	Transient	Equilibrate 1750	1750-2014

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RESTART	Full at 2014*	Continuation of	2015-2099
		HISTORICAL	
CONDENSED	Condensed**	Equilibrate 2015	2015-2099
TEMRESTART	Condensed***	Average from	2015-2099
		HISTORICAL 2014	

196 * Maximum cohorts for a grid in 2014 is 1020

197 ** Maximum cohorts for a grid is 7, because primary vegetation is treated separately from 198 secondary

199 *** Maximum cohorts for a grid is 5 (e.g. mixed potential vegetation with two cohorts, cropland, 200 pasture, urban)

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Experime	HISTORIC	HISTCON	HISTCON	RESTAR	CONDENS	TEMRESTA
<u>nt</u>	AL	<u>ST</u>	<u>D</u>	<u>T</u>	ED	<u>RT</u>
LULCC	Transient	Constant	Transient	Constant	Constant	Constant
		<u>2014</u>		2014	<u>2014</u>	2014
Number	Transient	Condensed	Condensed	Full at	Condensed	Condensed*
Cohorts		**	***	2014*	**	**
Initializati	Equilibrate	Equilibrate	Equilibrate	<u>Continuati</u>	Equilibrate	Average
on	<u>1750</u>	<u>1750</u>	1750	<u>on of</u>	2015	from
				Historical		HISTORIC
						AL
Time	1750-2014	1750-2014	1750-2014	2015-	2015-2099	2015-2099
Period				2099		
<u>CO</u> ₂	<u>CRU4.04</u>	<u>CRU4.04</u>	<u>CRU4.04</u>	<u>RCP8.5</u>	<u>RCP8.5</u>	<u>RCP8.5</u>
<u>N</u>	Tian et al.	Tian et al.	Tian et al.	<u>Constant</u>	<u>Constant</u>	Constant
Depositio	<u>2010</u>	<u>2010</u>	<u>2010</u>	2014	<u>2014</u>	<u>2014</u>
<u>n</u>						
N	Felzer et al.	Felzer et	Felzer et al.	Constant	Constant	Constant
<u>Fertilizati</u>	<u>2018</u>	<u>al. 2018</u>	<u>2018</u>	2014	<u>2014</u>	<u>2014</u>
on						
(crops/tur						
<u>f)</u>						
Ozone	Felzer et al.	Felzer et	Felzer et al.	Constant	Constant	Constant
(AOT40)	2004	al. 2004	2004	2014	2014	2014

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203 ** Maximum cohorts for a grid is 7, because primary vegetation is treated separately from secondary

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205 *** Maximum cohorts for a grid is 5 (e.g. mixed potential vegetation with two cohorts, cropland, 206 pasture, urban)

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209 The model is run monthly at a spatial resolution of 0.5° x 0.5°. Input datasets include transient

210 climate (surface air temperature, diurnal temperature range, precipitation, fractional cloud cover

211 to derive net irradiance at the surface and photosynthetically active radiation (PAR), vapor

212 pressure), climatological wind speed (as in (Felzer et al., 2011)), and annual atmospheric CO₂

from 1901-2014 based on CRU4.04 (Harris et al., 2014). Gridded transient climate data are not 213

