



# Ideas and perspectives: Alleviation of functional limitation by soil organisms is key to climate feedbacks from northern soils

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**Abstract.** Northern soils play an important role in Earth's climate system as they store large amounts of carbon that, if released, could strongly increase greenhouse gas levels in our atmosphere. Most research to date has focused on how the turnover of organic matter in these soils is regulated by abiotic factors and few studies have considered the potential role of biotic regulation. Here, we claim that soil organisms' presence or absence is key to understanding and predicting future climate  
15 feedbacks from northern soils. We propose that the arrival of soil organisms with currently 'missing traits', *i.e.*, properties that the present community does not have, can alleviate functional limitation and result in greatly enhanced decomposition rates, in parity with effects predicted due to increasing temperatures. We base this argument on a series of emerging evidence suggesting that the dispersal of until-then absent micro-, meso- and macro-organisms (*i.e.*, microbes and invertebrate soil fauna) into new regions and newly-thawed soil layers can drastically affect soil functioning. These new observations make us  
20 question the current view that neglects organism driven 'alleviation effects' when predicting the future feedbacks between northern ecosystems and our planets' climate. We therefore advocate for an updated framework in which soil biota and their traits become essential when predicting the fate of soil functions in warming northern ecosystems.

## 1 Introduction

Arctic soils store close to half of worldwide soil carbon (Hugelius et al., 2014; Strauss et al., 2017) and the potential feedbacks  
25 between the about 1300 Pg-C stored in northern soils and our planets climate system are causing concern (IPCC, 2021). To date, the prevailing view is that low temperatures are a primary control of this carbon store, especially of the 822 Pg-C stored frozen in permafrost. This view is well-supported by studies highlighting the top-down control of temperature, next to substrate quality and oxygen availability, on microbial processes (Conant et al., 2008; Razavi et al., 2017). It is also well-established that the widespread presence of permafrost, a soil property closely linked to temperature, is currently constraining  
30 decomposition in northern soils (Goulden et al., 1998). However, northern soils currently lack many species of soil organisms



with properties that are driving decomposition at lower latitudes (Hodkinson and Wookey, 1999; Golovatch and Kime, 2009; Sfenthourakis and Hornung, 2018; Briones, 2014; Aerts, 2006), potentially leaving open niches. As decomposition processes are driven by the functional dissimilarity among the decomposers present (Heemsbergen et al., 2004), these open niches imply that the absence of certain decomposer soil fauna (such as woodlice, millipedes or geoeengineering earthworm species but also  
35 microbial decomposers) and their functions may hamper decomposition rates of soil organic matter in the Arctic. We here refer to this idea as ‘functional limitation’. Consequently, the arrival of these organisms with additional traits or properties, such as bioturbation or an increased rate of litter fragmentation, would increase the functional diversity and could greatly stimulate decomposition.

To what extent such functional limitations in northern soils in the past contributed to the build-up and persistence of large  
40 carbon pools it not well known. In fact, there is no general consensus if the functionally-limited soil communities of northern soils are a result of the harsh climate, or simply due to slow northward dispersal rates of certain soil organisms after the last glaciation. Based on the current distribution of soil fauna in the northern hemisphere, climatic conditions do indeed seem to act as a prime regulator of soil organism traits (Golovatch and Kime, 2009; Kuznetsova and Gongalsky, 2012; Maynard et al., 2019; Sfenthourakis and Hornung, 2018). Nevertheless, studies suggest that the absence of some soil fauna species in the  
45 Arctic is rather due to limited natural dispersal vectors than present day environmental constrains and that large areas of the Arctic might be suitable for establishment of certain decomposing soil organisms already now or in the near future (Blume-Werry et al., 2020; Coulson, 2015; Wackett et al., 2018).

In this opinion piece, we propose that increasing temperatures are opening up new niches for soil organisms in northern soils, both laterally and vertically, and that the effect which newly arriving organisms may have on decomposition processes could  
50 be substantial. This is based on the knowledge that soil organisms are important components of the decomposition process everywhere (Lavelle, 1997; García-Palacios et al., 2013; Griffiths et al., 2021) and for example earthworms, millipedes, isopods, and collembola all can substantially increase mass loss or CO<sub>2</sub> emissions (e.g., (Addison and Parkinson, 1978; Cárcamo et al., 2000; Des Marteaux et al., 2020), especially if functionally diverse species combinations are present (Heemsbergen et al., 2004; Delgado-Baquerizo et al., 2020). Moreover, we highlight that slow historic dispersion of soil fauna  
55 can, at least partly, explain their current absences in northern soils. Soil macro-fauna in particular disperses at slow rates, thus that there is a time-lag, so called ‘invasion debt’ (Rouget et al., 2016), before soils that were previously constrained by glaciation or frozen soils develop food-webs that contain all major functional properties. Our perspective, illustrated in Fig. 1, introduces a framework stating that northern soils are currently in a trait-limited decomposition stage that could be alleviated by lateral, northward dispersal into currently unoccupied areas as well by vertical (downward) dispersal of missing soil  
60 organisms into newly-thawed soil layers. This implies that once soil organisms with currently missing functions or functional properties arrive and more complex food webs develop, decomposition rates may be much higher than suggested from warming of contemporary tundra soils alone (Aerts, 2006; van Geffen et al., 2011; Heemsbergen et al., 2004; Frouz, 2018). In that case, models based on assumptions of how contemporary northern soils respond to climatic variables may fail to foresee important future shifts in tundra soil functions that would arise when soil organisms with traits central for decomposition processes arrive.



65 In this perspective paper we highlight data showing that some functional traits are absent from northern soils and consequences of introduction of these missing properties. Thereto, we provide examples of experiments with additions of soil organisms to estimate the impacts of novel soil organisms arriving in northern soils on the current, ‘trait-limited’ decomposition rates.

## 2 A dispersal constrained community of soil organisms in northern soils

70 Soils are often considered to harbor most functions, due to the omnipresence and large diversity of soil organisms and the generally large functional redundancy assumed amongst them (Nannipieri et al., 2003). While the assumption of functional redundancy in soils is now questioned by soil ecologists, many scientists still generally assume that soil functioning is primarily determined by its physical and chemical composition and thus, that organisms are simply just there if right physiochemical conditions are met. For example, estimates of the climate feedback from northern soils (e.g., Koven et al., 2015; Schuur et al., 2015), rely strongly on incubation studies. Implicitly, this assumes that the incubated microbial and faunal communities carrying out decomposition processes are functionally representative of the communities present in the field after thawing and in warming soils. However, northern soils and particularly permafrost soils are likely to deviate from this assumption. Permafrost soils are indeed not only deprived of most viable fauna – although on rare occasions it has been possible to isolate viable animals such as nematodes (Shatilovich et al., 2018) or rotifers (Shmakova et al., 2021) as well as plants (Yashina et al., 2012) – but also of numerous microbial taxa, resulting in distinct microbial communities (e.g., Johnston et al., 2019; 80 Monteux et al., 2018). Similarly, the biogeographical history of the Arctic, including glaciations, effectively eradicated certain groups of soil organisms from the non-frozen topsoil as well (Briones, 2014). In other words, due to the unique past and current environmental filtering of inland ice-sheets and frozen soils, few would argue against the view that northern soils currently lack both micro-, meso- and macro-organisms that are present in most other soils. It is also likely that warming soils, including thawing permafrost, will open numerous new niches for such soil organisms to establish.

