

Sediment Quality Assessment in an industrialized Greek coastal marine area (West Saronikos Gulf)

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Abstract. Eight sediment cores from the coastal marine area of West Saronikos Gulf have been analyzed for grain size and geochemistry. The concentrations of eight metals (Al, Fe, Mn, Cu, Cr, Ni, Pb and Zn) along with total organic carbon (TOC) and carbonate content were measured. The cores are fairly homogeneous, in terms of carbonates and the down-core variability of % TOC, is characterized by high surficial values that decrease with depth. Metal concentrations from both geological (Al, Mn, Cr, Ni) and anthropogenic origin (Cu, Pb, Zn), are higher in the muddy than the sand fraction of sediments. The spatial distribution of Al, Fe, Mn, Cu, Pb and Zn in surface sediments presents increasing concentrations from the northeast to the southwest part of the study area, from the shallow to the deeper parts in contrast to Cr and Ni which are increased in the northern nearshore stations. Based on the vertical distributions, the metal to Al ratios of Cu, Pb and Zn show a constant decrease over depth along most cores, indicating the anthropogenic effects to surface sediments, while Fe/Al is constant. Spearman's correlation analysis performed among the fine grain metal contents, demonstrated a strong positive correlation ($r > 0.5$, $p < 0.05$) between Al, Fe, Mn, Cu, Pb and Zn. The calculated enrichment factors indicate minimal to moderate pollution. The concentrations of Cr at most surface sediments are higher than the ERL value (81 mg Kg^{-1}) but below the ERM value (370 mg Kg^{-1}) and the concentrations of Ni are always higher than the ERM value (51.6 mg Kg^{-1}). In contrast, the concentrations of Cu, Pb, Zn, at most surface sediments, are below ERL values. The mean effects range medium quotients (mERMq) of surface sediments, based on the overall metal concentrations indicated that the surface sediments of most cores, are moderately toxic. The levels of Cr, Ni, Mn and Zn at most stations are decreased in 2017, but the concentrations of Pb and Cu are increased in 2017, compared to a previous study of 2007. The concentrations of Cu, Pb and Zn in the surface sediments of West Saronikos Gulf are lower than levels reported for Inner Saronikos Gulf, Elefsis Bay and other polluted hot spot areas of Greece, owing to a lower degree of urban and industrial development.

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1 Introduction

Sediment cores are one of the most easily accessed natural archives, used to evaluate and reconstruct historical pollution trends in aquatic environments. The cores provide data to characterize sediment physical properties and their geochemistry and composition. Vertical profiles of heavy metals can present sedimentation rate, changes in diagenetic processes and evolution of human pressures. Metals released into aquatic systems, undergo several processes, such as adsorption, photolysis, chemical oxidation and microbial degradation. Sedimentation depends on the contaminant physicochemical properties, the sediment physical properties, the adsorption capabilities and the partitioning constant at the water-sediment interface. Trace metals removed from the water column are adsorbed on particulate matter and eventually deposited on bottom sediments (Bigus et al, 2014).

Sediments are repositories for metals such as chromium, lead, copper, nickel, zinc and manganese that present as discrete compounds, ions held by cation-exchanging clays, bound to hydrated oxides of iron and manganese, or chelated by insoluble

41 humic substances. Solubilization of metals from sedimentary or suspended matter depends on the presence of complexing
42 agents. Metals that are held by suspended particles and sediments are less available than those in true solution (Manahan,
43 2011).

44 Saronikos Gulf (Greece) is a marine area of the Aegean Sea between the peninsulas of Attiki and Argolida. The environmental
45 interest in Saronikos gulf arises from the fact that it is the marine border of the most urbanized areas of Greece, i.e., the
46 country's capital (Athens), the industrial zones of Attiki (Elefsis, Thriasio, Sousaki) and the large port city of Piraeus. As a
47 result, there have been on going environmental monitoring and oceanographic studies of Saronikos since the 80's. There are
48 several published works focusing on the eastern part of Saronikos due to the presence of extended and intensive anthropogenic
49 activities (Scoullou, 1986; Pavlidou et al., 2004; Scoullou et al., 2007; Kontogiannis, 2010; Paraskevopoulou et al., 2014;
50 Pangiotoulas et al., 2017; Karageorgis et al., 2020; Prifti et al., 2022). In contrast, fewer studies have focused on North West
51 Saronikos despite the presence of a less extensive industrial zone hosting a major oil refinery and coastal touristic activities
52 along with a particular geological (volcanic/hydrothermal) background (Paraskevopoulou, 2009; Keleperdis et al., 2001).

53 The main aim of this work is to assess the levels and the distribution of several heavy metals (Al, Fe, Mn, Pb, Zn, Ni, Cr, Cu)
54 in sediment cores of West Saronikos, in order to discern between the relative contribution of geological and anthropogenic
55 origins and to identify the major sources of metal pollution. The second aim is to determine the evolution of marine pollution
56 in the area by comparing the results with those of a similar study ten years ago, conducted at the Laboratory of Environmental
57 Chemistry (Department of Chemistry, National and Kapodistrian University of Athens). The last aim of this work is to assess
58 and highlight the differences between the concentrations of heavy metals in the surface sediments of West Saronikos in
59 comparison to sediments in studies conducted in East Saronikos, Elefsis bay and other areas of Greece.

60 **2 Study area**

61 Saronikos Gulf is situated at the central Aegean Sea (north east Mediterranean) between 37°30'N-38°00'N and 24°01' E-
62 23°00' E formed between the peninsulas of Attiki and Argolida. The length of its coastline is 270 km, the surface is 2.866 km²
63 and the mean water depth 100 m. To the north, a shallow (30 m depth) embayment is formed, known as Elefsis bay. The
64 islands of Salamina and Aigina and the plateau between them, divide the gulf into two basins: the western basin (Western
65 Saronikos Gulf) with maximum depths of 220 m in the north and 440 m in the south and the eastern basin which has a smooth
66 bathymetry with depths of 50-70 m to the north (Inner Saronikos Gulf) reaching 200 m to the southeast, from where the gulf
67 opens to the Aegean Sea (Outer Saronikos Gulf). To the west, the narrow Isthmus of Corinth connects Korinthiakos Gulf with
68 Saronikos Gulf in the Aegean Sea (Kontogiannis, 2010; Paraskevopoulou, 2014).

69 The gulf is subjected to a strong seasonal cycle of heating and cooling, with air temperatures between 0 - 40 ° C, which causes
70 the formation of a seasonal pycnocline from May to November. In winter, the water column is homogenized down to 120 m.
71 However, in the western part, vertical mixing never reached the sea bottom (440 m) in the years after 1992 and dissolved
72 oxygen concentration has approached nearly anoxic conditions (D.O. < 1mL/L) (Paraskevopoulou et al., 2014).

73 A few circulation studies have been conducted in Saronikos in order to discern the possible movement of the treated wastewater
74 effluent discharged at 65m south of Psittalia island in the eastern basin. The circulation is reported as strongly dependent on
75 local winds with predominant northerly direction throughout the year. However, westerly and southerly winds may also occur
76 in fall, winter and spring. Under the predominant northerly wind regime in summer and fall during the presence of the seasonal
77 pycnocline there is a distinctly different two-layer circulation above and below 60m. In the upper layer there is a general
78 eastward anticyclonic (clockwise) flow from the west to the east basin. Below 60m the flow is reversed from the northeast
79 through the Salamina-Aigina passage to the southwest towards the deeper part of the west basin following a cyclonic
80 (counterclockwise) path. During winter when the water column is fully mixed down to 90m the general flow is anticyclonic
81 from the west to east. Nearshore on the northwest Sousaki coast the currents reported are directed from the north to the south.

82 In spring during the pycnocline formation there is no continuous flow structure spanning both the west and east basins. In the
83 west during spring there is a rather strong north to south flow (SoHelME, 2005; Kontoyiannis, 2010).
84 Saronikos gulf, as a whole, is subjected to intense anthropogenic pressure, as it is the marine border of the cities of Athens and
85 Piraeus with 3-4 million inhabitants. Several point and non-point pollution sources are present mainly on the northeastern
86 coasts. One of the most important point sources is the Athens/Piraeus wastewater treatment plant (WWTP) on the small island
87 of Psittalia, among the largest in Europe, with a population equivalent (p.e.) coverage of 5.6 million. Other point sources along
88 the coasts include the port of Piraeus, marinas, touristic facilities, fish farms and the effluents of smaller towns and settlements.
89 Finally, pollution pressure arises from increased marine traffic, since Piraeus is one of the largest and busiest Mediterranean
90 ports, and from the heavy vehicle traffic and the heating systems in the extended urban areas (Paraskevopoulou et al., 2014).
91 The coastal marine area of the north part of West Saronikos Gulf is affected by a few types of industries established there
92 during the 1970's, that include a major oil refinery unit at the center of Susaki area, a cable manufacturer, soya mills, sulfur /
93 fertilizers manufacturing for agricultural use and the increased sewage load from nearby coastal villages due to the summer
94 tourist season (Kelepertsis et al., 2001; Paraskevopoulou, 2009). The settlements on the coast of West Saronikos, to the best
95 of our knowledge, are not yet connected to wastewater treatment facilities and a projected treatment plant in Aghioi Theodoroi
96 is under construction. The Susaki area, which extends parallel to the northern coast of West Saronikos for about 8 Km, is
97 known for its volcanic activity which took place during Pliocene-Quaternary. Most of the volcanic materials were transported
98 by fluvial processes and deposited in the alluvial plains and coastal regions. The formations observed are peridotites and
99 serpentinites, neogene deposits and Quaternary deposits. As a result, elevated values of Cr, Ni, Co, Mn, Fe are found in the
100 soils and sediments of this area and can be explained by the existence of the ultrabasic rocks (Kelepertsis et al., 2001). The
101 southwest deeper part of West Saronikos (Epidavros basin) could be affected by the transport of organic and inorganic
102 pollutants from the eastern basin due to the periodical northeast to southwest water flow below the pycnocline (Psyllidou-
103 Giouranovits and Pavlidou, 1998; Kontoyannis, 2010).

104 3. Materials and methods

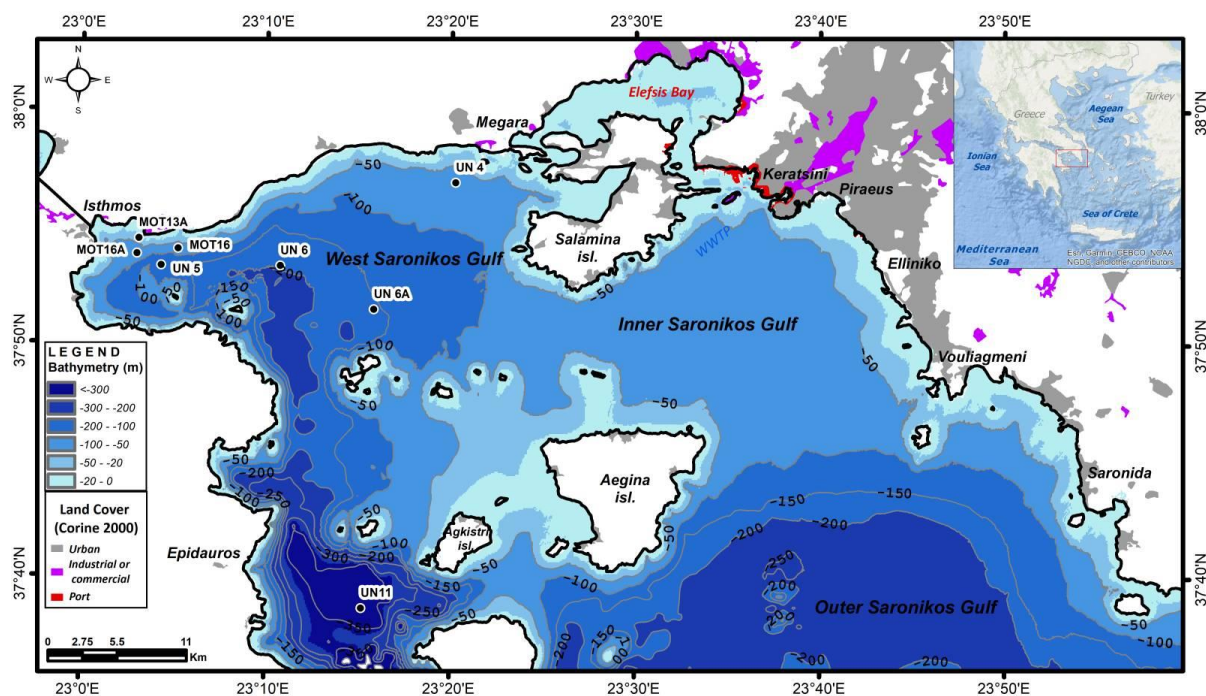
105 Eight short sediment cores (12-32 cm) were obtained at corresponding number of stations with varying depths (50-420 m) in
106 the area of West Saronikos Gulf with the use of a box corer. The sampling was conducted on 18 October 2017 with the Greek
107 Oceanographic vessel RV *Aegaeo*. The West Saronikos Gulf study area and the specific location of stations are presented in
108 Fig. 1.

109
110 **Table 1. Coordinates, water column depth and core length at each sampling station.**

Station	Latitude N (dec.minutes)	Longitude E (dec.minutes)	Depth (m)	Core length (cm)
MOT13A	37° 54.602	23° 03.184	50	12
MOT16A	37° 53.995	23° 03.080	100	32
UN5	37° 53.459	23° 04.393	140	32
MOT16	37° 54.179	23° 05.312	85	20
UN6	37° 53.455	23° 10.857	193	26
UN6A	37° 51.610	23° 15.932	165	24
UN4	37° 57.057	23° 20.331	79	22
UN11	37° 38.800	23° 15.338	420	32

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112 Stations MOT13A, MOT16A, UN5, MOT16, UN6 (near the Susaki area) and UN4 (Megara basin), at the north-western part,
113 are affected by the coastal industrial zone, urbanization and touristic activities. The offshore station UN6A, at the middle of
114 Megara basin, is probably less disturbed by anthropogenic activities. Finally, station UN11, in Epidavros basin at the southwest
115 part of the study area, is influenced by trawling and aquaculture and potential transport of organic and inorganic pollutants
116 from eastern Saronikos due to periodical circulation patterns in the water layers below 60m.

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119 **Figure 1: Study area of West Saronikos Gulf and sampling stations (The map was constructed using Arc Map with bathymetry**
120 **representation data from the European Marine Observation and Data Network-EMODnet: <https://portal.emodnet-bathymetry.eu/>**
121 **and land cover data from Corine 2000).**
122

123 The cores were frozen immediately after sampling and cut in 1cm layers down to the top 10cm of each core and 2cm layers
124 after that. The separated layers were stored frozen until further processing. The initial step of analysis is removal of water
125 content with the use of a Lab Conco freeze-dryer. Subsequently, grain size analysis via dry sieving was performed using the
126 1mm and 63 μ m Retsch stainless steel sieves. The gravel (>1mm), sand (> 63 μ m) and muddy (<63 μ m) fractions were separated
127 for the calculation of the respective percentages to the total sediment. The gravel fraction was discarded. Total organic carbon,
128 carbonates and the concentrations of heavy metals were determined separately in both sand and muddy sediment, when the
129 fraction percentage was more than 10 % of the total sediment or in the prevailing fraction only (above 90%). Table 1 presents
130 the coordinates and depth of each sampling station along with the corresponding core length.

131 The total organic carbon (TOC) content was measured using the standard Walkley (1947) method as modified by Jackson
132 (1958) and Loring & Rantala (1992), which is based on the exothermic reaction (oxidation) of the sediment with potassium
133 dichromate (K₂Cr₂O₇) and concentrated sulfuric acid (H₂SO₄), followed by back-titration with ferrous ammonium sulfate
134 (FeSO₄) and ferroine indicator.

135 The carbonate content was determined by calculating the weight difference of the sample before and after the strong
136 effervescence caused by adding hydrogen chloride (HCl) 6 M to the sediment causing an exothermic reaction followed by HCl
137 gas and CO₂ emission (Loring and Rantala, 1992).

138 The total metal contents were extracted via complete dissolution of sediment samples with an acid mixture of HNO₃-HClO₄-
139 HF (ISO-14869-1:2000) (Peña-Icart et al., 2011). Then, the total metal concentrations were determined by Flame or Graphite
140 Furnace Atomic Absorption Spectroscopy (FAAS-Varian SpectrAA-200 and GFAAS- Varian SpectrAA 640Z) (Skoog et al.,
141 1998). In order to evaluate the precision and accuracy of the method for total metal analysis a certified reference material
142 (PACS-3, NRC-CNRC) was carried through the analytical procedure along with the sediment samples in every digestion batch
143 and 1 or 2 layers of each core were also analyzed in duplicate. Accuracy was calculated as % recovery (percentage
144 ratio of the measured to the certified value). The precision was evaluated using the % RSD-Relative Standard Deviation
145 (percent ratio of the standard deviation to the average concentration of the replicates) calculated for each metal by each of the
146 duplicate measurements (repeatability estimation) and the multiple measurements of the reference material (reproducibility
147 estimation). The quality data for the total metal method are presented in the Appendix (Table A1). The precision and recoveries
148

149 generally fall into the ranges 3-10% RSD and 80-120 % recovery, depending on analyte level. Such results of analytical
 150 performance are anticipated for multistage analysis of solid samples and are roughly recommended by the US Environmental
 151 Protection Agency (US EPA) and the Association of Official Analytical Chemists (AOAC) (US EPA, 1996; AOAC
 152 International, 2016). For every core the %RSD for each metal was calculated as an indication of the downcore variability and
 153 the results are presented in Fig. A1 – A15 (Appendix A).