214 available prior to 1901, so climate variables from 1750 – 1849 are taken from the MPI-ESM-P 215 past 1000-year simulation and 1850-1900 from the MPI-ESM-P historical simulation (Schmidt et 216 al., 2014)(Schmidt et al., 2014). The downscaling and bias correction is similar as to what was 217 done in Felzer and Jiang (Felzer and Jiang, 2018), but starting in 1750 instead of 1700, using the 218 CRU4.04 data from 1901-1930. The resultant U.S. mean climate from 1750 is shown in Fig. S3. 219 Surface ozone (Felzer et al., 2004), nitrogen deposition (Tian et al., 2010), and soil 220 texture and elevation datasets are similar to those used in Felzer et al. (2011). 221 222 The future climate data (2015-2099) are taken from the Multivariate Adaptive Constructed 223 Analogs (MACA) statistically downscaled Coupled Model Intercomparison Project 5 (CMIP5) 224 data (Abatzoglou and Brown, 2012)(Abatzoglou and Brown, 2012), using the National Center for Atmospheric Research (NCAR) Community Climate System Model version 4 (CCSM4) 225 226 RCP8.5 emissions scenario (r6i1p1 ensemble). The downscaled resolution is at 4 km but has been extrapolated to the half-degree TEM grid for this study by averaging over all the 4 km 227 228 values within the larger half-degree grid cell. Net irradiance is used instead of clouds for the 229 future data. The TEM cloud scheme was adjusted for the historical cloud data to bias-correct to 230 ensure continuity of net irradiance between the historical and future data. The CRU4.04 data 231 does not include irradiance, which is why it was necessary to use clouds for the historical period, but since net irradiance is more directly used by the model, that was chosen for the future period. 232 The results (Fig. S3) show a continuity for climate during the transition between the historical 233 234 CRU4.04 and future RCP8.5 in 2014 for all the variables. Future RCP8.5 CO₂ data are taken 235 from CMIP5 recommendations (Meinshausen et al., 2011) Meinshausen et al. (2020). The ozone 236 and N deposition values are kept at their 2014 levels (which are held constant after 2000 for 237 ozone). 238 239 The decision to base climate prior to 1900, prior to the gridded historical data, was made to 240 capture more realistic climate variations during the period from 1750 to 1900, such as the Little Ice Age (LIA), which lasted through the 19th century (Bradley and Jonest, 1993; Mann, 2002). 241 The temperature record from the MPI-ESM-P model does show signs of temperature climbing 242 out of a cold peak after 1818 but remaining cool throughout the rest of the century (Figure S3), 243 244 which is consistent with Northern Hemisphere proxy records (Mann et al., 2008). Since this 245 study is for the conterminous U.S., it does not show as strong an LIA signal as would be 246 expected from records in the North Atlantic. The decision to then use historical CRU4.04 247 climate rather than modeled climate from 1901-2014 is to more accurately capture the true 248 interannual variability, which would be entirely lost by using output from a climate model. All 249 three datasets have been downscaled and bias corrected to produce a seamless record of climate 250 from 1750-2099. 251 252 The model is initially calibrated for specific PFTs without disturbance, though with agricultural 253 and urban management where necessary, to determine coefficients for the flux equations before 254 extrapolation to the entire U.S. Note that each experiment is not calibrated individually. The 255 HISTORICAL run is first equilibrated based on repeated use of the 1750-1779 climate in order

to establish initial conditions of carbon and nitrogen stocks (which are required to numerically
solve the fundamental model equations), and then the transient runs are started from 1750 to
2014. The CONDENSED run is first equilibrated based on repeated use of the 2016-204519862015 climate from the HISTORICAL run, and the transient runs are from 2015 to 2099. Results

of NEP or <u>Net Carbon Exchange (NCE)</u> fluxes are reported as TgCyr⁻¹, while cumulative NCE, a

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261 measure of net carbon accumulation over some time periods, is reported as PgC. NCE is the NEP

- 262 plus carbon lost through land-use conversion or by decomposition of agricultural or timber 263 harvest products. Model input, forcing data and output results are publicly available at
- 264 http://go.lehigh.edu/landlegacy.

265 266 3. Results

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268 The historical (1750-2014) NEP (from HISTORICAL) starts to increase in the 1870s (Fig. 1),

consistent with the time period when CO₂ levels start to increase and there is a slight warming, 269

270 though there is also a decrease in precipitation during this period (Fig. S3). The separation of

271 NCE from NEP signifies the results of LULCC, which become more pronounced after the 1850s

272 when timber harvest begins and pasture and cropland increase at the expense of forest, pasture, 273

and grassland (Fig. S2a). The cumulative NEP is 87 PgC, while the cumulative NCE is -42 PgC. So climate and CO₂ conditions cause the land to be a net carbon sink, but LULCC makes the 274

275 land a net carbon source.