85 Palaeoecological reconstructions have shown that plants have a remarkable capacity to rapidly, *i.e.*, on a decadal time-scale, colonize formerly glaciated areas (Nota et al., 2022), but less is known about the colonization rate of soil organisms after deglaciation or permafrost thaw. From studies of glacier forelands, where soil organisms can establish in open niches via short-range dispersal, we know that mature soil fauna communities can establish within a century (Kaufmann et al., 2002). However, rates of long-range dispersal across hundreds of kilometers into northern soils are unknown and likely much lower. It has been suggested that earthworm disperse naturally with a rate of 5 to 20 m yr<sup>-1</sup> (Chkrebti et al., 2015; Wackett et al., 2018; Cameron et al., 2008; Cameron and Bayne, 2015) and that this slow dispersal from glacial refugia can explain their absence in previously 90 glaciated American forest and the Arctic. These slow dispersal rates of earthworms are likely an important factor constraining their presence, considering that several species can survive and establish in northern soils once introduced by humans (Blume-Werry et al., 2020; Wackett et al., 2018). Similarly, several introduced species of collembola (Coulson, 2015; Enríquez et al., 2019), tapeworms and mites (Coulson, 2015) have been shown to thrive under arctic conditions, further indicating that these species were not constrained by the arctic climate *per se*, but rather by their ability to access tundra soil by their own means.



In contrast, large surface-dwelling animals, such as millipedes (Golovatch and Kime, 2009) and isopods (Sfenthourakis and Hornung, 2018) follow distribution patterns in the Arctic that suggest temperature itself limits their range, rather than the glaciation history. However, this distribution only suggest that their dispersal is fast enough to colonize niches in the Arctic  
100 over Holocene time-scale and thus, it is not self-evident that they can respond at time-scales of relevance for the ongoing climate change, *i.e.* centennial time-scale, unless introduced by humans.

### 3 Evidence for alleviation of functional limitation with novel soil organisms

If contemporary, or near-future, climatic conditions in northern soils do allow novel soil organisms to establish in previously 'trait-limited soils', it is highly relevant to assess how soil organic matter turnover may change when missing traits arrive and  
105 more niches are filled. It has been shown several times (Wall et al., 2008; García-Palacios et al., 2013) that larger soil organisms had limited influence on decomposition processes in northern soils compared to other areas. Yet, such studies are inherently limited by the simplified food web present in northern soils right now and cannot account for the potential contribution to the decomposition process of soil fauna species that are currently absent (Frouz, 2018). For example, when litterbags with subarctic leaf litter were incubated in other ecosystems, access of larger soil fauna did increase mass loss (Makkonen et al., 2012). That  
110 is, the same mesh size does not exclude or include the same soil macrofauna in northern sites as elsewhere. Estimating the effect of an alleviation of functional limitation through filling presently empty niches, therefore, requires the experimental addition of soil organisms to achieve a complex soil food web without missing traits and empty niches. Indeed, several recent studies suggest that the additions of novel traits of soil organisms can have a substantial impact on northern soil organic matter cycling (Blume-Werry et al., 2020; Monteux et al., 2022, 2020; Marushchak et al., 2021). Below, we exemplify that soil  
115 organisms on the micro-, meso-, and macro-scale can alleviate functional limitations and by doing so have profound consequences on northern plant communities and biogeochemical cycling.

Earthworms are probably the best-known example of invasive soil macrofauna. They are incredibly powerful ecosystem engineers that alter the physical and biogeochemical properties of the soil through increased litter decomposition and soil mixing (Fahey et al., 2013) and change soil microbial and faunal communities (Ferlian et al., 2018), thereby affecting  
120 ecosystem functioning and ultimately plant communities (Craven et al., 2017). Though most focus of earthworm research has been on invasions in North American temperate and boreal forests, these processes are likely very relevant in northern soils. Geoengineering, *i.e.* endogeic and anecic, earthworms are generally absent from northern soils, but have been found in isolated patches across the Arctic where they can not only survive but spread out after human introduction (*e.g.*, Blume-Werry et al., 2020; Tiunov et al., 2006; Wackett et al., 2018). Geoengineering earthworms are known to rapidly deplete thick organic layers  
125 in boreal forests through increased decomposition and mixing (Lejoly et al., 2021), likely resulting in carbon release to the atmosphere (Fahey et al., 2013), making this a concern for northern soils as well. Indeed, litter decomposition and organic matter turnover seem to be stimulated immensely when earthworms arrive in tundra soils. Blume-Werry et al. (2020) showed in an earthworm addition experiment into tundra mesocosms that geoengineering earthworms rapidly and substantially



increased plant nitrogen content and plant growth above- and belowground in different tundra plant communities. Late season  
130 root growth in the first year of the experiment, for example, was almost twice as high when earthworms were present. Changes  
in vegetation greenness and nitrogen concentration were even of a similar magnitude or larger, respectively, than 3 °C of  
warming (Blume-Werry et al., 2020). In subarctic microcosms, earthworm addition increased both litter mass loss and CO<sub>2</sub>  
fluxes, as did the addition of isopods and millipedes (van Geffen et al., 2011). Thus, upon the arrival of novel soil macrofauna  
135 new functions seem to be introduced in the soil food web that remove current bottlenecks in organic matter turnover, with thus  
far unquantified consequences for carbon and nutrient cycling.

Soil mesofauna, such as collembola, nematodes, rotifers, and tardigrades, affect litter decomposition rates through their feeding  
activities by inoculating litter with microbes, increasing the surface area of litter substrates, and increase microbial activity  
through grazing (Beare et al., 1992; Lussenhop, 1992). They thus also represent organisms potent enough to cause substantial  
impacts on soil processes when they colonize new areas or soils. In a mesocosm study mimicking a drained thermokarst or  
140 thaw slump scenario, Väisänen et al. (2020) found that within a year microarthropods –but not enchytraeids– were able to  
settle into newly-thawed permafrost at densities one order of magnitude below those found in the surrounding active layer.  
Monteux et al. (2022) assessed how mesofauna, collembola, affect carbon dioxide emissions from newly-thawed permafrost  
soils. Collembola are ubiquitous throughout the Arctic and can be found at very high densities (e.g., 130 000 individuals per  
square meter in high arctic Greenland, Sørensen et al., 2006), and are therefore likely to colonize newly-thawed permafrost  
145 where it is not water-saturated. In the study by Monteux et al. (2022), carbon dioxide production from permafrost soils  
increased by 26% when collembola were present. While about half of this effect could be attributed to collembola respiration  
itself, the remaining 13% directly resulted from increased soil organic matter decomposition. Presence of collembola also  
increased CO<sub>2</sub> emissions from topsoils by up to 400% in a high arctic site (Addison and Parkinson, 1978). These findings  
imply that standard incubation studies of permafrost or active layer soil without additional soil fauna, could be strongly  
150 underestimating the potential carbon emissions of these soils upon thawing.