154 Most statistical treatment of data and the vertical distribution graphs were performed by Microsoft Excel 2010. The horizontal
 155 distributions of metals were visualized with the software package Ocean Data View (ODV) 2017. The software IBM-SPSS
 156 Statistics 2020 was used for statistical comparisons between stations regarding the trace metal concentrations and Spearman
 157 correlation analysis to identify significant relationships between different heavy metals, total organic carbon and carbonates.
 158 The comparison between stations was done by performing parametric (One-Way Anova) and non-parametric (Kruskal Wallis,
 159 Kolmogorov-Smirnof) tests between the mean values and standard deviations calculated for each metal and each core by the
 160 results of the top 5 cm. The comparison between two groups of values, e.g. a variable between two cores or concentrations of
 161 variables above-below a certain layer in a single core was either done by two-sided t-test or by the non-parametric equivalents
 162 Mann-Whitney and Kolmogorov-Smirnov tests. In all comparative tests statistical results were calculated at the 95%
 163 confidence level ($p < 0.05$).

164 4 Results

165 4.1 Geochemical results

166 Table 2 summarizes the main findings from the determination of geochemical parameters. The grain size in cores from stations
 167 MOT16A, UN5, UN6, UN6A, UN11, is dominated by mud, while the percentage of sand fraction is generally low and below
 168 10% in most core layers. On the other hand, in cores MOT13A, MOT16, UN4 the percentages of both sand (39-66%) and mud
 169 (15-50%) are significant. The sediments at stations MOT13A and MOT16 are coarser with average mud percentages 25 and
 170 38% respectively, while at station UN4 especially in the top 10cm the average mud percentage is slightly increased (43%). As
 171 a result, the percentages of total organic carbon (TOC) and carbonates as well as the concentrations of heavy metals were
 172 determined in the mud fraction ($< 63\mu\text{m}$) of sediments in cores MOT16A, UN5, UN6, UN6A and UN11 and separately in the
 173 sand and mud fraction of sediments in cores MOT13A, MOT16 and UN4.

174

175 **Table 2. Summary statistics of grain size percentages, TOC and carbonate content along the collected cores.**

Station	% sand		% mud (silt and clay)		% TOC		% CO_3^{2-}	
	Min	Max	Min	Max	Min	Max	Min	Max
MOT13A	51	66	15	38	0.45	0.92	26	28
MOT16A	1	23	77	99	0.10	0.93	22	23
UN5	1	14	86	99	0.44	1.12	20	23
MOT16	47	65	29	45	0.33	1.28	20	22
UN6	1	32	68	99	0.57	2.44	19	23
UN6A	1	22	78	99	0.33	1.32	22	25
UN4	39	60	20	50	0.51	0.75	29	33
UN11	0	15	74	100	0.77	2.35	15	19

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177 Apart from small variations, the cores are fairly homogeneous, in terms of carbonates. The high percentages of % CO_3^{2-} in
 178 cores MOT13A and UN4 are associated with the coarse - grained samples and abundant presence of shell fragments. Figure 2
 179 presents the vertical distribution of % TOC in selected cores. Comparing between stations the lowest TOC values in the top
 180 5cm are measured in the coarser sediments of MOT13A (average 0.57%) and the highest in the deepest station UN11 (average
 181 1.74%). There was no statistical difference between TOC in the remaining cores and the average values for the top 5cm are
 182 the following: UN4 (0.68%), MOT16A (0.69%), MOT16 (0.96%), UN5 (0.96%), UN6A (1.09%) and UN6 (1.26%).

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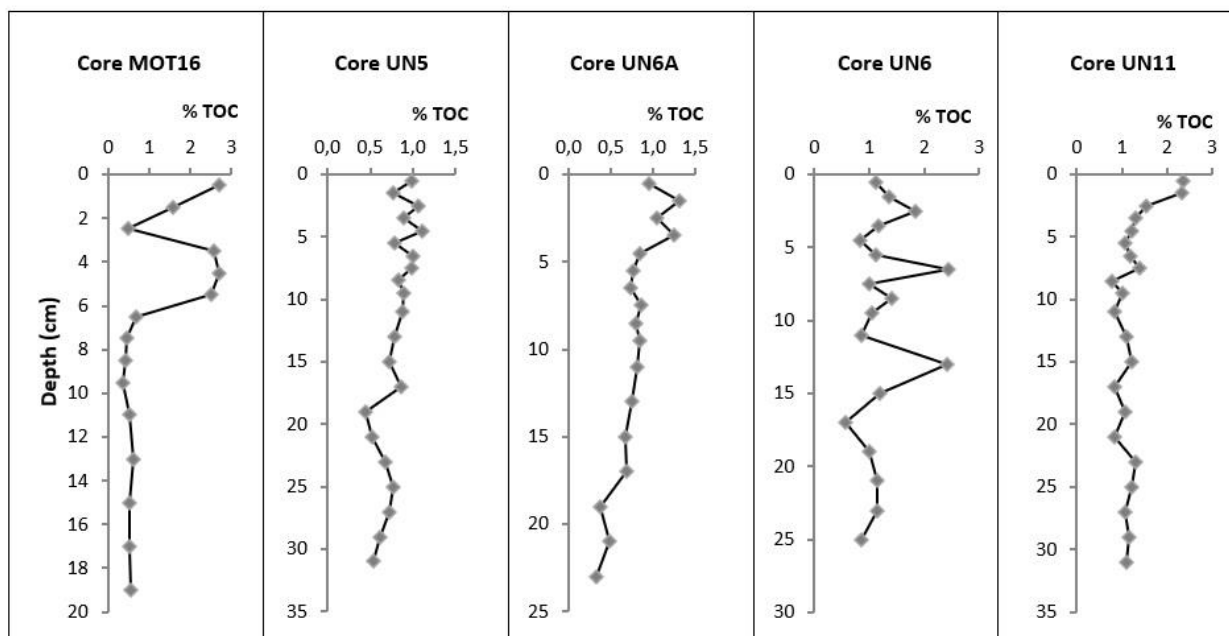


Figure 2: Vertical distribution of % TOC at selected cores [MOT16 profile depicts fine sediment (< 63 μm) TOC content].

The vertical TOC distribution in core MOT16 is differentiated than the corresponding in the adjacent station MOT13A. Below 6cm the levels of TOC are similar in both stations. However, in the top 6cm of core MOT16 average TOC in the total sediment is approximately 1% while below 6cm the corresponding average is 0,42%. This is caused by a striking increase in the TOC content of the muddy sediments, with an average of 2.4% TOC at the top 6cm approximately 5-fold higher than the corresponding in the deeper layers (0.5%). At the same time the TOC variability in the sandy fraction of the top 6cm of MOT16 is minimal, ranging between 0.3% and 0.4%. The vertical TOC distributions of the dominantly muddy cores (MOT16A, UN5, UN6, UN6A, UN11) present high surficial content and gradual increase with depth.

Table 3 presents TOC, carbonate and heavy metal contents in the two fractions (sand-mud) at the shallow coarse-grained cores MOT13A, MOT16, UN4. It is apparent that the % TOC content of the muddy material is higher than the corresponding content of the sandy fraction as anticipated. The % CaCO_3 content is increased in the sandy material of cores MOT13A, UN4 compared to that of the fine sediment and is approximately equal in the two sediment fractions of core MOT16.

The concentrations of Al, Cr, Cu, Mn, Pb, Ni and Zn are higher in the muddy sediments of MOT13A and UN4 than the corresponding values in sandy sediments. The same applies to Al, Mn, Cu, and Zn in MOT16. Unlike other metals the Fe content is more or less similar in both sediment fractions of cores MOT13A and UN4. The sediments of core MOT16 appear to be different with higher concentrations of Fe, Cr, Ni and Pb in the coarse-grained fraction.

The vertical distributions of the studied metals in mg kg^{-1} along the collected cores are presented in Figures A1–A15 (Appendix A). In cases of coarse-grained cores MOT13A, MOT16, UN4, the concentrations in the total sediment (calculated by the corresponding values in both fractions) are depicted. Table A2 (Appendix A) presents the ratios of eight heavy metals to Al in the surface and deeper sediment layer of the collected cores. The ratios presented for the cores MOT13A, MOT16, UN4 are calculated by the results in the corresponding fine-grained sediments. The vertical profiles of metal to aluminum ratios (in fine grained sediments) along the collected cores, are also given in Figures A1–A15 (Appendix A).

Table 4 summarizes the concentrations of eight heavy metals in the surface and deeper sediment layer of the collected cores. For the sake of direct comparison in the cases of cores MOT13A, MOT16, UN4 the fine grained concentrations are presented in Table 4. The sediments at the deeper parts of West Saronikos (cores UN6, UN6A and UN11) present elevated concentrations of Al, Fe, Pb and Zn. The sediments of UN11 are also particularly enriched in Mn. On the contrary the values of Cr and Ni are elevated in cores MOT13A, MOT16 and MOT16A.

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Table 3. Concentrations of metals (in mg Kg⁻¹), organic and inorganic carbon content in the two sediment fractions of coarse-grained cores (MOT13A, MOT16, UN4).

Variable/Core	MOT13A		MOT16		UN4	
	fine fraction	coarse fraction	fine fraction	coarse fraction	fine fraction	coarse fraction
% TOC	0.64-2.80	0.26-0.45	0.37-2.70	0.15-0.46	0.65-0.94	0.43-0.58
% CaCO ₃	22-23	27-30	21-23	18-22	26-29	32-36
Al	10561-18387	6615-9226	21677-28939	10610-18538	23271-34246	10449-19765
Cr	390-651	333-486	322-374	306-517	113-133	71.6-115
Ni	293-411	220-302	314-573	424-697	139-187	78.0-140
Fe	19878-21153	17430-20643	24130-31694	32265-36080	14149-17126	13501-17931
Mn	429-476	326-386	471-530	435-484	317-366	174-238
Cu	17.1-18.8	8.9-11.8	16.7-22.6	6.7-14.8	15.1-23.6	9.7-13.4
Pb	18.0-30.7	14.8-21.0	9.1-28.0	8.3-31.9	11.2-31.2	9.3-27.3
Zn	39.3-51.8	31.0-45.6	39.8-53.7	31.5-50.9	38.9-59.2	30.3-50.3

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Table 4. Concentrations in mg Kg⁻¹ of metals in the surface and the deeper sediment layer of cores. The concentrations at coarse-grained cores MOT13A, MOT16, UN4 refer to the fine sediment fraction (f < 63µm).

Core	Layer (cm)	Al	Cr	Ni	Fe	Mn	Cu	Pb	Zn
MOT13A	0-1	10561	651	411	20988	476	18.2	26.1	49.6
	10-12	14673	407	383	21153	444	18.1	22.5	44.6
MOT16A	0-1	27009	280	375	23315	578	22.6	20.4	48.3
	30-32	30070	256	382	26179	562	22.1	28.1	44.1
UN5	0-1	32705	199	305	25534	635	27.4	42.7	74.5
	30-32	37885	223	320	29170	520	26.0	24.7	57.5
MOT16	0-1	21677	369	377	25816	530	22.6	28.0	53.7
	18-20	23807	348	401	29752	482	16.9	9.1	40.4
UN6	0-1	43264	142	253	27301	954	36.7	52.9	92.1
	24-26	44547	153	274	29345	707	28.5	32.5	62.4
UN6A	0-1	39314	146	187	21838	570	26.3	38.4	73.8
	22-24	41303	161	209	23827	513	25.4	38.7	58.7
UN4	0-1	34246	132	162	15762	358	23.6	26.1	52.1
	20-22	31706	113	160	17112	365	15.1	12.0	40.3
UN11	0-1	54626	142	230	32177	3925	49.7	63.9	110
	30-32	48186	163	217	31573	1459	37.3	35.1	71.4

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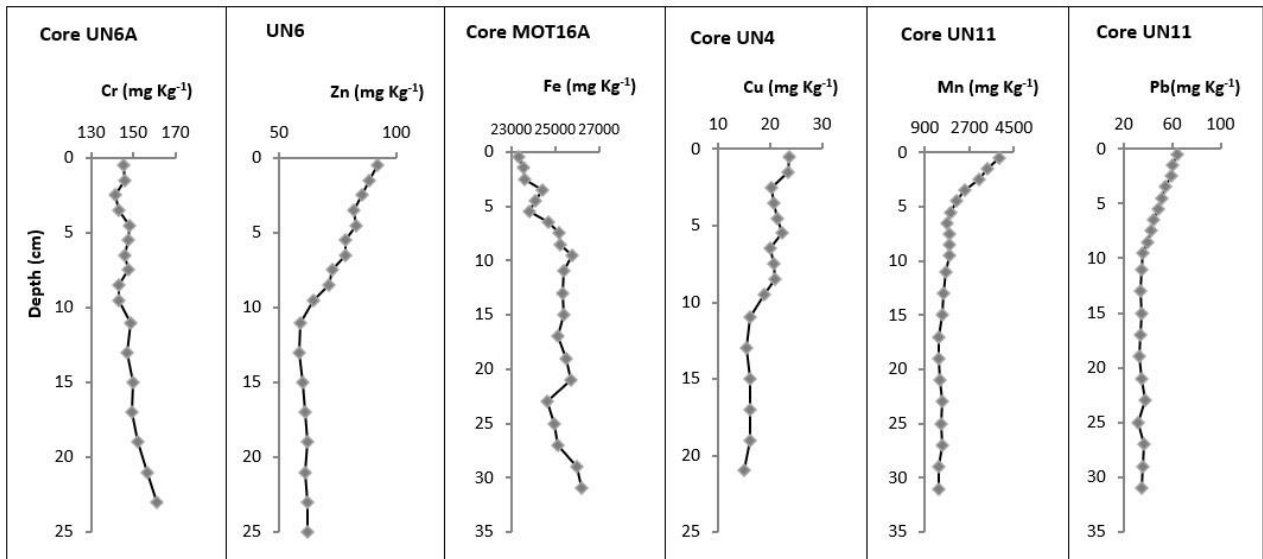
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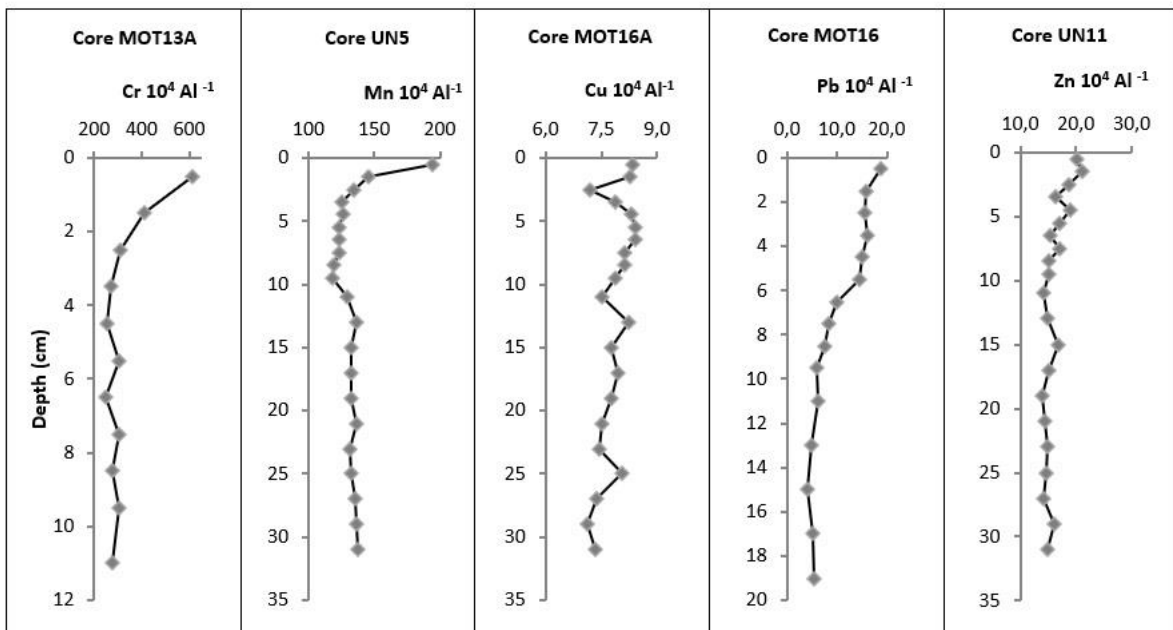
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The vertical distributions of Al (Fig. A1) for the total sediment (both fractions) present minimal variation. The down-core variability of Al is lower than 10 %, in the finer sediments (MOT16A, UN5, UN6, UN6A and UN11) and between 10-15 % in the coarser sediments (MOT13A, MOT16 and UN4). In the surface layer (0-1cm) of cores MOT13A and UN5 there is a sharp decrease of Al content. Furthermore, the average aluminum content (16709mgKg⁻¹) at the top 6cm of core MOT16 is statistically lower (two-sided t-test, p<0,05) than the corresponding average content in the layers below (20477mgKg⁻¹). On the contrary, at the top 12cm of core UN11 aluminum (average 52714mgKg⁻¹) is increased (two-sided t-test, p<0,05) compared to the deeper layers (average 48621 mgKg⁻¹). The other so-called lithogenic metals (Cr, Ni, Fe, Mn) generally present uniform vertical profiles with minimal variability, mostly below 10 % (Figures A2-A9). However, Cr, Ni, Fe and Mn in the surface sediments of MOT13A present an increasing trend, that remains pronounced after normalization to Al in the muddy fraction at the top 0-1cm and is concurrent with the 2-fold decrease of Al content. The vertical profiles of Cr and normalized Cr/Al in UN11 indicate a decrease of Cr in the upper sediment layers. This decrease is confirmed statistically and the average Cr (148mgKg⁻¹) and Cr/Al (28,0) values of the top 12cm are lower than the corresponding in the deeper layers (155 mgKg⁻¹and 32,0, respectively). The same trend of decrease in the top 0-1cm is seen in the normalized profiles of Fe, Ni and Mn in the fine-grained sediment fraction of core UN4. The average concentrations of Fe, Mn, and Ni in the top 12cm of UN11 are statistically higher than the corresponding in the deeper layers. The most pronounced difference is calculated for Mn, where the average upper layer content is 2396 mgKg⁻¹ while the deeper layer content is 1527 mgKg⁻¹. The same applies to Mn/Al ratio in UN11 but not to the ratios of Fe/Al and Ni/Al which are seemingly lower in the upper layers like Cr/Al but statistically equal. The down-core variability of Mn in all stations, except UN4, is typical of shelf sediments, with high surficial Mn concentrations and Mn/Al ratios that diminish with depth