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Figure 1: Net Carbon Exchange (NCE) and Net Ecosystem Productivity (NEP) for the
 HISTORICAL and RESTART runs. NCE includes fluxes from agricultural conversion and
 abandonment and decomposition of agricultural products.

285 The effect of including LULCC is evident in the difference between HISTORICAL and

HISTCONST (Fig. 2). While the final cumulative NEP is close by the year 2014, the use of

actual land-use transitions lowers the NEP, especially during the early years, consistent with the
 results of Felzer and Jiang (2018) that the effect of deforestation reduces the NEP, while the

289 larger area of mature forest do not contribute much to positive NEP. The vegetation and soil

carbon start out substantially higher in HISTORICAL, while without LULCC they remain

relatively constant out substantiany night in ThSTORICAL, while without ECECC they remain relatively constant in HISTCONST, which shows the effects of the other environmental changes

192 like climate, CO_2 , N deposition, and ozone. The HISTCOND compared to HISTORICAL shows

that the inclusion of forest demography does increase the cumulative NEP by 2014 and lowers

the vegetation and soil carbon estimates, as changing forest area is incorporated into existing

295 forests rather than separate cohorts to allow for forest regrowth.

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301 Figure 2: Historical experiments a) Cumulative NEP (TgCyr-1), b) Vegetation carbon (vegc) 302 and soil carbon (solc) in TgC. Experiments are HISTORICAL, HISTCONST, and HISTCOND 303 (see text).

In the future runs, the RESTART run is considered the "actual" to validate the others against, as 305 306 it is the run that includes effects of all the individual cohorts. The CONDENSED run is the 307 effect of condensing all the cohorts to single PFTs and the TEMRESTART is the result of 308 averaging the initial conditions for each of the cohorts in 2014. The NEP and NCE of the CONDENSED is lower than the RESTART and TEMRESTART, especially at the start of the 309 310 runs (Fig. 32), because reinitializing each grid is based on the assumption of NEP as close to 311 zero as possible. The cumulative result in 2099 is NEP of 76 PgC in the RESTART run, 80 PgC 312 in the TEMRESTART, and 630 PgC in the CONDENSED. The cumulative NCE of the 313 RESTART and TEMRESTART is close beyond the starting years, resulting in 20 and 18 PgC 314 respectively, while it is lower (9.67.8 PgC) for the CONDENSED run. NCE still differs from 315 NEP without LULCC because of crop decomposition, animal respiration, and crop residue 316 fluxes. Since there is no product decomposition in the CONDENSED run, the NCE is equal to 317 the NEP. NCE of the RESTART and TEMRESTART runs are much lower than NEP of those 318 runs because of product decomposition left over from the HISTORICAL run. -By the end of the 319 century there is no significant differences in the annual carbon fluxes or the cumulative NEP, but 320 the condensed run has significantly lower cumulative NEP and NCE than the other runs (Fig. 32 e,f). These results show that averaging the initial conditions is a good way to reduce cohort 321 322 complexity. The mapped patterns (Fig. 43) show that large positive NEP differences between 323 the CONDENSED and RESTART runs occur in the upper Midwest and central California, 324 which are dominated by cropland (Fig. S2b). This results from the reinitialization process in 325 which the NPP of cropland starts out larger than after accounting for transient conditions. 326 Forested areas in the Southeast are lower NEP in the CONDENSED, which would be expected 327 of more mature forests. Differences in the rest of the country are minor. The largest differences 328 in NCE are the negative differences in the Southeast corresponding to the NEP differences there. The lower NEP in the CONDENSED run is the result of larger heterotrophic respiration (R_h) 329



more than offsetting slightly larger Net Primary Productivity (NPP). Since NEP is the difference 330 331 332 between NPP and R_h , the net effect is a negative bias in NEP (Fig. <u>54</u>).

338 Figure 32: Comparison of NEP and NEC between the RESTART, CONDENSED, and

339 TEMRESTART runs, a) NEP, b) cumulative NEP, c) NCE, and d) cumulative NCE, e) NEP,







Figure 54: Carbon fluxes (NEP, NPP, Rh) for the RESTART and CONDENSED runs, , b)
 mean differences 2070-2099 (no significant differences for either of the three pairs).