Microbial communities in permafrost differ from those found in the overlying active layer (Doherty et al., 2020; Johnston et  
al., 2019; Monteux et al., 2018), and their functional potential for decomposition processes can also be drastically smaller than  
that of active layer communities. Consequently, if new microorganisms are added to thawed permafrost soils, they can alleviate  
missing functions and strongly increase carbon dioxide production (+38%, Monteux et al., 2020), but also initiate  
155 methanogenesis (Knoblauch et al., 2018) or nitrification (Monteux et al., 2020). In other words, such ecosystem processes  
seem not limited by the lack of adequate substrates, but rather by the absence of microorganisms harboring the specific genes  
needed to carry out these biochemical transformation processes. These findings are not constrained to laboratory incubations  
but can also be observed in more realistic field settings. For instance, nitrogen cycling gene abundances and process rates are  
very low immediately following permafrost thaw in Yedoma exposures, but substantially increase with ecosystem complexity  
160 as new functions are introduced by newly-arriving organisms (Marushchak et al., 2021). These new functions increased N<sub>2</sub>O  
production by 1 to 2 orders of magnitude, an effect which would be omitted by incubation studies focusing solely on the  
functionally limited microbial communities present in permafrost before thaw. Taken together with similar findings on



methane and carbon dioxide production (Knoblauch et al., 2018; Monteux et al., 2020), it seems evident that introduction of  
microbes with novel functions or increased efficiency can boost the emission of several greenhouse gases from thawing  
165 permafrost. The dynamics of microbial community assembly upon permafrost thaw are a growing field of research (see  
Ernakovich et al., 2022), yet their interaction with the rest of the soil food web remains largely unexplored.

#### 4 The contemporary mismatch between climate, plants, and soil organisms

Arctic regions are shaped by strong environmental filters, both in the past and present, which resulted in unmatched high  
allocation of plant biomass belowground relative to aboveground (Fig. 2a) and soil organisms adapted for survival rather than  
170 high functional performance (Crowther et al., 2019; Nielsen and Wall, 2013). Low functional performance and the resulting  
limited decomposition rates have led to a large build-up of soil organic matter in northern soils (Fig. 2a). However, as  
increasing temperature and changes in snowfall patterns are rapidly changing the arctic environment above- and belowground,  
new niches and opportunities are opening up for soil organisms to utilize the large energy sources stored at depth in northern  
soils. New niches can arise both through direct climate changes, or indirectly through vegetation changes (Kaufmann et al.,  
175 2002; Krab et al., 2019) which are widespread throughout the Arctic (Elmendorf et al., 2012; Myers-Smith et al., 2019). Sound  
predictions about the future fate of soil carbon in northern soils thus depend on correct understanding of processes controlling  
decomposition in the near future. We identified two different scenarios based on existing knowledge. The **first scenario**, which  
seems to be the theory most studies apply, assumes ‘business as usual’ (Scenario 1, Fig. 2b). Here, large-scale and dramatic  
changes in the belowground environment do not lead to a change in the presence, density or depth-distribution of soil  
180 organisms. With this conceptual view, the fate of C can simply be predicted by warming experiments or by using natural  
gradients existing in the contemporary Arctic. While warmer soils alone can increase activity and turnover rates of soil  
organisms, for example through an increase in density (Dollery et al., 2006), we assume that significant changes in the  
functional potential of the soil organisms only occur with changes in community composition (Crowther et al., 2019). Thus,  
our **second scenario** (Scenario 2, Fig 2b) highlights that the current soil food-webs might not be representative for the future  
185 and that the impacts of other functions, currently absent in the contemporary environment, need to be accounted for. Indeed,  
future northern soils may have a more functionally diverse soil community (Scenario 2, Fig. 2b) in which new properties arrive,  
both in topsoils and in deeper soil layers, and thus increase the rates of decomposition processes (Heemsbergen et al., 2004).  
As outlined in this perspective piece, there are numerous studies in support of his second scenario making it highly relevant to  
account for northward dispersion of soil fauna in future models. For example, woodlice distribution seems to be restricted  
190 south of the limit of 120 days per year with a temperature above 10°C (Sfenthourakis and Hornung, 2018), and might thus  
progress northwards along with that limit. Likewise, millipedes appear absent from most regions affected by permafrost  
(Golovatch and Kime, 2009) and therefore might be able to disperse northwards once permafrost recedes. As macro-  
decomposers breaking down large litter elements into smaller pieces, woodlice and millipedes provide important ecosystem  
functions to soil and can significantly speed up decomposition (Joly et al., 2018; Lavelle, 1997), and their dispersal into soils



195 where they are absent could therefore affect their biogeochemical cycling. Interestingly, Golovatch & Kime (2009) also show  
isolated millipede occurrence outside of their regular distribution range, illustrating that they can indeed survive beyond their  
current distribution range already now. This suggests that the current climate does not fully explain the absence of millipedes,  
and that dispersal limitation might be more important. Similarly, geoengineering earthworms are also mostly absent from  
previously glaciated areas but are successfully settling at and dispersing from points of anthropogenic introduction. Again, this  
200 suggests that the biogeographical history will play a smaller role in the future, as they will colonize more and more of these  
systems where they were until now absent (Blume-Werry et al., 2020; Wackett et al., 2018). If and how fast which groups will  
reach northern soils, whether on its own or via anthropogenic dispersal, is difficult to assess as there is an overall lack of data  
on dispersal abilities (Aerts, 2006; Hickling et al., 2006; David and Handa, 2010). Even the limits of spatial distribution are  
poorly defined for several important groups of soil fauna groups, as the northernmost range of their apparent distribution  
205 coincides with areas where samplings are scarce, and it is often not clear whether a given study did not find such or such group  
or did not look for it (Bastida et al., 2020; Lavelle et al., 2022).

Of the belowground changes in the Arctic, the widespread thawing of permafrost (Smith et al., 2022) is probably the most  
striking as it removes an obvious barrier for soil fauna dispersal and opens up new habitats. That is, there might not only be  
more species with new properties in the topsoil but also an increase in functionality deeper down in the newly-thawed soils  
210 (Fig. 2b). The thawing of permafrost happens overall as a thickening of the seasonally thawed layer above the permafrost,  
upon which the new soil volume is explored by roots of certain plant species (Blume-Werry et al., 2019; Finger et al., 2016).  
Microbial communities in the newly-thawed permafrost become similar to the active layer communities (Doherty et al., 2020;  
Monteux et al., 2018), although it is unclear to what extent this stems from downwards dispersal, influence of plant roots, or  
endogenic changes from the permafrost communities. While soil meso- and macrofauna are less likely to substantially colonize  
215 these deep, often water-logged thawing layers, permafrost does not solely thaw as a thickening of the active layer. Various  
thaw features can be observed, such as drained thermokarst, retrogressive thaw slumps or active layer detachments (Inglese et  
al., 2017; Olefeldt et al., 2016). In these circumstances, former permafrost becomes thawed and exposed to surface conditions,  
thus providing suitable new habitats for soil organisms. At least micro- and mesofauna appear able to establish in this newly-  
thawed permafrost (Väisänen et al., 2020), with hitherto unclear consequences, although their impact on bacterial community  
220 composition seems rather limited (Monteux et al., 2022, this special issue).

## 5 Conclusions

Here, we postulate a contemporary mismatch between climate changes, plant responses, and colonization by soil organisms  
across the Arctic, leading to a currently trait-limited decomposition. If the complexity and function of the food web are not  
explicitly manipulated, a potential functional alleviation is not captured by warming experiments thus inadvertently missing  
225 out on essential system shifts. We thus advocate for improved and accessible data on distribution of functional groups of soil  
decomposers, notably macro-detritivores and geoengineering earthworms, in the circum-arctic region. Ideally, this data would



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include whether populations out of their apparent climate range are relict or human-introduced, and experiments specifically testing the effects of alleviation of functional limitations by one or more of these functional groups such that the scientific community can better predict the true feedback potential from northern soils to the global climate.

#### **Author contributions**

GBW conceived the idea based on discussion with SM, JK, and EJK. GBW led the writing of the manuscript with substantial input from JK, EJK and SM. EJK was financially supported by VR (grant 2021-04458). All authors contributed critically to the drafts and gave final approval for publication.