242 Figure 3 presents selected vertical profiles of Mn and Pb along core UN11 and those of Cr, Fe, Cu, Zn at stations UN6A,
 243 MOT16A, UN4, UN6, respectively. Figure 4 presents selected vertical distributions of metal to Al ratios at cores MOT13A,
 244 UN5, MOT16A, MOT16 and UN11. The concentrations and ratios presented in Figures 3 and 4 refer to the muddy fraction of
 245 the sediments in the cases of the coarse-grained cores (MOT13A, MOT16, UN4).
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247
 248 **Figure 3: Vertical profiles of Cr, Fe, Cu, Zn at the muddy sediments of cores UN6A, MOT16A, UN4, UN6 respectively and Mn, Pb**
 249 **along core UN11.**
 250

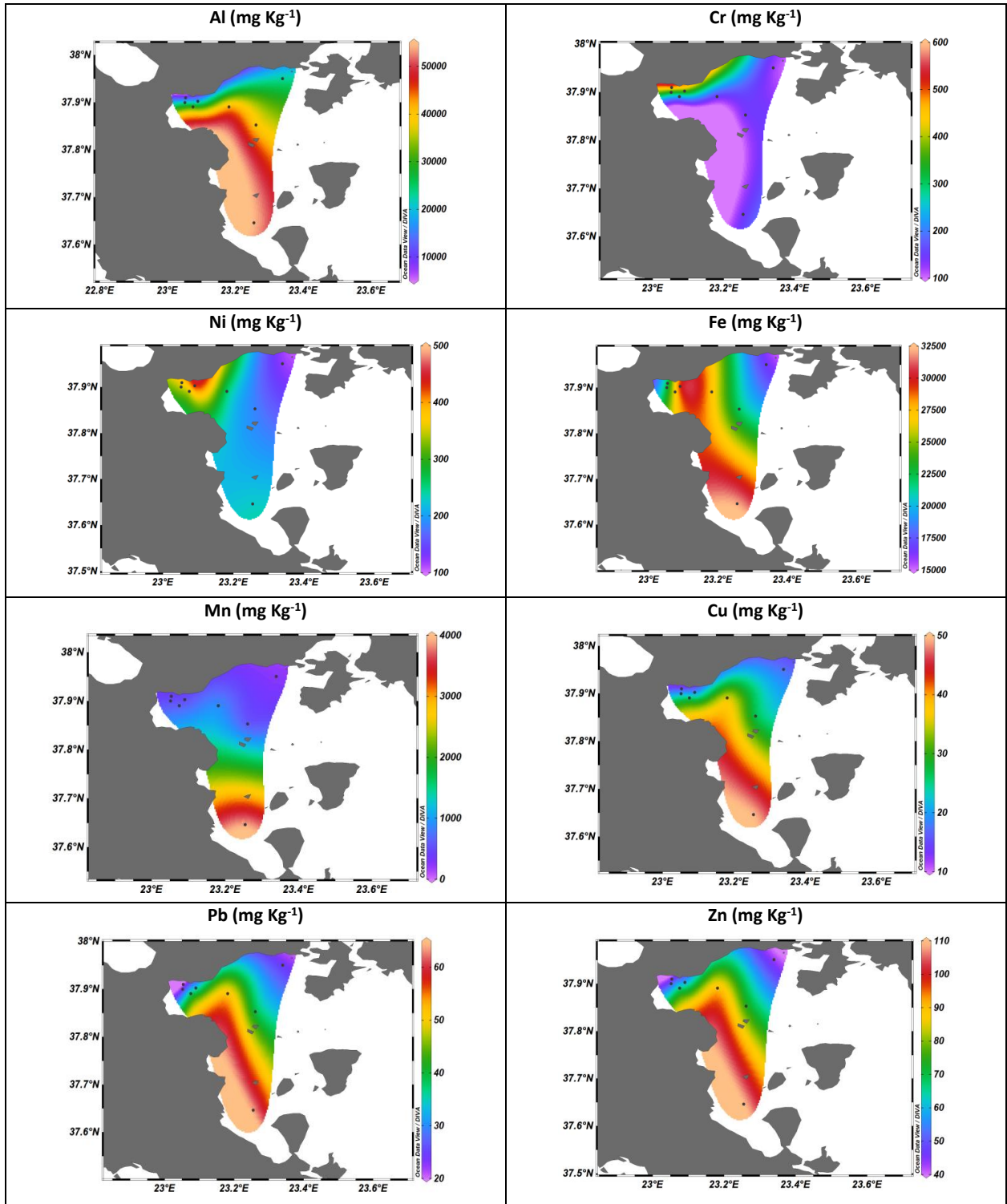


251
 252 **Figure 4: Vertical distributions of ratios to Al at cores MOT13A, UN5, MOT16A, MOT16 and UN11 of northwest Saronikos Gulf**
 253 **(profiles for MOT13A and MOT16 depict ratios calculated for the muddy fraction).**
 254

255 The concentrations of Cu, Pb and Zn as well as the normalized profiles (Fig. A10–A15) show a constant decrease over depth
 256 to background levels. That tendency is less pronounced with no statistical difference between upper and deeper sediment layers
 257 in very few cases and particularly for Cu (MOT13A, MOT16, MOT16A), Cu/Al (MOT13A), Pb (MOT13A, MOT16A), Pb/Al
 258 (MOT16A) and Zn (MOT16A). In all remaining cores the concentrations of Cu, Pb and Zn and the ratios to aluminum in the
 259 upper sediment layers (above 10cm) are statistically higher than the corresponding in the deeper layers.

260 **4.2 Horizontal distributions**

261 Figure 5 presents the horizontal distributions of heavy metals in the surface sediments (0-1 cm) of the study area. In cases of
262 coarse surface sediments of MOT13A, MOT16, UN4, the concentrations of total sediment fraction (sand and mud) were used.
263 The concentrations of Al, Fe, Mn, Cu, Pb and Zn are plotted increased from the northeast to the southwest area of West
264 Saronikos Gulf. On the other hand, the concentrations of Fe, Cr and Ni at the north part appear higher than those at the south
265 area.
266



268
269 **Figure 5: The horizontal distributions of heavy metals in the surface sediments of West Saronikos Gulf. In cases of coarse surface**
270 **sediments of stations MOT13A, MOT16, UN4, the concentrations of the total sediment (sand and mud) were used.**
271

272 In order to corroborate the above plots and produce solid statistical comparisons for the metal concentrations between stations
273 the average values and standard deviation from the results of the top 5 cm were calculated and examined used the IBM-SPSS
274 Statistics 2020 software. In most cases of metals and stations the 5 value groups followed the normal distribution therefore the
275 parametric One-Way Anova test was first applied. But since some cases did not follow the normal distribution non-parametric
276 tests (Kruskal-Wallis for multiple groups and Mann-Whitney / Kolmogorov-Smirnov for two groups) were also applied to
277 corroborate the statistical comparison outcome. The results summary is the following:

278 Al: UN11 >> UN6 = UN6A > UN5 > MOT16A > UN4 = MOT16 >> MOT13A

279 Fe: UN11 = MOT16 > UN6 = UN5 > MOT16A > MOT13A = UN6A > UN4

280 Mn: UN11 >> UN6 > MOT16A = UN6A = UN5 = MOT16 > MOT13A > UN4

281 Cr: MOT13A = MOT16 > MOT16A > UN5 > UN6A = UN11 = UN6 > UN4

282 Ni: MOT16 > MOT16A > UN5 = MOT13A = UN6 > UN11 > UN6A > UN4

283 Cu: UN11 > UN6 = UN5 > UN6A > MOT 16A > UN4 = MOT16 > MOT13A

284 Pb: UN11 = UN6 > UN5 = UN6A > UN4 = MOT16 > MOT 16A = MOT13A

285 Zn: UN11 > UN6 > UN5 = UN6A > MOT16 = UN4 = MOT 16A = MOT13A.

286 **5 Discussion**

287 **5.1 Geochemical findings and element interrelations**

288 The spatial grain size distribution falls into an expected pattern with coarser sediments closer to the coastline (stations
289 MOT13A, MOT16, UN4) and finer sediments in the more remote and deeper stations MOT16A, UN5, UN6, UN6A, UN11.

290 The observed spatial variability of TOC with higher content in the predominantly muddy cores is expected since fine-grained
291 sediments are known to contain elevated levels of organic material compared to coarse-grained sediments (Salomons &
292 Forstner, 1984). The pronounced increase of TOC in the muddy sediments of station MOT16 (top 6cm) could be attributed to
293 the treated wastewater effluents of the refinery which are discharged in the general vicinity of this station (Paraskevopoulou;
294 2009). The two deep stations UN6 (200m at Megara basin) and UN11 (440m at Epidavros basin) present the highest TOC
295 content. It has been stipulated that suspended matter from eastern Saronikos basin, which is affected by more pronounced
296 polluting activities is transported below the thermocline to the west basin (Psyllidou-Giouranovits and Pavlidou, 1998;
297 Kontoyannis, 2010). This could explain the gradual increase of TOC content in the more recently deposited upper layers of
298 sediment.

299 Fine-grained sediments are known to contain abundant geochemical phases, such as clay minerals, organic material and Fe-
300 Mn oxy-hydroxides with high affinity for trace metals due to increased surface adsorption and ionic attraction. Thus, the
301 elevated levels of Al and the so-called anthropogenic metals (Cu, Pb, Zn) in the muddy sediments of the study area can be
302 attributed to the predominant occurrence of aluminum-rich aluminosilicate and clay minerals in fine-grained sediments and
303 their concurrent efficacy to bind or sorb trace metals (Barjy et al., 2020; Karageorgis et al., 2005; Salomons & Forstner, 1984).
304 In contrast Cr and Ni which are also regarded as lithogenic elements along with Al, present a different distribution and are
305 elevated in the coarser near-shore sediments of cores MOT13A, MOT16 and MOT16A. This can be attributed to the existence
306 of ultrabasic rocks in the coastal Susaki area, in which, these metals are abundant (Kelepertsis et al., 2001). In the case of Fe
307 and Mn the lower contents at station UN4 can be attributed to the coarser sediments and the different geological setting with
308 absence of ultrabasic rocks.

309 The minimal depth variations observed at the vertical distributions of major (Al) and lithogenic elements (Fe, Cr, Ni) in the
310 study area are anticipated because of their terrigenous origin (Nolting et al., 1999; Karageorgis et al., 2005;), however some
311 of the discrepancies identified in specific cores cannot be readily explained due to lack of data for other elements such as Ti,
312 Si and Ca.

313 The down-core variability of Mn in all stations, except UN4, is typical of shelf sediments, with high surficial Mn concentrations
314 that diminish with depth, as reducing conditions develop. These variations are largely independent of lithological or carbonate
315 content fluctuations, and are attributed solely on the respiration of organic carbon and the redox-cycle of Mn (Karageorgis et
316 al., 2005; Sundby, 2006). The Mn enrichment is far more pronounced in the surface sediments of station UN11 and subtle in
317 the other muddy cores. In station UN11 the bottom waters have been hypoxic or near anoxic since the mid 90's (Kontoyiannis,
318 2010; Paraskevopoulou et al., 2014). A possible explanation is that the slow diffusion of dissolved oxygen from the more
319 oxidizing overlying waters and the upward diffusion of dissolved Mn (II) from the pore water of anoxic surface sediments to
320 the sediment/water interface (Ozturk, 1995), cause the oxidation of dissolved Mn (II) and its precipitation as Mn (IV) oxides
321 (Pohl and Hennings, 1999). The Mn enrichment in the surface layers of UN11 resembles concentrations in sediments from
322 suboxic parts of the Black Sea (Kiratli & Ergin, 1996; Chen et al., 2022).

323 The normalized profiles of Cu, Pb and Zn exhibit statistically higher metal ratios in the upper layers of the cores, which is a
324 typical indication of the effect of modern pollution sources to recently deposited sediments in contrast to pre-industrial
325 deposition of the deeper layers (Karageorgis. et al., 2005).

326 The concentrations of Al, Fe, Mn, Cu, Pb and Zn are increased from the northeast to the southwest area of West Saronikos
327 Gulf generally following the distribution of muddy sediments. The opposite increase of Cr and Ni to the north of the study
328 area is attributed to the ultrabasic geological substrate of the Susaki coastal area. The exception of increased Fe concentrations
329 in one of the northern sandy cores (MOT16) can also be attributed to the geology of the coastal region (Kelepertsis et al.,
330 2001).

331 Spearman's correlation analysis was carried out to determine the relationships between heavy metals and percentages of total
332 organic carbon (TOC) and carbonates in sediments of the collected cores. The concentrations of metals and the percentages of
333 organic and inorganic carbon used for the analysis correspond to the fine fraction in the sediments of stations MOT13A,
334 MOT16, UN4. Spearman's correlation coefficients are presented in Table A3 (appendix A) and Fig. A16 (appendix A).

335 Al is highly correlated ($r > 0.5$, $p < 0.05$) with Fe, Mn, Cu, Pb and Zn, which probably indicates an association between these
336 metals in the form of metal-clay complexes of continental origin (Barjy et al., 2020). On the other hand, there is a negative
337 correlation of Al with Cr and Ni.

338 Cr is highly correlated with Ni, but both of them show negative correlation with Cu, Pb and Zn, which can be attributed to
339 their different origin (Barjy et al., 2020) and poor correlation with Mn. Cr shows poor correlation with Fe, as well. The strong
340 correlation between Cr and Ni can be observed at sediments of the northwest part, too.

341 Fe, Mn, Cu, Pb and Zn show positive correlation with each other. Cu, Pb and Zn are highly correlated with each other ($r > 0.5$,
342 $p < 0.05$), which can be observed at sediments of the northwest part, too, suggesting that they have a common origin and
343 identical behavior during transport in the marine environment (Barjy et al., 2020).

344 The % TOC content presents moderate correlation with Al, Cu, Pb and Zn and negative correlation with Cr, Ni and %
345 carbonates. Moreover, it shows poor correlation with Fe and Mn. Finally, the percentage of carbonates content presents
346 negative correlation with all metals.

347

348 **5.2 Enrichment Factors**

349 The Enrichment Factors (EF) are used to distinguish between metals originating from anthropogenic activities and from natural
350 processes, assessing the degree of anthropogenic effect. Equation (1) was used for the calculations of EFs, where C_x is the
351 concentration of the analyzed metal and C_{EN} is the concentration of the normalizing element. Al was used as the reference
352 element.

$$353 \text{ EF} = \frac{(C_x/C_{EN})_{\text{sample}}}{(C_x/C_{EN})_{\text{background}}} \quad (1)$$

354 In Table 5, the categories of contamination according to the Enrichment Factor are presented (Diamantopoulou et al., 2019;
 355 Sutherland, 2000). In general, EFs use concentrations normalized to Al to account for the heterogeneity of the samples due to
 356 differences in texture and organic content (Gredilla et al., 2015).