While the more mature forestsed in CONDENSED would be expected to have lower NEP
(Besnard et al., 2018; He et al., 2012), they would also have more biomass. By the end of the
century regrowing forests in the RESTART run will still be younger than those in CONDENSED

run, and 85 years is not enough time to reach full equilibration in the model. The CONDENSED

- 361 vegetation carbon is 146% higher than the RESTART value by the year 2099, while the
- 362 TEMRESTART is only 5% higher (Fig. <u>65</u>). This large bias <u>The larger values</u> in the
- 363 CONDENSED run is due to the fact that the larger percentage of mature trees (since all trees are
- 364 considered mature in the CONDENSED run) result in much more biomass. Starting with

365 averaged initial conditions fixes most of the problem lowers the vegetation carbon so that it is 366 close to that of using the full cohorts. The soil carbon is $3\underline{1}2\%$ higher in the CONDENSED run, 367 while differences are minimal with the TEMRESTART run (Fig. 65). Note that the absolute 368 differences are larger with vegetation carbon, while the percent differences are more similar 369 since the soil carbon has lower absolute values. The mapped pattern of vegetation carbon 370 differences between the CONDENSED and RESTART runs (Fig. 76a) shows that the large positive bias results almost entirely from the eastern half of the U.S., especially in the forested 371 372 eastern portion, while the West exhibits smaller negative biases. The soil carbon differences 373 (Fig. 76b) are more scattered, with largest positive biases along the East coast and negative biases largest in the Southwest U.S. or Great Plains. 374





380 Figure 65: Vegetation and soil carbon in the RESTART, CONDENSED, and TEMRESTART 381 experiments, b) mean differences 2070-2099 (all three vegetation carbon and soil carbon differ

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to 9339 gCm-2) as differences between the CONDENSED and RESTART experiments.

The keys to these differences are the distribution of stand age in the forests and nutrients in the 391 soil during regrowth. Forest stand age in 2014 at the start of the future runs (when there is no 392 393 further disturbance) shows that while the largest bin of tree area is mature trees (> 500 years old), the next largest class is young trees less than 11 years old, with a majority of tree area less 394 395 than 71 year old, based on the disturbance history of the Hurtt et al. (2020) dataset (Fig. 87a). 396 However, the majority of mature forests are in the Western U.S. Most of the forests in the 397 eastern U.S. are under 30 years old (Fig. 87b, S4). The biomass is generally larger for the more 398 mature categories (Fig. 98a,b). More mature trees are therefore more important to determining 399 biomass than an even relatively large portion of younger trees. While biomass generally

400 increases with stand age, NEP peaks between 11-30 years (Fig. 98c,d). When classifying 401 vegetation carbon by PFT (Fig. 109a), the CONDENSED run values are larger than the 402 RESTART values for boreal forest and temperature coniferous, deciduous, mixed, and 403 broadleaved evergreen forests, as well as savanna (which is a mixture of grassland and trees). 404 The NEP differences between CONDENSED and RESTART runs (Fig. 109b) shows NEP is 405 generally lower in the CONDENSED runs since each cohort has been reinitialized at the start, 406 but the interannual variability (IAV) is much larger than the differences.





411 Figure 87: a) Stand-age frequency for U.S. and b) eastern forests. Bins represent 0-10, 11-20,

412 21-30, 31-40, 41-50, 51-60, 61-70, 71-80, 81-90, 91-100, 101-500, > 500 years.

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419 Figure 98: Vegetation carbon in the year 2014 for a) U.S. and b) eastern forests, and NEP in the year 2014 for c) U.S. and d) eastern forests. Most trees are not mature, but the mature trees

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433 pasture, CR = crops. Error bars are 10-year interannual variability, computed as 95% confidence

Figure <u>109</u>: a) Vegetation carbon by plant functional type in 2015 for the RESTART and for

shrublands, XF = xeric forests and woodlands, DE = deserts, SA = temperate savannas, BE =

temperate broadleaved evergreen forests, MS = Mediterranean shrublands, TL = turflawn, PA =

the CONDENSED experiments, b) NEP by PFT averaged 2015-2024. PFTs are:TU = tundra, BF