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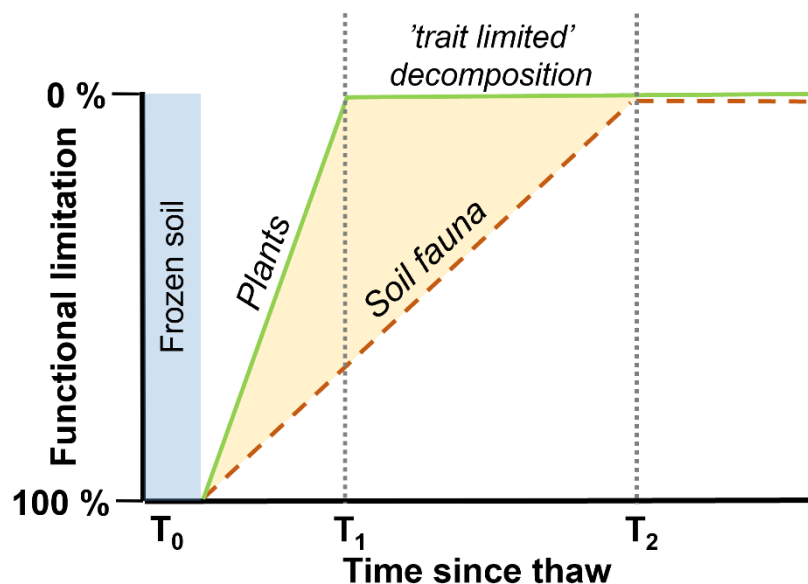
#### **Conflict of interest**

The authors declare that they have no conflict of interest.





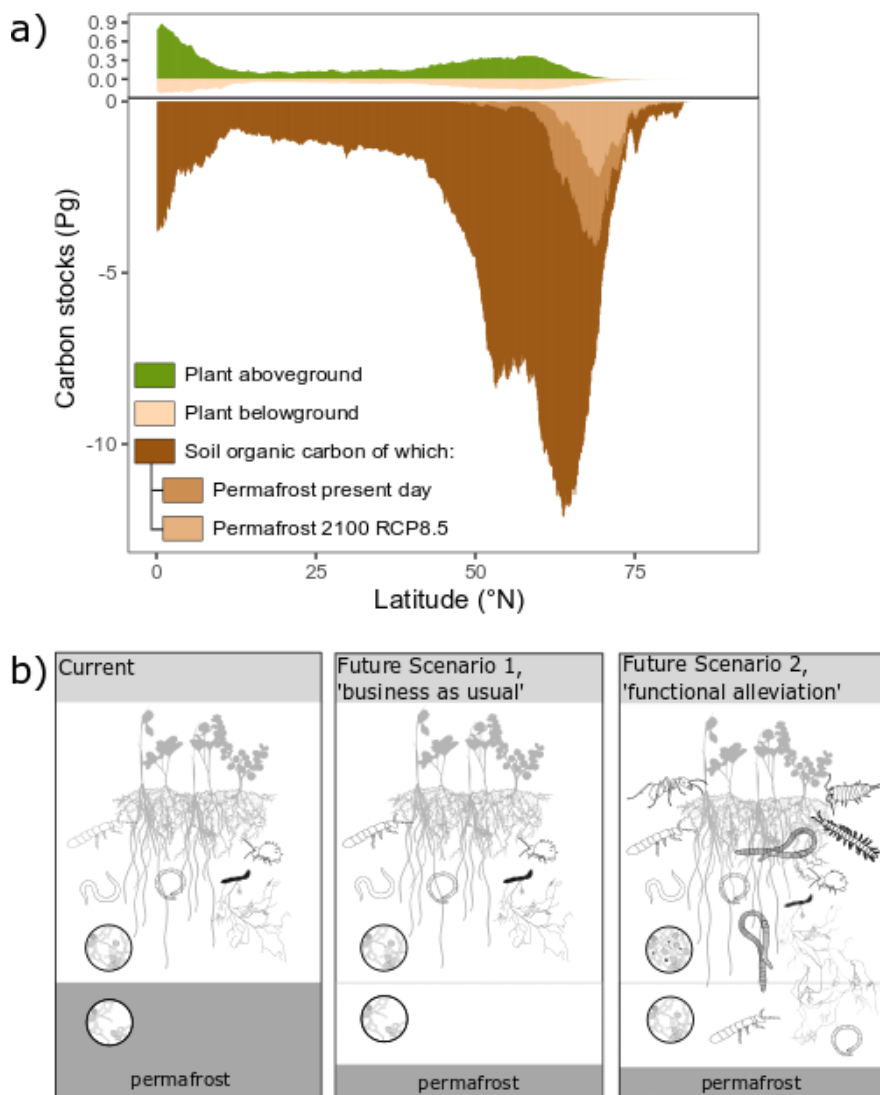
## Figures



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**Figure 1. Conceptual illustration of the theory underlying this perspective paper.** Following the retreat of glacier ice or thawing of permafrost, soil processes previously constrained by cryogenic processes are alleviated. At this point in time ( $T_0$ ), novel plants and soil biota can establish in previously unoccupied areas or soil layers. As plant niches are expected to be filled at a higher rate, all possible groups of plant functional traits are represented at time-stage  $T_1$  while functional traits or properties controlled by soil organisms arrive later ( $T_2$ ). During the outlined scenario soils evolve between time interval  $T_2$ - $T_1$  with a 'trait limitation', *i.e.* where key functional traits in the food web may be missing. Here, groups of organisms with specific functions may be lacking, not necessarily because of climatic drivers but possibly due to slow dispersion vectors. We propose here that northern soils are currently in  $T_1$ - $T_2$ , detritivore-limited decomposition, implying that once they arrive and fully complex food webs develop, decomposition rates will be much higher than suggested from warming of contemporary tundra soils alone.

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**Figure 2.** Latitudinal distribution of soil carbon, plant biomass above- and belowground, and an illustration of the current, functionally limited, arctic soil food web and two future scenarios, ‘business as usual’ and ‘functional alleviation’. a) Northern soils are characterized by disproportionately large amounts of belowground plant biomass and large stores of soil organic carbon. Climate change opens up new habitats both in latitude and depth as soils thaw and warm up, but current predictions assume no accompanying changes in the soil fauna and decomposition process. Above- and belowground plant carbon stocks for 2010 are from Spawn et al. (2020). Total soil organic carbon stocks are the sums over 0-2 m depth from SoilGrids250m 2.0 (Poggio et al., 2021); current and future permafrost stocks by Keuper et al., (2020) are obtained from applying CLM4.5 simulations (Koven et al., 2015) to SOC stocks from NCSCDv2 (Hugelius et al., 2014). Permafrost data are for deposits between 0-3 m and thus exclude about half of permafrost SOC contained in deeper deposits (Strauss et al., 2017; Hugelius et



265 al., 2014). b) The current food web in northern topsoils is characterized by dominance of micro- and mesofauna such as nematodes, enchytraeids and collembola and the soil matrix in which they live is often constrained vertically by permafrost. In a 'business as usual scenario', the same micro-and mesofauna continue to dominate the soil matrix despite vertical expansion of the soil matrix due to permafrost thaw. In the 'functional alleviation' scenario, functional traits are added to the both top- and lower soils through the establishment of soil organisms with additional properties.