357
 358 **Table 5. The categories of pollution according to the Enrichment Factor.**

EF	Contamination Degree
< 2	Depletion to minimal enrichment- no or minimal pollution
2 to 5	Moderate enrichment- moderate pollution
5 to 20	Significant enrichment- significant pollution
20 to 40	Very high enrichment- very strong pollution
>40	Extreme enrichment- extreme pollution

359
 360 Table 6 presents the Enrichment Factors of the surface sediments (0-1 cm) that were calculated based on the measured
 361 concentrations of heavy metals. The EFs were calculated at the fine fraction ($f < 63\mu\text{m}$) of sediments of cores MOT13A,
 362 MOT16, UN4. Most metals present minimal to moderate enrichment in almost all the cores analysed. Moderate enrichment is
 363 found for Cr, Ni, Mn and Pb in core MOT13A, Mn in UN11, and finally for Pb in UN5 and MOT16.

364
 365 **Table 6. Enrichment Factors of the surface sediments (0-1cm) of the study area. In cases of stations MOT13, MOT16, UN4, the EFs**
 366 **were calculated at the fine surface sediment fraction ($f < 63\mu\text{m}$).**

Core	EF Cr	EF Ni	EF Fe	EF Mn	EF Cu	EF Pb	EF Zn
MOT 13 A	3.09	2.07	1.92	2.07	1.94	2.24	2.14
MOT 16A	1.36	1.22	1.10	1.27	1.27	0.90	1.36
UN 5	1.20	1.28	1.17	1.64	1.42	2.32	1.74
MOT 16	1.28	1.14	1.05	1.33	1.61	3.71	1.60
UN6	0.98	0.98	0.99	1.43	1.36	1.73	1.57
UN 6A	1.00	0.99	1.01	1.23	1.15	1.09	1.39
UN 4	1.00	0.87	0.79	0.84	1.34	1.86	1.11
UN11	0.68	0.82	0.79	2.09	1.04	1.42	1.20

369 5.3 Sediment Quality Guidelines

370 Sediment Quality Guidelines (SQG) of effects range low (ERL) and effects range median (ERM) are used to assess the level
 371 of toxicity of metals in the surface sediments. Metal concentrations below the ERL value, indicate that effects on biota are
 372 rarely observed. Concentrations above the ERL but below the ERM, occasionally affect the biota and concentrations above
 373 the ERM frequently affect the biota. The ERL and ERM guideline values for trace metals (mgKg^{-1} , dry wt) and percent
 374 incidence of biological effects in concentrations ranges defined by the two values are presented in Table 7 (Long et al., 1995).

375
 376 **Table 7. ERL and ERM guideline values for trace metals (mgKg^{-1} , dry wt) and percent incidence of biological effects in concentration**
 377 **ranges defined by the two values.**

Metal	ERL (mg kg^{-1})	ERM (mg kg^{-1})	Percent incidence of effects		
			<ERL	ERL-ERM	>ERM
Cr	81	370	2.9	21.1	95.0
Cu	34	270	9.4	29.1	83.7
Pb	46.7	218	8.0	35.8	90.2
Ni	20.9	51.6	1.9	16.7	16.9
Zn	150	410	6.1	47.0	69.8

379
 380 In this study, the concentrations of heavy metals at the surface sediments were compared with the ERL and ERM criteria. For
 381 cores MOT13A, MOT16, UN4, the concentrations in the total sediment ($< 1\text{mm}$) were used. The concentrations of Cr in the
 382 surface sediments of cores UN4, UN5, MOT16A, UN6, UN6A and UN11 are higher than the ERL value (81 mg Kg^{-1}) but
 383 below the ERM value (370 mg Kg^{-1}) and the values at surface sediments of MOT13A and MOT16 are higher than the ERM

384 value. The concentrations of Ni at the surface sediments of the collected cores are higher than the ERM value (51.6 mg Kg⁻¹)
 385 and as a result, they frequently affect the biota (Long et al., 1995; Hahladakis et al., 2012).
 386 The concentrations of Cu and Pb at surface sediments of stations MOT13A, MOT16A, UN4, UN5, MOT16 and UN6A, are
 387 below the ERL values (34 mg Kg⁻¹ for Cu and 46.7 mg Kg⁻¹ for Pb), which indicates that effects on biota are rarely observed.
 388 On the other hand, the concentrations at surface sediments of cores UN6 and UN11 are higher than the ERL values but below
 389 the ERM values (270 mg Kg⁻¹ for Cu and 218 mg Kg⁻¹ for Pb), which means that they can occasionally affect biota . The
 390 concentrations of Zn are below the ERL (150 mg Kg⁻¹) value and the ERM value (410 mg Kg⁻¹), which indicates that effects
 391 on biota are rarely observed (Long et al., 1995; Hahladakis et al., 2012).

392 5.4 Mean effects range medium quotients

393 The mean effects range medium quotient (mERMq) is an index that is used to evaluate the possible biological effects of the
 394 coupled toxicity of all heavy metals in the surface sediments (Gredilla et al., 2015). Briefly, mERMq's were calculated by
 395 dividing the average concentration of each metal at the top 9cm, by its respective ERM (effects range median), to obtain the
 396 corresponding sediment quality guideline quotient (ERMq). Following this, mERMq's for each core were obtained as the
 397 average of ERMqs previously calculated. ERMqs indicate the pollutant concentration above which effects are expected to be
 398 frequent and have been only defined for very toxic elements (Gredilla et al., 2015).

399 In this study, Cr, Ni, Cu, Pb and Zn were considered in our calculations and the results are depicted in Table 8. In cases of
 400 cores MOT13A, MOT16, UN4, the concentrations of the total sediment (< 1mm) were used for the calculation of mERMq.
 401 Values of mERMq in the ranges of 0.0-0.1, 0.1-0.5, 0.5-1.5 and >1.5 correspond to the following probabilities of toxicity: 9
 402 % (non-toxic), 21 % (slightly toxic), 49 % (moderately toxic) and 76 % (highly toxic), respectively (Gredilla et al., 2015). The
 403 mERMq values obtained for the sediments varied from 0.62 to 2.00, which means that the sediments are moderately or highly
 404 toxic.

405 The concentrations of Cr, Ni, Cu, Pb, Zn in sediments of cores MOT13A, UN5, UN6, UN6A, UN4 and UN11 are moderately
 406 toxic and in sediments of MOT16A and MOT16 highly toxic. The concentrations of Cu, Pb, Zn in sediments of cores
 407 MOT13A, MOT16A, MOT16, UN4 are non-toxic, with mERMq ranging from 0.08 to 0.10 and those in sediments of UN5,
 408 UN6, UN6A, UN11 are slightly toxic, with mERMq range 0.16-0.21.

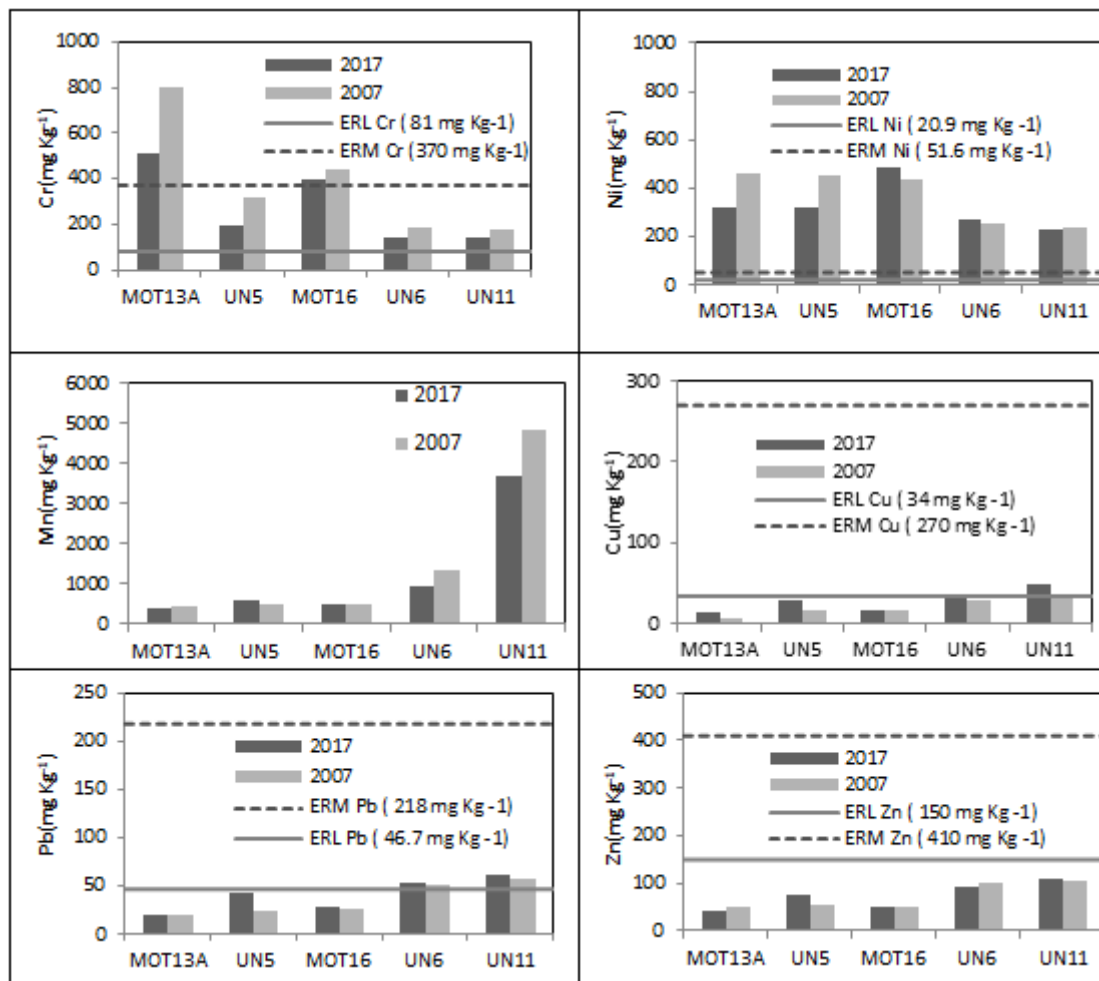
410 **Table 8. MERMqs values calculated for the surface sediments (0-9 cm) of the collected cores of West Saronikos Gulf, by dividing**
 411 **the average concentration (mg Kg⁻¹) of each metal (Cr, Ni, Cu, Pb, Zn) by its respective ERM (mg Kg⁻¹). In cases of cores MOT13A,**
 412 **MOT16, UN4, the concentrations of total sediment (< 1mm) were used.**

Core	mERMq (average)	toxicity of sediments
MOT13A	1.46	Moderately toxic
MOT16A	1.69	Highly toxic
UN5	1.46	Moderately toxic
MOT16	2.00	Highly toxic
UN6	1.21	Moderately toxic
UN6A	0.95	Moderately toxic
UN4	0.62	Moderately toxic
UN11	1.09	Moderately toxic

415 5.5 Evolution of marine pollution

416 The total concentrations of eight heavy metals in the surface sediments were compared with those of a similar study ten years
 417 ago (Paraskevopoulou, 2009). In cores MOT13A, MOT16, UN4, the concentrations of the total sediment fraction (<1mm)
 418 were used and the results are depicted in Fig.6. The levels of Cr, Ni and Mn at most sediments, are decreased in 2017, compared

419 to the study of 2007. On the other hand, the levels of Pb and Cu are increased in 2017, compared to the study of 2007. Moreover,
 420 the levels of Zn, at most sediments, are decreased in 2017 compared to the study of 2007.
 421



422
 423 **Figure 6: Levels of heavy metals in surface sediments of 2017 and 2007 plotted with sediment quality guidelines. For the coarse**
 424 **sediments of stations MOT13A, MOT16, UN4, the concentrations of total sediment (< 1mm) were used.**
 425

426 **5.6 Comparison of metal concentrations in West Saronikos Gulf with other areas of Saronikos Gulf**

427 The concentrations of heavy metals in the surface sediments of West Saronikos Gulf from the present study are compared with
 428 data from the other sub-areas of Saronikos. Specifically the data reviewed are: a) from surface sediments collected in Elefsis
 429 Bay (EB), Inner Saronikos Gulf (ISG) and Outer Saronikos Gulf (OSG) during the same sampling of October 2017 and
 430 analyzed in the Laboratory of Environmental Chemistry using the same methodologies (Xarlis; 2018; Vrettou,2019) and b)
 431 data from surface sediments in Elefsis Bay (EB), Inner Saronikos Gulf (ISG) and Outer Saronikos Gulf (OSG) from various
 432 samplings conducted by the Hellenic Centre for Marine Research and analyzed by X-ray Fluorescence (Karageorgis et al.,
 433 2020; Karageorgis et al., 2020a). Summary results of this comparative review are provided in Table 9. The location of selected
 434 stations from the other sub-areas of Saronikos are roughly given in Fig. A17 (Appendix). Furthermore, metal concentrations
 435 in sediments from various areas of Greece analyzed by X-ray Fluorescence (Kanellopoulos et al., 2022) were also reviewed.
 436 It is observed that high Al concentrations are recorded at West Saronikos as well as the Outer Gulf, related to the settling of
 437 finer aluminosilicates in deeper waters. Moreover, high values of Al contents at Elefsis Bay (S1 and neighbouring stations,
 438 Fig. A17) are attributed to the terrigenous inputs from ephemeral streams discharging into the bay (Karageorgis et al., 2020).
 439 Cr and Ni contents show maxima in the northwestern stations (MOT13A, MOT16, MOT16A) offshore Susaki, due to the
 440 geological contribution from ophiolite complexes (Kelepertsis et al., 2001). Ni and Cr concentrations throughout the northwest

441 area of Saronikos are higher than the mean background contents (117 mg Kg⁻¹ for Ni and 142 mg Kg⁻¹ for Cr) estimated from
 442 sediments sampled at various areas of Greece and similar to concentrations reported for Aliveri Bay and the Asopos river
 443 basin, also related to occurrence of ultrabasic rocks (Kanellopoulos et al., 2022).
 444 The concentrations of Mn and Fe are similar throughout the sub-areas of Saronikos Gulf, with the exception of Epidavros
 445 basin, where the maximum values are recorded due to the settling of finer aluminosilicates in deeper waters and possibly the
 446 implications of bottom water hypoxia/anoxia already discussed in previous sections (Kontoyannis, 2010; Karageorgis et al.,
 447 2020).

448

450 **Table 9. Comparison of heavy metals concentrations in mg Kg⁻¹ between areas of Saronikos Gulf.**

Area / Stations	Al	Cr	Ni	Fe	Mn	Cu	Pb	Zn	References
West Saronikos									Present work
MOT13A, MOT16, MOT16A	5697- 27009	280-552	344-484	20740- 21191	422-578	13.5-22.6	20.0-30.3	44.1-52.1	
UN5, UN6, UN6A	32705- 43264	142-199	187-305	21838- 27301	570-954	26.3-36.7	38.4-52.9	73.8-92.1	
UN4	22702	123	123	16682	270	17.5	24.5	43.6	
UN11	54626	142	230	32177	3925	49.7	63.9	110	
Outer Saronikos	14000- 42000	81.5-133	29.5-106	6407- 27139	293- 1159	12.8-22.9	15.9-58.9	32.4-110	
Inner Saronikos	2900 37700	30.7-184	9.6-87.1	3235- 20795	61.5-442	7.9-68.5	8.2-73.6	15.2-170	Karageorgis et al., 2020a; Vrettou, 2019; Xarlis,2018
Psitatalia (S7)	26780	-	83.2	20623	239	103	102	251	Panagopoulou, 2018; Xarlis, 2018
Elefsis Bay	35000 61000	108-176	41.2-119	29721- 30499	282-579	32.1-137	61.0-171	188-521	Karageorgis et al., 2020a; Xarlis,2018

451

452 The so-called anthropogenic metals Cu, Pb and Zn show maxima at Elefsis Bay, while their concentrations at the other parts
 453 of Saronikos are reported lower. The levels at Epidavros basin (UN11) are the highest among the stations of West Saronikos,
 454 which can be attributed to the transport of pollutants from the eastern basin (Psyllidou-Giouranovits and Pavlidou, 1998;
 455 Dassenakis et al., 2003). However, the West Saronikos levels of Cu, Pb and Zn remain lower than polluted stations of Inner
 456 Saronikos, such as OS2 (near the port of Pireaus) and S7 (Psittalia, outfall of the Athens Waste Water Treatment Plant)
 457 presented depicted in Gig. A17. Furthermore, Cu, Pb and Zn concentrations at Megara and Epidavros basins are similar to
 458 those at surface sediments of Malliakos and Pagassitikos Gulf (Kanellopoulos et al., 2022) and the concentrations found in
 459 most stations of Inner and Outer Saronikos (Karageorgis et al., 2020; Panagopoulou,2018; Vrettou,2019; Xarlis,2018) but
 460 higher than less affected island areas such as Chios Port, Milos and Andros (Kanellopoulos et al., 2022). In general the West
 461 Saronikos sediments, affected by a relatively small industrial zone, are much less contaminated by Zn compared to the
 462 concentrations well above 150 mg Kg⁻¹ found in specific locations of Elefsis Bay, Inner Saronikos, Thessaloniki Bay, Ierissos
 463 and Lavrio (Karageorgis et al., 2020, Kanellopoulos et al., 2022) with extensive pollution sources such as the Elefsis industrial
 464 zone, Piraeus Port, the Athens Waste Water Treatment Plant outfall, major rivers of Northern Greece and current or historical
 465 mining operations.