= boreal forest, MF = mixed temperate forest, CF = temperate coniferous forest, DF = temperate deciduous forest, TG = tall grasslands, SG = short grasslands, TS = tropical savanna, AS = arid

intervalstandard deviation of year 2015-2024 for each PFT in each of the two runs. 434

435 The inorganic nitrogen in the soil is crucial for regrowth following disturbance. Illustrated here 436 437 (Fig. $1\underline{10}$) is the amount available when disturbance occurs just before forest regrowth vs the 438 final vegetation carbon for that cohort in the year 2100, since no further disturbance occurs in the future. Only values of inorganic nitrogen < 10000 mgN/m2 are shown, because larger values of 439 available nitrogen are not limiting to forest growth. It is evident that larger amounts of initial 440 inorganic nitrogen generally lead to greater forest growth, although there is a wide range in the 441 slope of that relationship. There are also many cohorts that have low growth regardless of initial 442 nitrogen levels, so they are limited by other climate or environmental factors. This is only 443 illustrated for the more mesic forests of the eastern U.S. where moisture is less limiting. The 444 445 final amount of available inorganic nitrogen in 2100 will be compensated by the fact that mature 446 forests provide more nutrients because of the greater litter but also use more nutrients due the 447 higher biomass. 448



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Figure 110: Inorganic nitrogen available for plant uptake immediately following disturbance
before forest regrowth vs the final vegetation carbon by 2100. Shown here are value so
inorganic nitrogen less than 10000 mgN/m2.

The soil moisture is based on a bucket model and accounts for the excess of precipitation over
evapotranspiration, with runoff resulting if the bucket (whose capacity equals the difference
between field capacity and wilting point) is overflowed. The soil moisture of the CONDENSED
run over the last 30 years is too large by 1.0%, whilenot statistically different from RESTART,

460 while the TEMRESTART is has a positive bias of 1.8% by the end of the century higher during

461 that time period (Fig. 124a). The evapotranspiration flux of the condensed CONDENSED run is

too low <u>compared to RESTART</u> while it is too high in the TEMRESTART run, but the runoff

fluxes are nearly identical between the three runs (Fig. 124 b,c).



Figure 121: a) Soil moisture b) runoff, and c) evapotranspiration between the RESTART,
CONDENSED, and TEMRESTART experiments, d) mean differences 2070-2099 for soil
moisture, and e) mean difference 2070-2099 for runoff and evapotranspiration (letters based on
ANOVA analysis with P<0.05).

473

474 4 Discussion

476 The measured stand age frequency in the U.S. is given in Pan et al. (2011) for different regions 477 of the U.S. The eastern regions are dominated by younger trees, the Rocky Mountains by more 478 mature trees as well as a peak in very young trees, and the West coast more younger and mid-age trees. Lu et al. (2015), using a similar LULCC dataset as used here based on Hurtt et al. (2011) 479 land use transitions, specifically corrected that dataset to better represent the data from Pan et al. 480 (2011). The resulting correction was younger forest stand ages in the eastern U.S. after 1850, 481 with overall younger stand ages in the conterminous U.S. as a whole. In fact, the stand age 482 483 distribution for the NE U.S. before the correction (Fig. S2 in (Lu et al., 2015) shows most forest older than 70 years, whereas the Pan et al. (2011) data show most forests are younger. The more 484 recent land use dataset developed from Hurtt et al. (2020) actually shows a majority for forests in 485 486 the eastern U.S. less than 70 years old (Fig. 87), but for the conterminous U.S. the frequency of 487 mature forests is larger because of forests in the western U.S. 488