## References

- Addison, J. A. and Parkinson, D.: Influence of Collembolan Feeding Activities on Soil Metabolism at a High Arctic Site, *Oikos*, 30, 529, <https://doi.org/10.2307/3543348>, 1978.
- 270 Aerts, R.: The freezer defrosting: global warming and litter decomposition rates in cold biomes: *Global warming and litter decomposition*, *Journal of Ecology*, 94, 713–724, <https://doi.org/10.1111/j.1365-2745.2006.01142.x>, 2006.
- Bastida, F., Eldridge, D. J., Abades, S., Alfaro, F. D., Gallardo, A., García-Velázquez, L., García, C., Hart, S. C., Pérez, C. A., Santos, F., Trivedi, P., Williams, M. A., and Delgado-Baquerizo, M.: Climatic vulnerabilities and ecological preferences of soil invertebrates across biomes, *Mol Ecol*, 29, 752–761, <https://doi.org/10.1111/mec.15299>, 2020.
- 275 Beare, M. H., Parmelee, R. W., Hendrix, P. F., Cheng, W., Coleman, D. C., and Crossley, D. A.: Microbial and Faunal Interactions and Effects on Litter Nitrogen and Decomposition in Agroecosystems, *Ecol Monogr*, 62, 569–591, <https://doi.org/10.2307/2937317>, 1992.
- Blume-Werry, G., Milbau, A., Teuber, L. M., Johansson, M., and Dorrepaal, E.: Dwelling in the deep – strongly increased root growth and rooting depth enhance plant interactions with thawing permafrost soil, *New Phytol*, 223, 1328–1339, <https://doi.org/10.1111/nph.15903>, 2019.
- 280 Blume-Werry, G., Krab, E. J., Olofsson, J., Sundqvist, M. K., Väisänen, M., and Klaminder, J.: Invasive earthworms unlock arctic plant nitrogen limitation, *Nature Communications*, 11, 10, <https://doi.org/10.1038/s41467-020-15568-3>, 2020.
- Briones, M. J. I.: Soil fauna and soil functions: a jigsaw puzzle, *Front. Environ. Sci.*, 2, <https://doi.org/10.3389/fenvs.2014.00007>, 2014.
- 285 Cameron, E. K. and Bayne, E. M.: Spatial patterns and spread of exotic earthworms at local scales, *Can. J. Zool.*, 93, 721–726, <https://doi.org/10.1139/cjz-2014-0197>, 2015.
- Cameron, E. K., Bayne, E. M., and Coltman, D. W.: Genetic structure of invasive earthworms *Dendrobaena octaedra* in the boreal forest of Alberta: insights into introduction mechanisms, *Molecular Ecology*, 17, 1189–1197, <https://doi.org/10.1111/j.1365-294X.2007.03603.x>, 2008.
- 290 Cárcamo, H. A., Abe, T. A., Prescott, C. E., Holl, F. B., and Chanway, C. P.: Influence of millipedes on litter decomposition, N mineralization, and microbial communities in a coastal forest in British Columbia, Canada, *Can. J. For. Res.*, 30, 817–826, <https://doi.org/10.1139/x00-014>, 2000.
- Chkrebtii, O. A., Cameron, E. K., Campbell, D. A., and Bayne, E. M.: Transdimensional approximate Bayesian computation for inference on invasive species models with latent variables of unknown dimension, *Computational Statistics & Data Analysis*, 86, 97–110, <https://doi.org/10.1016/j.csda.2015.01.002>, 2015.
- 295 Conant, R. T., Drijber, R. A., Haddix, M. L., Parton, W. J., Paul, E. A., Plante, A. F., Six, J., and Steinweg, J. M.: Sensitivity of organic matter decomposition to warming varies with its quality, *Global Change Biology*, 14, 868–877, <https://doi.org/10.1111/j.1365-2486.2008.01541.x>, 2008.
- Coulson, S. J.: The alien terrestrial invertebrate fauna of the High Arctic archipelago of Svalbard: potential implications for the native flora and fauna, *Polar Research*, <https://doi.org/10.3402/polar.v34.27364>, 2015.
- 300 Craven, D., Thakur, M. P., Cameron, E. K., Frelich, L. E., Beauséjour, R., Blair, R. B., Blossey, B., Burtis, J., Choi, A., Dávalos, A., Fahey, T. J., Fisichelli, N. A., Gibson, K., Handa, I. T., Hopfensperger, K., Loss, S. R., Nuzzo, V., Maerz, J. C.,



- Sackett, T., Scharenbroch, B. C., Smith, S. M., Vellend, M., Umek, L. G., and Eisenhauer, N.: The unseen invaders: introduced earthworms as drivers of change in plant communities in North American forests (a meta-analysis), *Glob Change Biol*, 23, 1065–1074, <https://doi.org/10.1111/gcb.13446>, 2017.
- 305 Crowther, T. W., van den Hoogen, J., Wan, J., Mayes, M. A., Keiser, A. D., Mo, L., Averill, C., and Maynard, D. S.: The global soil community and its influence on biogeochemistry, *Science*, 365, eaav0550, <https://doi.org/10.1126/science.aav0550>, 2019.
- David, J.-F. and Handa, I. T.: The ecology of saprophagous macroarthropods (millipedes, woodlice) in the context of global change, *Biological Reviews*, 85, 881–895, <https://doi.org/10.1111/j.1469-185X.2010.00138.x>, 2010.
- 310 Delgado-Baquerizo, M., Reich, P. B., Trivedi, C., Eldridge, D. J., Abades, S., Alfaro, F. D., Bastida, F., Berhe, A. A., Cutler, N. A., Gallardo, A., García-Velázquez, L., Hart, S. C., Hayes, P. E., He, J.-Z., Hseu, Z.-Y., Hu, H.-W., Kirchmair, M., Neuhauser, S., Pérez, C. A., Reed, S. C., Santos, F., Sullivan, B. W., Trivedi, P., Wang, J.-T., Weber-Grullon, L., Williams, M. A., and Singh, B. K.: Multiple elements of soil biodiversity drive ecosystem functions across biomes, *Nat Ecol Evol*, 4, 210–220, <https://doi.org/10.1038/s41559-019-1084-y>, 2020.
- 315 Des Marteaux, L. E., Kullik, S. A., Habash, M., and Schmidt, J. M.: Terrestrial Isopods *Porcellio scaber* and *Oniscus asellus* (Crustacea: Isopoda) Increase Bacterial Abundance and Modify Microbial Community Structure in Leaf Litter Microcosms: a Short-Term Decomposition Study, *Microb Ecol*, 80, 690–702, <https://doi.org/10.1007/s00248-020-01527-4>, 2020.
- Doherty, S. J., Barbato, R. A., Grandy, A. S., Thomas, W. K., Monteux, S., Dorrepaal, E., Johansson, M., and Ernakovich, J. G.: The Transition From Stochastic to Deterministic Bacterial Community Assembly During Permafrost Thaw Succession, *Front. Microbiol.*, 11, 596589, <https://doi.org/10.3389/fmicb.2020.596589>, 2020.
- 320 Dollery, R., Hodkinson, I. D., and Jónsdóttir, I. S.: Impact of warming and timing of snow melt on soil microarthropod assemblages associated with *Dryas*-dominated plant communities on Svalbard, *Ecography*, 29, 111–119, <https://doi.org/10.1111/j.2006.0906-7590.04366.x>, 2006.
- Elmendorf, S. C., Henry, G. H. R., Hollister, R. D., Björk, R. G., Boulanger-Lapointe, N., Cooper, E. J., Cornelissen, J. H. C., Day, T. A., Dorrepaal, E., Elumeeva, T. G., Gill, M., Gould, W. A., Harte, J., Hik, D. S., Hofgaard, A., Johnson, D. R., Johnstone, J. F., Jónsdóttir, I. S., Jorgenson, J. C., Klanderud, K., Klein, J. A., Koh, S., Kudo, G., Lara, M., Lévesque, E., Magnússon, B., May, J. L., Mercado-Díaz, J. A., Michelsen, A., Molau, U., Myers-Smith, I. H., Oberbauer, S. F., Onipchenko, V. G., Rixen, C., Martin Schmidt, N., Shaver, G. R., Spasojevic, M. J., Þórhallsdóttir, Þ. E., Tolvanen, A., Troxler, T., Tweedie, C. E., Villareal, S., Wahren, C.-H., Walker, X., Webber, P. J., Welker, J. M., and Wipf, S.: Plot-scale evidence of tundra vegetation change and links to recent summer warming, *Nature Clim Change*, 2, 453–457, <https://doi.org/10.1038/nclimate1465>, 2012.
- 330 Enríquez, N., Pertierra, L. R., Tejedo, P., Benayas, J., Greenslade, P., and Luciáñez, M. J.: The importance of long-term surveys on species introductions in Maritime Antarctica: first detection of *Ceratophysella succinea* (Collembola: Hypogastruridae), *Polar Biol*, 5, 2019.
- 335 Ernakovich, J. G., Barbato, R. A., Rich, V. I., Schädel, C., Hewitt, R. E., Doherty, S. J., Whalen, E. D., Abbott, B. W., Barta, J., Biasi, C., Chabot, C. L., Hultman, J., Knoblauch, C., Vetter, M. C. Y. L., Leewis, M., Liebner, S., Mackelprang, R., Onstott, T. C., Richter, A., Schütte, U. M. E., Siljanen, H. M. P., Taş, N., Timling, I., Vishnivetskaya, T. A., Waldrop, M. P., and Winkel, M.: Microbiome assembly in thawing permafrost and its feedbacks to climate, *Global Change Biology*, *gcb.16231*, <https://doi.org/10.1111/gcb.16231>, 2022.