466 6 Conclusions

467 The heavy metal pollution of West Saronikos Gulf has not been sufficiently studied, despite the scientific interest of this area,
 468 in contrast to the numerous studies of the eastern coast. The distribution of metals in the sediment samples of West Saronikos
 469 indicates that the area is enriched in metals from both geological and anthropogenic origins. The concentrations of all metals
 470 (Al, Mn, Cr, Ni, Cu, Pb, Zn) in muddy sediments are higher than those measured in sandy sediments. The cores are fairly
 471 homogeneous, in terms of carbonates and the down-core variability of % TOC, is characterized by high surficial values that
 472 decrease with depth.

473 The Cr and Ni concentrations at the northwest part of the study area are higher than those measured at the southwest area and
 474 their values are very stable with depth of most sediment cores, which can be attributed to the geological background of the
 475 adjacent coast. Al, Fe and Mn are increased from the northeast to the southwest part of the study area. The concentrations of
 476 Al and Fe increase with depth in most cores, while the values of Mn decrease with depth. Generally, concentrations of Fe and
 477 Mn at surface sediments are affected by oxic and hypoxic conditions and the settling of finer aluminosilicates.
 478 The vertical distributions of Cu, Pb and Zn present a constant decrease over depth along most cores, which can be attributed
 479 to their anthropogenic origin. Moreover, their levels at most sediments are higher than those measured ten years ago. Finally,
 480 the Cu, Pb, Zn concentrations in West Saronikos Gulf surface sediments are comparable with those at Outer Saronikos Gulf
 481 and lower than those from Inner Saronikos Gulf, Elefsis Bay and other pollution hot spots of Greece, which can be attributed
 482 to the smaller industrial zone and sparse urbanized settlements of the West Saronikos coast.
 483 The concentrations of metals that are measured higher than the ERL values and the indication for moderately or highly toxic
 484 sediments by the calculation of mERMq signify that more research is required, in order to investigate probable effects on the
 485 marine ecosystem. Continuous monitoring, updating of the results of the present study complemented with detailed
 486 geochemical analysis for major elements and minerals identification as well as core dating, metal speciation and study of
 487 bioaccumulation should be conducted, to assess the impacts of heavy metal pollution on the marine environment of West
 488 Saronikos Gulf.

489 7 Appendices

490 Appendix A

491 **Table A1. Quality control data for total metal analyses.**

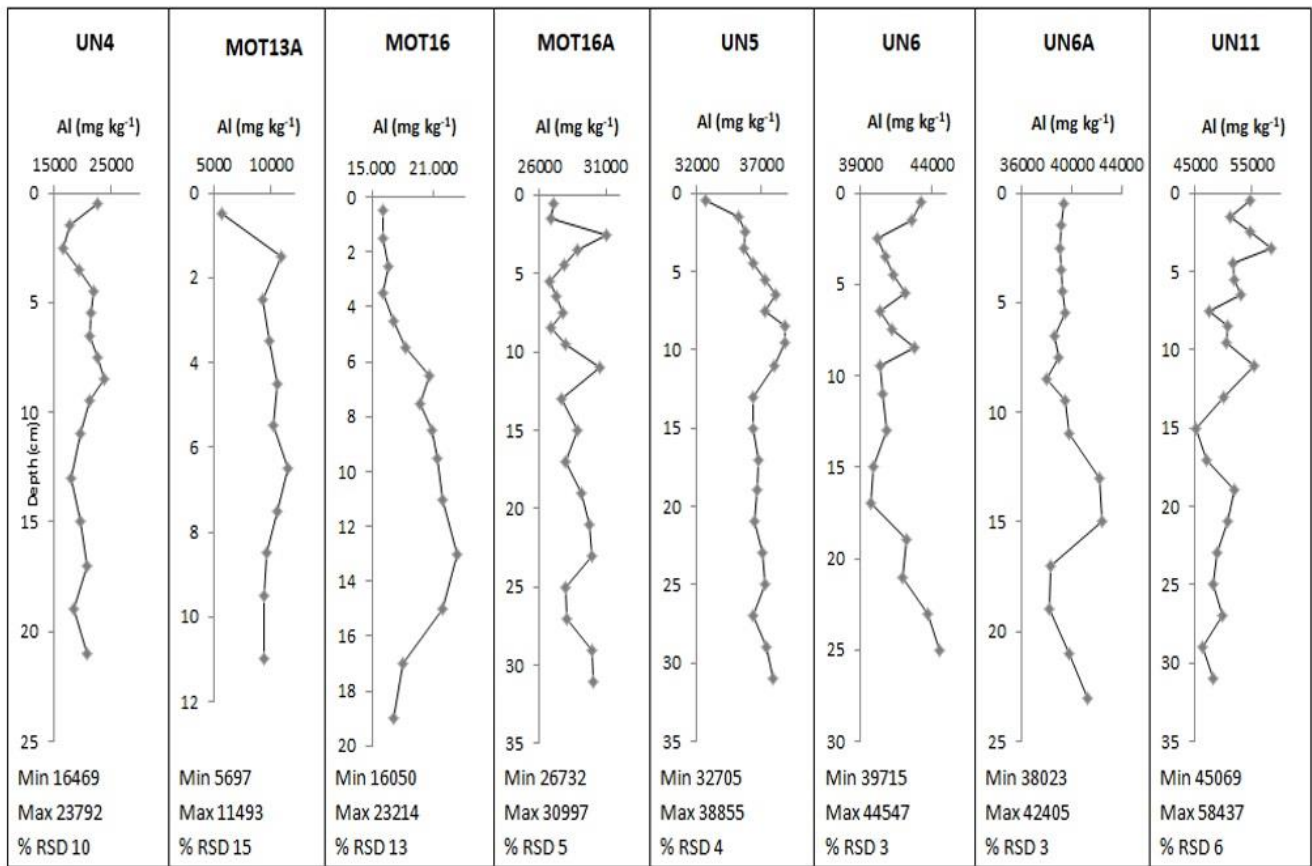
Quality parameter	Al	Cr	Cu	Fe	Mn	Ni	Pb	Zn
Daily duplicates %RSD (range)	0.8-4.6	1.2-6.2	0.4-6.0	0.4-5.1	0.2-6.6	0.8-9.2	1.1-8.2	0.5-4.0
Daily duplicates %RSD (average)	2.3	3.6	3.4	1.3	1.5	3.4	3.8	1.9
%RSD for PACS reproducibility (n=15)	6.2	8.3	3.2	2.6	7.4	9.7	6.2	4.0
Accuracy (% Recovery range)	85-106	84-114	87-96	86-95	85-113	85-115	86-109	86-99
Accuracy (%Recovery average)	96	96	91	89	99	101	95	91

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 494
 495
 496
 497

Table A2. The ratios of eight heavy metals to Al in surface and deeper layer sediments of the collected cores. In cases of coarse-grained sediment cores MOT13A, MOT16, UN4, the ratios in fine sediment fraction ($f < 63 \mu\text{m}$) are calculated.

Core	Layer (cm)	Cr 10^4 Al^{-1}	Ni 10^4 Al^{-1}	Fe Al^{-1}	Mn 10^4 Al^{-1}	Cu 10^4 Al^{-1}	Pb 10^4 Al^{-1}	Zn 10^4 Al^{-1}
MOT 13 A	0-1	616	389	1.99	451	17.3	24.7	46.9
	10-12	277	261	1.44	303	12.4	15.4	30.4
MOT 16A	0-1	104	139	0.86	214	8.35	7.55	17.9
	30-32	85.0	127	0.87	187	7.35	9.35	14.7
UN 5	0-1	61.0	93.1	0.78	194	8.39	13.1	22.8
	30-32	58.9	84.5	0.77	137	6.85	6.52	15.2
MOT 16	0-1	170	174	1.19	244	10.4	12.9	24.8
	18-20	146	168	1.25	203	7.11	3.81	17.0
UN6	0-1	32.7	58.4	0.63	220	8.48	12.2	21.3
	24-26	34.2	61.6	0.66	159	6.41	7.29	14.0
UN 6A	0-1	37.0	47.5	0.56	145	6.7	9.76	18.8
	22-24	39.0	50.6	0.58	124	6.14	9.38	14.2
UN 4	0-1	38.6	47.4	0.46	104	6.9	7.63	15.2
	20-22	35.7	50.5	0.54	115	4.75	3.80	12.7
UN11	0-1	26.1	42.0	0.59	719	9.1	11.7	20.2
	30-32	33.8	45.1	0.66	303	7.75	7.29	14.8

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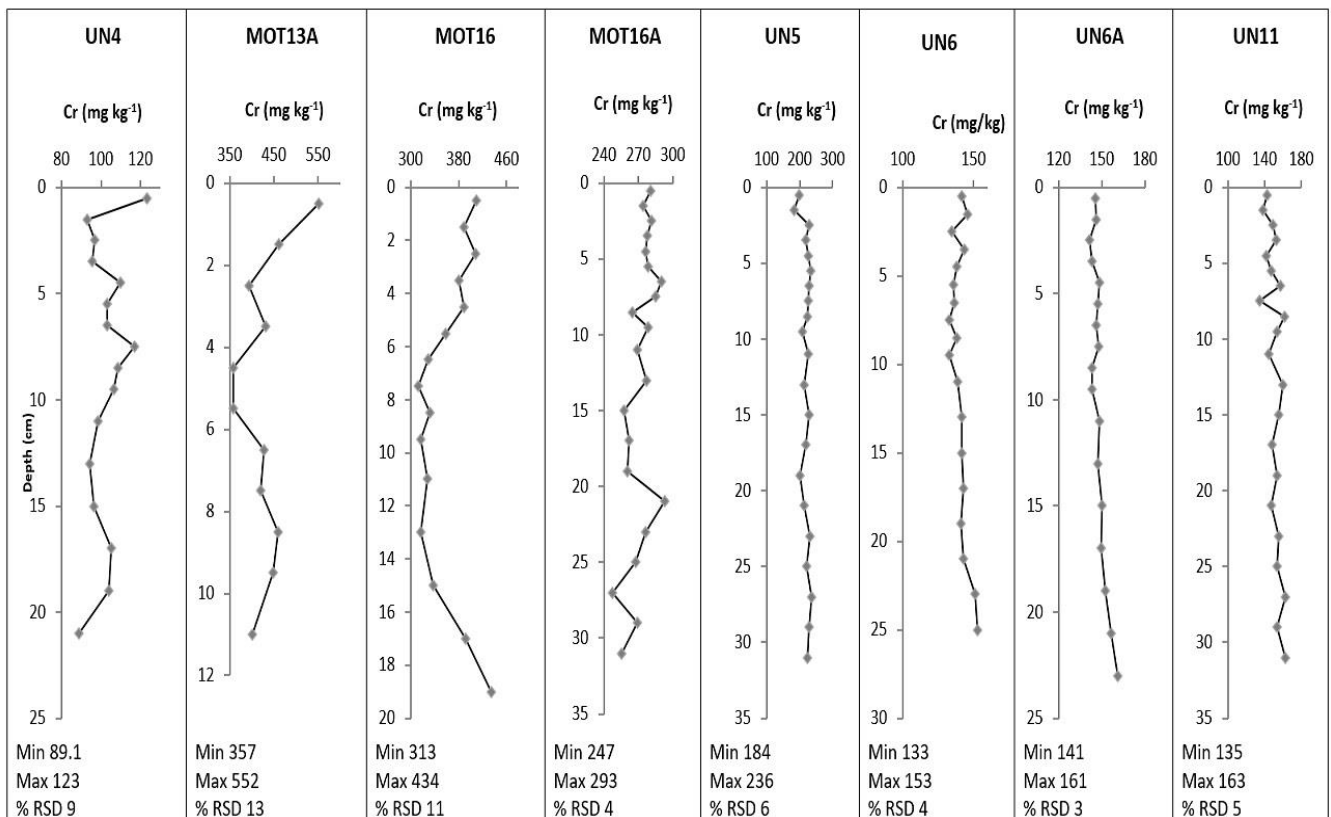


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Figure A1: Vertical distributions of Al in mg kg⁻¹ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction (< 1 mm).

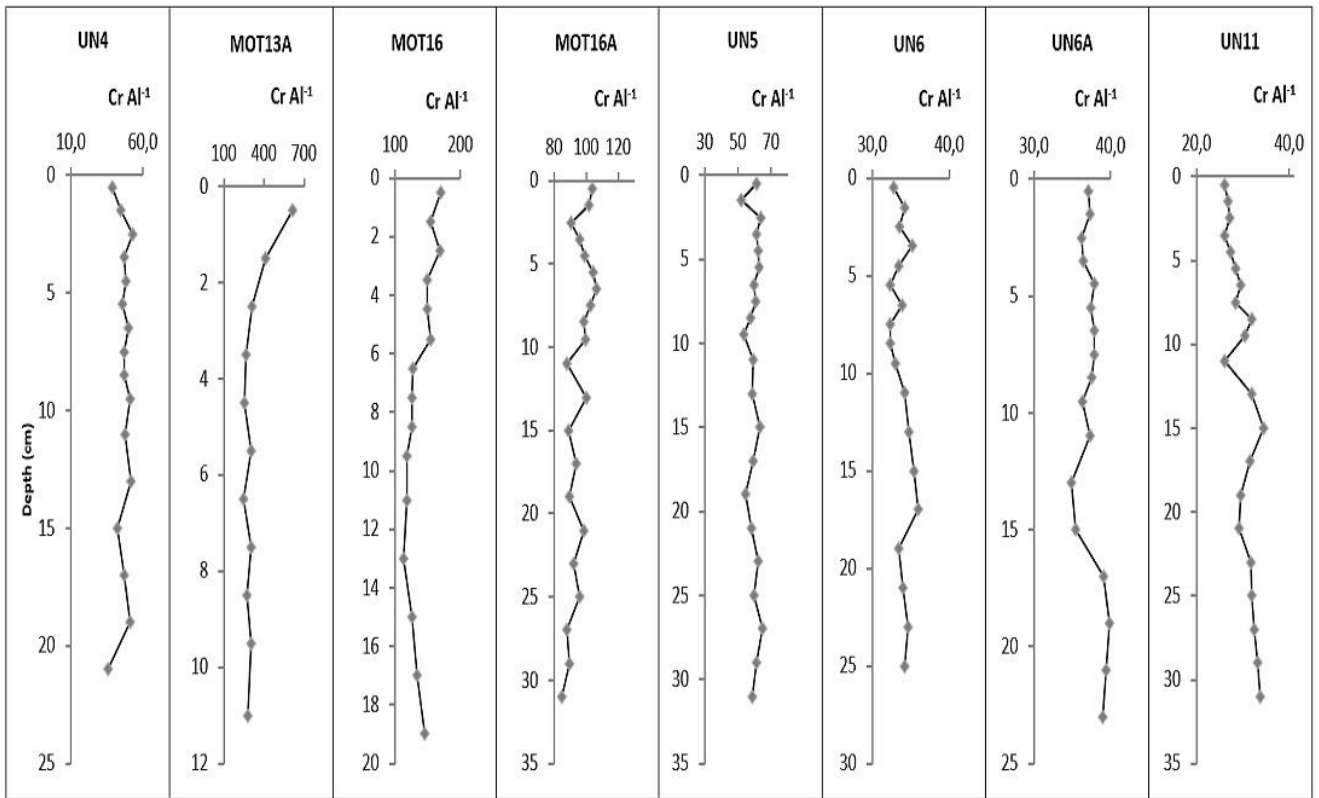


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Figure A2: Vertical distributions of Cr in mg kg⁻¹ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction (< 1 mm).