489 The total biomass increases with age, such that more mature trees have higher amounts of

490 vegetation carbon (Chapin Iii et al., 2011; Pan et al., 2002), consistent with the results presented 491 here (Fig. <u>98</u>a,b). The slight decrease in biomass for some of the more mature stand age classes 492 can represent the differences between geographic areas in which different classes dominate, as biomass for similar trees will be larger under more favorable climate conditions. For example, 493 more mature trees in intermountain forests in the Western U.S. may be expected to have less 494 biomass than less mature trees in the more mesic Eastern U.S. In the eastern U.S. the 101-500 495 year class, for example, the reduction in biomass is due to trees in the northeast (Fig. S5). Note 496 497 that there is no explicit mortality modeled in TEM-Hydro, so biomass in mature forests is not 498 decreasing because of increased mortality, which is another cause for reduced biomass in old 499 stands (Xu et al., 2012). The mapped differences at the end of the 21^{st} century (Fig. 43, 76) 500 represent the aging of all forests in the experiments, so the age distribution in the RESTART run 501 will nowwould be shifted upward by 70 years, so all the forests will be in the upper age 502 categories in both RESTART and CONDENSED runs. Positive biomass differences in the 503 eastern U.S. (Fig. 76a) may represent the even more mature status of the forests in the CONDENSED runs in that region. Forests in the CONDENSED run would be expected to have 504 505 lower NEP since they are more mature, which is generally true of forests, especially in the 506 Southeast U.S. (Fig. 43a), but by the end of the century all the forests have matured more in the 507 RESTART run as well, so differences are more muted with time.

509 NEP generally peaks between 20 and 30 years stand age, yet remains positive for hundreds of

years (Luyssaert et al., 2008). The TEM-Hydro results from the HISTORICAL run show
 maximum NEP occurring between 11-30 years for temperate forests across the U.S. or up to 40

years in the eastern U.S. (Fig. $\frac{98}{28}$ c,d), with NEP generally remaining positive except for very old

trees when including the Western U.S. In fact for the conterminous U.S. as a whole, Lu et al.

514 (2015) found that the Pan et al. (2011)-corrected data, with much younger stand age distribution,

had a cumulative NCE of 323 TgC/yr from 2001-2005 vs 173 TgC/yr with the uncorrected data

516 derived from Hurtt et al. (2011). The RESTART and TEMRESTART runs show continued

517 carbon uptake in the future (Fig. $\underline{32}$), consistent with Krause et al. (2020)) who point out that

regrowth, as well as climate change and elevated CO₂, will continue to promote carbon uptake even in the absence of future land use change. Houghton et al. (2012) also explains that future

520 carbon uptake is dominated by land legacy effects.

508

521 522 The interannual variability of fluxes, like NEP and NCE is very large (4930-5797 TgC/yr 523 standard deviation, or over (6370-75 gC/m2/yr for the three runs for NEP), so the differences 524 between the experiments are all within the interannual variability. These values are consistent with other measured values. For a range of 24 eddy covariance sites, standard deviation of 525 annual Net Ecosystem Exchange (NEE) ranged from about 20-280 gC/m2/yr, accounting for 526 50% of annual NEE (Niu et al., 2017). IAV from site-level FLUXNET sites mostly in North 527 America and Europe ranged from 15 to 400 gC/m2/yr (with a mean of 130 gC/m2/yr), with 528 lower values in more northern sites, and a lower range of values from global upscaling and 529 inversion models (Marcolla et al., 2017). Climate drivers, particularly temperature and moisture, 530 are considered the primary drivers for this large IAV (Piao et al., 2020). In any case, differences 531 between the experiments in this study (Fig. 32) are all much smaller than the IAV, but the 532 533 different experiments are well correlated, so the differences represent a shift of the entire time series, rather than a change in IAV. 534

536 The effect of nutrient loading on abandoned land, such as fertilization on abandoned cropland, 537 can increase the final growth of the forest, but final growth rates are dependent upon many other 538 environmental factors as well, which is why the relationship does not hold true everywhere, and above a certain level of nutrient availability, the system is not nitrogen limited, so it does not 539 matter at all (Fig. 110). Other studies have confirmed that increased nutrients availability, in the 540 form of lower C:N and C:P or high P, promotes radial stem growth (Mausolf et al., 2018) or tree 541 ring width (Von Oheimb et al., 2014), which is consistent with the biomass results from this 542 543 study. The greater nutrient availability, by directly increasing GPP, would also result in more litter and therefore more litter decomposition and higher rates of net nitrogen mineralization, also 544 545 consistent with Von Ohemib (2014). However there is also a legacy effect of reduced resiliency 546 to drought, having to do with changes in soil structure, which would not occur in the model 547 development here.