- 340 Fahey, T. J., Yavitt, J. B., Sherman, R. E., Maerz, J. C., Groffman, P. M., Fisk, M. C., and Bohlen, P. J.: Earthworm effects on the incorporation of litter C and N into soil organic matter in a sugar maple forest, *Ecol Appl*, 23, 1185–1201, <https://doi.org/10.1890/12-1760.1>, 2013.
- Ferlian, O., Eisenhauer, N., Aguirrebengoa, M., Camara, M., Ramirez-Rojas, I., Santos, F., Tanalgo, K., and Thakur, M. P.: Invasive earthworms erode soil biodiversity: A meta-analysis, *J Anim Ecol*, 87, 162–172, <https://doi.org/10.1111/1365-2656.12746>, 2018.
- 345 Finger, R. A., Turetsky, M. R., Kielland, K., Ruess, R. W., Mack, M. C., and Euskirchen, E. S.: Effects of permafrost thaw on nitrogen availability and plant-soil interactions in a boreal Alaskan lowland, *J Ecol*, 104, 1542–1554, <https://doi.org/10.1111/1365-2745.12639>, 2016.
- Frouz, J.: Effects of soil macro- and mesofauna on litter decomposition and soil organic matter stabilization, *Geoderma*, 332, 161–172, <https://doi.org/10.1016/j.geoderma.2017.08.039>, 2018.
- 350 García-Palacios, P., Maestre, F. T., Kattge, J., and Wall, D. H.: Climate and litter quality differently modulate the effects of soil fauna on litter decomposition across biomes, *Ecol Lett*, 16, 1045–1053, <https://doi.org/10.1111/ele.12137>, 2013.
- van Geffen, K. G., Berg, M. P., and Aerts, R.: Potential macro-detritivore range expansion into the subarctic stimulates litter decomposition: a new positive feedback mechanism to climate change?, *Oecologia*, 167, 1163–1175, <https://doi.org/10.1007/s00442-011-2051-8>, 2011.
- 355 Golovatch, S. I. and Kime, R. D.: Millipede (Diplopoda) distributions: A review, *Soil Org*, 81, 565–597, 2009.
- Goulden, M. L., Wofsy, S. C., Harden, J. W., Trumbore, S. E., Crill, P. M., Gower, S. T., Fries, T., Daube, B. C., Fan, S.-M., Sutton, D. J., Bazzaz, A., and Munger, J. W.: Sensitivity of Boreal Forest Carbon Balance to Soil Thaw, *Science*, 279, 214–217, <https://doi.org/10.1126/science.279.5348.214>, 1998.
- 360 Griffiths, H. M., Ashton, L. A., Parr, C. L., and Eggleton, P.: The impact of invertebrate decomposers on plants and soil, *New Phytologist*, 231, 2142–2149, <https://doi.org/10.1111/nph.17553>, 2021.
- Heemsbergen, D. A., Berg, M. P., Loreau, M., van Hal, J. R., Faber, J. H., and Verhoef, H. A.: Biodiversity Effects on Soil Processes Explained by Interspecific Functional Dissimilarity, *Science*, 306, 1019–1020, <https://doi.org/10.1126/science.1101865>, 2004.
- 365 Hickling, R., Roy, D. B., Hill, J. K., Fox, R., and Thomas, C. D.: The distributions of a wide range of taxonomic groups are expanding polewards, *Global Change Biology*, 12, 450–455, <https://doi.org/10.1111/j.1365-2486.2006.01116.x>, 2006.
- Hodkinson, I. D. and Wookey, P. A.: Functional ecology of soil organisms in tundra ecosystems: towards the future, *Applied Soil Ecology*, 11, 111–126, [https://doi.org/10.1016/S0929-1393\(98\)00142-5](https://doi.org/10.1016/S0929-1393(98)00142-5), 1999.
- 370 Hugelius, G., Strauss, J., Zubrzycki, S., Harden, J. W., Schuur, E. A. G., Ping, C.-L., Schirmer, L., Grosse, G., Michaelson, G. J., Koven, C. D., O’Donnell, J. A., Elberling, B., Mishra, U., Camill, P., Yu, Z., Palmtag, J., and Kuhry, P.: Estimated stocks of circumpolar permafrost carbon with quantified uncertainty ranges and identified data gaps, *Biogeosciences*, 11, 6573–6593, <https://doi.org/10.5194/bg-11-6573-2014>, 2014.
- Inglese, C. N., Christiansen, C. T., Lamhonwah, D., Moniz, K., Montross, S. N., Lamoureux, S., Lafrenière, M., Grogan, P., and Walker, V. K.: Examination of Soil Microbial Communities After Permafrost Thaw Subsequent to an Active Layer Detachment in the High Arctic, *Arct. Antarct. Alp.*, 49, 455–472, <https://doi.org/10.1657/AAAR0016-066>, 2017.
- 375