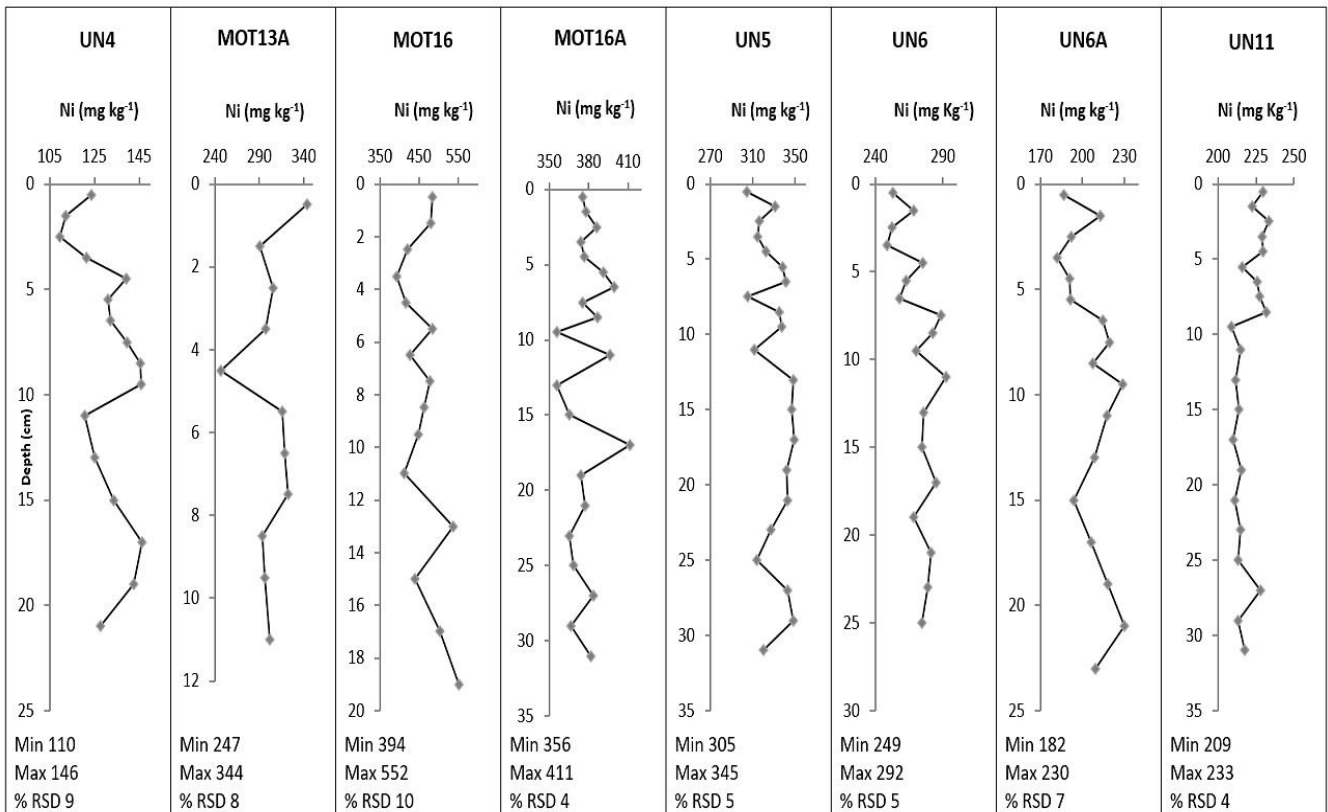


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Figure A3: Vertical distributions of $\text{Cr Al}^{-1} \times 10^4$ in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are calculated at the fine sediment fraction ($< 63 \mu\text{m}$).

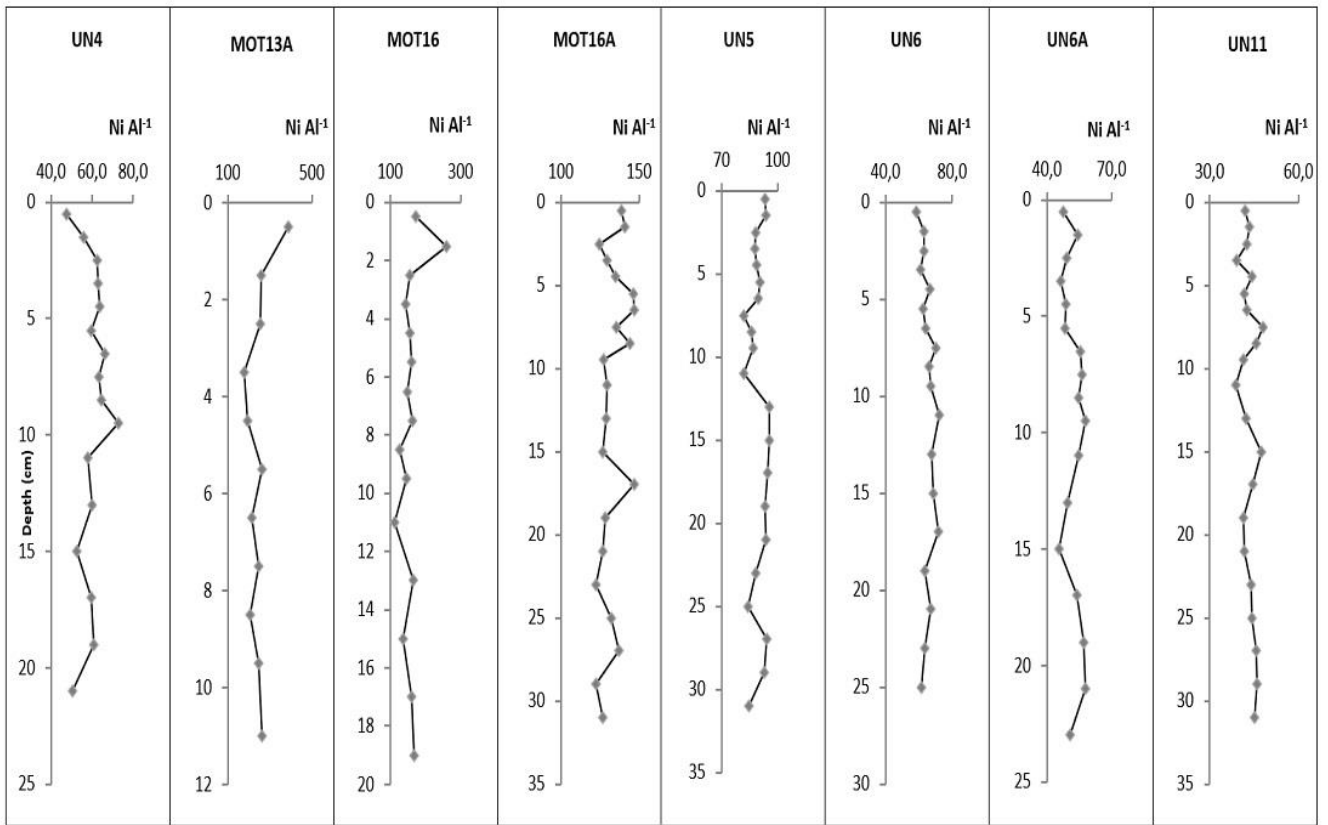


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Figure A4: Vertical distributions of Ni in mg kg^{-1} in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction ($< 1 \text{mm}$).

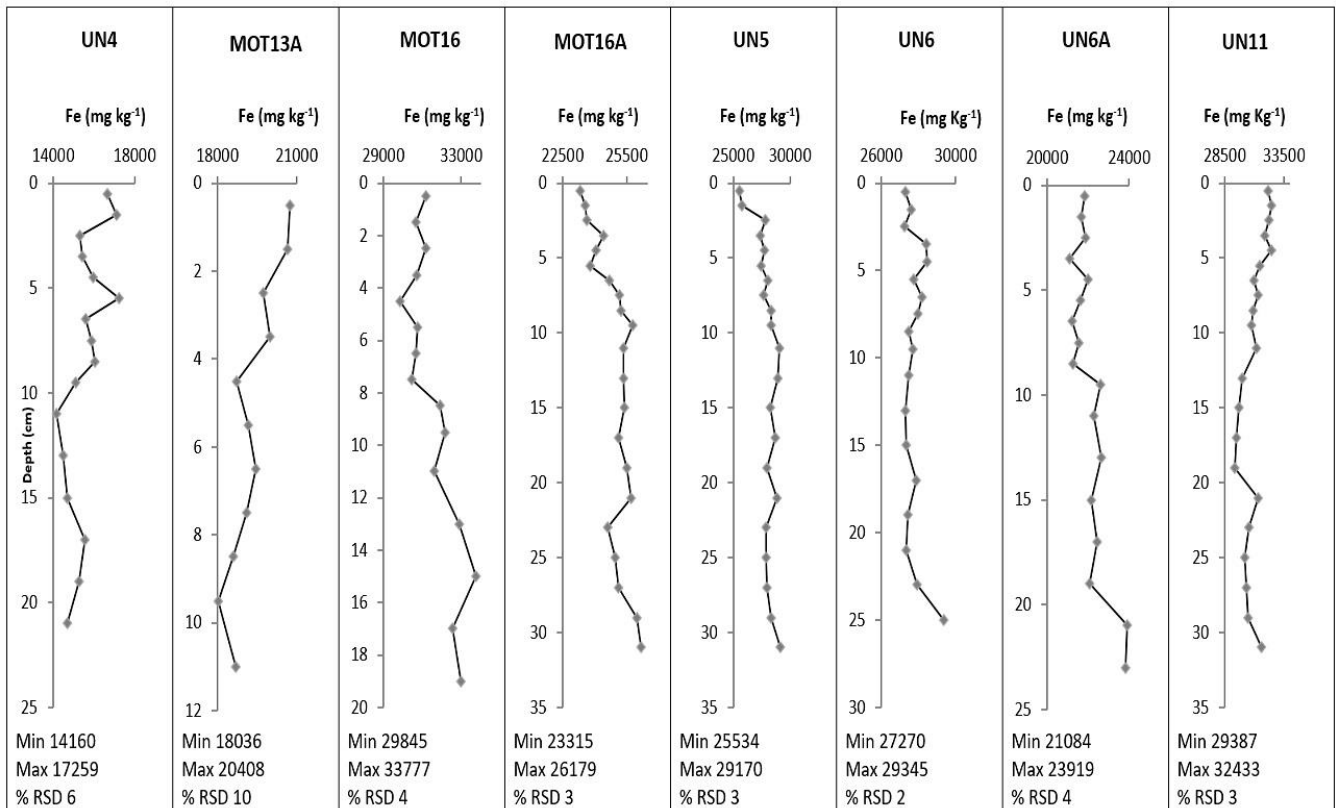


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Figure A5: Vertical distributions of Ni Al⁻¹ × (10⁴) in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are calculated at the fine sediment fraction (< 63 μm).

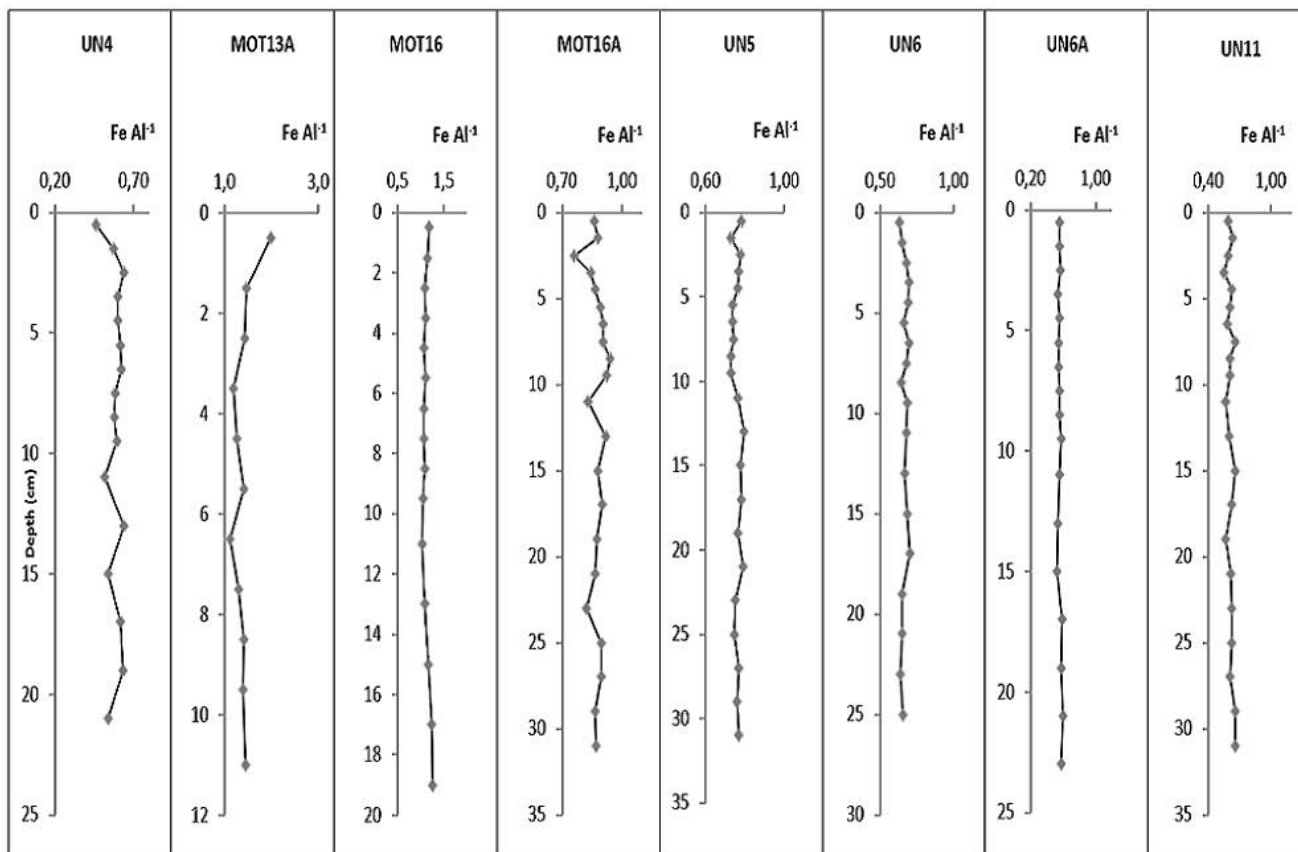


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Figure A6: Vertical distributions of Fe in mg kg⁻¹ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction (f < 1 mm).

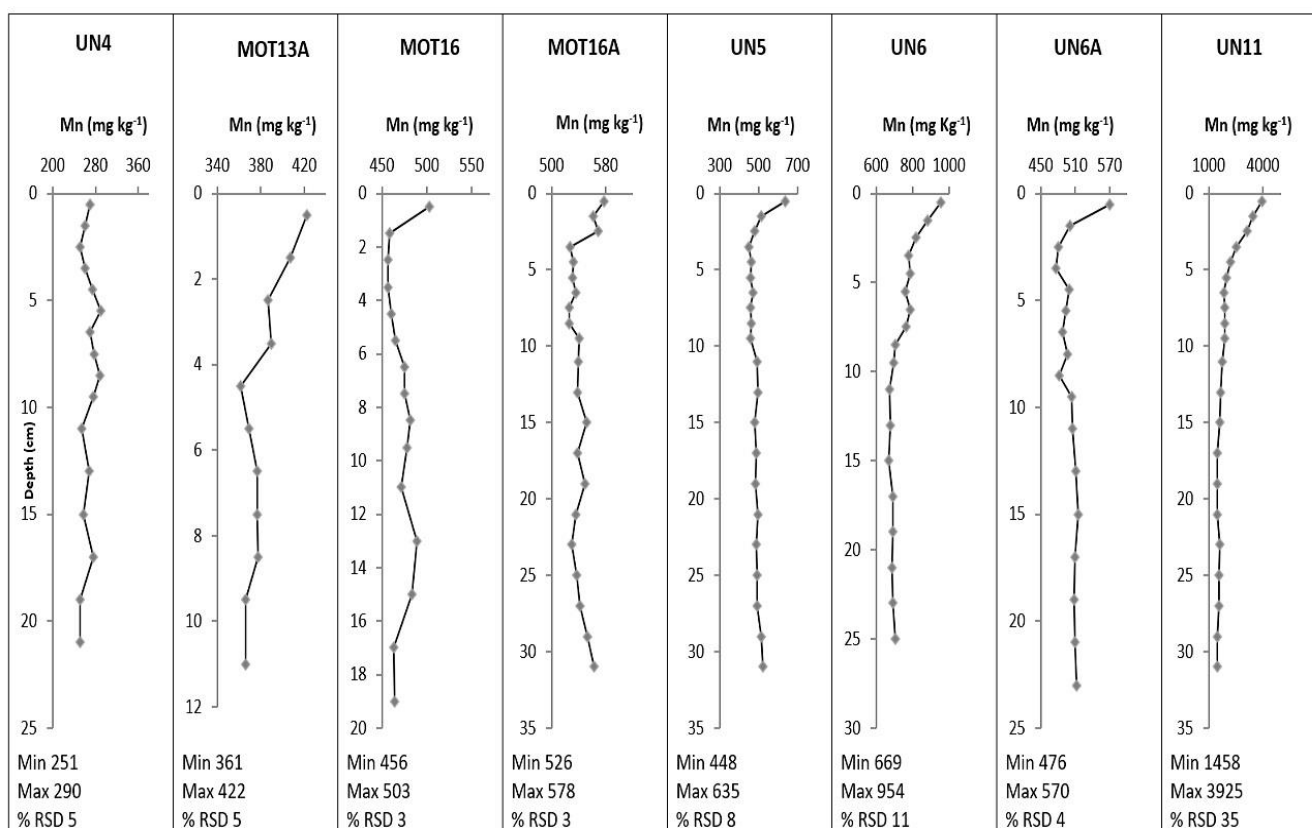


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Figure A7: Vertical distributions of Fe Al⁻¹ in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are calculated at the fine sediment fraction (< 63 μm).