549 Most other terrestrial ecosystem models do not include the effect of forest demography. The 550 551 552 Dynamic Global Vegetation Models (DGVM) included in Trends in Net Land-Atmosphere Exchange (TRENDY-v2) (Li et al., 2017) mostly include annual changes in PFTs to represent LULCC. They include the conversion and product fluxes resulting from these changes, and 553 often include the effects of mortality and regrowth within existing grids, but do not incorporate 554 the effects of forest regrowth due to LULCC. Two of the models (VISIT and JSBACH) (Kato et 555 al., 2013; Reick et al., 2013) include elaborate methods of applying the LULCC transition 556 matrices to ensure the correct redistribution of PFTs and correct carbon fluxes. Shevliokova et 557 al. (2009) does use a tiling approach to consider forest stand age and reduce the large number of 558 cohorts used here. The HISTCOND run was designed specifically to explore the effects of forest 559 demography by trying to emulate the effect of just redistributing annual land-use fractions, 560 without including the effect of forest demography or keeping track of soil nutrients. As seen in 561 the results, it does substantially overestimate the carbon stocks and underestimate the NEP 562 compared to the run that includes the full effects of forest demography. 563

564 Restarting from averaged initial conditions more closely approximates the full cohort approach 565 with a large computation advantage by avoiding the need for reinitializing and enabling the use 566 of condensed cohorts, but with the corrected initial conditions. In the fluxes (Fig. 32),

535

cumulative NEP of TEMRESTART is higher than the RESTART run, but cumulative NCE of 567 the TEMRESTART is nearly the same as the RESTART run in the latter half of the century. 568 The vegetation carbon of TEMRESTART diverges slightly from RESTART, while the soil 569 carbon barely diverges at all (Fig. 65). 570 571 572 To address the issue of discontinuity between using clouds as input for the historical period (1750-2014) and net irradiance for the future (2015-2099), an additional FUTURE run was 573 implements to use clouds for the future period as well. The reason for using clouds historically is 574 575 because net irradiance is not available from CRU4.04 dataset. The model, and actual 576 ecosystems, are affected more directly by net irradiance than clouds. The model code is 577 designed to convert clouds to net irradiance if net irradiance is unavailable {Raich, 1991 #408} 578 which means there can be considerable error in the net irradiance values calculated from cloud 579 data. Therefore it is most accurate to correct the historical cloud data to the bias-corrected 580 MACA net irradiance, which is what was done in this study. The additional run involved using 581 total cloud fraction output directly from the same r6i1p1 NCAR CCSM4 RCP8.5 simulation. 582 Note that since these data are not available from MACA, they were bias corrected and 583 downscaled to the corrected cloud data using the period 2006-2014 and a similar method as used to bias correct and downscale the MPI model output to CRU. The results are all statistically 584 insignificant differences in NEP, NCE, cumulative NEP, cumulative NCP, vegetation carbon, 585 586 and soil carbon. 587 (Raich et al., 1991) 588 589 Water variables depend upon precipitation (which is similar between the runs, but can be rain or 590 snowmelt) and evapotranspiration, which ultimately depends upon environmental conditions (i.e. 591 solar radiation, vapor pressure deficit), stomatal conductance, and soil texture (Felzer et al., 592 2011; Shuttleworth and Wallace, 1985). The CONDENSED run exhibits a low bias in 593 evapotranspiration, which is primarily due to low values in pasture grids (Fig. 124). Pasture in 594 the CONDENSED run has higher leaf area index (LAI) then in the RESTART run, due to reinitializing from equilibrium conditions, and that reduced the net irradiance, which limits the 595 amount of soil evaporation. The effect of LAI on soil evaporation in the Shuttleworth Wallace 596 or Penmon Monteith approaches takes the form of an exponential decay, resulting in a much 597 598 sharper dropoff in evaporation with smallerl changes in low LAI than large LAI, which is why the effect is predominant in low height vegetation like pastures. The soil moisture is slightly too 599 large in the RESTART run even though it starts off at the correct value, which also results in a 600 larger evapotranspiration rate. The larger biases in the evapotranspiration flux do not lead to 601 larger biases in the soil moisture stock. While evapotranspiration depends upon vapor pressure 602 603 deficit, net irradiance, stomatal conductance and surface roughness, and its value affects the soil 604 moisture, the amount of soil moisture also affects the amount of water available for evapotranspiration. Increasing vegetation cover has competing effects of reducing soil moisture 605 by shading the ground and increasing evapotranspiration, yet the relative effect of the two 606 depends upon range of the LAI change. 607 608