- IPCC: Climate Change 2021: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change, Cambridge University Press, 2021.
- Johnston, E. R., Hatt, J. K., He, Z., Wu, L., Guo, X., Luo, Y., Schuur, E. A. G., Tiedje, J. M., Zhou, J., and Konstantinidis, K. T.: Responses of tundra soil microbial communities to half a decade of experimental warming at two critical depths, *Proc. Natl. Acad. Sci. U.S.A.*, 116, 15096–15105, <https://doi.org/10.1073/pnas.1901307116>, 2019.
- 380 Joly, F., Coq, S., Coulis, M., Nahmani, J., and Hättenschwiler, S.: Litter conversion into detritivore faeces reshuffles the quality control over C and N dynamics during decomposition, *Funct Ecol*, 32, 2605–2614, <https://doi.org/10.1111/1365-2435.13178>, 2018.
- Kaufmann, R., Fuchs, M., and Gosterxeier, N.: The Soil Fauna of an Alpine Glacier Foreland: Colonization and Succession, *Arct. Antarct. Alp.*, 34, 242–250, 2002.
- 385 Keuper, F., Wild, B., Kummu, M., Beer, C., Blume-Werry, G., Fontaine, S., Gavazov, K., Gentsch, N., Guggenberger, G., Hugelius, G., Jalava, M., Koven, C., Krab, E. J., Kuhry, P., Monteux, S., Richter, A., Shahzad, T., Weedon, J. T., and Dorrepaal, E.: Carbon loss from northern circumpolar permafrost soils amplified by rhizosphere priming, *Nat. Geosci.*, 13, 560–565, <https://doi.org/10.1038/s41561-020-0607-0>, 2020.
- 390 Knoblauch, C., Beer, C., Liebner, S., Grigoriev, M. N., and Pfeiffer, E.-M.: Methane production as key to the greenhouse gas budget of thawing permafrost, *Nature Clim Change*, 8, 309–312, <https://doi.org/10.1038/s41558-018-0095-z>, 2018.
- Koven, C. D., Lawrence, D. M., and Riley, W. J.: Permafrost carbon–climate feedback is sensitive to deep soil carbon decomposability but not deep soil nitrogen dynamics, *Proc Natl Acad Sci USA*, 201415123, <https://doi.org/10.1073/pnas.1415123112>, 2015.
- 395 Krab, E. J., Monteux, S., Weedon, J. T., and Dorrepaal, E.: Plant expansion drives bacteria and collembola communities under winter climate change in frost-affected tundra, *Soil Biol. Biochem.*, 138, 107569, <https://doi.org/10.1016/j.soilbio.2019.107569>, 2019.
- Kuznetsova, D. M. and Gongalsky, K. B.: Cartographic analysis of woodlice fauna of the former USSR, *ZooKeys*, 176, 1–11, <https://doi.org/10.3897/zookeys.176.2372>, 2012.
- 400 Lavelle, P.: Faunal Activities and Soil Processes: Adaptive Strategies That Determine Ecosystem Function, in: *Advances in Ecological Research*, vol. 27, Academic Press, 93–132, [https://doi.org/10.1016/S0065-2504\(08\)60007-0](https://doi.org/10.1016/S0065-2504(08)60007-0), 1997.
- Lavelle, P., Mathieu, J., Spain, A., Brown, G., Fragoso, C., Lapied, E., De Aquino, A., Barois, I., Barrios, E., Barros, M. E., Bedano, J. C., Blanchart, E., Caulfield, M., Chagueza, Y., Dai, J., Decaëns, T., Dominguez, A., Dominguez, Y., Feijoo, A., Folgarait, P., Fonte, S. J., Gorosito, N., Huerta, E., Jimenez, J. J., Kelly, C., Loranger, G., Marchão, R., Marichal, R., Praxedes, C., Rodriguez, L., Rousseau, G., Rousseau, L., Ruiz, N., Sanabria, C., Suarez, J. C., Tondoh, J. E., De Valença, A., Vanek, S. J., Vasquez, J., Velasquez, E., Webster, E., and Zhang, C.: Soil macroinvertebrate communities: A world-wide assessment, *Global Ecol. Biogeogr.*, 31, 1261–1276, <https://doi.org/10.1111/geb.13492>, 2022.
- Lejoly, J., Quideau, S., and Laganière, J.: Invasive earthworms affect soil morphological features and carbon stocks in boreal forests, *Geoderma*, 404, 115262, <https://doi.org/10.1016/j.geoderma.2021.115262>, 2021.
- 410 Lussenhop, J.: Mechanisms of Microarthropod-Microbial Interactions in Soil, in: *Advances in Ecological Research*, vol. 23, Elsevier, 1–33, [https://doi.org/10.1016/S0065-2504\(08\)60145-2](https://doi.org/10.1016/S0065-2504(08)60145-2), 1992.



- Makkonen, M., Berg, M. P., Handa, I. T., Hättenschwiler, S., van Ruijven, J., van Bodegom, P. M., and Aerts, R.: Highly consistent effects of plant litter identity and functional traits on decomposition across a latitudinal gradient, *Ecol Lett*, 15, 1033–1041, <https://doi.org/10.1111/j.1461-0248.2012.01826.x>, 2012.
- 415 Marushchak, M. E., Kerttula, J., Diáková, K., Faguet, A., Gil, J., Grosse, G., Knoblauch, C., Lashchinskiy, N., Martikainen, P. J., Morgenstern, A., Nykamb, M., Ronkainen, J. G., Siljanen, H. M. P., van Delden, L., Voigt, C., Zimov, N., Zimov, S., and Biasi, C.: Thawing Yedoma permafrost is a neglected nitrous oxide source, *Nat Commun*, 12, 7107, <https://doi.org/10.1038/s41467-021-27386-2>, 2021.
- 420 Maynard, D. S., Bradford, M. A., Covey, K. R., Lindner, D., Glaeser, J., Talbert, D. A., Tinker, P. J., Walker, D. M., and Crowther, T. W.: Consistent trade-offs in fungal trait expression across broad spatial scales, *Nat Microbiol*, 4, 846–853, <https://doi.org/10.1038/s41564-019-0361-5>, 2019.
- Monteux, S., Weedon, J. T., Blume-Werry, G., Gavazov, K., Jassey, V. E. J., Johansson, M., Keuper, F., Olid, C., and Dorrepaal, E.: Long-term in situ permafrost thaw effects on bacterial communities and potential aerobic respiration, *ISME J*, 12, 2129–2141, <https://doi.org/10.1038/s41396-018-0176-z>, 2018.
- 425 Monteux, S., Keuper, F., Fontaine, S., Gavazov, K., Hallin, S., Juhanson, J., Krab, E. J., Revaillet, S., Verbruggen, E., Walz, J., Weedon, J. T., and Dorrepaal, E.: Carbon and nitrogen cycling in Yedoma permafrost controlled by microbial functional limitations, *Nat. Geosci.*, 13, 794–798, <https://doi.org/10.1038/s41561-020-00662-4>, 2020.
- Monteux, S., Mariën, J., and Krab, E. J.: Dispersal of bacteria and stimulation of permafrost decomposition by Collembola, *Biogeosciences Discuss.*, 2022, 1–26, <https://doi.org/10.5194/bg-2022-98>, 2022.
- 430 Myers-Smith, I. H., Grabowski, M. M., Thomas, H. J. D., Angers-Blondin, S., Daskalova, G. N., Bjorkman, A. D., Cunliffe, A. M., Assmann, J. J., Boyle, J. S., McLeod, E., McLeod, S., Joe, R., Lennie, P., Arey, D., Gordon, R. R., and Eckert, C. D.: Eighteen years of ecological monitoring reveals multiple lines of evidence for tundra vegetation change, *Ecol Monogr*, 89, e01351, <https://doi.org/10.1002/ecm.1351>, 2019.
- 435 Nannipieri, P., Ascher, J., Ceccherini, M. T., Landi, L., Pietramellara, G., and Renella, G.: Microbial diversity and soil functions, *Eur. J. Soil Sci.*, 54, 655–670, <https://doi.org/10.1046/j.1351-0754.2003.0556.x>, 2003.
- Nielsen, U. N. and Wall, D. H.: The future of soil invertebrate communities in polar regions: different climate change responses in the Arctic and Antarctic?, *Ecol Lett*, 16, 409–419, <https://doi.org/10.1111/ele.12058>, 2013.
- 440 Nota, K., Klaminder, J., Milesi, P., Bindler, R., Nobile, A., van Steijn, T., Bertilsson, S., Svensson, B., Hirota, S. K., Matsuo, A., Gunnarsson, U., Seppä, H., Välijanta, M. M., Wohlfarth, B., Suyama, Y., and Parducci, L.: Norway spruce postglacial recolonization of Fennoscandia, *Nat Commun*, 13, 1333, <https://doi.org/10.1038/s41467-022-28976-4>, 2022.
- Olefeldt, D., Goswami, S., Grosse, G., Hayes, D., Hugelius, G., Kuhry, P., McGuire, A. D., Romanovsky, V. E., Sannel, A. B. K., Schuur, E. A. G., and Turetsky, M. R.: Circumpolar distribution and carbon storage of thermokarst landscapes, *Nat Commun*, 7, 13043, <https://doi.org/10.1038/ncomms13043>, 2016.
- 445 Poggio, L., de Sousa, L. M., Batjes, N. H., Heuvelink, G. B. M., Kempen, B., Ribeiro, E., and Rossiter, D.: SoilGrids 2.0: producing soil information for the globe with quantified spatial uncertainty, *SOIL*, 7, 217–240, <https://doi.org/10.5194/soil-7-217-2021>, 2021.
- Razavi, B. S., Liu, S., and Kuzyakov, Y.: Hot experience for cold-adapted microorganisms: Temperature sensitivity of soil enzymes, *Soil Biol. Biochem.*, 105, 236–243, <https://doi.org/10.1016/j.soilbio.2016.11.026>, 2017.