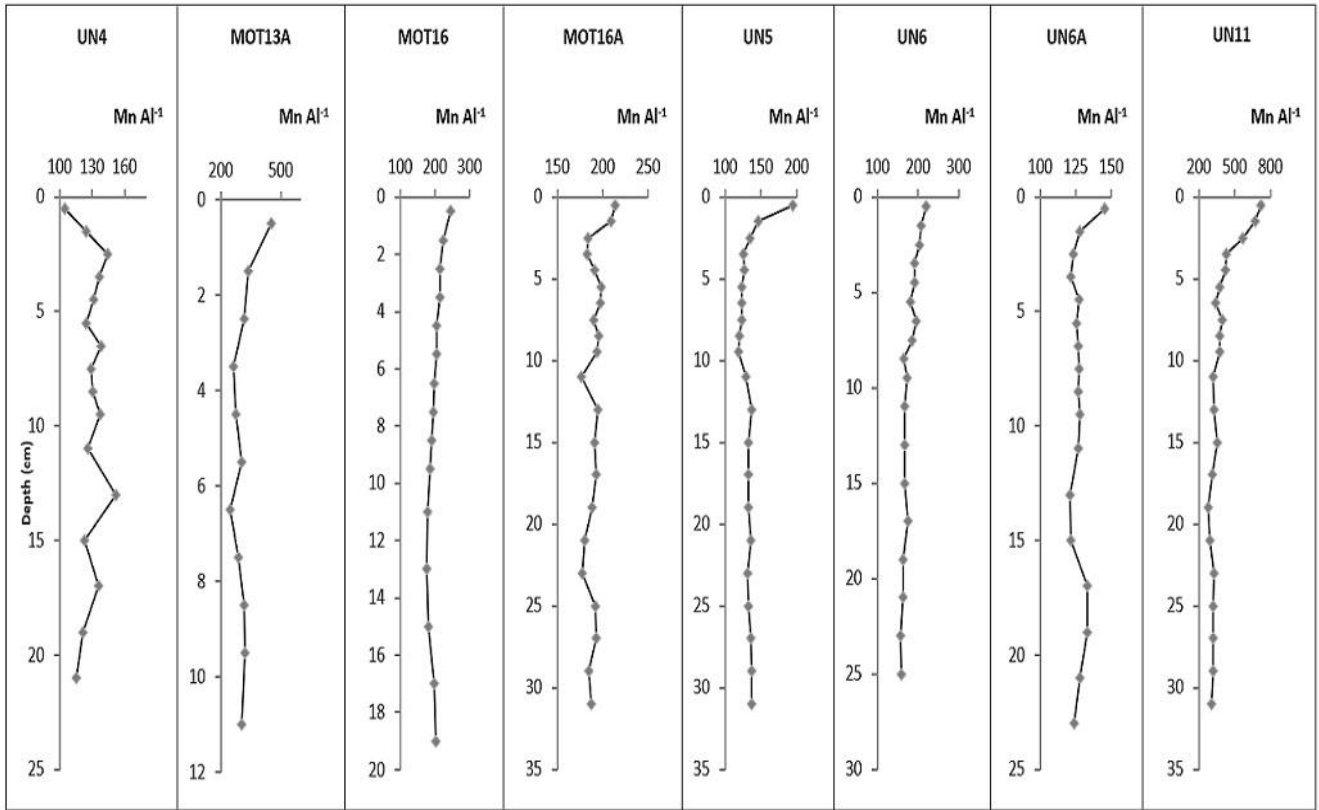


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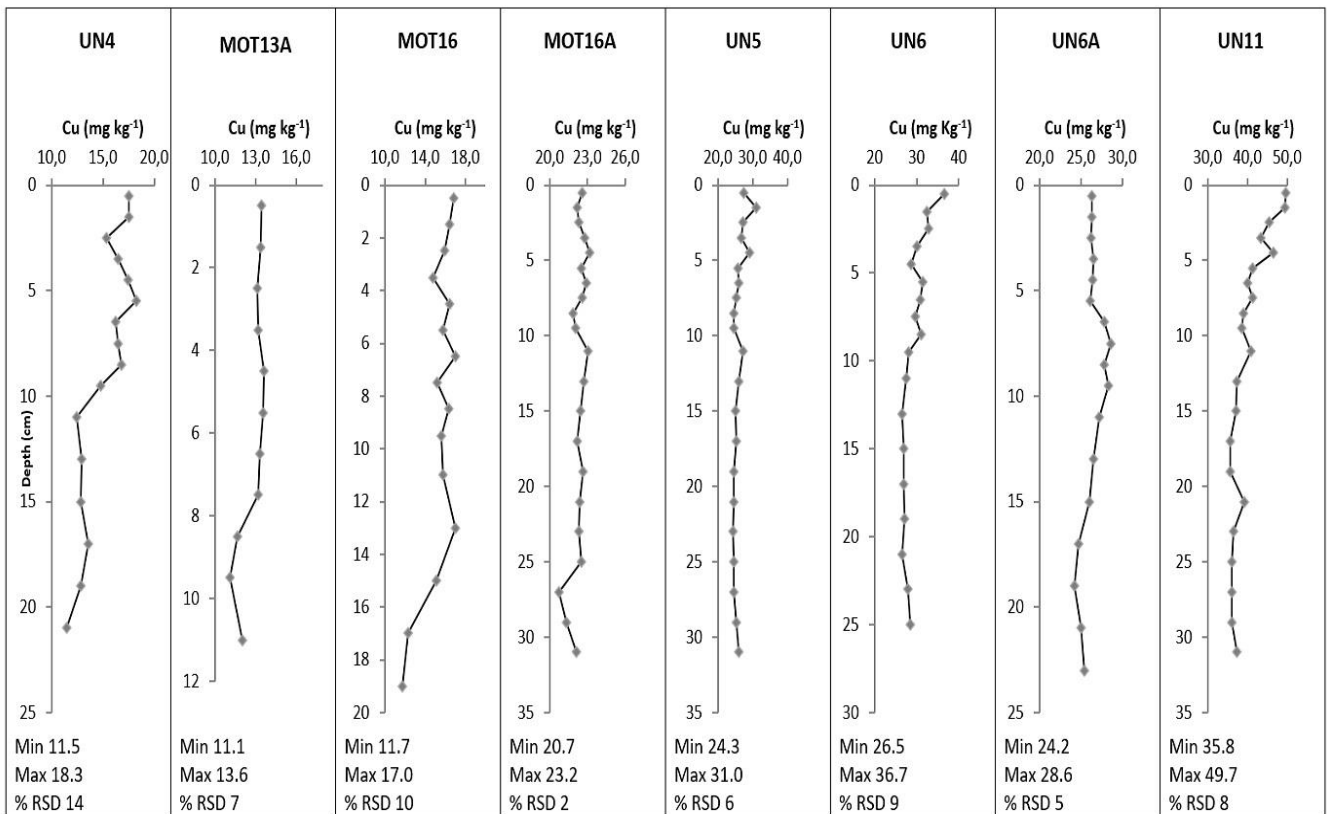
Figure A8: Vertical distributions of Mn in mg kg⁻¹ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction (< 1 mm).



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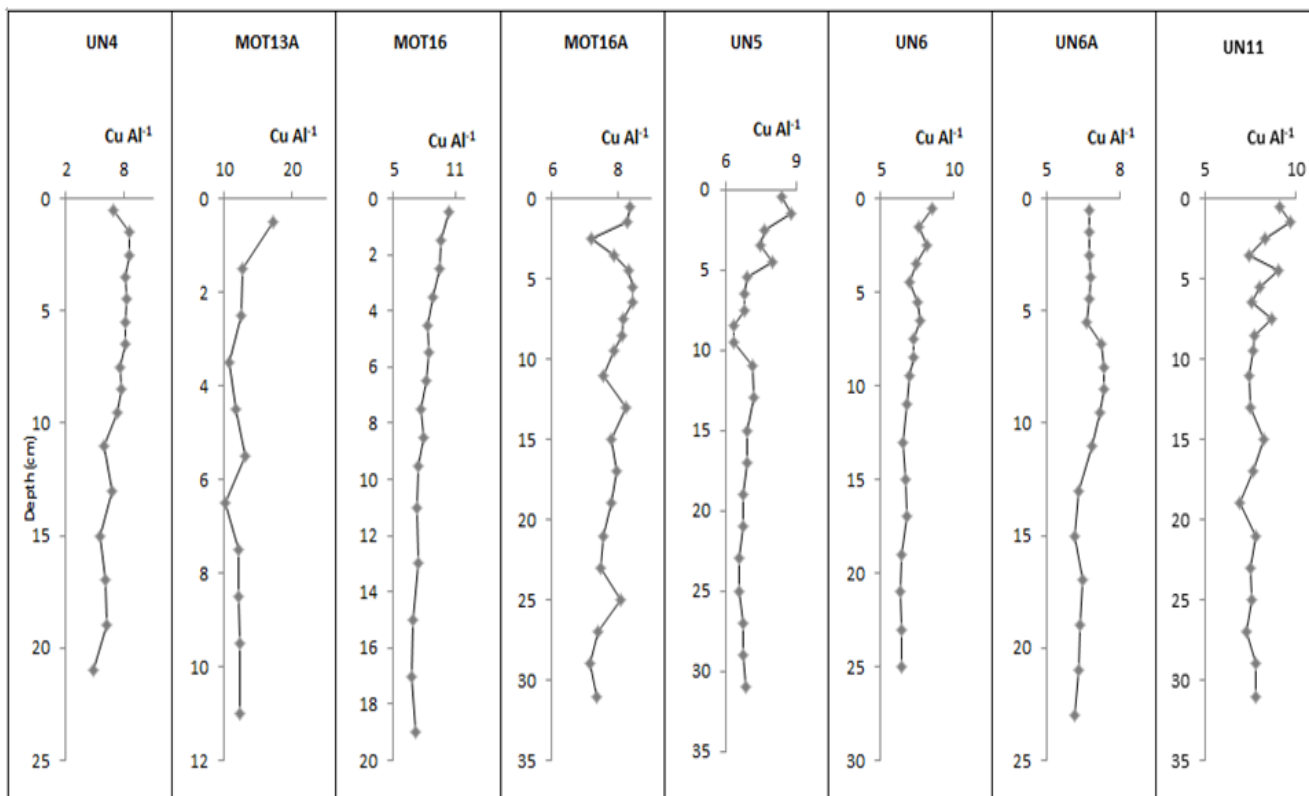
524 **Figure A9: Vertical distributions of Mn Al⁻¹ × (10⁴) in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are**
 525 **calculated at the fine sediment fraction (< 63 μm).**

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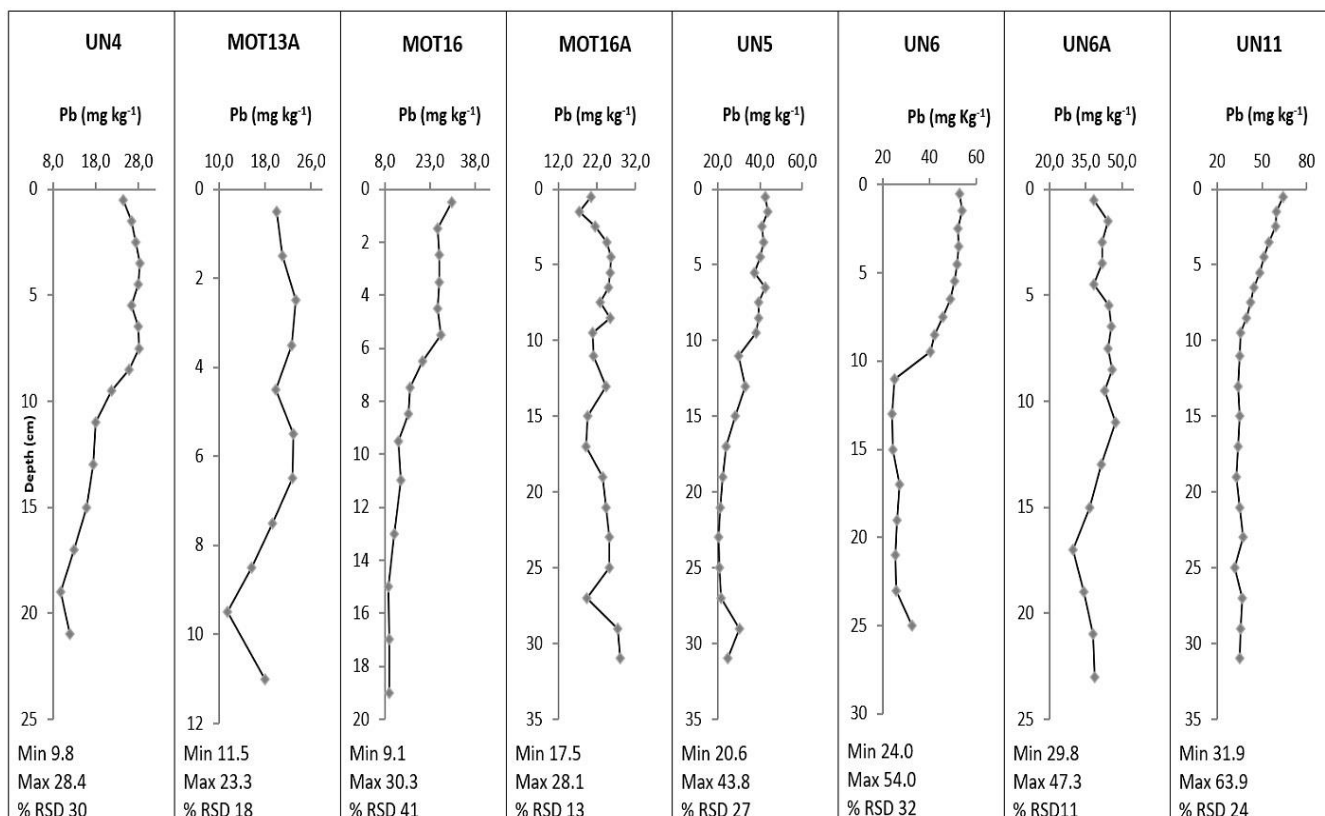
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528 **Figure A10: Vertical distributions of Cu in mg kg⁻¹ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4**
 529 **refer to the total sediment fraction (< 1 mm).**



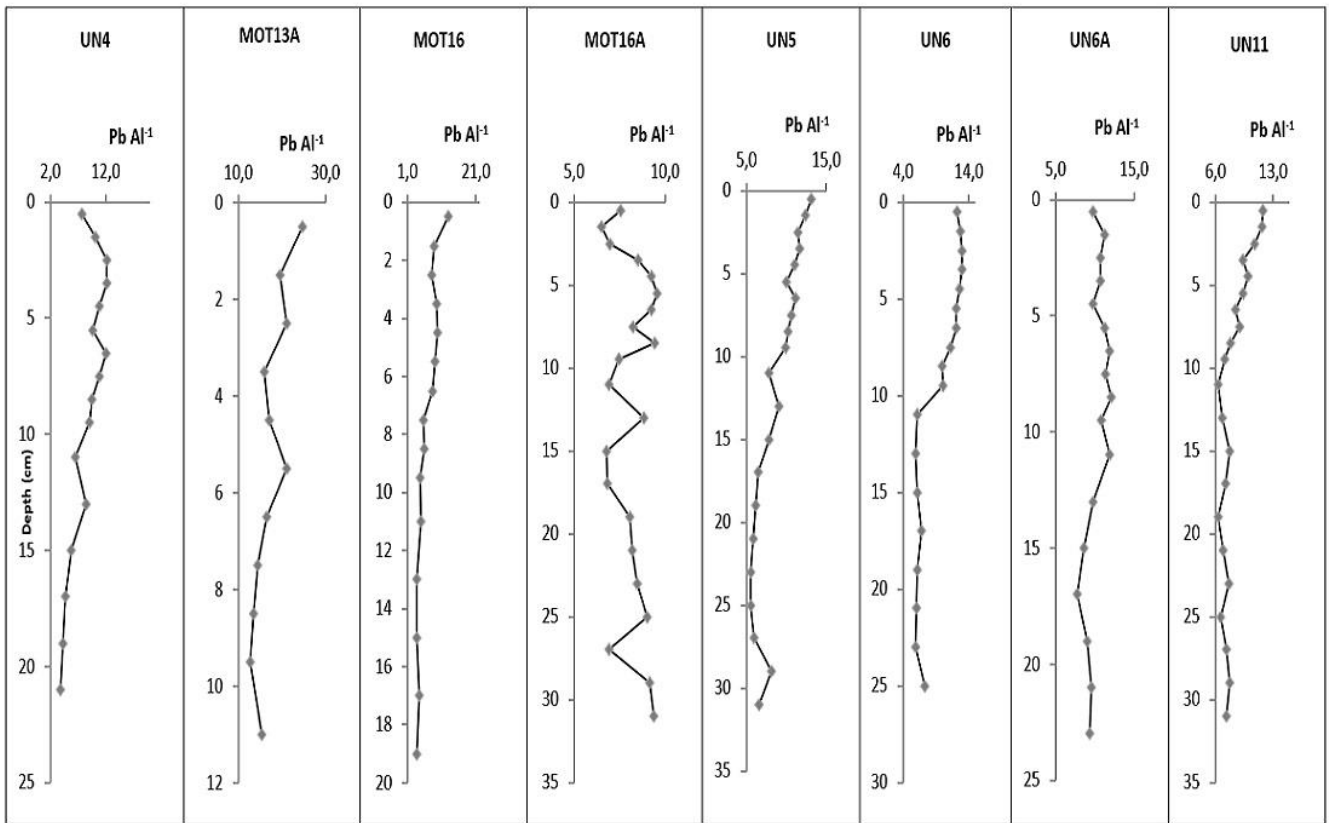
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Figure A11: Vertical distributions of $\text{Cu Al}^{-1} \times (10^4)$ in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are calculated at the fine sediment fraction ($< 63 \mu\text{m}$).