609 5 Conclusions610

611 This study explores the role of past land use and land cover legacy on the future carbon and 612 water dynamics of terrestrial ecosystems in the conterminous U.S. While most models of 613 thesimulating the future start with current LULCC by reinitializing initial conditions, the actual 614 value of the initial conditions will-should be different because ecosystems are not in a state of 615 equilibriumation, but are changing due to past disturbances and climate change. This study 616 determines whether it is nevertheless possible to use a single realization for each PFT if the 617 initial conditions are set correctly based on a past run that includes land use and land cover 618 legacy effects.

620 The NEP, a measure of carbon sequestration, is too low compared to using all the cohorts when reinitializing initial conditions because the assumption of mature forests rebalances the NEP to 621 become more neutral through enhanced heterotrophic decomposition. There are some offsetting 622 geographic differences across the U.S. when accounting for all ecosystems. The NCE 623 differences are somewhat reduced, however, due to continued product decomposition in runs that 624 625 account for transient changes to LULC in the past. Cumulatively, condensed cohorts have a 626 negative bias in both NEP and NCE, which becomes a positive bias in the case of NEP and is 627 eliminated in the case of NCE by the end of the century when initializing correctly 628 (TEMRESTART). This is evident in the positive biaslarger values in the biomasses (vegetation and soil carbon) relative to RESTART, which are too large for the CONDENSED cohorts but 629 630 greatly improved with TEMRESTART. When PFTs are condensed into single cohorts, the forests are all assumed to be mature forests, which leads to an overestimate of the biomass. The 631 NEP of mature forests is generally less than that of younger forests, though the actual biases 632 between the CONDENSED and RESTART runs but the end of the century are more muted as 633 the forests have had a chance to mature more in both. Correcting for initial conditions reduces 634 635 the bias in vegetation carbon and eliminates the bias in soil carbon. Starting with the correct initial conditions do not have a large impact on the water variables, as they are more dependent 636 637 on environmental factors, though the vegetation cover does have some lesser-minor effects. 638 639 Besides In addition to forest stand age, the initial nutrient loading of the soil is also an important 640 factor for future forest regrowth. With low levels of nitrogen, higher starting values often lead to a larger overall biomass as the forest develops, though there are other environmental factors 641 (e.g.like climate) that are important. Past agricultural use could deplete the soil of nutrients if 642 cropland was abandoned at a time period before chemical fertilization was frequently used (i.e. 643 before the 1950s), or could enhance the soil nutrients if abandoned from heavily fertilized soil. 644 These effects will be accounted for if the correct initial soil conditions are determined. 645 646

647 This study illustrates the importance of accounting for the correct forest stand age and initial soil 648 nutrient conditions in order to model the future carbon sink. While Although starting model runs 649 in the 1700s or earlier is computationally expensive, it is possible to average values from such a 650 run for each PFT to allow a run to start in the present with correct initial condition and achieve a 651 more realistic result more consistent with a detailed representation of land-use cohorts. While this research assumed constant LULC for the future, the next step is to use the corrected initial 652 conditions as a basis for future LULCC. A similar approach can be used to start land use 653 654 transitions at any particular year based on the complete history of land use transitions from 850 A.D. to serve as starting conditions for one of the SSP scenarios. Modeling groups need to 655 consider this effect of past LULC legacy to accurately estimate future carbon biomass and 656 657 fluxes. 658

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660

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