- 450 Rouget, M., Robertson, M. P., Wilson, J. R. U., Hui, C., Essl, F., Renteria, J. L., and Richardson, D. M.: Invasion debt - quantifying future biological invasions, *Diversity Distrib.*, 22, 445–456, <https://doi.org/10.1111/ddi.12408>, 2016.
- Schuur, E. A. G., McGuire, A. D., Schädel, C., Grosse, G., Harden, J. W., Hayes, D. J., Hugelius, G., Koven, C. D., Kuhry, P., Lawrence, D. M., Natali, S. M., Olefeldt, D., Romanovsky, V. E., Schaefer, K., Turetsky, M. R., Treat, C. C., and Vonk, J. E.: Climate change and the permafrost carbon feedback, *Nature*, 520, 171–179, <https://doi.org/10.1038/nature14338>, 2015.
- 455 Sfenthourakis, S. and Hornung, E.: Isopod distribution and climate change, *ZooKeys*, 801, 25–61, <https://doi.org/10.3897/zookeys.801.23533>, 2018.
- Shatilovich, A. V., Tchesunov, A. V., Neretina, T. V., Grabarnik, I. P., Gubin, S. V., Vishnivetskaya, T. A., Onstott, T. C., and Rivkina, E. M.: Viable Nematodes from Late Pleistocene Permafrost of the Kolyma River Lowland, *Dokl Biol Sci*, 480, 100–102, <https://doi.org/10.1134/S0012496618030079>, 2018.
- 460 Shmakova, L., Malavin, S., Iakovenko, N., Vishnivetskaya, T., Shain, D., Plewka, M., and Rivkina, E.: A living bdelloid rotifer from 24,000-year-old Arctic permafrost, *Curr. Biol.*, 31, R697–R713, <https://doi.org/10.1016/j.cub.2021.04.077>, 2021.
- Smith, S. L., O’Neill, H. B., Isaksen, K., Noetzli, J., and Romanovsky, V. E.: The changing thermal state of permafrost, *Nat Rev Earth Environ*, 3, 10–23, <https://doi.org/10.1038/s43017-021-00240-1>, 2022.
- 465 Sørensen, L. I., Holmstrup, M., Maraldo, K., Christensen, S., and Christensen, B.: Soil fauna communities and microbial respiration in high Arctic tundra soils at Zackenberg, Northeast Greenland, *Polar Biol*, 29, 189–195, <https://doi.org/10.1007/s00300-005-0038-9>, 2006.
- Spawn, S. A., Sullivan, C. C., Lark, T. J., and Gibbs, H. K.: Harmonized global maps of above and belowground biomass carbon density in the year 2010, *Sci Data*, 7, 112, <https://doi.org/10.1038/s41597-020-0444-4>, 2020.
- 470 Strauss, J., Schirrmeister, L., Grosse, G., Fortier, D., Hugelius, G., Knoblauch, C., Romanovsky, V., Schädel, C., Schneider von Deimling, T., Schuur, E. A. G., Shmelev, D., Ulrich, M., and Veremeeva, A.: Deep Yedoma permafrost: A synthesis of depositional characteristics and carbon vulnerability, *Earth-Science Reviews*, 172, 75–86, <https://doi.org/10.1016/j.earscirev.2017.07.007>, 2017.
- 475 Tiunov, A. V., Hale, C. M., Holdsworth, A. R., and Vsevolodova-Perel, T. S.: Invasion Patterns of Lumbricidae Into the Previously Earthworm-free Areas of Northeastern Europe and the Western Great Lakes Region of North America, *Biol Invasions*, 8, 1223–1234, <https://doi.org/10.1007/s10530-006-9018-4>, 2006.
- Väisänen, M., Krab, E. J., Monteux, S., Teuber, L. M., Gavazov, K., Weedon, J. T., Keuper, F., and Dorrepaal, E.: Meshes in mesocosms control solute and biota exchange in soils: A step towards disentangling (a)biotic impacts on the fate of thawing permafrost, *Appl. Soil Ecol.*, 151, 103537, <https://doi.org/10.1016/j.apsoil.2020.103537>, 2020.
- 480 Wackett, A. A., Yoo, K., Olofsson, J., and Klaminder, J.: Human-mediated introduction of geoengineering earthworms in the Fennoscandian arctic, *Biol Invasions*, 20, 1377–1386, <https://doi.org/10.1007/s10530-017-1642-7>, 2018.
- 485 Wall, D. H., Bradford, M. A., John, M. G. S., Trofymow, J. A., Behan-Pelletier, V., Bignell, D. E., Dangerfield, J. M., Parton, W. J., Rusek, J., Voigt, W., Wolters, V., Gardel, H. Z., Ayuke, F. O., Bashford, R., Beljakova, O. I., Bohlen, P. J., Brauman, A., Flemming, S., Henschel, J. R., Johnson, D. L., Jones, T. H., Kovarova, M., Kranabetter, J. M., Kutny, L., Lin, K.-C., Maryati, M., Masse, D., Pokarzhevskii, A., Rahman, H., Sabará, M. G., Salamon, J.-A., Swift, M. J., Varela, A., Vasconcelos, H. L., White, D., and Zou, X.: Global decomposition experiment shows soil animal impacts on decomposition are climate-dependent, *Glob Change Biol*, 14, 2661–2677, <https://doi.org/10.1111/j.1365-2486.2008.01672.x>, 2008.

<https://doi.org/10.5194/bg-2022-215>  
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Yashina, S., Gubin, S., Maksimovich, S., Yashina, A., Gakhova, E., and Gilichinsky, D.: Regeneration of whole fertile plants from 30,000-y-old fruit tissue buried in Siberian permafrost, *PNAS*, 109, 4008–4013, <https://doi.org/10.1073/pnas.1118386109>, 2012.