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Figure A12: Vertical distributions of Pb in mg kg^{-1} in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction ($< 1 \text{mm}$).

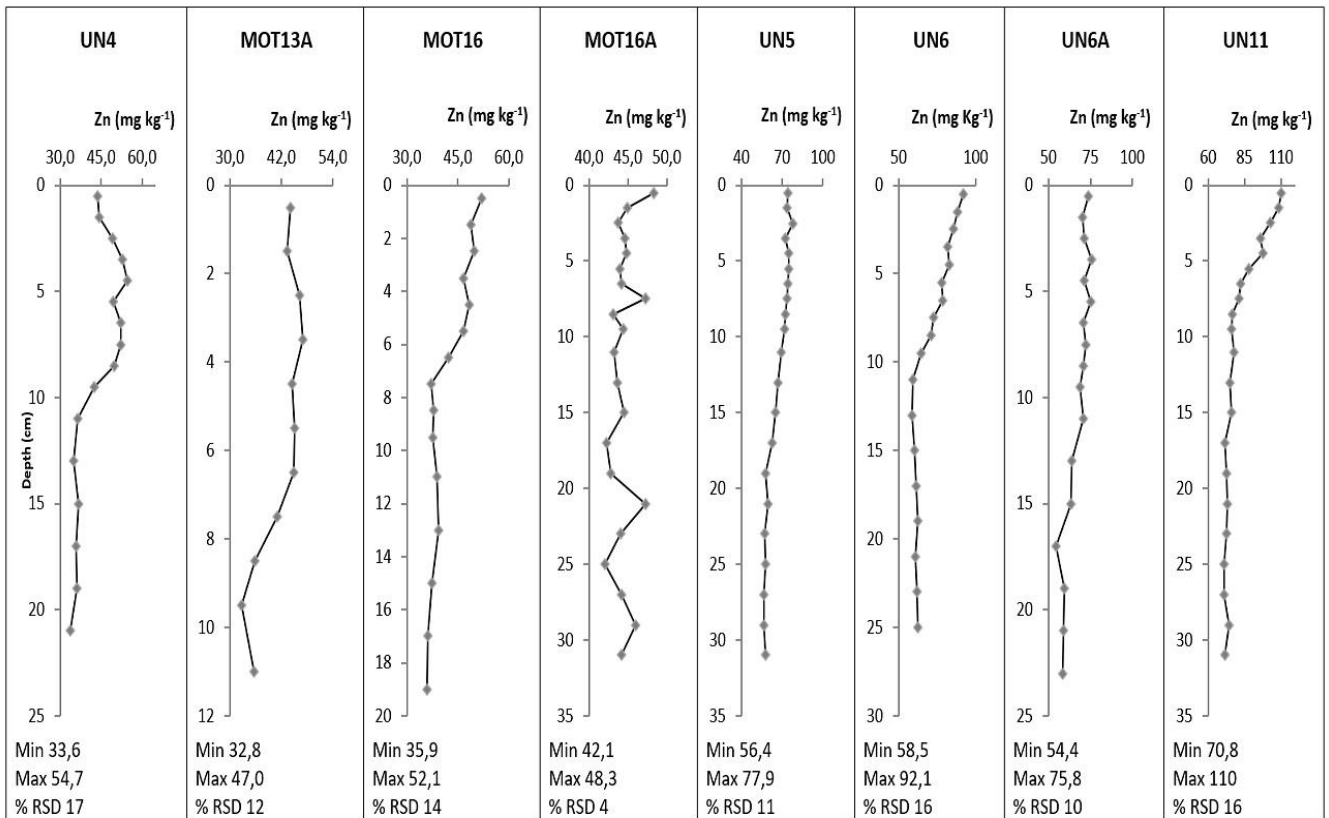


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Figure A13: Vertical distributions of $Pb Al^{-1} \times 10^4$ in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are calculated at the fine sediment fraction ($< 63 \mu m$).

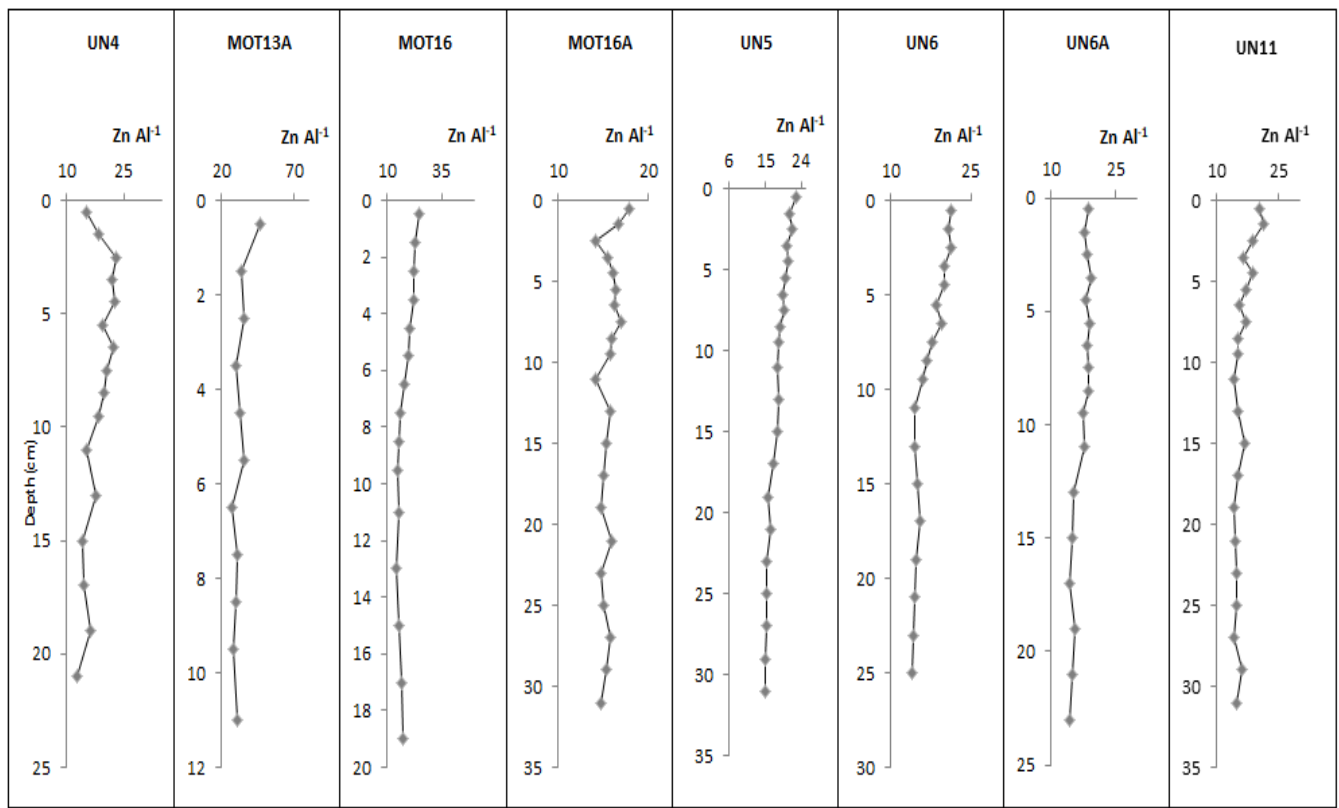


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Figure A14: Vertical distributions of Zn in $mg kg^{-1}$ in sediment cores. The concentrations in coarse cores MOT13A, MOT16, UN4 refer to the total sediment fraction ($< 1 mm$).



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544 **Figure A15: Vertical distributions of $Zn Al^{-1} \times (10^4)$ in sediment cores. The ratios in coarse cores MOT13A, MOT16, UN4 are**
 545 **calculated at the fine sediment fraction ($< 63 \mu m$).**

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547 **Table A3. Spearman's correlation coefficient matrix for Al ($mg Kg^{-1}$), Cr ($mg Kg^{-1}$), Ni ($mg Kg^{-1}$), Fe ($mg Kg^{-1}$), Mn ($mg Kg^{-1}$), Cu**
 548 **($mg Kg^{-1}$), Pb ($mg Kg^{-1}$), Zn ($mg Kg^{-1}$), TOC (% Total Organic Carbon), carbonates (% $CaCO_3$) (N=140 sediment samples).**

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Correlations										
Spearman's rho	Al	Cr	Ni	Fe	Mn	Cu	Pb	Zn	TOC	% $CaCO_3$
Al	1.000									
Cr	-0.521**	1.000								
Ni	-0.453**	0.841**	1.000							
Fe	0.624**	0.081	0.179*	1.000						
Mn	0.735**	-0.108	0.029	0.694**	1.000					
Cu	0.924**	-0.419**	-0.342**	0.633**	0.746**	1.000				
Pb	0.676**	-0.440**	-0.433**	0.244**	0.397**	0.779**	1.000			
Zn	0.790**	-0.452**	-0.441**	0.459**	0.479**	0.882**	0.894**	1.000		
TOC	0.244**	-0.198*	-0.272**	0.096	0.155	0.351**	0.428**	0.483**	1.000	
% $CaCO_3$	-0.472**	-0.215*	-0.222**	-0.766**	-0.597**	-0.566**	-0.358**	-0.516**	-0.282**	1.000

** . Correlation is significant at the 0.01 level (2-tailed).

* . Correlation is significant at the 0.05 level (2-tailed).

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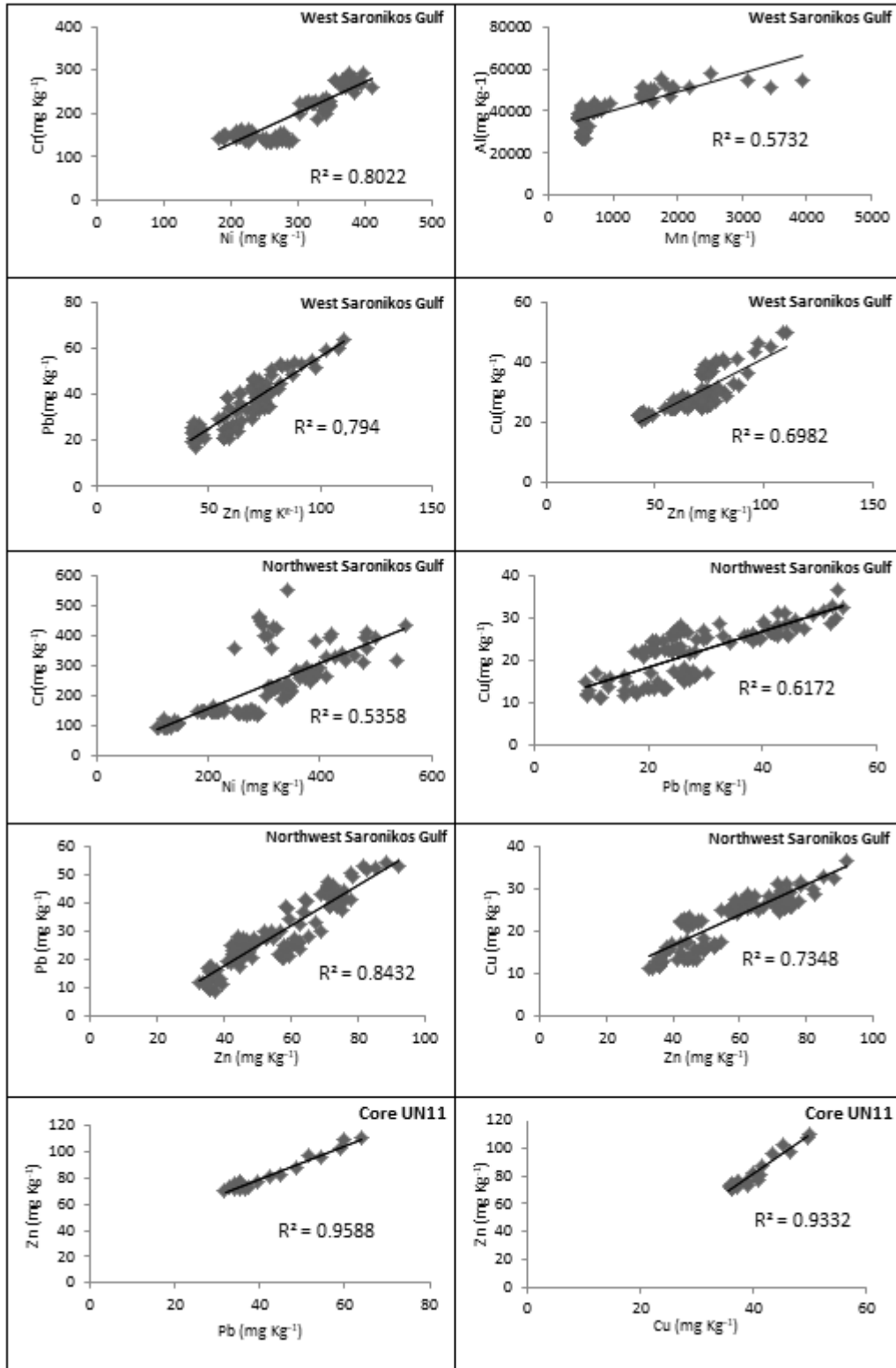
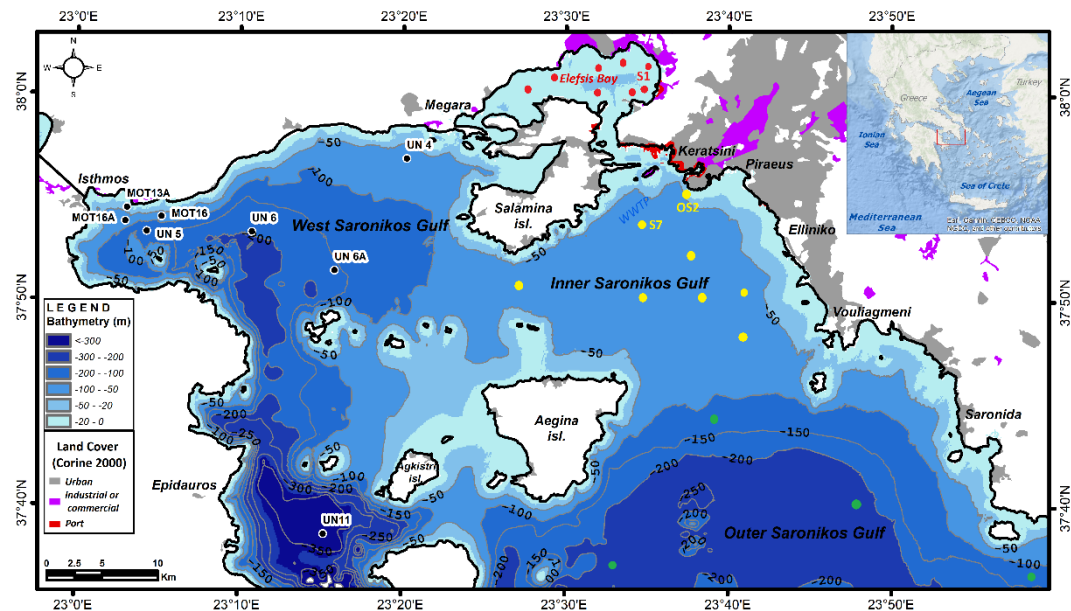


Figure A16: Correlations of heavy metals for the core samples of West Saronikos Gulf.

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578 **Figure A17:** Map of Saronikos with stations in Elefsis Bay (red circles), Inner Saronikos (yellow circles) and Outer Saronikos (green
 579 circles) from which metal concentrations in sediments were review for the comparative discussion to the present study results in
 580 West Saronikos (black circles).

581

582 8 Data availability

583 Datasets and their sources are fully detailed in the manuscript.

584

585 9 Team list

586 Not applicable.

587

588 10 Author contribution

589 Georgia Filippi and Vasiliki Paraskevopoulou conducted the chemical analyses in the Laboratory of Environmental Chemistry
 590 of the National and Kapodistrian University of Athens. Georgia Filippi wrote the paper, with contributions and reviews from
 591 all co-authors. Konstantinos Lazogiannis constructed the map of the study area and contributed to compilation and
 592 improvements of figures. Manos Dassenakis was the supervisor of the laboratory work and this article.

593

594 11 Competing interests

595 The authors declare that they have no conflict of interest.

596

597 12 Special issue statement

598 The statement on a corresponding special issue will be included by Copernicus, if applicable.

599

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 603 and National Funds (Hellenic General Secretariat for Research and Technology) in the framework of the project ARISTEIA
 604 I, 640 “Integrated Study of Trace Metals Biogeochemistry in the Coastal Marine Environment”, within the “Lifelong Learning
 605 Programme”. This project gave us the opportunity to reach the heavy metal pollution of West Saronikos Gulf. This paper
 606 summarizes the results of this study.

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